Multimodal data integration to model, predict, and understand
 changes in plant biodiversity: a systematic review

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10 Running head: Multimodal data integration in biodiversity

- 11 Abstract
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13 The integration of multimodal data to analyze, model, and predict changes in plant biodiversity is 14 critical for addressing global conservation challenges. This systematic review examines the current 15 landscape of plant biodiversity data, focusing on the identification, classification, and evaluation of 16 key open-access data sources and integration methodologies. We highlight the strengths and 17 limitations of major biodiversity platforms, emphasizing their contributions to species occurrence, 18 trait data, taxonomic checklists, and environmental variables. The review also explores 19 computational approaches for multimodal data integration, including metadata standardization 20 frameworks (e.g., Darwin Core) and advanced modeling techniques like Species Distribution Models 21 (SDMs) and deep learning methods. Additionally, we assess environmental drivers of biodiversity 22 change—such as climate variables, land-use patterns, and elevation gradients—and the tools 23 available for their analysis. Despite significant advancements in biodiversity informatics, challenges persist in achieving interoperability across datasets, addressing spatial and temporal biases, and 24 25 integrating remote sensing with in situ observations. By identifying gaps in current methodologies 26 and proposing pathways for improvement, this review contributes to advancing biodiversity 27 monitoring strategies. These efforts align with global conservation goals outlined by the Convention 28 on Biological Diversity and the United Nations Sustainable Development Goal 15. Ultimately, the 29 findings underscore the importance of harmonized data integration frameworks to enhance 30 predictive modeling capabilities and inform effective conservation policies.

31 Keywords: Biodiversity, Multimodal data, Databases, Data integration, Big data, Smart data,
32 Mathematical ecology

34 **1 Introduction**

35 Biodiversity includes the full spectrum of life on Earth, encompassing genetic, species, and 36 ecosystem diversity (Convention on Biological Diversity, 2011). Its conservation is crucial for 37 sustaining ecosystems, ensuring food security, and protecting human health. Forests are vital to 38 global biodiversity, serving as habitats that support diverse life forms (Pontius et al., 2020). 39 However, recent studies highlight an imbalance in research focus, with greater attention given to 40 animal diversity while plant biodiversity hotspots in developing countries remain understudied 41 (Wani et al., 2024). This knowledge gap, along with the threats facing plant biodiversity, hampers 42 the development of effective conservation strategies, particularly in the face of climate change and 43 human activity (FAO, 2022).

44 High-quality, comprehensive global plant biodiversity data are essential for developing adequate 45 biodiversity conservation strategies. Although plant biodiversity data have been collected for over 46 a century (Tejada-Gutierrez et al., 2024), satellite technologies have only recently enabled large-47 scale monitoring (Chen et al., 2021; Cohen et al., 2020; Khare et al., 2021; Negassa et al., 2020; Xu 48 et al., 2022). Many of these datasets, now openly accessible through online platforms (Saran et al., 49 2022), provide diverse information, including plant phenotypes, geodiversity, climate variables, and 50 molecular sequences (Enquist et al., 2016; Fernández et al., 2020; Keppel et al., 2021). These data 51 are generated using multimodal acquisition methods, ranging from in situ measurements to remote 52 sensing. Yet, their heterogeneity and diverse formats complicate access, integration, and 53 interpretation (Williamson & Leonelli, 2023).

54 Effective multimodal data integration enhances biodiversity change detection, refines existing 55 databases, and informs conservation strategies to mitigate biodiversity loss (Cai *et al.*, 2023;

Wäldchen & Mäder, 2017). Computational approaches are poised to play an increasingly important
role in that integration, improving data usability for biodiversity monitoring and conservation
(Ganjdanesh *et al.*, 2022; Saran *et al.*, 2022; Schellenberger Costa *et al.*, 2023).

59 This paper presents a comprehensive review of the current state of plant biodiversity data, 60 emphasizing the identification, classification, and critical evaluation of major data sources and 61 integration methodologies employed to study spatial and temporal patterns of plant biodiversity. 62 We pay special attention to open-access tools and resources that support biodiversity research, with 63 an analysis of their strengths and limitations to identify areas requiring further development. By 64 advancing strategies for biodiversity monitoring, this work contributes to the goals of the 65 Convention on Biological Diversity (Pörtner et al., 2021; Robinson, 2020; United Nations, 1992) and 66 supports the achievement of the United Nations Sustainable Development Goal 15 (United Nations, 67 2018).

68 **2 Selection of web platforms**

69 As a first step to identify plant biodiversity platforms, we searched Web of Science, Scopus, and Google Scholar using combination of the keywords biodiversity, databases, data, platforms, and 70 71 informatics. We applied specific filters in each portal and refined the results using additional criteria 72 (see Supplementary Table S1 for search strategy details). From this process, we selected 21 review 73 articles, including also a recent publication describing the ForestForward platform (Tejada-74 Gutiérrez et al., 2024), developed in our research group and directly aligned with the objectives of 75 this study. To identify and organize the biodiversity databases (DBs) mentioned in those articles, we 76 used Google's NotebookLM (2024). We identified 107 web platforms containing biodiversity data (Supplementary Table S2). Figure 1 summarizes the protocol and criteria we used to select theplatforms to include in our analysis, which are described below.

79 First, we applied a selection criterion based on taxonomic coverage. We selected only platforms 80 with information across all biological kingdoms or fully covering the *Plantae* kingdom. Second, we 81 applied a geographic resolution criterion, selecting only platforms containing DBs with global 82 coverage, and excluding DBs limited to a specific region or country. Third, we included only 83 platforms cited in more than one article. Fourth, we prioritized DBs focused on species occurrence, 84 taxonomy, and traits, unless they were recent (less than one year), contained valuable data on 85 species occurrence, taxonomy, traits, or employed innovative approaches, such as time series 86 analysis. Fifth, we revised accessibility of the platforms that passed the first four filters (n= 13, 87 marked in Supplementary Table S1), to ensure that the website was active at the time of access 88 (May 2025). After applying all the selection criteria, we include 12 DBs in our sample (see 89 Supplementary Table S2).

3 Plant Biodiversity platforms

In this section, we review the selected biodiversity data portals, and the information they contain.
We also briefly describe how each portal integrates data and infrastructure provided by the other
portals. Table 1 provides a summary of the data available in each portal, while Figure 2 summarizes
the interactions and dependencies between portals.

The Global Biodiversity Information Facility (GBIF) is the leading portal for biodiversity and climate change research (Heberling *et al.*, 2021; Saran *et al.*, 2022) and provides approximately 538 million plant occurrence species records, organized in thousands of individual datasets (accessed on: May 2025). Temporal coverage is from 1600s to the present. GBIF integrates data from multiple sources 99 such as Tropicos and World Flora Online (WFO), and supports portals such as Map of Life (MOL), 100 Catalogue of Life (CoL), and Encyclopedia of Life (EOL) through open standards such as the Darwin 101 Core Standard (DwC) (Wieczorek *et al.*, 2012). The individual datasets in GBIF lack uniformity and 102 more than half of the records are incomplete, for example lacking information about either 103 genus/species, date of registration, or location (Tejada-Gutierrez *et al.*, 2024).

The Yale Center for Biodiversity and Global Change hosts the MOL portal, which provides 44 plant datasets (accessed on: May 2025). The interface of the portal organizes data by species richness, rarity, and includes alien species checklist for some plant species, with temporal coverage from the 1500s to present. MOL collaborates with the International Union for Conservation of Nature Red List of threatened Species, Global Register of introduced and invasive species (GRIIS), and GBIF (Jetz *et al.*, 2012).

110 The CoL portal provides a global taxonomic checklist (TC) with accepted names and synonyms, hosting approximately 2 million living and 147,000 extinct species (accessed on: May 2025). 111 112 Temporal coverage varies by datasets, but species names are kept up to date. CoL is an essential 113 source that was designed specifically to link global taxonomic information, without considering how 114 to integrate geographic technologies. Its data is not organized to enable spatial analyses of 115 biodiversity (Saran et al., 2022). CoL integrates taxonomy data from the Plants of the World Flora 116 Online (POWO) and WFO (Bánki et al., 2018). It also provides access to additional information such 117 as GBIF distribution data, literature (Biodiversity Heritage Library- BHL), and trait data (Hobern et 118 al., 2021).

BioTIME provides time series of species abundances organized by biomes. BioTIME provides over
12 million records on a global scale (accessed on: May 2025), with a temporal coverage from 1800s

to 2016. BioTIME principally integrates data from GBIF, Ocean Biodiversity Information System
(OBIS), and Ecological Data Wiki (Dornelas *et al.*, 2018).

WFO provides approximately 1.6 million plant names, covering a temporal scale from 1700s to present (accessed on: May 2025). WFO's stated goal is to support Target 1 of the Convention on Biological Diversity (2012). WFO collaborates closely with GBIF, CoL, POWO, and Taxonomic Expert Networks (Borsch *et al.*, 2020).

Plants of the World Online (POWO) is hosted by the Royal Botanic Gardens Kew and provides approximately 1.4 million plant names (accessed on: May 2025), with a temporal coverage encompassing the past 250 years (Saran *et al.*, 2022; Zuanny *et al.*, 2024). POWO updates its data weekly, integrating data from World Checklist of Vascular Plants (WCVP), GBIF, CoL, and WFO.

The Missouri Botanical Garden hosts Tropicos, which provides approximately 1.33 million scientific
 names (accessed on: May 2025), with a temporal coverage from 1700s to present. Specimen data is
 accessible through the GBIF database (Teisher & Stimmel, 2024).

The ForestForward portal provides approximately 172 million records, with a temporal coverage from 1600s to the present (access on: May 2025). ForestForward integrates data from GBIF, using a semi-automated Extract, Transform, and Load (ETL) process to curate data. This ensures data quality for spatial-temporal analysis of plant biodiversity (Tejada-Gutierrez *et al.*, 2024).

The Smithsonian Institution of National Museum of Natural History hosts EOL, which provides trait data on animals, plants, fungi, protists, and bacteria of approximately 2 million species, covering a temporal scale from 1600s to present (Baker *et al.*, 2014; Parr *et al.*, 2014) (accessed on: May 2025).
Plant traits include ecological adaptations, propagations methods, multimedia information,

morphological and physiological characteristics. EOL collaborates with portals such as BHL, Barcode
of Life, CoL, and GBIF. The portal depends on GBIF for occurrence data (Saran *et al.*, 2022).

The Max Planck Institute for Biogeochemistry in Jena hosts the Plant Trait Database (TRY), which provides approximately 15 million trait records (accessed on: May 2025), including morphological, physiological and phenological traits (Kattge *et al.*, 2011, 2020). TRY integrates data from sources such as Life History Traits of the Northwest European Flora (LEDA), Global Plant Trait Network (GlopNet), Trait Database of the German Flora (BiolFlor), and collaborates with The Global Inventory of Floras and Traits (GIFT) and the Botanical Information and Ecology Network (BIEN) (Kattge *et al.*, 2020).

The GIFT is a global repository of plants with approximately 360,000 species across 3,627 geographic
regions (accessed on: May 2025). Trait data include information about morphology, phenology, and
physiology. GIFT collaborates with Global Naturalized Alien Flora (GloNAF), TRY, and MOL (Weigelt *et al.*, 2019).

The BIEN portal provides approximately 81 million records, 900,000 of which contain trait data about morphological characteristics (Maitner *et al.*, 2017). The last DBs version, BIEN 4, expanded ecological plot observations by 23% (accessed on: May 2025). BIEN 4 integrates datasets from sources such as GBIF, The Royal Botanical Garden of Sydney, and NeoTropTree (Enquist *et al.*, 2016).

4 Biodiversity data integration: approaches and challenges

160 The vast and heterogeneous nature of biodiversity datasets complicates their integration across 161 portals and computational systems. Estimates suggest that data preparation consumes 162 approximately 80% of research time due to these complexities (Huber *et al.*, 2021). This underscores the need for efficient computational methods for multimodal data integration (Ganjdanesh *et al.*,
2022). This section examines the primary strategies used by biodiversity portals to integrate diverse
data types from heterogeneous sources.

166 4.1 Metadata standardization

Biodiversity databases employ a limited but interconnected set of approaches to standardize and integrate data, with metadata standardization playing a critical role. Most portals adopt the Darwin Core (DwC) standards, developed by the Biodiversity Information Standards (TDWG), to ensure consistency in key data fields such as species names, locations, and observation dates. This standardization facilitates seamless data exchange across portals (Borsch *et al.*, 2020; Enquist *et al.*, 2016; Govaerts *et al.*, 2021; Hobern *et al.*, 2021; Parr *et al.*, 2016; Weigelt *et al.*, 2019; Wieczorek *et <i>al.*, 2012).

Some portals also support alternative metadata standards, including CoL Data Package (COLDP), Annual Checklist Exchange Format (ACEF), Ecological Metadata Language (EML), VegX, and Humboldt Core (Döring & Ower, 2019; Hobern *et al.*, 2021; Guralnick *et al.*, 2018; Jones *et al.* 2019; Jetz *et al.*, 2012; Weigelt *et al.*, 2019; Wieczorek *et al.*, 2012) (Supplementary Table S3). Additionally, stable unique identifiers—such as Digital Object Identifiers (DOIs) and WFO IDs—link species occurrences to images, morphological data, and passport information (Döring & Ower, 2019; Hobern *et al.*, 2021; Palese *et al.*, 2019; Steidinger *et al.*, 2019).

Despite these efforts, integration workflows vary across portals, leading to inconsistencies, data duplication, and omissions. These challenges hinder the incorporation of datasets, particularly those containing images, species descriptors, or specific trait data (Reichman *et al.*, 2011; Saran *et al.*,

184 2022). Although workflow standardization and data interchange are improving, inconsistencies in185 data quality and structure remain significant obstacles.

186 **4.2 Primary data integration**

Scientific names serve as the primary key for integrating biodiversity data. However, taxonomic nomenclature changes over time, leading to issues such as homonyms, synonyms, and ambiguous spellings (Boyle *et al.*, 2013; Feng *et al.*, 2022; Franz *et al.*, 2008). To address these challenges, biodiversity portals employ different taxonomic resolution approaches (Table 2).

Taxonomic Backbone Systems (TBS) integrate name-based information from multiple sources, providing a unified framework for taxonomic searches, cross-referencing, and data consistency.
Some portals, such as GBIF, WFO, and POWO, use TBS. GBIF's Backbone Taxonomy incorporates all names in GBIF datasets, primarily referencing the CoL. WFO combines a taxonomic backbone, supported by Botalista, with active curation by Taxonomic Expert Networks (TENs) (Borsch *et al.*, 2020; Palese *et al.*, 2019). POWO relies on the World Checklist of Vascular Plants (WCVP) as its taxonomic backbone, which is also integrated into CoL and GBIF (Govaerts *et al.*, 2021).

198 **Checklist Approaches (CA)** focus on regional species or specific taxonomic groups, minimizing 199 interconnectivity between DBs (Feng *et al.*, 2022). Some portals use **Taxonomic Name Resolution** 200 **Services (TNRS)** to resolve nomenclature conflicts. Portals such as BIEN, TRY, and GIFT implement 201 TNRS, integrating global and regional checklists, including WCVP, WFO, Tropicos, USDA's Plants 202 Database, the Global Compositae Checklist, and the International Legume Database and 203 Information Service (Boyle *et al.*, 2013; Denelle *et al.*, 2023; Kattge *et al.*, 2020; Maitner *et al.*, 2018; 204 Weigelt *et al.*, 2019).

Some portals collaborate with taxonomic experts and integrate multiple checklists for comprehensive data management. Platforms such as CoL, Tropicos, and WFO serve as primary checklists within major databases like GBIF, EOL, MOL, and TNRS (Boyle *et al.*, 2013; Hobern *et al.*, 2021; Jetz *et al.*, 2012; Parr *et al.*, 2016). Additionally, CoL and GBIF developed **ChecklistBank** (https://www.checklistbank.org/), an open portal that integrates taxonomic checklists (Hobern *et al.*, 2021).

Some biodiversity portals implement custom integration workflows tailored to specific research needs. MOL overlays geographical information to link species distributions with environmental data (Jetz *et al.*, 2012). BioTIME employs a structured table system with five levels (records, sample, plot, site, and study) to connect species abundance data with spatial and temporal information. Typographic errors and misspellings in species names are corrected during curation process (CP) (Dornelas *et al.*, 2018). ForestForward utilizes an automated ETL process with Python scripts to integrate thousands of GBIF datasets (Tejada-Gutierrez *et al.*, 2024).

In general, and despite their methodological differences, most portals use semi-automated or fully automated processes to remove duplicate and incomplete records, standardize taxonomic data, and curate missing or ambiguous information (Döring & Ower, 2019; Hobern *et al.*, 2021; Tejada-Gutierrez *et al.*, 2024). Many also rely on data and standards derived from GBIF, reinforcing its central role in biodiversity data integration.

223 4.3 Trait Data Standardization

Another key challenge in biodiversity data integration is the standardization of trait data. Variability in measurement units, methodologies, and trait terminology complicates interoperability and often exceeds the capabilities of standards such as DwC (Deans *et al.*, 2015; Feng *et al.*, 2022; Wieczorek *et al.*, 2012). To address this, biodiversity portals typically adopt either Ontology-Based Integration
or Standardized Trait Terminology (Table 3).

229 For example, EOL's TraitBank employs DwC metadata standard (MS) and its packaging format DwC-230 A to organize and share data. Additionally, TraitBank integrates ontologies to describe traits in a 231 structured and interoperable way. These include the Plant Trait Ontology (TO), Phenotypic Quality 232 Ontology (PATO), and the Environment Ontology (ENVO), and other vocabularies from Open 233 Biological and Biomedical Ontologies (OBO) (Baker et al., 2014; Parr et al., 2016). TRY enhances 234 interoperability using the Extensible Observation Ontology (OBOE) schema and trait vocabulary (TV) 235 and protocols (P) as the Thesaurus of Plant Characteristics (TOP) and the Plant Trait Handbook 236 (Garnier et al., 2017; Pérez-Harguindeguy et al., 2013). Similarly, GFIT and BIEN integrate with TRY 237 to ensure consistency in trait data representation (Kattge et al., 2020) (Supplementary Table S3).

238 Effective multimodal data integration requires a multidisciplinary approach, as biodiversity

239 datasets encompass complex variables, including species occurrences, traits, environmental

240 parameters, and remote sensing information. Addressing these standardization challenges is

241 crucial for advancing biodiversity analysis, which remains constrained by data inconsistencies and

242 limited interoperability (Enquist *et al.*, 2019; Gotelli, 2008).

243 **5 Multimodal Data in Biodiversity Research**

The plant-related information contained in biodiversity databases is essential for documenting and analyzing how that biodiversity changes. However, identifying the factors that correlate with or drive these changes requires integrating additional multimodal datasets, including genetic, morphological, climatic, soil, land-use, and ecological parameters (O'Connor *et al.*, 2020; Soltis & Soltis, 2016). This section reviews key environmental and climate-related data sources used to
assess and predict biodiversity changes.

250 **5.1 Environmental drivers of biodiversity change**

251 Environmental variables, such as temperature, precipitation, and solar radiation, play a crucial role in shaping biodiversity by influencing species fitness, adaptation, and evolution (Guan et al., 2023; 252 253 McGaughran et al., 2021; Segovia, 2023; Zhang et al., 2018; Zheng et al., 2024). Temperature 254 fluctuations, particularly extreme highs and lows, impact key physiological processes, including 255 photosynthesis, respiration, germination, and growth (Minoli et al., 2019; Najeeb et al., 2019; 256 Pimentel, 2022). Freezing temperatures can induce stress, affecting cellular survival, gene 257 expression, and overall plant development (Guan et al., 2023). Meanwhile, precipitation patterns influence plant productivity, biome distribution, and species diversity, especially in arid 258 259 environments, highlighting the importance of spatial resolution in biodiversity studies (Korell et al., 260 2021).

261 Land-use and land-cover changes are also major drivers of biodiversity loss and carbon stock 262 alterations (O'Connor et al., 2020). Deforestation accelerates greenhouse gas emissions and 263 disrupts ecosystems, often leading to declines in native species and increased prevalence of invasive 264 species, particularly when forests are converted to agricultural land (Calvin et al., 2023; Montràs-265 Janer et al., 2024). Conversely, climate change itself influences land-use patterns as communities 266 adapt to shifting environmental conditions (Calvin et al., 2023; IPCC, 2018; Pörtner et al., 2021). 267 Given their predictive power in climate change research, greenhouse gas emissions are among the 268 most frequently analyzed environmental parameters (Calvin et al., 2023; Conradi et al., 2024; Isbell 269 et al., 2023).

270 Altitude significantly affects climate variables such as temperature, precipitation, and wind patterns, 271 shaping vegetation distribution and biodiversity across different elevation zones (Umair et al., 272 2023). Variations in elevation, slope, and land relief create distinct biomes, each supporting 273 different plant communities (Hakkenberg et al., 2023). Research has identified three primary 274 elevation-related biodiversity patterns: (1) a low-elevation diversity plateau, (2) an intermediate 275 peak in mid-elevation zones, and (3) a steady decline in diversity at high elevations (Umair et al., 276 2023). Microclimate and biome-level assessments are critical for understanding spatial changes in 277 plant biodiversity.

278 **5.2 Data Sources and Tools for Environmental Analysis**

Environmental data are available from multiple sources with varying temporal and spatial resolutions. Climate Data Records (CDRs)—derived from satellite observations, climatological stations, and radar systems—offer long-term, homogeneous datasets for detecting climate trends (National Research Council, 2004). Remote sensing technologies provide crucial data on land-use, land-cover, and topographic changes over time (O'Connor *et al.*, 2020). Table 4 summarizes the primary sources for climate and environmental data used in biodiversity and climate change research (Al-Zu'bi *et al.*, 2022; Bedair *et al.*, 2023; Lundstad *et al.*, 2023).

Predictive tools such as the Köppen-Geiger climate classification and Shared Socioeconomic Pathways (SSPs)—developed within the Coupled Model Intercomparison Project (CMIP)—remain underutilized in biodiversity studies. These frameworks offer significant potential for integrating geographical, ecological, and climate change information to forecast biodiversity changes (Beck *et al.*, 2018, 2023; Calvin *et al.*, 2023). Leveraging these predictive approaches could enhance the accuracy of biodiversity projections and improve conservation strategies.

292 6 Plant biodiversity data analysis: approaches and challenges

293 6.1 Measuring Biodiversity

Large-scale biodiversity assessments often rely on diversity indices that condense complex ecological information into a single value (Table 5). Biodiversity is typically characterized by two key components: **species richness**, which represents the total number of species in a community, and **evenness**, which describes the relative distribution of individuals among species (Enquist *et al.*, 2019; Gotelli, 2008; Robinson, 2020).

299 Species richness is a fundamental but challenging metric, as identifying all species within a 300 community is often impractical, and its value is highly dependent on sample size (Enquist et al., 301 2019; Hurlbert, 1971). To account for this, indices such as Margalef (1958) and Menhinick (1954) 302 estimate richness by relating the number of species to the total number of individuals observed. 303 Other indices, including Shannon-Wiener (Shannon & Weaver, 1949) and Simpson (1949), integrate 304 both richness and evenness to provide a more comprehensive assessment of community structure. 305 Changes in these indices can be used to evaluate how environmental factors, such as land-use 306 modifications, impact biodiversity (IPCC, 2018).

Beyond species diversity, understanding species composition is crucial for assessing ecosystem health. For instance, invasive species frequently establish themselves in areas altered by climate change, outcompeting native species and reducing local biodiversity (Calvin *et al.*, 2023; Segan *et al.*, 2016; Wang *et al.*, 2019). Monitoring the abundance of invasive species is essential for developing conservation policies aimed at protecting native biodiversity and managing vulnerable ecosystems (Bellard *et al.*, 2014; Costello, 2023; de Lima *et al.*, 2020; Guo *et al.*, 2023). However, research on invasive species management has declined over the past two decades (Heberling *et al.*, Additionally, Enquist *et al.* (2019) reported that approximately 36.5% of the 435,000 known
 plant species are classified as rare, underscoring the need to incorporate species composition into
 biodiversity assessments. Addressing these challenges requires access to high-quality, reliable, and
 traceable biodiversity data.

318 Portals such as ForestForward, which integrates and standardizes over 4,000 plant datasets from 319 GBIF, are efforts in the right direction to improve biodiversity data quality and accessibility. By 320 curating original datasets and removing low-quality entries with incomplete information, ForestForward enhances data reliability for biodiversity research (Tejada-Gutierrez et al., 2024). 321 322 However, like many similar platforms, it lacks comprehensive environmental variable datasets 323 necessary for understanding the factors influencing biodiversity patterns (Heberling et al., 2021). 324 Expanding such portals to include environmental data would significantly improve their utility for 325 biodiversity analysis and conservation planning.

326 6.2 Understanding Changes in Biodiversity

327 Understanding the effects of climate change on plant biodiversity is essential for modeling future 328 scenarios and assessing the impact of human activities on ecosystems (Bellard et al., 2014; Mantyka-329 Pringle et al., 2015). Advancing biodiversity research requires a comprehensive analysis of the 330 spatial-temporal dynamics embedded in ecological data (Saran et al., 2022). However, few 331 biodiversity portals provide analytical tools for data exploration and modeling due to 332 heterogeneous data sources, biases, and methodological constraints, often necessitating custom 333 analytical pipelines. This section reviews key methodologies for analyzing spatial-temporal changes 334 in plant biodiversity (Figure 3). Some portals have developed specialized tools for biodiversity trend 335 analysis. POWO's 'expowo' R package enables biodiversity assessments across taxonomic levels above genus or within specific geographic regions but is restricted to POWO datasets (Zuanny et al.,

337 2024). Similarly, ForestGEO's CTFS R package facilitates the analysis of tree and stem measurements

from forest dynamics plots (Anderson-Teixeira *et al.*, 2015; Condit, 2012; Davies *et al.*, 2021).

339 Integrative approaches often combine multiple data sources and analytical techniques. Wani et al. 340 (2024) analyzed plant biodiversity in the Indian Himalayan Region (IHR) by aggregating data from 341 the Botanical Survey of India, eFlora of India, the India Biodiversity Portal, and Flowers of India. They 342 applied Sørensen's similarity index ('vegan' R package by Oksanen et al., 2022) to compare species 343 richness across states and used Chi-square tests on Pearson's contingency tables ('vcd' R package 344 by Meyer et al., 2020) to assess trait distribution between the Himalaya and Indo-Burma biodiversity 345 hotspots. Additionally, global plant growth trends were analyzed using logistic regression (nnte' R 346 package by Venables & Ripley, 2002) with trait data from GFIT (Weigelt et al., 2019).

Species Distribution Models (SDMs) are widely used to predict biodiversity patterns based on environmental factors. The use of GBIF data is frequent in SDMs to characterize species occurrence, identify conservation areas, and assess climate change impacts (Araújo *et al.*, 2019; Elith & Leathwick, 2009; Ferraille *et al.*, 2023; Jetz *et al.*, 2012, 2019). However, SDMs are susceptible to geographical and environmental biases, as biodiversity data is often collected opportunistically (Baker *et al.*, 2022; Botella *et al.*, 2020).

To address SDM limitations, the 'ibis.iSDM' R package (Jung, 2023) integrates multiple data sources into unified models. Deep Neural Networks (DNNs) implemented using PyTorch (Paszke *et al.*, 2019; Van Rossum & Drake, 2009), have demonstrated higher predictive accuracy than traditional SDMs (Brun *et al.*, 2024). Probabilistic modeling approaches offer alternative solutions, using inhomogeneous Poisson process (IPP) models for presence-only data and Bernoulli generalized

linear models (GLM) for presence/absence data to mitigate spatial sampling bias (Fithian *et al.*,
2014).

Bayesian mixed-effects models (Johnson et al., 2024) address hierarchical and correlative non-360 361 independence (e.g., temporal, spatial, and phylogenetic autocorrelations) using integrated nested 362 Laplace approximation (INLA). This approach is similar to the methodologies employed by Browning 363 et al. (2022) and Dinnage et al. (2020) in their biodiversity studies. Alternative techniques include 364 simple linear regression for long-term trends and Generalized Additive Models (GAMs) for nonlinear 365 short-term patterns. Temporal correlations are managed through ARIMA models, LOESS smoothing, 366 splines, and piecewise regression (Dornelas et al., 2013). Zou et al. (2024) demonstrated the 367 effectiveness of GAMs in controlling environmental covariates in forest leaf type distribution 368 analyses.

In general, correlative models tend to more accurately predict biodiversity trends. Modern biodiversity models increasingly incorporate remote sensing data to map biodiversity based on forest structure attributes (Hakkenberg *et al.*, 2023). Generalized Linear Mixed Models (GLMMs) and Generalized Least Squares (GLS) ('MASS' and 'nlme' R packages) have been applied to analyze structural drivers of alpha, beta, and gamma diversity (Pinheiro & Bates, 2022; Venables & Ripley, 2002).

A comprehensive approach to biodiversity modeling is demonstrated by Steidinger *et al.* (2019), who evaluated global mutualistic symbioses between forest trees and soil microorganisms. Their study integrated environmental variables, functional traits, remote sensing, and forest plot data from The Global Forest Biodiversity Initiative (GFBI), employing Random Forest algorithms for predictive modeling. These techniques enhance biodiversity research by leveraging large-scale datasets to infer ecological relationships.

Despite these advances, challenges persist in integrating remote sensing, satellite imagery, and environmental datasets. Most studies focus on species-level patterns, often neglecting ecosystem functions and species interactions (Heberling *et al.*, 2021). Future research should emphasize multisource data integration and interdisciplinary modeling approaches to improve biodiversity predictions and conservation strategies.

386 **7 Final considerations**

The integration of multimodal data offers a promising approach to model, predict, and understand plant biodiversity changes over time and geography (Dornelas *et al.*, 2013; Paul & Shanta Kumar, 2020; Wauchope *et al.*, 2021). The use of environmental variables could provide valuable insights into species abundance trends, ecosystem dynamics, and responses to climate change, but the lack of high-resolution, long-term, and harmonized datasets implied significant challenges (Balzan *et al.*, 2018; Beck *et al.*, 2023).

393 Several challenges present the integration of remote sensing data with biodiversity data. These 394 include technological difficulties in accessing heterogeneous data with unified protocols, 395 discrepancies and uncertainties between remote sensing and in situ data, spatial misalignment in 396 biodiversity datasets, and the discontinuity of remote sensing data. These issues complicate the 397 prediction of biodiversity patterns from these datasets (Guralnick & Neufeld, 2005; Hakkenberg et 398 al., 2023; Mo et al., 2023; Saran et al., 2022). Addressing these integration challenges is essential 399 for leveraging the full potential of remote sensing and geographic information systems in advancing 400 biodiversity research.

Additionally, technical incompatibility, incomplete or outdated datasets, and discrepancies in spatial
 and temporal resolutions complicate the synthesis of biodiversity data across portals (Huber *et al.*,

2021; IPBES, 2019). Robust computational tools and standardized protocols are necessary to solve
interoperability across databases, ensuring high-quality data integration (Halbritter *et al.*, 2024;
Heberling *et al.*, 2021; Latz *et al.*, 2024). By addressing these challenges, researchers can analyze
biodiversity patterns and predict changes with improved precision, using quality data from sources
such as ForestForward.

408 Integrating and standardizing historical datasets remains important for gaining insights into the 409 evolution of biodiversity over the past century, even if remote sensing and real-time satellite data 410 from programs like Copernicus Sentinel is increasingly being used to monitor biodiversity. Even 411 though the classical datasets are sparser, they register information over a very prolonged time 412 period that is absent from remote sensing data. In addition, the accuracy of observations made with 413 the newer technologies is harder to validate and depends on several factors: training data quality, 414 canopy coverage - open landscapes with small plants are harder to monitor-, and habitat size, 415 among others (Paz et al., 2022; Schweiger & Laliberté, 2022).

Modern approaches base on artificial intelligence, such as Large Language Models, have been implemented in medicine to address the limitations of unstructured data (Choi *et al.*, 2024; Shool *et al.*, 2025). Similarly, Physics-Informed Graph Neural Networks have been implemented to integrate and analyze spatiotemporal environmental data with discontinuous records (Liu & Pyrcz, 2023). Recently, ecological studies adopted these approaches to study species distribution and natural phenomena using images (Espitalier *et al.*, 2025; Taccari *et al.*, 2024).

In conclusion, quantifying and predicting plant diversity changes through time, contributes to identifying causes that affect the species diversity and to detect critical points to diversity conservation (Johnson *et al.*, 2017; Reba & Seto, 2020; Tittensor *et al.*, 2014). Understanding how plant diversity responds to environmental parameters is essential for developing effective

426 conservation strategies, managing ecosystems, and attaining environmental and human health 427 (FAO, 2021; Humphreys *et al.*, 2019). Combining modern tools with traditional methods for 428 multimodal data integration can enhance our understanding and prediction of changes in plant 429 biodiversity on a large scale.

430 Author Contributions

Emilce Soledad Martinez: conceptualization, investigation, visualization, writing - original draft
preparation, writing - review and ending. Eva Tejada-Gutiérrez: writing - review and ending. Jordi
Mateo-Fornes: writing - review and ending. Francesc Solsona: writing - review and ending, Raquel
Defacio: writing - review and ending. Rui Alves: conceptualization, funding acquisition,
investigation, project administration, resources, supervision, visualization, writing - original draft
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439 **Competing Interests**

440 The authors declare no competing interests.

441 Data Availability Statement

442 Data sharing is not applicable to this article as no new data were created or analyzed in this study.

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958 Figure legend

959 Figure 1. Database selection flowchart, adapted from the PRISMA statement flowchart (Page *et al.*,

960 2021).

- 961 Figure 2. Interconnectedness of biodiversity platforms.
- 962 Figure 3. Biodiversity analysis methodology and tools. EV= environmental variables, INLA=
- 963 integrated nested Laplace approximation, IPP= inhomogeneous Poisson process model, DNNs=
- 964 Deep Neural Networks, SDM= Species Distribution Modeling, iSDM= integrative Species Distribution

965 Modeling.

Tables

Dortal	Tayon	Data type	Download			Time coverage (years)					
Forta	Taxon		Reg	Options	Format	1500	1600	1700	1800	1900	2000
BIEN	Plantae	O, Tr, G, F, ED	yes	API**	NS						
BioTIME	all	Ts	yes	D	CSV						
CoL	all	TC	no	API, ColDP, DwC-A	CSV, TSV, TextTree						
EOL	all	T, Tr, R	no	API	JSON, CSV						,
ForestForward	Plantae	0, G	no	D	CSV						
GBIF	all	ED, O, T, G, S, I	yes	API, DwC-A	CSV						
GIFT	Plantae	O, TC, G, Tr, L, E, B, S, Sc	no	API**	NS					*	
MOL	all	0, TC, G, M	yes	D	CSV						
POWO	Plantae	TC, G, D, R, L, I	no	DwC-A	CSV						
Tropicos	Plantae	TC, O, G, R, I	yes	API, DwC-A	CSV						
TRY	Plantae	Tr, T, E, ED, G, R	yes	D	CSV, XLSX, PDF, PNG, TXT						
WFO	Plantae	TC, S, D, I, G, R	no	API, DwC-A	CSV, TXT						

968 Table 1. Open access platforms with plant data on a global scale.

969 O= species occurrences, T= taxonomy, TC= Taxonomic checklist, G= geographic distribution, S= species status, I= images, D= descriptions, R= references, L= lifeform, Tr= trait data, E= environmental

970 variables, ED= event date, F= phylogenies, Ts= time series of species abundances, B= bioclimatic data, Sc= Socio-economy data, M= distribution maps of specie richness, rarity, and alien species, Reg=

971 registration, API= application programming interface, D= directly from the portal, CoLDP= CoL data package, DwC-A= Darwin Core Archive, NS= not specified. *Species names derived from heterogeneous

972 resources and published over a timespan of about one hundred years (Weigelt et al., 2019). **Required: BIEN R package (Enquist et al., 2016) and GIFT R package (Denelle et al., 2023). Italicized letters

973 indicate incompleteness of information.

Standardization/Integration Taxonomic/Nomenclature Checklist/Tool used Portal Reference Framework conflict resolution WFO, BIEN DwC, VegX TNRS Enquist et al. (2016) Tropicos BioTIME Table system (csv, SQL) CP Dornelas et al. (2018) COLDP, DwC-A, ACEF, WCVP, WFO, Döring & Ower (2019; CoL Multiple CA Unique Identifiers (DOIs) ChecklistBank Hobern et al. (2021) CoL, Tropicos, EOL DwC, DwC-A, DwC-M, DC Multiple CA Parr et al. (2016) WFO Tejada-Gutiérrez et al. FF ETL GBIF (2024) CoL, WCVP, GBIF DwC, EML TBS Wieczorek et al. (2012) Tropicos, WFO WCVP, WFO, Denelle et al. (2023); GIFT DwC, HC TNRS Tropicos Weigelt et al. (2019) Jetz et al. (2012); DwC, EML, HC CoL MOL CA Wieczorek et al. (2012) POWO TBS CoL, GBIF DwC Govaerts et al. (2021) WFO, TRY EAV, OBOE TNRS Kattge et al. (2020) Tropicos DwC, Unique Identifiers Borsch et al. (2020); WFO (WFO ID), DwC-A, Dublin TBS, TENs Botalista Palese et al. (2019) Core

975	Table 2. Standardization Frameworks and Taxonomic Conflict Resolution Mechanisms in Global
976	Biodiversity Databases.

979 Language, ETL= Extract, Transform, Load processes, EAV= Entity-Attribute- Value, FF= ForestForward, HC= Humboldt Core, VegX=
 980 Vegetation data exchange.

 ⁹⁷⁷ ACEF= Annual Checklist Exchange Format, CA= Checklist Approach, COLDP= CoL Data Package, CP= curation process, DC= Dublin core,
 978 DwC= Darwin core, DwC- A= Darwin core archive, DwC-M= Darwin core MeasurementOrFact extension, EML= Ecological Metadata

-			-
Integration Standards/ Ontologies	Key Features	Portal	References
OBOE, TOP, PTH	MS, TV, P	TRY, BIEN, GFIT	Garnier <i>et al.</i> (2017); Pérez-Harguindeguy <i>et</i> <i>al.</i> (2013); Kattge <i>et al.</i> (2020)
OBO (TO, PATO, ENVO)	SVS	EOL (TraitBank)	Baker <i>et al</i> . (2014) ; Parr <i>et al.</i> (2016)

982 Table 3. Integration of trait data standards and ontologies in global biodiversity databases.

983 ENVO= Environment Ontology, MS= Metadata Schema, OBO= Open Biological and Biomedical Ontologies, OBOE= Extensible

984 Observation Ontology, PTH= Plant Trait Handbook, PATO= Phenotypic Quality Ontology, P = Protocols, SVS= Structured Vocabulary

985 Systems, TO= Plant Trait Ontology, TOP= Thesaurus of Plant Characteristics, TV= Trait Vocabulary.

986	Table 4.	Sources	for	environmental	data.

Course	Resolution Source Data (km) Web site	Web site	Time period (years)			
Source	Data	(km)	Web site	1800	1900	2000
GloH2O	pp, Tº, W, sH, Lc	10	https://www.gloh2o.org		1980	
NOAA	Tº, pp, w, p, g	25- 250	https://www.noaa.gov/	1880		
NASA	Tº, pp, g, ae, p, Lu, Lc, tp	25- 100 *	https://www.earthdata.nasa.gov/	1880		
ECMWF	Tº, pp, w, r, g, H, Lc	10	https://cds.climate.copernicus.eu/#!/home		1950	
WorldClim	Tº, pp, a, tp, B	1-5	https://www.worldclim.org/		1950	2000**
Climate Explorer	Tº, pp, p, W, g	250 *	https://climexp.knmi.nl/selectfield_obs2.cgi ?id=someone@somewhere	1850		
ESA	SI	0.01-0.06	https://dataspace.copernicus.eu/explore- data/data-collections/sentinel- data/sentinel-2			2015
CHELSA	Tº, pp, B	1	https://chelsa-climate.org/		1979	2013
USGS	SI	0.03	https://www.usgs.gov/landsat- missions/landsat-data-access		1972	

987 GloH2O= Köppen-Geiger Maps, NOAA= National Centers for Environmental Information, NASA= National Aeronautics and Space 988 Administration, ECMWF= European Centre for Medium-Range Weather Forecasts (The Copernicus programme), ESA= European Space

Administration, Eckiwi - European Centre for Median Nange Weather Forecasts (The Opeanicus programme), Eak- European Space
 Agency- Sentinel 2 (Copernicus programme), USGS= United State Geological Survey- Landsat (NASA), CHELSA= Climatologies at high
 resolutions for earth's land surface areas.

991 Pp= precipitation, T° = air temperature, w= wind, W= water levels/sea levels, sH= soil moisture, Lc= land-cover, Lu= land-use, p= 992 atmospheric pressure, g= greenhouse gases (CO2, CH4), ae = aerosols, tp= topography, r= radiation, H= humidity, a= altitude, B=

atmospheric pressure, g= greenhouse gases (CO2, CH4), ae = aerosols, tp= topography, r= radiation, H= humidity, a= altitude, B=
 bioclimatic variables, SI= satellite images.

994 * Varies by dataset, ** Projections up to 2100.

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Index Name	Equation	Key Features	Reference
Margalef	$R_M = \frac{(S-1)}{\ln\left(n\right)}$	Corrects richness by sample size using the natural log of total individuals.	Margalef (1958)
Menhinick	$R_{Mn} = \frac{S}{\sqrt{n}}$	Relates number of species to the square root of total individuals.	Menhinick (1964)
Shannon- Wiener	$H' = -\sum_{i=1}^{S} (p_i)$ $\ln \ln p_i$	Consider both richness and evenness. Sensitive to rare species.	Shannon & Weaver (1949)
Simpson	$D_{Si} = \sum_{i=1}^{S} p_i^2$	Probability that two individuals randomly selected belong to the same species.	Simpson (1949)

1001 Table 5. Summary of common diversity indices





