

Short periods dominate mast seeding across diverse tree species

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19 **Abstract**

20 Mast seeding, synchronous and highly variable reproduction among perennial plants, impacts
21 ecosystem dynamics and species interactions. However, the extent of periodicity in mast
22 seeding, defined as cyclical but not strictly regular intervals between reproduction, remains
23 poorly understood. Here, we used autoregressive analyses on seed production data from 556
24 populations across 20 tree species, with time series ranging from 10 to 62 years, to quantify
25 the prevalence and length of masting periods. We found widespread periodicity, predominantly
26 characterised by a short period of 2–3 years. Although periodicity was common, the signal was
27 often weak, indicating limited explanatory and predictive power. This period length aligns with
28 theoretical predictions that balance the ecological benefits of predator satiation, with costs such
29 as missed reproductive opportunities. Extended period lengths (>4 years) were uncommon (2%),
30 suggesting that longer periods may be less ecologically advantageous or subject to specific local
31 conditions. Climate and elevation have limited and species-specific effects on period length and
32 strength, implying local adaptation in cue sensitivity and resource accumulation. Our findings
33 emphasise the adaptive value of short periods in mast seeding, likely reflecting consistent
34 evolutionary constraints on reproductive timing across diverse ecological conditions.

35 *keywords:* | masting | masting period | seed production | periodicity | autoregressive model |
36 resource pulses

38 **Introduction**

39 Mast seeding, or masting, is synchronous and highly variable reproduction among years by a
40 population of perennial plants (Kelly, 1994; Bogdziewicz *et al.*, 2024). The seed pulses that result
41 from masting influence plant recruitment, demographic processes, and ecosystem dynamics,
42 including effects on seed consumers, their predators, and associated parasites (Ostfeld *et al.*,
43 2000; Hackett-Pain *et al.*, 2022; Seget *et al.*, 2022), as well as nutrient cycling and abundance
44 of mycorrhizal fungi (Müller-Haubold *et al.*, 2015; Michaud *et al.*, 2024). Although masting
45 can occur without periodicity — populations may vary and synchronise even when reproductive
46 peaks are irregularly spaced — whether masting follows predictable periods has long interested
47 ecologists and foresters (Elton, 1924). However, periods have been difficult to test largely due
48 to analytical constraints (Bogdziewicz *et al.*, 2023). Exceptions include Allen *et al.* (2012),
49 who applied ordinal time series analysis and detected a 7-year periodicity in seed production
50 of New Zealand mountain beech (*Nothofagus solandri*), and Shibata *et al.* (2020), who used
51 autoregressive models to show that warming reduced the period length in masting of Japanese oak
52 (*Quercus crispula*) from 3–4 years to just 2 years. These studies show that, despite the inherent
53 difficulties, it is possible to detect masting periods using a range of statistical approaches, which

54 have also revealed potential links with climate. However, broad-scale analyses comparing
55 periodicity across populations of multiple species and climates remain limited.

56 Prior research has mostly centred on the related concept of masting return intervals, which
57 refers to the average interval between large seed production events. Return intervals provide a
58 descriptive summary of the spacing between high-seeding years but do not capture the underlying
59 temporal dynamics of reproduction. Return interval studies have often reported longer masting
60 periods, sometimes exceeding 5–10 years, particularly in tropical and temperate tree species.
61 For example, return intervals for three North American oaks varied from 2 to 4 years (Sork *et al.*,
62 1993), while in tropical forests, Igarashi *et al.* (2024) reported 2 to 10 year return intervals in 18
63 dipterocarp species. In four temperate forest species, Nussbaumer *et al.* (2016) reported masting
64 frequency (the reciprocal of return interval) ranging from 1.64 to 10 years. Some studies report
65 ranges of typical masting return intervals, noting that return intervals are irregular over time
66 (Övergaard *et al.*, 2007; Broome *et al.*, 2007; Wagner *et al.*, 2010). Moreover, numerous manuals
67 and monographs used by foresters and wildlife managers include tables of species-specific
68 masting return intervals (Burns *et al.*, 1990; Young & Young, 1992), because such information
69 can optimise harvesting schedules, guide forest restoration, and support wildlife conservation
70 dependent on mast resources (Kettle *et al.*, 2010; Köhnke *et al.*, 2020; Bregnard *et al.*, 2021).
71 While the identification of return intervals may be of practical value, such assessments inevitably
72 depend on the selected threshold defining a mast year (Bogdziewicz *et al.*, 2024). Because return
73 intervals depend on arbitrary thresholds used to define mast years, they cannot reliably capture
74 variation in seed production behaviour across species or environmental gradients, including
75 elevation. Here, we return to the original concept of masting periods, analysing autoregressive
76 patterns (Bjørnstad *et al.*, 2008; Shibata *et al.*, 2020) in seed production across 556 populations
77 of 20 tree species, offering a broad-scale evaluation of periods across and within species.
78 Understanding these periods provides insight into temporal patterns of seed production, with
79 implications for the evolutionary drivers of masting and broader ecosystem dynamics, such as
80 resource availability and consumer–producer interactions.

81 From an evolutionary perspective, the masting period reflects the degree of reproductive de-
82 lay, shaped by the cost–benefit balance of interannual variation in seed production (Bogdziewicz
83 *et al.*, 2024). Major costs include missed reproductive opportunities, which can reduce popula-
84 tion growth rates (Vacchiano *et al.*, 2021), and increased density-dependent seedling mortality
85 associated with concentrating reproduction in large, intermittent events (Visser *et al.*, 2011;
86 Huang *et al.*, 2021; Seget *et al.*, 2022). These constraints select against prolonged periods of re-
87 productive delay (Bogdziewicz *et al.*, 2024), particularly in environments with low productivity
88 and high background mortality, where the risks associated with delaying reproduction may be
89 greater (Waller, 1979). In contrast, two major benefits, known as economies of scale, can favour
90 delayed reproduction (Bogdziewicz *et al.*, 2024). First, alternating between years of low and
91 high seed production allows plants to starve and then overwhelm scarce seed consumers, thereby
92 reducing seed predation rates (Zwolak *et al.*, 2022). This mechanism is particularly effective

93 against consumers with low mobility, high dietary specialization, and short lifespans, such as
94 many insect species (Kelly & Sork, 2002; Zwolak *et al.*, 2022). For example, populations of
95 seed-predating micromoths are often highly vulnerable to even a single year of seed scarcity
96 (Yasaka *et al.*, 2003; Żywiec *et al.*, 2013). Secondly, large and synchronised flowering enhances
97 pollination success by increasing floral density (Venner *et al.*, 2016; Szymkowiak *et al.*, 2025;
98 Crone & Rapp, 2025). While reproductive delay and interannual variation in flowering do not
99 directly improve pollination rates, they allow resource accumulation that helps populations ex-
100 ceed the flowering threshold needed for efficient pollination (Kelly *et al.*, 2001). Where species
101 or populations cannot maintain high flowering effort every year, selection may favour delayed
102 flowering that enables resource build-up (Kelly *et al.*, 2001; Bogdziewicz *et al.*, 2020; Kelly,
103 2020).

104 From a proximate perspective, periodicity arises from interactions between resource dynam-
105 ics and weather cues (Satake & Bjørnstad, 2008; Kelly *et al.*, 2025). Resource budget models
106 propose that plants must accumulate sufficient resources before high seed production occurs
107 (Crone & Rapp, 2014). Then, high seed production occurs when adequate resources align
108 with favourable weather cues, such as - in the case of species inhabiting boreal and temperate
109 regions - warm conditions in preceding years (Bisi *et al.*, 2016; Nussbaumer *et al.*, 2018; Hirsch
110 *et al.*, 2025). Because plant response to the weather cue depends on the levels of accumulated
111 resources, resource dynamics play the role of both promoter and suppressor of reproduction,
112 enabling plants to maintain periodic reproduction despite variability in cue frequency (Monks
113 *et al.*, 2016; Kelly *et al.*, 2025). Specifically, low resource levels can suppress reproduction
114 even when strong cues occur consecutively, thereby preventing successive high-seeding years
115 and reducing the risk of seed overexploitation by consumer populations (Kelly *et al.*, 2000,
116 2013, 2025). Conversely, high accumulated resource levels enhance plant sensitivity to weather
117 cues, allowing even moderate cues to trigger large seed production, thus preventing excessively
118 delayed reproductive episodes (Kelly *et al.*, 2025). These processes can stabilise the periodicity
119 of masting, optimising spacing between large-seeding years to maximise fitness benefits. How-
120 ever, environmental conditions influencing resource accumulation and weather cue frequency
121 introduce stochastic variation into these patterns. Consequently, quantitative assessments of
122 periodicity across multiple species are required to better understand how resource dynamics and
123 cue frequency interact at broader ecological scales.

124 To explore how environmental conditions influence reproductive dynamics, we tested whether
125 local climate and elevation affect masting periodicity. We expected warmer and wetter condi-
126 tions characterised by higher mean annual temperature (MAT) and precipitation (MAP), or
127 positive climatic water balance (CWB), to facilitate faster resource accumulation, resulting in
128 shorter periods. Conversely, populations inhabiting colder or drier environments were expected
129 to show longer periods, as resource build-up between reproductive events proceeds more slowly.
130 With increasing elevation, we also expected longer periods due to the combined effects of
131 lower temperatures, slower nutrient cycling, and shorter growing seasons that constrain resource

132 renewal.

133 Building on earlier methods (Bjørnstad *et al.*, 2008; Shibata *et al.*, 2020), we apply second-
134 order autoregressive models to quantify periodicity and evaluate how period length varies across
135 species and ecological contexts. We then relate these patterns to local climate and elevation,
136 testing whether variation in environmental conditions corresponds to systematic differences in
137 period length. Importantly, our focus on periodicity does not imply that masting events occur at
138 strictly regular or predictable intervals. Aside from the two-year “alternate bearing” observed
139 in a few species (Garcia *et al.*, 2021), there is little reason to expect strict periodicity in masting
140 time series. Instead, periodicity arises from interactions between stochastic weather cues and
141 internal plant resource dynamics, with period length shaped by selection processes.

142 **Materials and Methods**

143 **Seed production data** We obtained data from MASTREE+, a database that collects annual,
144 population-level records in perennial plants’ reproductive effort (Hackett-Pain *et al.*, 2022; Foest
145 *et al.*, 2024). The species selected for analysis were those that met the following criteria: data
146 were recorded on a continuous scale and included counts of seeds, fruits, or cones; data spanned
147 a minimum of ten distinct locations, each representing either a stand or a patch (excluding
148 regional-scale records); and each time series comprised at least ten years of data.

149 The final dataset included 20 species, encompassing 556 unique time series (Table S1).
150 All species displayed substantial interannual variation in seed production in our dataset, with
151 population-level coefficients of variation ranging from approximately 0.86 to 1.94. Elevation
152 data were included for species with available information and sufficient variation, defined as
153 multiple elevation points where differences in elevation across populations exceeded 100 meters.
154 This threshold was chosen to ensure meaningful differentiation in elevation values, enabling the
155 analysis to capture changes in climatic conditions. The dataset with elevation covered 10 species
156 and 141 populations.

157 **Climate data** Mean annual temperature (MAT), mean annual precipitation (MAP), and cli-
158 matic water balance were calculated for each unique location based on monthly data (1960–2020)
159 from the corresponding 1/24° (4 km) resolution in the TerraClimate dataset (Abatzoglou *et al.*,
160 2018). The CWB values were taken directly from TerraClimate, where they are computed as
161 precipitation minus reference evapotranspiration estimated using the Penman-Monteith equa-
162 tion. To evaluate the climatic representativeness of our sampled populations, we compared their
163 mean annual temperature (MAT) and precipitation (MAP) values to the full climatic range of
164 each species, defined as all grid cells within its native distribution. Climatic variables were ob-
165 tained from WorldClim 2.1 (Fick & Hijmans, 2017). To assess the climatic representativeness of
166 our sampled populations, we compared their mean annual temperature (MAT) and precipitation
167 (MAP) values to the full climatic range of each species based on species distribution poly-

168 gons. These range maps were compiled from EUFORGEN for European taxa (EUFORGEN,
169 2020), GBIF occurrence-based polygons for globally distributed species (Global Biodiversity
170 Information Facility, 2025), and digitized maps from the *Atlas of United States Trees* (Little,
171 1971–1978; Petry, 2024) for North American species. The average (\pm SD) coverage was 36.7%
172 (13.7) for MAT and 35.5% (18.8) for MAP (Fig. S1).

173 **Analysis** To estimate the period length, we calculated the second-order autoregression (AR2)
174 coefficients that relate current seed production (t) to past observations. The coefficients $1+a_1$ and
175 a_2 describe the statistical dependence of current seed production (t) on values in preceding years
176 within a second-order autoregressive process, capturing temporal autocorrelation rather than
177 fixed biological delays. The AR(2) framework, first formalised by Royama (Tom Royama, 1992)
178 and subsequently applied across diverse ecological time series, including insect outbreaks, rodent
179 population cycles, ungulate dynamics, and tree reproduction (Bjørnstad *et al.*, 1995; Cornulier
180 *et al.*, 2013; Ahrestani *et al.*, 2016; Shibata *et al.*, 2020), provides a simple and comparable way
181 to quantify the presence and strength of cyclic temporal structure in population-level data. This
182 structure may reflect interactions among underlying processes such as resource dynamics and
183 weather cues, but the model itself remains agnostic to their specific mechanisms. Species-level
184 differences in reproductive schedules, including seed maturation time, are therefore reflected in
185 variation in the estimated period length rather than in the coefficients themselves.

186 Each time series was fitted using the second-order autoregressive model

$$x_t = (1 + a_1) x_{t-1} + a_2 x_{t-2} + \varepsilon_t,$$

187 where x_t represents log-transformed annual seed production in year t , and x_{t-1} and x_{t-2} denote
188 values from one and two years prior, respectively. ε_t is the residual error term. Following
189 Royama (1992) and Bjørnstad *et al.* (2008), the coefficients a_1 and a_2 were estimated using
190 the Yule–Walker method implemented in the `stats::ar()` function in base R (R Core Team,
191 2020), ensuring consistent estimation across species.

192 Population dynamics were considered periodic when the estimated combinations of $1 + a_1$
193 and a_2 fell within the periodic region of the parabola diagram — that is, inside the left triangular
194 stability region and below the parabolic boundary that defines oscillatory dynamics (see Fig. 1),
195 following the stability criteria of Bjørnstad *et al.* (2008). Within this framework, gradients in
196 period length can arise through two main parameter changes (for examples, across populations,
197 see Fig. S2). First, increasing the $1 + a_1$ coefficient, while a_2 remains within the periodic region,
198 elongates the period length. The position of points in the parabola diagram determines the
199 period length: points between lines 2 and 2, as well as 2 and 3, indicate a two-year period;
200 between 3 and 4, a three-year period, and so on (Bjørnstad *et al.*, 1995, 2008; Cornulier *et al.*,
201 2013; Ahrestani *et al.*, 2016; Shibata *et al.*, 2020). Second, increasing a_2 while $1 + a_1$ remains
202 negative shifts points leftward within the periodic region of the parabola diagram (Fig. 1). This

203 change affects the strength of periodicity: more negative values of a_2 indicate stronger cycles,
 204 whereas higher values correspond to weaker or more irregular periodicity, or even non-periodic
 205 dynamics.

206 Points falling outside the triangle or falling inside the top right side of the triangle above
 207 the parabola region (i.e. the white space), are considered as non-periodic (Fig. 1). Because
 208 negative values of the $1 + a_1$ term are common in masting time series (Pearse *et al.*, 2020; Foest
 209 *et al.*, 2025), many populations naturally fall within the periodic region of the AR(2) parameter
 210 space, reflecting genuine ecological structure. Additionally, we note that although period length
 211 takes integer values in years, it is a ratio-scale quantitative variable rather than an ordinal one:
 212 differences between successive values (e.g., 2→3 years and 3→4 years) are equivalent, and
 213 ratios are meaningful.

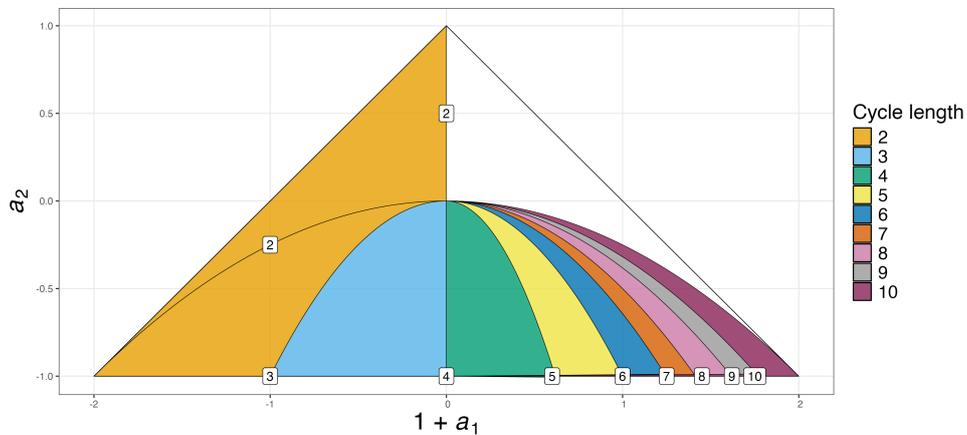


Figure 1: Parabola plot showing correspondence of combinations of $1 + a_1$ and a_2 coefficients values to period length. Coloured regions indicate oscillatory dynamics, with colours corresponding to different period lengths; white space indicates non-periodic dynamics. See Bjørnstad *et al.* (1995, 2008).

214 We used Bayesian second-order autoregressive models to analyse the relationships between
 215 climate variables and the autoregressive coefficients $1 + a_1$ and a_2 for each species separately.
 216 All predictors (MAT, MAP, and CWB) were centred and scaled prior to modelling to ensure
 217 comparability. Because these predictors were correlated, we fitted separate models for each cli-
 218 matic variable per species. Models were specified in brms (Bürkner, 2021) using the `bf()` and
 219 `brm()` functions, which allowed both autoregressive coefficients to be estimated simultaneously
 220 as multivariate responses. Default priors were used, and models were run with four MCMC
 221 chains of 10,000 iterations each (3,000 warm-up). To complement these analyses, we addition-
 222 ally fitted zero-truncated Poisson (ZTP) models treating period length as a discrete response,
 223 implemented in `g1mmTMB` (Brooks *et al.*, 2017). All analyses were performed in R v.4.2.3 (R
 224 Core Team, 2020).

Results

Percentage of periodic populations Based on temporal autoregression coefficients, almost all populations (N = 525 out of 556) across the 20 analysed species exhibited periodic masting behaviour, with an overall prevalence of periodicity of approximately 95%. For eight species, all populations were periodic, including four conifers (*Picea engelmannii*, *Araucaria araucana*, *Abies amabilis*, *Abies alba*) and four deciduous species (*Quercus douglasii*, *Quercus cerris*, *Fagus sylvatica*, *Betula pubescens*).

For species that included both periodic and non-periodic time series, the proportion of periodic populations ranged from 90.0% in *Alnus incana* to 96.7% in *Quercus petraea*. Other species with relatively lower percentages of periodic populations included *Quercus robur* (90.9%) and *Pinus sylvestris* (91.1%). A detailed summary of species-level proportions is provided in the Supplementary Information (Fig.S3).

Variation in period length across species The average period length across species was 2.65 years (± 0.49 SD) (Fig. 2), and was not correlated with the length of time series (Fig. S4). Note that even long time-series >40 years did not show differences in mean period length, although it was notable that longer period lengths were restricted to relatively short time-series (Fig.S4). Among the species analysed, *Quercus lobata* had the shortest period (2 ± 0 SD), showing highly consistent values across populations (all 11 populations had period = 2). In contrast, *Quercus cerris* had the longest period (3.27 ± 0.14 SD), also with relatively low variability. Most species exhibited period lengths closely centred around the overall mean, corresponding to a two- to three-year period. We found little evidence of longer period length, with only 15 time series showing periods longer than 4 years, mainly associated with *Pinus ponderosa* and *Pinus sylvestris* (Fig.2).

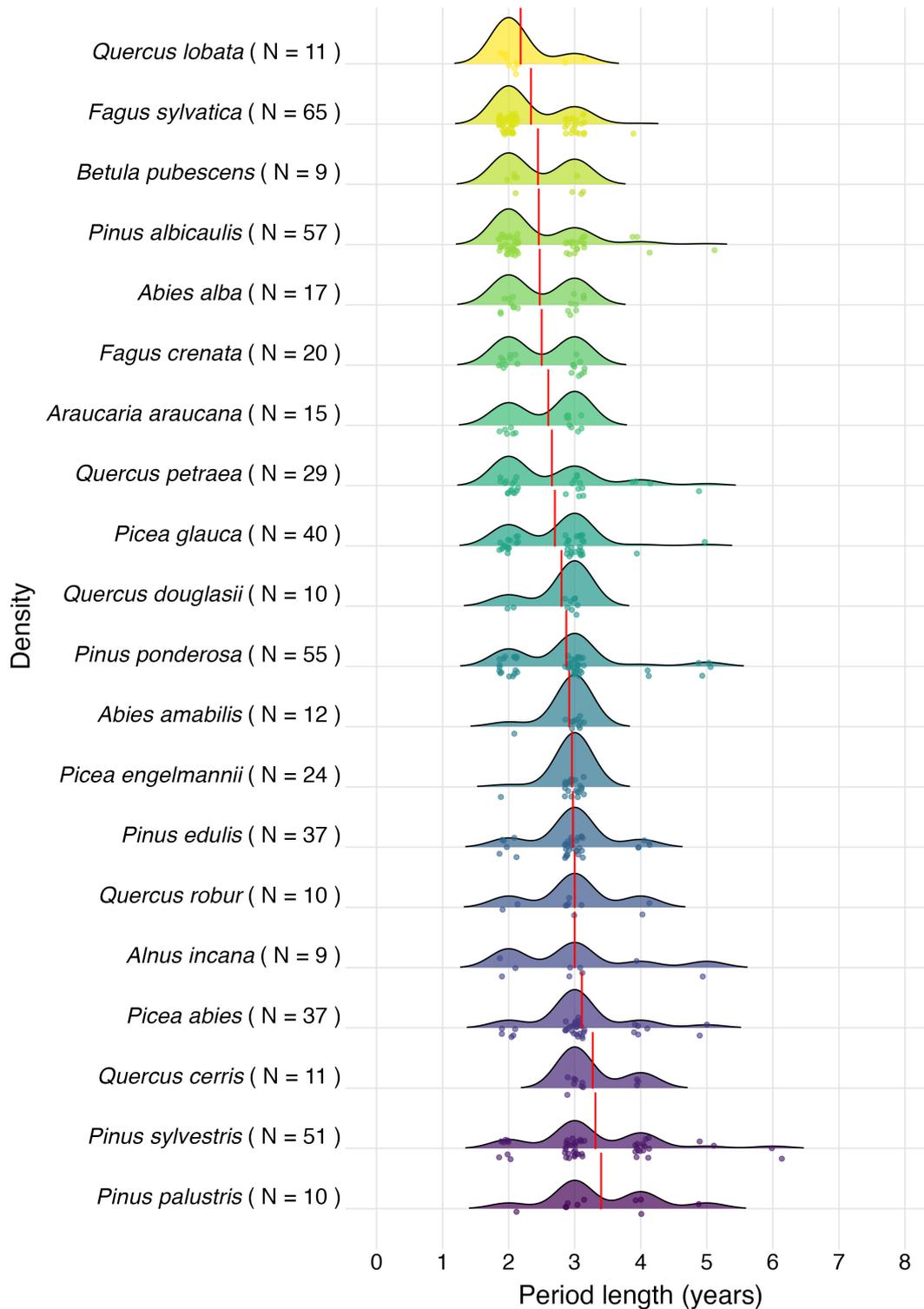


Figure 2: Inter- and intra-specific variation in period length (N = 525 periodic populations, 20 species). Each row represents the distribution of estimated period lengths (years) across populations of each species. Period length for each population was determined based on the position of the population $1 + a_1$ and a_2 coefficients on the parabola plot (Fig. 1). Numbers in brackets indicate the number of studied populations. Each point corresponds to a single population. All presented values are integers, and jitter was added to decrease overlap. Red vertical lines represent the species-level means. Non-periodic populations are excluded from the graph. The $1 + a_1$ and a_2 coefficients for each population are given in Fig. S5.

248 **Period length and climate** *Alnus incana*, *Fagus crenata*, *Fagus sylvatica*, *Picea abies*, *Picea*
249 *engelmannii*, *Picea glauca*, *Pinus albicaulis*, *Pinus palustris*, and *Quercus lobata* showed associ-
250 ations between mean annual temperature (MAT), mean annual precipitation (MAP), or climatic
251 water balance (CWB) and the autoregressive coefficients $1 + a_1$ and a_2 . The remaining 11 species
252 did not exhibit consistent relationships with these climatic variables. The diversity of slope
253 directions and significance levels indicates that these associations are species-specific.

254 Higher MAT was associated with longer periods in only one species, *P. abies*, as indicated
255 by a positive slope of MAT on $1 + a_1$. In contrast, higher MAT was associated with shorter
256 periods in three species, *P. albicaulis*, *P. engelmannii*, and *A. incana*, as indicated by its negative
257 effects on $1 + a_1$ (Fig. 3A). A positive a_2 slope with MAT in *P. abies* indicates that populations
258 occurring in warmer conditions tend to show weaker periodicity and less regular reproductive
259 dynamics.

260 Period length was also associated with MAP. Longer periods tended to occur under higher
261 precipitation in four species—*P. albicaulis*, *F. sylvatica*, *F. crenata*, and *A. araucana*, as indicated
262 by positive slopes of MAP on $1 + a_1$. Conversely, in five species, *Q. lobata*, *P. palustris*, *P.*
263 *engelmannii*, *P. glauca*, and *A. incana*, shorter periods were associated with higher MAP, as
264 indicated by its negative slopes on $1 + a_1$ (Fig. 3B). Positive slopes of a_2 in *F. crenata* and *A.*
265 *araucana* indicate that although wetter conditions are associated with longer periods, periodicity
266 tends to be weaker and reproduction less regular with increasing MAP.

267 Higher CWB values were associated with longer periods in *P. albicaulis* and *A. araucana*, as
268 indicated by positive slopes of CWB on $1 + a_1$. However, in *A. araucana*, as well as *F. crenata*,
269 the positive slope of a_2 suggests that, although wetter conditions were linked to longer periods,
270 periodicity tended to weaken. In contrast, *Q. lobata*, *P. palustris*, *P. glauca*, *P. engelmannii*,
271 and *A. incana* showed the opposite pattern, with shorter periods associated with more humid
272 conditions (Fig. 3C).

273 Although associations emerged in several species, most 95% credible intervals overlapped
274 zero, indicating that climatic variables were only weakly associated with variation in period
275 length across taxa. Complementary zero-truncated Poisson models yielded consistent results,
276 showing similarly weak associations between period length and climate across species (Fig. S6).

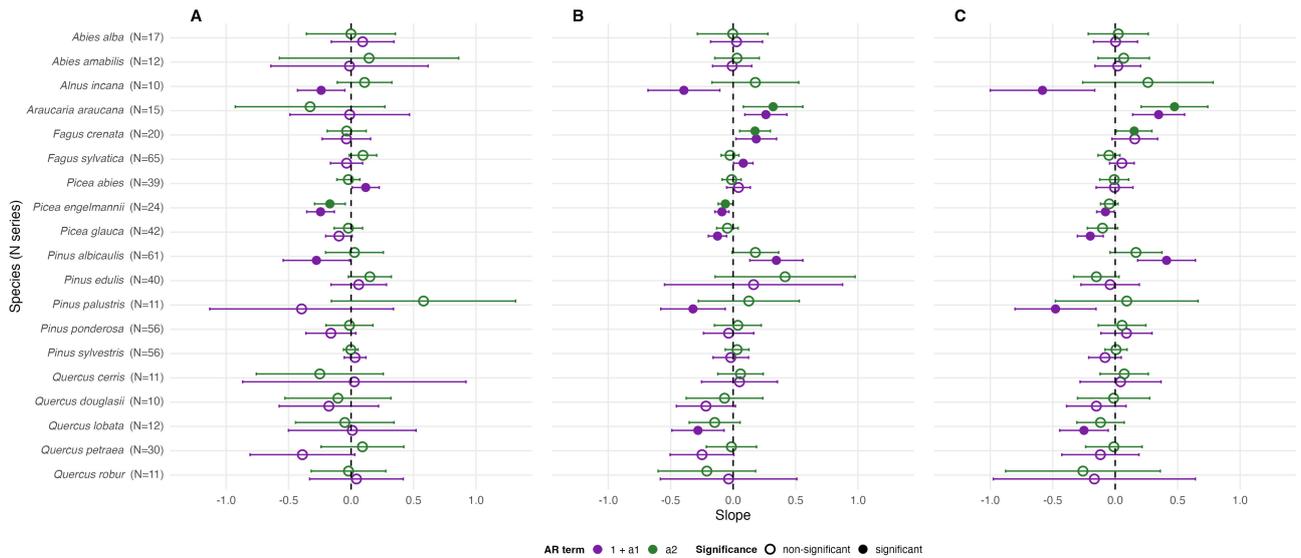


Figure 3: Associations between autoregressive coefficients ($1 + a_1$) and a_2 and local climate (MAT: mean annual temperature, MAP: mean annual precipitation and CWB: climatic water balance) across 525 populations representing 19 species. Results for *B. pubescens* are not shown due to model convergence issues. Panel A: MAT, B: MAP, and panel C: CWB. Rows represent individual species. Each point shows the estimated slope for a standardised climatic predictor, with error bars indicating 95% confidence intervals. Significant effects are shown as filled circles, while non-significant effects are shown as empty circles. Positive slopes of $1 + a_1$ are associated with longer period lengths, whereas negative slopes are associated with shorter period lengths. Positive slopes of a_2 indicate weaker or less regular periodicity, while negative slopes indicate stronger or more regular periodicity. To aid interpretation of these slopes, we provide a supplementary Figure S7 illustrating how fitted $1 + a_1$ and a_2 values from the same models move across the parabola in *Fagus sylvatica* along MAT and MAP gradients.

277 **Period length and elevation** The associations between elevation and autoregressive coefficients $1 + a_1$ and a_2 were generally weak, though statistically significant in seven species (Fig. S8).
 278
 279 At higher altitudes, estimated periods tended to be longer in four species—*A. araucana*, *F. syl-*
 280 *vatica*, *P. engelmannii*, and *P. albicaulis*, as indicated by positive relationships with $1 + a_1$, and
 281 with shorter periods in two species, *P. abies* and *P. glauca*. The strength of periodicity tended
 282 to weaken with increasing elevation in *F. crenata* and *P. engelmannii*, as reflected by positive
 283 relationships with a_2 , but became stronger in *P. glauca*, which showed a negative relationship
 284 with a_2 . In other species, the associations between elevation and both $1 + a_1$ and a_2 were not
 285 significant (Fig. S8). Overall, although several species exhibited significant associations, these
 286 patterns were weak, suggesting that elevation is only weakly linked to variation in masting peri-
 287 od. Complementary zero-truncated Poisson models of period length (Fig. S9) showed similarly
 288 limited associations.

289 Discussion

290 Our findings show that periodicity is widespread across diverse tree taxa. Specifically, the 20
 291 species — including angiosperms and gymnosperms from boreal, temperate, and Mediterranean
 292 biomes — consistently fell into parameter space that indicated periodicity, with a relatively short

293 mean period length of 2–3 years. We interpret this as evidence that common ecological factors,
294 most likely the economies of scale and the inherent costs of delayed reproduction, play key roles
295 in governing masting behaviour across broad phylogenetic and environmental settings. This
296 suggests that the selective pressures favouring reproductive delay converge to a similar outcome
297 across species.

298 The mean period length of 2–3 years fits well with predictions based on economies of
299 scale. Interannual variation in seed production can effectively reduce seed predation by insects,
300 whose short life spans mean that a delay of just 1–2 years is sufficient to lock them into periods
301 of starvation and satiation (Zwolak *et al.*, 2022). Prolonging the delay would increase the
302 costs, such as missed reproductive opportunities (Rees *et al.*, 2002; Visser *et al.*, 2011; Tachiki
303 & Iwasa, 2010), possibly with little additional benefit to reproductive success. In contrast,
304 extending the period length specifically to starve and satiate consumers with longer lifespans,
305 such as generalist vertebrates, is unlikely to be effective. These animals are typically mobile and
306 capable of exploiting alternative resources or migrating to areas of higher seed availability during
307 periods of scarcity, making predator satiation less viable as a strategy against them (Curran &
308 Leighton, 2000; Bogdziewicz *et al.*, 2022; Zwolak *et al.*, 2022). Similarly, while accumulating
309 resources to reach a critical flower density is essential for successful pollination (Kelly *et al.*,
310 2001), extending the delay beyond what is necessary increases the costs of missed reproductive
311 opportunities. Recent findings further suggest that the interaction between resource levels and
312 cue strength allows plants to fine-tune this delay, providing a proximate mechanism to maintain
313 the period length at the desired level (Monks *et al.*, 2016; Ascoli *et al.*, 2017; Kelly *et al.*, 2025).
314 Importantly, reliance on stochastic weather cues remains crucial, as regular periods could enable
315 seed predators to predict and exploit seed pulses, for example, through diapause (Maeto & Ozaki,
316 2003; Péliesson *et al.*, 2012).

317 Period length is only one among many metrics used to describe mast seeding patterns,
318 alongside more frequently used synchrony among individuals and interannual variation in seed
319 production (CV) (Koenig *et al.*, 2003; Lamontagne & Boutin, 2007; Qiu *et al.*, 2023). Natural
320 selection acts on physiological traits influencing reproductive synchrony and variability at the
321 proximate level, such as sensitivity to environmental cues (Kelly *et al.*, 2013; Bogdziewicz
322 *et al.*, 2020). These physiological traits produce seed production patterns that enhance fitness
323 at the ultimate level, primarily through economies of scale (Kelly, 1994; Pearse *et al.*, 2016;
324 Pesendorfer *et al.*, 2021). Various combinations of reproductive synchrony and individual
325 variability can generate a similar population-level period length, meaning that this pattern arises
326 from group-level dynamics rather than direct selection on period length itself. For example,
327 low synchrony but high variability among individual trees could yield similar periodicity as
328 high synchrony with moderate variability (Koenig *et al.*, 2003), a pattern reflected in our
329 data by the weak correlation between CV and period length (Fig. S10). The specific ecological
330 context, particularly seed predator mobility, generation time, and feeding behaviour, shapes these
331 combinations (Koenig *et al.*, 2003; Bogdziewicz *et al.*, 2021). Relatively immobile predators,

332 like micromoths, may be satiated by seed production from single trees, reducing selective
333 pressure for synchrony (Nilsson & Wastljung, 1987; Satake *et al.*, 2004). In turn, mobile
334 predators might be attracted to large seed crops, potentially selecting against high interannual
335 variability or synchrony (Koenig *et al.*, 2003). Despite these different selective contexts, period
336 lengths frequently converge around 2–3 years, reflecting common costs and benefits associated
337 with masting strategies.

338 Notably, we found little evidence supporting long period length. Only 15 time series
339 exhibited periods exceeding four years, and these were generally derived from relatively short
340 datasets compared to the period lengths identified. This scarcity of long periods contrasts
341 with some reports in the literature indicating longer masting periods (Maki, 1952; Wagner
342 *et al.*, 2010). Such discrepancies might arise from masting periods being identified based on
343 arbitrary thresholds to define mast years (Bogdziewicz *et al.*, 2024). Conversely, our findings of
344 predominantly short periods (2–3 years) align with the results of (Qiu *et al.*, 2023), who identified
345 periodicity averaging around three years in 142 species. Furthermore, the detected periods, when
346 present, were often weak (see, e.g., Fig.S2) suggesting that these inherent periods account for only
347 a modest fraction of total variability in seed production. Consequently, the periodicity identified
348 here might offer limited utility for predicting future seed crops. Nevertheless, forecasting models
349 that increasingly link seed production to weather variability (Journé *et al.*, 2023; Wion *et al.*,
350 2025) could potentially benefit from incorporating patterns indicating a higher likelihood of
351 masting events as more time passes since the last occurrence.

352 The effects of local climate on period length were generally weak or absent, as observed in
353 seven species lacking clear climatic associations. This finding is consistent with recent work
354 showing that intraspecific variation in masting behaviour is often unrelated to climate, with no
355 consistent support for the environmental stress hypothesis across species or masting metrics
356 (Foest *et al.*, 2025). One possible explanation is that climate ranges within species were too
357 narrow to detect consistent effects. However, most species in our dataset exhibited substantial
358 within-species variation in both MAT and MAP, and sampled populations span large portions
359 of the species' climatic ranges, including those with weak or absent association, making this
360 explanation unlikely (Figure S1). Superficially, our results appear to contrast with a recent large-
361 scale synthesis reporting longer masting periods in drier climates (Qiu *et al.*, 2023). However,
362 period lengths in that study varied only narrowly (most 2.5–3.5 years), so any aridity effect is
363 small and consistent with convergence toward a common period across climates.

364 This limited effect of climate on period length may instead reflect local adaptation, with the
365 thresholds for climatic cues triggering high-seeding years varying across sites (Kon *et al.*, 2005;
366 Foest *et al.*, 2025). For example, species responding to warm temperatures might not exhibit
367 lower cue frequencies at colder sites, as the "warm" threshold may adjust to local climate norms
368 (Foest *et al.*, 2024). Similarly, local adaptation may affect how climate influences resource
369 dynamics, making absolute climatic values less predictive of resource accumulation rates (Piper
370 & Fajardo, 2024). When climate influences period length, temperature and precipitation can

371 function as reproductive cues or vetoes (Kelly *et al.*, 2013; Bogdziewicz *et al.*, 2019). Warmer
372 temperatures may shorten period length by frequently triggering seed production (Shibata *et al.*,
373 2020), although colder conditions, such as late frosts, could lengthen masting period by inhibiting
374 reproduction (Inouye, 2008; Schermer *et al.*, 2020). Similarly, precipitation variability is likely
375 to influence reproductive dynamics; drought conditions can correspond with longer masting
376 periods by limiting resources, whereas excessive precipitation may reduce pollination success,
377 leading to longer resource accumulation periods between mast events (Espelta *et al.*, 2008;
378 Fleurot *et al.*, 2024). Thus, the relationship between climate and period length appears complex
379 and influenced by species-specific adaptations and local environmental contexts.

380 Finally, we note that tropical tree species were not represented in our dataset, and therefore
381 our inferences should not be extended to tropical forests. In a recent synthesis, Qiu *et al.* found
382 that low volatility was common in tropical clades, and that periodicity was often not strongly
383 expressed for tropical taxa (Qiu *et al.*, 2023).

384 In summary, our analyses highlight the widespread occurrence of periodicity in masting
385 across the 20 tree species studied, predominantly converging toward a 2–3 year period length.
386 This consistent periodicity highlights the adaptive balance between the advantages of predator
387 satiation, primarily targeting specialised insect seed predators, and the risks associated with
388 prolonged reproductive delays, such as resource loss or missed reproductive opportunities. The
389 limited effect of climatic variables on periodicity across species suggests that local adapta-
390 tion likely modulates plant sensitivity to weather cues, complicating broad-scale predictions of
391 masting behaviour under changing climates. Interestingly, the rarity of longer masting periods
392 challenges previous assumptions derived from return interval studies, indicating that prolonged
393 periods may be ecologically exceptional rather than typical.

394

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623 **Supporting Information**

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625 M. Bogdziewicz

626

627 **Title:** Short periods dominate mast seeding across diverse tree species

628

Table S1: Summary of per-species sample size in our study. Reproduction (seed/cones, depending on species) was monitored annually; thus time series length is given in years.

Species name	Number of populations	Mean time series length	Maximal time series length
<i>Abies alba</i>	17	18	25
<i>Abies amabilis</i>	12	11	12
<i>Alnus incana</i>	10	18	22
<i>Araucaria araucana</i>	15	16	18
<i>Betula pubescens</i>	9	13	17
<i>Fagus crenata</i>	23	17	24
<i>Fagus sylvatica</i>	65	29	43
<i>Picea abies</i>	39	22	40
<i>Picea engelmannii</i>	24	36	41
<i>Picea glauca</i>	42	24	57
<i>Pinus albicaulis</i>	61	22	33
<i>Pinus edulis</i>	40	16	20
<i>Pinus palustris</i>	11	47	62
<i>Pinus ponderosa</i>	57	16	31
<i>Pinus sylvestris</i>	57	21	45
<i>Quercus cerris</i>	11	25	31
<i>Quercus douglasii</i>	10	30	41
<i>Quercus lobata</i>	12	27	41
<i>Quercus petraea</i>	30	13	23
<i>Quercus robur</i>	11	16	27

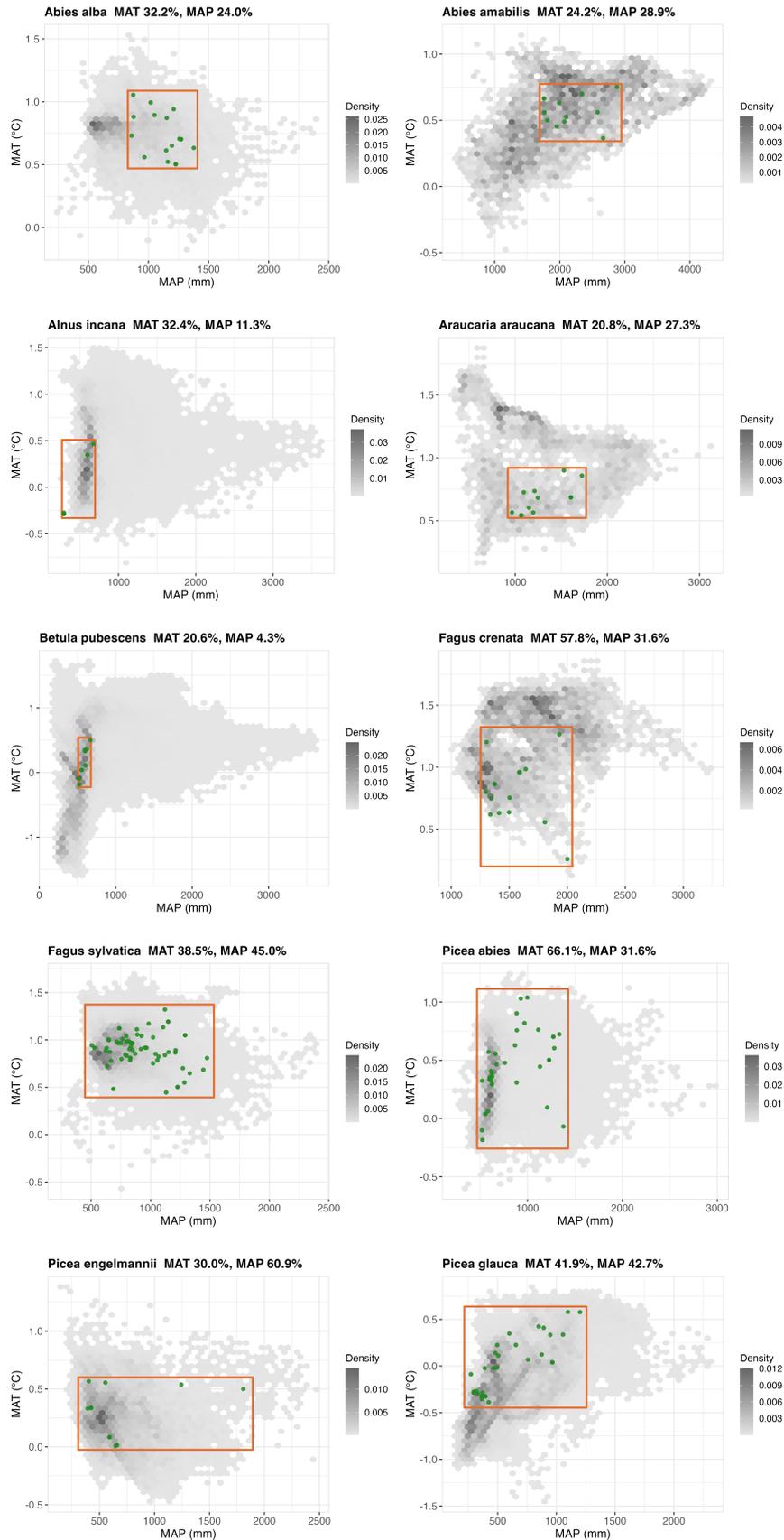


Figure S1: Climatic coverage of sampled populations relative to each species' full climatic distribution (A). Each green point represents one population; grey shading represents species-level climatic density estimated from WorldClim 2.1 (1 km resolution). Orange rectangles delineate the range of mean annual temperature (MAT) and precipitation (MAP) represented in our sampling. Numbers in panel titles indicate the percentage of the total species climatic range covered by the sample (MAT, MAP respectively).

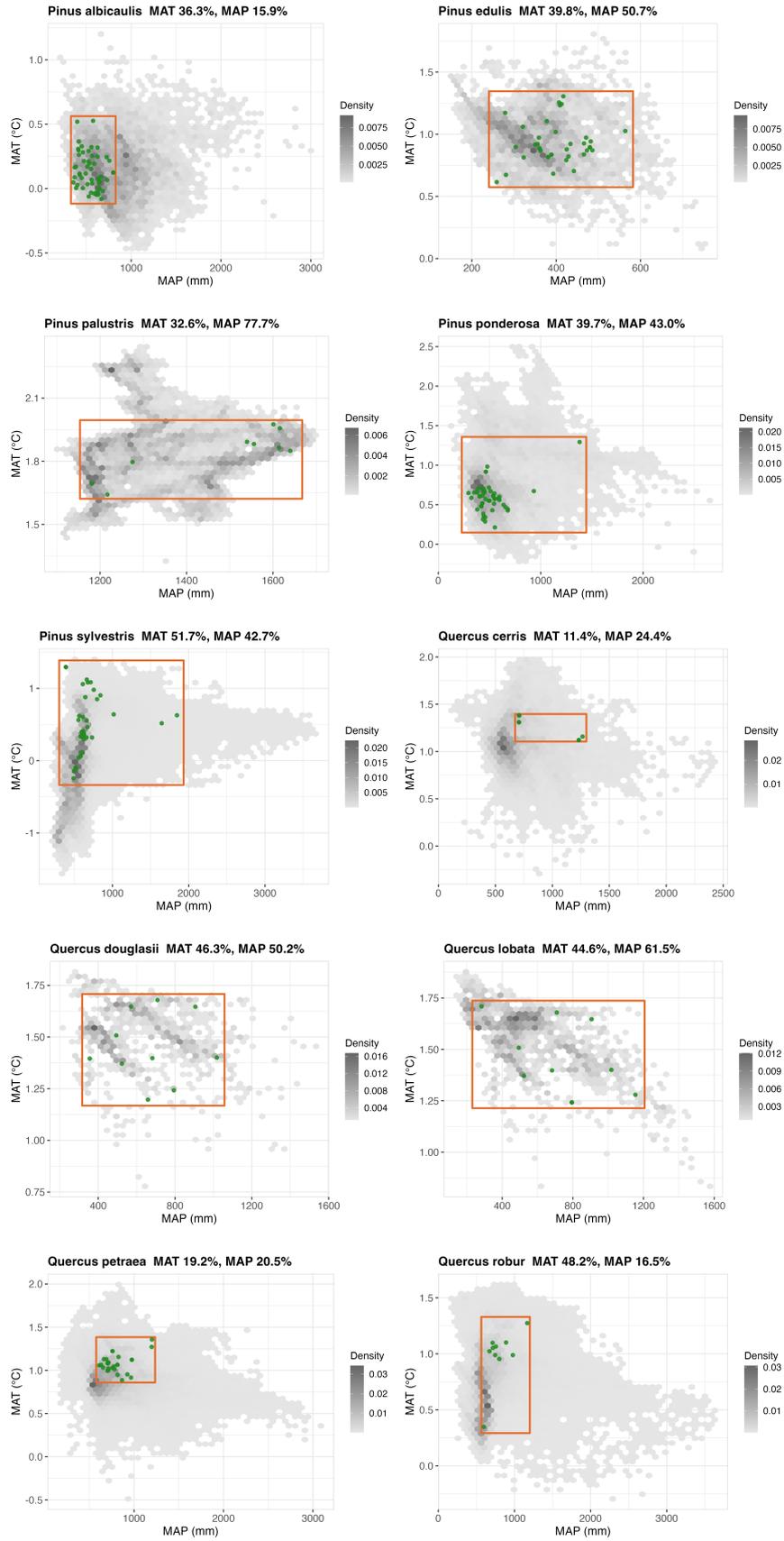


Figure S1: Climatic coverage of sampled populations (continued, B).

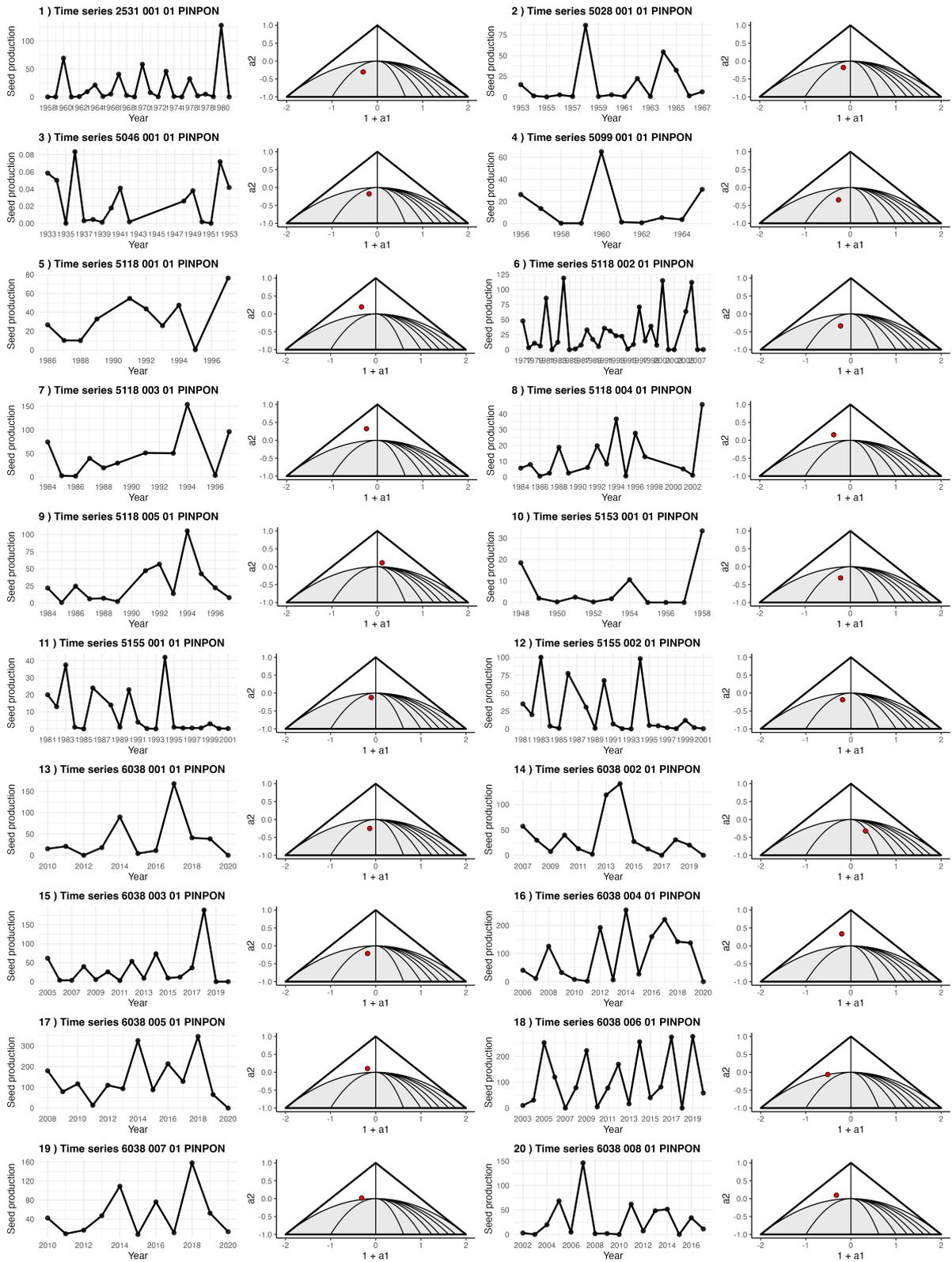


Figure S2: Time series of seed production and respective parabola plots for 20 populations of *Pinus ponderosa*. Each row contains two populations, with time series graphs (first and third columns) showing interannual variation in seed production and parabola plots (second and fourth columns) showing the period length of given time series based on $1 + a_1$ and a_2 coefficients from AR2 model.

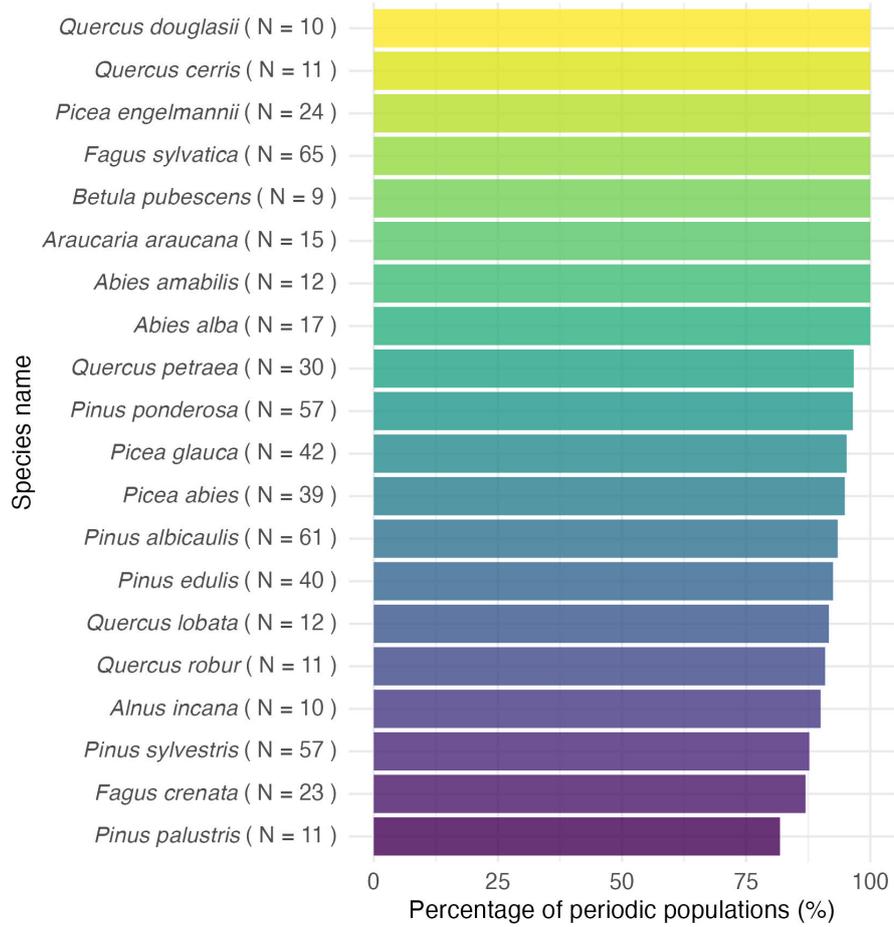


Figure S3: Percentage of periodic populations across species studied (N = 525 periodic populations, 20 species). The x-axis shows the percentage of periodic populations for each species. Whether a time series was categorised as periodic or not is based on whether a_1 and a_2 coefficients fell within the periodic region of the parabola plot (Fig. 1). The minimum time series length for each population was 10 years, with an average of 20 years (Table S1).

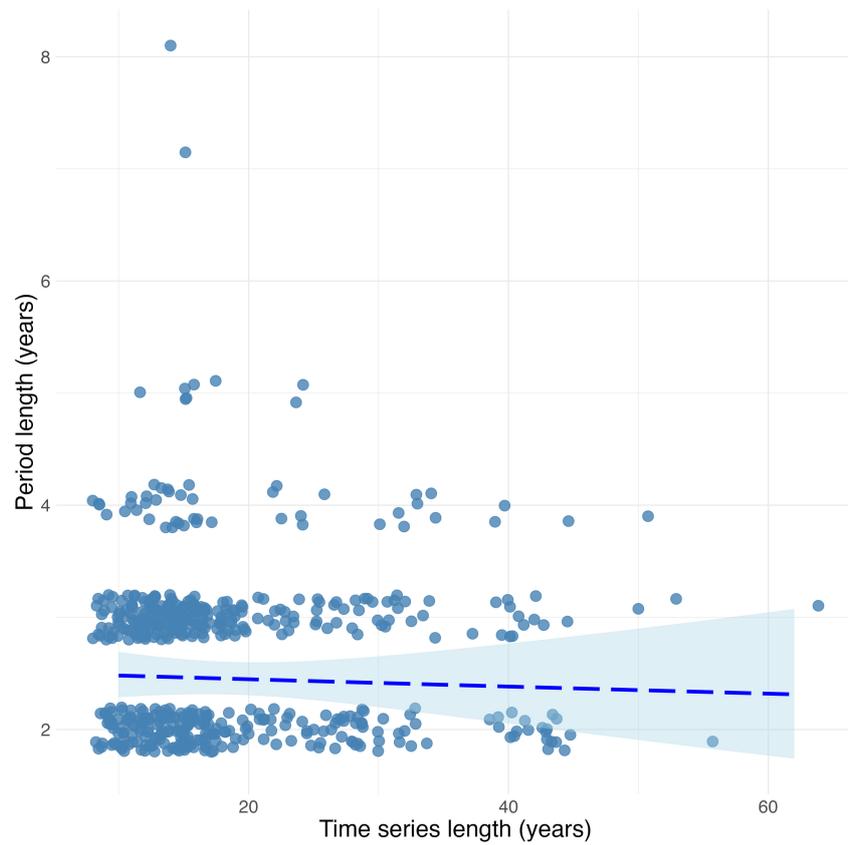


Figure S4: Relationship between period length and time series length. Each point represents an individual time series. A zero-truncated Poisson generalized linear model (GLM) was fitted to examine the relationship between period length as a response variable and time series length as an explanatory variable. The model showed no significant effect of time series length on period length ($p = 0.681$). The blue line represents model predictions, with 95% confidence intervals.

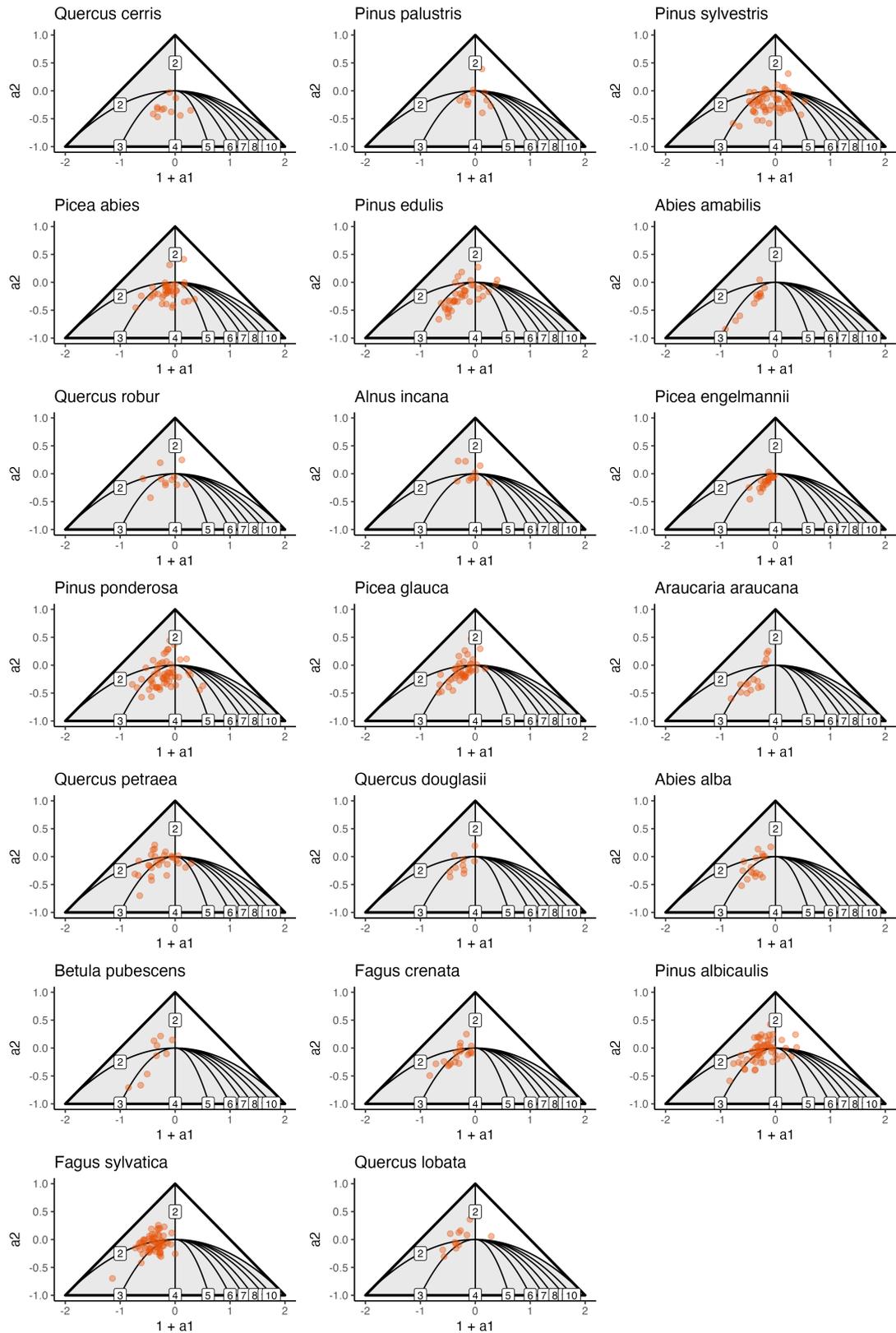


Figure S5: Temporal autoregression coefficients across studied species ($N = 20$). Each dot shows one population.

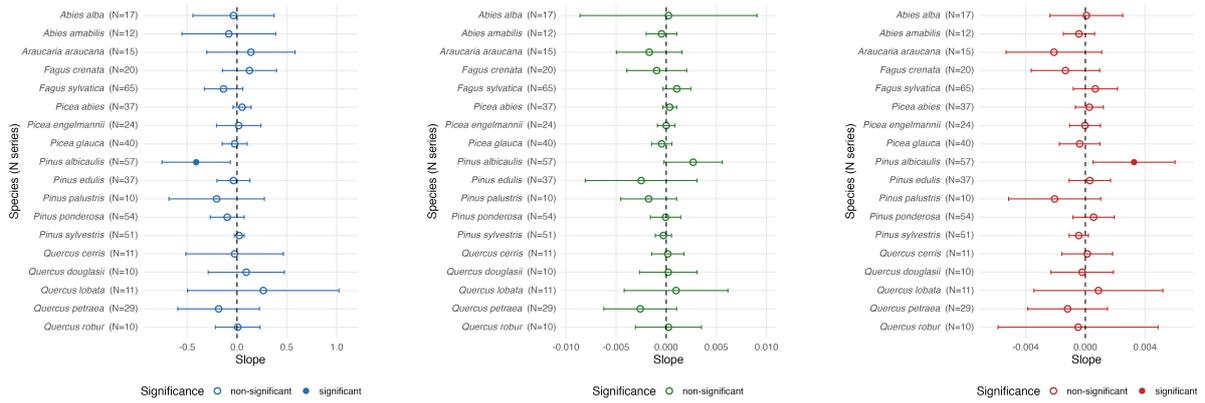


Figure S6: Relationship between climatic variables and masting period length from per-species truncated Poisson models. Each panel shows species-level slopes ($\pm 95\%$ confidence intervals) from truncated Poisson regressions fitted with period length as the response and (A) mean annual temperature (MAT), (B) mean annual precipitation (MAP), or (C) climatic water balance (CWB) as predictors. Positive slopes indicate a tendency toward longer masting periods with increasing climatic values, while negative slopes indicate shorter periods. Models were fitted separately for each species using the `glmTMB()` function from the `glmTMB` R package. Significant effects are shown as filled circles, while non-significant effects are shown as empty circles.

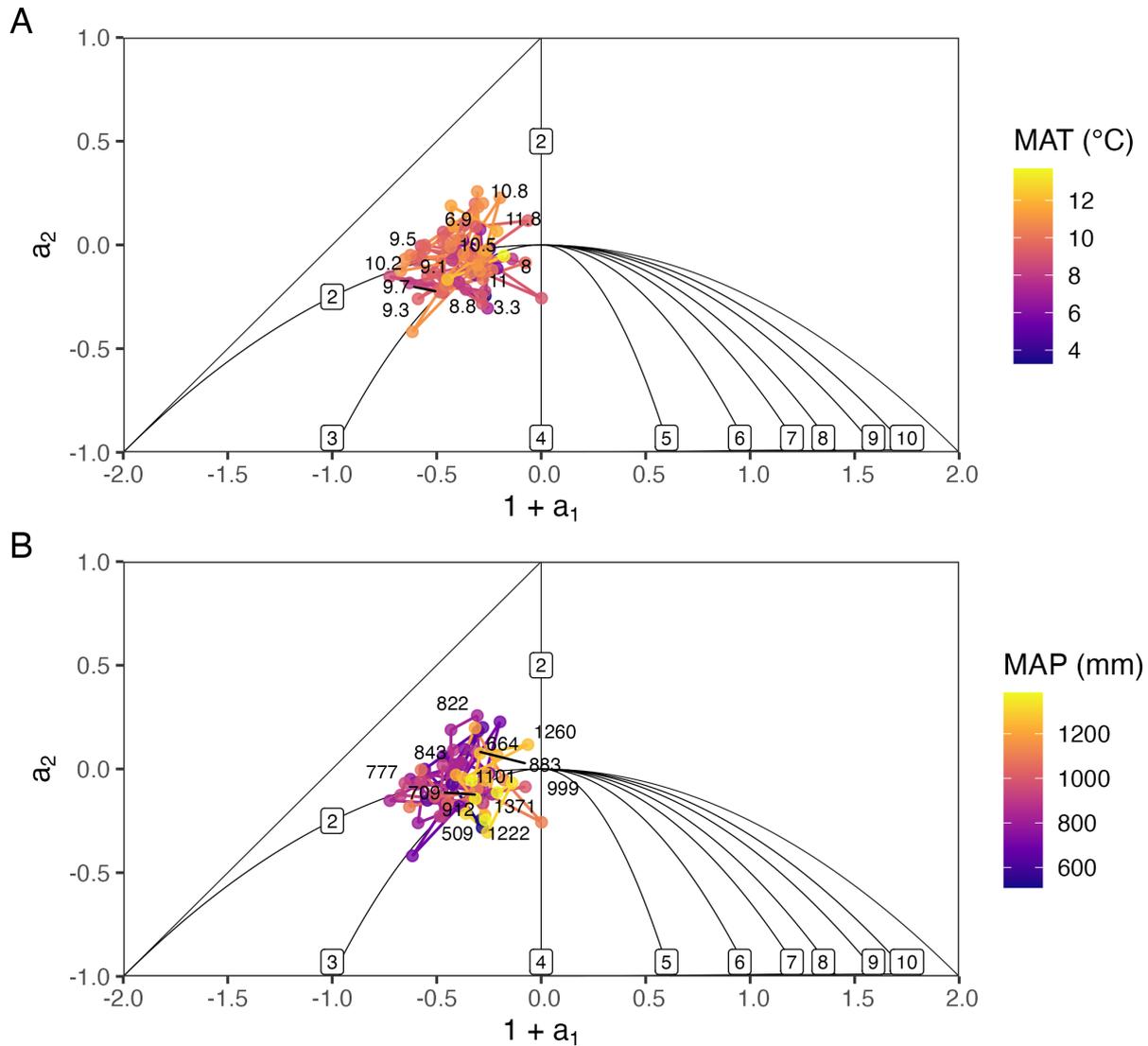


Figure S7: Climatic variation in fitted $1 + a_1$ and a_2 values for *Fagus sylvatica*. The panels show how fitted $1 + a_1$ and a_2 coefficients from the same climate-AR(2) models used in Fig. 4 vary along real climatic gradients for *Fagus sylvatica*. Populations are ordered from cold to warm in the MAT panel (A) and from dry to wet in the MAP panel (B), and plotted within the parabola plot according to their fitted coefficients. Consistent with Fig. 3, the MAT effect for *Fagus sylvatica* was non-significant, and fitted $1 + a_1$ and a_2 values show no systematic movement across the parabola. In contrast, MAP showed a significant positive effect, and fitted values display a clear directional shift across the parabola with increasing precipitation. This visualisation shows the same pattern as the whisker plot: no directional shift with MAT, and a gradual positive progression with MAP.

A

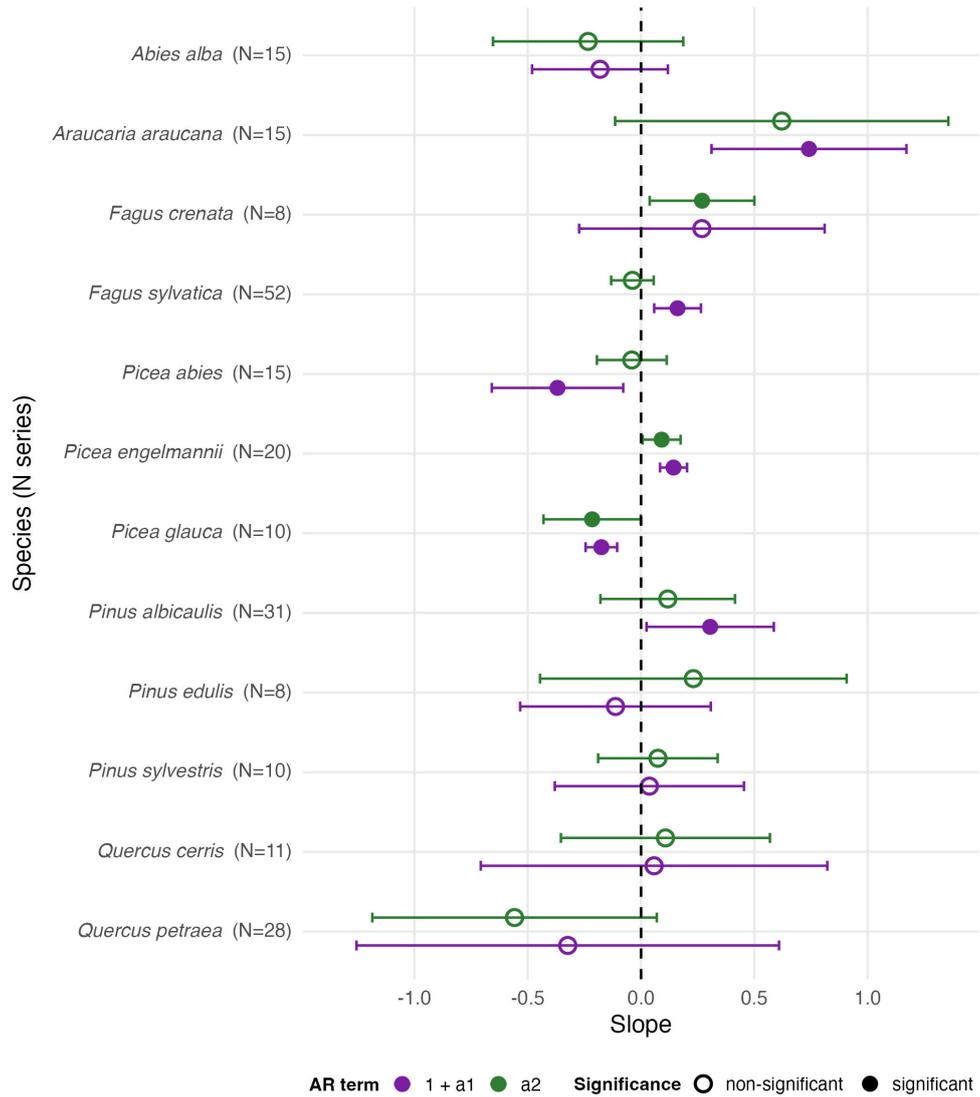


Figure S8: Relationship between temporal autoregression coefficients $1 + a_1$ and a_2 and elevation (N = 223 populations, 10 species). Each row represents a separate species. Each point represents the coefficient of the linear regression of the relationship, with error bars indicating the 95% confidence intervals. Significant effects are shown as filled circles, while non-significant effects are shown as empty circles. Positive slopes of $1 + a_1$ are associated with longer period length, whereas negative slopes are associated with shorter period length. Positive slopes of a_2 indicate weaker or less regular periodicity, while negative slopes indicate stronger or more regular periodicity.

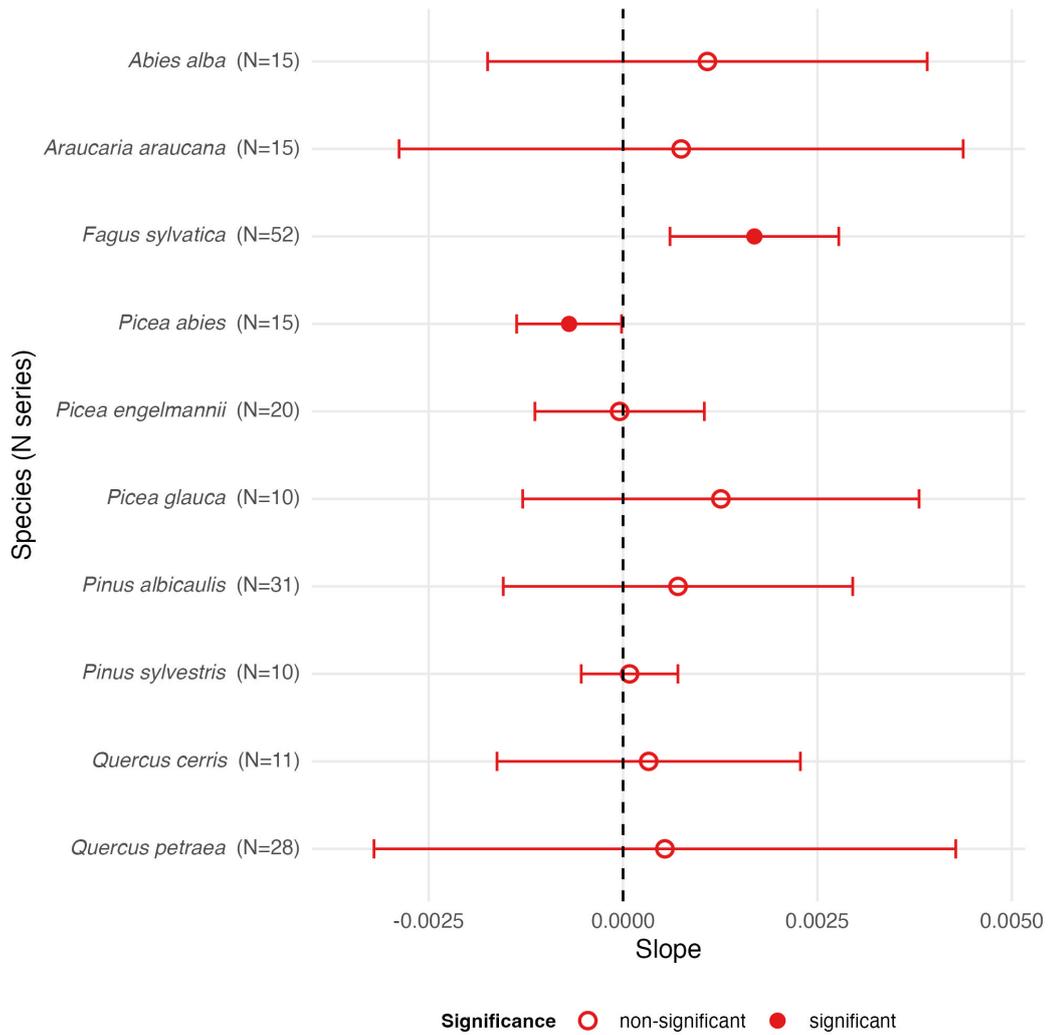


Figure S9: Relationship between masting period length and elevation derived from per-species truncated Poisson models. Each row represents a separate species. Points show model-estimated slopes with 95% confidence intervals. Positive values indicate a tendency toward longer masting periods with increasing elevation, while negative values indicate shorter periods. Significant effects are shown as filled circles, while non-significant effects are shown as empty circles.

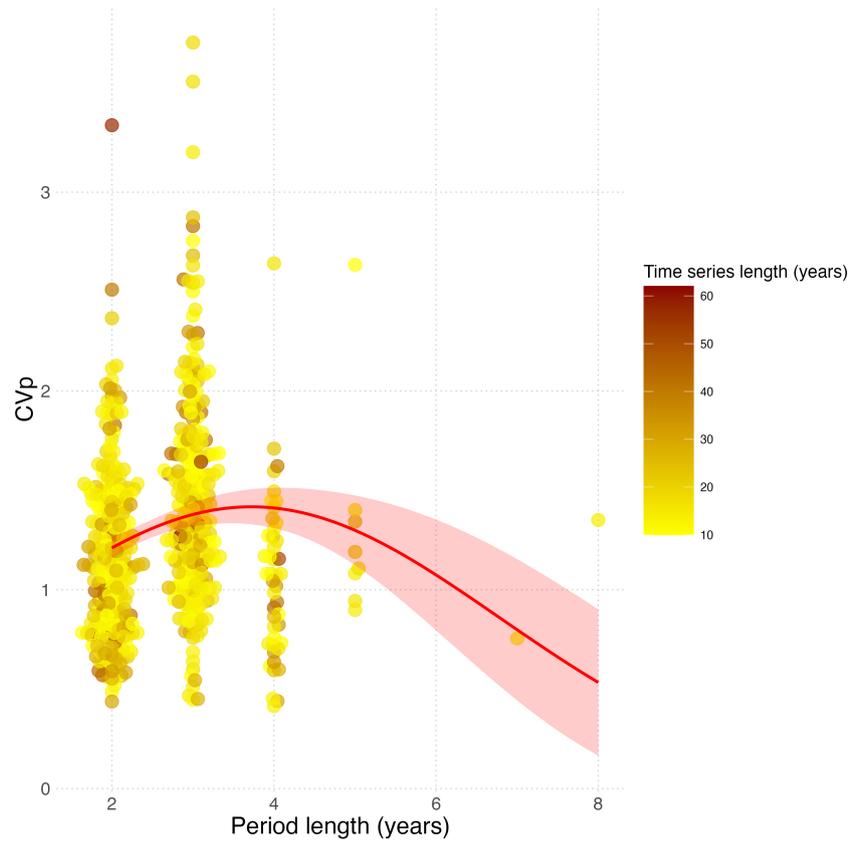


Figure S10: Relationship between period length and CVp (coefficient of variation for each population). Each point represents an individual time series, coloured by time series length. CVp is calculated as the standard deviation of seed production in a population divided by the mean seed production in that population. A Gamma generalized linear model (GLM) with a log link was fitted, including a quadratic term for period length and time series length as explanatory variables, with CVp as the response variable. Period length ($p < 0.001$) had a significant effect, while time series length was not significant ($p = 0.678$). The model had a marginal R^2 of 0.03, indicating that period length explained only a small proportion of the variation in CVp. The red line represents model predictions, with 95% confidence intervals.