1	The implementation of network meta-analysis in Ecology; a case study using
2	crop yield data
3	https://github.com/maxanochirim/network-meta-analysis-using-crop.yield.data.git
4	
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19 ABSTRACT

20 Network meta-analysis (NMA) is a method commonly used in medical research that allows for the 21 comparison of multiple interventions in a single, coherent analysis. In this study, we explore how 22 NMA can be applied in ecological studies – specifically, in comparing the effectiveness of multiple 23 interventions in field experiments. Our study aims to provide a general and non-technical 24 introduction of network meta-analysis to ecologists, particularly on key assumptions and methods. 25 Using an example, we demonstrate how NMA can serve as a tool to compare the effectiveness 26 of different interventions used in enhancing crop yield. We conducted a systematic review and 27 extracted data from meta-analytical studies that explored the response of yield to an intervention. 28 Each study structured data as a pairwise comparison between an intervention and a control. 29 Using yield as the measure of effectiveness, we evaluated four interventions – Liming, Straw return, Super Absorbent Polymers, and considered no intervention as the control. All 30 31 measurements came from field experiments, and we analyzed data from 3733 independent 32 studies that were included in these meta-analytical reviews. The results of our analysis 33 demonstrate the potential of NMA as a valuable statistical method in ecological research, 34 providing more precise comparisons of multiple interventions. However, we emphasize the need 35 for careful consideration of important assumptions such as transitivity and consistency when 36 implementing NMA in ecological studies. This study offers a novel approach to synthesizing 37 ecological data, contributing to improved decision-making in agriculture, ecology, and 38 environmental sciences.

39

Keywords: Network meta-analysis, Multi-treatment analysis, Mixed-treatment comparison,
Evidence synthesis, Systematic reviews, Policy, Decision making, Crop yield, Interventions,
Agriculture, Ecology, Environmental science

43 1. INTRODUCTION

44 1.1. What is Network Meta-Analysis (NMA)

45 The advent of meta-analysis in ecology in the early nineties (Arnqvist & Wooster, 1995; 46 Fernandez- Duque & Valeggia, 1994; Jarvinen, 1991) provided researchers with a statistical 47 method that helped to systematically synthesize empirical findings (Bilotta et al., 2014; H. M. 48 Cooper et al., 2019). In contrast to narrative reviews and other qualitative methods that were 49 previously used to summarize study findings (Gurevitch et al., 2018; Slavin, 1995), the 50 quantitative synthesis of empirical findings allowed for data from multiple studies to be drawn 51 together. Additionally, and since the former did not require a clear formulation of research 52 questions nor a quantitative synthesis of the data (Vetter et al., 2013), meta-analysis proved 53 extremely useful and was considered a powerful statistical method for summarizing research 54 findings. Furthermore, the quantitative synthesis of results across studies offers better precision 55 as well as a higher statistical power as it helps in identifying sources of variation in study outcomes 56 (Gurevitch et al., 2018). Subsequently, the use of meta-analysis (which is a part of the larger field 57 of research synthesis) became widely accepted (Figure S1) and adapted across many scientific 58 disciplines despite its initial skepticism (Cadotte et al., 2012; Gurevitch et al., 2018; Hillebrand & 59 Cardinale, 2010; Hunt, 1997; Whittaker, 2010). It offered many advantages including 60 transparency, replicability, generalizability, and the ability to better quantify effect size by 61 analyzing the result from multiple independent studies (Conn et al., 2012; Hernandez et al., 2020).

In ecological and environmental contexts, the application of these methods has not only helped in the identification of research gaps (Thomsen et al., 2012), but has also allowed for the estimation of the direct effects of major environmental drivers – including climate change, invasive species, and habitat fragmentation (Aguilar et al., 2006; Clewley et al., 2012; Jactel et al., 2012). Additionally, these approaches have allowed for comparisons of these effects across different scales, taxa, and ecosystems (Powell et al., 2011; Rodríguez-Castañeda, 2013), thus helping to

68 enhance the reliability and generalization of ecological conclusions (Cadotte et al., 2012; Maki et al., 2018; Nakagawa et al., 2023). Likewise in the field of plant ecology, the quantitative synthesis 69 70 of research results has had a tremendous impact, proving extremely useful in objectively 71 summarizing findings from numerous studies (Koricheva & Gurevitch, 2014). This has led to the 72 development of standardized protocols, usage, and organizations such as the Collaboration for 73 Environmental Evidence – CEE that guide both scientists and policymakers in providing the best 74 evidence for high quality environmental decision-making processes (Koricheva & Gurevitch, 75 2014; Nakagawa et al., 2023; Pullin & Stewart, 2007; Vetter et al., 2013), a move similar to already 76 established standards like The Cochrane Collaboration in the field of medical sciences.

77 Despite all of this, meta-analysis still has its limitations. For example, when one would like to make 78 inferences about all the indicated treatments for the same condition and sample characteristics, 79 the standard meta-analytic methodology experiences a significant drawback. Furthermore, since 80 meta-analysis is focused on comparing only two interventions at the same time, the use of this 81 method implies that all studies must have a common treatment. Likewise, the growing multiplicity 82 of available scientific literature on a topic, particularly in ecology produces a plethora of 83 information/interventions which could unfortunately complicate decision-making processes (Dias 84 & Caldwell, 2019; Koricheva et al., 2013; Roberts et al., 2006). Additionally, ecological systems 85 are oftentimes highly interconnected, with multiple drivers of change acting simultaneously. The 86 limiting ability of meta-analytical methods to adequately mirror the complexity of natural 87 ecosystems allows room for the desire of a statistical method that can optimally address questions 88 about how different environmental factors/interventions interact and compare in their effects. 89 There is therefore growing a new perspective to make inferences about competing treatments for 90 the same condition, one such as the network meta-analysis.

91 First coined by Lumley (2002), the methodology of network meta-analysis, (sometimes called 92 multi-treatment analysis or mixed-treatment comparison, and hereafter referred to as NMA)

93 focuses on combining both the direct and indirect information across a network of randomized 94 studies to infer about the relative effectiveness of multiple interventions (Dias, 2018; Lumley, 95 2002; Madden et al., 2016). In a scenario with multiple interventions, as is the case with most ecological data, it is clear that not all possible pairwise comparisons will have been 96 97 separately/directly carried out. The method of NMA solves this issue by combining in a single 98 coherent analysis, the result of those studies/interventions where a direct comparison has been 99 made (direct evidence) together with those studies/interventions where there is no pairwise 100 comparison (indirect evidence). The term 'indirect' is used because it relies on evidence against 101 the common comparator or other all relevant comparators of interest, and not on 'direct' head-to-102 head evidence. Thus, with NMA, all interventions can be compared with one another, including 103 comparisons not evaluated within any of the primary studies further strengthening inferences 104 concerning the relative efficacy of treatments (Lu & Ades, 2004). This type of method incorporates 105 all available evidence into a general statistical framework for the comparison of all available 106 treatments thus resolving the limitations of traditional pairwise meta-analyses" (Tu, 2014).

107 With NMA, it becomes possible to statistically analyze the result from multiple independent studies 108 with different treatments (Higgins & Whitehead, 1996; Salanti, 2012; Salanti & Schmid, 2012); 109 thus providing a more effective comparative review method and subsequently forming the basis 110 for coherent, evidence- based treatment decisions. A major advantage of NMA is that correlations 111 of estimated treatment effects are automatically considered when an appropriate model is used 112 thus improving the precision for the estimated effect sizes. It facilitates simultaneous inference 113 regarding all treatments in order to select the best possible treatment for example (Lu & Ades, 114 2004; Figure 1). Additionally, NMA makes it possible for the comparisons of interventions that 115 are not possible in a single study because all treatments of interest may not be included in any 116 given study. Other advantages of a network meta-analysis include: (i) the preservation of 117 randomization: Because a network meta-analysis does not directly compare treatment arms 118 across studies but rather the relative differences between treatment comparisons, it preserves 119 randomization (i.e., takes into consideration the fact that samples were randomized to 120 interventions within studies but not across studies), (ii) the maximization of all available evidence. 121 (iii) the production of treatment ranking which is useful in decision-making processes or in 122 revealing the most effective components of complex interventions.



123

Figure 1. A hypothetical scenario where NMA combines all available evidence including those that have not previously been compared to each other.

126

127 By focusing on the use of network meta-analytic approaches in ecological studies, we argue that 128 this relatively new approach and subsequent improvements will aid both researchers and 129 policymakers in their ability to promptly and effectively craft policies that address pressing 130 environmental issues. The development of user-friendly software, protocols, and guidelines such 131 as the PRISMA-NMA extension (Preferred Reporting Items for Systematic Reviews and Meta-132 Analyses; Hutton et al. (2015); http://www.prisma-statement.org/nma) is evidence that the method 133 of NMA is growing in acceptance and popularity. Already, it is being embraced by national and 134 international policy-making bodies as a tool to best answer complex policy-relevant questions 135 (Salanti & Schmid, 2012).

When conducting a network meta-analysis, it is however important to note that analytical methods are more complex and can be quite challenging (Tu, 2014). In addition, confusion about how to choose models, fit them to the data, or interpret the results can be expected for users who are not yet familiar with the different statistical models (Dias, 2018; Madden et al., 2016).

140 1.2. Research Objectives

141 While studies comparing more than two competing interventions are guite common in several 142 research fields including medicine, pharmacology, social psychology, and education, the 143 implementation of the NMA methodology is not yet common practice in ecological studies. A 144 search for published literature on NMA using the keyword "network meta-analysis" in the literature 145 database ISI Web of Science revealed that the field of ecology contributed less than 1% out of 146 the approximately 27 thousand entries (Figure S2). Our main research objective therefore was 147 to investigate the feasibility of applying the methodology of indirect comparisons (network meta-148 analysis) in the scientific field of ecology as a way of further improving the precision of results that 149 could be obtained from ecological studies. We aim to provide a general and non-technical 150 introduction of network meta-analysis to ecologists, and the general research community in the 151 field of environmental management.

We showed an example of how NMA works by developing a research question that seeks to find out if we could implement a multi-treatment meta-analysis to uncover the effects of different interventions used in the yield production of croplands. To the best of our knowledge, we believe this is the first detailed guideline on the application of NMA in ecology, particularly in how this method can be utilized to assess the efficacy of interventions used in improving crop yield.

157

158 **2. METHODS**

159 2.1. Basic Concepts and Assumptions of a Network Meta-Analysis

160 2.1.1. Concepts and terminology

161 The methodological advantage behind a network meta-analysis is a very simple one: "indirect 162 comparisons". Assuming we have two independent studies (study 1: directly comparing treatment A versus C [μ^{D}_{AC}], and study 2 directly comparing treatment B versus C [μ^{D}_{BC}] but no studies 163 164 directly comparing A versus B [µ^D_{AB}]), we can indirectly compare treatment A to treatment B via 165 the common comparator C. This is done by statistically combining the information from all A 166 versus C (AC) and B versus C (BC) studies represented here as: $\mu_{AB}^{I} = \mu_{AC}^{D} - \mu_{BC}^{D}$ (Figure 2). 167 With a network meta-analysis, it is also possible to further improve the precision of treatment estimates assuming a scenario exists where both the direct $[\mu^{D}_{AB}]$ and indirect estimates $[\mu^{I}_{AB}]$ are 168 169 available for the same comparison. In this situation, a 'mixed' effect size [µ^M_{AB}] is calculated by 170 taking the weighted average of μ^{D}_{AB} and μ^{I}_{AB} (Bucher et al., 1997).

171



Figure 2. An example of a network of three treatments (ABC) compared in two studies (solid black lines),where an indirect comparison can be made (dashed grey line).

175 2.1.2. Statistical assumptions

Since indirect and mixed comparisons are generally considered observational in nature (CataláLópez et al., 2014), two major assumptions need to be considered and where possible, met,
before a network meta-analysis study can be considered as valid.

179 1. Assumption of transitivity: The main assumptions that underpin the validity of indirect and 180 mixed comparisons is that there are no significant differences between the studies making 181 different comparisons other than the treatments that are being compared (Cipriani et al., 182 2013). Since the aim of a network meta-analytical study is to compare two treatments via a 183 third one, it assumes that indirect comparison validly estimates unobserved head-to-head 184 comparison (Salanti, 2012). With our earlier example comparing treatment A versus treatment 185 B via treatment C, our common comparator which is C is regarded as 'transitive'. This is 186 because it allows a valid comparison of the treatment to which it is linked (Salanti, 2012). With 187 transitivity, similarity is not required for all characteristics of studies and samples across the 188 evidence base i.e., valid indirect comparisons can be obtained even when studies are 189 dissimilar in characteristics which are not effect modifiers (Dias, 2018). In technical terms, 190 what this means is that indirect comparisons can still be made between a study conducted in 191 pot experiments and another study that was conducted in field conditions provided that the 192 study samples were not shown to have a modifying effect on the result from any of the study. 193 Additionally, this assumption requires that treatments/interventions should be comparable 194 among themselves in practice. For example, assuming we were comparing two treatments A 195 and B for plant yield. If treatment A needs a precondition before it can be implemented (e.g. 196 only as a second-line treatment or perhaps only to samples with certain conditions) and 197 Treatment B doesn't need these preconditions, then the law of transitivity is violated because 198 in practice, the observational samples would be different for each treatment (we would not be

able to randomly assign the samples to just any treatment in a fresh study). Furthermore, samples that are included in a network should be able to be randomized (sample randomization) to any of the treatments/interventions. i.e. (in principle), one should be able to apply any of the treatments/interventions randomly to all participants in the network. The assumption of transitivity could be violated if interventions have different indications.

204

4 Questions to be asked regarding the assumption of transitivity:

- i. In the planned (ecological) study, can the two treatments that want to be comparedindirectly form a common node?
- 207 ii. Are the missing treatments 'missing at random' or is it directly associated with the true
 208 relative effectiveness of the interventions (intervention effectiveness bias)?
- iii. Is the choice of the comparator random? If the choice of the comparison is associated,
 directly or indirectly, with the relative effectiveness of the interventions, then the
 assumption of transitivity is violated.
- 212 2. <u>Assumption of consistency</u>: Consistency is the extension of transitivity over a loop of evidence 213 (Cipriani et al., 2013). With consistency, the major assumption is that the direct and indirect 214 estimates/sources of evidence agree i.e., both the direct and indirect evidence are estimating 215 the same underlying treatment effect. This assumption can be measured/evaluated 216 statistically with the use of a simple z-test, often called the Bucher method (Bucher et al., 217 1997). It can be evaluated only when there is direct and indirect evidence existing in the 218 evidence network for a particular comparison of interventions (Dias et al., 2010; Dias & 219 Caldwell, 2019; Higgins et al., 2012). When the direct comparisons of means are different 220 from indirect comparisons, then the network is said to be inconsistent (Cipriani et al., 2013; 221 Lu & Ades, 2004, 2006). The assumption of consistency is a prerequisite in calculating a valid 222 mixed estimate. A significant f-test for the design and treatment interaction is an indication of 223 inconsistency.
- 224 Questions to be asked regarding the assumption of consistency:

i. Do the ecological studies use comparable methodologies (e.g., measurement ofspecies abundance, or habitat quality) to ensure consistency in effect estimates?

- ii. Is there significant variability in ecological contexts that might lead to inconsistentresults between direct and indirect comparisons?
 - iii. Do studies reporting direct comparisons conflict with indirect evidence due to context-specific factors?

231 3. Exchangeability assumption: This assumes that the distribution of effect modifiers (variables 232 that influence treatment effects) is balanced across studies. i.e., that two sets of studies e.g. 233 AC and BC, do not differ with respect to the distribution of effect modifiers. For example: if 234 'site' was an effect modifier in all AC studies (irrespective of their distribution – heterogeneous 235 studies), then in order to make a valid indirect comparison of AB; 'site' should also be an effect 236 modifier (distributed in a similar proportion) in all BC studies - AC and BC studies should 237 therefore cover the entire spectrum of the observed effect modifier. It is therefore important to 238 identify apriori possible effect modifiers and compare their distributions across comparisons 239 when synthesizing evidence from many comparisons. Adjustments can be used to improve transitivity (through network meta-regression or a subgroup analysis) if an imbalanced 240 241 distribution of effect modifiers is identified. Adjustment should take place only for study or 242 sample characteristics that are categorized as effect modifiers.

243 Questions to be asked regarding the assumption of exchangeability:

- i. Are key ecological variables (e.g., temperature, precipitation, soil) influencing the
 treatment effects similarly distributed across the studies in the network?
- 246 ii. Are the interventions or treatments implemented in similar ecological settings, or247 are there systematic differences?
- iii. Are there outliers among the studies (e.g., studies conducted in extreme
 environments or with highly specialized species) that might bias the overall effect
 estimate?

251 2.1.3. Statistical approaches to fitting a network meta-analysis

252 Both Bayesian and frequentist approaches can be used in fitting a network meta-analysis model 253 (Dias & Caldwell, 2019; Hong et al., 2013). Bayesian methods for NMA require selecting a prior 254 probability distribution that describes the range and probability of plausible values for the 255 parameters of interest (e.g., treatment effect). Using Bayes theorem, this is then combined with a 256 likelihood statement that provides information on the collected data (Dias & Caldwell, 2019). On 257 the other hand, the frequentist approach to network meta-analysis does not require prior 258 knowledge or beliefs. Relying solely on the data collected, the frequentist approach calculates 259 probabilities and estimates based on how likely the observed data would be under different 260 assumptions about the parameter of interest (e.g., treatment effect) (Rücker, 2012). The results 261 from both analyses however are guite similar (Dias & Caldwell, 2019; Hong et al., 2013) with the 262 main difference being the way results are presented. Results from a frequentist approach are 263 presented as estimated relative effects and a corresponding 95% Confidence Interval (CI), while 264 results from a Bayesian NMA analysis are presented as summaries of the effect (typically, mean 265 or median) and a 95% credible interval (CrI) (Dias & Caldwell, 2019). Both approaches can be 266 implemented in commonly used statistical software such as R. Bayesian approaches to network meta-analysis are conducted with packages such as; "gemtc", "bnma", "pcnetmeta", "multinma", 267 268 "nmaINLA", "bayesmeta", "BUGSnet" while the package "netmeta" developed by Balduzzi et al. 269 (2023) is used for frequentist methods.

270 2.2. An example of network meta-analysis in ecology: a case study using crop yield data
271 2.2.1. Data collection

In this systematic review and network analysis, we searched the database used in Takola et al. (unpublished) to identify published studies of interventions applied in the production of various crop yields. We compiled a dataset of meta-analytical studies from around the globe that had investigated the response of yield production to an agricultural intervention (**Figure 3**). These

276 strategies - ranging from; straw return, liming, super absorbent polymers, alternative fertilization 277 options, substitution of mineral fertilizers with manure nitrogen, removal of topsoil, etc. are among 278 some of the most common agricultural management strategies for balancing yield increase (a full 279 list of interventions can be found in **Table S1**). Our initial dataset (from Takola et al., unpublished) 280 consisted of 11 meta-analytical studies, with 13 interventions and 8814 yield data (Figure 4). 281 Study information extracted include (i) study identifiers (e.g. title, abstract, authors, publication 282 year); (ii) study characteristics (e.g. study design, study location); (iii) participant characteristics 283 (e.g. sample size, standard deviation); (iv) intervention details; and (v) outcome data (e.g. effect 284 sizes of intervention and control). Our primary outcome of interest was yield production measured 285 in kilogram per hectare (kg/ha).



Figure 3. Locations of sites included in the network meta-analysis. (see Appendix for a complete list of references)





Figure 4. A PRISMA flow diagram showing the systematic review process ('identification', 'screening',
 'eligibility', and 'included') of selecting publications relevant to our network meta-analysis.

292 2.2.2. Inclusion criteria

293 To help ensure consistency in our analysis, we defined a set of criteria to screen out irrelevant

studies and only included studies that (i) were randomized controlled trials (ii) were paired i.e.,

compared an active intervention with a control (iii) had no missing data e.g. SD, sample size (n),

296 effect sizes, etc. (iv) used interventions not applied as a second-line treatment.

After applying our selection criteria (**Table 1**), we extracted data from 3 meta-analytical studies of

interest (**Table 2**). These studies comprising of 3 paired interventions and 3733 yield data met all

299 inclusion criteria and were then used for the network meta-analysis.

300 Table 1. Summary of eligibility criteria

Inclusion criteria

Exclusion criteria

Population	Studies with no missing da size (n).	ta. E.g. SD, sample	No SD	
	Studies that were o	pen-access		
	Studies that had suppler available	mentary datasets		
Intervention	Interventions for imp	proving yield		
Study design	Randomized contr	rolled trials		
ComparisonPaired. i.e., Intervention vs Control group (noInterventionintervention)pa				
	Yield values			
Outcome	Yield valu	es		
Outcome	Yield valu	es		
Outcome	Yield valu	es		
Outcome Table 2. Interventions No.	Yield valu s of Interest Intervention	es Dataset		
Outcome Table 2. Interventions No. 1.	Yield values s of Interest Intervention Straw return	es Dataset D331		
Outcome Table 2. Interventions No. 1. 2.	Yield value s of Interest Intervention Straw return Liming	es Dataset D331 D973		
Outcome Table 2. Interventions No. 1. 2. 3. S	Yield value s of Interest Intervention Straw return Liming Super Absorbent Polymer (SAP)	es Dataset D331 D973 D652		

303 2.2.3. Data analysis

304 To make a comparison of these selected interventions on yield production, we conducted a 305 network meta-analysis in R (version 4.3.2) using the netmeta package. This approach adopts 306 frequentist methods and calculates point estimates and their corresponding confidence intervals 307 based on weighted least squares regression (Rücker, 2012). We calculated the effect of the 308 different agricultural interventions on yield production. As a summary measure of effect size, we 309 estimated standardized mean differences (SMD) of each intervention relative to the control group 310 using pairwise and a random effects meta-analytical model. In choosing a random-effects model 311 for our network meta-analysis, we assume that heterogeneity exists among studies and that all 312 effect sizes did not come from one population i.e., each study has different overall means and 313 they don't have one true overall mean. By choosing a random-effects model, our analysis further 314 accounts for variability both within and between studies. The alternative to this is the 'fixed-effect' 315 model commonly referred to as the 'common-effect' model which assumes that all effect sizes 316 (from different studies) come from one population (Nakagawa et al., 2023). From the generated 317 estimates and confidence intervals, probability scores (P-scores) were calculated and these were 318 used to hierarchically rank each intervention according to their effects on yield using methods 319 developed by Rücker & Schwarzer, 2015. Estimates, confidence intervals, and P-scores then 320 allowed us to construct, forest plots and league tables which are useful in visualizing the 321 comparisons.

Our data input was based on a wide-arm format with 3733 rows (where each row corresponds to a single study with multiple or double comparisons) and it was subsequently transformed during the analytical process into the standard contrast-based format using the auxiliary *pairwise* function of the *netmeta* package. Our dataset has a continuous outcome wherein all variables containing information on group sample sizes (argument n), means (mean), and standard deviations (SD) are provided – hence the reason why we chose the SMD as an effect size. SMD is used as a

328 summary statistic when studies assess the same outcome but use different measurement 329 methods (9.2.3.2 The Standardized Mean Difference, n.d.). In such cases, it becomes necessary 330 to standardize the results of the studies to a common scale before combining them. SMD (often 331 times referred to as Hedges' g or Cohen's d) is considered a comparative measure because it is 332 typically used when comparing two groups (Nakagawa et al., 2023). In the *netmeta* package, the 333 method by Crippa & Orsini, (2016) is used to guarantee consistent SMDs and standard errors for 334 multi-arm studies (Balduzzi et al., 2023).

$$SMD = \frac{Difference in mean outcome between groups}{Standard deviation of outcome among participants}$$

Similarly, a workflow detailing the steps involved in performing a network meta-analysis, along
with the related functions of *netmeta* in R is provided in **Figure S3**. All codes used in our study
are also provided in the supplement section.

339 **3. RESULTS**

340 *3.1.* **Data summary**

341 This dataset contains cleaned data extracted from a systematic review assessing the effect of 342 different interventions on yield production in different crop farms (maize, wheat, cotton, oat, etc.). 343 They are a combination of three pairwise meta-analyses comparing the effects of "Straw return", 344 "Super Absorbent Polymer", and "Liming" respectively, with a Control (no treatment). The primary 345 outcome was "Yield production". Relative treatment effects were expressed as SMD. Data on this 346 outcome were available for 3733 (total number of samples) from 3733 single pairwise comparison 347 studies; all of which are two-arm studies comparing an active treatment against the absence of a 348 treatment (control), thus providing information for two means (y1, y2), two standard deviations 349 (sd1, sd2), two group sample sizes (n1, n2), along with two treatment labels (t1, t2).

350 3.2. Summary of Networks

The number of treatments of interest in our network (also called nodes or vertices) is 4 (n=4) and d which is the number of designs is 3. Additionally, each study contributes a number of pairwise comparisons (m) and the total sum of all pairwise comparisons across studies in our network is 3732. K (which is the number of independent studies) simultaneously corresponds to m and this is because there are only two-arm studies in our network. Assuming there was at the least, one study evaluating more than two treatments, then m will be greater than k.

357 3.3. Net graph

358 In the network graph (Figure 5), we get a graphical presentation of the network structure with 359 each treatment represented as a point (node) in the plane. It shows a network of interventions 360 compared in a yield production study. Furthermore, treatments are connected by a line (edge) if 361 at least one direct pairwise comparison exists with the thickness of the edges being proportional 362 to the number of studies directly comparing treatments. There are 3 edges in the plot, suggesting 363 that 3 of the 6 pairwise comparisons had direct evidence, while the remaining 3 (SAP versus 364 Straw return, SAP versus Liming, and Liming versus Straw return) had only indirect evidence. 365 Our network graph also visualizes the number of studies contributing to each pairwise 366 comparison. From the net graph, we immediately see from the line width that the comparison of 367 Super Absorbent Polymer (SAP) versus Control has the largest number of studies (1424). 368 Furthermore, all studies were two-armed and only had direct comparisons.





Figure 5. Network graph without crossings for yield production network meta-analysis. Line widths areproportional to the number of studies directly comparing treatments

372 3.4. Forest plot

Forest plots provide a graphical display of the observed effect, confidence interval, and often times the weight of each study (Harrer et al., 2021). It can also be used as a way to better visualize the uncertainty in our network. With "Control" as the comparison group and from the forest plot that was produced, we visually see that the intervention "Super Absorbent Polymer" works better in comparison with the other interventions in improving crop yield, while the intervention "Straw return" does not have a strong impact on yield (**Figure 6 and 7**).



Figure 6. Forest plot for yield production network meta-analysis; with control as reference.

Comparison	Studies	Evidence	12	Random Effects M	odel SMD	95%-CI
Liming:Control Direct estimate Indirect estimate Mixed treatment estimate	1187	1.00	47%	*	1.24	[1.12; 1.37]
				*	1.24	[1.12, 1.57]
Straw return:Control Direct estimate	1121	1.00	54%		1.07	[0.94; 1.21]
Mixed treatment estimate				\$	1.07	[0.94; 1.21]
Super Absorbent Polym		ontrol				
Direct estimate	1424	1.00	89%		+ 6.19	[5.98; 6.41]
Indirect estimate Mixed treatment estimate					♦ 6.19	[5.98; 6.41]
Liming:Straw return						
Direct estimate	0	0			0.47	
Indirect estimate Mixed treatment estimate				÷	0.17	[-0.01; 0.36] [-0.01: 0.36]
						[,
Liming:Super Absorben	t Polymer	(SAP)				
Indirect estimate	0	0			-4.95	[-5.20: -4.70]
Mixed treatment estimate				♦	-4.95	[-5.20; -4.70]
Straw return: Super Abs	orbent Poly	(mer (SAP)				
Direct estimate	0	0				
Indirect estimate				+	-5.12	[-5.37; -4.86]
Mixed treatment estimate				\$	-5.12	[-5.37; -4.86]
Figure 7. Forest plot for y	ield produ	ction netw	ork r	meta-analysis (active	intervention v	ersus all other

383 treatments).

384 3.5. Treatment estimates (SMD) – Assessment of interventions on yield

The network estimates for the random effects models are provided, using Control as the reference group. We produced network estimates, lower and upper confidence limits for all observed and unobserved pairwise comparisons. The results in **Table 4** below summarizes the standardized mean differences (SMDs) of Yield (in kg/ha) for each treatment compared to the "Control" along with their 95% Confidence Intervals.

	SMD	95%-CI	Z	p-value
Control				
Liming	1.24	[1.12; 1.37]	18.89	< 0.0001
Straw return	1.07	[0.94; 1.21]	15.88	< 0.0001
Super Absorbent Polymer (SAP)	6.19	[5.98; 6.41]	55.72	0

390 Table 4. Summary of standardized mean differences of yield

391 The random effects NMA shows strong evidence that all treatments have a significant positive 392 effect on yield production compared to the control, as indicated by p-values < 0.0001. Liming 393 (SMD = 1.24, 95% CI [1.12, 1.37]) had a moderate positive effect compared to the control, with a 394 statistically significant result (p < 0.0001). The intervention - liming, increases crop yield by 1.24 395 kg/ha, compared with an absence of it. In addition, we can also say that there is a 95% probability 396 that this increase is between 1.12 and 1.37 kg/ha. Straw return (SMD = 1.07, 95% CI [0.94, 1.21]) 397 also showed a moderate positive effect relative to the control, and the result was statistically 398 significant (p < 0.0001). SAP showed the largest effect size (SMD = 6.19), indicating a very strong 399 impact compared to the other treatments.

400 3.6. Assessing heterogeneity (within designs) and inconsistency (between designs)

$$tau^2 = 4.1014$$
; $tau = 2.0252$; $l^2 = 78.2\%$ [77.6%; 78.9%]

402 *I*² measures the proportion of total variation in effect estimates that is due to heterogeneity rather
403 than chance. An I² of 78.2% suggests substantial heterogeneity, indicating that the treatment
404 effects vary considerably across studies.

	Q	d. f.	p-value
Total	17123.15	3729	0
Within designs	17123.15	3729	0
Between designs	0.00	0	

405 Table 5. Quantifying heterogeneity and inconsistency

The results of the Q statistic tests for heterogeneity among studies (Total Q) show a p-value of < 0.0001, thus indicating significant heterogeneity. Additionally, the significant p-value of the Q statistic examining heterogeneity within the groups of studies that share the same design (within designs) indicates substantial variability even within these groups. The tests of inconsistency (Between designs) typically assess differences between groups of studies with different designs. The results of our study suggest that all heterogeneity is captured within designs since there are no degrees of freedom provided for this test.

Based on the results of the tests for heterogeneity and inconsistency **(Table 5)**, we can deduce that there is significant variation among studies (this is supported by the high value of l² and the significance of the Q value. Given this, we performed the following recommended steps as recommended by N. J. Cooper et al., 2009. 417 **Investigating sources of Heterogeneity:** With the use of (i) Subgroup Analysis, there is 418 the possibility of conducting analyses within subgroups of studies to explore whether 419 certain study characteristics would account for the observed heterogeneity, and (ii) the 420 use of meta-regression techniques to identify factors that might explain the variability in 421 effect sizes across studies. In this study however, we did not explore the causes of 422 heterogeneity due to limitations of time, resource, and available information on important 423 covariates at a global scale (such as field size, management history, management 424 intensity, etc.).

425 3.7. Hierarchy/Ranking of competing treatments

Table 6 shows treatment rankings and the probability that each intervention is the 'best' or 'worst' 426 427 in improving crop yield. Here ranks are reported for effectiveness, such that rank 1 means that 428 the intervention is most effective. Control/No Intervention has a rank of 4 (*P-score* = 1.0000). That 429 is, on average, the absence of no intervention was ranked approximately fourth out of the four 430 available treatments (i.e., worst) for improving crop yield. Conversely, Super Absorbent Polymer 431 (SAP) was ranked first out of all four treatments and had a higher probability of being the most 432 effective treatment to improve crop yield. The area under the cumulative ranking curve (SUCRA), 433 a summarization method that gives an index of the overall performance of the treatments (Salanti 434 et al., 2011) can then be derived from the ranking probability of each treatment. SUCRA ranges 435 from 0 to 1 and the higher the SUCRA value, the greater the likelihood that a treatment is better 436 than the other treatments in the network. On the other hand, the closer the SUCRA value to 0, 437 the greater the likelihood that a treatment is worse than the other treatments.

438 Table 6. Treatment ranking based on P-scores and SUCRA values

Intervention	P-score	SUCRA values

Control	1.0000	1.0000	
Straw return	0.6548	0.6543	
Liming	0.3452	0.3457	
Super Absorbent (SAP)	Polymer 0.0000	0.0000	

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SUCRA values based on 1000 simulations

440 **4. DISCUSSION**

441 As experimental evidence in Ecology increases, so does the need for methodologies and 442 statistical models to analyze them. Meta-analyses are powerful tools that synthesize evidence 443 and help decision-making (Bilotta et al., 2014). Currently, pairwise meta-analyses are the most 444 commonly used statistical technique to quantitatively synthesize research findings and to 445 compare treatments. However, they have limitations as they can only compare two treatment 446 options at a time. Network meta-analyses are a tool to compare multiple treatments because they 447 combine in a single coherent analysis, studies where a direct comparison has been made (direct 448 evidence) together with studies where there is no pairwise comparison (indirect evidence). In 449 addition, they provide a helpful framework to present a comprehensive, and reproducible 450 synthesis of the evidence (Bilotta et al., 2014). However, the uptake of this method by ecological 451 research has been rather slow.

In this study, we present a network meta-analysis and implement it on data from agricultural field experiments. We provide a general introduction to network meta-analysis (including its advantages over the traditional pairwise meta-analysis), showcase the steps involved in conducting/correctly reporting one, and discuss the major assumptions that guide a standard NMA, such as consistency and transitivity. In this NMA, we were able to compare different

457 interventions used in improving crop yield. Doing so also helped us make comparisons of 458 interventions that had not been previously addressed in any individual primary study (indirect 459 evidence). The dataset we used contains experiments with the following treatments: Super 460 Absorbent Polymer, straw return, liming, and a control group (no intervention). Our results show 461 that the most effective treatment for yield enhancement is the Super Absorbent Polymer. The 462 hierarchy of the treatments was as follows: Super Absorbent Polymer, liming, straw return, and 463 control.

464 The benefits of Super Absorbent Polymer on crop yield have already been summarized in 465 previous reviews by Zheng et al. (2023). When compared to a control group (no intervention), the 466 addition of SAP significantly increased (p < 0.01) crop yields by 12.8% (CIs: 12.1 - 13.4%). The 467 effect sizes of crop yield under liming in comparison to control treatments (no liming) showed that 468 liming similarly had a positive influence on crop yield (Enesi et al., 2023; Li et al., 2019; Liao et 469 al., 2021; average of 12.9%). Likewise, Wang et al. (2021) reported an annual increase of 5.83% 470 in the yield of agricultural products like corn when an optimal scheme of straw return was 471 implemented relative to straw removal. Assuming we were to look at the result of our NMA individually (i.e. Treatment vs Control), these pairwise meta-analyses comparisons confirm our 472 473 results that all interventions significantly contribute to improving crop yield. Additionally, they could 474 offer some insight into the validity of the treatment rankings obtained during our network meta-475 analysis. To the best of our knowledge, the global network meta-analysis study published by 476 Herrmann et al. (2022) on the promotion of crop growth, yield, and quality by bioeffectors is the 477 first study that utilizes NMA methods on agricultural field experiments. We however recommend 478 that subsequent studies that plan on implementing this method follow reporting protocols and 479 already established guidelines such as the one suggested by Hutton et al. (2015) when 480 conducting or reporting the results of an NMA.

481 Network meta-analysis (NMA) can be a valuable tool for ecologists when synthesizing evidence
482 from studies that compare multiple interventions or management strategies in controlled

483 experimental settings. For example, lattice square agricultural experiments, where different 484 combinations of crops, fertilizers, or farming methods are systematically tested across multiple 485 plots, provide structured data suitable for NMA. Similarly, controlled field experiments designed 486 to evaluate different ecological restoration techniques, pest management strategies, or habitat 487 interventions often involve overlapping comparisons that align well with NMA assumptions, such 488 as transitivity and consistency. On the other hand, the applicability of NMA is more limited in 489 monitoring or impact studies, as these often lack the controlled settings and standardized 490 comparisons needed for a robust analysis. Monitoring studies frequently involve highly variable 491 contexts, such as natural ecosystems with diverse species interactions or long-term impacts influenced by unmeasured confounders (e.g., climate variability, historical land use). These 492 493 complexities make it challenging to meet the assumptions of exchangeability and consistency, 494 reducing the reliability of indirect comparisons in such cases.

495 A common criticism of the implementation of network meta-analyses in ecological contexts is that 496 experiments are not randomized because the researcher is not typically blind to the control and 497 treatment groups. Randomization is essential in medical studies because it helps to make causal 498 inferences between the treatment and the effect. Although the ecologist cannot be blind to field 499 treatments, we argue that causal inference is facilitated from the variation of the contexts in which 500 each treatment is applied (i.e. temporal and spatial contexts). Regarding the assumption of 501 transitivity which essentially seeks to ensure that studies are comparable to each other with 502 respect to any potential effect-modifying characteristics, we emphasize the need to carefully 503 assess the quality of studies to ensure they are comparable before they are included in a network 504 meta-analysis. Violating this crucial assumption can lead to inconsistencies in a network, which 505 can lead to inaccuracies in the result of the analysis.

506 In conclusion, the use of NMA has continued to grow over the last decades with expansions into 507 different research fields. We showcase with our study, how this method can be implemented on 508 data from agro-ecological experiments, but there is a plethora of other contexts in which this

509	method can be applied: for example, (i) in the estimation of the effects of environmental drivers -
510	including climate change, invasive species, and habitat fragmentation, (ii) in the comparison of
511	interventions used in nature conservation or forest management (iii) long-term experiments in
512	grassland ecosystems (such as The Jena Experiment). More importantly, the resulting hierarchy
513	of treatments, based on their effectiveness, is a very valuable and important tool to inform
514	decision-making. For example, policy-makers can evaluate specific conservation measures as
515	well as their interactions, to effectively design protected areas. Overall, network meta-analysis is
516	a novel tool at the disposal of ecologists, in their effort to find the best nature-based solutions.
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- 534 List of abbreviations
- 535 NMA Network Meta-analysis
- 536 CEE Collaboration for Environmental Evidence
- 537 PRISMA NMA Preferred Reporting Items for Systematic Reviews and Meta-Analyses
- 538 ISI Institute for Scientific Information
- 539 CI Confidence Interval
- 540 Crl Credible Interval
- 541 SD Standard Deviation
- 542 SAP Super Absorbent Polymer
- 543 PICO Population, Intervention, Comparison and Outcome
- 544 SMD Standardized Mean Differences
- 545 SUCRA Surface Under The Cumulative Ranking Curve

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554	DECLARATIONS
555	Ethics approval and consent to participate
556	Not applicable
557	Consent for publication
558	Not applicable
559	Availability of data and materials
560	• The datasets generated and analysed during the current study are publicly available online
561	at the following repository - https://github.com/Helmholtz-UFZ/network-meta-analysis-
562	using-crop.yield.data
563	Competing interests
564	The author(s) declare that they have no competing interests
565	Funding
566	Not applicable
567	Acknowledgements
568	Not applicable
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726	corn prod	luction in Nor	theast China:	An integrated reg	ional evaluatio	on with m	neta-anal	ysis and
727	system	dynamics.	Resources,	Conservation	and Rec	ycling,	167,	105402.
728	https://doi	.org/10.1016/j	j.resconrec.202	21.105402				
729	Whittaker	, R. J. (2010).	Meta-analyses	s and mega-mista	kes: Calling tir	ne on me	eta-analys	sis of the
730	species	richness-	productivity	relationship.	Ecology,	<i>91</i> (9),	252	22–2533.
731	https://doi	.org/10.1890/	08-0968.1					
732	Zheng, H	., Mei, P., Wa	ng, W., Yin, Y.,	Li, H., Zheng, M.,	Ou, X., & Cui	Z. (2023). Effects	of super
733	absorbent	t polymer on a	crop yield, wate	er productivity and	d soil propertie	es: A glob	al meta-	analysis.
734	Agricultur	al Water Mana	agement, 282,	108290. <u>https://do</u>	bi.org/10.1016	/j.agwat.2	2023.108	<u>290</u>
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744 SUPPLEMENT SECTION





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Figure S1. Number of publications on meta-analysis (search web of science using string "meta-analysis"
until March 2024). The number of papers using meta-analytical methods has increased exponentially over
the years.

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Figure S2. Pie chart showing the percentages of scientific contributions/articles on network meta-analysis according to research area (search web of science using string "network meta-analysis" until March 2024)



Figure S3. Workflow to perform a (component) network meta-analysis with the R package *netmeta*. Adapted

757 from (Balduzzi et al., 2023)

760 2. **Tables**

761 Table S1. Summary of retrieved datasets with accompanying interventions

Dataset	Title of the meta-analytical	Intervention	Citation
ID	study paper		
D221	Agricultural management	Strow roturo	Liu D. Song C. Vin Z. Eong C. Liu Z. 8
0331	Agricultural management		Liu, D., Song, C., Ain, Z., Fang, C., Liu, Z., α
			Au, F. (2023). Agricultural management
	increase, carbon		strategies for balancing yield increase,
	sequestration, and emission		carbon sequestration, and emission
	reduction after straw return		reduction after straw return for three major
	for three major grain crops in		grain crops in China: A meta-analysis.
	China: A meta-analysis		Journal of Environmental Management, 340,
			117965.
			https://doi.org/10.1016/j.jenvman.2023.1179
			<u>65</u>

D973	Potential benefits of liming to acid soils on climate change mitigation and food security	Liming	Wang, Y., Yao, Z., Zhan, Y., Zheng, X., Zhou, M., Yan, G., Wang, L., Werner, C., & Butterbach-Bahl, K. (2021). Potential benefits of liming to acid soils on climate change mitigation and food security. Global Change Biology, 27(12), 2807–2821.
			https://doi.org/10.1111/gcb.15607
D1120	The adaptive capacity of maize-based conservation agriculture systems to climate stress in tropical and subtropical environments: A meta-regression of yields	Straw addition	Steward, P. R., Dougill, A. J., Thierfelder, C., Pittelkow, C. M., Stringer, L. C., Kudzala, M., & Shackelford, G. E. (2018). The adaptive capacity of maize-based conservation agriculture systems to climate stress in tropical and subtropical environments: A meta-regression of yields. Agriculture, Ecosystems & Environment, 251, 194–202. https://doi.org/10.1016/j.agee.2017.09.019

D921A	Integrated biochar solutions	Integrated	Xia, L., Cao, L., Yang, Y., Ti, C., Liu, Y.,
&	can achieve carbon-neutral	biochar	Smith, P., van Groenigen, K. J., Lehmann,
D921B	staple crop production	solutions,	J., Lal, R., Butterbach-Bahl, K., Kiese, R.,
		Straw addition	Zhuang, M., Lu, X., & Yan, X. (2023).
			Integrated biochar solutions can achieve
			carbon-neutral staple crop production.
			Nature Food, 4(3), 236–246.
			https://doi.org/10.1038/s43016-023-00694-0
D309	Improving yield and nitrogen	Green	Ding, W., Xu, X., He, P., Ullah, S., Zhang, J.,
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