

1 **Forest composition and diversity buffer microclimates and enhance**
2 **productivity**

3
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29 **Author Contributions:** J.C.-B. designed and established the FAB2 experiment. M.H.P designed
30 this study with input from J.C.-B. M.H.P. and H.D constructed data loggers and operationalized
31 sensors for microclimate and soil moisture. M.H.P and A.S. collected physiological and spectral
32 data. A.S. collected forest inventory data. J.A.G.Q. collected and processed LiDAR point clouds
33 and ran models for spectral analysis. MHP performed statistical analyses and modeling with
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36 **Data Availability Statement**

37 Data and code are available at

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40 **Conflict of Interest**

41 The authors declare no conflict of interest.

42 **ABSTRACT**

43 Trees can buffer forest ecosystems against climate extremes, creating microenvironments that
44 support diverse plant communities. We asked how diverse tree communities could influence the
45 above- and below-ground moisture environment, thereby linking tree diversity and ecosystem
46 function. To address this question, we integrated remotely sensed forest structure, growth
47 inventories, microclimate measurements, and leaf-level physiology in a forest biodiversity
48 experiment. We found that fast-growing, early successional tree species buffered daily
49 amplitudes of vapor pressure deficits (VPD) within forest plots, increasing the annual wood
50 production of slow-growing, shade tolerant tree species. In assemblages with more stable
51 microclimates, several tree species showed increased leaf water content and reduced
52 physiological stress. Plots with reduced amplitude in VPD had greater complementarity and
53 overyielding relative to those with higher variability. A mechanistic understanding of how tree
54 diversity and community composition contribute to VPD amplitude and overyielding can inform
55 conservation and restoration efforts in a rapidly changing world.

56

57 1 | INTRODUCTION

58

59 Biodiversity supports the cycling of energy, nutrients, and water through ecosystems (Schuldt et
60 al. 2018; Tilman et al. 2014). Climate change, habitat loss, and a suite of compounding factors
61 are causing rapid declines in biodiversity (Bellard et al. 2012; Isbell et al. 2023; Segan et al.
62 2016). Informed stewardship of landscapes can sustain ecosystems under increasingly extreme
63 environments (Hong et al. 2022; Wright et al. 2017).

64

65 Relationships between biodiversity and ecosystem function are well established in both
66 experimental and observational systems (Liu et al. 2018; Mori et al. 2018; Schnabel et al. 2021;
67 Tilman et al. 1996), yet the mechanisms by which forest ecosystems become productive beyond
68 the sum of their parts remain poorly understood (Williams et al. 2017). We have only begun to
69 discern how biodiversity changes local environments and contributes to ecosystem functioning
70 (Guzmán et al. 2025; Isbell et al. 2015; Uργοiti et al. 2022; Zheng et al. 2024). Improving our
71 understanding of the mechanisms by which biodiversity results in overyielding will fill critical
72 gaps in ecological theory and enhance urgently needed sustainable forestry, conservation, and
73 restoration efforts.

74

75 Coexistence theory explains how diverse species can survive together in an environment with
76 shared resources (Chesson 2000; HilleRisLambers et al. 2012; Levine et al. 2017; Mayfield and
77 Levine 2010; Tilman 1980). Studies have described how organisms engage in direct competition
78 or respond to their abiotic environment, often without simultaneously testing how organisms
79 shape their local environment and thereby influence neighbors and drive ecosystem functions.

80 We designed the present study to examine how organisms alter their local microclimate, which
81 in turn impacts the functioning of organisms. This mechanistic approach to studying coexistence
82 and biodiversity-ecosystem function examines the feedback loops between abiotic and biotic
83 factors in an ecosystem (Goldberg 2012). Interactions between species and the local environment
84 can influence individual growth and scale up to the ecosystem level (Goldberg 2012; Hacker and
85 Gaines 1997). Complementary among tree crowns of different species has been shown to result
86 in overyielding, a critical net biodiversity effect based on productivity (Loreau and Hector 2001;
87 Williams et al. 2017). In addition, complementarity in resource usage and facilitation between
88 species can enhance the survival and growth of trees, particularly under drought conditions
89 (Blondeel et al. 2024; Kothari et al. 2021).

90
91 Warmer and drier conditions are increasing in many parts of the world, increasing vapor pressure
92 deficits (VPD) and negatively impacting the growth and reproduction of numerous plant species
93 (López et al. 2021). The ability of plants to reduce fluctuations in VPD may be critical to the
94 survival of species that are adapted to live in cool and wet environments, and those at the rear-
95 edges of their growing zones (Maclean et al. 2015; Sánchez-Salguero et al. 2017). Plants can act
96 as ecosystem engineers in their local environments through shading and restricting air
97 movement, reducing variation in VPD and decreasing solar irradiation available to neighbor
98 plants (Connell 1983; Gilad et al. 2004; Holmgren et al. 1997; Wright et al. 2014; Zhang et al.
99 2024). Recent research shows how tree composition, diversity, and structure can buffer
100 microclimates (Schnabel et al. 2025, Wang et al. 2025). What is not yet clear is how tree
101 diversity and composition influence microclimate variation, and how this variation influences
102 productivity in forest communities (De Frenne et al. 2021; Losapio et al. 2021).

103
104 Long term, large-scale biodiversity experiments provide opportunities for testing the
105 mechanisms linking biodiversity and ecosystem function. In this study, we utilized the
106 ecosystem-scale Forest and Biodiversity Experiment 2 (FAB2) (Cavender-Bares et al. 2024)
107 after 8 years of growth as a platform to examine potential drivers of overyielding in forest
108 ecosystems (Fig. 1). We asked how tree diversity and community composition may influence
109 above-ground water availability (VPD), which in turn may influence tree physiology and
110 overyielding. First, we hypothesized that tree diversity and **fast-growing early successional tree**
111 (FAST) species – acting as ecosystem engineers – increase forest cover due to their resource-
112 acquisitive strategies (Zheng et al. 2024). Second, we hypothesized that more forest cover
113 decreases local light availability and reduces microclimate variability. Third, we hypothesized
114 that buffered microclimates – which we define as dampened daily amplitude in VPD – arising
115 from greater forest cover reduce leaf-level stress, increase productivity, and contribute to
116 complementarity.

117

118 **2 | METHODS**

119

120 **2.1 | Site Description and Study Species**

121 We conducted research in the Forest and Biodiversity experiment (FAB2) at Cedar Creek
122 Ecosystem Science Reserve, located in Anoka County, Minnesota (45.406299, -93.191309). The
123 soil is characterized as sandy with limited nutrient availability and is classified as a mixed frigid
124 Lamellic Udipsamment on the Zimmerman series (SoilWeb: An Online Soil Survey 2025).
125 Established in 2016, the long-term experiment includes 12 tree species native to Minnesota and

126 was designed to tease apart the influences of species richness, functional diversity, functional
127 variability, phylogenetic diversity, and phylogenetic variability on forest ecosystem processes
128 (Cavender-Bares et al. 2024). Phylogenetic species variability differs from phylogenetic diversity
129 by capturing evolutionary differences independently of species richness. The represented
130 angiosperm species included *Tilia americana*, *Acer rubrum*, *Acer negundo*, *Quercus rubra*, *Q.*
131 *ellipsoidalis*, *Q. alba*, *Q. macrocarpa*, and *Betula papyrifera*. The gymnosperm species included
132 *Juniperus virginiana*, *Pinus banksiana*, *P. resinosa*, and *P. strobus*. More details on how
133 dimensions of tree diversity were calculated and the experimental design of FAB2 can be found
134 in Cavender-Bares et al. 2024. For this study, we studied 148 100 m² (10 m × 10 m) plots that
135 were originally planted with 100 trees each, including 36 monocultures, 46 bi-cultures, 46 6-
136 species polycultures, 10 6-species polycultures, and 10 12-species polycultures. We chose a
137 subset of those plots spanning the range of multiple dimensions of tree diversity to install sensors
138 and conduct physiology measurements (Supplementary Table 1).

139
140 The proportion of fast-growing early successional tree (FAST) species was selected as a measure
141 of community composition to understand the potential role of FAST species as ecosystem
142 engineers that could shape local environmental conditions, thereby influencing the growth and
143 physiology of slower growing tree species (Kothari et al. 2021). FAST proportion was calculated
144 as the fraction of wood volume in a plot composed of pines (*Pinus*) and birches (*Betula*), as these
145 species were able to grow swiftly despite the nutrient-poor, sandy soil of the experimental site
146 (Fig. 1, Goldberg 1982). To understand the role of gymnosperms in determining ecosystem
147 function, the proportion of gymnosperms was calculated as the fraction of wood volume in a plot
148 composed of gymnosperm species.

149

150 **2.2 | Forest Structure and Environment**

151

152 To determine how tree diversity and composition affected forest structure, we performed an
153 Uncrewed Aerial Vehicle LiDAR survey (UAV-LiDAR) over FAB2 on July 9, 2024 to derive
154 tridimensional point clouds of LiDAR returns. Following the methods of Guzmán et al. 2025,
155 plant fractional cover (FC) was computed for each plot as a ratio of LiDAR returns from
156 vegetative structure to LiDAR returns from the ground. FC is a descriptor of light attenuation
157 and vegetation density, where values close to 1 represent plots with higher vegetation density or
158 lower light availability at the forest floor than values close to 0 (Boucher et al. 2023; Guzmán Q.
159 et al. 2025; Venier et al. 2019).

160

161 To study associations of forest composition, diversity and FC with microclimate, we deployed
162 microclimate sensors to monitor air and soil conditions within experimental plots. Between April
163 and May 2024, we installed SensorPush HT.w microclimate sensors (SensorPush, NY, USA) in
164 the centers of 75 selected plots (Supplementary Table 1). The sensors measured air temperature,
165 relative humidity, and vapor pressure deficit (VPD) at 1-minute intervals and were housed within
166 AcuRite solar radiation shields (AcuRite, WI, USA) attached to shepherd hooks (Fig. 1). The
167 sensors were suspended at an average height of 84.2 cm, aligning with the average heights of
168 slower growing tree species in monocultures. The minimum, mean, median, 95th percentile,
169 range, and amplitude for each microclimate metric were calculated across the field sampling
170 days (July 16 - July 19, 2024). Sampling days were chosen to represent the peak growing season.
171 The 95th percentile was used to approximate the maximum because logged maximum

172 microclimate values may have been influenced by excess solar irradiation and ground re-
173 radiation. The amplitudes of temperature, relative humidity, and VPD (VPD_{amp}) were calculated
174 by fitting a sine curve to the daily fluctuations of each microclimate metric and extracting the
175 amplitude coefficients. VPD_{amp} and range of VPD were highly correlated ($R^2 = 0.99$, $p < 0.001$),
176 but VPD_{amp} was selected as a representative measure of microclimate due to its strong
177 correlations with FC.

178

179 On April 12, 2024, we installed 24 GS 1 and TEROS 10 soil moisture sensors at the centers of
180 experimental plots to pair with aboveground microclimate measurements (Supplementary Table
181 1) (METER, WA, USA). Soil moisture calibration curves were created for both sensor types by
182 correlating sensor voltage readings to known volumetric water contents (VWC) (Supplementary
183 Table 2). Data loggers attached to the sensors recorded VWC on 30-minute intervals from April
184 12 to September 3, 2024 (Chen et al. 2025). Statistics of minimum, mean, median, maximum,
185 and range (VWC_{range}) were calculated for VWC in each plot, corresponding to the month of July
186 2024. VWC statistics were averaged over the month due to staggered data logger malfunctions.

187

188 To determine how FC was associated with light availability in the plots, we used an AccuPAR
189 LP-80 photometer (METER, WA, USA) to measure photosynthetically active radiation (PAR).
190 PAR measurements were taken between 10:00 and 15:00 h on cloudless days from July 11 – 22,
191 2024 within experimental plots in which soil moisture was monitored. PAR was measured at the
192 height of microclimate sensors (~84.2 cm) in the northeast and southwest quadrants of each plot.
193 The within-plot PAR measurements were divided by PAR measurements of open sky conditions
194 outside of the plot, resulting in ratios characterizing PAR transmission through the forest canopy.

195

196 **2.3 | Plant Physiological Measurements**

197

198 We studied four slow-growing deciduous tree species (*Tilia americana*, *Acer rubrum*, *Quercus*
199 *alba*, *Quercus rubra*) to understand how changes in microenvironmental conditions could
200 influence tree physiology. During the peak of the 2024 growing season, we measured sun-
201 exposed leaves in the upper canopies of three trees of each focal species per plot, unless fewer
202 than three individuals were alive in the selected plot. Measurements were averaged to generate
203 an average metric for each species per plot.

204

205 Gas exchange measurements (photosynthesis, stomatal conductance) were taken on July 17 and
206 July 19, 2024 between 9:00 and 13:00 h using a LI-COR 6800 (LI-COR, NE, USA). The LI-
207 COR 6800 was set with light levels at $1200 \mu\text{mol m}^{-2} \text{s}^{-1}$, flow rate at 500 m s^{-1} , fan speed at
208 10000 rpm, relative humidity at 60%, temperature at $20 \text{ }^\circ\text{C}$, and reference CO_2 at $400 \mu\text{mol}$
209 mol^{-1} . One leaf from each of the three trees was collected before sunrise, scanned, dried, and
210 weighed to calculate leaf mass per area (LMA). Leaf area was determined using ImageJ software
211 and the magick and LeafArea packages in R (Katabuchi 2015; Ooms 2016; R Core Team 2023;
212 Schneider et al. 2012).

213

214 Reflectance spectra measurements on leaves (400-2500 nm) were taken in the field using a PSR+
215 Spectroradiometer on July 16, 2024 (Spectral Evolution, MA, USA). The Carter Miller Stress
216 Index (CMSI), which gives information on water stress and the decline of photosynthetic
217 activity, was calculated using spectral reflectance values (R_{694} / R_{760} ; Carter and Miller 1994).

218 Based on methods in the literature, partial least squares regression models were developed from
219 leaf reflectance spectra of woody species across central Minnesota using the pls package
220 (Appendix S1; Kothari et al. 2023; Mevik and Wehrens 2007; Serbin et al. 2014). The
221 coefficients from the partial least squares regression models were used to derive leaf-level traits
222 of lignin and water content for selected trees.

223

224 **2.4 | Growth and Net Biodiversity Effects**

225

226 We used growth inventories of FAB2 to understand the productivity of tree species in plots with
227 differences in diversity, composition, and microclimate. For this study, we analyzed trees
228 measured at the ends of the 2023 and 2024 growing seasons. Growth metrics included height,
229 diameter at root collar, and diameter at breast height. Because diameter at breast height was not
230 measured in 2024 due to funding limitations, we developed species-specific regression models to
231 predict $\log_{10}(\text{diameter at breast height})$ using $\log_{10}(\text{diameter at root collar})$ and $\log_{10}(\text{height})$. The
232 wood volume of each tree was estimated using geometric equations for conoid and conoidoid
233 volumes as described in Cavender-Bares 2025 and Guzmán et al. 2025. Tree inventory data were
234 quality controlled by reviewing trees identified as having outlier volume and annual wood
235 productivity estimates.

236

237 Net biodiversity effects of forest plots were calculated for the 2024 growing season based on
238 equations from Loreau and Hector 2001 that partition biodiversity effects into complementarity
239 and selection. Following the methods of Guzmán et al. 2025, we used calculated tree volumes to
240 estimate annual wood productivity (AWP, $\text{m}^3 \text{y}^{-1}$) and net biodiversity effects of productivity,

241 hereafter called overyielding ($\text{m}^3 \text{y}^{-1} \text{ha}^{-1}$). The average AWP of monoculture plots was
242 considered the “expected” AWP for each species. The difference between observed and expected
243 AWP ($\text{AWP}_{\text{O-E}}$) was calculated for polycultures of each tree species.

244

245 **2.5 | Statistical Analysis**

246

247 To test our hypotheses, we ran multiple linear regressions in R between metrics of diversity,
248 community composition, forest structure, microenvironment, physiology, and overyielding (R
249 Core Team 2023). Quantile regressions were performed to model the relationships between plot
250 diversity metrics and FC using the quantreg package in R (Koenker 2025). Twenty-fifth quantile
251 regressions allowed us to examine how diversity related to FC while accounting for zero-inflated
252 diversity data (Fig. 2). Generalized additive models were used to observe changes in variance of
253 FC residuals as they related to diversity and were fit using the mgcv package (Wood 2011).

254

255 We regressed FC with VPD_{amp} to understand how forest structure influenced microclimate
256 variation. FC was transformed to reduce skew and linearize the relationship between FC and
257 VPD_{amp} . Transformed fractional cover (TFC) was calculated using the equation:

258

$$259 \quad \text{Transformed Fractional Cover} = -\text{sqrt}(1 - \text{FC}) \quad (1)$$

260

261 We examined how microclimate influenced the productivity of different tree species. Due to
262 limited replication of microclimate sensors in polycultures, we used FAST proportion, TFC, and

263 VPD_{amp} to create a multiple linear regression that predicted VPD_{amp} for all 148 100 m² plots in
264 the experiment. The R² for this regression was 0.8446, with p < 0.001.

265

$$266 \quad \text{VPD}_{\text{amp}} = 0.79986 - (0.43435 * \text{TFC}) - (0.26989 * \text{FAST Proportion}) \quad (2)$$

267

268 This model was used to infill VPD_{amp} values in plots without microclimate sensors. Infilled
269 VPD_{amp} values were used to examine relationships between VPD_{amp} and AWP_{O-E} of tree species
270 (Supplementary Fig. 2).

271

272 To understand the relationships between tree diversity, community composition, forest structure,
273 microclimate, and net biodiversity effects, we built a structural equation model using the lavaan
274 package in R (Rosseel 2012). Phylogenetic species variability describes evolutionary
275 dissimilarity and was previously shown to explain variation in FC (Guzmán et al. 2025) and was
276 therefore selected as a representative metric of diversity. FAST proportion was chosen as a
277 measure of community composition, TFC was included as a descriptor of forest structure,
278 VPD_{amp} was incorporated as a metric of microclimate, and overyielding was included as a
279 measure of net biodiversity effects.

280

281 **3 | RESULTS**

282

283 **3.1 | Tree diversity and community composition influence forest structure, reducing** 284 **variability in microclimates**

285

286 Greater tree diversity resulted in more canopy cover (i.e., fractional cover; FC) (Fig. 2). Quantile
287 regression analysis at the 25th percentile showed that tree communities with higher species
288 richness, functional diversity, phylogenetic diversity, and phylogenetic variability were
289 associated with increased FC ($p < 0.001$). Greater tree diversity also reduced variation in FC ($p <$
290 0.001). Plots with higher proportions of fast-growing early successional (FAST) species also
291 maintained higher FC (Supplementary Fig. 1, $p < 0.001$). In addition to increasing FC, FAST
292 proportion contributed to the reduction of microclimate variation (VPD_{amp}). FAST species had
293 significantly lower VPD_{amp} in monocultures compared to non-FAST species ($p < 0.001$), and
294 greater FAST proportion in plots significantly reduced VPD_{amp} ($R^2 = 0.76$, $p < 0.001$)
295

296 FC buffered the microclimate against large fluctuations in VPD_{amp} and decreased local light
297 availability. Increases in FC decreased the 95th percentile and mean VPD in plots ($p < 0.001$).
298 The daily amplitudes of temperature and relative humidity—drivers of VPD_{amp} —were both
299 strongly influenced by transformed fractional cover (TFC) ($p < 0.001$). However, a multiple
300 linear regression revealed that trends in VPD_{amp} were primarily driven by the amplitude of
301 relative humidity ($R^2 = 0.98$, $p < 0.001$) compared to temperature ($p > 0.05$). TFC was strongly
302 negatively correlated with VPD_{amp} (Fig. 3, $R^2 = 0.75$, $p < 0.001$). Included together in a multiple
303 linear regression, TFC and FAST proportion robustly predicted VPD_{amp} ($R^2 = 0.84$, $p < 0.001$).
304 Increases in FC decreased understory light availability ($R^2 = 0.70$, $p < 0.001$). A multiple linear
305 regression including VPD_{amp} and FC predicted PAR transmission well ($R^2 = 0.81$, $p < 0.001$).
306 There was a strong positive correlation between VPD_{amp} and PAR transmission within the
307 canopy, highlighting a potential tradeoff between microclimate buffering and light availability
308 (Fig. 3, $R^2 = 0.77$, $p < 0.001$).

309

310 **3.2 | Buffered microclimates and tree diversity are associated with more annual wood**
311 **production of some species**

312

313 VPD_{amp} showed a strong negative relationship with overyielding ($R^2 = 0.69$, $p < 0.001$),
314 indicating that lower values of VPD_{amp} were associated with more overyielding in forest plots
315 (Fig. 3). *T. americana*, *A. negundo*, and *P. banksiana* showed significant declines in annual
316 wood production (AWP_{O-E}) with increasing VPD_{amp} ($p < 0.05$). The AWP_{O-E} of the remaining
317 species was not strongly influenced by VPD_{amp} (Supplementary Fig. 2, $p > 0.05$).

318

319 Tree diversity also influenced the annual wood production of some species. Increases in
320 phylogenetic species variability resulted in AWP_{O-E} gains for *P. strobus* ($p < 0.05$), and *T.*
321 *americana* ($p = 0.051$). Increased phylogenetic diversity resulted in significantly decreased
322 AWP_{O-E} of *Q. ellipsoidalis* and *Q. macrocarpa* ($p < 0.05$). The AWP_{O-E} of *A. rubrum* increased
323 with higher phylogenetic diversity ($p < 0.05$), although a monoculture plot with exceptionally
324 high AWP caused AWP_{O-E} to be negative.

325

326 **3.3 | Deciphering mechanisms underlying Biodiversity-Ecosystem Function**

327

328 Overyielding resulted from the interactions between local biotic and abiotic factors in the forest
329 diversity experiment. Structural equation modeling showed that the proportion of FAST species
330 – rather than phylogenetic species diversity – was responsible for increased canopy cover (Fig. 4,
331 $p < 0.001$). Increases in FAST proportion and canopy cover reduced microclimate variation

332 (VPD_{amp}) within the forest plots during the peak growing season ($p < 0.001$). Buffered
333 microclimates influenced the annual wood production of tree species, as described above.
334 Overall, higher variation in microclimate negatively influenced overyielding ($p < 0.05$),
335 indicating a net positive relationship between buffered microclimates and overyielding (Fig. 4).
336 A linear regression between VPD_{amp} and overyielding corroborated the relationship (Fig. 3, $R^2 =$
337 0.69, $p < 0.001$). Canopy cover and FAST proportion had non-significant correlations with
338 overyielding ($p > 0.05$).

339

340 Although both complementarity and selection effects were present in the plots, trends in
341 overyielding were driven by complementarity effects (Supplementary Fig. 3). Compared to plots
342 with less than 50% FAST proportion, plots in which FAST species composed 50% or more of
343 the wood volume displayed higher complementarity effects ($p < 0.01$), more negative selection
344 effects ($p < 0.05$), and greater net biodiversity effects ($p < 0.001$).

345

346 **3.4 | Species exhibit physiological responses to variation in local environments modified by** 347 **tree diversity and composition**

348

349 Differences in plot-level environmental conditions caused responses in the physiology of four
350 focal tree species: *A. rubrum*, *Q. alba*, *Q. rubra*, and *T. americana*. Physiological responses to
351 variations in the local environment were largely species-specific, but some general trends were
352 also observed. Greater VPD_{amp} significantly decreased leaf water content in most focal species
353 (Fig. 5, $p < 0.01$ for *A. rubrum*, *Q. alba*, *Q. rubra*, $p < 0.1$ in *A. rubrum*), and greater leaf water
354 content was significantly correlated with a lower Carter Miller Stress Index (CMSI) in most focal

355 species ($p < 0.05$ for *A. rubrum*, *Q. alba*, *Q. rubra*, $p = 0.06$ in *T. americana*). Greater VPD_{amp}
356 was associated with higher LMA in *A. rubrum*, *Q. rubra* ($p < 0.05$), and *T. americana* ($p = 0.05$).
357 Increases in LMA were tightly linked to photosynthesis in *T. americana* ($p = 0.01$) and *Q. rubra*
358 ($p < 0.1$). Increased LMA of species was also highly correlated with greater leaf lignin content in
359 *Q. alba* ($p < 0.01$) and *T. americana* ($p < 0.1$). Leaf lignin content was negatively correlated with
360 CMSI in *A. rubrum* ($p < 0.01$) and *T. americana* ($p < 0.05$).

361
362 Higher proportions of gymnosperm species in the plot lowered mean and minimum VWC values
363 ($R^2 = 0.51$, $p < 0.001$; $R^2 = 0.52$, $p < 0.001$) and increased the range of VWC (VWC_{range})
364 (Supplementary Fig. 4, $R^2 = 0.20$, $p < 0.05$). Maximum VWC was not significantly influenced by
365 higher proportions of gymnosperm species ($p > 0.05$). VWC_{range} decreased moderately with more
366 PAR transmission in the plot ($R^2 = 0.15$, $p < 0.1$), but was not correlated with tree diversity
367 metrics or VPD_{amp} ($p > 0.05$). No strong correlations were observed between VWC_{range} and leaf
368 water content or LMA. However, increases in VWC_{range} significantly increased the average
369 stomatal conductance of *A. rubrum* (Fig. 5, $p < 0.05$). Increases in stomatal conductance
370 contributed to greater photosynthesis in only *T. americana* ($p < 0.01$).

371 372 **4 | DISCUSSION**

373
374 We found that tree species composition increased fractional cover more than diversity, fractional
375 cover decreased microclimate variation within the forest canopy, and more stable microclimates
376 contributed to increased forest productivity. The results from this study supported our three
377 hypotheses. Fast-growing early successional tree (FAST) species served as ecosystem engineers

378 to increase forest cover. Greater forest cover decreased light availability and reduced
379 microclimate variability, buffering against extreme microclimates, which in turn reduced
380 physiological stress in some species and increased overyielding. These findings highlight the
381 importance of forest composition and structure in shaping microclimates and demonstrate the
382 roles of facilitation and biotic-abiotic interactions in diverse ecosystems (Basnet et al. 2025;
383 Goldberg 2012; Wright et al. 2017). Increased complementarity and productivity in more diverse
384 plots with high proportions of FAST species contributed to higher overyielding.

385

386 **4.1 | Consequences of tree diversity and community composition**

387

388 We showed how tree diversity and especially community composition influenced forest
389 fractional cover (FC), supporting previous findings (Guzmán et al. 2025). The proportion of
390 FAST species had a stronger effect on FC than multiple dimensions of tree diversity,
391 highlighting the importance of community composition on forest structure. However,
392 phylogenetic species variability was significantly associated with overyielding, indicating that
393 diversity can influence ecosystem functions beyond microclimate buffering.

394

395 Considering the influence of tree diversity on the difference between observed and expected
396 annual wood production (AWP_{O-E}) for individual species can provide insight into potential BEF
397 mechanisms. The significant positive associations between AWP_{O-E} and phylogenetic species
398 variability suggest that the productivity of some species, such as *P. strobus* and *T. americana*,
399 was positively influenced by large evolutionary differences between these trees and their
400 neighbors. Phylogenetic diversity – but not phylogenetic variability – includes species richness

401 as a factor. Consequently, the significant relationships between AWP_{O-E} and phylogenetic
402 diversity show that a greater number of species in a community increased the productivity of *A.*
403 *rubrum*, but decreased the productivity of *Q. ellipsoidalis*, *Q. macrocarpa*. These findings
404 highlight the importance of understanding multiple aspects of diversity, such as species
405 variability and richness, which can have distinct consequences for different species.

406

407 Not all species increased AWP_{O-E} in response to greater tree diversity. Heightened interspecies
408 competition for resources could explain decreases in AWP_{O-E} of *Q. ellipsoidalis* and *Q.*
409 *macrocarpa*. These drought-tolerant *Quercus* species grew less in plots with more species,
410 indicating a preference for open canopies, consistent with their shade intolerance (Kothari et al.
411 2021) and frequent establishment in savanna ecosystems (Peterson and Reich 2001).

412

413 **4.2 | Buffered microclimates promote the growth of shade-tolerant and water-limited tree** 414 **species, contributing to overyielding**

415

416 Complementarity increased in plots with higher FAST proportion, which contributed to higher
417 fractional cover, and consequently decreased local light availability and reduced microclimate
418 variability. The reduced microclimate variation contributed to overyielding in forest plots. This
419 finding supports prior studies of gap-filling and crown complementarity between different tree
420 species (Williams et al. 2017) and highlights the facilitative effect of FAST species on other tree
421 species through microclimate buffering.

422

423 The negative relationship between microclimate variability (VPD_{amp}) and overyielding indicates
424 that reduced VPD_{amp} is associated with increased productivity of trees. With ongoing climate
425 change, microclimate buffering will become more important as species are pushed to the edges
426 of their climate envelopes (Koo et al. 2017). Previous research has highlighted the importance of
427 topography for climate refugia (Ackerly et al. 2020; Suggitt et al. 2018) and microclimate
428 buffering by plant cover (Bernath-Plaisted et al. 2023; De Frenne et al. 2021; De Lombaerde et
429 al. 2022).

430

431 In line with successional theory, FAST species may alter their local climate and environment in
432 ways that promote their own survival, including buffering of microclimate (Connell and Slatyer
433 1977). A FAST species, *P. banksiana*, showed significant declines in AWP_{O-E} associated with
434 increases in VPD_{amp} , indicating vulnerability of the species to drought (McCollum and Ibáñez
435 2020). Similar to the results from the higher density FAB1 experiment (Kothari et al. 2021),
436 shade tolerant tree species in FAB2 grew more under the canopy cover of their taller
437 counterparts. FAST proportion influenced the local microclimate by creating canopy cover that
438 provided shade and reduced evaporative demand (Lin 2010). Reduced VPD_{amp} was associated
439 with increased AWP_{O-E} of *T. americana* and *A. negundo*, which are shade tolerant angiosperms
440 found in moist sites and riparian ecosystems (Crow 1990; Niinemets and Valladares 2006; Ward
441 et al. 2002). Sitewide soil water limitation, vulnerability to drought, and shade tolerance
442 characteristics could explain how these tree species benefited from buffered microclimates that
443 alleviated atmospheric drought stress.

444

445 **4.3 | Buffered microclimates reduce leaf-level stress of focal species**

446
447 Reduced microclimate variation was associated with changes in leaf-level physiology indicating
448 lower levels of water and light stress. Variability in leaf form and function can be important
449 contributors to growth and overyielding (Chitwood and Sinha 2016; Rozendaal et al. 2006;
450 Williams et al. 2021). Although we did not observe strong correlations between the leaf traits
451 and growth in this study, possibly due to low sample size, the decreased leaf-level stress
452 resulting from reduced VPD_{amp} may have indirectly contributed to the growth of the focal trees.
453 Buffered microclimates resulted in higher leaf water content and reduced drought-related
454 photosynthetic stress for all focal species, as indicated by the CMSI (Carter and Miller 1994).
455 Trees living in microclimates buffered by FAST species had more water in their leaves and were
456 less stressed than their counterparts in more open environments.

457
458 Dampened VPD_{amp} significantly decreased LMA in multiple species (Fig. 5). Reduced VPD_{amp}
459 was strongly correlated with lower PAR transmission through the forest canopy, thus decreases
460 in LMA may have been responses to coupled declines in drought and light stress. Leaves with
461 higher LMA invest more in structural support mechanisms such as lignin and are better able to
462 withstand drought and other physical stressors (Wright et al. 2004). *Q. alba* and *T. americana*
463 increased leaf lignin content in leaves with higher LMA. Despite *Q. rubra* and *A. rubrum* not
464 displaying within-species relationships between LMA and leaf lignin, the overall positive
465 relationship between LMA and leaf lignin held true (Fig. 5). This result simultaneously
466 emphasizes species-specific differences in leaf traits and the ability of some species to shift their
467 resource allocation strategies in response to changes in their environment.

468

469 In this system, soil moisture variation (VWC_{range}) did not influence leaf-level physiology as
470 much as microclimate variation. Given that the proportion of gymnosperms did not significantly
471 influence the maximum VWC but significantly lowered the mean and minimum VWC in plots,
472 the relationship between gymnosperms and VWC_{range} may have been driven by increased water
473 uptake by well-established gymnosperm roots (Bond 1989). Redistribution of soil water by
474 gymnosperm roots may have alleviated drought stress in *A. rubrum*, increasing the species'
475 stomatal conductance (Caldwell et al. 1998). For *A. rubrum*, increases in stomatal conductance in
476 gymnosperm-dominated plots did not translate to increased photosynthesis, perhaps due to light
477 limitation. Despite the lack of intraspecific relationships between stomatal conductance and
478 photosynthesis in *Q. rubra*, *Q. alba*, and *A. rubrum*, the positive correlation across species
479 between these variables remained strong. These results highlight the differences between species
480 in physiological responses to environmental stress as well as consistent mechanisms governing
481 plant function that persist across diversity treatments and environmental conditions.

482

483 **4.4 | Conclusions**

484

485 Plant traits, evolutionary histories, and coexistence theory all inform understanding of species
486 interactions and growth dynamics. We show that plants exhibit complementarity in how they
487 influence and respond to the microenvironmental conditions they collectively shape. In this
488 study, we hypothesized and found strong connections between diversity, community
489 composition, forest cover, and microclimate variability (VPD_{amp}), with influences on plant
490 growth and overyielding. We thus demonstrated a clear mechanistic relationship between
491 decreased VPD_{amp} and overyielding, caused by tree diversity and composition. Our findings,

492 combined with previous work explaining how reduction in light stress increases productivity of
493 shade tolerant tree species (Kothari et al. 2021), indicate that tree diversity and composition drive
494 overyielding through amelioration of both atmospheric drought and light stress. Specifically, we
495 show that in young experimental forests, fast-growing early successional tree (FAST) species
496 shape the microenvironment of their neighbors, influencing species growth rates, increasing
497 complementarity, and contributing to overyielding. These relationships could shift over time,
498 with slower-growing later successional species dominating and influencing both microclimate
499 and productivity in the future (Zheng et al. 2024).

500

501 In the face of a changing climate, FAST species may play an important role in buffering
502 microclimates in young forests, benefiting the FAST species themselves and facilitating the
503 growth of species that are less tolerant of drought or light stress. While FAST species may differ
504 between sites, they play an important role in early successional forests as ecosystem engineers,
505 influencing local microenvironments and ecosystem function (Gilad et al. 2004; Losapio et al.
506 2021). This study creates a window into the complex relationships between overyielding,
507 microclimate, forest structure, community composition, and tree species diversity: a critical step
508 in understanding and protecting our biodiverse ecosystems.

509

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525

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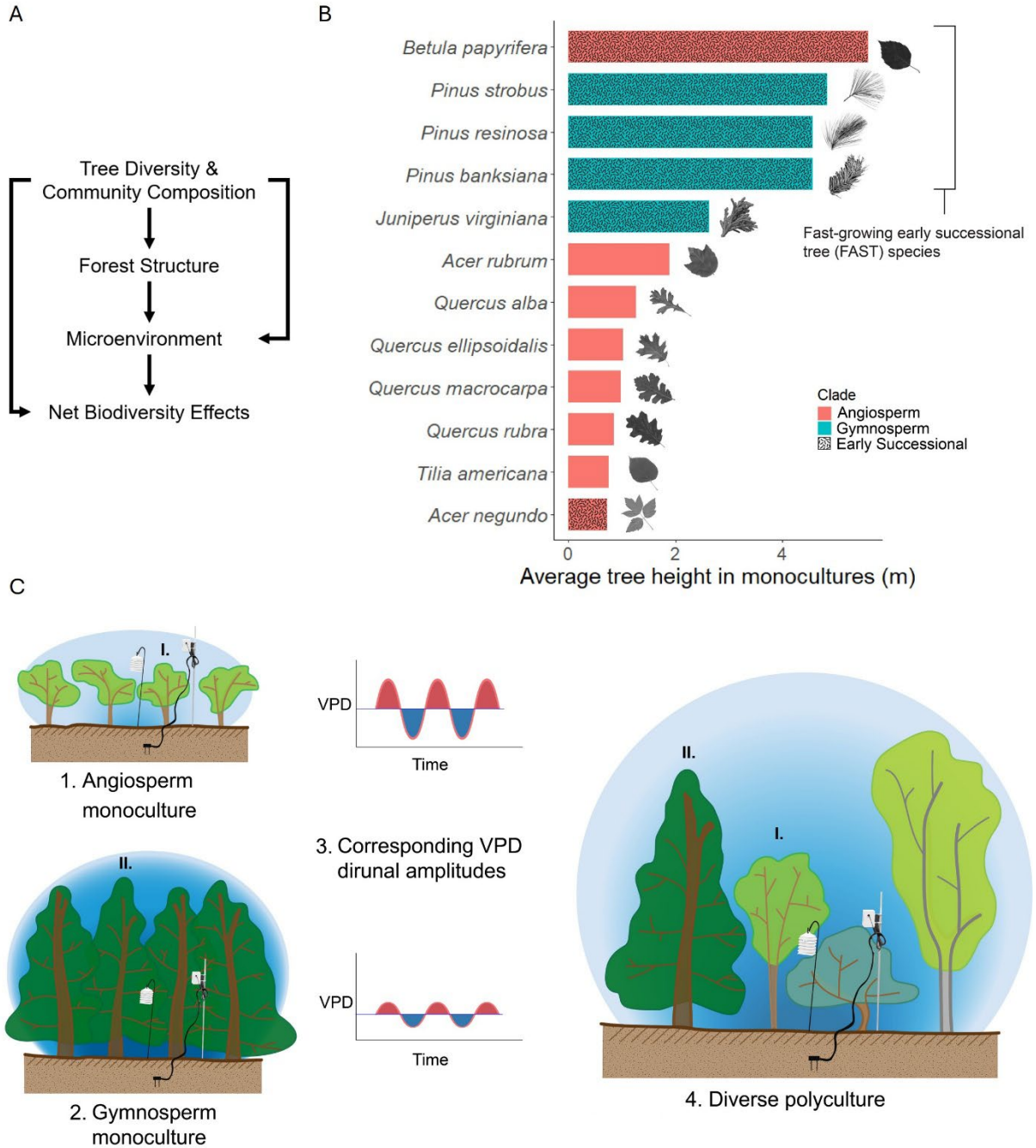
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FIGURES

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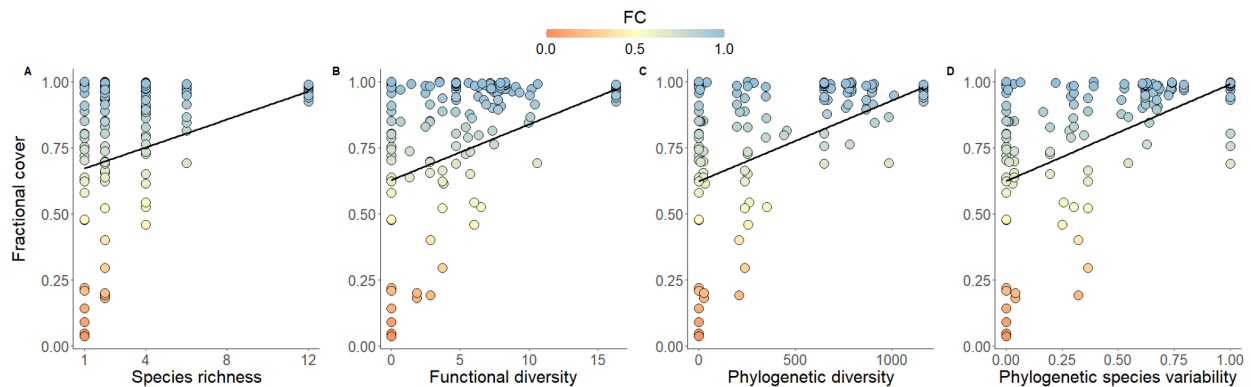
Figure 1. This study used a forest biodiversity experiment to examine the mechanisms linking

787

biodiversity and ecosystem functioning. (A) We hypothesized that forest structure and its effects

788 on microenvironment served as links between tree diversity, community composition, and net
789 biodiversity effects. (B) We predicted that the fast growing early successional (FAST) species
790 would act as ecosystem engineers, decreasing local light availability and buffering against
791 extremes in microclimate. (C) Sensors were installed in forest plots to measure microclimate and
792 soil moisture. We predicted that (1) plots with shorter angiosperm trees would experience more
793 (3) extreme daily fluctuations in vapor pressure deficit (VPD) than (2) taller gymnosperm
794 monocultures with more canopy cover to buffer local VPD. (4) Diverse polycultures would bring
795 together tree species that would create variation in local environments that could influence the
796 physiology of individual trees (I, II) and contribute to net biodiversity effects through increases
797 in complementarity. Leaf scans courtesy of Jake Grossman.

798



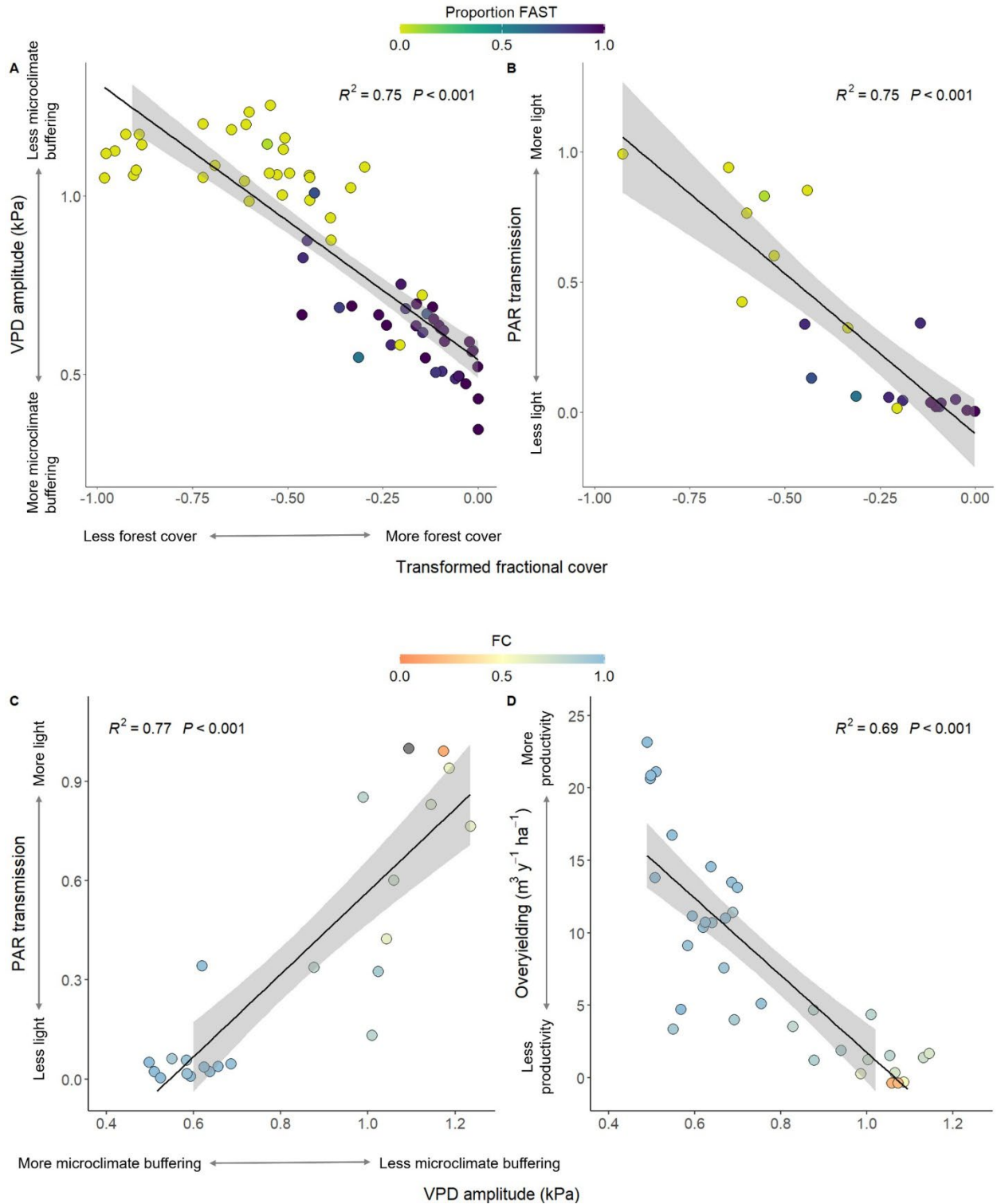
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800 **Figure 2.** Quantile regressions at the 25th percentile show that different metrics of tree diversity
801 are significantly correlated with increases in fractional cover (FC) ($p < 0.001$). In addition,
802 variation in fractional cover decreases with increasing tree diversity ($P < 0.001$).

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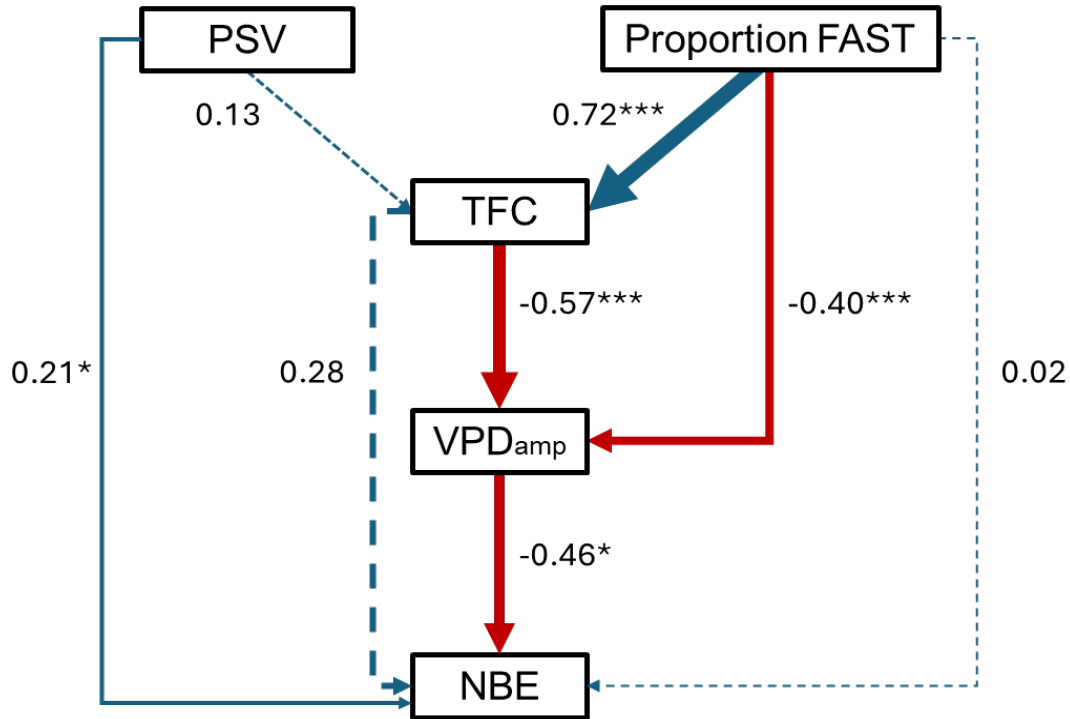
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807 **Figure 3.** Linear regressions show the strong relationships between transformed fractional cover

808 (TFC), VPD amplitude (VPD_{amp}), PAR transmission, and overyielding. A higher FAST

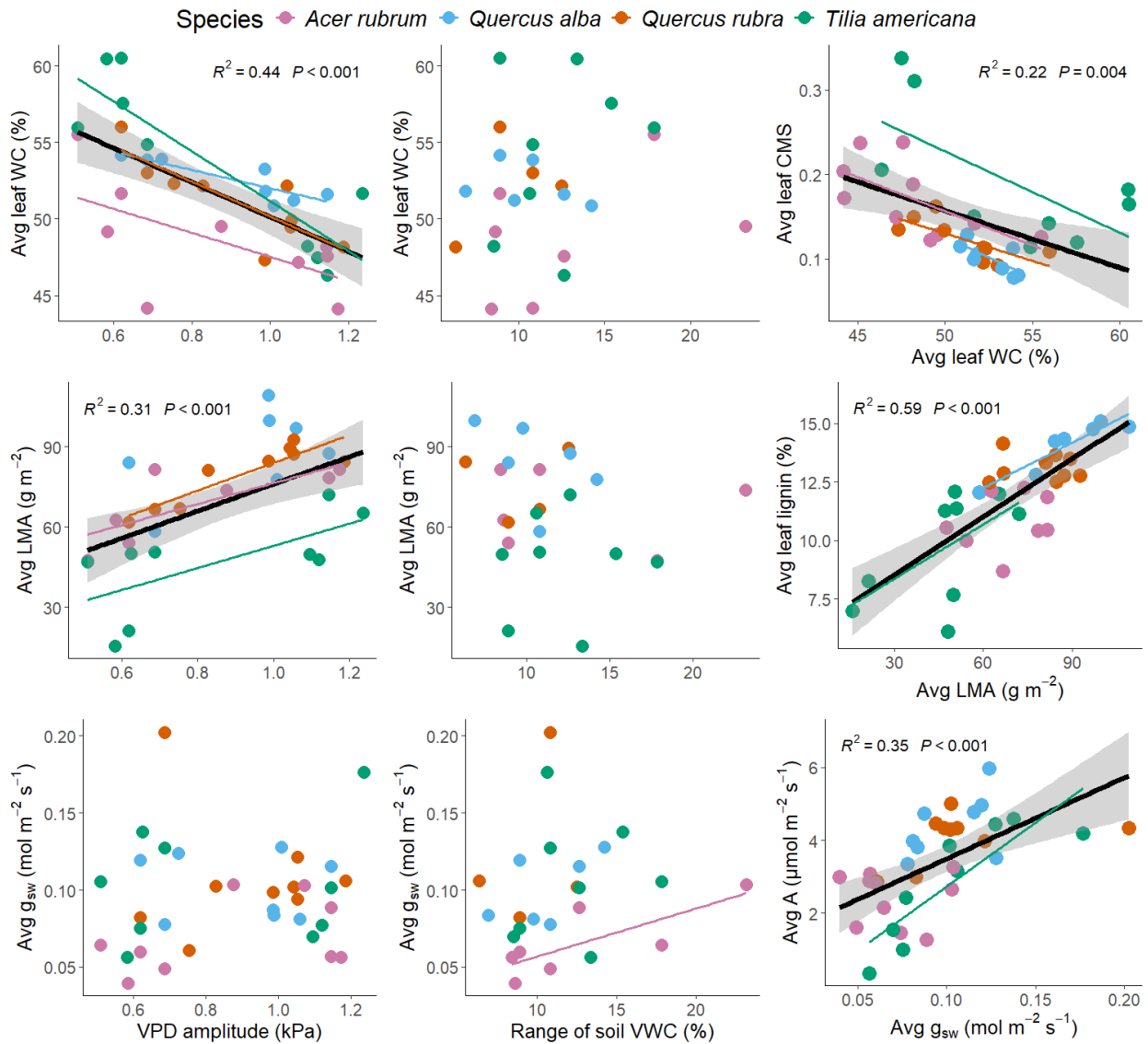
809 proportion in a forest plot contributes to increased TFC. Increased TFC is linked with declines in
810 (A) VPD_{amp} and (B) PAR transmission through the forest canopy. Higher VPD_{amp} is associated
811 with (C) more PAR transmission and (D) decreased overyielding. Connecting these panels, we
812 reason that FAST species contribute to buffered microclimates that are associated with increased
813 wood productivity. FAST species also lower PAR transmission through the canopy, lowering
814 light related stressors, but also potentially posing a challenge of light limitation for some tree
815 species.

816



817

818 **Figure 4.** A structural equation model was built to understand the mechanisms linking
 819 biodiversity and ecosystem function. Blue lines indicate positive relationships, and red lines
 820 indicate negative relationships between variables. Solid lines and asterisks indicate significant
 821 correlations between variables. Line thickness denotes the size of standardized estimates of paths
 822 between variables. Phylogenetic species variability (PSV) was chosen as a diversity metric, and
 823 FAST proportion was selected as a metric of community composition. Transformed fractional
 824 cover (TFC) is a metric of forest structure, and VPD amplitude (VPD_{amp}) is a measure of
 825 microclimate. Overyielding is the net biodiversity effect (NBE) in this model. Greater FAST
 826 proportion significantly increased TFC and decreased VPD_{amp} ($p < 0.001$). TFC also greatly
 827 reduced VPD_{amp} ($p < 0.001$). Higher VPD_{amp} was associated with lower NBE ($p < 0.05$), and
 828 greater PSV positively influenced NBE ($p < 0.05$). This model shows that tree diversity,
 829 community composition, and the interactions between trees and their local environment can
 830 influence tree growth and overyielding.



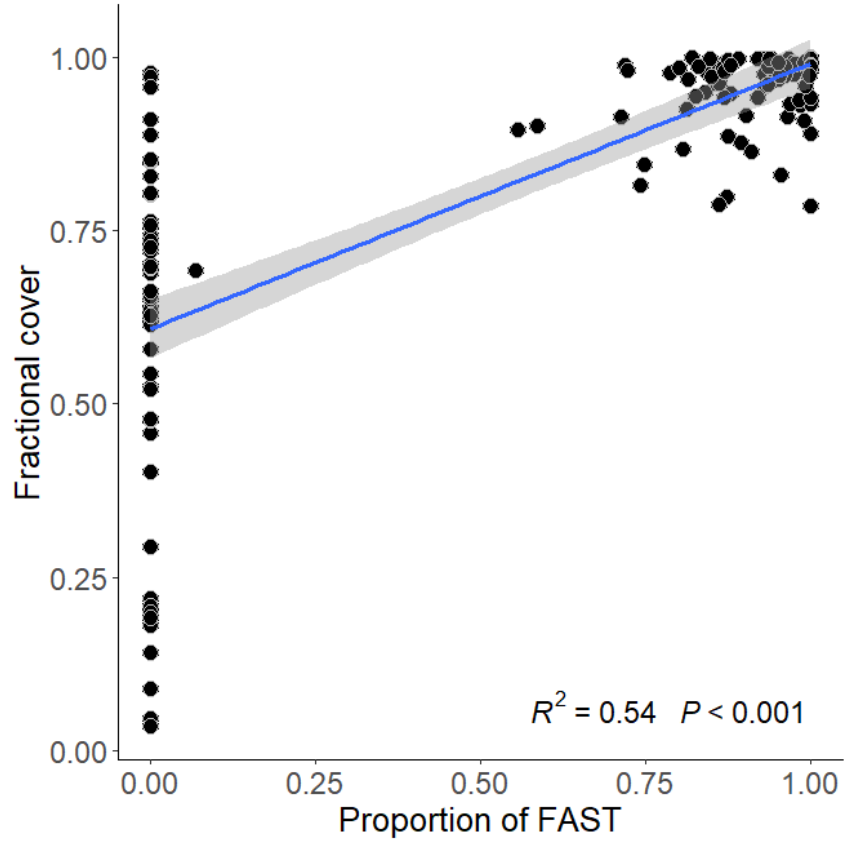
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833 **Figure 5.** Relationships between micro-environmental metrics and leaf physiology of four focal
 834 species. To understand how environmental changes in the forest plots influenced tree physiology,
 835 we selected four species to study in-depth: *Acer rubrum*, *Quercus alba*, *Quercus rubra*, and *Tilia*
 836 *americana*. Black regression lines show significant regressions ($p < 0.01$) between variables for
 837 all four species combined across plots representing different diversity and composition
 838 treatments. Displayed R^2 and p values correspond to these regression lines. Individual colored

839 lines are linear regressions of each species and are included for relationships in which $p < 0.1$.
840 Environmental variables VWC and VPD were measured at the plot level. VWC is expressed as
841 % water volume per unit volume of soil. All of the remaining physiological variables are species
842 averages per plot. Leaf-level water content (WC) is the relative water content in a fresh leaf by
843 mass (%). Carter Miller Stress Index (CMS) is a unitless ratio. Avg LMA refers to average leaf
844 mass per area, Avg A refers to average photosynthetic gas exchange of a leaf ($\mu\text{mol m}^{-2} \text{s}^{-1}$), and
845 Avg g_{sw} refers to average stomatal conductance to water of a leaf ($\text{mol m}^{-2} \text{s}^{-1}$). Leaf lignin
846 content is estimated as % dry mass. Significant regressions across plots ($p < 0.01$) include:
847 VPD_{amp} and leaf water content, leaf water content and CMS, VPD_{amp} and LMA, LMA and leaf
848 lignin, leaf lignin and CMS, LMA and photosynthesis, and stomatal conductance and
849 photosynthesis.

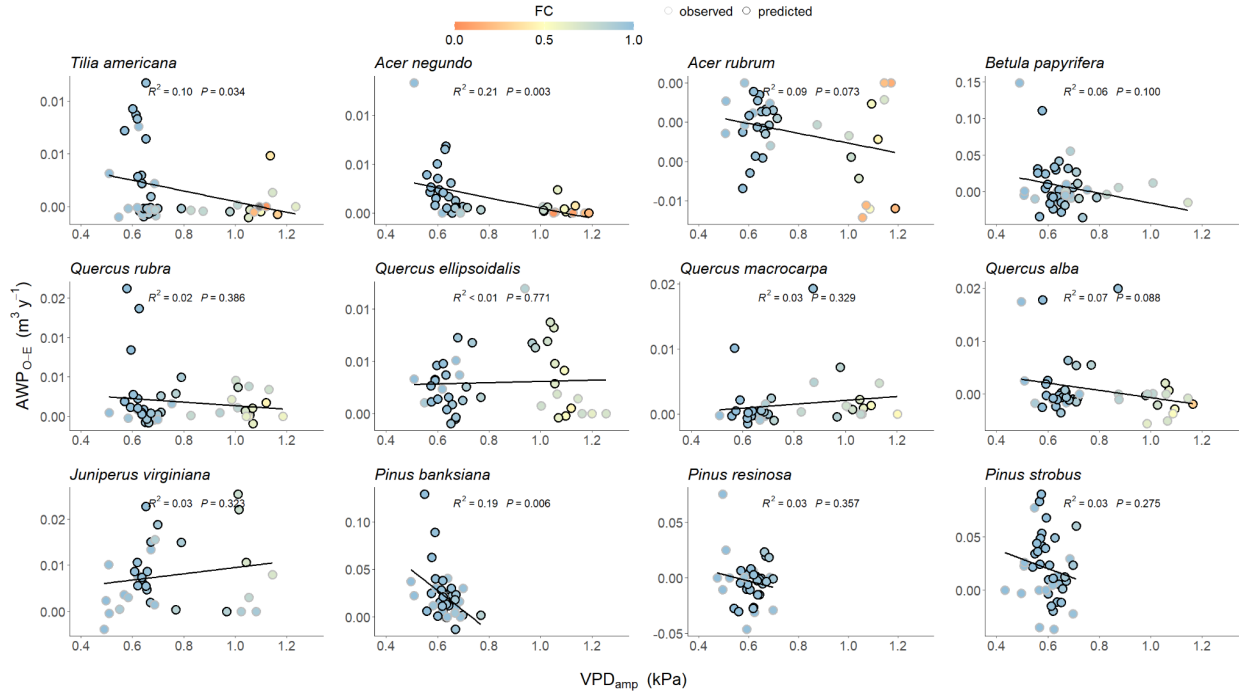
1 SUPPLEMENTARY MATERIALS

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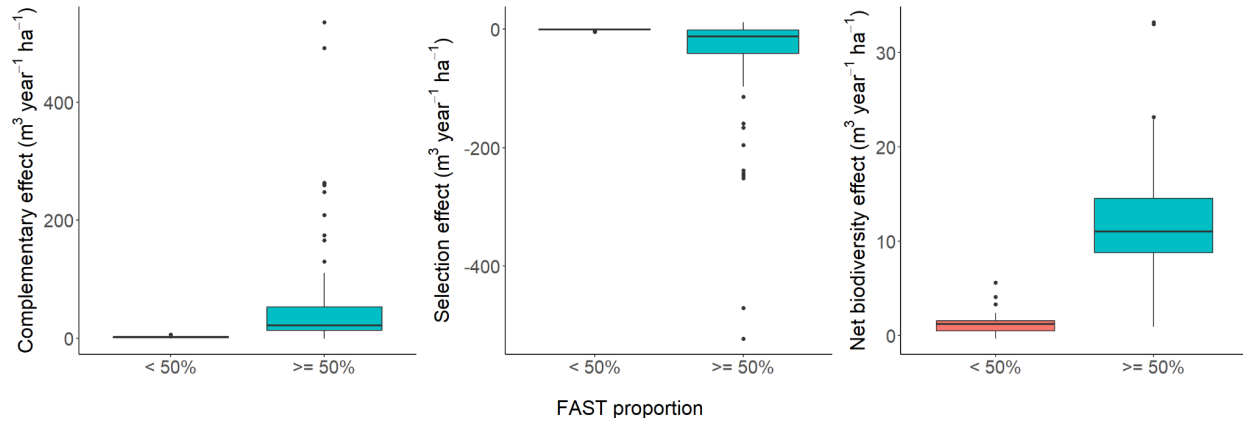
4 **Supplementary Figure 1.** The FAST proportion was significantly correlated with fractional
5 cover in a forest plot ($p < 0.001$). When FAST species had high abundance in a plot, they
6 contributed a substantial proportion of the plot wood volume, leading to the split in data
7 observed in this graph.



8

9 **Supplementary Figure 2.** Increases in VPD_{amp} significantly decreased the annual wood
 10 production (AWP_{O-E}) of tree species that were shade tolerant and/or water limited, including *T.*
 11 *americana*, *A. negundo*, and *P. banksiana*. Each point represents a forest plot. Plots with more
 12 fractional cover (FC) were generally associated with reduced VPD_{amp} . A multiple linear
 13 regression model was developed to predict VPD_{amp} for plots in which this metric was not
 14 measured (Equation 2). Plots in which VPD_{amp} was observed are encircled in grey, and plots for
 15 which VPD_{amp} was predicted are encircled in black.

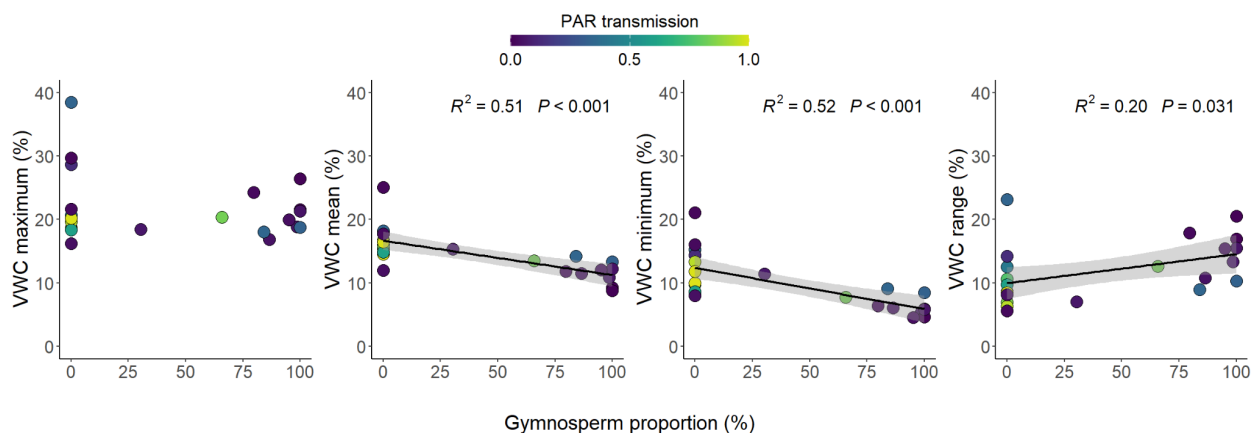
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18 **Supplementary Figure 3.** Overall trends in overyielding (net biodiversity effect) were driven by
 19 complementarity effects. Plots in which FAST species composed 50% or more of the wood
 20 volume displayed greater complementary ($p < 0.01$), less selection ($p < 0.05$) and more
 21 overyielding ($p < 0.001$) than plots with less than 50% FAST proportion.

22



23

24 **Supplementary Figure 4.** The proportion of gymnosperms significantly influenced the mean,
 25 minimum, and range of volumetric water content (VWC), but not the maximum VWC, in forest
 26 plots. PAR transmission was moderately correlated with the range of VWC ($p < 0.1$).

27 Acknowledging that VPD_{amp} and VWC_{range} were not measured on the same timescales, we found
 28 that the two variables were largely decoupled in this system. Given that the proportion of
 29 gymnosperms did not significantly influence the maximum VWC but significantly lowered the

30 mean and minimum VWC in plots, the relationship between gymnosperms and VWC_{range} may
31 have been driven by increased water uptake by the well-established gymnosperm roots in the
32 shallow rooting zone (Bond 1989). The aboveground wood volume of gymnosperm species in
33 the experiment was substantial, suggesting that belowground root volume was similarly sizable
34 (Annighöfer et al. 2022).

35

36 **Supplementary Table 1.** The table describes the experimental design of this study.

37 Microclimate and soil moisture monitoring sensors were distributed in plots that spanned large
38 gradients of the different diversity metrics considered in FAB2. Plots in which sensors were
39 installed are noted with “1”, and the total number of plots are summed at the bottom of the table.
40 One control plot, an unplanted, unmown open area, was included in the experimental design for
41 quality control purposes but was not included in the analyses presented in this paper. The
42 gymnosperm-angiosperm ratio was used as a rough estimate of gymnosperm proportion based on
43 the original planting ratios (i.e. 1 gymnosperm species out of 4 total species in a plot = 0.25
44 gymnosperm-angiosperm ratio).

45

46 **Supplementary Table 2.** The table explains how volumetric water content (VWC) calibration
47 curves were fit to two soil moisture sensors used in the experiment. In 2023, soil from an
48 unplanted section of the FAB2 experiment was harvested and placed in a drying oven. After the
49 soil was completely dry, the soil was placed in a large bucket. Water was incrementally added to
50 the soil in known quantities and thoroughly mixed. Two different sensors (GS1, Teros 10) were
51 placed in the soil at regular intervals to obtain voltage readings. VWC was calculated as the

52 volume of water divided by the total volume of soil. The volume of soil was calculated as the
53 area of the bucket base ($\pi d^2/4$) multiplied by the height of the soil. Linear regression lines were
54 fitted in R v. 4.3.1 to obtain coefficients that could be used to convert voltage readings into VWC
55 for each sensor (R Core Team 2023).

56

57 **Appendix 1.** We used predicted models on full-range reflectance spectra (350–2500 nm)
58 measurements to derive two leaf traits: lignin and water content. These models were constructed
59 on pair measurements of traits and reflectance spectra (lignin: $n = 117$, and water content: $n =$
60 357) of samples collected on woody species across central Minnesota using Partial Least Square
61 Regressions within a machine learning framework. This framework involves methods for model
62 calibration and validation as described in Burnett et al. 2021. Data were split using a stratified
63 sampling according to the distribution of a trait in a histogram to randomly select 60% samples
64 for model calibration and the remaining 40% for independent model validation. On the
65 calibration samples, 1000 repeated PLSR were applied using 80% of the samples each time.
66 These repeated models were used to identify the optimal numbers of components required to
67 avoid underfitting and overfitting for predictions (i.e., lignin = 6; water content = 8). Using the
68 optimal number of components, PLSR coefficients were extracted and used to compute leaf
69 traits. These models were assessed using R^2 and relative root mean squared error (rRMSE)
70 between measurements and predictions on validation datasets (lignin: $R^2 = 0.65$, rRMSE = 0.26;
71 water content: $R^2 = 0.6$, rRMSE = 0.15). All statistical analyses for these models were performed
72 in R v. 3.6.3 (R Core Team 2020) using the package pls v. 2.7.1 (Mevik and Wehrens 2007) for
73 PLSR modelling.

74