

1 **estar: An R package to measure ecological stability**

2 Ludmilla Figueiredo<sup>1,2</sup>, Cédric Scherer<sup>1</sup>, Stephanie Kramer-Schadt<sup>1,3</sup>, Juliano Sarmiento Cabral<sup>2,4,5</sup>, Sonia  
3 Kéfi<sup>6,7</sup>, Paul J. Van den Brink<sup>8</sup>, Viktoriia Radchuk<sup>1</sup>

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5 <sup>1</sup>Department of Ecological Dynamics, Leibniz Institute for Zoo and Wildlife Research (IZW), Alfred-  
6 Kowalke-Straße 17, 10315 Berlin, Germany

7 <sup>2</sup>Ecosystem Modeling group, Center for Computational and Theoretical Biology (CCTB), University of  
8 Würzburg, Würzburg, Germany

9 <sup>3</sup>Institute of Ecology, Technische Universität Berlin, Rothenburgstrasse 12, 12165 Berlin, Germany

10 <sup>4</sup> School of Biosciences, College of Life and Environmental Sciences, University of Birmingham,  
11 Birmingham B15 2TT, United Kingdom

12 <sup>5</sup> Ecological Modelling group, Department of Plant Biodiversity, Bonner Institute for Organismal Biology,  
13 University of Bonn, 53115 Bonn, Germany

14 <sup>6</sup> ISEM, CNRS, Univ. Montpellier, IRD, EPHE, Montpellier, France

15 <sup>7</sup> Santa Fe Institute, 1399 Hyde Park Road, Santa Fe, NM 87501, USA

16 <sup>8</sup> Aquatic Ecology and Water Quality Management group, Wageningen University, PO Box 47, 6700 AA  
17 Wageningen, The Netherlands

18

19 Corresponding author: Ludmilla Figueiredo

20 Current address: German Centre for Integrative Biodiversity Research (iDiv) Halle-Jena-Leipzig,  
21 Puschstrasse 4, 04103 Leipzig, Germany

22 e-mail: ludmilla.figueiredo@idiv.de

23 **Abstract**

24 1. Assessing ecological stability across populations or communities is a prime goal in biodiversity  
25 monitoring and conservation research. Quantifying stability is not trivial because its different aspects can be  
26 measured with various metrics. However, to date, no software enables measuring different stability metrics  
27 on ecological time-series data.

28 2. We present the `estar` R package that standardises and facilitates the use of ten established stability  
29 properties that have been used to assess systems' responses to press or pulse disturbances at different  
30 ecological levels (e.g. population and community).

31 3. `estar` provides two sets of functions. The first set corresponds to functions that can be applied to  
32 univariate data, i.e., a time series of a system's state variable (e.g., individual body mass, population  
33 abundance, or species richness). The metrics included in this set are: invariability, resistance, extent and rate  
34 of recovery, and persistence. The second set of functions can be applied to multivariate data represented by  
35 the time series of the abundances of all species in a community. The functions in this set measure the  
36 stability of a community at short and long time scales. In the short term, community's response to a pulse  
37 (sudden) perturbation is measured by maximal amplification, reactivity and initial resilience (i.e. initial rate  
38 of return to equilibrium). In the long term, stability can be measured as asymptotic resilience and intrinsic  
39 stochastic invariability.

40 4. The package includes vignettes demonstrating the use of all functions and an introduction to the  
41 multivariate autoregressive state-space models necessary for the second set of functions. `estar` constitutes  
42 a toolbox with standardised, ready-to-use functions that bridge dichotomies in definitions and enable  
43 comparisons across state variables, taxa and scales.

44 **Introduction**

45 Measuring stability at different levels of ecological organisation, from individuals to populations and  
 46 communities, is of utmost importance for biodiversity monitoring and conservation because of widespread  
 47 human-driven perturbations such as climate change, pollution, species invasions, and their multiple effects  
 48 on biodiversity. Consensus emerges that ecological stability can be broadly defined as the ‘overall ability of  
 49 a system (...) to retain its function and structure in the face of perturbations’ (Noy-Meir, 1973 *apud* Van  
 50 Meerbeek et al., 2021), but since its introduction in Ecology in the 1950s, stability has been shown to be a  
 51 multidimensional concept (Donohue et al., 2013; Grimm & Wissel, 1997; Pimm, 1984; Van Meerbeek et al.,  
 52 2021). This multidimensional character of stability manifests itself in at least four ways. First, different  
 53 stability properties are not easily comparable because they are uncorrelated and capture different aspects of  
 54 the system’s response trajectory at different temporal scales (Donohue et al., 2016; Kéfi et al., 2019).  
 55 Secondly, the same stability property can be measured at different organisational levels (Hillebrand &  
 56 Kunze, 2020) or for different → *state variables* (e.g. total community biomass or Shannon index). Thirdly,  
 57 different metrics have been proposed to calculate a single stability property. For example, resistance is often  
 58 measured immediately after a perturbation but can also be measured when the difference between the state  
 59 variable of the disturbed system and its pre-disturbed state is the largest (also known as “maximum  
 60 attenuation”, Capdevila et al., 2020). Further, different stability properties are not necessarily correlated  
 61 (Arnoldi et al., 2018; Domínguez-García et al., 2019; Downing et al., 2020; Neubert & Caswell, 1997;  
 62 Radchuk et al., 2019), making it necessary to estimate several of them when studying a system’s stability.  
 63 These four issues hamper synthesis across ecological stability studies. Beyond the necessity of quantifying  
 64 multiple properties of stability, there is one practical concern: no software permits quantifying stability in the  
 65 diverse ways it has been measured so far. Therefore, the time is ripe for a tool to measure different stability  
 66 properties (or dimensions) in a standardised, comparable and reproducible way.

67 In addition to the multifaceted nature of the stability concept, there is a large divide between the metrics used  
 68 by empiricists and theoreticians (Donohue et al., 2016). While empirical studies often measure the temporal  
 69 invariability of population and community state variables, theoretical studies mainly quantify asymptotic

70 stability properties derived from species interaction matrices ( $\rightarrow$  *species interaction matrix*, Donohue et al.,  
 71 2016), which are rarely available for observational studies. Assessing interaction strength in empirical  
 72 systems is a rather laborious task and can be accomplished with various methods (e.g. controlled pairwise  
 73 species experiments; Carrara et al., 2015). However, estimations from different methods are usually not  
 74 comparable (Carrara et al., 2015b). In that context, the first-order Multivariate Autoregressive Models  
 75 (MARs) are a flexible method, as they can be applied to time-series data on community composition to  
 76 derive species interaction matrices, which can be used to derive asymptotic stability properties (Downing et  
 77 al., 2020; Ives et al., 2003). Still, the use of MARs outside of freshwater plankton community studies  
 78 remains limited (Hampton et al. 2013). Therefore, a user-friendly tool that integrates MARs and allows the  
 79 user to derive asymptotic stability properties has the potential to close the gap between empirical and  
 80 theoretical stability research.

81 Here, we present **estar**, an R package designed to facilitate the use of ten stability properties. We present  
 82 the properties in two groups, distinguished by the format of the input data: i) if the input is in the form of a  
 83 time series of a single state variable measured at any organisational level we talk about **univariate**  
 84 **properties** (invariability, resistance, extent of recovery, rate of recovery, and persistence; Table 1), and ii) if  
 85 the input is a matrix of species abundances over time or a species interaction matrix, i.e. multivariate data,  
 86 we talk about **multivariate properties** (maximal amplification, initial resilience, asymptotic resilience,  
 87 intrinsic stochastic invariability, and intrinsic deterministic invariability; Table 2). We showcase the use of  
 88 the package to measure the stability of a real-world freshwater invertebrate community perturbed by an  
 89 insecticide (van den Brink et al., 1996).

## 90 **Package overview**

91 Although ecological stability attracts much research interest, we still lack R and Python software to readily  
 92 calculate the variety of stability metrics available in the literature. At the community level, the *codyn*  
 93 package (Hallett et al., 2020) provides four functions to measure stability in terms of species variances and  
 94 covariance. The package MAR1 (Scheef & Holmes, 2023) offered functionality to both fit MARs and derive

95 asymptotic stability properties from the output of those models, but it was taken down from R's centralized  
 96 repository (CRAN) in 2019. The package MARSS, which can be used to estimate  $\rightarrow$  *species interaction*  
 97 *matrices*, does not provide formulas to calculate the stability properties (*sensu* Ives et al., 2003). Moreover,  
 98 while MARSS has been used in aquatic communities (Hampton et al., 2013; Ruhí et al., 2015; Tolimieri et  
 99 al., 2017), it remains largely unused in terrestrial systems. At the population level, the R package *popdemo*  
 100 (Stott et al., 2021) allows performing transient analyses and thus calculating demographic resilience, a  
 101 concept that was recently introduced to population ecology from community ecology (Capdevila et al.,  
 102 2020). Further, although many stability properties (e.g. resistance and recovery rate) apply to a single time  
 103 series (univariate data) and seem conceptually simple, no single software allows the quantification of all of  
 104 them simultaneously. Thus, a tool is needed to derive asymptotic stability properties at the community level  
 105 and stability properties from univariate time-series data (applicable to any level of organization). To address  
 106 this, we provide the **estar** package. It has been submitted to CRAN and can currently be downloaded from  
 107 <https://anonymous.4open.science/r/estar-251E> and installed from source: `devtools::install.packages("estar",`  
 108 `repos = NULL, type = "source")`.

109

## 110 **Univariate stability properties**

111 Univariate stability properties are calculated from the time series of a state variable measured in a system  
 112 disturbed by a pulse disturbance (or white noise, in the case of invariability; Fig. 1-2, Table 1). These  
 113 univariate properties can be calculated using a variable measured at any organisation level, for example,  
 114 individual stress hormone levels, species richness, and Shannon-Weaver index.  $\rightarrow$  *Baseline* is a central  
 115 concept for quantifying multiple stability properties. The stability values depend on whether the state  
 116 variable of the disturbed system is normalised relative to the baseline (Ingrisch & Bahn, 2018). Our functions  
 117 give the user flexibility to specify the baseline and whether the state variable should be normalised by the  
 118 baseline (e.g. by using the log ratio between the values measured on the disturbed and undisturbed systems).

119 We implemented functions to calculate five univariate stability properties. *Invariability* ( $I$ ,  
 120 `invariability()`, Fig. 2a), the most common stability measure (Donohue et al., 2016), measures the  
 121 system's response to white noise. *Resistance* ( $R$ , `resistance()`, Fig. 2b) is the magnitude of change in  
 122 the state variable following a disturbance (Pimm, 1984). It can be calculated as the maximum magnitude of  
 123 change in the state variable or the magnitude at a user-defined time step after the disturbance (e.g. at the first  
 124 time step to capture the system's initial resistance). *Recovery* is the return of the system's state variable to the  
 125 baseline state (Medeiros et al., 2021; Van Meerbeek et al., 2021). `estar` contains two properties for it: the  
 126 *extent of recovery*, i.e. how close the system returns to the baseline ( $Er$ , `recovery_extent()`, Fig. 2c)  
 127 and the *rate of recovery*, i.e. how fast it returns to the baseline ( $Rr$ , `recovery_rate()`, Fig. 2d). Finally,  
 128 *persistence* ( $P$ , `persistence()`, Fig. 2e) is the proportion of time a variable stays within one standard  
 129 deviation from the baseline's mean during the user-defined period (Pimm, 1984). Descriptions of the variants  
 130 of each metric and an evaluation of the functions' performance can be found in the vignette "Univariate  
 131 metrics".

### 132 **Multivariate stability properties**

133 The multivariate properties measure the community's responses to a perturbation, both in the short-term and  
 134 in the long-term. The long-term (asymptotic) rates are usually estimated for theoretical systems (Arnoldi et  
 135 al., 2018; Neubert & Caswell, 1997), while empirical studies mainly assess the short-term (transient) rates  
 136 (Arnoldi et al., 2018). The long-term data are rarely available for empirical systems, limiting the possibility  
 137 of deriving dominant eigenvalues from  $\rightarrow$  *Jacobian matrices* to calculate the asymptotic resilience of  
 138 theoretical systems (Table 2).

139 In contrast to the single time series of the state variable that is required as data input for univariate stability  
 140 properties, multivariate properties are calculated from species interaction matrices (denoted as  $\rightarrow$   $\mathbf{B}$ ). Our  
 141 package requires the estimation of  $\mathbf{B}$  from the user-supplied time series data on species abundances in the  
 142 community (termed "multivariate data"). We provide the function (`extractB()`) to format the output of  
 143 the multivariate autoregressive state space (MARSS) model fitted by the R package MARSS (Holmes et al.,

144 2012; Holmes, Scheuerell, et al., 2024; Holmes, Ward, et al., 2024). MARSS models estimate  $\mathbf{B}$  while  
 145 accounting for uncertainty in the observation process that generated the data. We exemplify how MARSS  
 146 models work and comment on possible pitfalls and the functions' performance in the "MARSS in *estar*"  
 147 vignette.

148 We provide three functions to characterise a community's transient response to a pulse disturbance (Fig. 3).  
 149 Two of these functions calculate the metrics describing the "amplification envelope" — the curve describing  
 150 the upper bound response to a perturbation (Neubert & Caswell, 1997). The amplification envelope can be  
 151 interpreted as an estimation of initial instability (Arnoldi et al., 2016) and is characterised by measures of  
 152 *reactivity* ( $R_a$ , `reactivity()`, Table 2) and *maximal amplification* ( $A_{\max}$ , `max_amp()`).  
 153 Complementing the amplification envelope, the *initial resilience* ( $R_0$ , `init_resil()`) characterises the  
 154 system's initial rate of return to equilibrium (Table 2, Arnoldi et al., 2018).

155 To measure stability in the long term, we provide a function to calculate *asymptotic resilience* ( $R_{\infty}$ ,  
 156 `asympt_resil()`, Table 2, Fig. 3), the long-term rate of return to equilibrium, complementary to  $R_0$   
 157 (Arnoldi et al., 2018). Further, to measure instability under press disturbance or white noise, `estar`  
 158 provides the *intrinsic stochastic invariability* ( $I_{\infty}$ , `stoch_var()`), a theoretical equivalent of the univariate  
 159 measure of invariability (Arnoldi et al., 2016), albeit uncorrelated to it (Downing et al., 2020).

#### 160 **Example: Stability of an aquatic macroinvertebrate community**

161 To exemplify the use of `estar`, we applied it to data from a species-rich community of freshwater  
 162 macroinvertebrates that was disturbed by the insecticide chlorpyrifos in a previously conducted eco-  
 163 toxicological experiment (van den Brink et al., 1996). We analysed the effect of two concentrations of the  
 164 insecticide on the abundance of three functional groups of aquatic macroinvertebrates (herbivores,  
 165 carnivores, and detritivores; Supplementary information S3). According to the univariate properties,  
 166 detritivores recovered better than carnivores and herbivores under the lower insecticide concentration, as  
 167 quantified by the extent of recovery and rate of recovery (at week 28, Fig. 4-a, b). Under the higher

168 insecticide concentration, carnivores had the highest rate of recovery because by week 28 their abundance  
169 increased to higher values in relation to their pre-disturbance state, whereas detritivores stayed at lower  
170 abundances. Finally, resistance (calculated as the magnitude of abundance change in the first week) was  
171 particularly low for the detritivores, demonstrating that resistance and recovery correlate negatively in this  
172 system.

173 Regarding the multivariate stability properties, the higher reactivity and maximal amplification values at  
174 high insecticide concentration reflect the strong decreases in herbivore and detritivore abundances (Table 3).  
175 Similarly, both the initial and asymptotic resilience were lower at a higher concentration of chlorpyrifos,  
176 indicating a slower rate of return to equilibrium compared to the communities subjected to a lower  
177 concentration. Nonetheless, we obtained much higher stochastic invariability for the community under  
178 higher insecticide concentration.

## 179 **Conclusion**

180 The `estar` package provides functions for calculating different stability properties, grouped in two sets:  
181 those applicable to time series 1) of single state variables and 2) of community compositional data. With the  
182 functions applied to time series of single state variables, we offer a flexible tool that quantifies stability at  
183 different levels of organisation, from individual to community. With the functions applied to community  
184 compositional data, we offer a tool for easy computing of the stability properties that were hitherto mainly  
185 used by theoreticians. Thus, `estar` closes the gap between empiricists and theoreticians in stability  
186 research. Finally, this package will facilitate cross-system comparisons of stability and further our  
187 understanding of how stable systems are across organisation levels, spatiotemporal scales and environmental  
188 conditions.

189

190 **Glossary**

191 *Baseline*: → *state variable* value(s) the user defines to represent an undisturbed system. It can be measured  
 192 before a disturbance affects the focal system, from a separate undisturbed system, or from a single point  
 193 considered to be representative of an undisturbed system (Fig. Error: Reference source not found-b).

194 *Jacobian matrix*: a functional matrix or derivative matrix of a differentiable function of all first partial  
 195 derivatives in the case of total differentiability. The Jacobian is used, for example, to approximate  
 196 multidimensional functions in mathematics or, in the community ecology context, for functions of interaction  
 197 strength. → *Species interaction matrix*.

198 *Press disturbances*: a press disturbance affects a system permanently and continuously (Ryo et al., 2019),  
 199 e.g. climate change or ongoing pollution from leakage.

200 *Pulse disturbance*: an event that suddenly affects a system and recedes quickly after reaching a peak, e.g.  
 201 fires or storms (Ryo et al., 2019).

202 *Species interaction matrix* (**B**): a matrix quantifying the strength of density dependence between the  $n$   
 203 species in a community, i.e. the effect of one species' density on the per capita growth rate of another (or of  
 204 its own, Hampton et al., 2013). It is often referred to as “community matrix” (or “a discrete-time version of a  
 205 → *Jacobian matrix*”; Downing et al., 2020).

206 *State variable*: a variable describing a system, e.g. population abundance (whereby the population is a  
 207 system) or species richness (whereby the community is a system).

208

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300 Table 1: Definitions of the univariate stability metrics and their variants, along with the options of baseline  
301 (*b*) used, and the options of response from which the metric is calculated.  $v_b$  refers to an independent  
302 baseline;  $v_p$ , to a baseline defined by pre-disturbance values; and  $v_d$ , to the state variable values in the  
303 disturbed time series;  $t$  stands for time. Notes: <sup>a</sup> usually the first time step following disturbance, <sup>b</sup>  
304 summarised as the mean or median of values of a user-defined time period, <sup>c</sup> usually the last time step of the  
305 time series, <sup>d</sup> time steps of disturbed systems and baseline match. User-defined values do not have a default  
306 value in the function.

Metric, function name	Formal definition (several definitions are possible for the same metric)	Baseline	Response	Notation of metric
Invariability <i>(I, invariability())</i> , also referred to as “temporal stability”, Hillebrand et al., 2018)	The inverse of the standard deviation of residuals of the linear model where the response is predicted by time.	$v_b, v_p$	Log-ratio $l = \log(v_d/v_b)$	$\frac{1}{\sigma(\varepsilon)}$ , where $\sigma(\varepsilon)$ is the standard deviation of $e$ , the residuals of the linear model $l = \beta t + \varepsilon$ , with $t$ referring to the time, $\beta$ to the regression coefficients of the model, and $\varepsilon$ to its errors.
		none	$v_d$	Same as above, but the response in the linear model is the state variable, $v_d = \beta + \varepsilon$
Resistance <i>(R, resistance())</i>	The inverse of the coefficient of variation of response.	$v_b$	Log-ratio, $l$	$\frac{1}{CV(l)}$
		none	$v_d$	$\frac{1}{CV(v_d)}$
		$v_b$	Log-ratio	$\log(v_d/v_b)$
	Log response ratio between the state variable’s value in the disturbed and in the baseline systems, on the user-defined time step <sup>a</sup> .	$v_p^b$	Log-ratio	$\log(v_d/v_p)$
		$v_b$	Log-ratio	$\max(\log(v_d/v_b))$

	in the disturbed and the baseline systems, over user-defined time interval.	$v_p$	Log-ratio	$\max(\log(v_d/v_p))$
	Absolute difference between the state variable's value in the disturbed and in the baseline systems, on the user-defined time step <sup>a</sup> .	$v_b$	Difference	$ v_d - v_b $
		$v_p$	Difference	$ v_d - v_p $
	Maximal difference between state variable's value in the disturbed and in the baseline systems, over user-defined time interval.	$v_b$	Difference	$\max( v_d - v_b )$
		$v_p$	Difference	$\max( v_d - v_p )$
Extent of recovery ( $E_r$ , <code>recovery_extent()</code> )	Log-response ratio between the state variable in the disturbed system and the baseline taken on the user-defined time step $t_{post}$ when the recovery is assumed to have taken place <sup>c</sup> .	$v_b$	Log-ratio	$\log(v_d/v_b)$
		$v_p$	Log-ratio	$\log(v_d/v_p)$
	Difference between the state variable in the disturbed system and the baseline, taken on the user-defined time step $t_{post}$ when the recovery is assumed to have taken place <sup>c</sup> .	$v_b$	Difference	$v_d - v_b$
		$v_p$	Difference	$v_d - v_p$
Rate of recovery ( $R_r$ , <code>recovery_rate()</code> - also called "engineering resilience" by Pimm (1984))	The slope of a linear model, where the response is predicted by time.	$v_b$	Log-ratio, $l$	$l = R_r.t + b$
		none	$v_d$	$v_d = R_r.t + b$

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<p>Persistence (<i>P</i>, persistence())</p>	<p>The proportion of the user-defined time frame <math>t_a</math> during which the response stayed within the limits of an interval determined by the baseline mean <math>\pm</math> sd (<math>t_p</math>).</p>	<p><math>v_b</math></p>	<p><math>v_d</math></p>	<p><math>t_p/t_a</math></p>
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309 Table 2: Definitions of the multivariate stability metrics, based on species interactions matrix (**B**).

Metric, function name	Formal definition and interpretation	Equation
Reactivity ( $R_a$ , <code>reactivity()</code> )	Maximum initial amplification rate of a perturbation (Neubert & Caswell, 1997).	$\lambda_{dom}(H(B))$ , where $\lambda_{dom}$ is the dominant eigenvalue, and $H$ is the Rayleigh quotient.
Maximal amplification ( $A_{max}$ , <code>max_amp()</code> )	The factor by which the perturbation that grows the largest is amplified, calculated as the Euclidian norm of the species interaction matrix (Neubert & Caswell, 1997).	$A_{max} = \max_{t \geq 0} (\max_{x_0 \neq 0} \frac{\ e^{B^T} x_0\ }{\ x_0\ })$ , where $\ e^{B^T} \cdot\ $ is the matrix norm of <b>B</b> and $x_0$ is the vector of initial abundances (Domínguez-García et al., 2019).
Initial resilience ( $R_0$ , <code>init_resil()</code> )	Initial resilience is calculated as the initial rate of return to equilibrium (Downing et al., 2020). The larger its value, the more stable the system, as its “worst case” initial rate of return to equilibrium is faster (Downing et al., 2020).	$R_0 = -\log(\sqrt{\lambda_{dom}(B^T B)})$
Asymptotic resilience ( $R_\infty$ , <code>asyp_resil()</code> )	The slowest/long-term asymptotic rate of return to equilibrium after a pulse perturbation (Arnoldi et al., 2016; Downing et al., 2020). $R_\infty$ is a positive real number. The larger its value, the more stable the system, as its rate of return to equilibrium is faster (Downing et al., 2020).	$R_\infty = -\log( \lambda_{dom}(B) )$

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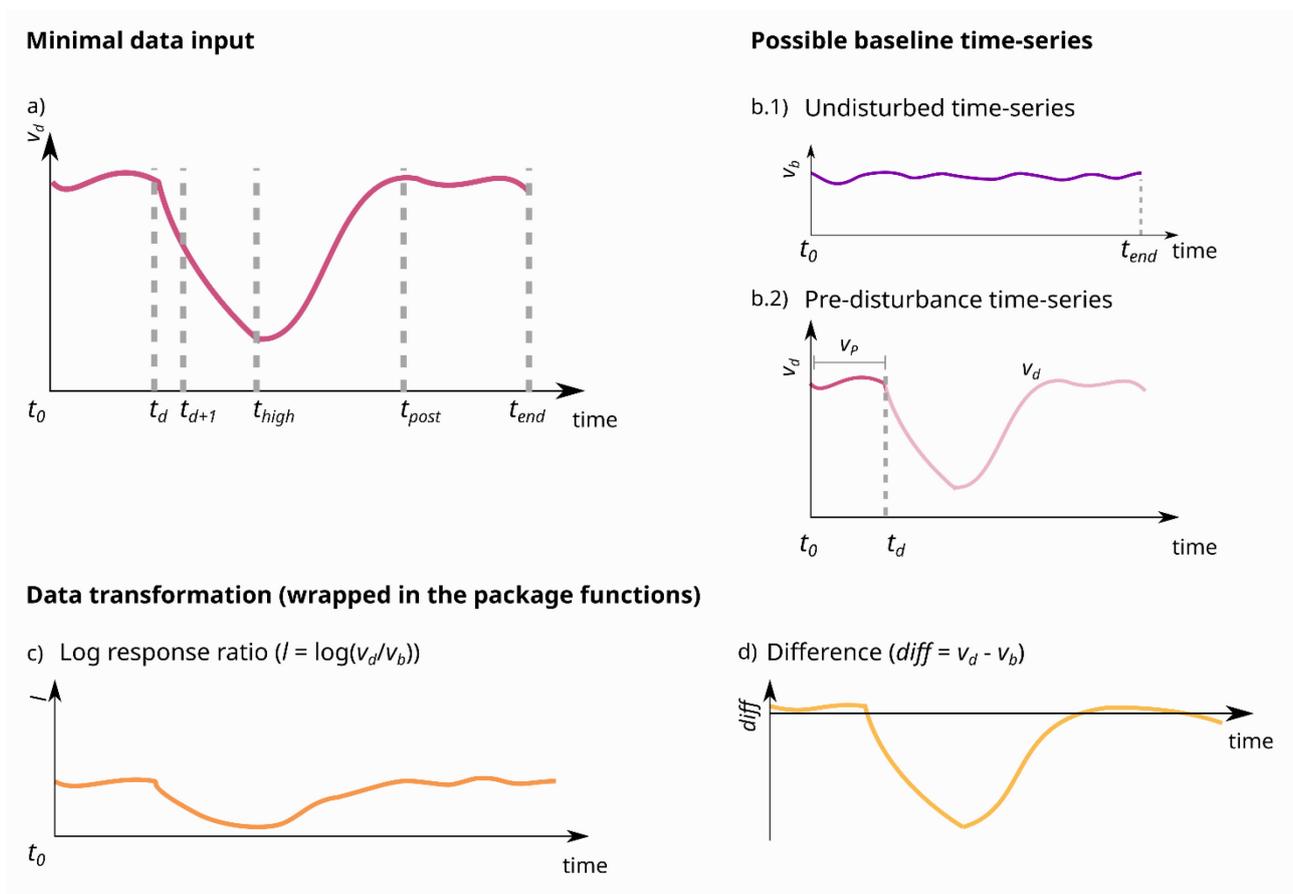
Intrinsic stochastic invariability  $(I_S, \text{stoch\_var}())$  Inverse of the maximal response variance to white noise. The larger its value, the more stable the system, as its rate of return to equilibrium is higher.

$I_S = \frac{1}{\|\hat{B}^{-1}\|}$  where  $\|\hat{B}^{-1}\|$  where the spectral norm of the inverse of matrix  $\|\hat{B}^{-1}\| = B \otimes I + I \otimes B$ , with  $I$  being an identity matrix.

311 Table 3: Multivariate metrics calculated by **estar** for the aquatic macroinvertebrate community subjected to two concentrations of the chlorpyrifos  
 312 insecticide.

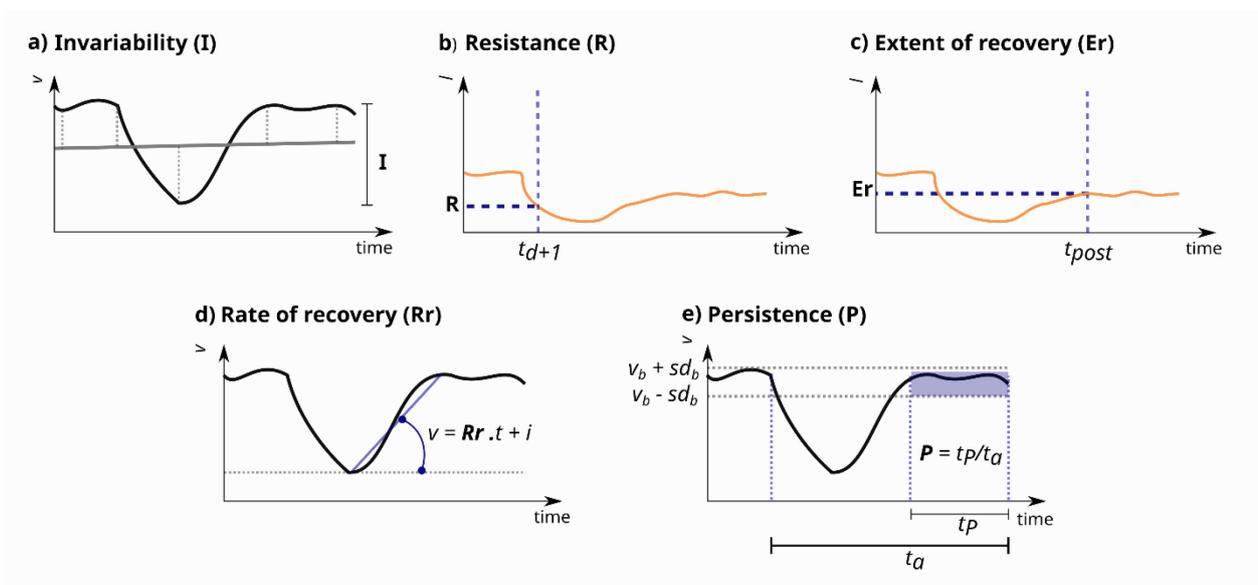
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	<b>Insecticide concentration</b>	
	0.9 µg/L	6 µg/L
<b>Multivariate metrics</b>		
<i>Reactivity</i>	0.593	0.972
<i>Maximal amplification</i>	2.308	3.14
<i>Stochastic invariability</i>	7.510	119.061
<i>Initial resilience</i>	0.350	0.275
<i>Asymptotic resilience</i>	0.651	0.173



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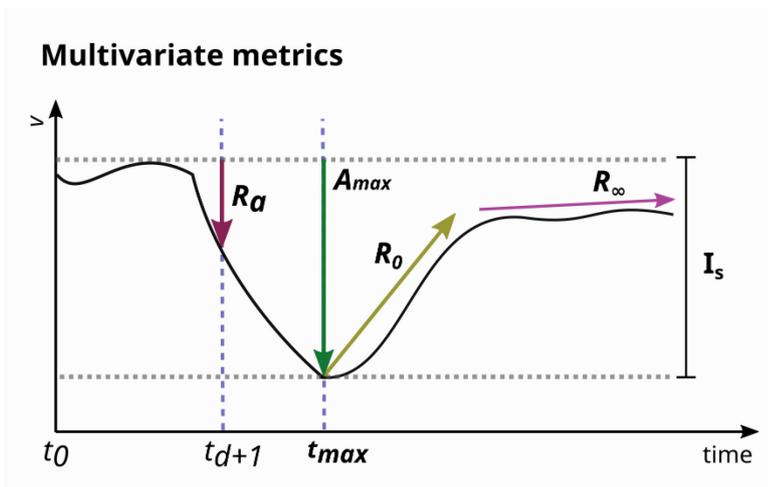
315 Figure 1: Schematic representation of possible univariate inputs (a and b) and the two types of  
 316 transformation applied for some metrics (c and d). Relevant time steps:  $t_d$  time when pulse disturbance is  
 317 applied to the system,  $t_{d+1}$  first time step after disturbance,  $t_{high}$  time step where the absolute distance  
 318 between state variable value in the disturbed system and baseline is the highest,  $t_{post}$  time step where  
 319 recovery is considered to have happened,  $t_{end}$  end of time series. a) The data for which a metric is to be  
 320 calculated must constitute a time series of a state variable in the disturbed system ( $v_d$ ). b) Most functions  
 321 require a time series of the same state variable in the baseline system. This baseline can be taken from a  
 322 separate undisturbed system ( $v_b$ , b.1) or from the pre-disturbance values of the disturbed system ( $v_p$ , b.2).  
 323 It can be a single value or a summary of the time series. The stability metrics that require both the  
 324 disturbed and baseline time series are applied to either the log response ratio ( $l$ ) between the values in  
 325 these two time series (c) or the difference between them (d).



326

327 Figure 2: Schematic representation of how some of the variants of univariate properties are calculated with  
 328 estar: a) invariability ( $I$ ) as the coefficient of variation of the state variable in the disturbed state; b)  
 329 resistance ( $R$ ) as the log-ratio ( $l$ ) between the disturbed and baseline time series in the first time step after  
 330 disturbance ( $t_{d+1}$ ); c) extent of recovery ( $Er$ ) as log-ratio between the disturbed and baseline time series at  
 331 the user-defined time step the system is expected to have recovered ( $t_{post}$ ); d) rate of recovery ( $Rr$ ) as the  
 332 slope of the linear model fitted to the disturbed time series; and e) persistence ( $P$ ) calculated as the  
 333 proportion of time during which the disturbed time series stayed in the interval defined as  $\pm 1$  standard  
 334 deviation from the baseline ( $v_b \pm sd_b$ ) over the total user-specified time period ( $t_a$ ). The metrics are  
 335 calculated from a time series of a state variable ( $v$ ; black line) and the log response ratio ( $l$ ; orange line). All  
 336 variants are demonstrated in the “Univariate properties” vignette.

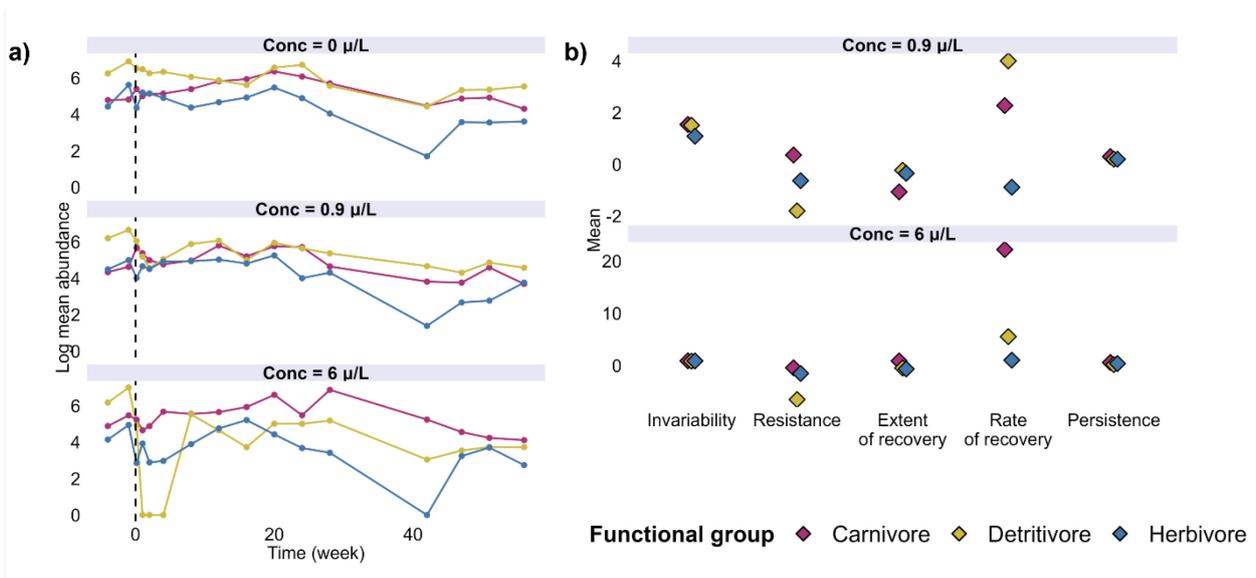
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339 Figure 3: Schematic representation of the multivariate properties: reactivity ( $R_a$ ), calculated at the first time  
 340 step following disturbance ( $t_{d+1}$ ); maximal amplification ( $A_{max}$ ); initial resilience ( $R_0$ ); asymptotic resilience  
 341 ( $R_\infty$ ); and intrinsic stochastic invariability ( $I_s$ ). To facilitate comprehension, we illustrate metrics in relation  
 342 to a state variable ( $v$ ) disturbed at time  $t_d$ , but the metrics, in fact, are calculated from the community's  
 343 species interactions matrix and not a single state variable. All properties are demonstrated in the "MARSS  
 344 models" vignette.

345



346

347 Figure 4: Example of estar application to measure the stability of three macroinvertebrate functional  
 348 groups to two concentrations of the chlorpyrifos insecticide. a) Log of mean abundance over the 60 weeks  
 349 of the experiment. Insecticide was applied at week 0 (vertical dashed line). b) Univariate metrics calculated  
 350 by estar under two concentrations of the insecticide (0.9µg/L and 6µg/L).