

1 Integrating Phylogeny and Functional Traits into Evaluations of Nestedness in  
2 Island Floras

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4 Running title: Nestedness in Island Floras

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6 \*Fabio Mologni<sup>1</sup>

7 Peter J. Bellingham<sup>2</sup>

8 Ewen K. Cameron<sup>3</sup>

9 Anthony E. Wright<sup>4</sup>

10 Yanping Wang<sup>5</sup>

11

12

13 1. University of Vienna, Vienna, Austria

14 2. Manaaki Whenua—Landcare Research, Lincoln, Aotearoa New Zealand

15 3. Auckland War Memorial Museum, Auckland, Aotearoa New Zealand

16 4. Canterbury Museum, Christchurch, Aotearoa New Zealand

17 5. Nanjing Normal University, Nanjing, China

18

19 \* Corresponding author

20 Postal address: Department of Botany and Biodiversity Research, University of Vienna, Rennweg 14,  
21 1030, Vienna, Austria

22 Phone: +43 670 1812022

23 Email: fabio.mologni@univie.ac.at, fabio.mologni@gmail.com

24 ORCID: <https://orcid.org/0000-0003-4750-9974>

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## 28 **Abstract**

### 29 **Questions**

30 Understanding the composition and structure of island floras is crucial for making informed  
31 conservation decisions. Island floras are often nested, i.e. lower species richness assemblages  
32 are frequently subsets of those higher in richness. However, the circumstances under which this  
33 occurs on islands are often unclear. Moreover, research in island biogeography rarely integrates  
34 phylogeny and functional traits in nestedness studies. Here, we integrated phylogenetic and  
35 functional traits in the investigation of relationships between taxonomic nestedness and island  
36 characteristics to gain insight into the processes that shape the composition and structure of  
37 island floras. We asked whether (i) native plant species assemblages are nested according to  
38 their phylogeny and functional traits and (ii) whether nested patterns are related to island  
39 characteristics.

### 40 **Location**

41 264 islands offshore from northern Aotearoa New Zealand.

### 42 **Methods**

43 We combined field surveys and published data for 775 native plant species across the islands.  
44 We compiled information about their phylogeny at the fine (species) and coarse level (i.e. ferns  
45 and allies, conifers, monocots, and dicots). Then, for each species, we determined three plant  
46 functional traits (growth forms, dispersal modes, and species maximum height). We quantified  
47 nestedness by organizing species incidence matrices using the NODF (i.e. nestedness metric  
48 based on overlap and decreasing fill) and fixed-fixed null models. Finally, we correlated island  
49 nestedness ranks with three island characteristics (area, isolation, and exposure to ocean-borne  
50 disturbances).

### 51 **Results**

52 All plant categories were nested, except for ferns and water-dispersed species. Nestedness  
53 ranks were consistently related to island area, strongly supporting the selective extinction  
54 hypothesis. Relationships with isolation and exposure to ocean-borne disturbances were less  
55 strong and varied by phylogeny and functional traits, suggesting weaker support for selective  
56 immigration and habitat nestedness.

## 57 **Conclusions**

58 These findings highlight the importance of nestedness studies and the integration of phylogeny  
59 and functional traits for identifying conservation priority areas. The overarching effect of island  
60 area in shaping insular plant composition underscores the key role of large islands for  
61 conserving plant diversity.

62

63 *Keywords:* Aotearoa New Zealand, Island biogeography, Island Conservation, Island flora,  
64 Native species, Nestedness, NODF, Plants, Species composition, Taxonomic Nestedness

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66

## 67 **Introduction**

68 Islands harbour a disproportionately high number of endemic plant species, many of which are  
69 under significant threat (Caujapé-Castells et al. 2010; Fernández-Palacios et al. 2021; Schrader  
70 et al. 2024). Thus, understanding the structure and composition of island plant communities is  
71 essential for making informed conservation decisions. Island floras frequently exhibit a nested  
72 structure, where species-poor assemblages are subsets of richer ones (Hu, Feeley, Wu, Xu,  
73 and Yu 2011; Morrison 2013; Patterson and Atmar 1986; Traveset, Kueffer, and Daehler 2014).  
74 However, the circumstances under which this occurs and the implications for island  
75 conservation remain poorly understood.

76 Nested patterns can be generated by three main ecological processes. First, extinction  
77 rates on small islands are often higher than on larger islands (MacArthur and Wilson 1967),  
78 especially for species with larger minimum area requirements (Higgins, Willig, and Strauss  
79 2006; Tjørve and Turner 2009) or with smaller population sizes (Matthews, Cottee-Jones, and  
80 Whittaker 2015). If extinctions are non-random, then communities on small, species-poor  
81 islands become subsets of larger, species-rich islands, a process known as selective extinction  
82 (Atmar and Patterson 1993; Chen, Zhan, and Wang 2022; Higgins et al. 2006; Lomolino 1996;  
83 Millien, Zhan, Li, Wang, and Wang 2024). Second, colonization rates are often lower on more  
84 isolated islands, especially for species with limited capacity for dispersal (Higgins et al., 2006). If  
85 colonization events are non-random, then communities on isolated, species-poor islands

86 become subsets of less isolated, species-rich islands, a process known as selective  
87 colonization (Chen et al. 2022; Higgins et al. 2006; Hu et al. 2011; Lomolino 1996; Millien et al.  
88 2024). Third is habitat nestedness (Chen et al. 2022; Higgins et al. 2006; Millien et al. 2024). In  
89 circumstances where habitats are nested, specialists disappear as their habitat does, while  
90 generalists do not (Higgins et al. 2006). Islands are typically exposed to disturbances originating  
91 from the ocean (e.g. strong wind, salt spray, storms, etc.), which will favour specialists of coastal  
92 habitats (Mologni 2022; Mologni et al. 2021; Morrison and Spiller 2008; Neufeld, Starko, and  
93 Burns 2017). As exposure to these disturbances declines, so will the number of specialists,  
94 promoting nested patterns.

95         Traditionally, research has focused on taxonomic nestedness. Similar to the theory of  
96 island biogeography, this approach treats species as equivalent and independent of one  
97 another (MacArthur and Wilson 1967). More recent research that recognizes species are neither  
98 equivalent nor independent has led to evaluations of phylogenetic (Johnson, Adler, and Cherry  
99 2000; Sanmartín, Van Der Mark, and Ronquist 2008; Valente, Phillimore, and Etienne 2018;  
100 Weigelt and Kreft 2013) and functional island biogeography (Barajas Barbosa et al. 2023;  
101 Mologni et al. 2022; Schrader, Wright, Kreft, and Westoby 2021; Walentowitz, Troiano,  
102 Christiansen, Steinbauer, and Barfod 2022; Whittaker et al. 2014). Phylogenetic and functionally  
103 informed approaches have gradually extended to studies of nestedness (Chen et al. 2022;  
104 Millien et al. 2024; Zhan, Li, Chen, and Wang 2024). Implementing this approach has revealed  
105 distinct patterns in the nestedness of amphibians (Chen et al., 2022) and mammals (Millien et  
106 al., 2024; Zhan et al., 2024), both dependent on their phylogenetic relationships and functional  
107 traits. However, little is known about how phylogeny and traits influence nested patterns of  
108 native plant communities on islands.

109         One way to integrate plant phylogeny in taxonomic nestedness studies on islands is by  
110 using broad phylogenetic categories, such as monocots, dicots, ferns, and conifers. While some  
111 of these groups are highly diverse and occupy a wide range of habitats, others are not. For  
112 instance, many fern species are confined to shaded, humid habitats (Mehltreter 2010), hence,  
113 islands without forest cover or highly exposed to the ocean might be too open or too lacking in  
114 humidity for many of them. Similarly, several traits can be integrated in taxonomic nestedness  
115 studies. For example, identifying traits associated with plant body size might help characterize  
116 the role of island area and selective extinction in shaping island plant communities. Most woody  
117 species are bigger than herbs (Pérez-Harguindeguy et al. 2013) and have higher minimum area  
118 requirements. Instead, other traits could be associated with selective colonization. Vectors vary  
119 among species, and while some, such as the wind, allow species to disperse over extended

120 distances, others, such as ants, will limit the distance a species can disperse to (Arjona,  
121 Nogales, Heleno, and Vargas 2018; Mologni et al. 2022; Negoita et al. 2016). Similarly, taller  
122 plants can disperse over longer distances (Thomson, Moles, Auld, and Kingsford 2011) but they  
123 might also have higher minimum area requirements due to their larger body size.

124         There are several nestedness metrics (Almeida-Neto, Guimarães, Guimarães, Loyola,  
125 and Ulrich 2008; Atmar and Patterson 1993; Ulrich, Almeida-Neto, and Gotelli 2009; Wright,  
126 Patterson, Mikkelsen, Cutler, and Atmar 1998). The nestedness metric based on overlap and  
127 decreasing fill ('NODF' hereafter) is generally considered the most appropriate (Almeida-Neto et  
128 al. 2008; Chen et al. 2022; Matthews et al. 2015; Ulrich, Almeida-Neto, et al. 2009). That is  
129 because it quantifies nestedness not exclusively for a whole presence/absence matrix (NODF),  
130 but also for columns (i.e. NODFc, or species composition across sites) and rows (i.e. NODFr, or  
131 species incidence) separately, allowing to determine the independent contribution of each  
132 component (Almeida-Neto et al. 2008; Chen et al. 2022). Quantifying column nestedness allows  
133 for assessing whether communities are nested among islands, while row nestedness for  
134 evaluating if specialists and rare species are present on islands that have generalists and  
135 widespread species (Traveset et al. 2014). Moreover, the NODF focuses on the overlap of  
136 species occurrences between sites rather than on unexpected presences or absences in the  
137 matrix. This reduces the risk of over- and underestimating nested patterns, respectively (Ulrich,  
138 Almeida-Neto, et al. 2009; Wright et al. 1998). The NODF is also more robust to variations in  
139 matrix fill (i.e. the proportion of non-zero elements), size (i.e. the total number of elements), and  
140 shape (i.e. the number of columns and rows), enabling the comparison of markedly different  
141 matrices (Almeida-Neto et al. 2008).

142         Here, we investigated patterns of taxonomic nestedness of island floras in relation to  
143 island characteristics (i.e. area, isolation, and exposure to ocean-borne disturbances). We  
144 integrated phylogenetic and functional traits to gain insight into the processes that shape the  
145 composition and structure of island native floras. We focused on 264 islands offshore from  
146 northern Aotearoa New Zealand. We combined field surveys and published data for 775 native  
147 plant species. We compiled information about their phylogeny and functional traits and explored  
148 nestedness patterns using species incidence matrices and correlation techniques. We asked  
149 whether (i) plant species assemblages are nested according to their phylogeny and functional  
150 traits and (ii) whether nested patterns are related to island characteristics.

151

152 **Methods**

153 *Study system*

154 The islands in the study system are located between 34–38° S and 172–179° E, spanning over  
155 600 km, and vary markedly in their characteristics (Figure 1). The smallest island is only 21 m<sup>2</sup>  
156 and the least isolated are tidal islands (i.e. separated from the mainland only at high tide).  
157 Conversely, the largest island is over 277 km<sup>2</sup> and the most isolated is over 50 km from the  
158 mainland. Geologically, the islands are nearly evenly split between volcanic (n=136, 51.5%) and  
159 sedimentary (n=122, 46.2%), with a small number having a mixed composition (n=6, 2.3%).  
160 Except for the Poor Knights Islands/Tawhitirāhi and Rangitoto, all were connected during the  
161 last glacial maximum (Fleming, 1979, Shane et al., 2013).

162 The climate of the islands is temperate and oceanic (McGlone, Buitenwerf, and  
163 Richardson 2016) and all but the smallest islands and those with <5 m maximum elevation  
164 originally supported warm temperate rain forests (Grubb, Bellingham, Kohyama, Piper, and  
165 Valido 2013; Peel, Finlayson, and McMahon 2007; Wilmshurst et al. 2014). Most islands have  
166 elevations <400 m, where the forests were comprised of evergreen angiosperms and conifers,  
167 the latter often emergent over angiosperms (Dawson 1988; Grubb et al. 2013). Today, old-  
168 growth forests are rare in the study system, and confined to the largest islands (e.g. Cameron &  
169 Young, 2019).

170 Most of the islands were deforested since human settlement of Aotearoa New Zealand  
171 in c. 1250 CE (Atkinson 2004; Bellingham et al. 2010; Daugherty, Gibbs, Towns, and Atkinson  
172 1990; Wilmshurst et al. 2014). Both burning and agriculture were abandoned on most islands,  
173 which are now in a process of secondary succession such that most islands are now dominated  
174 by second-growth warm temperate rain forests (Atkinson 2004; P J Bellingham 1984). Common,  
175 widespread trees in these second-growth forests include *Brachyglottis repanda* (Asteraceae),  
176 *Coprosma macrocarpa* and *C. repens* (Rubiaceae), *Corynocarpus laevigatus*  
177 (Corynocarpaceae), *Kunzea ericoides* s.l. (Myrtaceae), *Melicope ternata* (Rutaceae), *Melicytus*  
178 *novae-zelandiae* and *M. ramiflorus* (Violaceae), *Metrosideros excelsa* (Myrtaceae), *Piper*  
179 *excelsum* (Piperaceae), *Pittosporum crassifolium* (Pittosporaceae), *Planchonella costata*  
180 (Sapotaceae), and *Pseudopanax lessonii* (Araliaceae) (Atkinson 2004; Wardle, Bellingham,  
181 Bonner, and Mulder 2009). Two of the largest islands (Hauturu-a-Toi/Little Barrier Island, 30.79  
182 km<sup>2</sup>, and Aotea/Great Barrier Island, 277.21 km<sup>2</sup>), with elevations up to 722 m, support  
183 floristically distinct old-growth original forests above 500 m (Cameron and Young 2019; Ogden  
184 and Perry 2023). Active and recent volcanic islands (i.e. Whakaari/White Island and Rangitoto)

185 are undergoing primary succession, in which *M. excelsa* is the dominant colonising tree  
186 (Clarkson and Clarkson 1994; Shane et al. 2013). Non-native plant species, most introduced  
187 since the 19th century, comprise approximately half of the total flora of the study system  
188 (Mologni et al. 2021). Presently, 41 islands (15.5% of all islands) are inhabited.

189

### 190 *Species lists and plant functional traits*

191 We collated lists of plant species native to Aotearoa New Zealand (>80% endemic; McGlone et  
192 al., 2001) for each of the 264 islands using both published material and field surveys (see  
193 Mologni et al., 2021, 2022 for more details). After removing duplicates and aggregating records  
194 at the species level, a total of 810 native plant species were recorded. We removed 35 species  
195 due to taxonomic uncertainty (e.g. genus only noted in species lists), thus only 775 were used in  
196 the analyses. A further 25 and 3 species were removed from dispersal modes and plant height  
197 analyses due to a lack of data (Table 1). Lists were combined into a presence/absence matrix.  
198 Plant species lists follow consistent vascular plant taxonomy (Allan Herbarium 2023b, 2023a).

199 To account for phylogeny, we organized species matrices using both fine (species level)  
200 and coarse phylogenetic levels (i.e. dicots, monocots, conifers, and ferns and allies, Table 1).  
201 To account for plant functional traits, we organized species according to their growth forms and  
202 dispersal modes and determined species maximum height. We used five categories for growth  
203 forms: graminoids, forbs, woody species, climbers and lianas, and epiphytes (Table 1). We used  
204 five categories for dispersal modes: unspecialized, short-distance, wind-, animal-, and water-  
205 dispersed (Table 1). Data for each species' maximum height were obtained in July 2024 mostly  
206 from the New Zealand Plant Conservation Network (2024) and the Manaaki Whenua - Landcare  
207 Research (2005, see Appendix 2 for the full list), with additions from expert knowledge (PJB).  
208 For species that lacked data for maximum height, we used average height data instead. For  
209 hybrids, we used the average of the parent species' maximum or average height. We then  
210 organized species in three categories: short (maximum height < 1 m), medium (1–10 m), and  
211 tall (> 10 m, Table 1). We excluded categories with a sample size below 50 species, namely  
212 conifers (n = 12), climbers and lianas (n = 35), epiphytes (n = 8), and short-distance (n = 34,  
213 Table 1).

214

### 215 *Island characteristics*

216 We quantified three island characteristics: area, isolation, and exposure to ocean-borne  
217 disturbances (e.g. storms, waves, and salt spray). We measured island area as the surface of

218 an island viewed from above (km<sup>2</sup>) using available sources or manually digitizing it (Mologni et  
219 al. 2021). Isolation was measured as a series of concentric belts of different radii surrounding  
220 each island (radii = 250, 500, 1000, 1500, 2000, 2500, 3000 m, see Carter, Perry, and Russell  
221 2020; Diver 2008; Mologni et al. 2021; Negoita et al. 2016; Weigelt and Kreft 2013). We  
222 quantified the proportion of land within each radius and subtracted this value from the total area,  
223 to ensure that greater values indicate greater isolation. We utilized the radius that performed  
224 best (Mologni et al. 2021). We estimated exposure to ocean-borne disturbances by drawing two  
225 lines starting from the centre of an island and connecting to the edge of the nearest landmasses  
226 (the mainland or islands larger than 50 km<sup>2</sup>). The angle between these two lines represents the  
227 degree of exposure of an island to ocean-borne disturbances (Burns and Neufeld 2009). All  
228 spatial analyses were carried out using ArcGIS 10 and QGIS 2 (ESRI 2011; QGIS Development  
229 Team 2018).

230

### 231 *Statistical analyses*

232 For all phylogenetic and trait categories, we calculated Nestedness based on overlap and  
233 decreasing fill (NODF, Almeida-Neto et al., 2008). Nestedness was assessed on the entire  
234 matrix (NODF), on columns (i.e. species composition, NODFc) and rows (i.e. species incidence,  
235 NODFr). We employed a fixed-fixed null model with a 'quasi-swap' algorithm and 999  
236 simulations to test for significant nestedness (Millien et al. 2024; Ulrich, Almeida-Neto, et al.  
237 2009). Fixed-fixed null models maintain row and column totals fixed, which reduces the number  
238 of possible arrangements in the randomized matrices and thus the likelihood of identifying  
239 statistically significant nested patterns (Matthews et al. 2015). This is a more conservative  
240 approach which reduces the risk of type I errors (Ulrich, Almeida-Neto, and Gotelli 2009).

241 Relationships between island nestedness ranks and the three island characteristics were  
242 quantified using partial Spearman correlation coefficients (Millien et al. 2024). This method  
243 enables the comparison of more than 3 predictors and the assessment of the independent effect  
244 of each one (Frick, Hayes, and Heady III 2009; Millien et al. 2024; Shipley 2016). We extracted  
245 island nestedness ranks from the nested incidence matrix and correlated them with island  
246 characteristics (i.e. area, isolation, exposure to ocean-borne disturbances). Nestedness ranks  
247 assign higher values to species-poor sites and species that occur less frequently. If, for  
248 instance, species disappear as area declines so that communities on small islands are subsets  
249 of communities on large islands, this will result in a negative relationship between nestedness  
250 ranks and island area. Plant categories whose distribution is not significantly nested, or nested

251 only according to rows (i.e. species incidence, NODFr) will be excluded from this analysis. A  
252 Bonferroni correction was applied to account for multiple tests and reduce type I error rates.

253 This approach was applied to all native species (fine phylogenetic level) and all subsets  
254 separately (coarse phylogenetic level, growth forms, dispersal modes, and plant height  
255 categories). Spearman correlation coefficients were calculated for all comparisons with more  
256 than one predictor, but no correlation exceeded the threshold of 0.7 (Appendix 3, Figure S1). All  
257 analyses were conducted in R. Nestedness was quantified using the package *vegan* (Oksanen  
258 et al. 2024), nestedness ranks using the package *bipartite* (Dormann, Gruber, and Fründ 2008),  
259 and Spearman partial correlation coefficients using the package *ppcor* (Seongho Kim 2012).

260

## 261 **Results**

262 Most native species groups were significantly nested based on at least one of the 3 nestedness  
263 metrics ( $P < 0.039$ , full matrix, rows, and columns, Table 2). All species (fine phylogenetic level),  
264 dicots, forbs, wind-dispersed, short and medium-height species were nested based on both the  
265 full matrix and columns (Table 2). Graminoids, unspecialized, and medium-height species were  
266 nested only according to columns (i.e. species composition, Table 2). Tall species were nested  
267 based on both the full matrix and rows, while woody species and animal-dispersed species were  
268 nested only according to rows (i.e. species incidence, Table 2). Monocots were nested  
269 according to both columns and rows, but not across the entire matrix (i.e., both species  
270 composition and incidence, Table 2). Ferns and allies and water-dispersed species were not  
271 significantly nested ( $P > 0.05$ , Table 2).

272 Island area was negatively related to island nestedness ranks of all plant categories that  
273 were significantly nested according to the full matrix (NODF) or columns (i.e. species  
274 composition, see Figure 2 & Appendix 3, Table S1). Isolation was positively related to the island  
275 nestedness ranks of all species (fine phylogenetic level), monocots among the coarse  
276 phylogenetic categories, graminoids among growth forms, wind-dispersed and unspecialized  
277 species among dispersal modes, and short and medium-height species among plant height  
278 categories. The island nestedness ranks of dicots, forbs, and tall species were not significantly  
279 related to isolation ( $P > 0.05$ ). Exposure to ocean-borne disturbances was positively related to  
280 the island nestedness ranks of all phylogenetic groups (all species, monocots, dicots), forbs  
281 among growth forms, wind-dispersed species among dispersal modes, and all plant height  
282 categories (short, medium, and tall). The island nestedness ranks of graminoids and  
283 unspecialized species were not significantly related to exposure to ocean-borne disturbances.

284

285 **Discussion**

286 Assemblages of the highly endemic native plant species across 264 offshore islands in  
287 Aotearoa New Zealand are largely nested, regardless of phylogenetic and functional trait  
288 categories, except for ferns and water-dispersed species assemblages. With light propagules  
289 (e.g. spores) and the ability to take advantage of water currents, these plant categories can  
290 disperse over long distances (Thomson et al. 2011), potentially reaching most islands and  
291 overriding nested patterns. Woody and animal-dispersed species were nested only according to  
292 rows (i.e., species incidence), indicating that widespread species occur across most sites, while  
293 rarer species are restricted to sites with higher diversity. This pattern is likely influenced by  
294 differences in plant traits, such as variations in the dispersal capacity of carriers for animal-  
295 dispersed species. The nestedness ranks of plant categories nested according to the full matrix  
296 or columns (i.e. species composition) were generally correlated with island area, isolation, and  
297 exposure to disturbance, with variations depending on phylogenetic and functional trait  
298 categories.

299         Nestedness ranks of native plant species were consistently related to island area,  
300 irrespective of the phylogenetic or functional group considered. Island area is a well-established  
301 predictor of plant diversity (Arrhenius 1921; MacArthur and Wilson 1967). Larger islands provide  
302 more space, more diverse habitats, and more resources, supporting higher species richness  
303 (Tjørve and Turner 2009). Previous work in the same study system found that larger islands  
304 contain more species, regardless of plant functional traits (Mologni et al. 2021, 2022). In the  
305 current study, we observed that native plant assemblages on smaller islands were consistently  
306 subsets of those of larger islands. This pattern aligns with the selective extinction hypothesis,  
307 which posits that as islands get smaller, species with larger area requirements or smaller  
308 population sizes are more likely to become extinct, generating nested patterns (Higgins et al.  
309 2006; Matthews et al. 2015; Tjørve and Turner 2009). In addition, the effect of island area was  
310 at least twice as great as that of other predictors, suggesting selective extinction might be the  
311 dominant process in the study system (see Figure 2). However, the effect size of island area  
312 was comparable across all groups. If both large- and small-bodied species (e.g. woody species  
313 and graminoids, short and medium-height species) exhibit similar responses to island area, then  
314 minimum area requirements might be less important in plants than in animals (Millien et al.  
315 2024).

316           The pronounced effect of area on island floras has strong conservation implications. In  
317 the longstanding debate on whether it is best to protect a single large area versus several small  
318 ones (Wright and Reeves 1992), our overall results support protecting larger islands (Diamond  
319 1975; Donaldson, Wilson, and Maclean 2017; Fahrig et al. 2022; Tjørve and Tjørve 2008).  
320 However, options to protect larger islands in Aotearoa New Zealand are limited simply because  
321 there are few of them and most of the islands >100 ha in our study region are either wholly or  
322 largely managed for nature conservation (e.g., Armitage, 2011; Veitch & Wade, 2019). Their  
323 long-term protection gives the best prospect of safeguarding the greatest number of native plant  
324 species. Nationally and in our study region over 70% of protected islands are <100 ha (Towns  
325 and Ballantine 1993). The smaller islands in our study region need long-term protection to  
326 preserve the diversity of plant categories that are not nested, such as ferns, water-dispersed,  
327 and tall species, as well as endemic plant species that are confined to or are most abundant on  
328 small islands with high densities of nesting seabirds and with strong maritime climatic influences  
329 (Cameron 1990; Gillham 1960; Norton, Delange, Garnock-Jones, and Given 1997).

330           A second process linked to nested patterns is selective immigration, when, due to  
331 varying dispersal capabilities, communities on isolated islands become subsets of those on less  
332 isolated islands. However, the effect of isolation was less consistent than that of area, and three  
333 categories did not respond to it: dicots, forbs, and tall species. One explanation is that these  
334 categories might simply be more homogenous in their dispersal capabilities. For instance, tall  
335 species are likely to be consistently better dispersers than short and medium-height species  
336 (Thomson et al. 2011).

337           Similarly, the effect of islands' exposure to ocean-borne disturbances varied across  
338 groups. This suggests that habitat nestedness, or the systematic disappearance of specialists  
339 following their habitat, does not apply to all cases. Among growth forms, graminoids were the  
340 only category unrelated to exposure. Some grasses deposit high amounts of silica in their  
341 leaves (Lanning and Eleuterius 1983), increasing resistance to sand, salt, and other particulate  
342 matter carried by the wind, making graminoids more resistant to ocean-borne disturbances.  
343 Among dispersal modes, only wind-dispersed species were related to exposure. Strong oceanic  
344 winds will likely blow propagules away, preventing colonization (Carlquist 1974; Cody and  
345 Overton 1996). If so, this pattern might be better explained by selective immigration than habitat  
346 nestedness. Alternatively, ocean-borne disturbances might simply increase the chance of  
347 extinction events, resulting in the depauperate floras of exposed islands being a subset of the  
348 floras found on sheltered islands. Other studies have also shown that pressure from  
349 disturbances results in nestedness (e.g. in breeding bird communities, Wang, Ding, Chen, and

350 Zheng 2013). Ocean-borne disturbances are likely to play a role in explaining the effect of island  
351 area as well, given that species on smaller islands are more prone to extinction (MacArthur and  
352 Wilson 1967).

353 Most of the islands in northern Aotearoa New Zealand were connected to the mainland  
354 during the last glacial maximum (Fleming, 1979, Shane et al., 2013). As such, relictual plant  
355 populations might homogenize the flora and mitigate nested patterns. More recently, human  
356 activities have heavily modified these islands. The land was often burned or cleared for  
357 agriculture. Historic deforestation and subsequent secondary successions, which are likely to  
358 have been influenced at least in part by dispersal of species among islands and from the  
359 mainland, are also likely to have homogenized the flora (Christensen and Peet 1984) and  
360 mitigated nested patterns. Finally, a large number of non-native species were introduced and  
361 they now amount to half of the total flora (Mologni et al. 2021), some of which are invasive and  
362 highly competitive. Habitat loss and invasive species most likely drove more than one native  
363 species to extinction, influencing nestedness patterns. Unfortunately, we could not account for  
364 these factors. Data are lacking for relictual populations and habitat loss, while non-native and  
365 native species have similar distributions within this study system (Mologni et al. 2021), rendering  
366 any interpretation potentially circular.

367 In this study, we focused only on ecological processes such as selective extinction and  
368 immigration. However, nested patterns can also be generated by stochastic processes,  
369 prominently passive sampling (i.e. rare species are underrepresented in the community  
370 compared to abundant species) and collecting artifacts (i.e. the disproportional sampling of rare  
371 species or small assemblages, Higgins, Willig, and Strauss 2006; Hu, Feeley, Wu, Xu, and Yu  
372 2011). Unfortunately, we lack data to investigate the role of these processes. Additionally, we  
373 used a fixed-fixed null model and the Bonferroni correction, both very conservative approaches.  
374 While this reduces the risk of type I errors, they inevitably increase the risk of type II errors.  
375 Similar trade-offs also apply to other methods, and we thus advise caution in interpreting the  
376 results.

377 Overall, native plant communities were nested across northern Aotearoa New Zealand  
378 islands but varied according to plant phylogeny and functional traits. This aligns with previous  
379 work on animals (Chen et al., 2022; Millien et al., 2024; Zhan et al., 2024), and highlights the  
380 importance of incorporating phylogeny and functional traits into studies of island nestedness of  
381 plants (Schrader et al. 2021). Furthermore, island floras were largely nested according to area,  
382 which emphasises the need to protect large islands, since, at least in our study system, some of

383 the larger islands are also inhabited, highly modified, and more invaded (Mologni et al. 2021;  
384 Mologni, Bellingham, Cameron, and Wright 2024). Nestedness studies can suggest probable  
385 processes that determine insular community composition and aid in identifying conservation  
386 priority islands.

387 **Data availability statement**

388 The data in this research were collected from islands owned by Māori or from islands over  
389 which they have customary authority. The authors recognise their sovereignty and authority to  
390 control data about their lands under the CARE Principles for Indigenous Data Governance. Data  
391 for island characteristics are available from the Manaaki Whenua data repository at  
392 <https://doi.org/10.7931/ndkt-zw49>. Island locations (longitude and latitude) are not publicly  
393 available due to private ownership and issues of data sovereignty of concern to Māori.

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395 **Code availability statement**

396 The code used for this research is available at as supplementary material.

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401 **Tables**

402 **Table 1** - Description of phylogenetic and functional trait categories utilized to investigate  
 403 taxonomic nestedness across 264 Aotearoa New Zealand offshore islands. Numbers with an  
 404 asterisk indicate categories excluded due to the small sample size (n<50).

Category	Description	Number
<b>Phylogenetic categories</b>		
<i>Fine level</i>	All species	775
<i>Coarse level</i>		775
Ferns and allies	Ferns and lycophytes	157
Conifers	Conifers	12*
Monocots	Monocotyledons	220
Dicots	Dicotyledons	386
<b>Functional trait categories</b>		
<i>Growth forms</i>		775
Graminoids	Grasses, sedges, and rushes	123
Forbs	Herbaceous, non-graminoid	393
Woody species	Trees and shrubs	216
Climbers and lianas	Herbaceous or woody climbers	35*
Epiphytes	Plants that grow upon other plants	8*
<i>Dispersal modes</i>		750
Water-dispersed	Buoyant propagules, e.g., corky tissues, air pockets (hydrochory)	61
Unspecialized	No evident or prevalent morphological adaptations (unspecialized)	110
Short-distance	Morphological adaptations for short-distance dispersal only (ballochory, myrmecochoy)	34*
Animal-dispersed	Fleshy fruits or adhesive barbs (endo and epizoochory)	176
Wind-dispersed	Plumes, wings, dust diaspores (anemochory)	369

<i>Plant height</i>		772
Short	Maximum height below 1 m	361
Medium	Maximum height between 1 and 10 m	335
Tall	Maximum height above 10 m	76

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428 **Table 2** - Nestedness parameters for matrices including all species (phylogenetic fine level) and  
 429 phylogenetic (coarse level) and functional (growth forms, dispersal modes, plant height)  
 430 categories. In bold are significant nested patterns. In grey are categories that are not  
 431 significantly nested for any parameter.

Category	NODF	p-value	NODFr	p-value	NODFc	p-value
<b>Phylogenetic categories</b>						
<i>Fine level</i>	<b>55.27</b>	<b>0.003</b>	64.29	0.303	<b>54.231</b>	<b>0.003</b>
<i>Coarse level</i>						
Ferns and allies	56.24	0.509	53.52	0.439	63.94	0.959
Monocots	54.84	0.659	<b>53.96</b>	<b>0.005</b>	<b>56.12</b>	<b>0.005</b>
Dicots	<b>55.07</b>	<b>0.007</b>	64.31	0.529	<b>50.76</b>	<b>0.003</b>
<b>Functional trait categories</b>						
<i>Growth forms</i>						
Graminoids	53.83	0.533	53.00	0.089	<b>57.70</b>	<b>0.007</b>
Forbs	<b>57.71</b>	<b>0.001</b>	65.14	0.473	<b>54.37</b>	<b>0.001</b>
Woody species	59.79	0.201	<b>62.86</b>	<b>0.039</b>	55.19	0.899
<i>Dispersal modes</i>						
Wind-dispersed	<b>59.52</b>	<b>0.015</b>	64.58	0.253	<b>56.91</b>	<b>0.025</b>
Water-dispersed	61.95	0.249	62.33	0.391	54.69	0.163
Unspecialized	48.84	0.053	48.23	0.209	<b>52.36</b>	<b>0.033</b>
Animal-dispersed	66.44	0.137	<b>70.73</b>	<b>0.005</b>	56.77	0.393

<i>Plant height</i>						
Short	<b>56.65</b>	<b>0.001</b>	61.60	0.053	<b>54.00</b>	<b>0.003</b>
Medium	<b>56.83</b>	<b>0.019</b>	63.84	0.413	<b>52.48</b>	<b>0.011</b>
Tall	<b>51.74</b>	<b>0.033</b>	<b>50.19</b>	<b>0.047</b>	70.63	0.495

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434 **Figures**

435 **Figure 1** Map and example photos of the 264 islands offshore northern Aotearoa New  
436 Zealand investigated.

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438 **Figure 2** Relationships between the nestedness rankings of phylogenetic and functional  
439 trait categories and island characteristics across 264 islands offshore northern Aotearoa New  
440 Zealand. On the x-axis are species categories. Categorizations were phylogenetic (fine - all  
441 species; coarse - monocots and dicots), growth forms (graminoids and forbs), dispersal modes  
442 (wind-dispersed and unspecialized) and plant height categories (short, medium, and tall). On  
443 the y-axis are rho values, which quantify the strength and direction of the relationship between  
444 two variables. A Bonferroni correction was applied to reduce type 1 error rates. The alpha level  
445 was set at 0.005, dividing 0.05 by the number of categories ( $n = 10$ ).

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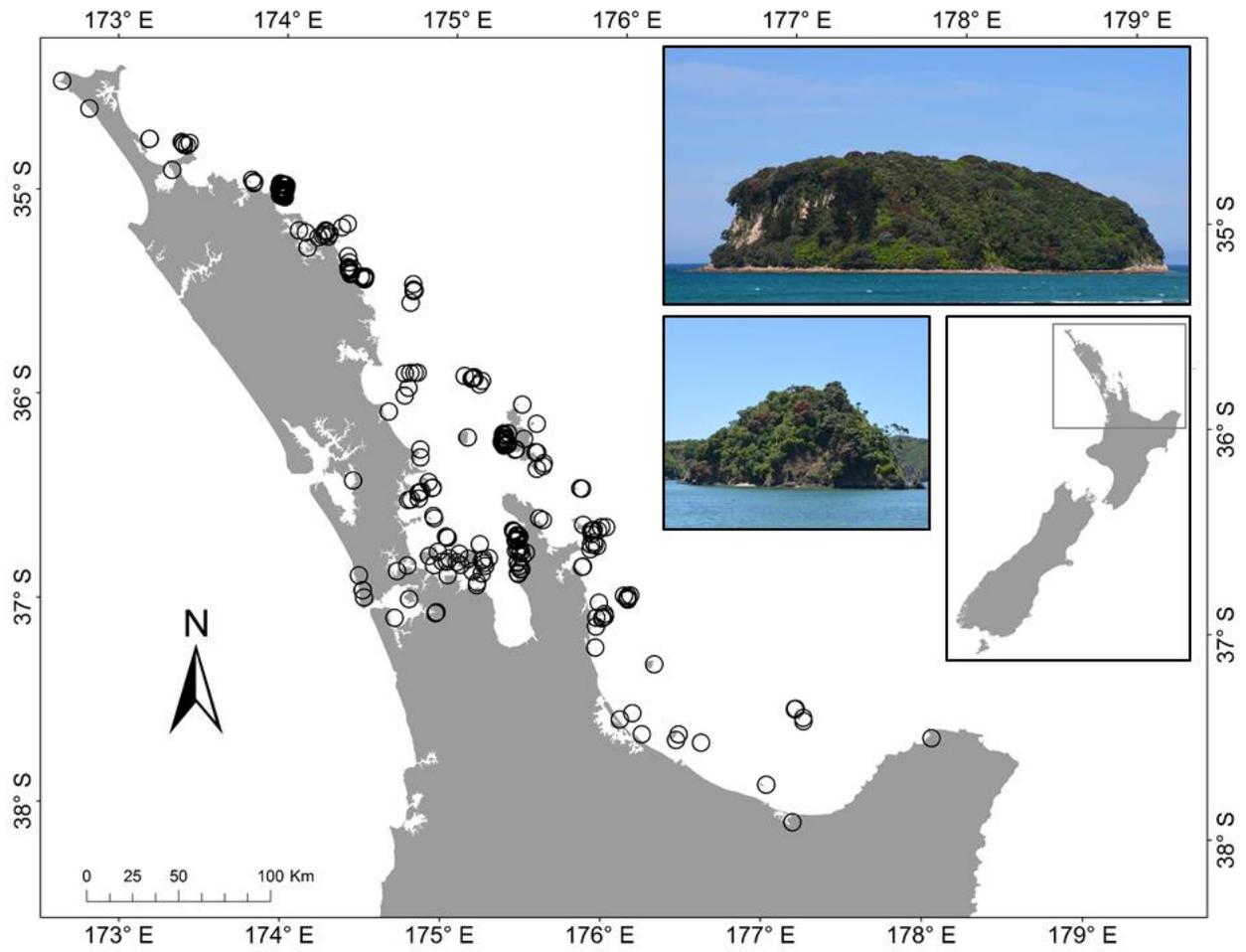
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463 **Figure 1**

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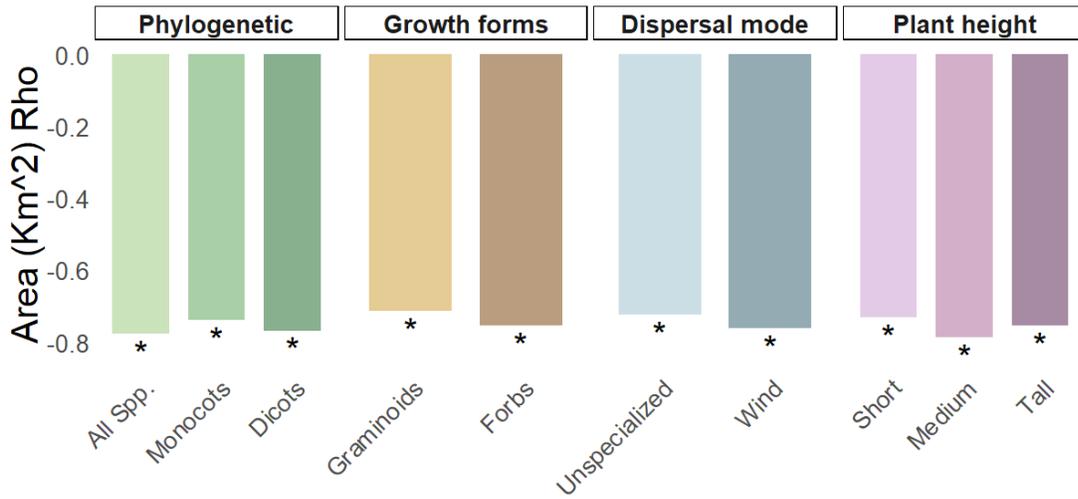
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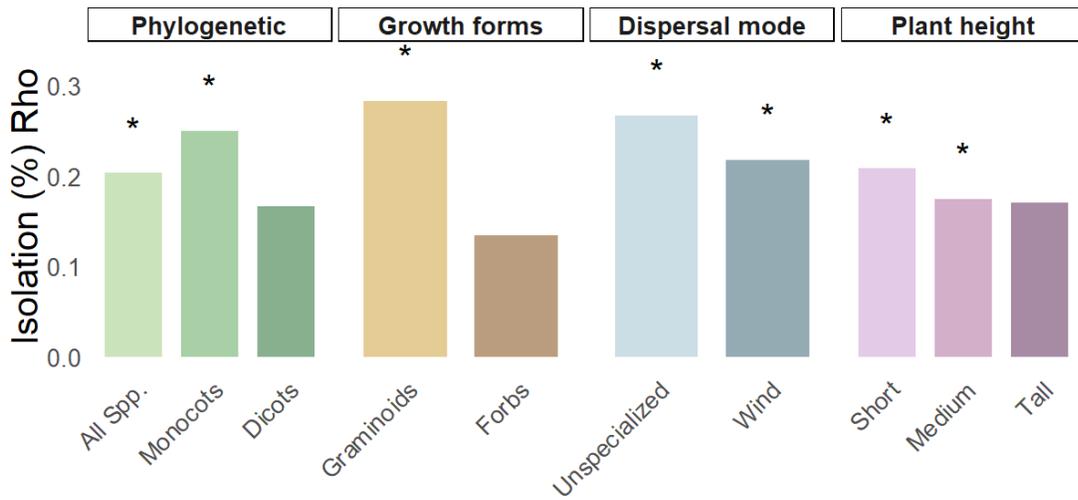
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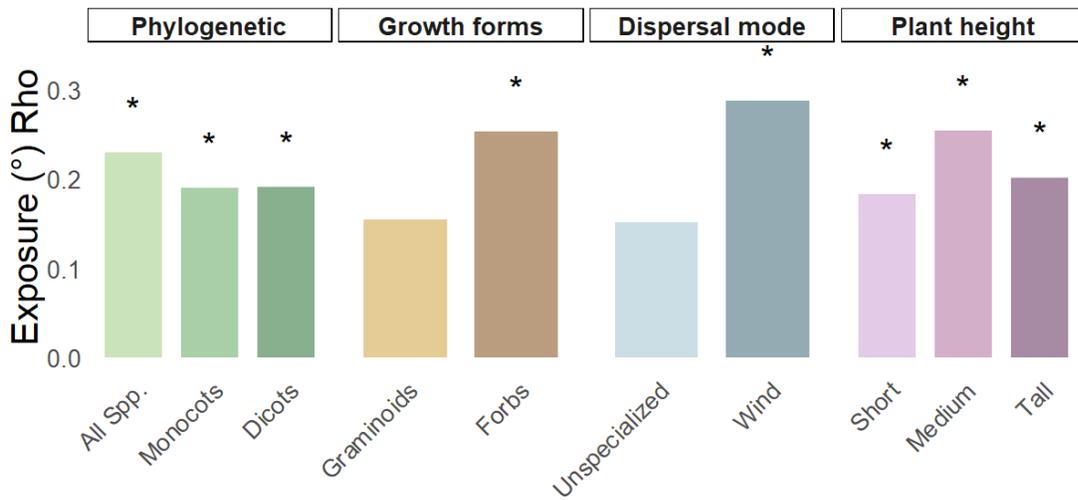
476 **Figure 2**



Island nestedness ranks of plant categories



Island nestedness ranks of plant categories



Island nestedness ranks of plant categories

478 **References**

- 479 Allan Herbarium. (2023a). Checklist of the New Zealand Flora – Ferns and Lycophytes.  
480           Manaaki Whenua Landcare Research.
- 481 Allan Herbarium. (2023b). Checklist of the New Zealand Flora – Seed plants. Manaaki Whenua  
482           Landcare Research.
- 483 Almeida-Neto, M., Guimarães, P., Guimarães, P. R., Loyola, R. D., & Ulrich, W. (2008). A  
484           consistent metric for nestedness analysis in ecological systems: reconciling concept and  
485           measurement. *Oikos*, 117(8), 1227–1239. [https://doi.org/10.1111/j.0030-](https://doi.org/10.1111/j.0030-1299.2008.16644.x)  
486           1299.2008.16644.x
- 487 Arjona, Y., Nogales, M., Heleno, R., & Vargas, P. (2018). Long-distance dispersal syndromes  
488           matter: diaspore–trait effect on shaping plant distribution across the Canary Islands.  
489           *Ecography*, 41(5), 805–814. <https://doi.org/10.1111/ecog.02624>
- 490 Armitage, D. (ed.). (2011). *Great Barrier Island*. Revised edition. Canterbury University Press.  
491           Christchurch, New Zealand.
- 492 Arrhenius, O. (1921). Species and Area. *Journal of Ecology*, 9(1), 95–99.  
493           <https://doi.org/10.2307/2255763>
- 494 Atkinson, I. A. E. (2004). Successional processes induced by fires on the northern offshore  
495           islands of New Zealand. *New Zealand Journal of Ecology*, 28(2), 181–193.
- 496 Atmar, W., & Patterson, B. D. (1993). The measure of order and disorder in the distribution of  
497           species in fragmented habitat. *Oecologia*, 96(3), 373–382.  
498           <https://doi.org/10.1007/BF00317508>
- 499 Barajas Barbosa, M. P., Craven, D., Weigelt, P., Denelle, P., Otto, R., Díaz, S., ... Kreft, H.  
500           (2023). Assembly of functional diversity in an oceanic island flora. *Nature*, 619(7970),  
501           545–550. <https://doi.org/10.1038/s41586-023-06305-z>
- 502 Bellingham, P J. (1984). Forest regeneration on Lady Alice Island, Hen and Chickens group.  
503           *Tane*, 30, 31–42.

504 Bellingham, P. J., Towns, D. R., Cameron, E. K., Davis, J. J., Wardle, D. A., Wilmshurst, J. M.,  
505 & Mulder, C. P. H. (2010). New Zealand island restoration: seabirds, predators, and the  
506 importance of history. *New Zealand Journal of Ecology*, 34(1), 115–136.

507 Burns, K. C., & Neufeld, C. J. (2009). Plant extinction dynamics in an insular metacommunity.  
508 *Oikos*, 118(2), 191–198. <https://doi.org/10.1111/j.1600-0706.2008.16816.x>

509 Cameron, E. K. (1990). Flora and vegetation of Middle Island, Mercury Islands Group, eastern  
510 Coromandel, northern New Zealand. *Journal of the Royal Society of New Zealand*,  
511 20(3), 273–285. <https://doi.org/10.1080/03036758.1990.10416822>

512 Cameron, E. K., & Young, M. E. (2019). In Vegetation and vascular flora. In: Wade L, Veitch D  
513 (eds) Hauturu—the history, flora and fauna of Te Hauturu-o-Toi Little Barrier. (pp. 166–  
514 213). Auckland: Massey University Press.

515 Carlquist, S. J. (1974). *Island biology*. New York: Columbia University Press.  
516 <https://doi.org/10.5962/bhl.title.63768>

517 Carter, Z. T., Perry, G. L. W., & Russell, J. C. (2020). Determining the underlying structure of  
518 insular isolation measures. *Journal of Biogeography*, 47(4), 955–967.  
519 <https://doi.org/10.1111/jbi.13778>

520 Caujapé-Castells, J., Tye, A., Crawford, D. J., Santos-Guerra, A., Sakai, A., Beaver, K., ...  
521 Kueffer, C. (2010). Conservation of oceanic island floras: Present and future global  
522 challenges. *Perspectives in Plant Ecology, Evolution and Systematics*, 12(2), 107–129.  
523 <https://doi.org/10.1016/j.ppees.2009.10.001>

524 Chen, C., Zhan, C., & Wang, Y. (2022). Do functional and phylogenetic nestedness follow the  
525 same mechanisms as taxonomic nestedness? Evidence from amphibians in the largest  
526 archipelago of China. *Journal of Animal Ecology*, 91(12), 2424–2436.  
527 <https://doi.org/10.1111/1365-2656.13824>

528 Christensen, N. L., and R. K. Peet. (1984). Convergence During Secondary Forest Succession.  
529 *Journal of Ecology*, 72(1), 25–36. <https://doi.org/10.2307/2260004>

530

531 Clarkson, B. D., & Clarkson, B. R. (1994). Vegetation decline following recent eruptions on  
532 White Island (Whakaari), Bay of Plenty, New Zealand. *New Zealand Journal of Botany*,  
533 32(1), 21–36. <https://doi.org/10.1080/0028825X.1994.10410404>

534 Cody, M. L., & Overton, J. McC. (1996). Short-term evolution of reduced dispersal in island plant  
535 populations. *Journal of Ecology*, 84(1), 53–61. <https://doi.org/10.2307/2261699>

536 Daugherty, C. H., Gibbs, G. W., Towns, D. R., & Atkinson, I. A. E. (1990). The significance of  
537 the biological resources of New Zealand islands for ecological restoration. *Conservation*  
538 *Sciences Publication (New Zealand)*, (2). Retrieved from  
539 <https://agris.fao.org/search/en/providers/122578/records/6471d3902a40512c710e72f6>

540 Dawson, J. (1988). *Forest vines to snow tussocks: the story of New Zealand plants*. Wellington:  
541 Victoria Univ. Press.

542 Diamond, J. M. (1975). The island dilemma: Lessons of modern biogeographic studies for the  
543 design of natural reserves. *Biological Conservation*, 7(2), 129–146.  
544 [https://doi.org/10.1016/0006-3207\(75\)90052-X](https://doi.org/10.1016/0006-3207(75)90052-X)

545 Diver, K. C. (2008). Not as the crow flies: assessing effective isolation for island biogeographical  
546 analysis. *Journal of Biogeography*, 35(6), 1040–1048. [https://doi.org/10.1111/j.1365-](https://doi.org/10.1111/j.1365-2699.2007.01835.x)  
547 [2699.2007.01835.x](https://doi.org/10.1111/j.1365-2699.2007.01835.x)

548 Donaldson, L., Wilson, R. J., & Maclean, I. M. D. (2017). Old concepts, new challenges:  
549 adapting landscape-scale conservation to the twenty-first century. *Biodiversity and*  
550 *Conservation*, 26(3), 527–552. <https://doi.org/10.1007/s10531-016-1257-9>

551 Dormann, C. F., Gruber, B., & Fründ, J. (2008). Introducing the bipartite Package: Analysing  
552 Ecological Networks, 8.

553 ESRI. (2011). *ArcGIS desktop: release 10*.

554 Fahrig, L., Watling, J. I., Arnillas, C. A., Arroyo-Rodríguez, V., Jörger-Hickfang, T., Müller, J., ...  
555 May, F. (2022). Resolving the SLOSS dilemma for biodiversity conservation: a research

556 agenda. *Biological Reviews*, 97(1), 99–114. <https://doi.org/10.1111/brv.12792>

557 Fernández-Palacios, J. M., Kreft, H., Irl, S. D. H., Norder, S., Ah-Peng, C., Borges, P. A. V., ...

558 Drake, D. R. (2021). Scientists' warning – The outstanding biodiversity of islands is in

559 peril. *Global Ecology and Conservation*, 31, e01847.

560 <https://doi.org/10.1016/j.gecco.2021.e01847>

561 Fleming, C. A. (1979). *The geological history of New Zealand and its life*. Auckland: Auckland

562 University Press.

563 Frick, W. F., Hayes, J. P., & Heady III, P. A. (2009). Nestedness of desert bat assemblages:

564 species composition patterns in insular and terrestrial landscapes. *Oecologia*, 158(4),

565 687–697. <https://doi.org/10.1007/s00442-008-1168-x>

566 Gillham, M. E. (1960). Plant communities of the Mokohinau Islands, northern N.Z. *Transactions*

567 *of the Royal Society of New Zealand*, 88(1), 79–98.

568 Grubb, P. J., Bellingham, P. J., Kohyama, T. S., Piper, F. I., & Valido, A. (2013). Disturbance

569 regimes, gap-demanding trees and seed mass related to tree height in warm temperate

570 rain forests worldwide: Gap-demanders in warm temperate rain forests. *Biological*

571 *Reviews*, 88(3), 701–744. <https://doi.org/10.1111/brv.12029>

572 Higgins, C. L., Willig, M. R., & Strauss, R. E. (2006). The role of stochastic processes in

573 producing nested patterns of species distributions. *Oikos*, 114(1), 159–167.

574 <https://doi.org/10.1111/j.2006.0030-1299.14720.x>

575 Hu, G., Feeley, K. J., Wu, J., Xu, G., & Yu, M. (2011). Determinants of plant species richness

576 and patterns of nestedness in fragmented landscapes: evidence from land-bridge

577 islands. *Landscape Ecology*, 26(10), 1405–1417. [https://doi.org/10.1007/s10980-011-](https://doi.org/10.1007/s10980-011-9662-7)

578 [9662-7](https://doi.org/10.1007/s10980-011-9662-7)

579 Johnson, K. P., Adler, F. R., & Cherry, J. L. (2000). Genetic and phylogenetic consequences of

580 island biogeography. *Evolution*, 54(2), 387–396. [https://doi.org/10.1111/j.0014-](https://doi.org/10.1111/j.0014-3820.2000.tb00041.x)

581 [3820.2000.tb00041.x](https://doi.org/10.1111/j.0014-3820.2000.tb00041.x)

582 Lanning, F. C., & Eleuterius, L. N. (1983). Silica and ash in tissues of some coastal plants.  
583 *Annals of Botany*, 51(6), 835–850. <https://doi.org/10.1093/oxfordjournals.aob.a086534>

584 Lomolino, M. V. (1996). Investigating causality of nestedness of insular communities: selective  
585 immigrations or extinctions? *Journal of Biogeography*, 23(5), 699–703.  
586 <https://doi.org/10.1111/j.1365-2699.1996.tb00030.x>

587 MacArthur, R., & Wilson, E. (1967). *The theory of island biogeography*. Princeton, U.S.A.: Princeton University Press.

589 Manaaki Whenua - Landcare Research. (2005). *Ecological traits of New Zealand Flora*.  
590 Retrieved July 15, 2024, from <https://ecotraits.landcareresearch.co.nz/SearchForm.aspx>

591 Matthews, T. J., Cottee-Jones, H. E. W., & Whittaker, R. J. (2015). Quantifying and interpreting  
592 nestedness in habitat islands: a synthetic analysis of multiple datasets. *Diversity and*  
593 *Distributions*, 21(4), 392–404. <https://doi.org/10.1111/ddi.12298>

594 McGlone, M. S., Buitenwerf, R., & Richardson, S. J. (2016). The formation of the oceanic  
595 temperate forests of New Zealand. *New Zealand Journal of Botany*, 54(2), 128–155.  
596 <https://doi.org/10.1080/0028825X.2016.1158196>

597 McGlone, M. S., Duncan, R. P., & Heenan, P. B. (2001). Endemism, species selection and the  
598 origin and distribution of the vascular plant flora of New Zealand. *Journal of*  
599 *Biogeography*, 28(2), 199–216. <https://doi.org/10.1046/j.1365-2699.2001.00525.x>

600 Mehlreter, K. (2010). Fern Conservation. In Mehlreter K, Walker LR, Sharpe JM (eds). *Fern*  
601 *Ecology* (pp. 323–359). Cambridge: Cambridge University Press.

602 Millien, V., Zhan, C., Li, Y., Wang, J., & Wang, Y. (2024). A global assessment of nested  
603 patterns in insular mammal assemblages. *Global Ecology and Biogeography*, 33(9),  
604 e13885. <https://doi.org/10.1111/geb.13885>

605 Mologni, F. (2022). Different levels of disturbance influence the distributional patterns of native  
606 but not exotic plant species on New Zealand small islands. *Frontiers of Biogeography*,  
607 14(2), e54598. <https://doi.org/10.21425/F5FBG54598>

608 Mologni, F., Bellingham, P. J., Cameron, E. K., Dinh, K., Wright, A. E., & Burns, K. C. (2022).  
609 Functional traits explain non-native plant species richness and occupancy on northern  
610 New Zealand islands. *Biological Invasions*, 24(7), 2135–2154.  
611 <https://doi.org/10.1007/s10530-022-02762-1>

612 Mologni, F., Bellingham, P. J., Cameron, E. K., & Wright, A. E. (2024). Time since first  
613 naturalization is key to explaining non-native plant invasions on islands. *Journal of*  
614 *Biogeography*, 51(8), 1340–1350. <https://doi.org/10.1111/jbi.14825>

615 Mologni, F., Bellingham, P. J., Tjørve, E., Cameron, E. K., Wright, A. E., & Burns, K. C. (2021).  
616 Similar yet distinct distributional patterns characterize native and exotic plant species  
617 richness across northern New Zealand islands. *Journal of Biogeography*, 48(7), 1731–  
618 1745. <https://doi.org/10.1111/jbi.14110>

619 Morrison, L. W. (2013). Nestedness in insular floras: spatiotemporal variation and underlying  
620 mechanisms. *Journal of Plant Ecology*, 6(6), 480–488. <https://doi.org/10.1093/jpe/rtt002>

621 Morrison, L. W., & Spiller, D. A. (2008). Patterns and processes in insular floras affected by  
622 hurricanes. *Journal of Biogeography*, 35(9), 1701–1710. [https://doi.org/10.1111/j.1365-](https://doi.org/10.1111/j.1365-2699.2008.01910.x)  
623 [2699.2008.01910.x](https://doi.org/10.1111/j.1365-2699.2008.01910.x)

624 Negoita, L., Fridley, J. D., Lomolino, M. V., Mittelhauser, G., Craine, J. M., & Weiher, E. (2016).  
625 Isolation-driven functional assembly of plant communities on islands. *Ecography*, 39(11),  
626 1066–1077. <https://doi.org/10.1111/ecog.01551>

627 Neufeld, C. J., Starko, S., & Burns, K. C. (2017). Disturbance and diversity in a continental  
628 archipelago: a mechanistic framework linking area, height, and exposure. *Ecosphere*,  
629 8(10), e01957. <https://doi.org/10.1002/ecs2.1957>

630 New Zealand Plant Conservation Network. (2024). New Zealand Plant Conservation Network.  
631 Retrieved from <https://www.nzpcn.org.nz/>

632 Norton, D. A., Delange, P. J., Garnock-Jones, P. J., & Given, D. R. (1997). The role of seabirds  
633 and seals in the survival of coastal plants: lessons from New Zealand *Lepidium*

634 (Brassicaceae). *Biodiversity & Conservation*, 6(6), 765–785.  
635 <https://doi.org/10.1023/B:BIOC.0000010401.93153.29>

636 Ogden, J., & Perry, G. L. W. (2023). Ranges of woody plant species and ferns on forested  
637 elevational gradients on Aotea-Great Barrier Island, New Zealand: the role of zones of  
638 permanent and temporary establishment. *New Zealand Journal of Ecology*, 47(1), 3512.  
639 <https://doi.org/10.20417/nzjecol.47.3512>

640 Oksanen, J., Simpson, G. L., Blanchet, F. G., Kindt, R., Legendre, P., Minchin, P. R., ...  
641 Weedon, J. (2024). *vegan: Community Ecology Package*.  
642 <https://doi.org/10.32614/CRAN.package.vegan>

643 Patterson, B. D., & Atmar, W. (1986). Nested subsets and the structure of insular mammalian  
644 faunas and archipelagos. *Biological Journal of the Linnean Society*, 28(1–2), 65–82.  
645 <https://doi.org/10.1111/j.1095-8312.1986.tb01749.x>

646 Peel, M. C., Finlayson, B. L., & McMahon, T. A. (2007). Updated world map of the Köppen-  
647 Geiger climate classification. *Hydrology and Earth System Sciences*, 11(5), 1633–1644.  
648 <https://doi.org/10.5194/hess-11-1633-2007>

649 Pérez-Harguindeguy, N., Díaz, S., Garnier, E., Lavorel, S., Poorter, H., Jaureguiberry, P., ...  
650 Cornelissen, J. H. C. (2013). New handbook for standardised measurement of plant  
651 functional traits worldwide. *Australian Journal of Botany*, 61(3), 167–234.  
652 <https://doi.org/10.1071/BT12225>

653 QGIS Development Team. (2018). QGIS Geographic Information System. Retrieved from  
654 <http://qgis.osgeo.org/>

655 Sanmartín, I., Van Der Mark, P., & Ronquist, F. (2008). Inferring dispersal: a Bayesian approach  
656 to phylogeny-based island biogeography, with special reference to the Canary Islands.  
657 *Journal of Biogeography*, 35(3), 428–449. [https://doi.org/10.1111/j.1365-](https://doi.org/10.1111/j.1365-2699.2008.01885.x)  
658 [2699.2008.01885.x](https://doi.org/10.1111/j.1365-2699.2008.01885.x)

659 Schrader, J., Weigelt, P., Cai, L., Westoby, M., Fernández-Palacios, J. M., Cabezas, F. J., ...

660 Kreft, H. (2024). Islands are key for protecting the world's plant endemism. *Nature*,  
661 634(8035), 868–874. <https://doi.org/10.1038/s41586-024-08036-1>

662 Schrader, J., Wright, I. J., Kreft, H., & Westoby, M. (2021). A roadmap to plant functional island  
663 biogeography. *Biological Reviews*, 96(6), 2851–2870. <https://doi.org/10.1111/brv.12782>

664 Seongho Kim. (2012, October 29). ppcor: Partial and Semi-Partial (Part) Correlation.  
665 <https://doi.org/10.32614/CRAN.package.ppcor>

666 Shane, P., Gehrels, M., Zawalna-Geer, A., Augustinus, P., Lindsay, J., & Chaillou, I. (2013).  
667 Longevity of a small shield volcano revealed by crypto-tephra studies (Rangitoto  
668 volcano, New Zealand): Change in eruptive behavior of a basaltic field. *Journal of*  
669 *Volcanology and Geothermal Research*, 257, 174–183.  
670 <https://doi.org/10.1016/j.jvolgeores.2013.03.026>

671 Shipley, B. (2016). *Cause and Correlation in Biology: A User's Guide to Path Analysis,*  
672 *Structural Equations and Causal Inference with R* (2nd ed.). Cambridge University  
673 Press. <https://doi.org/10.1017/CBO9781139979573>

674 Thomson, F. J., Moles, A. T., Auld, T. D., & Kingsford, R. T. (2011). Seed dispersal distance is  
675 more strongly correlated with plant height than with seed mass: Dispersal distance and  
676 seed mass. *Journal of Ecology*, 99(6), 1299–1307. [https://doi.org/10.1111/j.1365-](https://doi.org/10.1111/j.1365-2745.2011.01867.x)  
677 [2745.2011.01867.x](https://doi.org/10.1111/j.1365-2745.2011.01867.x)

678 Tjørve, E., & Tjørve, K. M. C. (2008). The species-area relationship, self-similarity, and the true  
679 meaning of the z-value. *Ecology*, 89(12), 3528–3533. <https://doi.org/10.1890/07-1685.1>

680 Tjørve, E., & Turner, W. R. (2009). The importance of samples and isolates for species-area  
681 relationships. *Ecography*, 32(3), 391–400. [https://doi.org/10.1111/j.1600-](https://doi.org/10.1111/j.1600-0587.2008.05515.x)  
682 [0587.2008.05515.x](https://doi.org/10.1111/j.1600-0587.2008.05515.x)

683 Towns, D. R., & Ballantine, W. J. (1993). Conservation and restoration of New Zealand Island  
684 ecosystems. *Trends in Ecology & Evolution*, 8(12), 452–457.  
685 [https://doi.org/10.1016/0169-5347\(93\)90009-E](https://doi.org/10.1016/0169-5347(93)90009-E)

686 Traveset, A., Kueffer, C., & Daehler, C. C. (2014). Global and regional nested patterns of non-  
687 native invasive floras on tropical islands. *Journal of Biogeography*, 41(4), 823–832.  
688 <https://doi.org/10.1111/jbi.12243>

689 Ulrich, W., Almeida-Neto, M., & Gotelli, N. J. (2009). A consumer's guide to nestedness  
690 analysis. *Oikos*, 118(1), 3–17. <https://doi.org/10.1111/j.1600-0706.2008.17053.x>

691 Ulrich, W., Almeida-Neto, M., & Gotelli, N. J. (2009). A consumer's guide to nestedness  
692 analysis. *Oikos*, 118(1), 3–17. <https://doi.org/10.1111/j.1600-0706.2008.17053.x>

693 Valente, L., Phillimore, A. B., & Etienne, R. S. (2018). Using molecular phylogenies in island  
694 biogeography: it's about time. *Ecography*, 41(10), 1684–1686.  
695 <https://doi.org/10.1111/ecog.03503>

696 Veitch, C. R., & Wade, L. (2019). Hauturu: the history, flora and fauna of Te Hauturu-o-Toi Little  
697 Barrier Island. Massey University Press.

698 Walentowitz, A., Troiano, C., Christiansen, J. B., Steinbauer, M. J., & Barfod, A. S. (2022). Plant  
699 dispersal characteristics shape the relationship of diversity with area and isolation.  
700 *Journal of Biogeography*, 49(9), 1599–1608. <https://doi.org/10.1111/jbi.14454>

701 Wang, Y., Ding, P., Chen, S., & Zheng, G. (2013). Nestedness of bird assemblages on urban  
702 woodlots: Implications for conservation. *Landscape and Urban Planning*, 111, 59–67.  
703 <https://doi.org/10.1016/j.landurbplan.2012.11.008>

704 Wardle, D. A., Bellingham, P. J., Bonner, K. I., & Mulder, C. P. H. (2009). Indirect effects of  
705 invasive predators on litter decomposition and nutrient resorption on seabird-dominated  
706 islands. *Ecology*, 90(2), 452–464. <https://doi.org/10.1890/08-0097.1>

707 Weigelt, P., & Kreft, H. (2013). Quantifying island isolation - insights from global patterns of  
708 insular plant species richness. *Ecography*, 36(4), 417–429.  
709 <https://doi.org/10.1111/j.1600-0587.2012.07669.x>

710 Whittaker, R. J., Rigal, F., Borges, P. A. V., Cardoso, P., Terzopoulou, S., Casanoves, F., ...  
711 Triantis, K. A. (2014). Functional biogeography of oceanic islands and the scaling of

712 functional diversity in the Azores. *Proceedings of the National Academy of Sciences*,  
713 111(38), 13709–13714. <https://doi.org/10.1073/pnas.1218036111>

714 Wilmshurst, J. M., Moar, N. T., Wood, J. R., Bellingham, P. J., Findlater, A. M., Robinson, J. J.,  
715 & Stone, C. (2014). Use of pollen and ancient DNA as conservation baselines for  
716 offshore islands in New Zealand. *Conservation Biology*, 28(1), 202–212.  
717 <https://doi.org/10.1111/cobi.12150>

718 Wright, D. H., Patterson, B. D., Mikkelsen, G. M., Cutler, A., & Atmar, W. (1998). A comparative  
719 analysis of nested subset patterns of species composition. *Oecologia*, 113(1), 1–20.  
720 <https://doi.org/10.1007/s004420050348>

721 Wright, D. H., & Reeves, J. H. (1992). On the meaning and measurement of nestedness of  
722 species assemblages. *Oecologia*, 92(3), 416–428. <https://doi.org/10.1007/BF00317469>

723 Zhan, C., Li, B., Chen, C., & Wang, Y. (2024). Taxonomic, phylogenetic, and functional  
724 nestedness of mammal assemblages in the Zhoushan Archipelago, China. *Current*  
725 *Zoology*, 70, 728–738. <https://doi.org/10.1093/cz/zoae006>

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728 **Appendix 1** - Full list of sources consulted to collect data for plant height  
729  
730  
731 *Edgar E, Connor HE 2010 Flora of New Zealand Volume V Grasses. 2nd edition. Manaaki*  
732 *Whenua Press, Lincoln.*  
733 *Heenan PB 2017 A taxonomic revision of Cardamine L. (Brassicaceae) in New Zealand.*  
734 *Phytotaxa 330, 1–154.*  
735 *Heenan PB, Mitchell AD, De Lange PJ 2004 Arthropodium bifurcatum (Asparagaceae), a new*  
736 *species from northern New Zealand. New Zealand Journal of Botany 42, 233–246.*  
737 *McGlone MS, Richardson SJ 2023. Sexual systems in the New Zealand angiosperm flora. New*  
738 *Zealand Journal of Botany 61, 201–231.*  
739 *Merrett MF, Clarkson BD 2000 Reinstatement of Alseuosmia quercifolia (Alseuosmiaceae) from*  
740 *New Zealand. New Zealand Journal of Botany 38, 153–164.*  
741 *Raven PH, Raven TE 1976 The Genus Epilobium (Onagraceae) in Australasia: A Systematic*  
742 *and Evolutionary Study. New Zealand Department of Scientific and Industrial Research Bulletin*  
743 *216: 1–321.*  
744 *Allan HH 1961 Flora of New Zealand Volume 1 (Vol. 1). Manaaki Whenua Press.*  
745 *Brownsey PJ 1977a A taxonomic revision of the New Zealand species of Asplenium. New*  
746 *Zealand Journal of Botany 15: 39–86.*  
747 *Brownsey PJ 1977b Asplenium hybrids in the New Zealand flora. New Zealand Journal of*  
748 *Botany 15: 601–637.*  
749 *Brownsey P, Perrie L 2014 Lygodiaceae. In: Breitwieser, I.; Heenan, P.B.; Wilton, A.D. Flora of*  
750 *New Zealand—Ferns and Lycophytes. Fascicle 2. Manaaki Whenua Press, Lincoln.*  
751 *<https://doi.org/10.7931/J26Q1V5T>*  
752 *Brownsey P, Perrie L (2015a). Loxsomataceae. In: Breitwieser, I.; Heenan, P.B.; Wilton, A.D.*  
753 *Flora of New Zealand—Ferns and Lycophytes. Fascicle 6. Manaaki Whenua Press, Lincoln.*  
754 *<https://doi.org/10.7931/B1WC7J>*

755 Brownsey P, Perrie L 2015b. Psilotaceae. In: Breitwieser, I.; Heenan, P.B.; Wilton, A.D. *Flora of*  
756 *New Zealand—Ferns and Lycophytes. Fascicle 6. Manaaki Whenua Press, Lincoln.*  
757 <https://doi.org/10.7931/B1MW2B>

758 Brownsey P, Perrie L 2017 Lindsaeaceae. In: Breitwieser, I.; Wilton, A.D. *Flora of New*  
759 *Zealand—Ferns and Lycophytes. Fascicle 17. Manaaki Whenua Press, Lincoln.*  
760 <https://doi.org/10.7931/B1D59W>

761 Brownsey P, Perrie L 2021a Dryopteridaceae. In: Breitwieser, I.; Heenan, P.B.; Wilton, A.D.  
762 *Flora of New Zealand—Ferns and Lycophytes. Fascicle 31. Manaaki Whenua Press, Lincoln.*  
763 <https://doi.org/10.7931/5SNH-WV42>

764 Brownsey P, Perrie L 2021b Pteridaceae. In: Breitwieser, I.; Heenan, P.B.; Wilton, A.D. *Flora of*  
765 *New Zealand—Ferns and Lycophytes. Fascicle 30. Manaaki Whenua Press, Lincoln.*  
766 <https://doi.org/10.7931/DTKJ-X078>

767 Brownsey PJ, Smith-Dodsworth JC 2000 *New Zealand ferns and allied plants. Revised edition.*  
768 *David Bateman, Auckland.*

769 Carter SN, Miller S, Meyer SJ, Gemmill CEC 2018 A New Species of *Pittosporum* Described  
770 *from the Poor Knights Islands, Northland, Aotearoa/New Zealand. Systematic Botany, 43: 633–*  
771 *643. https://doi.org/10.1600/036364418X697355*

772 Garnock-Jones PJ 2023 *Veronica. Flora of New Zealand—Seed Plants. Fascicle 9.*  
773 *Manaaki Whenua - Landcare Research 2005 Ecological traits of New Zealand Flora.*  
774 <https://ecotraits.landcareresearch.co.nz/SearchForm.aspx>

775 Moore LB, Edgar E 1970 *Flora of New Zealand Volume II. Manaaki Whenua Press.*

776 *New Zealand Plant Conservation Network 2024 New Zealand Plant Conservation Network.*  
777 <https://www.nzpcn.org.nz/>

778 *Royal Botanic Gardens and Domain Trust, Sydney 2024 PlantNET (The NSW Plant Information*  
779 *Network System). https://plantnet.rbgsyd.nsw.gov.au*

780 *Royal Botanic Gardens Victoria 2024 VicFlora—Flora of Victoria. https://vicflora.rbg.vic.gov.au/*

781 *Webb CJ, Sykes WR, Garnock-Jones PJ 1988 Flora of New Zealand Volume IV. Manaaki*

782 *Whenua Press.*

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786 **Appendix 2** - Supplementary analyses

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788 **Table S1** - Spearman partial correlation coefficients for each plant category and island  
 789 characteristics. Columns represent rho and p-values for each variable. In bold are significant  
 790 relationships. A Bonferroni correction was applied to reduce type 1 error rates. The alpha level  
 791 was set at 0.005, dividing 0.05 by the number of categories (n = 10).

Category		Area (km <sup>2</sup> )		Isolation (%)		Exposure (°)	
		Rho	p	Rho	p	Rho	p
Phylogenetic - Fine	All species	<b>-0.777</b>	<b>&lt;0.0001</b>	<b>0.203</b>	<b>0.0010</b>	<b>0.229</b>	<b>&lt;0.0001</b>
Phylogenetic - Coarse	Monocots	<b>-0.740</b>	<b>&lt;0.0001</b>	<b>0.249</b>	<b>&lt;0.0001</b>	<b>0.190</b>	<b>0.0020</b>
	Dicots	<b>-0.768</b>	<b>&lt;0.0001</b>	0.166	0.0069	<b>0.191</b>	<b>0.0020</b>
Functional - Growth forms	Graminoids	<b>-0.713</b>	<b>&lt;0.0001</b>	<b>0.282</b>	<b>&lt;0.0001</b>	0.154	0.0124
	Forbs	<b>-0.755</b>	<b>&lt;0.0001</b>	0.134	0.0302	<b>0.252</b>	<b>&lt;0.0001</b>
Functional - Dispersal modes	Unspecialised	<b>-0.723</b>	<b>&lt;0.0001</b>	<b>0.267</b>	<b>&lt;0.0001</b>	0.151	0.0141
	Wind	<b>-0.762</b>	<b>&lt;0.0001</b>	<b>0.218</b>	<b>0.0004</b>	<b>0.287</b>	<b>&lt;0.0001</b>
Plant height	Short	<b>-0.731</b>	<b>&lt;0.0001</b>	<b>0.208</b>	<b>0.0007</b>	<b>0.182</b>	<b>0.0031</b>
	Medium	<b>-0.786</b>	<b>&lt;0.0001</b>	<b>0.174</b>	<b>0.0047</b>	<b>0.254</b>	<b>&lt;0.0001</b>
	Tall	<b>-0.755</b>	<b>&lt;0.0001</b>	0.170	0.0058	<b>0.201</b>	<b>0.0011</b>

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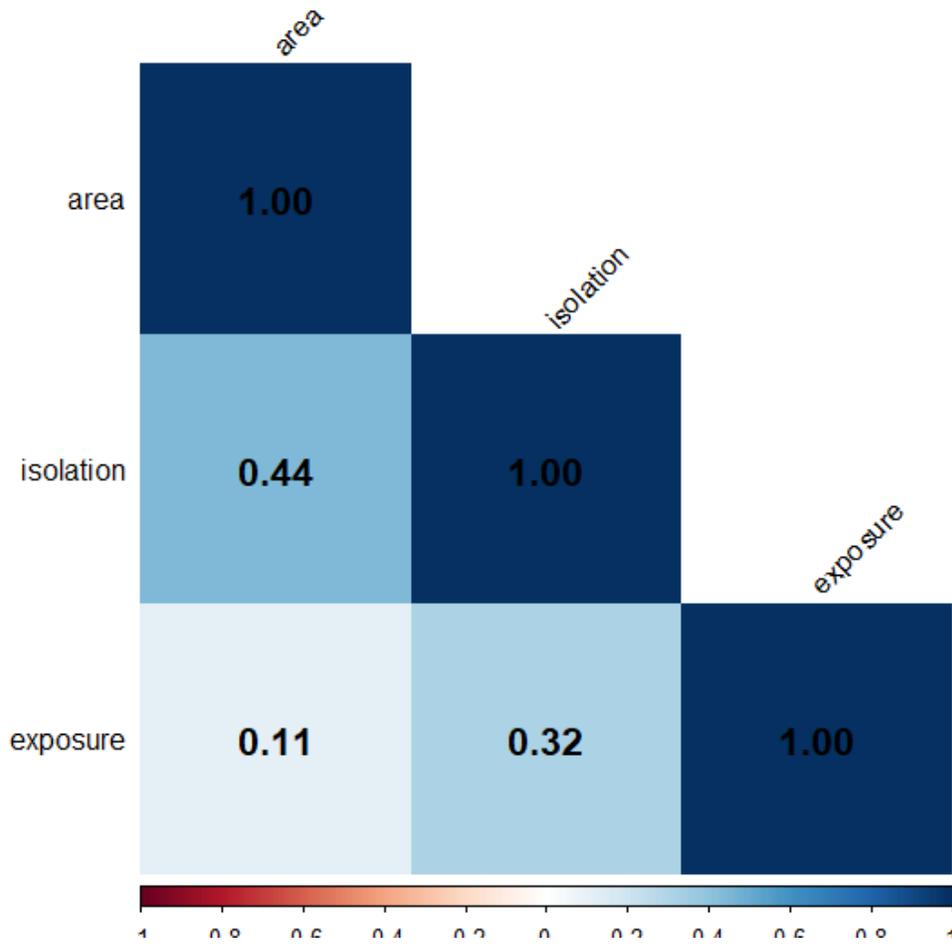
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798 **Figure S1** Correlation matrix for the island characteristics used in the models.  
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