

# A systems-modelling approach to predict biological responses to extreme heat

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## 26    **Abstract**

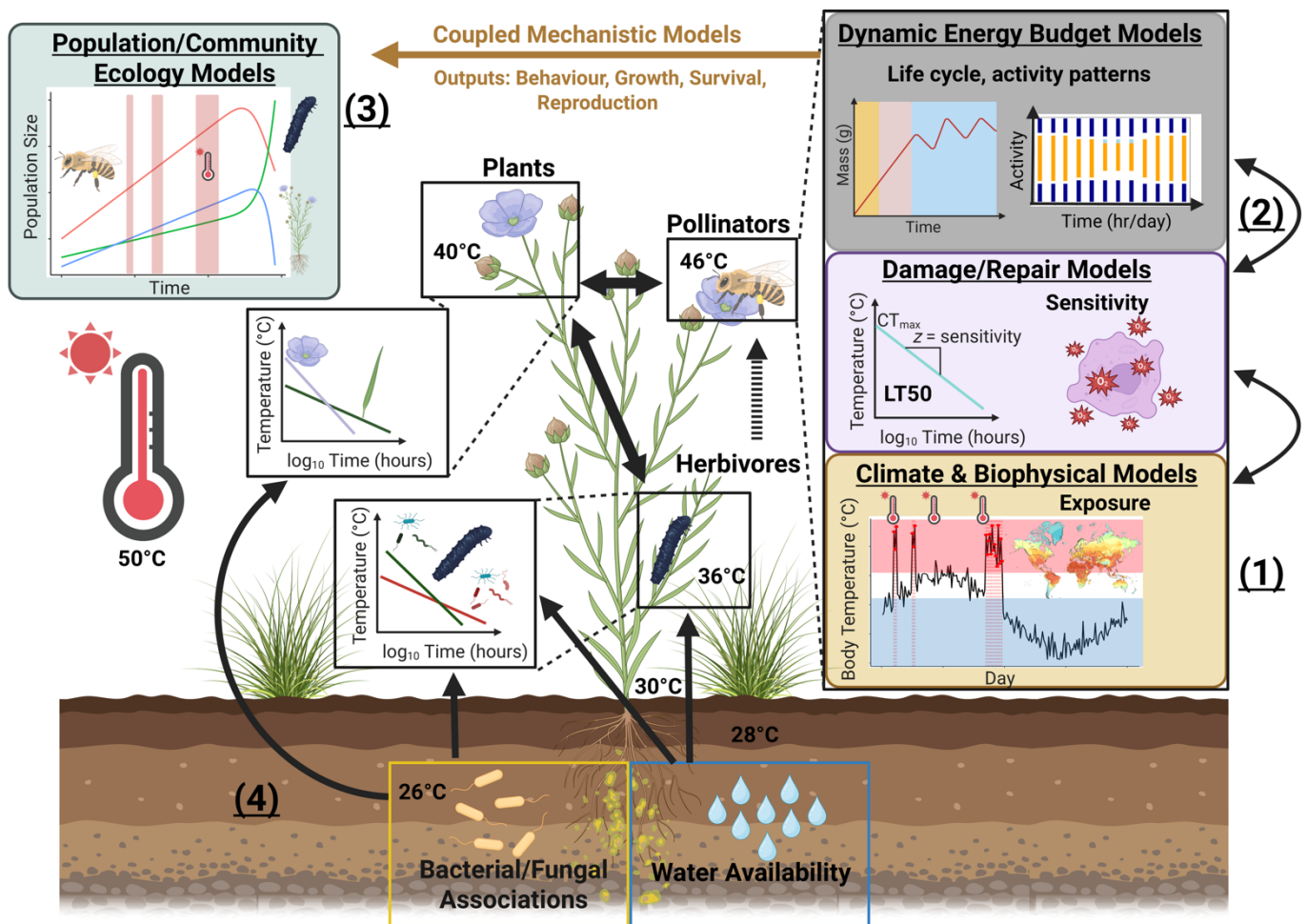
27    Anthropogenic climate change is leading to more frequent and extreme heat waves. These large-scale events  
28    are radically re-shaping interactions among organisms – impacting biodiversity, community composition and  
29    ecosystem services crucial to natural systems and food security. Predicting heat wave impacts on interacting  
30    species requires an understanding of the processes driving differential exposure and sensitivity of organisms  
31    to extreme heat events in a life-cycle context. To achieve this predictive capacity, we need to integrate  
32    models across scales while capturing species-specific responses at the individual level. We review and  
33    demonstrate how existing models in disparate fields can be linked to achieve an increased understanding of  
34    how individuals and communities will respond to extreme heat, now and into the future.

## 35    **A systems-modelling approach to understand the biotic impacts of heatwaves**

36    Climate change is leading to warmer and more variable thermal environments globally [1,2]. Greater thermal  
37    variability is resulting in organisms experiencing extreme heat waves that lead to thermal stress impacting  
38    organismal growth, survival and reproduction, with cascading effects on population dynamics, species  
39    interactions, community composition and ecosystem structure and function [3,4]. Climate variability has  
40    already been linked to dramatic global declines in pollinator abundance [5] and crop yields [6–8], but the  
41    causes underlying such declines, and their ramifications through communities and society, are not fully  
42    understood. The impacts of extreme heat on species and communities are driven by a combination of direct  
43    effects of heat stress on the physiology and fitness of organisms within a given species [9–11] and indirect  
44    effects on interactions (both positive and negative) among species [e.g., 12]. Understanding the dynamic  
45    interplay between direct and indirect effects of extreme heat, and how these are mediated by environmental  
46    factors (e.g., water and food availability, microbial community), has been hampered by inadequate coupling  
47    of models that predict short-term physiological damage on a given species' fitness with population and  
48    community dynamic models across species [11,13,14]. A systems-thinking approach is now needed to tackle  
49    the multifaceted nature with which extreme heat manifests within a community. This approach will require  
50    interdisciplinary collaboration to integrate biophysical, physiological, population and community ecology  
51    processes so that we can capture and model the dynamic feedbacks, nonlinearities and interactions between  
52    species that drive responses across scales [15,16].

53    Using plants and insects as examples, we review broad classes of models across subfields of ecology and  
54    evolution and discuss how they can be linked to effectively model the biotic impacts of heatwaves from  
55    individuals to communities (Figure 1). We focus on plant and insect communities given the strong  
56    interconnections between them, the ease with which they can and have been studied, and their importance to

socioecological systems [17]. Physiological models can now be seamlessly integrated with biophysical models to characterize temperatures experienced by organisms and simulate the effects of extreme heat events on the entire life cycle of species within ecological communities [18–21] (Figure 1). Physiological models that incorporate estimates of thermal sensitivity across species can capture the delicate balance between damage and repair of physiological systems [10,11,13,22], yielding predictions of the immediate and cumulative impacts of extreme heat on growth, survival and reproduction. Importantly, mechanistic physiological models provide outputs at the individual level (e.g., energy and water requirements, waste production, activity constraints, vital rates) that can be integrated into population and community ecology models to capture how extreme heat events perturb interactions among species and across whole, complex communities [23,24]. We demonstrate how such a coupling can be made within and across species and highlight the opportunities it presents to develop a greater understanding of the ways in which extreme heat stress manifests across species. We discuss the challenges of coupling models across diverse species in communities, and of capturing the eco-evolutionary feedbacks that will be necessary for accurate predictions in the future.



**Figure 1- A coupled systems modelling framework for predicting the effects of extreme heat on ecological communities.** A systems modelling approach first integrates (1) climate/biophysical models to predict the different microclimates species experience. (2) Microclimates experienced by organisms then provide inputs for physiological models that integrate temperature exposure with explicit physiological processes that simulate development, behaviour, survival (e.g., 50% survival thresholds, LT50) and reproduction in response to microclimate dynamics across life history stages. Physiological models can be built around the unique life cycles of the diverse species in communities, capturing lagged responses to extreme heat, phenological mismatches and mechanistically informed responses to extreme heat for species of the system in question. Outputs (behaviour, growth rate, biomass accumulation, survival and reproduction) from physiological models of individual organisms can then be integrated into (3) population and community ecology models (such as those developed with in the Modern Coexistence Theory Framework – see text) to predict population growth and community composition under a specific type of change (either to the environment or suite of interacting species). Environment-mediated feedback loops (4) influence organism thermal exposure and sensitivity, including water availability and microbial communities, which can vary across different life stages and organs depending on the microclimates occupied. Coupling existing models will allow for quantitative predictions to be made on how extreme heatwaves perturb biological systems and for the development and implementation of strategies to enhance biological system resilience to extreme heat. Created with BioRender.com.

## 71 **From weather to microclimates: predicting community wide exposure to** 72 **extreme heat**

73 Extreme weather events such as heatwaves are predicted by Global Circulation Models (GCMs), typically at  
74 large spatial scales (e.g. 0.5° grid cells), and can be regionalized by downscaling to smaller spatial scales.  
75 However, species within communities experience heatwaves differently because of the varied microhabitats  
76 they occupy. For example, while air temperature may be 50°C, an organism a metre below the soil surface  
77 may be exposed to temperatures ~20°C cooler (Figure 1). Therefore, to understand how extreme heat affects  
78 a community of organisms, we must characterize the **microclimates** experienced by individuals of different  
79 species under a given atmospheric event. Microclimate predictions are made by taking the atmospheric  
80 conditions as independent forcing variables and combining them with detailed information on terrain (slope,  
81 aspect, hill shade), vegetation [plant-area index (PAI), stomatal behaviour, leaf reflectance] and soil  
82 hydrothermal properties to predict how radiation, wind speed, air temperature, humidity, soil temperature and  
83 soil moisture vary on small spatial scales [25–32]. Capturing the interaction between soil moisture and

84 temperature in microclimate models is critical because the same atmospheric heat wave can have different  
85 implications for microclimatic conditions depending on the recent history of rainfall and temperature [33].  
86 Downscaling to microclimate conditions requires the specification of key environmental properties (e.g. %  
87 shade, soil characteristics, surface albedo) at a fine scale (metres) for the local area of interest. Depending on  
88 the organism and question of interest, not all input variables are needed to parameterize a given microclimate  
89 calculation, as recent guides to microclimate modelling illustrate [e.g., 34,35,36].

90 Beyond parameterization, the benefit of a systems-thinking approach in the context of microclimate  
91 modelling is its ability to capture vegetation and soil dynamics which can dramatically shape microclimate  
92 conditions within and between species dynamically across time. While the large-scale feedbacks from plant  
93 and soil dynamics to heatwave development are captured in GCMs [e.g. 37], these dynamics are also  
94 important in determining thermal regimes at micro-scales. For example, plant water use dynamics can  
95 amplify or reduce within-canopy temperatures via effects on photosynthetic capacity, stomatal decoupling,  
96 cuticular conductance, leaf damage and plant mortality [20,38]. These impacts are important for the plants  
97 themselves and for thermal regimes available to other organisms. Forecasting future thermal regimes is  
98 challenging because it involves forecasting future vegetation dynamics, including changes in key vegetation  
99 properties such as leaf area index, as a function of plant growth, phenology, plant population dynamics and  
100 shifts in community composition. Despite the complexity and non-linearity of these interactions across the  
101 soil-plant-atmosphere continuum, there is existing capacity, and growing potential, to model them and  
102 determine individual-specific exposure to extreme heat. For example, a wide range of vegetation models is  
103 available for this purpose, from crop growth models that simulate growth and yield of crops over a season  
104 (e.g. Agricultural Production Systems sIMulator (APSIM); [39]) up to the dynamic global vegetation models  
105 (DGVMs) that simulate vegetation function and distribution at local to global scales [40].

## 106 **Translating exposure to organism temperature: biophysics to the rescue!**

107 Once microclimates are quantified, it is crucial to estimate the heat and water budgets of organisms to  
108 determine how microclimate variability within a system translates to realized organismal temperatures and  
109 hydration states. Heat and water budgets can be computed using a combination of species functional traits  
110 (e.g., body mass, metabolic rate, surface area, solar absorptivity) along with how heat energy and water are  
111 exchanged with the environment – dependent on the microclimate experienced. Such **biophysical models**  
112 have a long history [41,42] but have become more widely applied in ecology in the past 20 years, facilitated  
113 by developments in environmental datasets, microclimate modelling, and the emergence of high-level  
114 programming languages such as R [21]. Biophysical models of ectothermic animals make use of equations  
115 for energy and mass exchange between an organism and its environment and can account for complex

radiative heat transfer and the role of evaporative water loss across surfaces [21,25,43]. Equivalent models of leaves incorporate the dynamic role of stomata [20, 44] (**Box 1**). Biophysical models are powerful because, by translating microclimate conditions to organism body temperature and hydric states, simple regulatory decision-making models can be used [45,46]. For example, given an understanding of an organism's thermal activity windows, thermal optima, and critical thermal limits we can determine what microclimates the organism should chose to get close to its target body temperature [e.g., 45,46]. Such behavioural thermoregulatory decision making is crucial for mitigating the negative impacts of extreme heat. Incorporating it into models can inherently capture the trade-offs such a strategy entails through reduced foraging time [21,45]. Similarly, changes in stomatal behaviour can mitigate or exacerbate the extreme temperatures leaf tissue experiences while simultaneously shaping the microclimate conditions for other organisms [20,47] (**Box 1**). Future developments in biophysical modelling of organismal temperature under extreme heat will involve understanding the nuances of plastic physiological and behavioral responses. For example, lizards may pant when exposed to high temperatures [48], birds may 'wind surf' or seek thermal micro-refugia [49], and stomata can enact emergency cooling, departing from the typical responses to vapor pressure and light intensity [38], though this response may depend on prior soil moisture conditions [50]. Although biophysical models require many traits for parameterization, the models and parameters can be tailored to the question and organism of interest [21]. In addition, many traits can be approximated based on traits of similar species, and then validated and updated as more species-specific data become available. The application of biophysical models in concert with microclimate models can thus be used to infer the thermal conditions to which members of the same ecological community are exposed with different degrees of detail. Realised thermal exposure can then be linked to physiological models.

## **Capturing both physiological damage and repair to predict multi-species thermal sensitivity**

Translating how heat waves impact plants and animals not only depends on modelling temperature *exposure* but also the varying *sensitivity* of organisms to extreme heat [22,51]. Sensitivity to extreme heat can be captured by **thermal load sensitivity (TLS) / thermal death time (TDT)** models that explicitly account for how heat stress depends on both the body temperature experienced and its duration [13,22]. TLS models predict the relative accumulation of damage to cellular and sub-cellular systems that compromise physiological function. Without periods of recovery, where damage can be repaired, organisms accumulate damage over time, reducing growth and impacting survival and reproduction (**Box 2**). Typically, TLS models focus on endpoints that include survival (e.g., lethal temperatures,  $LT_{50}$  or  $LT_{80}$ ) or some measure of reduced fertility, but this need not be the case [13]. Endpoints are predicted by assuming the effect of time at a given

temperature decreases survival and/or fertility exponentially, and have been shown to have high predictive power [10,11,22,51]. For example, Ørsted *et al.* [11] show that both mortality and fecundity follow a clear exponential relationship with time in the Spotted Wing Drosophila (*Drosophila suzukii*), with survival and fecundity being compromised most for long thermal exposures at high temperatures. Importantly, they also show that reproductive sensitivity can be higher than mortality, with heat injury impacts accumulating faster for reproduction than survival [11]. TLS theory also applies well to photosynthetic function in plants [52,53], highlighting its generality. TLS models are crucial for predicting how heat waves affect organisms because accumulated damage to physiological systems can result in lagged responses to heat stress or exacerbate future stress (i.e., future heat waves / droughts) – a common feature of extreme heat events [13,54].

Environmental factors known to impact thermal sensitivity, such as water and drought stress [e.g., 55,56], along with nutritional and dietary changes [e.g., 57,58], can be incorporated into systems modelling approaches through their impacts on thermal sensitivity and tolerance [13] (Figure 1). Indeed, an exciting potential application of a systems-modelling approach is to explore how microclimatic conditions mediate changes to interactions between plant and animal microbial communities. Interactions between plants and microorganisms, such as plant growth-promoting rhizobacteria (PGPR), arbuscular mycorrhizal fungi (AMF), and bacterial or fungal endophytes, are known to enhance growth, defense, and heat tolerance in plants [59,60], and gut microbiota can improve heat tolerance in animals [61]. Additionally, modelling the intricate balance between damage and repair for a suite of different species can help identify susceptible species, life stages and tissues that are most at risk from extreme heat events due to direct sensitivity to extreme heat, allowing more accurate predictions of the varying levels of species sensitivity within a community [13].

### **Box 1: Building a systems modelling approach to capture multi-species exposure to extreme heat**

Here we outline how a systems-modelling approach can be developed, illustrated with a simple hypothetical community (Figure IA). The focus is on how climate, microclimate, and biophysical models can predict the differential exposure to extreme heat events across species, how species can alter each other's microclimates, and the incorporation of 'behavioural' responses of species in response to temperature (e.g., stomatal behaviour, thermoregulation). Predicted organismal temperatures are used from these models to then predict life-cycles for species, incorporate differences in thermal sensitivity, and ultimately estimate vital rates for population and community ecology models (**Box 2**).

The context is a heatwave event in January 2018/19 at Renmark, South Australia (Calperum Station). The historical SILO climate dataset (0.05° resolution, 1889 to present) [62] was used as input to the `micro_silo` function in *NicheMapR* [20,63], in conjunction with *microclima* [26,27] in the R statistical language to compute microclimates. Microclimates are calculated at various heights above ground (20cm-1.2m, relevant to our plants and insects) and soil depths (for burrowers) with varying levels of shade (we use 40% and 80%). Simulating these varied conditions captured the diversity of microclimates that different species in a community might experience and provided opportunities to build in behavioural plasticity.

With available microclimates predicted, we can compute leaf and body temperatures using biophysical models. However, plant temperatures will impact the temperatures experienced by the grasshoppers because plant stomatal conductance changes in response to extreme heat events. Stomatal responses of leaves to their environment can be captured with the help of the *plantecophys* package [44] and used in conjunction with the `ectotherm` function in *NicheMapR* to compute realistic leaf temperatures as stomata open and close in response to temperature and vapor pressure deficit (Figure IB). The combination of microclimate, leaf traits (e.g., shape, leaf conductance, emissivity) and stomatal behaviour thereby produces realistic leaf temperature estimates during the heatwave event.

Calculated leaf temperature can then become part of the microclimate of the insects (Figure IC). The `ectotherm` function of *NicheMapR* can be used to compute insect body temperature, given the microclimate (including plant leaf temperature) and insect traits. The insect temperature calculations incorporate their capacity to thermoregulate behaviourally. For example, Species B moves to cooler parts of the plant to minimize thermal stress, when possible (Figure IA), given a range of available microclimates and using information on their thermal preferences. All code to demonstrate this process is available at: [https://daniel1noble.github.io/thermal\\_tol\\_interactions/](https://daniel1noble.github.io/thermal_tol_interactions/).



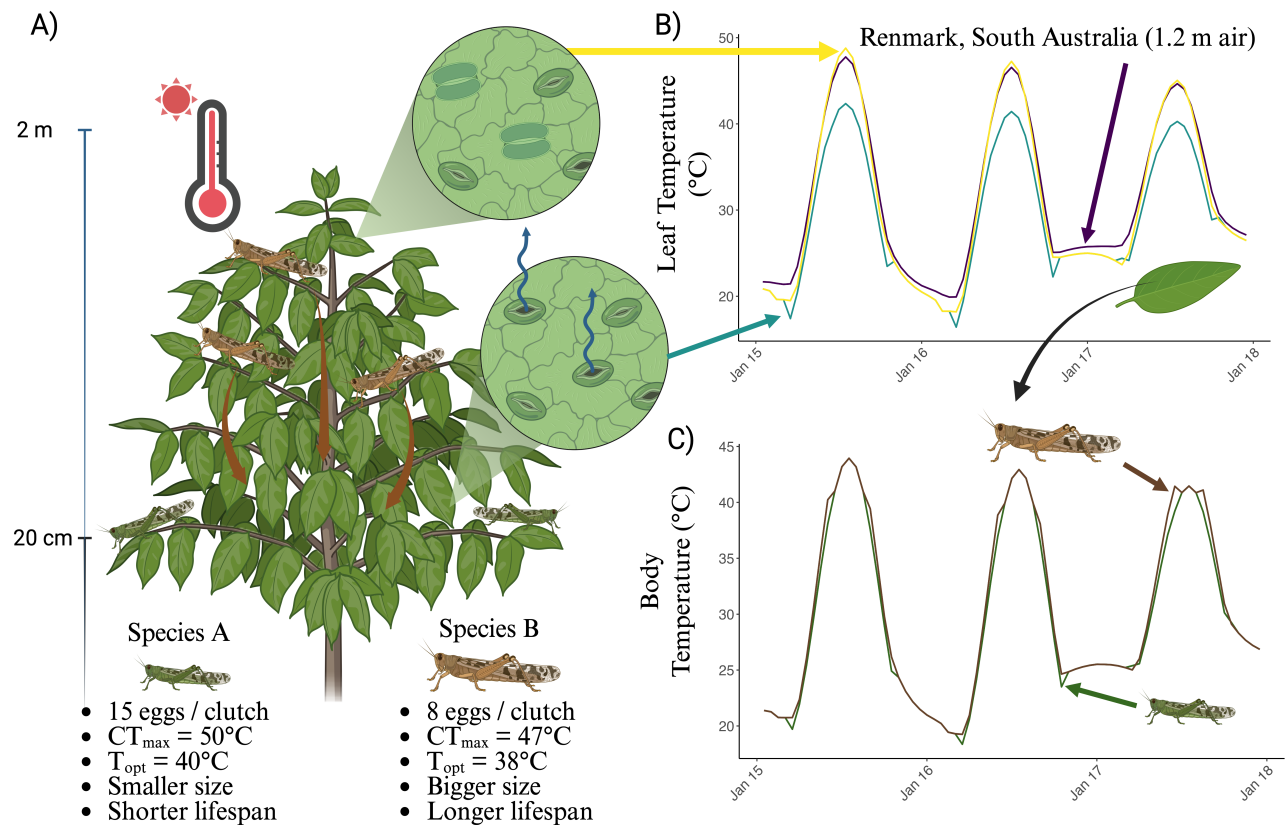


Figure I: Predicting and coupling microclimate and biophysical models of a plant and two grasshopper species that vary in their thermal tolerance and life history (A). Biophysical and thermoregulatory models use microclimate data to predict leaf (B) and grasshopper (C) temperatures through time (coloured lines) incorporating changes in insect and stomatal behaviour. Created with BioRender.com.

## Glossary

**Biophysical models:** Biophysical models capture the balances of heat, water, and other aspects of energy and mass exchange between organisms and their microclimatic environment to predict how organisms function, survive, and behave in varying environments.

**Coexistence:** Coexistence occurs when populations of two or more species are able to persist in each other's presence indefinitely at steady state under constant abiotic conditions.

**Functional traits:** Functional traits are the characteristics of organisms that influence their performance and fitness, such as body size, reproductive output, and metabolic rates. In the context of extreme heat, functional traits can be used to categorize species based on their responses to thermal stress.

**Modern Coexistence Theory:** A theoretical framework derived from population ecology describing the conditions under which species can coexist mathematically. It has resulted in the development of population growth models using criteria based on niche and fitness differences, mutual invasibility and feasibility domains.

**Microclimate:** Microclimates are the local conditions experienced by organisms, which can differ substantially from the broader atmospheric conditions. Microclimates are influenced by factors such as vegetation cover, soil properties, and topography, and can significantly affect the thermal and hydric environment experienced by organisms.

**Thermal load sensitivity (TLS) / thermal death time (TDT) models:** Models that calculate thermal stress load from a time series of body temperatures to quantify lethal and sublethal impacts on tissues, organs and whole organisms.

**Vital rates:** Demographic parameters that determine how populations change over time. They include measures of birth, death, growth, and reproduction.

## Physiological models that capture the effects of extreme heat across life

Accumulated damage to an organism from experiencing extreme heat events will impact survival probability in the future and needs to be integrated with a full account of growth, development and reproduction through life to scale effects to populations and communities. Dynamic Energy Budget (DEB) models are physiologically-explicit life-cycle models of energy and mass uptake and conversion that predict key life-history transitions, growth, reproduction and survival (senescence) through time [18,64]. DEB models consider organisms as a thermodynamic system (fully obeying energy-mass conservation) capturing the

exchange of food, water, respiratory gases and metabolic waste throughout the life cycle [21]. DEB models can translate short-term (hours) variation in factors such as temperature, toxins (e.g., toxicants, [65]) and resource availability (nutrients) [66,67] into long-term (days, months, years) effects on development, growth, reproduction and survival [19,21]. DEB theory, when combined with biophysical models more generally, allows the full water budget to be computed to account for heat stress impacts of a hydric nature [21]. Perhaps most importantly, DEB theory calculates the life cycle trajectory of an organism and so can account for the effects of timing of heat wave impacts relative to life stage. Applied to multi-species assemblages, DEB models can capture how different species' life cycles are affected by heat and interact with each other to predict how traits such as birth, age at maturity, reproductive events, energy flow and lifespan vary across species (**Box 2**). DEB models integrate many fundamental physiological processes and are parameter-sparse (one parameter per process). Moreover, the necessary parameters have already been estimated for > 7000 species already (although there are taxonomic biases) [68,69]. New analytical approaches using phylogenetic imputation methods show promise for calculating energy budgets for species without data. Such methods have, for instance, already been used to predict DEB parameters for over 1.3 million animal species [see, 70]. In addition, software packages such as *NicheMapR* [25,63] allow for simulations of life-cycles using parameters for diverse species. DEB model outputs, such as body mass and size, reproductive success and timing, and survival, can be used to parameterize vital rates needed for population and community models, with additional options to incorporate eco-evolutionary feedbacks (see **Box 3**).

Analogously, physiology-based vegetation models simulate plant growth over time as a function of carbon, water and nutrient uptake and use, following mass balance principles [71]. Model drivers typically include incident radiation, humidity, rainfall, and soil properties, as well as air temperature. The models capture the direct effects of temperature on key physiological processes, such as photosynthesis, respiration and phenology, as well as the indirect effects via feedbacks on vegetation water balance and soil nutrient availability. Short-term responses to temperature are captured well in these models through representations of enzyme kinetics, but representation of longer-term responses remains challenging because plants show high flexibility in their ability to acclimate to ambient temperatures [72]. Another area under active research is capturing plant damage and death from hot-dry weather extremes. Considerable progress has been made by representing the plant hydraulic system explicitly, enabling the risk of hydraulic failure to be predicted under conditions of low rainfall and high evaporative demand [73]. Direct damage to plant tissue from extreme heat has yet to be represented in such models, but the thermal death time framework outlined above offers a promising way forward.

## Mechanistic approaches that ‘speak’ to population and community ecology under extreme heat

Modelling tools developed as part of **Modern Coexistence Theory** (MCT) [16] offer a promising suite of approaches for combining biophysical, physiological, and population ecology models to predict whole community responses to extreme heat events. Population growth models are core to predicting species **coexistence** within this framework. In their most recent applications, these simple pairwise population growth models have been used to underpin horizontal network models that allow one to examine how species in a community interact and create stable communities [15,74]. Other extensions of MCT tools have integrated environmental variation [24], traits [75] and cross-trophic dynamics [76,77] in predicting the outcomes of species interactions in community contexts. Fundamentally, this approach relies on simple two-species individual-based population growth models [78,79], but key extensions involve the use of whole communities as one “species” [80,81] or the grouping of species to simplify interaction matrices while accounting for a small number of dominant species [77].

Population dynamic models that are used as part of MCT can calculate the individual fitness effect of interactions in relation to environments experienced by organisms [23,77,82]. As these models use the same fitness measures as DEB models, they offer a population modelling framework to add thermal biology to community diversity predictions. This is because population dynamical models can be combined, for example as networks [15] and compared between microsites with different biophysical properties and harbouring species with different thermal responses to their microenvironment. For example, Bimler *et al* [15] used this approach to determine which species were keystone species under shady and sunny portions of the same plant communities.

Modified **vital rate** functions (which can incorporate any measure of fitness) can be used in population growth models to compare population dynamics for individuals in areas with different biophysical properties and/or thermal tolerances. Indeed, the predicted traits from DEB models could be used to directly parameterise growth rate models (**Box 2**) or to predict **functional traits** which, at a community scale, can be used to categorise species with different response and effect profiles. Horizontal interaction networks can then be used to determine how important microclimates with different biophysical properties are for species interactions (defined by their fitness responses to their thermal environment) within the context of whole communities. These networks can be made to target particular types of interaction effects or responses to extreme heat events, or used to compare how species interact in areas with different thermal landscapes. The main limitation of these methods is that they are data hungry, but even this can be handled by categorizing

239 species by ‘traits’ or shared phylogenetic relationships, which are simplifications shown to be effective at  
240 reducing model complexity without sacrificing model accuracy [83].

## Box 2: Scaling up extreme heat effects from impacts on individuals to populations and communities

We use predicted organismal temperatures (Figure I in Box 1) to take stock of how thermal heat stress accumulation impacts survival probability and simulate life cycles for two interacting grasshopper species under their respective microclimates.

To start, using existing parameters for a dynamic energy budget (DEB) model for grasshoppers [84], we can use the ectotherm function in *NicheMapR* to simulate life cycles for each of the two species. We assume that both species have similar DEB parameters but vary in their life-history, thermal physiology, size and reproduction.

DEB models incorporate a simple Gompertz mortality function to capture senescence, but do not integrate the effects of thermal load accumulation in response to extreme heat events into survival functions. We assume reproduction is not impacted by heat for simplicity, but thermal load on reproduction can also be incorporated in a similar way [11]. By using thermal sensitivity parameters for mortality endpoints from TLS models (i.e.  $z$ , slope and  $\alpha$  = critical thermal limit), they can be used to predict the accumulation of damage through time during stressful temperatures (environmental temperatures  $T_e >$  critical temperatures  $T_c$ ) experienced under realistic conditions for each species by calculating heat injury (HI) accumulated between time point  $t_i$  and  $t_{i+1}$  across the insect's life (using an  $LT_{50}$  threshold)(Equation 1) ('red lines' in Figure IIA) [10,11,13,51].

$$HI = \sum_{i=1}^{T_e > T_c} \frac{100 \cdot (t_{i+1} - t_i)}{10^{\left(-\frac{1}{z} \max(T_i, T_{i+1}) + \alpha\right)}} \quad (1)$$

After calculating HI accounting for repair [85], and converting to the probability of mortality through time, we can calculate the total probability of survival (i.e., from senescence, thermal stress, and activity-based mortality) for each age/stage class. In combination with the DEB model outputs on total reproduction across life, we can build a simple age/stage-structured population matrix (day of life or developmental stage) (Figure IIB) to estimate population growth rate [ $r_{max} = \ln(\lambda)$ ] for each species under their respective microclimates.  $r_{max}$  can then be included in a two-species, density dependent Ricker model (a model commonly used as the basis for coexistence modelling in the MCT framework) to predict their population growth under competition (Equation 2).

$$N_{t+1}^i = N_t^i \cdot \exp \left[ r_{max,i} \left( 1 - \frac{N_t^i + \sum_{j \neq i} \alpha_{ij} N_t^j}{K_i} \right) \right] \quad (2)$$

where  $N_t^i$  is the population size of species  $i$  at time  $t$ ,  $r_{max_i}$  is the intrinsic growth rate of species  $i$ ,  $K_i$  is the carrying capacity of species  $i$ , and  $\alpha_{ij}$  is the competition coefficient that describes how much species  $j$  affects species  $i$ . Note that, because mechanistic physiological models are used, the population matrix can be updated each generation to reflect changes in vital rates (survival and reproduction) under different microclimates in the future.

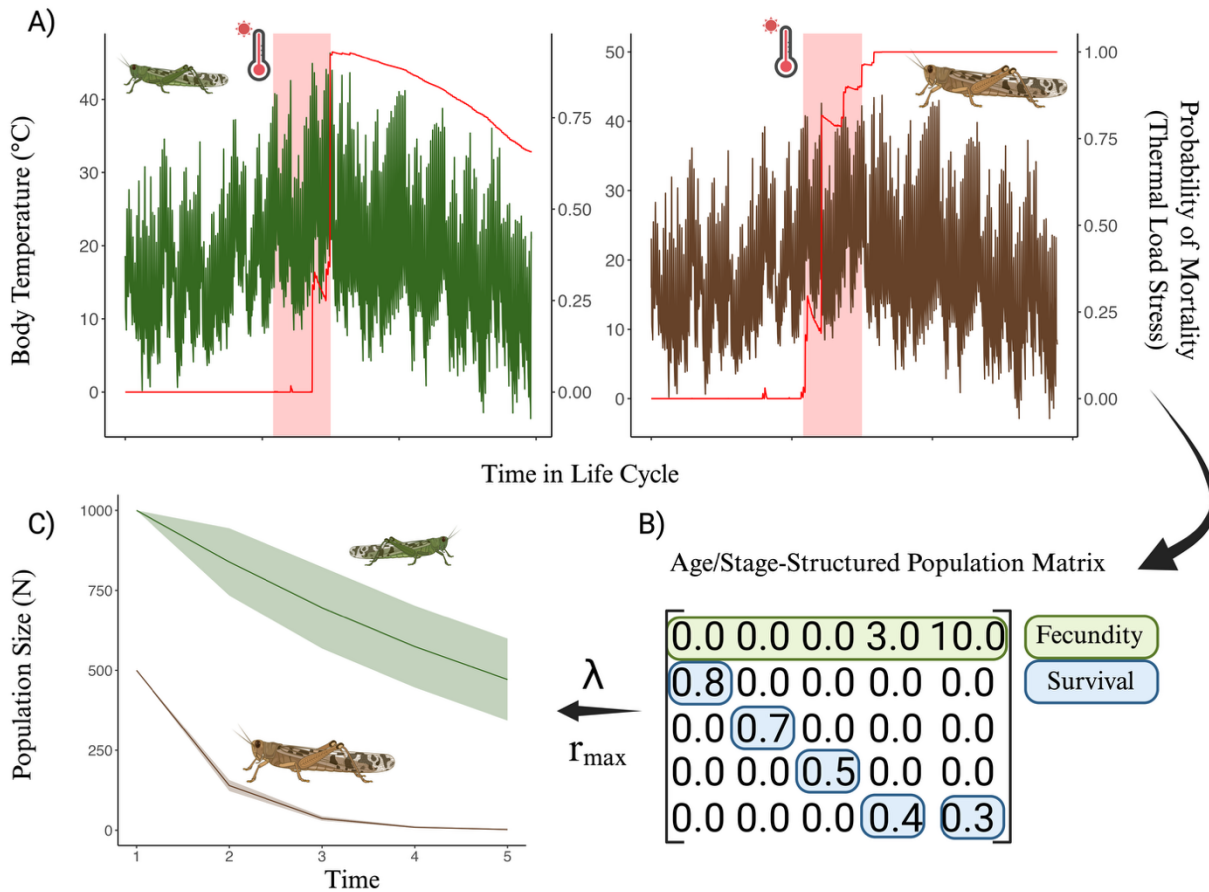


Figure II: A systems approach to using mechanistic physiological models to capture life-cycle vital rates and thermal load stress on survival to scale up to multi-species population dynamics. Created with BioRender.com.

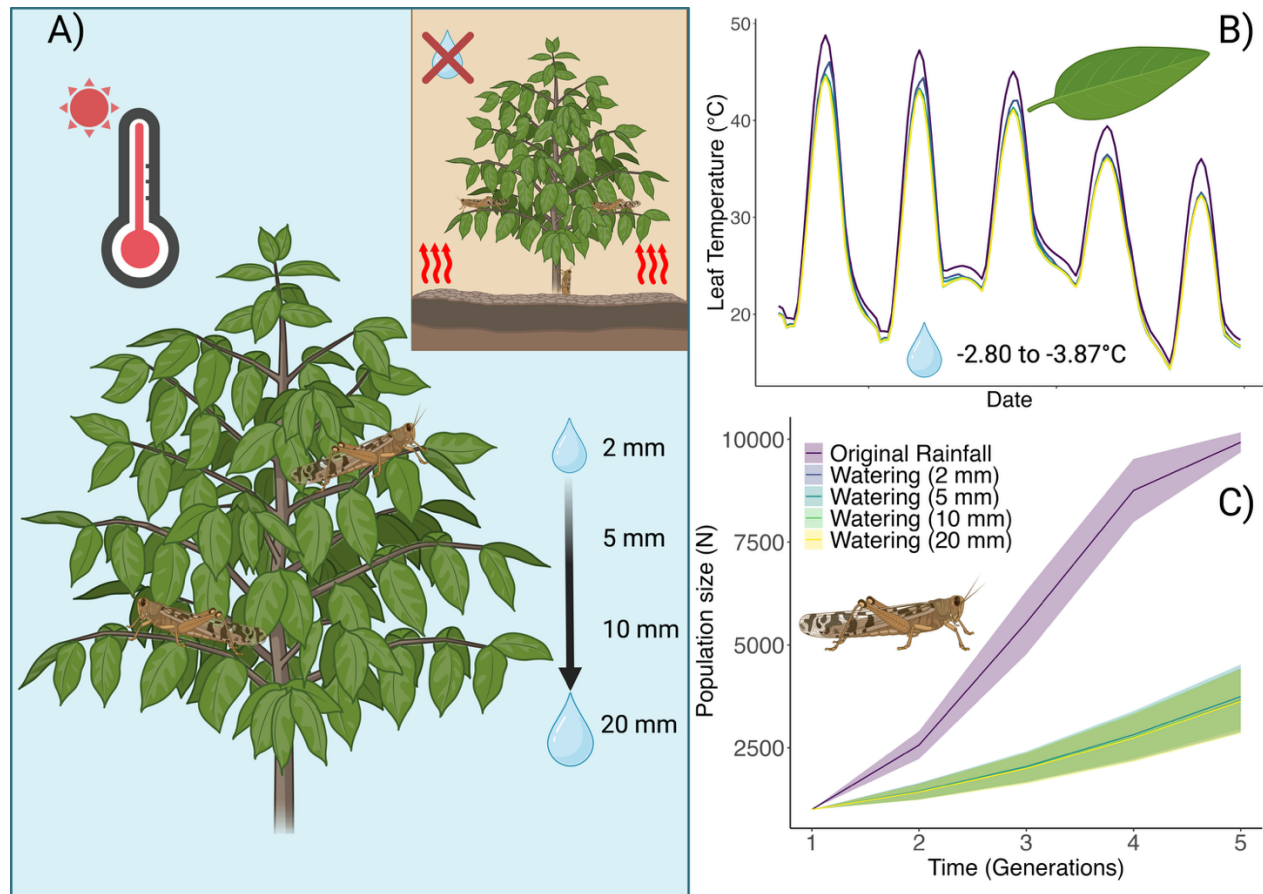
## ‘Hot’ solutions for a changing world

Extreme heat will have dramatic and widespread effects on biological and socioecological systems that humans rely upon. For example, under current global climate change scenarios, each degree Celsius increase in global mean temperature is estimated to reduce the global yield of wheat by 6.0% and soybean by 3.1% [86], exacerbating food shortages.

A systems-modelling approach can help us develop suitable interventions, predict the varied consequences of such interventions, and allow for adjustments that can improve decision making for mitigating the impacts of extreme heat in real world situations. As a simple example (Figure 2A), a systems modelling approach can allow for informed predictions about how to mitigate heat stress to crops while also considering the potential for pest outbreaks. In essence, we want to add a minimal amount of water to a system to cool leaves while avoiding depleting limited water supplies and/or creating better microclimate conditions for pests. Making use of real weather data to predict microclimates and organism temperatures, we can assess how adding 2, 5, 10, or 20 mm d<sup>-1</sup> of water to a system affects leaf temperatures and the life-cycle for a grasshopper species that is a known pest (Figure 2). We can see that adding 2-5 mm d<sup>-1</sup> of water can decrease leaf temperature by up to 3.87°C (Figure 2B), but adding more water does not result in significant gains in cooling. In addition, 2-5 mm of water added each day results in the grasshopper population being suppressed over time, whereas not adding water at all results in grasshopper population growth, leading to potential heatwave-associated outbreaks. Here, the optimal solution is adding 2-5 mm d<sup>-1</sup> of water to both cool plants and protect them from pests. Identifying and/or modifying the thermal suitability of landscapes, such as in our simple example, may dramatically benefit plants and animals under extreme heat and is a promising feature of a systems-thinking approach [21]. Additionally, as our understanding of the mechanisms associated with heat tolerance improve, genetic tools, environmental manipulations (e.g., promoting beneficial microbes) and targeted plant breeding can be used to enhance thermal tolerance, creating more resilience to extreme heat events [e.g., 60].

Systems modelling can also help understand how people's responses to the impacts of extreme heat could lead to accelerating changes in plants, animals and landscapes. Farmers or ecosystem managers may introduce new species, alter management practices or farm remaining unaffected areas more intensively. Thermal stress may ultimately result in people leaving areas, potentially making remaining communities more vulnerable through loss of resources and services [87]. Thermal risks to ecosystems may also precipitate declines in individual wellbeing, and these consequences may accrue differently across regions and peoples [88].





**Figure 2 - How a systems modelling approach to extreme heat can be used to inform solutions.** Heat stress to crops can result in foliage loss and crop failure, but also weaken crops to pests. Using a hypothetical case study of a crop and associated pest, we explore how different watering regimes affect plant temperatures and the flow on effects such regimes have for pest dynamics (A). Using a site near Renmark, South Australia, we use microclimate and biophysical models to estimate leaf temperatures (detailed in Boxes 1 & 2). We see that, compared to original rainfall conditions, certain watering regimes can significantly reduce leaf temperatures by nearly 4°C, providing protection to leaves (B). Modelling how these changed microclimate conditions translate to grasshopper body temperatures, we can then estimate changes to grasshopper life cycles and use the vital rates to predict population dynamics, showing, in this case, that even small amounts of water can indirectly suppress pest growth through lower body temperatures (C). Created with BioRender.com.

### Box 3: An important frontier: integrating eco-evolutionary feedback within a systems-modelling framework

Extreme climactic events, such as heatwaves and droughts, can result in intense episodes of selection which can lead to rapid evolution of heritable traits [89]. Given the intensity of selection associated with extreme climatic events, these events have the potential to radically alter the phenotypic and genetic variability available to selection [90], potentially impacting population resilience to future events. Predicting evolutionary responses from extreme heat events is challenging because of their rarity and stochasticity which makes quantifying the strength of selection difficult as it often can only be opportunistically measured [89].

Predicting evolutionary responses also relies on our ability to predict heatwaves and how organisms experience them in the future – a task that is becoming more feasible with new climate models [e.g., 91,92]. Evolutionary consequences of extreme heat events will depend on the characteristics of these events (e.g., frequency, timing, exposure intensity and duration), the traits under selection and their levels of genetic (co)variation, as well as the demographic and ecological context in which the event occurs [93–95]. Considering various forms of plastic responses (e.g., behavioural, developmental, and physiological plasticity) as environments change through time is also crucial because plasticity can weaken selection and shelter genetic variation while also promoting persistence [96,97]. Many of these processes can now be incorporated into a systems modelling framework (e.g, behavioural plasticity), to capture environment-phenotype feedbacks.

A more complete systems modelling approach that captures eco-evolutionary dynamics needs to treat ‘traits’ more broadly than is typically done in evolutionary ecology (e.g., body mass). For example, ‘traits’ can include the parameters within models that establish functional traits within a population (e.g., DEB parameters)[98,99], which may provide predictions for suites of resulting traits that emerge from such parameters. Furthermore, it is important to consider the evolution of suites of traits as can be done using the multivariate breeders equation [100]:

$$\Delta \bar{\mathbf{z}} = \mathbf{G} \cdot \boldsymbol{\beta} \quad (3)$$

where  $\Delta \bar{\mathbf{z}}$  is the vector of changes in the mean trait values,  $\mathbf{G}$  is the genetic covariance matrix and  $\boldsymbol{\beta}$  is the vector of standardised selection gradients that regress each trait on relative fitness (i.e.,  $\boldsymbol{\omega} = \boldsymbol{\alpha} + \boldsymbol{\beta}^T \cdot \mathbf{z} + \mathbf{e}$ ; [101]). Plasticity can be captured by mapping trait development to the environment and incorporating environmental variability into the breeders equation [see also, 102]. While the multivariate breeders equation can be useful for predicting short-term evolutionary responses (i.e., one or a few generations) it

likely has limited predictive power over longer timespans because of changes in heritability and selection through time, which is a future challenge [95,101].

## Concluding Remarks

Implementing a systems-modelling approach will no doubt be challenging, particularly for complex communities, and knowledge gaps remain (see **Outstanding Questions**). But, more than ever, we need mechanistic approaches that can capture key biological processes (physiology, behaviour, phenology, species interactions and eco-evolutionary dynamics - **Box 3**) to better predict biological responses to extreme heat [54,103]. Coupled mechanistic models within and across species are expected to have improved predictive power when projected to new environmental conditions and will better capture interacting processes [21,103]. Nonetheless, application of a systems modelling approach will require a solid foundation of species natural history, diverse modelling expertise, and a concerted effort to collect and collate trait and environmental data for model building and validation.

There are exciting new tools and data that are helping to overcome these challenges [26], and careful consideration of the key processes important to a system can help alleviate these challenges [21]. For example, large databases of physiological traits for both plants and animals now exist [e.g., 104,105] and advanced missing data approaches can help estimate likely parameters for data deficient species [70,74,106]. Ever more sophisticated and powerful computational pipelines (e.g. *NicheMapR*, *TrenchR*, *mcera5*, *terra* in R) make it easier to implement and connect models. Model selection and validation are important steps in any modelling process and will be more complicated when applied to entire systems; however, starting with a simple model, validating predictions with empirical data, and then adding complexity as needed can make building systems models more tractable [21]. Even simple models that compare counterfactual scenarios may provide important quantitative insights into the impacts of extreme heat on organisms, populations and communities [107–109]. Such insights can form the basis for prediction-driven solutions to mitigate the impacts of extreme heat on biodiversity. By working collaboratively we can develop a more quantitative and predictive understanding of the impacts that extreme heat will have on organisms, populations, and communities, and decide how to mitigate these impacts to preserve diverse systems now and into the future.

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## References

1. Barriopedro, D. *et al.* (2023) Heat waves: Physical understanding and scientific challenges. *Rev. Geophys.* 61, e2022RG000780
2. Perkins-Kirkpatrick, S. and Lewis, S.C. (2020) Increasing trends in regional heatwaves. *Nat. Comm.* 11, 3357
3. Evans, M.E.K. *et al.* (2025) Scaling plant responses to heat: From molecules to the biosphere. *Science* 388, 1167–1173
4. Margalef-Marrase, J. *et al.* (2020) Relationship between heatwave-induced forest die-off and climatic suitability in multiple tree species. *Glob. Chang. Biol.* 26, 3134–3146
5. Kazenel, M.R. *et al.* (2024) Heat and desiccation tolerances predict bee abundance under climate change. *Nature* 628, 342–348
6. Zhao, C. *et al.* (2017) Temperature increase reduces global yields of major crops in four independent estimates. *Proc. Natl. Acad. Sci. U. S. A.* 114, 9326–9331
7. Tito, R. *et al.* (2018) Global climate change increases risk of crop yield losses and food insecurity in the tropical andes. *Glob. Chang. Biol.* 24, e592–e602
8. Bernacchi, C.J. *et al.* (2025) Safeguarding crop photosynthesis in a rapidly warming world. *Science* 388, 1153–1160
9. Martinez, L.D. *et al.* (2023) Effect of simulated heat waves on the behaviour of two mirid predators. *J. Appl. Entomol.* 147, 486–498
10. Rezende, E.L. *et al.* (2020) Predicting temperature mortality and selection in natural *Drosophila* populations. *Science* 369, 1242–1245
11. Ørsted, M. *et al.* (2024) Thermal limits of survival and reproduction depend on stress duration: A case study of *Drosophila suzukii*. *Ecol. Lett.* 27, e14421

326 12. López-Goldar, X. *et al.* (2024) Heat waves induce milkweed resistance to a specialist herbivore via  
327 increased toxicity and reduced nutrient content. *Plant Cell Environ.* 47, 4530–4542

328 13. Arnold, P.A. *et al.* (2025) A framework for modelling thermal load sensitivity across life. *Glob.*  
329 *Chang. Biol.* 31, e70315

330 14. Trivedi, P. *et al.* (2022) Plant–microbiome interactions under a changing world: Responses,  
331 consequences and perspectives. *New Phytol.* 234, 1951–1959

332 15. Bimler, M.D. *et al.* (2024) Plant interaction networks reveal the limits of our understanding of  
333 diversity maintenance. *Ecol. Lett.* 27, e14376

334 16. Chesson, P. (2000) General theory of competitive coexistence in spatially-varying environments.  
335 *Theor. Popul. Biol.* 58, 211–237

336 17. Dicks, L.V. *et al.* (2021) A global-scale expert assessment of drivers and risks associated with  
337 pollinator decline. *Nat. Ecol. Evol.* 5, 1453–1461

338 18. Kooijman, S.A.L.M. (2010) Dynamic energy budget theory for metabolic organisation, Cambridge  
339 University Press.

340 19. Kearney, M. and Porter, W. (2009) Mechanistic niche modelling: Combining physiological and  
341 spatial data to predict species’ ranges. *Ecol. Lett.* 12, 334–350

342 20. Kearney, M.R. and Leigh, A. (2024) Fast, accurate and accessible calculations of leaf temperature  
343 and its physiological consequences with NicheMapR. *Methods Ecol. Evol.* 15, 1516–1531

344 21. Briscoe, N.J. *et al.* (2023) Mechanistic forecasts of species responses to climate change: The promise  
345 of biophysical ecology. *Glob. Chang. Biol.* 29, 1451–1470

346 22. Rezende, E.L. *et al.* (2014) Tolerance landscapes in thermal ecology. *Funct. Ecol.* 28, 799–809

347 23. Bimler, M.D. *et al.* (2018) Accurate predictions of coexistence in natural systems require the  
348 inclusion of facilitative interactions and environmental dependency. *J. Ecol.* 106, 1839–1852

349 24. Bowler, C.H. *et al.* (2022) Accounting for demographic uncertainty increases predictions for species  
350 coexistence: A case study with annual plants. *Ecol. Lett.* 25, 1618–1628

351 25. Kearney, M.R. and Porter, W.P. (2017) NicheMapR – an R package for biophysical modelling: The  
352 microclimate model. *Ecography (Cop.)* 40, 664–674

353 26. Maclean, I.M.D. *et al.* (2019) Microclima: An R package for modelling meso- and microclimate.  
354 *Methods Ecol. Evol.* 10, 280–290

355 27. Kearney, M.R. *et al.* (2020) A method for computing hourly, historical, terrain-corrected  
356 microclimate anywhere on earth. *Methods Ecol. Evol.* 11, 38–43

357 28. Maclean, I.M.D. *et al.* (2021) On the measurement of microclimate. *Methods Ecol. Evol.* 12, 1397–  
358 1410

359 29. Buckley, L.B. *et al.* (2023) TrenchR: An R package for modular and accessible microclimate and  
360 biophysical ecology. *PLOS Clim* 2, e0000139

361 30. Bramer, I. *et al.* (2018) Advances in monitoring and modelling climate at ecologically relevant scales.  
362 *Adv. Ecol. Res.* 58, 101–161

363 31. Levy, O. *et al.* (2016) A dynamically downscaled projection of past and future microclimates.  
364 *Ecology* 97, 1888–1888

365 32. Choi, F. *et al.* (2019) Mapping physiology: Biophysical mechanisms define scales of climate change  
366 impacts. *Conserv. Physiol.* 7, coz028

367 33. Lorenz, R. *et al.* (2010) Persistence of heat waves and its link to soil moisture memory. *Geophys. Res.*  
368 *Lett.* 37, L09703

369 34. De Frenne, P. *et al.* (2024) Ten practical guidelines for microclimate research in terrestrial  
370 ecosystems. *Methods Ecol. Evol.* 16, 269–294

371 35. Meyer, A.V. *et al.* (2023) A guide and tools for selecting and accessing microclimate data for  
372 mechanistic niche modeling. *Ecosphere* 14, e4506

373 36. Klinges, D.H. *et al.* (2025) Matching climate to biological scales. *Trends Ecol. Evol.* in press

374 37. Hirsch, A.L. *et al.* (2019) Amplification of australian heatwaves via local land-atmosphere coupling.  
375 *J. Geophys. Res.* 124, 13625–13647

376 38. Drake, J.E. *et al.* (2018) Trees tolerate an extreme heatwave via sustained transpirational cooling and  
377 increased leaf thermal tolerance. *Glob. Chang. Biol.* 24, 2390–2402

378 39. Holzworth, D. *et al.* (2018) APSIM next generation: Overcoming challenges in modernising a  
379 farming systems model. *Environ. Model. Softw.* 103, 43–51

380 40. Medlyn, B.E. *et al.* (2011) Reconciling the optimal and empirical approaches to modelling stomatal  
381 conductance. *Glob. Chang. Biol.* 17, 2134–2144

382 41. Porter, W.P. *et al.* (1973) Behavioral implications of mechanistic ecology : Thermal and behavioral  
383 modeling of desert ectotherms and their microenvironment. *Oecologia* 13, 1–54

384 42. Porter, W.P. and Gates, D.M. (1969) Thermodynamic equilibria of animals with environment. *Ecol.*  
385 *Monogr.* 39, 227–244

386 43. Riddell, E.A. *et al.* (2017) Physical calculations of resistance to water loss improve predictions of  
387 species range models. *Ecol. Monogr.* 87, 21–33

388 44. Duursma, R.A. (2015) Plantecophys—an R package for analysing and modelling leaf gas exchange  
389 data. *PLoS One* 10, e0143346

390 45. Wild, K.H. *et al.* (2025) Climate change and the cost-of-living squeeze in desert lizards. *Science* 387,  
391 303–309

392 46. Briscoe, N.J. *et al.* (2022) Too hot to hunt: Mechanistic predictions of thermal refuge from cat  
393 predation risk. *Conserv. Lett.* 15, e12906

394 47. Duursma, R.A. *et al.* (2019) On the minimum leaf conductance: Its role in models of plant water use,  
395 and ecological and environmental controls. *New Phytol.* 221, 693–705

396 48. Loughran, C.L. and Wolf, B.O. (2020) The functional significance of panting as a mechanism of  
397 thermoregulation and its relationship to the critical thermal maxima in lizards. *J. Exp. Biol.* 223, jeb224139

398 49. Sharpe, L.L. *et al.* (2022) In the hot seat: Behavioral change and old-growth trees underpin an  
399 australian songbird’s response to extreme heat. *Front. Ecol. Evol.* 10, 813567

400 50. Posch, B.C. *et al.* (2024) Intensive leaf cooling promotes tree survival during a record heatwave.  
401 *Proc. Natl. Acad. Sci. U. S. A.* 121, e2408583121

402 51. Ørsted, M. *et al.* (2022) Finding the right thermal limit: A framework to reconcile ecological,  
403 physiological and methodological aspects of CTmax in ectotherms. *J. Exp. Biol.* 225, jeb244514

404 52. Jagadish, S.V.K. *et al.* (2021) Plant heat stress: Concepts directing future research. *Plant Cell*  
405 *Environ.* 44, 1992–2005

406 53. Faber, A.H. *et al.* (2024) Application of the thermal death time model in predicting thermal damage  
407 accumulation in plants. *J. Exp. Bot.* 75, 3467–3482

408 54. Martínez-De León, G. and Thakur, M.P. (2024) Ecological debts induced by heat extremes. *Trends*  
409 *Ecol. Evol.* 39, 1024–1034

410 55. Marchin, R.M. *et al.* (2022) Extreme heat increases stomatal conductance and drought-induced  
411 mortality risk in vulnerable plant species. *Glob. Chang. Biol.* 28, 1133–1146

412 56. Youngblood, J.P. *et al.* (2025) Dehydration worsens heat tolerance of locusts and amplifies predicted  
413 impacts of climate change. *Funct. Ecol.* 39, 1194–1207

414 57. Mengutay, M. *et al.* (2013) Adequate magnesium nutrition mitigates adverse effects of heat stress on  
415 maize and wheat. *Plant Soil* 368, 57–72

416 58. Andersen, L.H. *et al.* (2010) Protein and carbohydrate composition of larval food affects tolerance to  
417 thermal stress and desiccation in adult drosophila melanogaster. *J. Insect Physiol.* 56, 336–340

418 59. Chouhan, D. *et al.* (2023) Plant-microbe interaction and their role in mitigation of heat stress. In  
419 *Rhizosphere biology*, pp. 127–147, Springer Nature Singapore

420 60. Márquez, L.M. *et al.* (2007) A virus in a fungus in a plant: Three-way symbiosis required for thermal  
421 tolerance. *Science* 315, 513–515

422 61. Hoang, K.L. *et al.* (2019) The effects of bacillus subtilis on caenorhabditis elegans fitness after heat  
423 stress. *Ecol. Evol.* 9, 3491–3499

424 62. Jeffrey, S.J. *et al.* (2001) Using spatial interpolation to construct a comprehensive archive of  
425 australian climate data. *Environ. Model. Softw.* 16, 309–330

426 63. Kearney, M.R. and Porter, W.P. (2020) NicheMapR – an R package for biophysical modelling: The  
427 ectotherm and dynamic energy budget models. *Ecography (Cop.)* 43, 85–96

428 64. Nisbet, R. *et al.* (2000) From molecules to ecosystems through dynamic energy budget models.  
429 *Journal of Animal Ecology* 69, 913–926

430 65. Jager, T. (2020) Revisiting simplified DEBtox models for analysing ecotoxicity data. *Ecol. Modell.*  
431 416, 108904

432 66. Kearney, M.R. *et al.* (2013) Balancing heat, water and nutrients under environmental change: A  
433 thermodynamic niche framework. *Funct. Ecol.* 27, 950–966

434 67. Kearney, M. *et al.* (2010) Modelling the ecological niche from functional traits. *Philos. Trans. R. Soc.*  
435 *Lond. B Biol. Sci.* 365, 3469–3483



436 68. Marques, G.M. *et al.* (2018) The AmP project: Comparing species on the basis of dynamic energy  
437 budget parameters. *PLoS Comput. Biol.* 14, e1006100

438 69. Kooijman, S.A.L.M. *et al.* (2021) Multidimensional scaling for animal traits in the context of  
439 dynamic energy budget theory. *Conserv. Physiol.* 9, coab086

440 70. Bruggeman, J. *et al.* (2009) PhyloPars: Estimation of missing parameter values using phylogeny.  
441 *Nucleic Acids Research* 37, W179–W184

442 71. Bonan, G. (2019) Climate change and terrestrial ecosystem modeling. Cambridge University Press

443 72. Kumarathunge, D.P. *et al.* (2019) Acclimation and adaptation components of the temperature  
444 dependence of plant photosynthesis at the global scale. *New Phytol.* 222, 768–784

445 73. De Kauwe, M.G. *et al.* (2020) Identifying areas at risk of drought-induced tree mortality across south-  
446 eastern australia. *Glob. Chang. Biol.* 26, 5716–5733

447 74. Bimler, M.D. *et al.* (2023) Estimating interaction strengths for diverse horizontal systems using  
448 performance data. *Methods Ecol. Evol.* 14, 968–980

449 75. Kraft, N.J.B. *et al.* (2015) Plant functional traits and the multidimensional nature of species  
450 coexistence. *Proc. Natl. Acad. Sci. U. S. A.* 112, 797–802

451 76. Buche, L. *et al.* (2024) Multitrophic higher-order interactions modulate species persistence. *Am. Nat.*  
452 203, 458–472

453 77. Buche, L. *et al.* (2025) A continuum from positive to negative interactions drives plant species’  
454 performance in a diverse community. *Ecol. Lett.* 28, e70059

455 78. Ricker, W.E. (1954) Stock and recruitment. *J. Fish. Res. Board Can.* 11, 559–623

456 79. Beverton, R.J.H. and Holt, S. (2012) On the dynamics of exploited fish populations. Springer

457 80. Levine, J.M. and HilleRisLambers, J. (2009) The importance of niches for the maintenance of species  
458 diversity. *Nature* 461, 254–257

459 81. Wainwright, C.E. *et al.* (2019) Distinct responses of niche and fitness differences to water availability  
460 underlie variable coexistence outcomes in semi-arid annual plant communities. *J. Ecol.* 107, 293–306

461 82. Mayfield, M. and Stouffer, D.B. (2017) Higher-order interactions capture unexplained complexity in  
462 diverse communities. *Nat. Ecol. Evol.* 1, 0062

463 83. Martyn, T.E. *et al.* (2021) Identifying “useful” fitness models: Balancing the benefits of added  
464 complexity with realistic data requirements in models of individual plant fitness. *Am. Nat.* 197, 415–433

465 84. Llandres, A.L. *et al.* (2015) A dynamic energy budget for the whole life-cycle of holometabolous  
466 insects. *Ecol. Monogr.* 85, 353–371

467 85. Arnold, P.A. *et al.* (2019) Sparse evidence for selection on phenotypic plasticity in response to  
468 temperature. *Philos. Trans. R. Soc. Lond. B Biol. Sci.* 374, 20180185

469 86. Shekhawat, K. *et al.* (2022) Beat the heat: Plant- and microbe-mediated strategies for crop  
470 thermotolerance. *Trends Plant Sci.* 27, 802–813

471 87. Dandy, J. *et al.* (2019) Leaving home: Place attachment and decisions to move in the face of  
472 environmental change. *Reg. Environ. Change* 19, 615–620

473 88. Leviston, Z. *et al.* (2018) Linkages between ecosystem services and human wellbeing: A nexus webs  
474 approach. *Ecol. Indic.* 93, 658–668

475 89. Baeckens, S. and Donihue, C.M. (2025) Evolutionary consequences of extreme climate events. *Curr.*  
476 *Biol.* 35, R850–R864

477 90. Urban, M.C. (2024) Climate change extinctions. *Science* 386, 1123–1128

478 91. Muñoz-Sabater, J. *et al.* (2021) ERA5-land: A state-of-the-art global reanalysis dataset for land  
479 applications. *Earth Syst. Sci. Data* 13, 4349–4383

480 92. Eyring, V. *et al.* (2016) Overview of the coupled model intercomparison project phase 6 (CMIP6)  
481 experimental design and organization. *Geosci. Model Dev.* 9, 1937–1958

482 93. Chevin, L.-M. and Hoffmann, A.A. (2017) Evolution of phenotypic plasticity in extreme  
483 environments. *Philos. Trans. R. Soc. Lond. B Biol. Sci.* 372, 20160138

484 94. Hoffmann, A.A. and Bridle, J. (2022) The dangers of irreversibility in an age of increased  
485 uncertainty: Revisiting plasticity in invertebrates. *Oikos* 2022, e08715

486 95. Hoffmann, A.A. and Sgrò, C.M. (2011) Climate change and evolutionary adaptation. 470, 479–485

487 96. Catullo, R.A. *et al.* (2019) The potential for rapid evolution under anthropogenic climate change.  
488 *Curr. Biol.* 29, R996–R1007

489 97. Ghalambor, C. *et al.* (2015) Non-adaptive plasticity potentiates rapid adaptive evolution of gene  
490 expression in nature. *Nature* 525, 372–375

491 98. Sousa, T. *et al.* (2010) Dynamic energy budget theory restores coherence in biology. *Philos. Trans. R.*  
492 *Soc. Lond. B Biol. Sci.* 365, 3413–3428

493 99. Kearney, M. *et al.* (2009) Integrating biophysical models and evolutionary theory to predict climatic  
494 impacts on species' ranges: The dengue mosquito *aedes aegypti* in australia. *Funct. Ecol.* 23, 528–538

495 100. Lande, R. (1979) Quantitative genetic analysis of multivariate evolution, applied to brain:body size  
496 allometry. *Evolution* 33, 402–416

497 101. Walsh, B. and Lynch, M. (2018) Evolution and selection of quantitative traits. Oxford University  
498 Press

499 102. McGlothlin, J. and Galloway, L. (2014) The contribution of maternal effects to selection response:  
500 An empirical test of competing models. *Evol.* 68, 549–558

501 103. Urban, M.C. *et al.* (2016) Improving the forecast for biodiversity under climate change. *Science* 353,  
502 aad8466–aad8466

503 104. Kattge, J. *et al.* (2020) TRY plant trait database - enhanced coverage and open access. *Glob. Chang.*  
504 *Biol.* 26, 119–188

505 105. Leiva, F.P. *et al.* (2025) ShareTrait: Towards interoperable and reusable individual trait-based data in  
506 ectotherms. *Funct. Ecol.* 39, 3124–3138

507 106. Weiss-Lehman, C.P. *et al.* (2022) Disentangling key species interactions in diverse and  
508 heterogeneous communities: A bayesian sparse modelling approach. *Ecol. Lett.* 25, 1263–1276

509 107. Dudley, J. *et al.* (2025) A causal inference framework for climate change attribution in ecology.  
510 *Ecol. Lett.* 28, e70192

511 108. Carroll, T. *et al.* (2023) Biodiversity change under adaptive community dynamics. *Glob. Chang. Biol.*  
512 29, 3525–3538

513 109. Dornhaus, A. *et al.* (2022) How can we fully realize the potential of mathematical and biological  
514 models to reintegrate biology? *Integr. Comp. Biol.* 61, 2244–2254