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Bridging data silos to holistically model plant macrophenology

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The data used to create graphs from Box 1 are openly available in Environmental Data Initiative (EDI) at [http://doi.org/\[doi in progress\]](http://doi.org/[doi in progress]), reference number [reference number in progress]. Additionally, the data derived in this article are available from USA-National Phenology Network at <http://doi.org/10.5066/F78S4N1V>, National Ecological Observatory Network at <https://www.neonscience.org/data>, Dryad at <https://datadryad.org/stash>, and EDI at <https://edirepository.org/>. These data were derived from the following resources available in the public domain:

Switzer J, Chamberlain S, Marsh L, Wong K (2024). `_rnpn`: Interface to the National 'Phenology' Network 'API'. R package version 1.2.8.0, <<https://CRAN.R-project.org/package=rnpn>>.

NEON (National Ecological Observatory Network). 2024. Plant phenology observations, DP1.10055.001 (RELEASE-2023). 2013-2021 for Region: Contiguous United States. Dataset accessed May 2023 via the USA National Phenology Network at <http://doi.org/10.5066/F78S4N1V>.

Park, Isaac et al. (2023). Herbarium-Derived Phenological Data in North America [Dataset]. Dryad. <https://doi.org/10.25349/D9WP6S>. Accessed Feb 2023

Park, D., A. Williams, E. Law, A. Ellison, and C. Davis. 2023. Assessing Plant Phenological Patterns in the Eastern United States Over the Last 120 Years ver 5. Environmental Data Initiative. <https://doi.org/10.6073/pasta/bfb70a1701ef23f686fcc73840e6ae17> (Accessed 2023-08).

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3 **Bridging data silos to holistically model plant macrophenology**
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29 Summary

- 30 ● Phenological shifts to global climate change impact ecosystem functions. There are
31 various data sources from which spatiotemporal, and taxonomic phenological data may
32 be obtained: mobilized herbaria, community-science initiatives, observatory networks,
33 and remote-sensing. However, analyses conducted to date have generally relied on single
34 sources of data, thus treating alternative data sources as isolated silos.
- 35 ● Siloed treatment of data in analyses may be due to the lack of harmonization across
36 different data sources, that offer partially non-overlapping information and are often
37 complementary. Such treatment precludes a deeper understanding of phenological
38 responses at macroecological scales. Here, we describe data harmonization as the direct
39 integration of disparate sources of phenological data using a common schema.
- 40 ● We highlight existing methods for data harmonization that can be applied to phenological
41 data: data-design patterns, metadata standards, and ontologies. We describe how
42 harmonized data from multiple sources can be integrated into analyses using existing
43 methods and discuss the use of automated extraction techniques.
- 44 ● Data harmonization is not a new concept in ecology but the harmonization of
45 phenological data is long overdue. We aim to highlight the need for better data
46 harmonization providing a roadmap for how harmonized phenological data may fill data
47 gaps while simultaneously integrated into analyses.

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49 **Keywords:** Data harmonization, Data management, Ontologies, Scales, SDMs

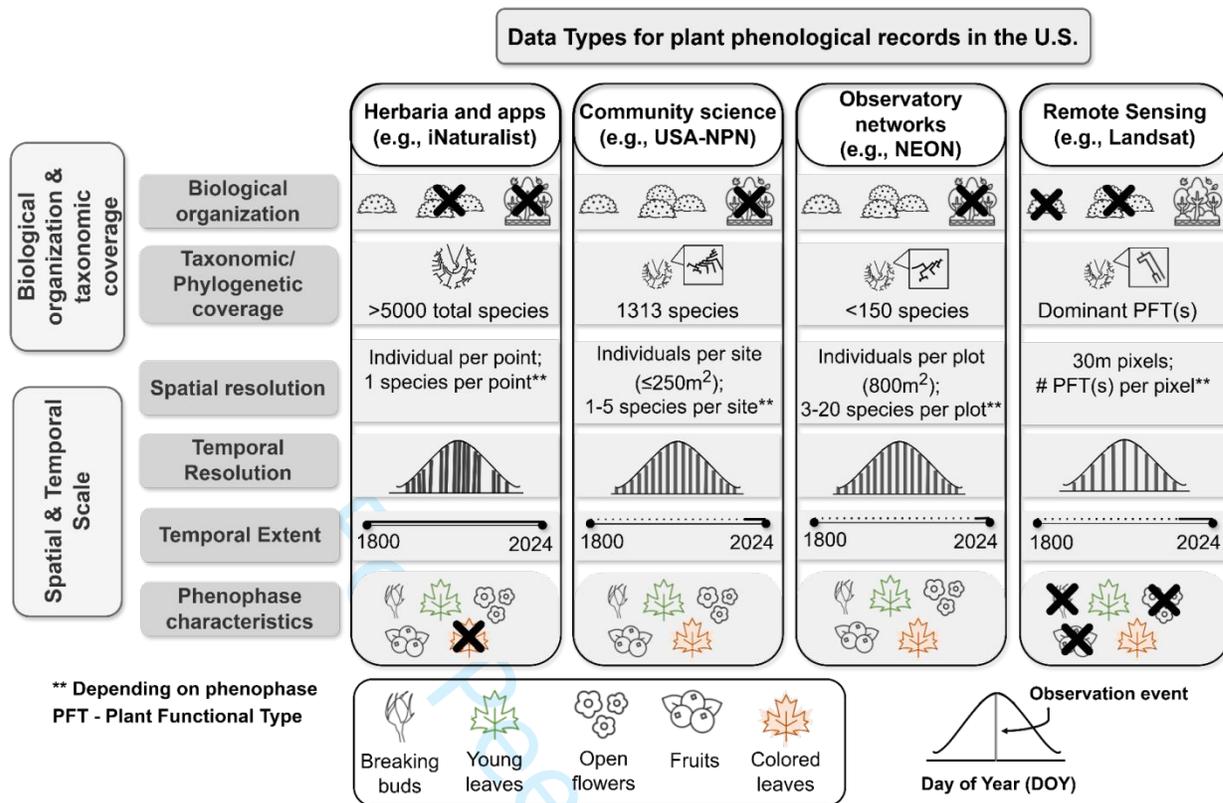
50 Introduction

51 Many biological interactions depend on phenological patterns that reflect ecological and
52 evolutionary responses to climatic conditions (e.g., Chmura *et al.*, 2019). For example, plant
53 phenology — the recurring seasonal timing of leaf-out, flowering, fruiting, and leaf senescence
54 — is a key set of genetically- and environmentally-controlled traits that are central to plant
55 reproduction, plant-pollinator interactions, and availability of resources to herbivores. Plant
56 phenology is also linked directly to ecosystem processes and services relevant to human society,
57 such as carbon sequestration, seasonal allergies, and food security (e.g., Fatima *et al.*, 2020; Gray

58 & Ewers, 2021; Cope *et al.*, 2022). Because plant phenology is very sensitive to ongoing rapid
59 environmental change, there is an urgent need to better quantify and predict plant phenological
60 dynamics (e.g., Gallinat *et al.*, 2021; Park *et al.*, 2021).

61 A major challenge to quantifying and predicting phenology lies in its scale-dependence
62 (D. S. Park *et al.*, 2021). Like many other ecological phenomena, inferences made about
63 phenology depend on how data are combined across space, time, or taxa (Levin, 1992; D. S. Park
64 *et al.*, 2021). Furthermore, environmental drivers of phenology (e.g., temperature, precipitation,
65 solar insolation) can vary across space and time (e.g., Chamberlain & Wolkovich, 2023), and
66 plastic organismal responses to these drivers can differ among individuals, populations, species,
67 and communities (Ramirez-Parada *et al.*, 2024). Studying the effects of phenology on ecological
68 processes at global scales is the objective of a growing field known as macrophenology; such
69 inferences differ qualitatively from studies focused on more local scales (Doi *et al.* 2017;
70 Gallinat *et al.* 2021). Understanding and analyzing this scale-dependence requires
71 macrophenological data that span scales of space, time, taxonomy, and levels of biological
72 organization.

73 Data sources at each of these scales do exist although they rarely have been analyzed
74 simultaneously. For plant phenological data, these include herbarium specimens, community
75 science initiatives, observatory networks, and remote-sensing platforms (e.g., Gray & Ewers,
76 2021; Reyes-González *et al.*, 2021; Davis *et al.*, 2022). Different data types capture disparate
77 scales as a result of their sampling design and effort. For instance, remote sensing may provide
78 continuous monitoring over a long period of time, whereas observatory networks may provide
79 periodic sampling with field surveys that vary in their temporal extent (Fig. 1). These scale
80 mismatches often hinder data harmonization—the direct integration of disparate data types under
81 a common schema. The Lack of harmonization across datasets limits our ability to assess
82 phenological responses to climate at various scales.



83

84 **Figure 1.** Data aspects across phenological data types with example datasets. Each data type
 85 offers a distinct level of biological, taxonomic, spatial, temporal, and phenophase information.
 86 Each level contains example-specific information comparing the similarities and differences
 87 across data types. Differences between data types (across columns) highlight gaps where data
 88 harmonization would benefit and maximize coverage within each level. We use a bell-curve to
 89 exemplify the sampling frequency within the duration of a phenophase. Young leaves also
 90 include fully opened green leaves.

91

92 We assert that data harmonization is critical for improving our understanding of the
 93 impacts of climatic change on macrophenology (Gallinat *et al.*, 2021). In this Viewpoint, we
 94 focus on plant phenological datasets collected predominantly in the United States, although
 95 similar data have been collected at sites worldwide. We explore the characteristics of these data
 96 and reveal the unintentional data silos that limit our ability to answer a range of important
 97 ecological and evolutionary questions about phenology. Additionally, we identify how bringing
 98 together multiple data sources will enable us to answer new questions. To move toward a
 99 common goal of phenological data harmonization, we provide a road map describing methods

100 for harmonization, how harmonization can help to fill gaps in phenological data across space and
101 time, and methods for integrating harmonized data into analyses. We end with a call for
102 harmonization of phenological data to rapidly advance phenological research.

103 **A multiplicity of data sources with different strengths and weaknesses**

104 Phenological data have provided invaluable insights into the varying effects of changing
105 climate on the timing of phenology (D. Li *et al.*, 2019; Zohner *et al.*, 2023), and recent papers
106 highlight potential new insights to be made from each independent data type (Davis *et al.*, 2022;
107 Dronova & Taddeo, 2022; Binley & Bennett, 2023; Zhu & Song, 2023). Because of gaps within
108 any single data type, we cannot get a complete picture of phenology without integrating across
109 data types. In this section, we draw attention to the strengths, weaknesses, and gaps of each data
110 type with respect to space, time, taxonomy, life history, and levels of biological organization.

111 *Herbarium Specimens*

112 Herbarium specimens provide irreplaceable historical climate baselines and they have
113 been mobilized to study phenological sensitivity to local climatic conditions and climatic change
114 (e.g., Chmura *et al.*, 2019; Lenters *et al.*, 2021; Davis, 2023) at broad spatial scales (e.g., Davis
115 *et al.*, 2015; Willis *et al.*, 2017; Zhu & Song, 2023). Many of the large herbarium collections in
116 the U.S. have increased the accessibility of phenology data contained in these specimens through
117 massive efforts to digitize millions of physical specimens and the information contained in their
118 labels with the centralization of data into repositories (e.g., the Global Biodiversity Information
119 Facility [GBIF], Southwestern Environmental Information Network [SEINet]) (GBIF; SEINet
120 Portal Network, 2023; Hedrick *et al.*, 2020). However, smaller collections or those from
121 countries with less digital infrastructure in place are less accessible or less frequently digitized,
122 leading to biogeographical biases in spatial coverage (Daru *et al.* 2018; Davis *et al.* 2022).
123 Despite the large taxonomic coverage at the species level, herbarium specimens provide
124 relatively coarse phenophase information, represent single “snapshots” of phenology in space
125 and time, and may exhibit sampling biases that make it unclear whether a specimen represents an
126 early, median, or late observation relative to its source population (Daru *et al.*, 2018; Ramirez-
127 Parada *et al.*, 2022; I. W. Park *et al.*, 2024; Fig. 1). Moreover, as specimens represent single

128 observations of individuals distributed widely in space and time, variation in phenology among
129 specimens represents both within- and among-population differences. Thus, identifying the level
130 of ecological organization associated with relationships between phenology and environmental
131 variables—and the mechanisms underlying such relationships—requires careful statistical design
132 and interpretation of results (Ramirez-Parada *et al.* 2024). Another limitation of herbarium-
133 derived data for use in phenological studies is the difficulty in identifying dates of occurrence for
134 phenological phases other than flowering and fruiting for most species. Additionally, the patchy
135 temporal and spatial coverage limits their use at global or local scales and depends on the
136 sampling biases (Daru *et al.* 2018).

137 *Community-science initiatives*

138 Community-science initiatives harness the power of volunteers to record phenological
139 data across broad spatial extents while providing high-resolution phenophase information from a
140 variety of taxa (Reyes-González *et al.*, 2021; Domingo-Marimon *et al.*, 2022). Such initiatives
141 vary in the degree of standardization used in data collection. For instance, image contributors for
142 community-sourced app-based records (e.g., iNaturalist) do not follow specific protocols for
143 capturing phenophases. In contrast, the USA National Phenology Network (USA-NPN) is an
144 example of a community-science initiative with volunteer engagement across the country
145 through their Nature's Notebook platform (Crimmins *et al.*, 2017; <https://www.usanpn.org/>). It
146 has a standardized protocol to facilitate repeated observations of specific individuals (or patches)
147 at a chosen site (Crimmins *et al.*, 2017; Fig. 1). These data provide estimates of date onset,
148 termination, and duration of multiple phenophases at high temporal resolution and national
149 coverage (Fig. 1). However, while these data encompass observations for thousands of species,
150 most correspond to a narrow set of indicator species for which specific observational protocols
151 have been developed. Furthermore, investment by volunteers leads to large variations in
152 taxonomic coverage and duration of observations (e.g., a single versus multiple years). These
153 data often contain observation bias and inconsistencies in protocol implementation that can limit
154 their application (Reyes-González *et al.*, 2021; Domingo-Marimon *et al.*, 2022). In some cases,
155 participants might record the phenological status of only one individual at one site many times
156 per year (sometimes less), but they might not sample multiple individuals at a given site, thus
157 greatly limiting population-level inferences.

158 *Observatory Networks*

159 Observatory networks provide systematic, long-term field data that follow individuals
160 throughout their phenological cycle (Gallinat *et al.*, 2021), thus providing opportunities to
161 quantify inter- and intraspecific variation in phenology across ecoregions. The U.S. National
162 Science Foundation's National Ecological Observatory Network (NEON) is one such long-term,
163 ecological monitoring network designed to collect data through 2049 (Elmendorf *et al.*, 2016).
164 NEON works closely with the USA-NPN and Phenocam networks to collect data and has
165 adopted their standards and protocols (Richardson *et al.*, 2007). This offers an exciting example
166 of how different phenological monitoring systems can coordinate efforts for standardized
167 observations (Richardson *et al.*, 2007). Similar to NPN, NEON provides high-resolution
168 phenological information through repeated measures in their field sampling design (Fig. 1).
169 NEON also records co-located information on a suite of other biological and physical variables
170 relevant to phenology (e.g., beetle pollinator abundance, climatic variables, carbon dioxide flux)
171 (Nagy *et al.*, 2021). Despite the continental scale and projected 30-year lifespan scale of NEON,
172 their NPN-style field observations are limited in i.) taxonomic coverage due to resource
173 constraints; ii.) spatial coverage as data are collected only at the several dozen established NEON
174 sites; and iii.) current temporal range due to the relatively recent establishment of the network in
175 2019 (Fig. 1). The Long-term Ecological Research Network has some sites that collect
176 phenological information, but these data are not collected with a standardized protocol and
177 synthesis of them is challenging (Mulder *et al.*, 2021; Schulze, 2023; but see Keenan *et al.*,
178 2014). Moving forward, NEON promises to be an irreplaceable long-term reference for fine-
179 resolution phenological data compatible with all data amassed by the USA-NPN through their
180 active partnership and delivering NEON data through the USA-NPN portal.

181 *Remote Sensing*

182 Remotely-sensed data are capable of capturing continental and interannual changes in
183 phenology, yet the spatial resolution of the data is often too coarse to discern phenological
184 changes at the species-level (Gallinat *et al.*, 2021; Reyes-González *et al.*, 2021). Satellites, near-
185 Earth imagery, and biologging devices have allowed for spatially continuous observations with
186 improving temporal resolution (Zarnetske *et al.*, 2019; Lechner *et al.*, 2020; Latifi *et al.*, 2023).

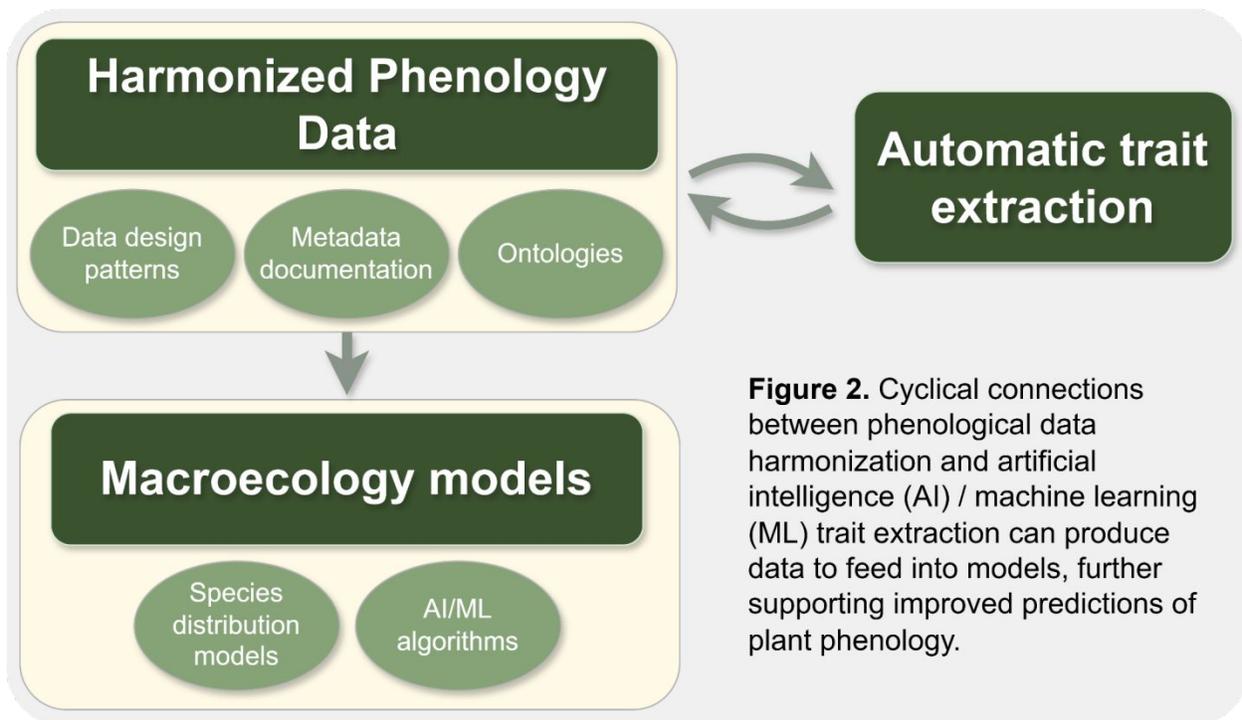
187 Remote-sensing techniques related to phenology have been applied successfully to detect the
188 start and end of the growing season (i.e., leaf out and leaf off) for either dominant tree taxa or
189 functional types, and primarily in temperate deciduous and tropical dry forests (Dronova &
190 Taddeo, 2022). Recent applications of deep-learning algorithms to high-resolution hyperspectral
191 and red-green-blue images (1-m and 0.25-m, respectively) from near-Earth (i.e., airborne) remote
192 sensing of NEON sites enable the segmentation and identification of individual tree crowns
193 (Weinstein *et al.*, 2024), paving the way for the detection of individual tree-crown phenology
194 from frequent near-Earth image acquisition (e.g., by drones). Furthermore, the National
195 Aeronautics and Space Administration's Surface Biology and Geology High Frequency Time
196 Series (SHIFT) near-Earth remote-sensing campaign in 2023 enabled the detection of
197 superblooms in the grasslands of coastal California from weekly flyovers (Angel *et al.*, 2024).
198 Although hyperspectral sensors have broadened the possibilities of remotely-sensed phenological
199 monitoring, such efforts remain limited to specific sites (e.g., NEON sites) or campaigns with
200 high spatial and temporal resolution. Another limitation is that the earliest remote sensing data
201 are limited to the 1970's (i.e. Landsat 1 products) and do not provide substantial pre-global
202 warming information comparable to point-based herbarium data.

203 **Moving forward: bridging data silos in macrophenology**

204 *Data Harmonization*

205 Data harmonization is not a new concept in ecology. For decades there has been
206 tremendous interest across the scientific community in pooling and harmonizing plant trait data.
207 For instance, the TRY database has excelled in aggregating trait data and supported extensive
208 advances in trait-based plant ecology, but lacks a common format that limits compatibility
209 between data sets (Kattge *et al.*, 2011, 2020). More recently, ecologists have recognized the
210 importance of considering intraspecific trait variation, emphasizing the coordination of open
211 science efforts around individual-level trait information (Violle *et al.*, 2012; Cope *et al.*, 2022).
212 Except by remote-sensing networks, phenological data are collected from individual organisms
213 and allow for exploration of intraspecific trait variation. This makes phenological data an
214 excellent test bed for developing and testing approaches for data harmonization of individual-
215 level traits. Various approaches exist for harmonizing ecological data that could be applied to

216 phenological data. Many of these approaches incorporate common terminology and structures
 217 (i.e., design patterns) representing relational tables tracking organismal information (e.g.,
 218 taxonomy, measurement [i.e., trait, number of individuals]) and other important metadata (e.g,
 219 geographic locations and differences in sampling methodologies (O'Brien *et al.*, 2021; Keller *et*
 220 *al.*, 2023)(Fig. 2).



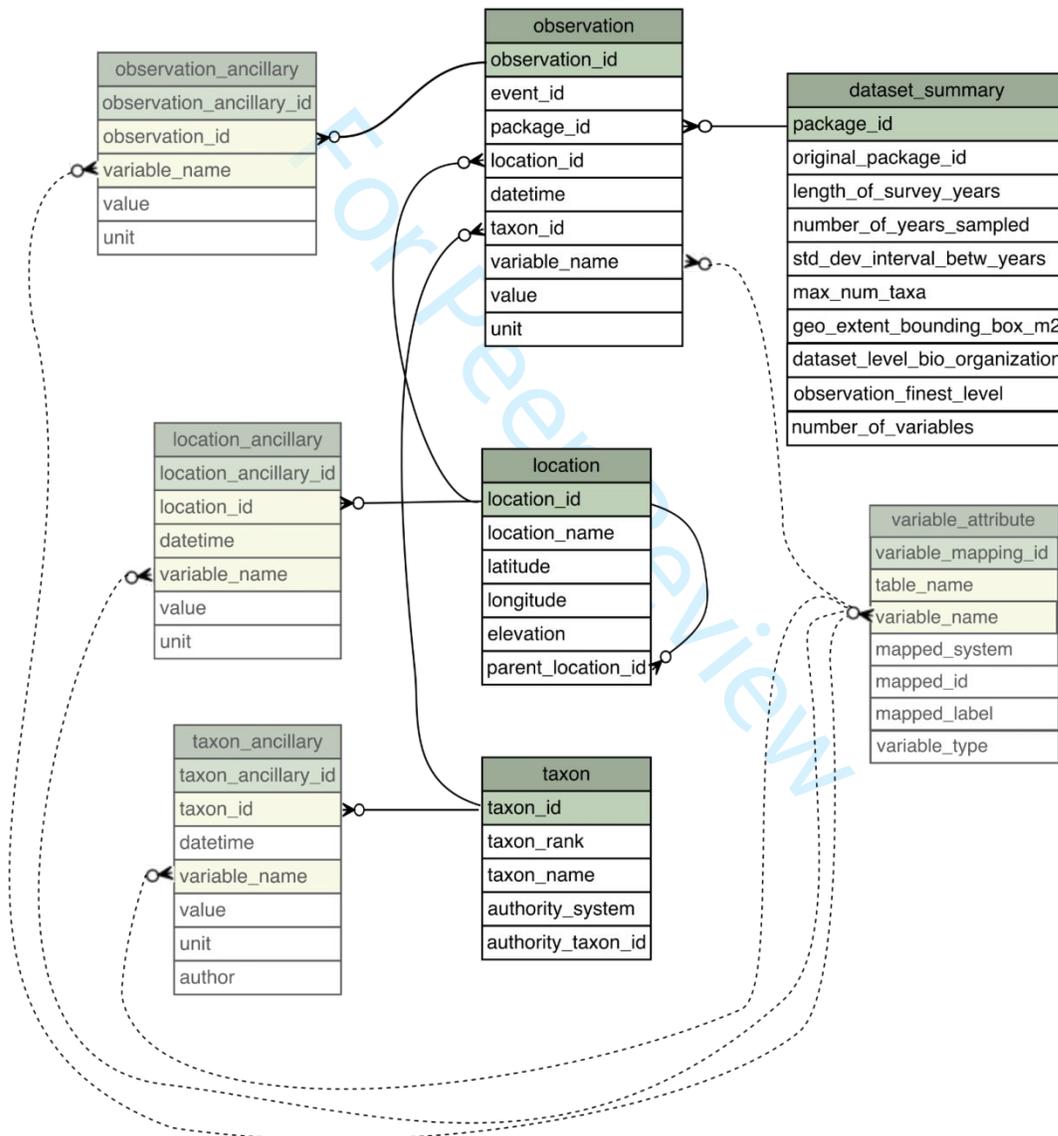
221
 222 **Figure 2.** Cyclical connections between phenological data harmonization and artificial
 223 intelligence (AI) / machine learning (ML) trait extraction can produce data to feed into
 224 models, further supporting improved predictions of plant phenology.

225
 226 Using common terminology and notation is a key aspect of harmonizing phenological
 227 data to make it easier for researchers (and algorithms) to discover and use data. Ontologies
 228 provide a structured, formal language for the standardization of terminology and concepts related
 229 to data management (see Stucky *et al.*, 2018; Schneider *et al.*, 2019; O'Brien *et al.*, 2021;
 230 Lenters *et al.*, 2021; Dumschott *et al.*, 2023; Keller *et al.*, 2023). To our knowledge, the most
 231 well-developed catalog for ontologies in plant phenology was made available by Stucky *et al.*
 232 (2018), who assembled a robust framework of aggregated vocabulary from global phenological
 233 records. Contributions towards such efforts are crucial for dispelling uncertainties in naming

234 conventions for phenophases. For example, intensity-based vocabulary may require a minimum
235 percentage of reproductive organs to be displayed as “open flowers” for an individual’s
236 phenophase to be identified as “flowering”, whereas qualitative assessments of flowering status
237 may simply require the presence of a single open flower. The ontology of Stucky *et al.* (2018)
238 could also be combined with more vocabulary pertaining more widely to ecological traits to
239 capture other important information (e.g., the location of observation). The Ecological Trait Data
240 Standard (ETS) is a common vocabulary to facilitate trait data harmonization and is implemented
241 in some trait data integration networks (Open Traits Network; Gallagher *et al.*, 2020). Notably,
242 ETS incorporates terms used by the Darwin Core Standard (DwC), a glossary of terms to
243 facilitate sharing information about biodiversity maintained by the Biodiversity Information
244 Standards Taxonomic Databases Working Group. Common vocabularies propagate a shared
245 understanding of phenological phenomena that lay the groundwork for the harmonization of data
246 into a common structure.

247 A robust data design pattern (i.e., relational structure) will empower the increased
248 integration of harmonized data into derived data products that may make it easier to account for
249 differences in sampling effort or biases in downstream analyses. Although there are various data
250 design patterns for plant traits (e.g., structural traits of palms incorporating the ETS [Lenters *et*
251 *al.* 2021]), there is not a well-adopted data design pattern for plant phenological traits. We
252 propose a phenological trait extension of the Ecological Community Data Design Pattern
253 (ecocomDP), which was developed for harmonizing community ecology biodiversity data
254 (O’Brien *et al.*, 2021). The original ecocomDP model is extended with two features: 1)
255 reconfiguring the table for mapping variables to external dictionaries to allow any variable
256 attribute (e.g., a trait) to be recorded and linked to an external dictionary of concepts, such as the
257 ontologies mentioned above, and 2) adding additional descriptive fields to the dataset summary
258 table (Fig. 3). Because ecocomDP already accommodates community-level analyses, this
259 extension would enable researchers to ask questions across levels of biological organization
260 (e.g., from individuals to populations to communities). Another advantage of incorporating trait
261 data into ecocomDP is that existing NEON and Long Term Ecological Research (LTER)
262 Network data from various taxa are already harmonized with ecocomDP, making it a good
263 candidate for the future incorporation of phenological traits as well as additional individual-level
264 traits. ecocomDP also employs concepts used by ETS and DwC, making data harmonized into its

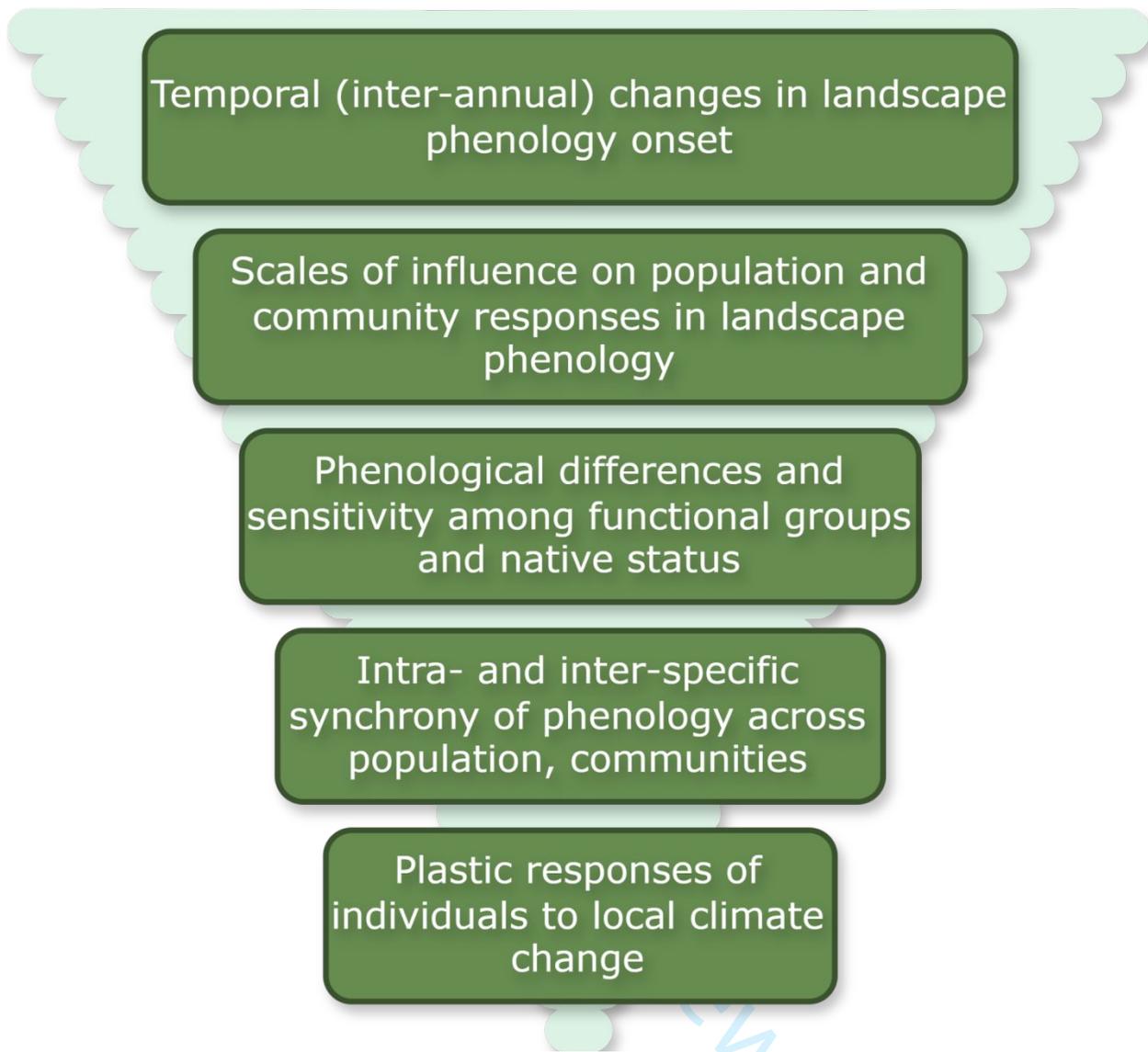
265 structure easily convertible to the DwC-Archive. Finally, ecocomDP strongly emphasizes
 266 metadata, which is essential to ensure that downstream users can determine the relevance of the
 267 data for their study objectives through filtering.



268
 269 **Figure 3.** Schema of the updated ecocomDP data design pattern with the extension to
 270 accommodate traits (O'Brien *et al.*, 2021). Added fields in the dataset summary table allow
 271 users to include the level of biological organization, level of observation, and the number of

272 variables associated with the trait. The variable attribute (previously `variable_mappling`) now
273 includes `variable_type` to indicate the type of trait measured (e.g. start and end dates).

274 With phenological measurements, the protocol for observing the phenophase is a key
275 aspect to consider. For instance, when combining community-science and observatory-network
276 data in analyses, it is incumbent to know the difference between observation methods because
277 each has different levels of observation uncertainty (Binley & Bennett, 2023). Accessible
278 information about sampling design would help compare levels of uncertainty between data types
279 and assess their degree of interoperability for specific research objectives (Fig. 4). Sharing
280 reproducible methods in open platforms such as `protocol.io` (<https://protocols.io>; Accessed June
281 3, 2024) will contribute towards the development of standard disciplinary formats that are useful
282 in metadata curation. Shared protocols between NEON and NPN are a notable example, but even
283 their metadata is found only on site-specific publications or their websites. Furthermore, any
284 protocols used in the collection or curation of herbarium specimens (including their digital
285 records) that are relevant to plant phenology should be included in the metadata or the specimen
286 label. This fits perfectly with the concept of the Global Metaherbarium and the Extended
287 Specimen concept (Davis, 2023). Protocol standards with accessible metadata information and
288 sampling disclosures will be key in supporting data integration and harmonization among data
289 sources while highlighting commonalities and differences in observations (Schneider *et al.*,
290 2019; Dumschott *et al.*, 2023; Keller *et al.*, 2023;).



291
 292 **Figure 4.** Hierarchy of general objectives in plant macrophenology, from broad to specific.
 293 The phenophase event of interest may determine the data types appropriate for achieving the
 294 objective.

295 *Methods for integrating harmonized phenological data into analyses*

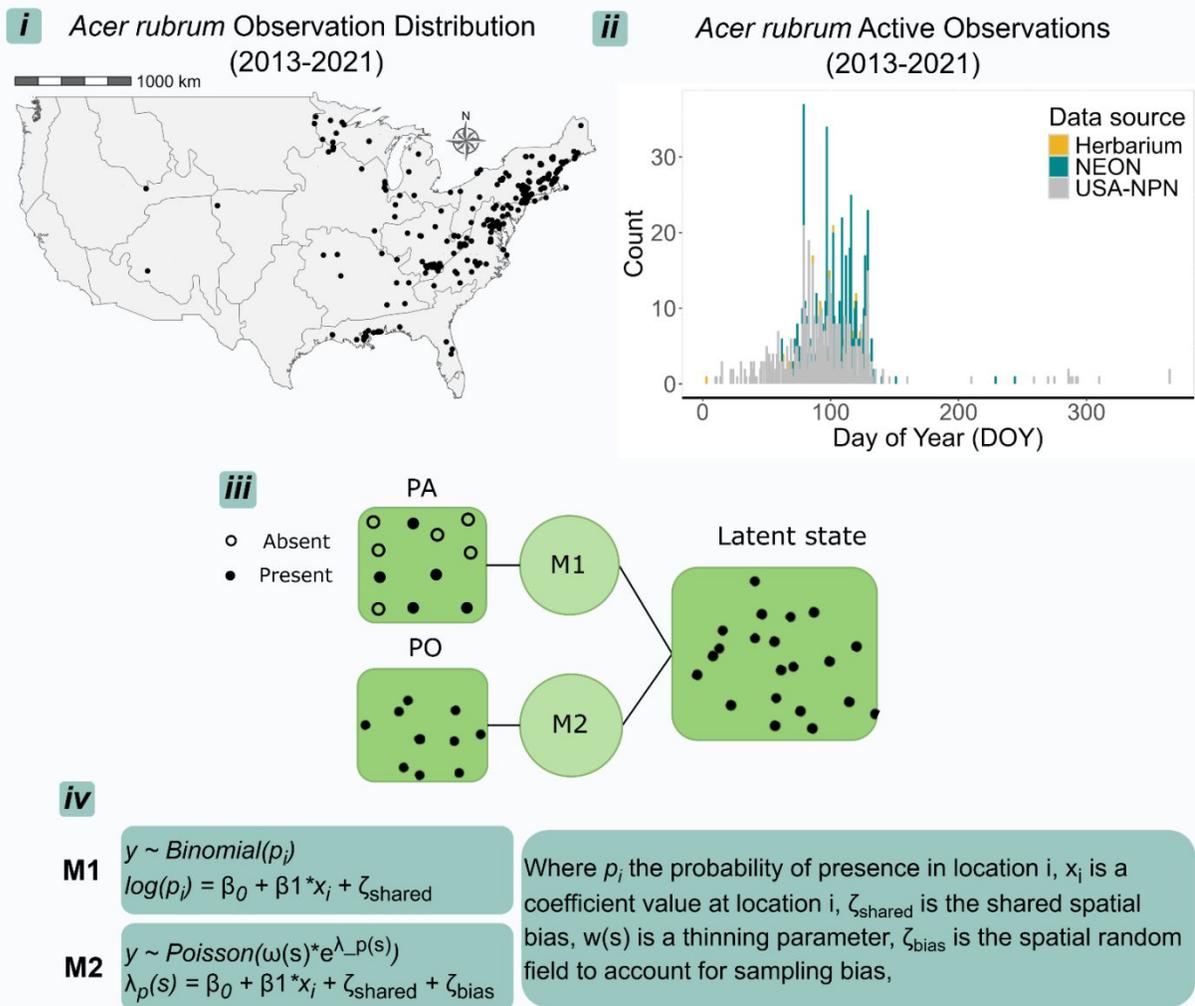
296 When modeling phenology we need to expand our perspective on where we can apply
 297 observations beyond classic phenological models (i.e., location-specific growing degree day
 298 models; Chamberlain & Wolkovich, 2023). With the integration of different data types into
 299 analyses, models must account for underlying biases from different data containing information
 300 on phenology and its drivers across spatial and temporal scales. Here we discuss macroecological

301 approaches to solving two challenges: 1) differences in how phenophases are recorded and 2)
302 spatial and temporal mismatches between phenological, geographic, and climatic data. These
303 hurdles greatly impede efforts of macroscale phenology studies as they limit the geographic
304 scope and questions that may be explored (Gallinat *et al.*, 2021). To address the first challenge of
305 differences in how phenophases are recorded, there are many methods phenologists could adapt
306 from species distribution modeling (SDM). Instead of modeling a response of species occurrence
307 or abundance, we can model the probability of occurrence of a phenophase throughout the year.
308 For instance, such a model could be used to create rasterized forecasts of species-level
309 phenological point observations from herbarium, community science, or observatory network
310 data into a spatially gridded dataset that is compatible with remote sensing data (Peng *et al.*,
311 2024). For example, Yoder *et al.* (2024) used herbarium and community-science data to create
312 gridded predictions of whether Joshua trees are expected to have masted or flowered for each
313 year and location. These rasterized predictions of mast events can then be compared to remotely-
314 sensed gridded data on leaf phenology (i.e., peak greenness).

315 Another approach to account for differences in how phenophases are recorded is
316 occupancy modeling (OM), which is widely used in the field of macroecology to model species
317 distributions and provides a rich methodology by addressing imperfect detection and
318 incorporating geographic location error from specimens for analyses of harmonized phenological
319 data with differences in sampling effort (Erickson & Smith, 2021). One difference in sampling
320 effort presented by phenological data is that some data types only record presence-only (PO)
321 information on phenophases (e.g., herbarium records; iNaturalist), whereas others record both
322 presences and absences (PA) of phenophases (e.g., NPN, NEON). With an OM framework,
323 differences in sampling effort can be accounted for by treating each data type as a designated
324 survey or method to account for differences in detection. Recent advances with integrated SDMs
325 that model PO and PA responses provide a powerful approach for combining data types with
326 differences in sampling effort (Miller *et al.*, 2019; Isaac *et al.*, 2020; Mäkinen *et al.*, 2024).
327 Integrated SDMs could be a powerful way to combine disparate phenological data types (Box 1)
328 and overcome the challenge of accounting for differences in how phenophases are recorded.

Box 1. ISDM roadmap for harmonized models

Understanding phenology at local and global scales is imperative towards connecting multi-scaled issues like climate change. Macrophenology can be better studied through leveraging data harmonization as it provides the potential to fill gaps in coverage across space and time. Phenological data types have different biases that should be considered when using them in a common analysis. For example, different data sources vary in the precision with which phenophases are observed (i.e., presence-only (PO) observations of flowering vs. presence-absence (PA) of flowering). Fortunately, frameworks exist for modeling data with varying levels of precision in detection, such as integrated species distribution models (ISDMs) that allow for both PO and PA responses. To illustrate how ISDMs could be used to simultaneously model different types of phenological data, we present a roadmap to constructing a harmonized model using open flowers of red maples (*Acer rubrum*) using open-source data types that differ in detection of flowering (herbarium [PO] and field observations from NEON and USA-NPN [PA]). To better understand the heterogeneity across data types, it is important to first understand where observations occur in space in time to guide model parameterization (*panel i, ii*). Visualizations of observations for each data source depict disparities in sampling effort across time. The main assumption for an ISDM is that the data observed are modeling the same ecological state where the true distribution is unknown (i.e. latent state; *panel iii*). The common parameters shared between each model address the assumption of observations pertaining to the same ecological state (*panel iv*). The ISDM incorporates shared spatial biases and any known sampling biases into the species-specific models to predict the probability of flowering occurring across space at a point in time (*panel iv*) to produce a probabilistic map of flowering across space at that time.



329

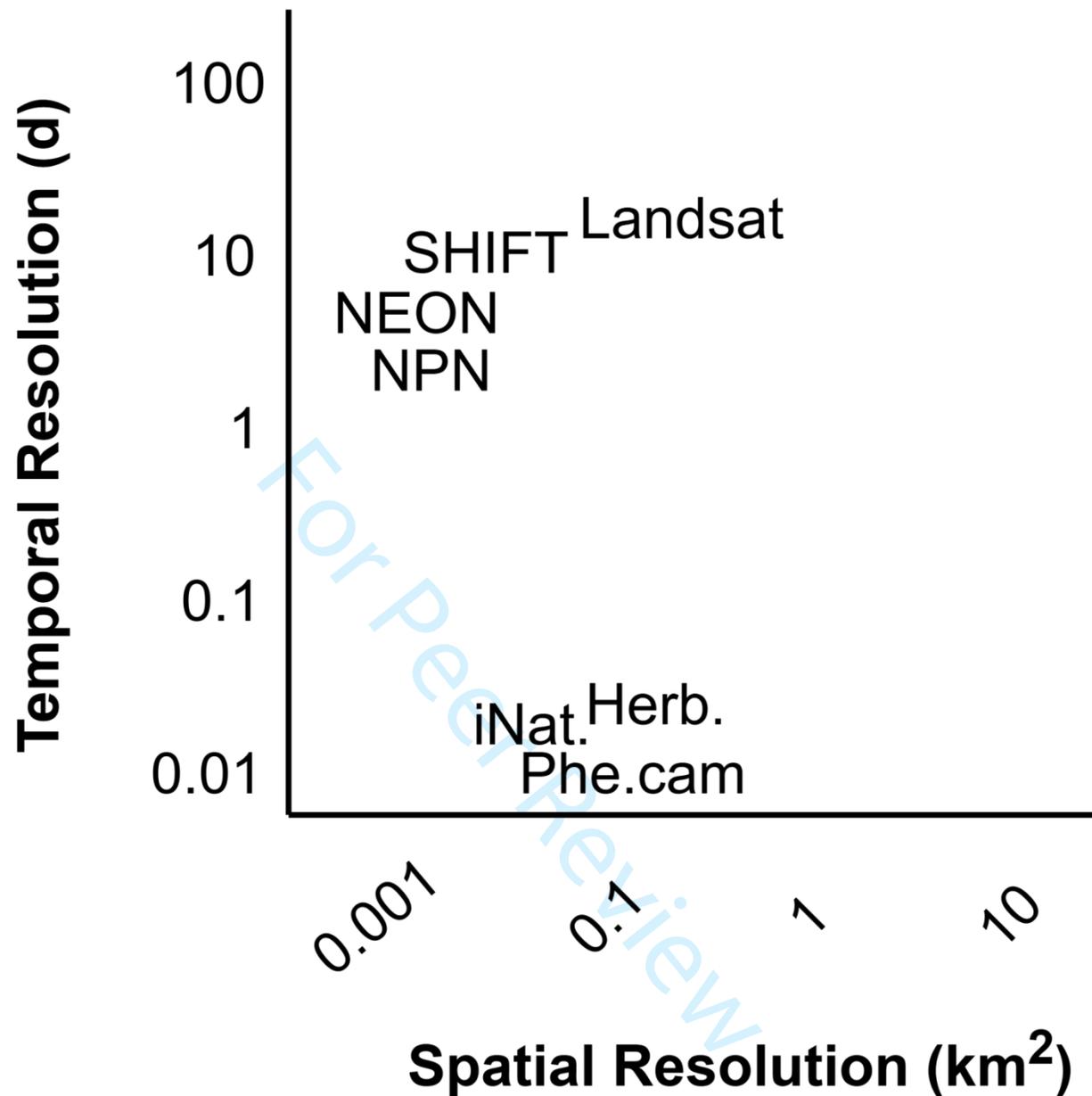
330

331

A second challenge is that phenological data and drivers of phenological responses are often measured at different temporal scales and spatial scales (and therefore, levels of biological

332 organization) (Fig. 5). Automated integration and synthesis tools have begun to be developed to
333 facilitate cross-scale phenological studies. For instance, the Pheno-Synthesis Software Suite
334 (PS3) summarizes ground-based phenological observations into gridded climate and
335 phenological indices (Morisette *et al.*, 2021). One consideration in using such software is
336 understanding what spatial and temporal resolutions and extents have the greatest influence on
337 phenology; to explore the optimal spatial scale between phenological data and its drivers (e.g.,
338 climate, land use topography), different grains (e.g., varying radii around a central phenological
339 observation point or pixel) and extents (e.g., continental, ecoregion, site for NPN or NEON) that
340 are then compared in analyses (Zarnetske *et al.*, 2019; Read *et al.*, 2020; Z. Li, 2022).

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342 **Figure 5.** Temporal and spatial resolution mismatches between phenological attributes.

343 Abbreviations: Herb. - Herbarium, Phe.cam - Phenocam, iNat. - iNaturalist.

344

345 A promising approach for exploring scales of space and time simultaneously is through
 346 interpretable machine learning (ML). Interpretable ML aims to understand what input data
 347 characteristics are most important in driving predictions of output data. Local interpretation with
 348 ML—wherein the prediction of a model for a single observation in space and time is considered,
 349 as opposed to trying to understand the overall predictive behavior of the model across the entire

350 dataset—is especially relevant for exploring spatio-temporal drivers of geo-referenced
351 phenological data. This allows for the visualization and estimation of interactions between
352 location features (i.e., spatial coordinates of phenological data points or grid cells) and other
353 model features (e.g., temperature data represented by different spatial resolutions or temporal
354 lags; Z. Li, 2022). An example of a local interpretation method comes from an extension of the
355 Shapley value in game theory (Shapley 1953), which evaluates how contributions of different
356 players collectively result in a contest's outcome. SHapley Additive exPlanations (SHAP), a
357 recent ML offshoot of Shapley values, quantifies how much each feature collectively contributes
358 to averaged model predictions (Štrumbelj & Kononenko, 2014). Historical phenological data
359 could be used as features in such a model to predict contemporary or future phenological
360 responses. The Shapley value and other local interpretability methods (e.g., Local Interpretable
361 Model-agnostic Explanation or LIME; Ribeiro *et al.*, 2016) offer an exciting new opportunity to
362 simultaneously explore spatiotemporal effects of drivers of harmonized phenological data.
363 Overall, existing modeling approaches from macrosystems ecology and data informatics pose
364 unique solutions to challenges by the integration of phenological data simultaneously into
365 analyses.

366 **Concluding remarks**

367 Predicting plant phenological responses to global change is important given its close ties
368 to ecosystem processes and human health. However, given the scale-dependence of plant
369 phenology, it is difficult to make informed predictions in the absence of data that spans space,
370 time, taxa, and levels of biological organization. Fortunately, such data are at our fingertips
371 through various efforts in recording plant phenology at different scales and with different
372 methods of observation, but the distinct types of phenological data need to be harmonized to
373 unlock their full potential. Efforts to bridge phenological data silos can benefit from successful
374 examples from other subdisciplines in ecology. Approaches to harmonize data can be adopted
375 from existing ecological data design patterns, metadata standards, and ontologies. Biogeographic
376 and macroecological studies offer many solutions for integrating disparate data with unique
377 sampling biases into models. They provide a rich methodology for tackling imperfect detection
378 and incorporating geographic location error from specimens (Erickson & Smith, 2021). Data
379 informatic approaches towards multimodal and big data fusion are other promising tools to

380 automate data extraction and harmonization while improving predictions of plant phenology
381 through pattern detection. Data interoperability is not a new concept in ecology, and
382 phenological data harmonization is long overdue.

383 **Glossary**

384 Data design pattern: A blueprint that captures the essential data characteristics so that a
385 centralized workflow can access, reformat, and structure data (O'Brien *et al.*, 2021).

386 Data harmonization: Direct integration of different plant phenological data categories (e.g.,
387 community science, herbarium, remote sensing) under a common schema

388 Data integration (or data interoperability): Disparate data sources that may be used in tandem
389 and are readably applicable in modeling or management frameworks (Brenskelle *et al.*, 2019;
390 Stucky *et al.*, 2018; Wilkinson *et al.*, 2016).

391 Data management: The organization and handling of data that supports its continuous discovery,
392 evaluation, and reuse (Wilkinson *et al.*, 2016).

393 Machine learning (ML): A subset of methodologies that use algorithms to automate learning
394 predictions about data (e.g., Deep Learning, random forest; “Artificial Intelligence (AI) vs.
395 Machine Learning,” n.d.; Pearson *et al.*, 2020).

396 Occupancy model (OM): A spatially explicit model that determines the occupation of an
397 organism using presence and absence information.

398 Ontology: Standardized vocabulary and a language framework using formal logic that relates
399 terms to concepts and allows for the integration of different data (Madin *et al.*, 2008; Stucky *et*
400 *al.*, 2018).

401 Plant phenology: The timing of recurring life stages (reproductive or growth) of a plant; with a
402 focus on angiosperms.

403 Phenophase: The phenological stage of a plant or animal, that details a particular life cycle stage
404 (e.g., leaf emergence, migration, breeding).

405 Species Distribution Model (SDM): A form of occupancy modeling that predicts species
406 distributions over space based on the attributes of the locations where they are currently known
407 to occur.

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417 **Conflict of Interest Statement**

418 L. G. A. and S. R. are in a working group with Daijiang Li, Kai Zhu, and Tong Qui who may appear as
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420 **Author Contributions**

421 L.G.A., A.M.E, C.C.D. conceived the initial ideas which were further developed and refined
422 with S.J.M., I.W.P., T.R-P., and S.R. L.G.A. designed and developed outlines which were further
423 refined by S.R., C.C.D., A.M.E., I.W.P., S.J.M., and T.R-P. L.G.A. and M.O'B drafted and
424 compiled figures which were further refined by S.R., C.C.D., A.M.E., I.W.P., S.J.M., T.R-P,
425 C.A.S, and E.R.S. L.G.A. led the writing of the manuscript and S.R., C.C.D., A.M.E., I.W.P.,
426 S.J.M, M.O'B, C.A.S., and E.R.S. contributed significantly to the subsequent revisions. S.R.
427 served as Ph.D. advisor for L.G.A. All authors contributed critically to the drafts and gave final
428 approval for publication.

429 **Data Availability**

430 The data used to create graphs from Box 1 are openly available in Environmental Data Initiative
431 (EDI) at [http://doi.org/\[doi in progress\]](http://doi.org/[doi in progress]), reference number [*reference number in progress*].

432 Additionally, the data derived in this article are available from USA-National Phenology
433 Network at <http://doi.org/10.5066/F78S4N1V>, National Ecological Observatory Network at
434 <https://www.neonscience.org/data>, Dryad at <https://datadryad.org/stash>, and EDI at
435 <https://edirepository.org/>. These data were derived from the following resources available in the
436 public domain:

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