

1 Tracking the hidden niches: Movement-based
2 insights into northern lapwing intraspecific variation
3 and conservation

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25 ABSTRACT

26 Global monitoring data reveal farmland bird population declines, primarily driven by agricultural intensification, chemical
27 inputs, and climate shifts. The northern lapwing (*Vanellus vanellus*) is a farmland bird that exemplifies this decline
28 across Europe. The aim of this study is to quantify intraspecific variation in habitat selection and evaluate the degree
29 of behavioural variance in resource use among individuals and assess its conservation implications. We tested for: (1)
30 inter-individual differences in resource use; (2) repeatability of habitat preferences; and (3) niche components driving
31 selection patterns. Using Movebank GPS data from 13 individuals (2021-2023) across Europe, we applied step
32 selection functions (SSFs) incorporating biotic and abiotic covariates. Repeatability estimates were derived from mixed-
33 effects models to quantify consistent individual-level selection across resource gradients. We also estimated variation
34 of habitat selection across seasons as well as within- and between-individual variation. Movement data showed high
35 individual heterogeneity in several niche axes (notably glyphosate exposure and human population density), while
36 some axes such as soil nitrogen were relatively homogeneous. Step selection functions indicated that soil water index
37 and the earthworm abundance factor influenced movement, but no single environmental variable explained movement
38 patterns consistently across all individuals. Lapwings tended to move more under certain moon conditions and selected
39 pulses, flowers, vegetables and sugar beet with additional seasonal shifts in preferences. Repeatability analyses
40 suggested that individual consistency in crop selection was generally low, with the highest repeatability for pulses,
41 vegetables, flowers ($R=0.31$) and much lower or negligible for most other crops. Conservation strategies should focus
42 on maintaining low-intensity farmland and grasslands, particularly in remote areas. Predator control through culling is
43 ineffective due to rapid recolonization; instead, managing nesting habitats and removing predator perching structures
44 in highly used nesting sites is recommended. Establishing dedicated nesting areas (lapwing plots) can further support
45 the species survival.

46 Keywords

47 wader, spatial personality, multifunctional landscape, individualised niches, NDVI, tillage, mowing

48 Introduction

49 Farmland birds are avian species that breed and forage in, or close to, agricultural landscapes.
50 Adapted to open habitats, they likely benefited historically from the expansion and intensification
51 of agriculture. They contribute to a range of ecosystem services, including seed dispersal,
52 biological pest control, nutrient cycling, maintenance of habitat quality, and cultural services such
53 as recreation. Farmland bird populations are also widely used as indicators of broader biodiversity
54 in agricultural landscapes (Kirk et al., 2020).

55 Multiple long-term monitoring schemes have shown that most farmland bird species undergo
56 significant population declines. Declines of farmland birds have been documented in all continents
57 and in different ecological contexts and scales (from regional to global). Country-level declines
58 have been reported e.g. for Germany (Kamp et al., 2021), Denmark (Heldbjerg et al., 2018) and
59 Spain (Traba & Morales, 2019). Meanwhile, BirdLife Australia reported a 56% decline in farmland
60 bird populations since 1970. Similarly, the abundance of farmland birds in North America has
61 declined by 74% from 1966 to 2013 (Stanton et al., 2018). The 2024 report of the PanEuropean
62 Common Bird Monitoring Scheme found a decline of 60% of the Farmland Bird Index since 1980.
63 The main causes of the decrease of farmland birds have been identified as the intensification and
64 modernization of agricultural activities (Reif et al., 2024; Rosin et al., 2016, 2020), the application
65 of pesticides (Fuentes et al., 2023), illegal capture and hunting, urbanization (Tschardtke &
66 Batáry, 2023) and climate change. The European Union, in an attempt to mitigate farmland bird
67 declines, introduced the agri-environment schemes (AES) as a measure to increase biodiversity
68 in agricultural landscapes within the Common Agricultural Policy (Sharps et al., 2023).

69 Natural communities of plants and animals in human-modified landscapes have been shaped by
70 both agricultural activities and climate shifts (Ellis et al., 2021). European agricultural landscapes
71 have changed significantly in the past decades. Biodiversity has declined and landscape structure
72 has become more homogeneous (European Commission, 2021; European Environment Agency,

73 2025; Harpke et al., 2025; Jeanneret et al., 2021; PECBMS, 2024; van Swaay et al., 2025). The
74 last Common Agricultural Policies (CAP) have introduced the concepts of agro-ecosystems, agri-
75 environment schemes and multifunctional landscapes (Lefebvre et al., 2015; Taylor & Van
76 Grieken, 2015), which represent practical measures for the promotion of species diversity in
77 agricultural landscapes.

78 The northern lapwing (*Vanellus vanellus*) is a medium sized bird that feeds and breeds in open
79 agricultural habitats, such as lowland wet meadows, marshes and wet grasslands and open
80 habitats with low vegetation and high-water content. Although it is highly popular among the public
81 (NABU, 2024; Sheldon et al., 2004a), it is facing severe declines in Europe (PECBMS, 2024).
82 The monitoring scheme of common birds has found a 62% decline of northern lapwings across
83 Europe (PECBMS, 2024). The species is characterized as Near Threatened (NT) by the IUCN
84 and it is included in the Annex 2 of the Birds Directive, Annex III of the Berne Convention, and
85 Annex II of the Bonn Convention. At a national scale, though, its conservation status varies. For
86 example, in Germany it is under the Federal Species Conservation Regulation (BArtSchV), while
87 in France and Greece, despite being protected by EU law, northern lapwing hunting is permitted
88 (Souchay & Schaub, 2016).

89 Recent technological advancements in individual tracking devices led to the generation of large
90 datasets of very high spatiotemporal resolution. The combination of ecological niche theory with
91 technologically advanced individual-level observations represents a new frontier in conservation
92 biology and aligns with recent calls for more broad integration across organizational and spatial
93 scales (Jeltsch et al., 2025). The ecological niche of a population can be decomposed to a
94 between-individual and within-individual components (Roughgarden, 1972). Both components
95 are instrumental for the conservation of a species. For instance, tracking data allow the
96 exploration of individual specialization in space use, also known as spatial personalities (Stuber
97 et al., 2022) which in turn leads to individual differences in resource use, also known as
98 individualized niches (Takola & Schielzeth, 2022). The methodologies that are used to collect

99 movement data can be either through GPS tracking devices or through satellites or camera
100 tracking or telemetry. These data are then usually made available through platforms and web-
101 based interactive maps.

102

103 The aim of this paper is to quantitatively estimate intraspecific niche variation in northern lapwing
104 in order to inform the conservation of the species. The research questions of our study were: i)
105 Do individuals differ consistently in their resource use patterns? ii) What is the magnitude of the
106 between-individual variation? iii) What are the conservation implications of intraspecific variation
107 in resource use? To address these research questions, we analysed the individual resource use
108 by combining biotic (e.g. earthworms, human presence) and abiotic (e.g. weather, soil and
109 vegetation) drivers. We used step selection functions to quantify how the aforementioned
110 variables affect movement behaviour of each individual, offering a mechanistic understanding of
111 resource use and niche specialisation at the individual level. Subsequently, we used the results
112 of the step selection functions to inform the exploration of consistent behavioural differences
113 through repeatability models. Mechanistic approaches of intraspecific variation can shed light into
114 community dynamics and interspecific interactions (Moran et al., 2022).

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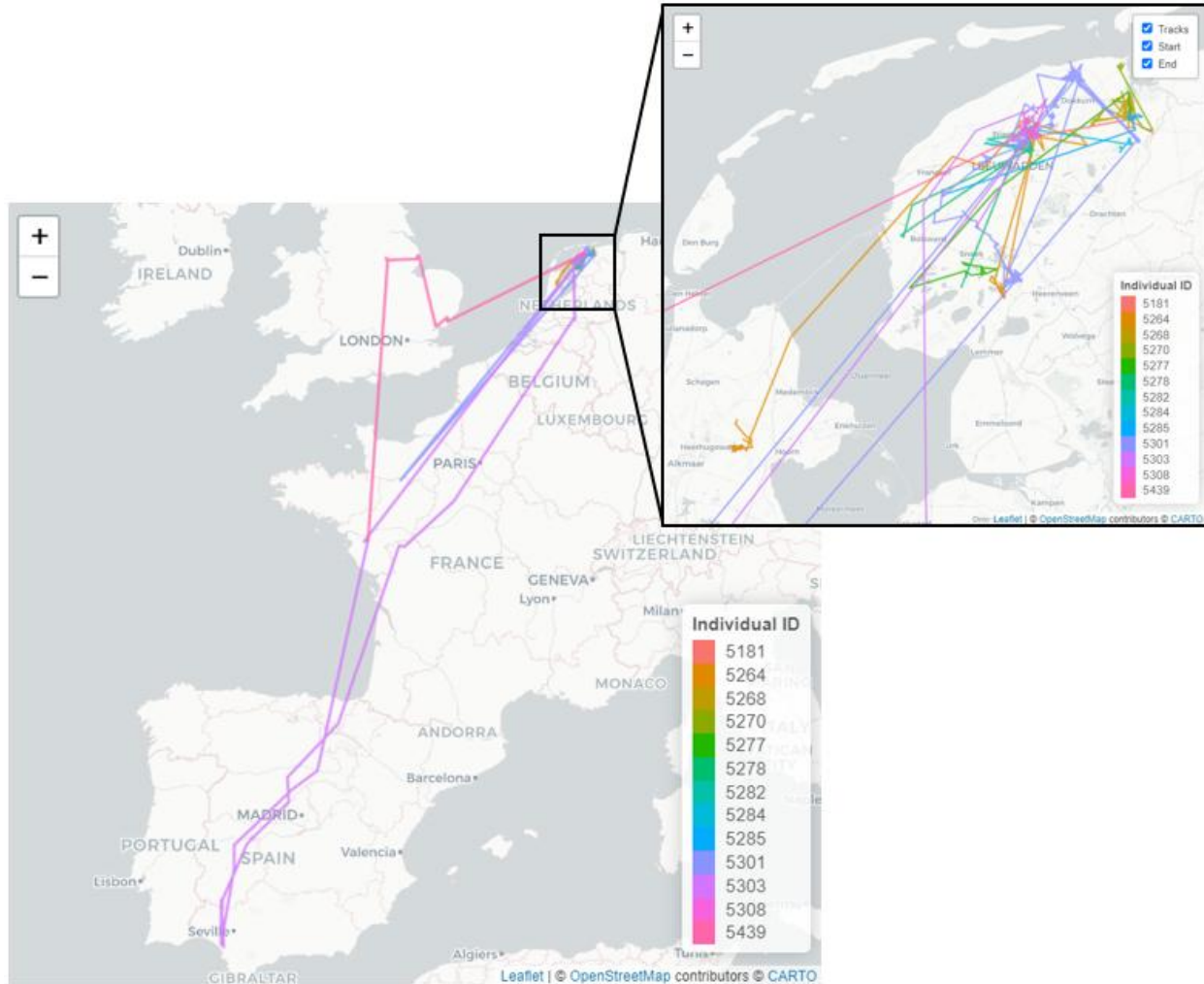
116 **Methods**

117 **Study area and tracking data**

118 We analysed all northern lapwing movement data that are publicly available for Europe through
119 the Movebank database (dataset ID: 1448409403, curator: Jelle Loonstra). The data covered an
120 area across the Netherlands, Belgium, France, and Spain (Fig. 1) and the tracking period
121 extended from April 2021 to March 2023. The temporal resolution ranged from 0.08 minutes to

122 5.6 hours. In order to standardize sampling effort across all individuals, we filtered the points with
123 a sampling rate of 180 minutes, considering ± 30 seconds of tolerance. The final dataset contained
124 a total of 14,701 data points from 13 individuals, since one was excluded after resampling. The
125 number of observations per individual ranged from 18 to 1,139 steps.

126 For each recorded position (hereafter used step), we generated a set of 10 random alternative
127 available positions (hereafter available steps) that the individual could possibly move towards
128 (Fig. S1). Each available random point was at most 10 km away from the used one, as a step
129 length of over 10 km was deemed unrealistic. In the case where the birds appeared to remain in
130 the same location for multiple time intervals (inactive or sleeping), the generation of random steps
131 was not possible, thus, we added jitter (0,001 m) to each repeated point. We removed all points
132 (used or available) that were over the sea and England.



133

134 Figure 1. Extend of the study area and GPS recorded positions of all individuals. An interactive
 135 version of this plot is provided in the Supplementary files.

136 Environmental data cube compilation

137 We compiled a set of environmental variables hypothesized to influence the movement patterns
 138 of northern lapwings and to characterize the ecological niche of the species. These variables
 139 included abiotic factors (air temperature at 2 meters above the ground, precipitation, wind speed,
 140 distance and phase of the moon, Normalized Difference Vegetation Index, crop type), soil factors
 141 (soil bulk density, soil water index, soil organic carbon, soil organic carbon stock, soil nitrogen),
 142 biotic factors (human population density, human footprint, and earthworm abundance and

143 richness) and landscape management information (cropping intensity, pesticide application rates)
144 (Table S1). For each environmental variable, we used the most recent data available, ensuring
145 that temporal alignment was maintained as closely as possible. In cases where exact temporal
146 matching was not possible, we incorporated environmental data from the closest year, under the
147 assumption that these variables had not changed significantly over time.

148 **Abiotic factors**

149 We extracted ERA5-Land wind data and aligned daily temperature and precipitation values to
150 each used and available observation using the associated timestamp for the corresponding hour
151 of the day. We obtained crop type data from the EU Crop Maps for 2018 and 2022, using the
152 2018 data to fill in gaps on the 2022 map, particularly in the northern part of the study area. We
153 removed all crops that appeared 30 times or less in the dataset.

154 We modelled night-time activity using the distance and phase of the moon for each used and
155 available position, calculated with the lunar R package (Lazaridis, 2022), considering the
156 respective timestamp of each observation. NDVI data were extracted for each used and available
157 location from the MODIS Terra Vegetation Indices (250 m resolution) dataset corresponding to
158 the closest available date.

159 **Soil factors**

160 Soil carbon, carbon stock, bulk density and nitrogen content data were extracted for each used
161 and available location from the SoilGrids dataset (Poggio et al., 2021).

162 **Biotic factors (human population density, human footprint, and earthworm 163 abundance and richness)**

164 We used human population density data from Eurostat for NUTS3 regions (Eurostat, 2025) along
165 with the Global Terrestrial Human Footprint dataset (Mu et al., 2022), which relies on more
166 permanent and rigid factors, such as infrastructure, and night-time lights, which remain relatively

167 stable over time. The earthworm abundance and richness factors (Orgiazzi & Panagos, 2018),
168 were derived from a study on earthworm richness and abundance (Rutgers et al., 2016) and are
169 available upon request from the European Soil Data Centre. Due to the calculation method of the
170 earthworm factor, negative values indicate high earthworm abundance or richness.

171 Landscape management data (cropping intensity, pesticide application 172 rates)

173 Cropping intensity was extracted from a fine resolution global dataset (Zhang et al., 2021), while
174 pesticide application rates were extracted from a European database (Porta et al., 2025).

175 We assumed that cropping cycles remain stable over short time periods, thus we used the GCI30
176 map from 2020 on the data of 2023. Similarly, we assumed that pesticide application rates
177 remained stable between 2018 and 2021.

178

179 Data analysis

180 To address collinearity among the predictor variables, we calculated correlations for every pair of
181 variables and an auxiliary regression with the numerical variables to estimate the Variance
182 Inflation Factor (VIF) for every variable separately (Fig. S2, Table S3). We then built a model for
183 each predictor group (food, weather, soil, humans) to examine intraspecific variation using step
184 selection functions. Inter-individual variation in use patterns of vegetation types was examined
185 with the aid of repeatability models. All analyses were performed in R and all variables were
186 scaled to a mean of 0 and standard deviation of 1.

187 Individual variation in niche axes

188 To explore the intraspecific variation among the birds in our sample, we adopted the following
189 workflow. First, we used individual resource utilisation curves (visualised as ridgeline plots and
190 heatmaps) to visually examine within- and between-individual variation for each variable. We then

191 implemented a step-selection function for each individual with the amt R package (Signer, 2018)
192 to estimate the importance of each variable. We fitted the step selection functions with a binary
193 response variable (used vs available), step ID as a random effect and environmental variables,
194 turning angle and step length as fixed effects. One individual (ID: 5284) was excluded from the
195 individual-based SSFs due to insufficient data, but included in all other analyses. Individuals 5308
196 and 5270 were excluded for the same reason from the weather models of individual-based SSFs.
197 In the case of crop type selection, where the individual-level subsets did not have enough data
198 for individual-level SSDs, we fitted an SSF for all individuals, using *Grassland* as the reference
199 level (because it was the most commonly used crop type) and an interaction between crop type
200 and temperature. We fitted this step selection function with a binary response variable, individual-
201 level effects for temperature, while crop types, turning angle and step length were used as fixed
202 effects. Based on the results of the SSF on crop types, we identified the crop types that have a
203 significant effect on the movement of northern lapwings and subsequently estimated the
204 magnitude of individual differences on crop type selection through the repeatability of usage for
205 each (Stoffel et al., 2017).

206 To analyse the temporal variation in activity, we used two approaches: nocturnal activity (since
207 some birds tend to be active also at night) and seasonal variation in resource use. The nocturnal
208 activity of northern lapwings was modelled using only the used locations (i.e. the original points
209 of the GPS tags). The response variable in the night models was step length, transformed as
210 $\log(\text{step length} + 1)$. First, we fitted a linear mixed-effects model to test whether nocturnal step
211 length varied with moon distance, including moon distance and moon phase as fixed effects and
212 allowing both the intercept and the slope of moon distance to vary among individuals. To visualise
213 between-individual heterogeneity, we generated individual-specific predicted relationships
214 between moon distance and $\log(\text{step length} + 1)$ (with moon phase held constant at new moon)
215 and contrasted these with the population-level prediction based on fixed effects only. Second, we
216 assessed whether individuals differed in their responses across moon phases by fitting a mixed

217 model with moon phase as a fixed effect and individual-specific random effects for each phase
218 (0+moon_phase|id), while holding moon distance constant at its centred mean (moon_dist = 0)
219 for prediction. We also used step selection functions to explore the crop type selection during
220 night time. Seasonal variability in the drivers of northern lapwing movements was examined by
221 fitting step selection functions for all individuals (due to inadequate size of individual-level
222 subsets), but with random slopes in order to allow for different responses.

223 Repeatability of resource selection

224 Repeatability is defined as the proportion of population-level variance that can be attributed to the
225 between-individual differences. Repeatability is an indicator of different phenotypes (or
226 personalities) within a population. A repeatability of 0 indicates that a trait is not consistent across
227 observations, whereas a repeatability of 1 indicates that an individual trait remains consistent
228 across measurements. In the context of habitat selection, high repeatability indicates that the
229 individuals' habitat preferences are consistent over time and across contexts (indicating spatial
230 'personalities') and low repeatability means that individual preferences vary substantially.
231 Repeatability models included individual ID as a random effect and a crop type as a fixed effect.

232 Results

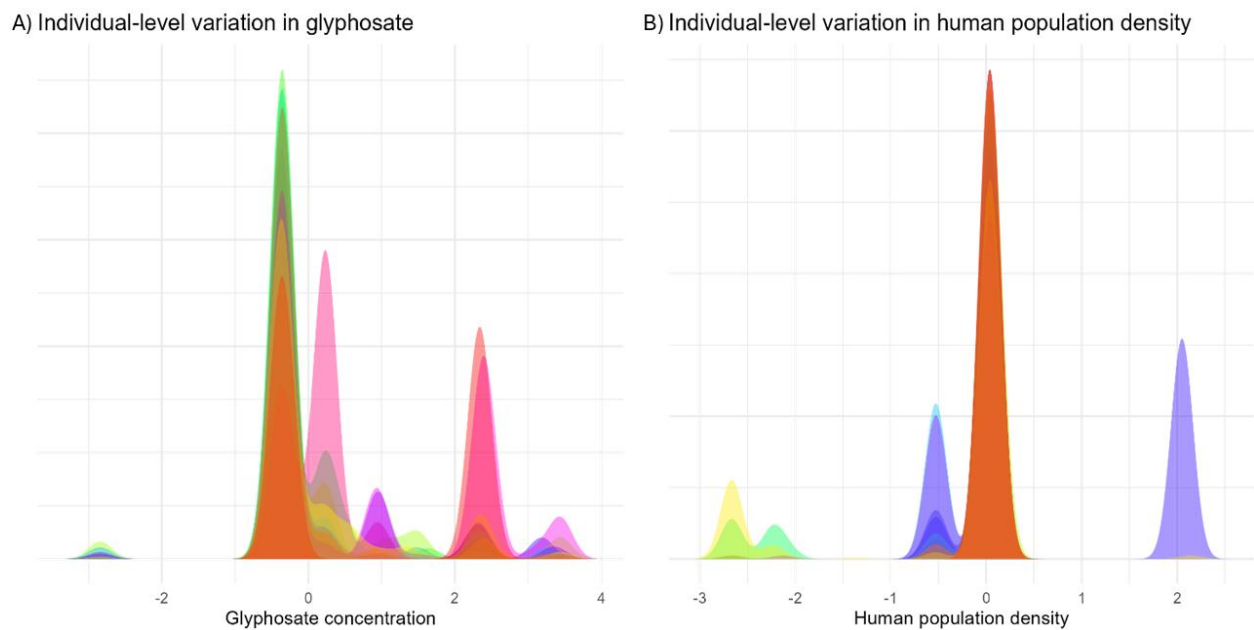
233 Data description

234 Our dataset included information on the movement of 13 northern lapwings, over the course of 2
235 years (2021-2023) in western Europe. We used GPS tracking data to derive used and available
236 steps (n = 40,777 points) and remote sensing techniques to characterize the environmental space
237 selected by the individuals.

238 **Individual variation in niche axes**

239 We detected both between-individual and within-individual variation in resource use patterns,
240 through a visual inspection of individual resource utilisation curves and heatmaps. For example,
241 individuals showed different patterns in their exposure to herbicides (glyphosates) and human
242 population density (Fig. 2). In other cases, such as the soil nitrogen, the patterns were more
243 homogeneous (Fig. S5).

244



245

246 Figure 2. Individual resource utilisation curves for glyphosate (A) and human population density
247 (B).

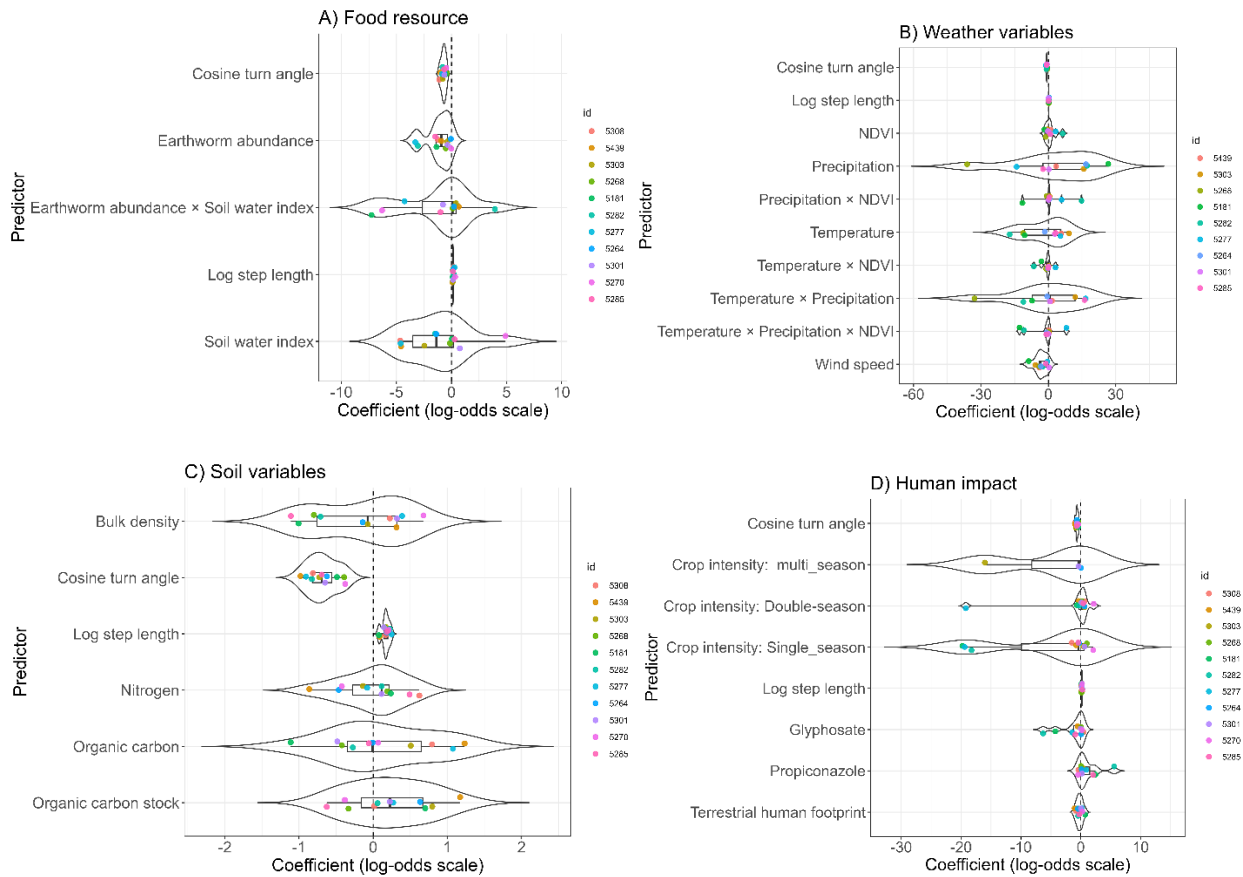
248

249 We used step selection functions to describe resource use variation by northern lapwings in
250 relation to various biotic and abiotic variables (Table S2). The step selection functions indicated
251 that the soil water index and the earthworm abundance factor were significantly affecting the
252 movement patterns. Overall, none of the tested variables was significantly associated with the
253 movement patterns of all northern lapwings (Fig. 3), however, some variables had significant

254 coefficients for certain individuals, indicating differentiation in their impact. For example, in the
 255 case of individual-level step selection functions, the NDVI had either a significantly positive (IDs:
 256 5439, 5264, 5285), a significantly negative (IDs: 5303, 5268, 5301) or a non-significant effect on
 257 the movement patterns of lapwings (Tables S4-S7).

258 Northern lapwings tended to move more during full-moon nights compared to new-moon nights
 259 (Fig. S3). In addition, earth-moon distance was positively correlated with step length (estimate =
 260 0.20, S.E. = 0.09) and positively correlated with full and waxing moon (estimate = 1.17, S.E. =
 261 0.54 and estimate = 0.97, S.E. = 0.56 correspondingly), indicating that northern lapwings tend to
 262 be more active when the moon is less bright but fuller, though there was high between-individual
 263 variation. During the night, individuals were strongly choosing pulses-flowers-vegetable and sugar
 264 beet crops while avoiding potato crops, artificial land and shrubland (Fig. S4).

265



266

267

268 Figure 3. Model estimates for the A) food resource B) weather C) soil D) human impact models.

269

270 We also detected seasonal variation for all models (Fig. S7). Northern lapwing preferences varied

271 across seasons in the visitation rates of crop types, while individual variation in the use patterns

272 of all resources was lowest during summer.

273

274 Repeatability of resource selection

275 We used a step selection function to identify the important crop types across all data. The step

276 selection function had significant effects for barley, maize, sugar beet, pulses-vegetables-flowers

277 and grasslands. We then fitted a repeatability model for each of these crops (Fig. S8). The highest

278 repeatability was observed in the category of pulses, vegetables and flowers ($R=0.31$, 95% CI: 0,

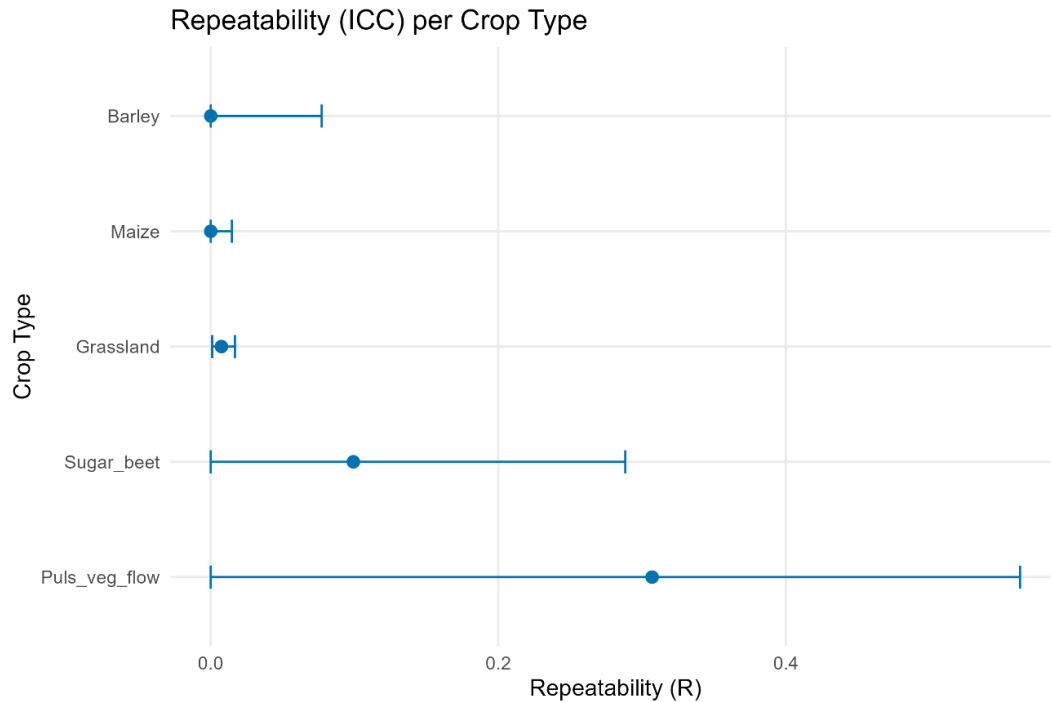
279 0.5), followed by sugar beet ($R=0.09$, 95% CI: 0, 0.28), and grasslands ($R=0.007$, 95% CI: 0, 0.01).

280 All other crop types, including maize, barley, potatoes, grassland, and water, showed negligible

281 repeatability estimates ($R < 0.01$) with confidence intervals overlapping zero (Fig. 4). Repeatability

282 for grasslands and water was also negligible, indicating uniform use patterns between all

283 individuals.



284

285 Figure 4. Repeatability of choosing each crop type across all northern lapwings in our study
 286 (n=13 individuals).

287

288 Discussion

289 In this study we analysed over 40,000 GPS locations of 13 northern lapwing individuals from 2021
 290 to 2023. We used step selection functions and repeatability models to associate lapwing
 291 movement with abiotic and biotic variables and estimate intraspecific niche variation. Overall,
 292 northern lapwings selected primarily grasslands, barley and maize, however there was high
 293 between-individual heterogeneity in the selection of sugar beet and pulses-vegetables-flowers,
 294 confirming previous observations from the UK that northern lapwings are feeding on sugar beet
 295 during the night (Sheldon et al., 2004b).

296 The main biotic variables that we tested were earthworm abundance and human presence.
 297 Earthworms are an important component of the northern lapwing diet (Sheldon et al., 2004a). A

298 recent study analysed dietary habits of northern lapwings in the Netherlands and Germany using
299 metabarcoding techniques and found that earthworms were not a major part of the species'
300 trophic niche (Legendijk et al., 2025). Our study showed that northern lapwings tended to select
301 areas with high earthworm abundance (marginally significant negative values in Fig. 3 indicate a
302 positive association, see also *Biotic factors (human population density, human footprint, and*
303 *earthworm abundance and richness)*), but no significant effect of the interaction between
304 earthworms and soil water content. Previous studies have found, similar to ours, that lapwings
305 show low tolerance to the presence of humans (Roche et al., 2016). However, the individual
306 resource utilisation curves revealed that one individual actually preferred human presence very
307 strongly.

308 The main abiotic variables that we tested, beyond crop types and pesticides, were related to soil
309 water content, crop intensity, soil composition, climate and moon luminosity. Propiconazole is a
310 fungicide while glyphosate is a herbicide for weed control. For example, propiconazole is
311 commonly used in cereal crops (especially wheat and barley) (European Food Safety Authority
312 (EFSA) et al., 2017), while glyphosate is commonly used in wheat, maize or rapeseed crops
313 (Antier et al., 2020). Although the overall effect of propiconazole on the birds' movements was
314 low and non-significant, there were three individual step selection functions (IDs: 5285, 5301 and
315 5439) where propiconazole had a significantly positive estimate. Regarding diurnal activity, our
316 results show that northern lapwings do not always remain inactive during the night. On the
317 contrary, we have shown that northern lapwings tend to be more active (longer steps) in nights
318 with full moon than in nights with new moon, confirming previous observations (Milsom et al.,
319 1990), but their activity was negatively correlated with moon luminosity (measured as the distance
320 between Earth and Moon), likely due to higher predation risk in bright nights. The moon model
321 showed a positive significant relationship between full moon and step length, but with a high
322 standard error. In addition, a large amount of the model's variance was explained by between-
323 individual variation, indicating that activity during night hours differs significantly among

324 individuals. The highest variance explained by among-individual differences was in the moon
325 model, which, in combination with the wide confidence interval of activity during full moon,
326 indicates that not all individuals are equally active at night.

327 Fine-scale observations are instrumental for the effective design of species conservation
328 measures. Field observations of occupancy rates have shown that, for northern lapwings, the
329 effectiveness of AES for lapwings can be enhanced by the presence of water, which is an integral
330 part of the species ecological niche (Hawkes et al., 2025). AES have proved to be beneficial for
331 northern lapwings, since they provide foraging and breeding sites (Chamberlain et al., 2009;
332 O'Brien & Wilson, 2011). However, when designing conservation interventions, it is important to
333 take into account that northern lapwings show preference for an intermediate level of crop
334 management, such as planting or rolling (Fraixedas et al., 2020; McCallum et al., 2018). In other
335 words, they tend to avoid non-managed and natural areas (even with low vegetation), while, in
336 the present study, they also avoid areas where there is multi-seasonal planting. Interestingly, our
337 results confirm previous findings that lapwings select for medium cropping intensity or tilled areas
338 (Düttmann et al., 2018; Horvat & Denac, 2019; McCallum et al., 2018) and improved grasslands
339 (Düttmann et al., 2018), but see (Taylor & Grant, 2004), although this is not the case for restored
340 meadows (Berg et al., 2002). In addition, nesting site selection is not always driven by the
341 perceived vegetation composition or the presence of wetlands (McCallum et al., 2018), implying
342 that there might be more complex and intrinsic factors for breeding habitat selection (Berg et al.,
343 2002).

344 Although northern lapwings are considered a well-studied species, there are still key gaps in our
345 understanding of the underlying mechanisms of their population declines. Moreover, lapwings
346 appear to be well studied in Europe, but not throughout the range of their distribution (Phoswa &
347 Downs, 2025). To this day, there are two main bodies of studies on northern lapwings: habitat
348 selection and breeding success. The relationship between these two, across ecological contexts
349 and scales, is the missing link. Future research can focus on linking fine-scale movement and

350 habitat use with fitness outcomes across countries and continents (also known as the ‘think
351 globally, measure locally’ principle).

352 Conservation recommendations

353 Based on the results of our study and the literature search, we suggest the following conservation
354 measures for the protection of the northern lapwing:

355 1. Habitat management

- 356 • Northern lapwings avoid humans (densely populated areas) and human infrastructure,
357 thus conservation measures should target habitat in remote areas (present study).
- 358 • The presence of even small (e.g. size of 10x10 meters) water bodies (rivers, lakes) is a
359 very important resource for northern lapwings (present study).
- 360 • Spring tillage benefits northern lapwings, but they tend to avoid autumn tillage (Milsom,
361 2005; Sheldon et al., 2004a).
- 362 • Plant spring crops when doing crop rotations particularly roots and cereals (Sheldon et al.,
363 2004a).
- 364 • Fewer machinery operations during breeding season (Schekkerman et al., 2009)
- 365 • Habitat management interventions supported by expert advice (Hunt et al., 2023).
366 Lapwing conservation requires an active involvement of the land managers (McCallum et
367 al., 2018).
- 368 • Liming is likely beneficial for northern lapwing breeding (McCallum et al., 2018).
- 369 • Northern lapwings select for fields with single- or double-season cropping intensity
370 (present study).

371 2. Predation

- 372 • Culling is a sub-optimal solution because predators like foxes can be replaced very fast
373 from neighboring populations (Porteus et al., 2024). Eliminating one predator does not
374 decrease predation overall (Teunissen et al., 2008).

- 375 • Northern lapwings are active also during nights with intermediate moon light (present
376 study).
- 377 3. Supporting breeding success
- 378 • Avoiding mowing and grass harvesting reduces chick mortality (Schekkerman et al.,
379 2009).
- 380 • Ground nests are more threatened by horses rather than cattle (Mandema et al., 2013).
- 381 • Set aside areas for lapwing nesting (aka lapwing plots) (Chamberlain et al., 2009).
- 382 • Provision of nesting sites should be coupled with limited nest destruction and
383 establishment of cropland-grassland mosaics (Milsom, 2005).
- 384 • Lapwings choose to nest away from potential predator perching spots (e.g. hedges)
385 (McCallum et al., 2018), thus the removal of vegetation (such as shrubs) can reduce
386 predation (Hunt et al., 2023).

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