

1 **Altered phenotypic responses of asexual Arctic *Daphnia* after 10 years of**
2 **rapid climate change**

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4 Running Title: Response of asexual *Daphnia* to climate change

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20 Abstract

21 Understanding the fates of organisms and ecosystems under global change requires
22 consideration of the organisms' rapid adaptation potential. In the Arctic, the recent
23 temperature increase strongly impacts freshwater ecosystems which are important
24 sentinels for climate change. However, a mechanistic understanding of the adaptive
25 capacity of their key zooplankton grazers, among them polyploid, obligate parthenogenetic
26 *Daphnia*, is lacking. Theory suggests low adaptation potential of asexual animals, yet
27 examples exist of asexuals persisting through marked environmental changes. Here, we
28 studied asexual *Daphnia pulicaria* from a meromictic lake in South-West Greenland. Its
29 oxycline hosts purple sulfur bacteria (PSB), a potential food source for *Daphnia*. We tested
30 two key phenotypic traits: (1) thermal tolerance as a response to rapid regional warming
31 and (2) hypoxia tolerance tied to grazing of PSB in the hypoxic/anoxic transition zone. We
32 resurrected *Daphnia* from dormant eggs representing a historical subpopulation from 2011,
33 sampled modern subpopulation representatives in 2022 and measured phenotypic variation
34 of thermal (time to immobilization - T_{imm}) and hypoxia tolerance (respiration rate and critical
35 oxygen limit - P_{crit}) in clonal lineages of both subpopulations. Whole genome sequencing of
36 the tested clonal lineages identified three closely related genetic clusters, one with clones
37 from both subpopulations and two unique to each subpopulation. We observed significantly
38 lower T_{imm} and P_{crit} and a trend for higher respiration rates in the modern subpopulation,
39 indicating a lower tolerance to both high temperature and hypoxia in comparison to the
40 historical subpopulation. As these two traits share common physiological mechanisms, the
41 observed phenotypic divergence might be driven by a relaxed selection pressure on hypoxia
42 tolerance linked to variation in PSB abundance. Our results, while contrary to our

43 expectation of higher thermal tolerance in the modern subpopulation, provide evidence for
44 phenotypic change within a decade in this asexual *Daphnia* population.

45

46

47 Keywords: freshwater, zooplankton, resurrection ecology, environmental change,
48 phenotypic adaptation, respiration rate, whole genome sequencing, thermal tolerance,
49 hypoxia tolerance, critical oxygen limit

50

51

52 Introduction

53 Environmental change is a powerful driver of selection and evolutionary adaptation
54 (MacColl 2011; Garant 2020). However, if changes occur too rapidly, the microevolutionary
55 capacity of organisms may fail to track it (Bürger and Lynch 1995; Chevin et al. 2010). Under
56 the current unprecedented rate of environmental change (IPCC 2023) it is unclear which
57 organisms will persist, and under which conditions, as ecosystems are being affected
58 globally (Parmesan 2006; Finn et al. 2023). It is therefore crucial to understand whether and
59 how organisms can adapt at pace with the rapidly changing environments (Visser 2008;
60 Hoffmann and Sgrò 2011).

61 Growing evidence highlights the significant potential of sexually reproducing organisms for
62 rapid adaptation to climate change (Bradshaw and Holzapfel 2006; Catullo et al. 2019).
63 However, the rapid adaptive capacity of asexual natural populations is still poorly
64 understood (Bast et al. 2018; Jaron et al. 2021). Evolutionary adaptation is facilitated by the
65 process of natural selection acting upon genetic variation (Agashe et al. 2023). To generate
66 this variation, sexual organisms benefit from genetic recombination, a major force in the
67 evolution and establishment of sexual reproduction (Crow 1994). Spontaneous mutations
68 additionally contribute to genetic variation, but are expected to produce mainly deleterious
69 effects, which should constrain mutation rate at an evolutionary low optimum (Kimura
70 1967). Without the ability for recombination, asexual lineages are thus bound to accumulate
71 non-beneficial mutations in the long term (known as Müller's ratchet), resulting in
72 mutational meltdown (Lynch et al. 1993). Conversely, recent studies of vertebrate and
73 invertebrate asexuals did not find evidence for a reduced fitness of long-lived clones
74 compared with sexual congeners despite the accumulation of mutations (Kočí et al. 2020;

75 Kearney et al. 2022) and uncovered that the asexual genome is much more dynamic than
76 previously thought including high rates of gene conversion, ameiotic recombination and
77 deletions (Xu et al. 2011; Tucker et al. 2013; Jaron et al. 2021).

78 The microcrustacean *Daphnia*, a freshwater keystone grazer, is an important model for
79 rapid evolution. Species of this genus generally reproduce by cyclical parthenogenesis (CP),
80 but obligate parthenogenesis (OP) exists, and in higher latitudes is often coupled with
81 polyploidy (Weider 1987; Dufresne and Hebert 1997; Decaestecker et al. 2009). Many
82 studies have evidenced the genetic and phenotypic responses of CP *Daphnia* over centuries
83 and even decades to cultural eutrophication (Frisch et al. 2014), dietary cyanobacteria
84 (Hairston et al. 1999), predation pressure (Cousyn et al. 2001; Chaturvedi et al. 2021),
85 parasites (Decaestecker et al. 2007), salinisation (Wersebe and Weider 2023), and warming
86 (Geerts et al. 2015; Yousey et al. 2018). However, studies in OP *Daphnia* that address the
87 question of rapid adaptation are lacking, especially in Arctic regions where climate change
88 has particularly severe effects.

89 A powerful method to directly observe evolutionary change in a population across different
90 time periods is known as resurrection ecology. It uses revived dormant stages from natural
91 populations archived in lake sediment (Burge et al. 2018; Weider et al. 2018) for
92 experimental study of otherwise inaccessible phenotypic traits of ancestral populations.

93 South-West Greenland is a suitable region to test potential contemporary microevolutionary
94 responses, as it has been undergoing particularly pronounced environmental change during
95 the last decades. Average June air temperatures in this region, representing the end of
96 spring thaw, have increased by 2.2°C in the last 30 years, while average summer
97 temperatures (July) increased by 1.1°C in the last 20 years (Saros et al., 2019). Associated

98 with this average increase of air temperature, there is a trend towards an earlier lake ice-
99 out (Šmejkalová et al. 2016). A recent study in the region of Kangerlussuaq, SW Greenland,
100 showed that the timing of ice-out has implications on the duration of spring mixing and
101 oxygenation of the water column (Hazuková et al. 2024). This area also harbours several
102 oligosaline meromictic lakes (Anderson et al. 2001). These oligotrophic, fishless lakes
103 contain Arctic *Daphnia pulicaria*, a member of the *Daphnia pulex* species complex
104 (Colbourne et al 1998). Following the pattern often observed towards higher latitudes
105 (Beaton and Hebert 1988), *Daphnia* populations in this area are polyploid and obligate
106 parthenogens (asexuals) (Dane et al. 2020).

107 We focused our study on one of these lakes, Braya Sø, where a single asexual clone has
108 dominated the population for at least the past 200 years (Dane et al. 2020), suggesting a
109 genetically nearly uniform population. An important characteristic of this lake is the
110 presence of phototrophic purple sulfur bacteria (PSB) that form a bacterial plate in the
111 upper part of the anoxic zone. The sediment record indicates that the abundance of PSB is
112 dynamic over time with periods of near absence (McGowan et al. 2008), and that their
113 abundance is correlated with the abundance of ephippia, the dormant stages of *Daphnia*
114 (Anderson et al., submitted). Studies have shown that anaerobic bacteria communities can
115 contribute significantly to crustacean zooplankton diets (Overmann et al. 1999; Kankaala et
116 al. 2010). Moreover, *Daphnia* could be relying on PSB as either a direct or an indirect food
117 source (Massana et al. 1994). To exploit PSB, *Daphnia* require tolerance to hypoxia, to allow
118 them to graze in or near the anoxic zone at least for short periods of time. Hypoxia
119 tolerance could also have implications for thermal tolerance as both traits are strongly
120 associated with the efficiency of oxygen metabolism (Pörtner 2002).

121 In this study, we asked whether phenotypic evolution in an asexual population is fast
122 enough to track rapid environmental change. Given the recent temperature increase in the
123 area, we tested the possible adaptation of modern clones to higher temperatures. We
124 assessed thermal tolerance by measuring time to immobilisation (T_{imm}) of multiple clonal
125 lineages of two *Daphnia* temporal subpopulations (modern and historical) from Braya Sø. To
126 assess the genotypic variation present in the two temporal subpopulations, we performed
127 whole genome sequencing. Members of the current Braya Sø population had lower
128 respiration rates than populations in neighbouring holomictic lakes (Karapli-Petritsopoulou
129 et al. 2024). To further explore hypoxia tolerance in Braya Sø, we measured respiration rates
130 and determined the critical oxygen limit (P_{crit}) in members of both temporal subpopulations.
131 Respiration rate (i.e. oxygen consumption) was expected to be low in this population to
132 allow feeding on PSB in the hypoxic to anoxic zone (Seidl et al. 2005; Karapli-Petritsopoulou
133 et al. 2024). Similarly, a low P_{crit} would signal high hypoxia tolerance.

134

135 Methods

136 Environmental data

137 To determine whether there has been a change in summer water temperature at Braya Sø
138 (Lake code: SS4; 66°59'24.1"N 51°01'39.8"W) over the study period, we used the known
139 relationship between ice-out date (IoD; as day of the year) and mean July water
140 temperature (determined by *in situ* data loggers; see Anderson and Brodersen 2001: mean
141 July water temperature = $-0.1018 * [IoD] + 3820.4$; $r^2 = 0.641$; $P < 0.01$). In the Kangerlussuaq
142 area regional surveys have shown that mean May air temperature [May-T] determines ice-
143 out date: $IoD = -2.33 * [May-T] + 165.24$; $r^2 = 0.797$; $P < 0.05$. May air temperature was taken for

144 the years 2005-2016 from the DMI meteorological station at Kangerlussuaq airport. For the
145 period 2000–2004; (n=5) ice-out date was inferred from *in-situ* data loggers (measuring at
146 either 1 or 2 hourly intervals) confirmed by remote automatic cameras (see Anderson and
147 Brodersen 2001). For the period 2017-2023, ice-out was determined using satellite imagery
148 (Sentinel-2 which has a 2-day return frequency at this latitude (66 °N)); it was not possible
149 to determine ice-out for the lake in 2022 due to cloud cover during the relevant time period.
150 Between 2000 and 2020 ice-out date varied by 29 days (minimum day 141; maximum day
151 170) which gives an inferred mean July water temperature range of 11.04 to 13.77 °C.

152 [Field sampling](#)

153 Sediment from Braya Sø, South-West Greenland was sampled in April 2022 by extracting
154 three sediment cores with a Hon-Kajak sediment corer (diameter 9 cm), from the deepest
155 part of the lake, one meter apart from each other. Sediment cores were sectioned in the
156 field (0.5 cm interval) and kept at 4 °C until processing. In July 2022, we sampled adult
157 *Daphnia* from the entire water column using a 200 µm plankton net (diameter 25 cm).
158 Temperature and oxygen profiles of the water column were measured on the day of
159 *Daphnia* sampling (YSI 650 MDS multiprobe). In previous years, a purple sulfur bacteria
160 population was present near the oxycline between 9 and 14 m depth (Anderson,
161 unpublished data). After determining the depth of the oxycline, we sampled the water
162 column at multiple locations in 50 cm increments between 9 and 14 m, using a 5 L Van-Dorn
163 bottle (height 41.5 cm, diameter: 13 cm), but were unable to detect the PSB population.

164 [Sediment core dating](#)

165 Radiometric dating of the core was performed at the St. Croix Watershed Research Station
166 Minnesota, USA , following the method for Lead-210 dating described in Appleby (2001).
167 Results are shown in Supplementary Table S1.

168 [Resurrection of ephippia from the sediment](#)

169 *Daphnia* ephippia were removed from the sediment and decapsulated using watchmaker's
170 forceps. Eggs from individual ephippia were separated into 2 ml SSS-Medium (Saebelfeld et
171 al. 2017). Hatching conditions were 10 °C and a photoperiod of 18:6 light:dark. The
172 resurrected clones used in this study were all hatched from eggs retrieved from the 1 - 1.5
173 cm sediment section, which was dated to be from 2011 (supplementary table S1). Once
174 hatched, juveniles were transferred to 100 ml jars with SSS-Medium at 14 °C and 24h
175 dimmed light.

176 [Clonal cultures](#)

177 The clones used in this study were each established from one *Daphnia* female cultured in an
178 incubator at 14 °C under 24 h dimmed light for at least six months before the start of
179 measurements, equivalent to ca. 9 generations. They were fed with 1 mg C*L⁻¹ of
180 *Tetrademus obliquus* (Turpin) M.J. Wynne, 2016 three times a week. For the experiments
181 of this study we used 11 clones representing a historical subpopulation resurrected from a
182 sediment section dated to 2011 and 12 clones representing the modern subpopulation
183 (sampled in summer 2022).

184 Whole genome sequencing

185 Genomic DNA extraction from all 11 historical and 12 modern clones was performed using
186 the MasterPure™ Complete DNA and RNA Purification kit (Biozym) according to the
187 manufacturer's instructions. Before DNA extraction, *Daphnia* were treated with a mixture of
188 antibiotics (50 mg/L ampicillin and 50 mg/L tetracycline dissolved in SSS-Medium) and fed
189 with Sephadex G-50 beads (5g*L⁻¹) for two days prior to DNA extraction. Paired-end (PE)
190 sequencing libraries were prepared with the Illumina DNA Prep PCR free method (~300 bp
191 insert size). Sequencing was done on an Illumina NovaSeq 6000 S4 Flowcell with 150bp
192 paired-end-reads and an average of 6 Gb per sample. Library preparation and sequencing
193 were performed at the Competence Centre for Genomic Analysis (CCGA) in Kiel, Germany.

194 Bioinformatic analyses

195 Following O'Grady et al. (2022), we used FASTQC 0.11.9 (Andrews 2015) for quality control
196 of PE reads and TrimGalore 0.6.6 (Krueger et al. 2023) for adapter trimming with a Phred
197 score cutoff of 20 and discarding paired reads shorter than 35 bp after trimming. Reads
198 were mapped to the *Daphnia pulex* reference genome (NCBI accession GCF_021134715.1)
199 with BWA-MEM 0.7.17 (Li and Durbin 2009) and default parameters. Duplicate and
200 supplementary reads were removed with *markduplicates* (Picard Toolkit 2.26; Broad
201 Institute 2024) and *samtools view* (Samtools 1.10; Danecek et al. 2021). Single nucleotide
202 polymorphisms (SNPs) were called using *freebayes* 1.3.2 (Garrison and Marth 2012) with a
203 ploidy setting of 3, and excluding regions with extremely high coverage (-g 10,000). For this
204 variant call, only alignments with a mapping quality > 40 and alleles with a base quality > 24
205 were included. The options `--min-alternate-fraction` and `--min-alternate-count` were set to
206 0.05 and 5, respectively. Next, we filtered variants with *vcfilter* from *vsflib* 1.0.3 (Garrison
207 et al. 2022) using the parameters `QUAL > 1`, `QUAL/AO > 10`, `SAF > 0` and `SAR > 0`, `RPR > 1` and

208 RPL > 1. The above analyses were conducted within the High-Performance Computing
209 infrastructure at ZEDAT, Freie Universität Berlin (Bennett et al. 2020).

210 The last stage of filtering and further analysis was performed in R (R Core Team 2023). To
211 obtain the final SNP set we selected polymorphic, biallelic SNPs with a minor allele
212 frequency (MAF) of 0.15 and an average read depth per sample of 10 to 1000 reads using
213 the R package *SeqArray* 1.26.2 (Zheng et al. 2017). From this set we calculated pairwise
214 genome-wide identity-by-state (IBS) similarity with the *snpGdsIBS()* function of the *SeqArray*
215 package modified for triploid genomes (details in Karapli-Petritsopoulou et al. 2024). The
216 package *SNPrelate* (Zheng et al. 2012) was then used to compute the hierarchical cluster
217 analysis. Principal component analysis (PCA) of SNPs was computed with the package
218 *smartsnp* (Herrando-Pérez et al. 2021)

219 [Time to immobilization](#)

220 *Daphnia* of the 23 clones described above were used for time to immobilisation (T_{imm}) tests.
221 These were conducted in two batches with two experimental runs each. Each batch
222 contained a mixture of randomly chosen clones of both subpopulations. Each run had three
223 replicates of each clone from the respective batch (six replicates per clone in total). T_{imm} was
224 measured on pre-adults, aged 10-18 days.

225 *Measurement setup*

226 To measure T_{imm} we used a custom-made apparatus following the design from Burton et al.
227 (2018). The apparatus consisted of 45 6 ml vials with 1.6 cm diameter and 4 cm height that
228 were fitted in a plexiglass box of 28.5 cm x 17.5 cm x 5.5 cm in dimensions in five rows of 9
229 vials. The box was sealed with silicon and connected to a refrigerated water circulator
230 through two main openings as inflow and outflow on two sides. The two openings were

231 separated into five smaller ones per side placed between the rows of vials to allow for
232 optimal water flow and temperature control.

233 *Determination of CT_{Max}*

234 To establish the temperature of the T_{imm} assay we first determined average CT_{Max} in mixed
235 clone pre-adults from Braya SØ. For this test we used the apparatus described above
236 connected to the refrigerated water circulator. *Daphnia* were placed individually in each vial
237 and were observed throughout the trial. Starting from 18 °C we increased the water
238 temperature by 1 °C every 6 minutes. The CT_{Max} was reached when all *Daphnia* exhibited
239 immobilisation due to thermal stress, and was determined as 36 °C for the Braya SØ
240 *Daphnia*.

241 *T_{imm} measurement*

242 We measured T_{imm} on pre-adult *Daphnia* at 34 °C (two degrees below CT_{Max}), following
243 Burton et al. (2018). *Daphnia* were placed individually into the vials in a randomised order
244 and the temperature was increased from 18 °C to 34 °C within 16 minutes to avoid a sudden
245 heat shock. *Daphnia* swimming activity was recorded using a Sony Alpha AIII camera with a
246 Tamron HA036 objective mounted on a camera tripod (Hitchy, Hitchy Handy Stativ). A light
247 table (HSK, A4 LIGHT PAD) was positioned beneath the plate to enhance visibility. After all
248 *Daphnia* had reached T_{imm} , they were removed from the apparatus and photographed
249 under a dissecting scope (ZEISS, STEMI 508) at a magnification of 25x using the image
250 analysis software MikroLive v.5 (<https://www.mikroskopie.de/>). Body length was measured
251 from the top of the head to the apical base of the spine. T_{imm} of individual *Daphnia* was
252 determined using the video material and defined as the moment when no movement was
253 detected for at least 30 seconds.

254 Respiration rate

255 *Respirometry setup*

256 To estimate the respiration rate and the critical oxygen limit (P_{crit}) of single *Daphnia* we
257 measured oxygen consumption at 18 °C for six hours using a closed respirometry system
258 (Loligo® Systems, Denmark) consisting of a 200 µl 24-well plate fitted with oxygen sensor
259 spots and an SDR SensorDish® Reader (PreSens Precision Sensing GmbH, Germany). The
260 plate was sealed with a PCR film. Before adding individual *Daphnia* to each well, they were
261 filled with SSS medium. Air bubbles were removed from the wells which were then topped
262 up with medium before sealing the plate. To achieve a stable temperature, the 24-well plate
263 was immersed in a sealed flow-through waterbath (Loligo® Systems, Denmark) connected to
264 a circulating refrigerated system (circulator: Thermo Scientific™ HAAKE A10, thermostat:
265 Thermo Scientific™ HAAKE SC100). The MicroResp™ software (Loligo® Systems, Denmark)
266 was used to conduct the measurements.

267 To account for diffusion of oxygen through the PCR film, we measured oxygen differences at
268 18 °C in a separate run without *Daphnia* over four hours by placing air-saturated medium in
269 the wells and corrected the measurements with the diffusion rate (see below).

270 Due to the low fecundity and long time to maturation (aprox. 20 days) of *Daphnia* from the
271 study lake, entering all clones in the measurements synchronously was not possible. We
272 therefore performed the measurements within six weeks in two batches: one including all
273 resurrected (historical) clones and one with the clones collected from the water column in
274 2022 (modern clones). To control for possible batch effects, the historical clone SS4-5 was
275 included in both batches as a reference. We used pre-adults of 11 to 20 days old for our
276 measurements.

277 To measure *Daphnia* oxygen consumption, we separated each batch into six runs that were
278 performed during six consecutive days. Each run included one replicate of each clone
279 present in the batch, four replicates of the reference and four blank wells containing
280 medium from the *Daphnia* jars to account for microbial (background) respiration, yielding a
281 total of six replicates per clone. The positions of the clones in the plate were randomized
282 before each run using the Microresp software. After six hours, we measured the body
283 length of individual *Daphnia* as described above. In cases when a replicate was missing from
284 a run due to handling, an extra replicate of the same clone was included into the next run.

285 [Statistical data analysis](#)

286 All statistical analyses were performed in R (v. 4.3.0; R Core Team 2023).

287 *T_{imm}*

288 *Daphnia* identified as males and female *Daphnia* with eggs were excluded from further
289 analysis, as their different physiological state might have biased the results. Due to this, two
290 clones from the historical subpopulation had to be excluded from further analysis because
291 less than three replicates remained. Additionally, the reference clone SS4-5 was randomly
292 subsampled to six replicates to be included in the analysis. We estimated dry weight from
293 body length of each *Daphnia* using the available Length-Weight formula for this lake's
294 population (Dry Weight = $9.015362 * \text{Length}^{2.86448}$ (Karapli-Petritsopoulou et al. 2024). We
295 fitted two sets of linear mixed models with ML (package lmerTest, v. 3.1-3; Kuznetsova et al.
296 2017, based on lme4 v. 1.1-33; Bates et al. 2015) and compared them within each set using
297 a Likelihood-Ratio test. The models included either genetic cluster or subpopulation as a
298 fixed factor with the addition of weight in two of the models. The experimental run and
299 clone nested either within genetic cluster or subpopulation were used as random factors:

300 *Set 1:*

301 model gen0: $T_{imm} \sim 1 + (1|Run) + (1|Genetic\ Cluster:Clone)$

302 model gen1: $T_{imm} \sim Genetic\ Cluster + (1|Run) + (1|Genetic\ Cluster:Clone)$

303 model gen2: $T_{imm} \sim Genetic\ Cluster + Weight + (1|Run) + (1|Genetic\ Cluster:Clone)$

304

305 *Set 2:*

306 model sub0: $T_{imm} \sim 1 + (1|Run) + (1|Subpopulation:Clone)$

307 model sub1: $T_{imm} \sim Subpopulation + (1|Run) + (1|Subpopulation:Clone)$

308 model sub2: $T_{imm} \sim Subpopulation + Weight + (1|Run) + (1|Subpopulation:Clone)$

309

310 The models with genetic cluster failed to converge due to the random effect of Clone. We
311 therefore removed the random effect to reduce complexity and reran Set 1 (i.e. model gen1:
312 $T_{imm} \sim Genetic\ Cluster + (1|Run)$). The models picked by the Likelihood-Ratio test were re-
313 fitted with REML for reporting.

314 We used the package multcomp (v. 1.4-25; Hothorn et al. 2008) to perform post-hoc tests
315 with the Holm method for genetic cluster. Data visualisation was performed using the
316 ggplot2 (v. 3.5.1; Wickham 2016) and patchwork (v. 1.2.0; Pedersen 2024) packages.

317 *Respiration rate calculation and statistical analysis*

318 Rates of diffusion, background respiration, and *Daphnia* respiration were calculated with
319 the respR package (v. 2.3.1; Harianto et al. 2019). The respiration rates were adjusted for
320 the mean diffusion rate measured over 24 wells in four hours and the background
321 respiration of each run averaged over the blank wells. We estimated dry weight of *Daphnia*
322 based on their body length using the formula described above and adjusted the respiration

323 rate dividing by dry weight to obtain weight-specific respiration rate. To correct for batch
324 effect, we used the reference clone SS4-5. For this, we calculated a bias weight as follows:

325
$$\text{Bias} = \frac{|\overline{ref_{run}} - \overline{ref}|}{\overline{ref}}$$

326 where $\overline{ref_{run}}$ is the mean mass-specific respiration rate of the reference clone of a given
327 run and \overline{ref} is the overall mean of the reference mass-specific respiration rate. The bias was
328 calculated separately for each of the six runs. To obtain standardised mass-specific
329 respiration rates (in the following $\text{Resp}_{\text{standard}}$), we multiplied the mass-specific respiration
330 rate of each experimental *Daphnia* with the run-specific bias. The reference clone was
331 randomly subsampled to six replicates to be included in the analysis.

332 To assess whether genetic cluster or subpopulation were significant factors in explaining our
333 data we performed two model comparisons between a null model with only random effects
334 and a full model containing either genetic cluster or subpopulation as a fixed effect. The
335 random factors were experimental run and clone nested within subpopulation or genetic
336 cluster, respectively. The models were fitted using the lmerTest (v. 3.1-3; Kuznetsova et al.
337 2017) and the comparisons were done with the Likelihood-Ratio test.

338 *Set 1*

339 model gen0: $\text{Resp}_{\text{standard}} \sim 1 + (1 | \text{Run}) + (1 | \text{Genetic Cluster:Clone})$

340 model gen: $\text{Resp}_{\text{standard}} \sim \text{Genetic Cluster} + (1 | \text{Run}) + (1 | \text{Genetic Cluster:Clone})$

341 *Set 2*

342 model sub0: $\text{Resp}_{\text{standard}} \sim 1 + (1 | \text{Run}) + (1 | \text{Subpopulation:Clone})$

343 model sub: $\text{Resp}_{\text{standard}} \sim \text{Subpopulation} + (1 | \text{Run}) + (1 | \text{Subpopulation:Clone})$

344

345 Data visualisation was performed with ggplot2 (v. 3.5.1; Wickham 2016) and patchwork (v.
346 1.2.0; Pedersen 2024).

347 *P_{crit} estimation and statistical analysis*

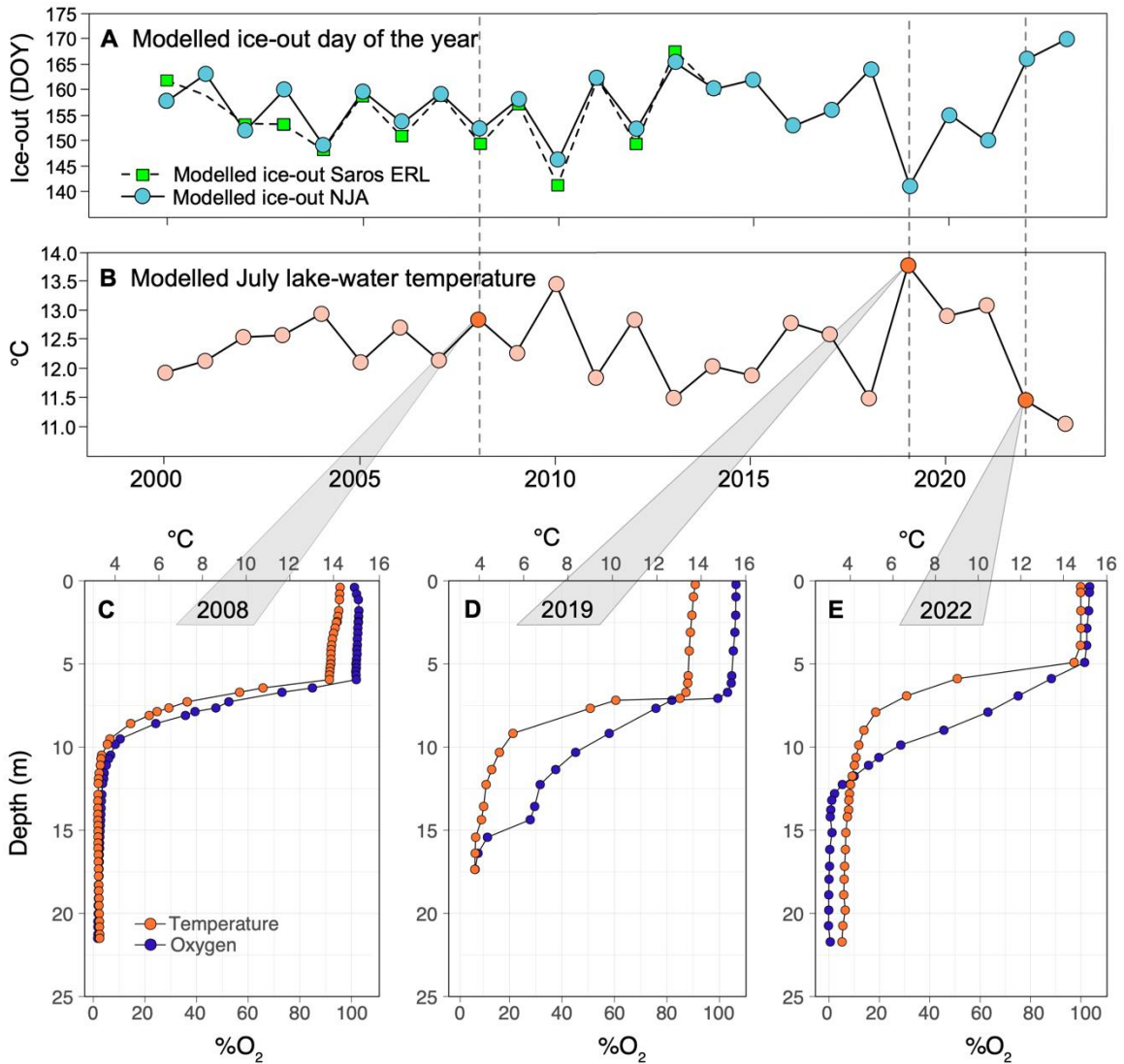
348 P_{crit} was estimated using the broken stick method from the respR package (v. 2.3.1; Harianto
349 et al. 2019). Many individuals did not reach P_{crit} and this resulted in an unbalanced dataset
350 between runs. Therefore, a linear mixed model approach was not possible. The P_{crit} of
351 genetic clusters and subpopulations was compared with one-way ANOVA and Welch's t-
352 test, respectively. Post-hoc tests were done using the Holm method. We tested the
353 relationship between weight adjusted respiration rates and P_{crit} with a linear regression
354 ($\log(P_{crit}) \sim \log(\text{weight_adj_rates})$).

355

356 Results

357 Environmental data

358 The modelled lake ice-out data show a pattern of increased variability after 2010 with both
359 extreme early and late dates present (Figure 1A). Extreme early and late ice-out dates
360 reflect higher and lower modelled July surface lake water temperatures, respectively (Figure
361 1B). Exemplary August lake profiles from an average year before the increase in variability
362 (2008) and two years after 2010 with an extremely early (2019) and an extremely late
363 (2022) ice-out date show a deeper oxycline in both years after 2010 with 2019 being the
364 deepest of the two (Figure 1D-E). The thermocline was at its deepest in 2019 followed by
365 2008 and 2022, matching the order from earlier ice-out to later.



366

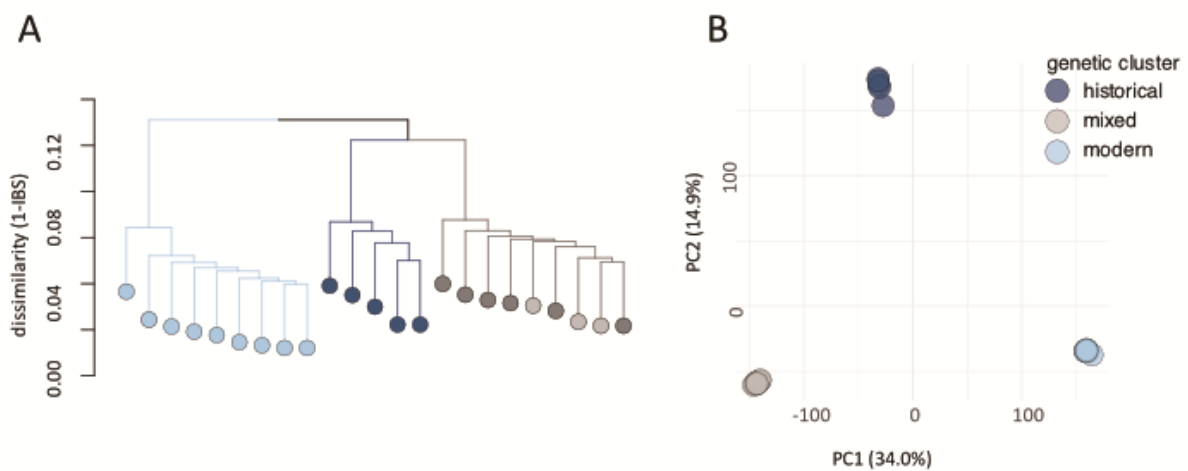
367 Figure 1. Change in ice-out, temperature and oxygen environment of Braya Sjø during the
 368 study period. (A) Modelled lake ice-out day of the year (DOY) estimated for this study is
 369 shown as blue circles from 2000 to 2023. For comparison, green squares with a dashed line
 370 show the ice-out day estimated by Saros et al. (2019). (B) Modelled average July surface lake
 371 water temperature. Dashed vertical lines in both (A) and (B) indicate the years 2008, 2019,
 372 and 2022, respectively. Depth profiles of August temperature (orange circles) and dissolved
 373 oxygen (DO % air saturation, purple circles) in Braya Sjø measured in the years 2008 (C),
 374 2019 (D) and 2022 (E).

375 Genomic results

376 A total of 365,466 SNPs remained after filtering and were used in further analysis. The
 377 hierarchical cluster analysis based on identity-by-state (IBS) identified three closely related
 378 genetic clusters: two clusters unique to each temporal subpopulation (a historical cluster
 379 with five of the 11 resurrected clones from 2011, and a modern cluster with nine clones

380 sampled in 2022), and a mixed cluster which included six of the resurrected and three of the
381 12 modern clones (Fig 2A, Figure S1). The historical and mixed genetic cluster were more
382 similar to each other with a dissimilarity range (1 - IBS) of 8.75% - 11.49%, while their
383 common dissimilarity range to the modern cluster was 8.91% - 13.33% (Table S2). These
384 ranges are relative to the total number of SNPs identified. Based on an estimated genome
385 size of 150 Mb, the highest dissimilarity of ~13% corresponds to a genome-wide IBS of
386 0.03%. The PCA confirmed the clusters with PCA1 separating mainly the mixed from the
387 modern cluster and explaining 34% of the variation, while PCA2 separated the historical
388 cluster from the other two and explained 14.9% of the variation (Fig. 2B).

389



390

391 Figure 2. Genetic clusters of clonal lineages based on identity-by-state (IBS). (A) Dendrogram
392 showing the results of hierarchical cluster analysis based on pairwise dissimilarity values (1-
393 IBS). From left to right the modern cluster is represented by light blue circles, the historical
394 cluster by dark blue and the mixed cluster by grey circles. The darker and light grey shade in
395 the mixed cluster represents clonal lineages from the historical and modern subpopulation
396 respectively. Each circle represents one clonal lineage. (B) Principal component analysis
397 (PCA) of all identified SNPs, using the same colour code as the dendrogram with one circle
398 per clone.

399

400 Time to immobilization (T_{imm})

401 Both predictor variables "subpopulation" and "genetic cluster" explained the data variation
 402 significantly better than the null model, while the addition of "Weight" as a fixed factor did
 403 not improve the model (Table 1a). Time to immobilization (T_{imm}) was significantly lower for
 404 the modern genetic cluster compared with the mixed and historical clusters (Fig. 3a, Table
 405 1b, Table S3). Additionally, the modern subpopulation showed a significantly lower T_{imm}
 406 compared to the historical subpopulation (Fig. 3b, Table 1c).

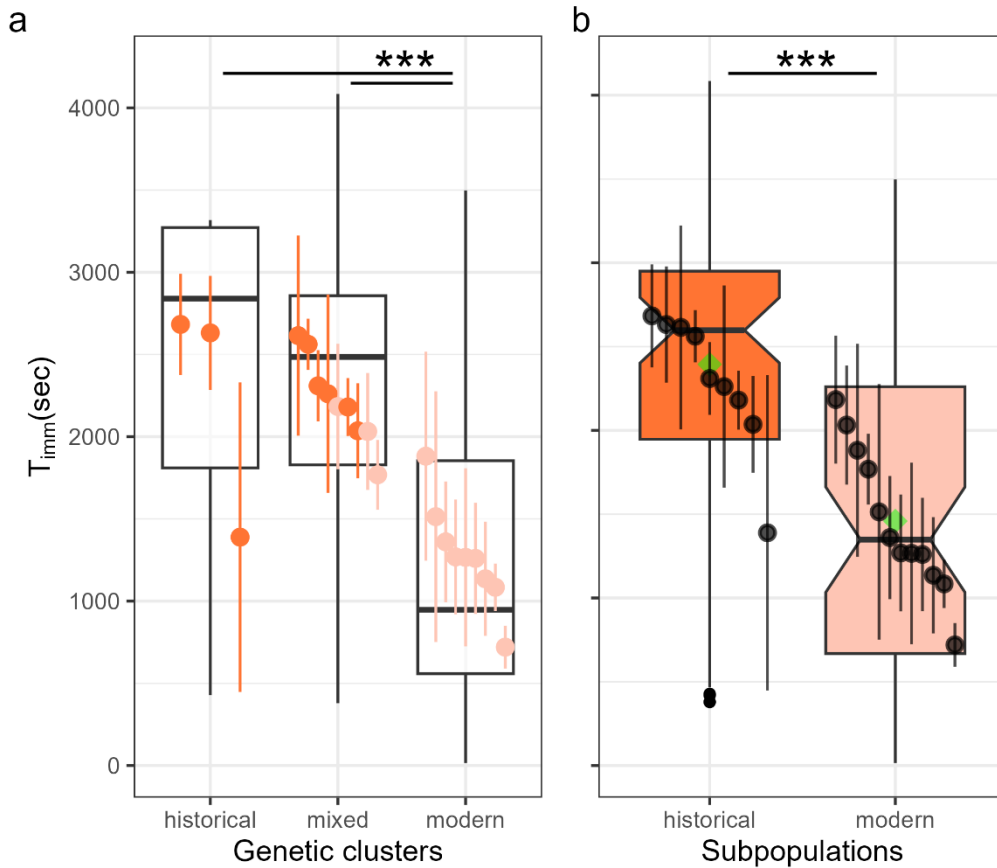
407 Table 1. Model comparison with Likelihood-Ratio Test (a) and model results for model gen1
 408 (b) and model sub1 (c) for time to immobilization (T_{imm}).

a. Model comparisons	npar	AIC	BIC	logLik	deviance	Chisq	Df	Pr(>Chisq)
<i>Set 1</i>								
null: $T_{imm} \sim 1 + (1 \text{Run})$	3	1869.76	1877.92	-931.88	1863.76			
gen1: $T_{imm} \sim \text{Genetic Cluster} + (1 \text{Run})$	5	1830.32	1843.91	-910.16	1820.32	43.45	2	3.675E-10
gen2: $T_{imm} \sim \text{Genetic Cluster} + \text{Weight} + (1 \text{Run})$	6	1832.18	1848.49	-910.09	1820.18	0.13	1	0.715
<i>Set 2</i>								
null: $T_{imm} \sim 1 + (1 \text{Run}) + (1 \text{Subpopulation:Clone})$	4	1855.03	1865.90	-923.51	1847.03			
sub1: $T_{imm} \sim \text{Subpopulation} + (1 \text{Run}) + (1 \text{Subpopulation:Clone})$	5	1842.44	1856.03	-916.22	1832.44	14.59	1	1.333E-04
sub2: $T_{imm} \sim \text{Subpopulation} + \text{Weight} + (1 \text{Run}) + (1 \text{Subpopulation:Clone})$	6	1844.36	1860.67	-916.18	1832.36	0.07	1	0.787
b. Model gen1: $T_{imm} \sim \text{Genetic Cluster} + (1 \text{Run}) + (1 \text{Genetic Cluster: Clone})$								
	Estimate	Std. Error	df	t value	Pr(> t)			
Intercept	2425.86	315.84	9.55	7.68	2.188E-05			
Genetic cluster mixed	-158.31	250.52	106.71	-0.63	0.529			
Genetic cluster modern	-1271.07	255.65	106.97	-4.97	2.539E-06			

c. Model sub1: $T_{imm} \sim \text{Subpopulation} + (1|\text{Run}) + (1|\text{Subpopulation:Clone})$

	Estimate	Std. Error	df	t value	Pr(> t)
Intercept	2361.71	253.45	5.39	9.32	1.594E-04
Subpopulation modern	-937.63	208.54	18.09	-4.50	2.765E-04

409



410

411 Figure 3. Time to immobilisation (T_{imm}) among genetic clusters (a) and between temporal
 412 subpopulations (b). The darker shades in both panels represent the historical subpopulation
 413 and the lighter shades the modern subpopulation. Point ranges in boxplots represent means
 414 per clone across replicates with SE. The boxplots are limited by the first and third quartiles,
 415 the middle line is the median and the whiskers are 1.5x interquartile ranges. Filled black
 416 circles at the bottom end of the whiskers represent outliers. In panel b green diamonds
 417 indicate mean values per subpopulation. Horizontal bars above the boxes represent post-
 418 hoc comparisons between genetic clusters (a, Holm method) and subpopulations (b, lmm
 419 model result). The statistical significance of the p-value is shown by asterisks (***)
 420 In panel a both comparisons share the same statistical significance.

421

422 **Respiration rate**

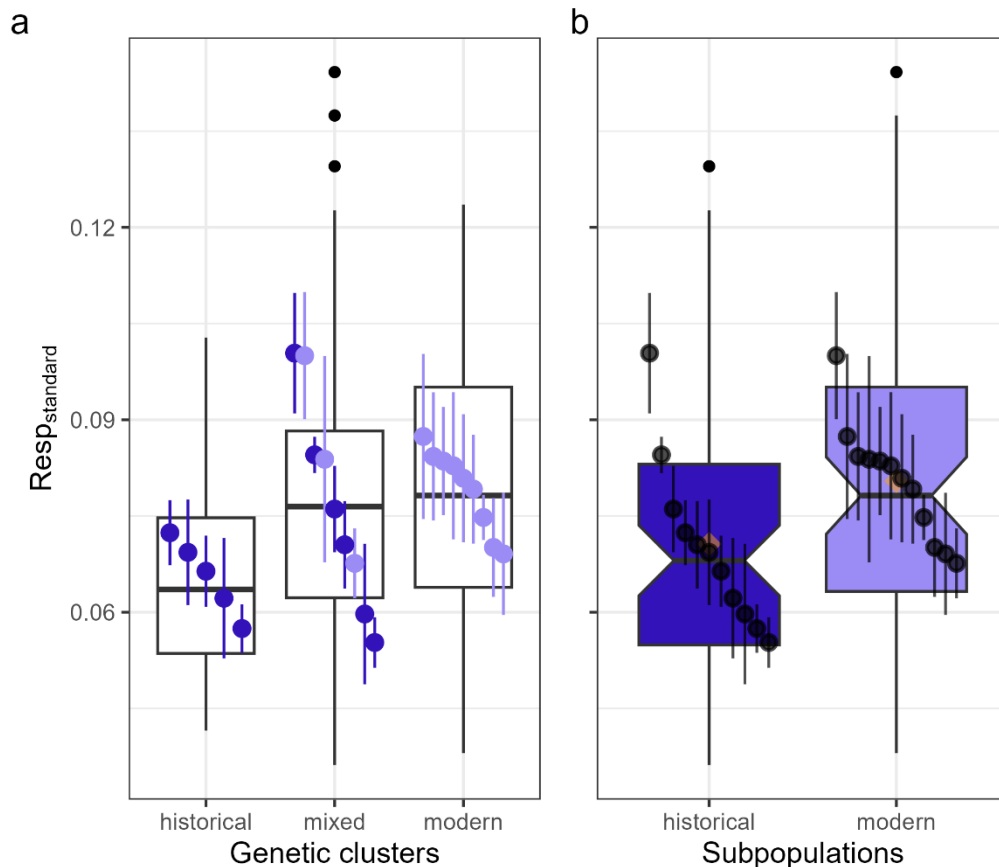
423 We observed a trend for respiration rates to be higher in the modern cluster (historical <
 424 mixed < modern) as well as in the modern compared with the historical subpopulation (Fig.
 425 4). These differences, however, were not statistically significant and the null models were
 426 selected by the Likelihood-Ratio test in both sets of comparisons (Table 2).

427

428 Table 2. Model comparison with Likelihood-Ratio test for respiration data.

Model comparison for genetic cluster	npar	AIC	BIC	logLik	deviance	Chisq	Df	Pr(>Chisq)
gen0: $\text{Resp}_{\text{standard}} \sim 1 + (1 \text{Run}) + (1 \text{Genetic Cluster:Clone})$	4	-655.82	-644.17	331.91	-663.82			
gen: $\text{Resp}_{\text{standard}} \sim \text{Genetic Cluster} + (1 \text{Run}) + (1 \text{Genetic Cluster:Clone})$	6	-655.08	-637.60	333.54	-667.08	3.25	2	0.197
Model comparison for subpopulation								
sub0: $\text{Resp}_{\text{standard}} \sim 1 + (1 \text{Run}) + (1 \text{Subpopulation:Clone})$	4	-655.82	-644.17	331.91	-663.82			
sub: $\text{Resp}_{\text{standard}} \sim \text{Subpopulation} + (1 \text{Run}) + (1 \text{Subpopulation:Clone})$	5	-656.07	-641.50	333.03	-666.07	2.24	1	0.134

429



430

431 **Figure 4.** Standardized respiration rate ($Resp_{standard}$) among genetic clusters (a) and between
 432 temporal subpopulations (b). The darker shade in both panels represents the historical
 433 subpopulation and the lighter shade the modern subpopulation. Point ranges show the
 434 mean of each clone across replicates and the standard error of the mean. The boxplot limits
 435 represent the first and third quartiles, the middle line indicating the median, whiskers are
 436 1.5x interquartile ranges. Filled black circles above the whiskers represent outliers. In panel
 437 b the mean values of the subpopulations are shown by orange diamonds.

438

439 P_{crit}

440 Most individuals did not reach the critical oxygen limit (P_{crit}) within the 6 h measurement

441 (Fig. 5a, 5c). This was mostly because low oxygen consumption prevented oxygen levels

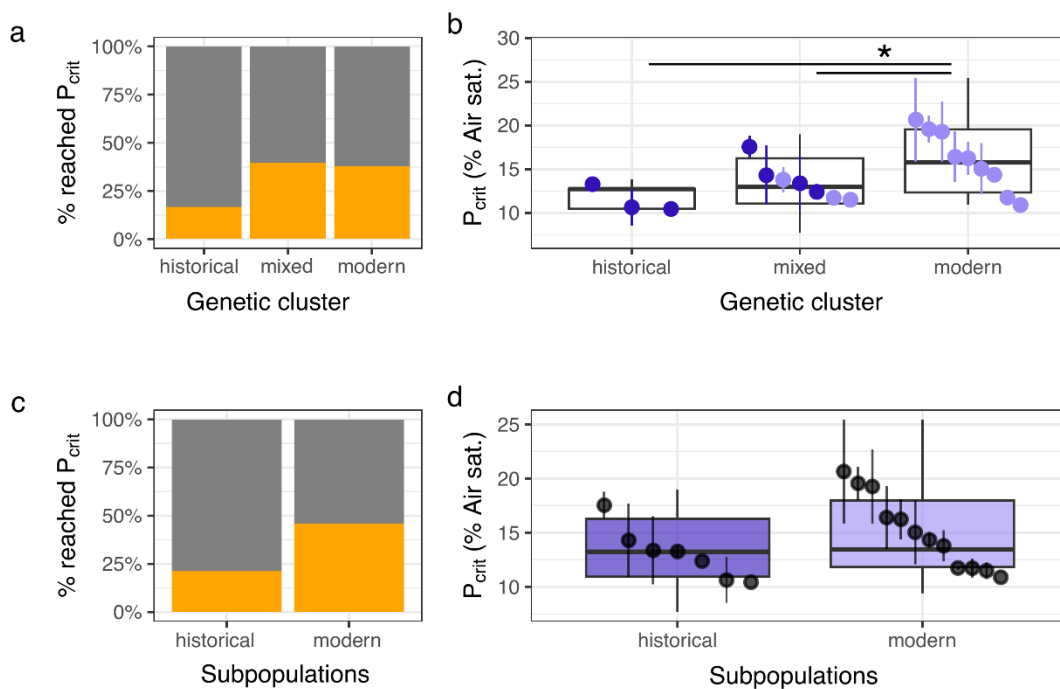
442 from approaching P_{crit} (see supplementary figure S2). Notably, fewer individuals from the

443 historical subpopulation reached P_{crit} , emphasizing the observed trend for lower respiration

444 rates of the historical subpopulation. The same pattern was observed for the historical

445 genetic cluster. Differences in P_{crit} were significant among genetic clusters ($F_{(2,43)} = 4.72$, $p =$

446 0.014) and the modern genetic cluster had a significantly higher P_{crit} than both other clusters
 447 (Fig. 5b, Table S4). P_{crit} did not significantly differ between subpopulations (Figure 5d; $t_{(25.52)}$
 448 = -1.1, $p = 0.28$). Finally, the overall regression between weight-adjusted respiration rates
 449 and P_{crit} was statistically significant with a positive relationship between the two (adj. $R^2 =$
 450 0.088, $F_{(1,44)} = 5.39$, $p = 0.024$; Figure S3), suggesting a significant association between the
 451 two variables.



452
 453 **Figure 5.** Critical oxygen limit in historical and modern subpopulations. Barplots of the
 454 percentage of replicates that reached the critical oxygen limit (P_{crit}) and boxplots of % air
 455 saturation values of P_{crit} per genetic cluster (a,b) and subpopulation (c,d). Percentages that
 456 reached P_{crit} are shown in orange. The historical and modern subpopulation are represented
 457 by darker and lighter purple shades, respectively. Point ranges are the means per clone with
 458 SE. The boxplots were computed as in Figure 4. Horizontal bars above the boxes represent
 459 post-hoc comparisons between genetic clusters (Holm method). The statistical significance
 460 of the p-value is shown by asterisks ($* < 0.05$). Both comparisons share the same statistical
 461 significance.

462

463 Discussion

464 We tested the adaptation potential of an asexual *Daphnia* population in response to rapid
465 environmental change, focussing on temperature rise and its relevance for hypoxia
466 tolerance. For this we applied a resurrection ecology approach to compare two temporal
467 snapshots of this population; resurrected *Daphnia* from a sediment layer dating to 2011 and
468 contemporary clonal lineages from the population sampled in 2022. Whole genome
469 sequencing of all clones participating in the experimental work of this study identified three
470 genetic clusters of which two are specific to each temporal subpopulation, and a third which
471 contains clones of both subpopulations. The genetic clusters and temporal subpopulations
472 provide similar interpretations of the results (except in the case of P_{crit}). We found that
473 modern and historical *Daphnia* differ in both thermal and hypoxia tolerance, albeit at
474 different magnitudes. Contrary to our prediction, T_{imm} was shorter for the modern *Daphnia*
475 (both subpopulation and genetic cluster), a proxy that indicates tolerance to high
476 temperatures. Additionally, we observed slightly higher respiration rates for the modern
477 subpopulation and significantly higher P_{crit} for the modern genetic cluster, both indicating a
478 lower tolerance to hypoxia. The combined pattern of low thermal and hypoxia tolerance
479 points to the shared physiological coping mechanisms that can enhance oxygen uptake (*e.g.*,
480 haemoglobin increase). We discuss these results in the light of the current climate change in
481 the study area that has significant effects on lake stratification and the lakes' oxygen
482 environment (Hazuková et al. 2024).

483 Our results suggest the occurrence of phenotypic change in an asexual *Daphnia* population
484 within a decade of environmental change among three closely related genetic clusters. The
485 high similarity seen among the clusters (highest pairwise dissimilarity between clusters was

486 13.33% within the SNP set, equivalent to 0.03% divergence genome-wide) could suggest
487 common ancestry via long-term molecular evolution, however the maximum number of
488 SNPs we found between clusters (13.33% corresponds to 48,717 SNPs) is elevated in
489 comparison to SNPs found between diploid Japanese obligate asexual lineages of *Daphnia*
490 *pulex* (~5,000), estimated to have diverged molecularly during the last ~300 years based on
491 nuclear substitutions (Ohtsuki et al. 2022). A higher SNP number could partly be explained
492 by the higher mutation rates occurring in polyploids in comparison to diploids (Meirmans et
493 al. 2018). Additionally, even though it is likely that polar regions including Greenland
494 harbour only obligate parthenogenetic *Daphnia* (Weider et al. 1999; Decaestecker et al.
495 2009; Haileselasie et al. 2016) and sexually reproducing *Daphnia* populations have not (yet)
496 been found in the study area, the possibility of local clonal recruitment through sexual
497 turnover cannot be completely excluded.

498 Two of the genetic clusters are specific to one of each time periods, and a third has
499 members of both periods. The most likely explanation would be an undetected, local
500 presence of members of the modern cluster at a low frequency in the historical
501 subpopulation, rising in frequency due to clonal selection. Alternatively, dispersal from
502 nearby lakes cannot be excluded. In both cases, the recent change in the clonal frequencies
503 contrast the results of a previous study (Dane et al. 2020) that showed the long-term
504 dominance of a single clone for at least the last 200 years until ~2008, identified by
505 microsatellite markers. A final possibility is that these genetic clusters are all closely related
506 to the dominant clone found previously, with the lower resolution of the microsatellite
507 analysis being unable to detect them.

508 A lower tolerance to high temperatures as observed in the modern clones (and more
509 pronounced in the modern genetic cluster) while temperature in the area is rising (Saros et
510 al. 2019), contradicts our expectations about the direction of phenotypic change. Climate
511 data point to an increased variability in yearly spring and summer temperatures rather than
512 a linear increase. While several of our August temperature profiles for SS4 do not show an
513 extreme difference in surface water temperatures, earlier ice-out is predicted to lead to a
514 deepening of surface water mixing, resulting in a shift of the thermocline and the oxycline to
515 greater water depths (Hazuková et al. 2024). Our data demonstrate such an expansion of
516 the higher surface temperatures during years of earlier ice-out, thus reducing the thermal
517 refugium for *Daphnia*. Higher temperatures near the water surface do not yet seem to be
518 directly affecting this population as the adaptations to hypoxia present in this population
519 may have shielded it against the recent higher temperature range. Nevertheless, further
520 warming and more frequent early ice-out events in future years, may catch up with the
521 population's thermal tolerance pushing it to a maladapted state.

522 The lower tolerance to hypoxia observed for the modern subpopulation and especially for
523 the modern genetic cluster may prevent *Daphnia* from successfully grazing near or in the
524 anoxic layer. It has been shown that *Daphnia* can utilize phototrophic purple sulfur bacteria
525 (PSB) as a food source (Jürgens et al. 1994; Massana et al. 1994). *Daphnia's* dependence on
526 PSB which are a supplement to the scarce algal food sources, may be stronger in Braya Sø or
527 other oligotrophic meromictic lakes occurring in the Kangerlussuaq area. However,
528 according to paleolimnological pigment analysis on the lake sediment of Braya Sø, the PSB
529 population has fluctuated in size in the past 500 years and seems to have been declining
530 since 2000 (Dane et al. 2020, Anderson et al., submitted). In line with these data, we were
531 unable to locate the PSB population during field sampling in August 2022, a season when

532 the bacterial plates of PSB are typically present. The presence of dense PSB bacterial plates
533 in August has for example been observed in Lake Cadagno, an alpine lake with climatic
534 conditions similar to those in West Greenland (Tonolla et al. 2003; Storelli et al. 2024). The
535 deepening of the oxycline observed in the 2019 and 2022 profiles might indicate an
536 important trend occurring since 2010, with early ice-out occurring more frequently.
537 According to Hazuková et al. (2024), earlier ice-out induces more prolonged spring mixing in
538 holomictic lakes before the summer stratification, expanding the presence of oxygen deeper
539 into the water column. Since phototrophic PSB rely on both sufficient light and anoxic
540 conditions (Overmann 2008), this shift in the oxycline is likely to destabilize the PSB
541 population, a situation that could be perpetuated in the following years.

542 The presence of fish in Braya Sø would be another reason for the *Daphnia* to descend closer
543 to the anoxic zone (Larsson and Lampert 2011). However, bigger *Daphnia* species like the
544 polyploid Arctic *D. pulicaria* and fish generally do not co-occur in lakes of West Greenland
545 (Jeppesen et al. 2017). Hence, a reduced or absent PSB layer could be a major reason for
546 tolerance to hypoxia no longer conferring an important benefit. The central mechanism for
547 adaptation to hypoxia is through elevated haemoglobin (Hb) production and oxygen affinity
548 (Paul et al. 2004; Seidl et al. 2005). Other mechanisms are changes in carbohydrate-
549 degrading enzymes, and a decrease in oxygen consumption and P_{crit} (Seidl et al. 2005; Zeis
550 et al. 2009). These adaptations involving elevated protein synthesis may confer costs (Pirow
551 et al. 2001), and therefore the relaxation of this selection pressure could explain a lower
552 hypoxia tolerance in the modern subpopulation.

553 Finally, a similar trend for the two measured traits (low tolerance to hypoxia combined with
554 low thermal tolerance and vice versa) suggests a relationship with the physiological

555 connection of the two traits as explained by the oxygen-limited thermal tolerance
556 hypothesis (Pörtner 2002). This hypothesis posits that the upper thermal limit of an
557 organism is set by oxygen limitation and transportation capacity and has found more
558 support in water-breathing arthropods in comparison to air-breathers (Verberk et al. 2016).
559 This has also been observed in *Daphnia* where elevated oxygen demand caused an increase
560 in Hb production with rising oxygen consumption at higher temperatures, similar to hypoxic
561 conditions (Fox and Phear 1953). Higher gene expression of haemoglobin has been
562 measured as a reaction to both hypoxia and raised temperature (Lamkemeyer et al. 2003;
563 Becker et al. 2011; Zeis 2020), while increased Hb levels were correlated with longer T_{imm}
564 (Yampolsky et al. 2014). This common mechanism suggests that a shift in hypoxia tolerance
565 in response to the environment would also affect thermal tolerance.

566 In conclusion, we observed phenotypic change over a short time period in an asexual
567 population of *Daphnia*. The direction of this change was, however, contrary to our
568 expectation. Even though temperature rise is one of the most prevalent factors of global
569 change, our findings suggest that at least until now temperature was not the driving stress
570 factor for this population. In contrast, the destabilisation of the PSB population and a thus
571 reduced demand for hypoxia tolerance may have outbalanced the costs of hypoxia
572 adaptation such as elevated haemoglobin production, leading simultaneously to a lower
573 thermal tolerance. The consequences of this phenotypic change could increase the
574 population's vulnerability to future warming. This finding further demonstrates the
575 complexity of the impacts of environmental change in ecosystems and how local
576 adaptations to unique environmental factors affect population responses. In addition, we
577 found three genetic clusters in this asexual population that may have resulted from

578 molecular evolution. Whether these genetic clusters are only found in this lake or have
579 dispersed from other regional lakes is unknown.

580 Our findings pose the question if the relatively small genome-wide differences among the
581 genomic clusters could explain the associated phenotypic divergence. In future research, it
582 will be important to disentangle the molecular underpinnings, in particular on gene
583 expression profiles, epigenetic modifications or ploidy-related effects which could underlie
584 the observed phenotypes. The presence of such mechanisms could be crucial for asexually
585 reproducing animals to keep up with the contemporary fast-paced environmental change,
586 ultimately allowing their survival.

587 [Author contributions](#)

588 **AKP:** Conceptualization (supporting); Formal analysis (lead); Investigation (lead);
589 Methodology (equal); Visualisation (equal); Writing-original draft preparation (lead);
590 Writing-Review & Editing (equal). **JJH:** Investigation (supporting); Methodology (supporting);
591 Writing-Review & Editing (supporting). **DB:** Conceptualization (supporting); Writing-Review
592 & Editing (equal). **NJA:** Conceptualization (supporting); Formal analysis (equal); Resources
593 (supporting); Visualisation (supporting); Writing-Review & Editing (equal). **DF:**
594 Conceptualization (lead); Formal analysis (equal); Funding acquisition (lead); Investigation
595 (supporting); Methodology (supporting); Project administration (lead); Resources (lead);
596 Supervision (lead); Visualisation (equal); Writing-original draft preparation (supporting);
597 Writing-Review & Editing (equal).

598

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612 References

- 613 Agashe, D., M. Sane, and S. Singhal. 2023. Revisiting the Role of Genetic Variation in Adaptation. *Am.*
614 *Nat.* **202**: 486–502. doi:10.1086/726012
- 615 Anderson, N. J., and K. P. Brodersen. 2001. Determining the date of ice-melt for low Arctic lakes
616 along Søndre Strømfjord, southern West Greenland. *Geol. Greenl. Surv. Bull.* **189**: 54–59.
617 doi:10.34194/ggub.v189.5156
- 618 Anderson, N. J., R. Harriman, D. B. Ryves, and S. T. Patrick. 2001. Dominant Factors Controlling
619 Variability in the Ionic Composition of West Greenland Lakes. *Arct. Antarct. Alp. Res.* **33**:
620 418–425. doi:10.1080/15230430.2001.12003450

621 Andrews, S. 2015. FASTQC A Quality Control tool for High Throughput Sequence Data. Babraham
622 Inst.

623 Appleby, P. G. 2001. Chronostratigraphic Techniques in Recent Sediments, p. 171–203. *In* W.M. Last
624 and J.P. Smol [eds.], *Tracking Environmental Change Using Lake Sediments: Basin Analysis,*
625 *Coring, and Chronological Techniques.* Springer Netherlands.

626 Bast, J. and others. 2018. Consequences of Asexuality in Natural Populations: Insights from Stick
627 Insects. *Mol. Biol. Evol.* **35**: 1668–1677. doi:10.1093/molbev/msy058

628 Bates, D., M. Mächler, B. Bolker, and S. Walker. 2015. Fitting Linear Mixed-Effects Models Using
629 lme4. *J. Stat. Softw.* **67**: 1–48. doi:10.18637/jss.v067.i01

630 Beaton, M. J., and P. D. N. Hebert. 1988. Geographical Parthenogenesis and Polyploidy in *Daphnia*
631 *pulex*. *Am. Nat.* **132**: 837–845. doi:10.1086/284892

632 Becker, D., B. F. Brinkmann, B. Zeis, and R. J. Paul. 2011. Acute changes in temperature or oxygen
633 availability induce ROS fluctuations in *Daphnia magna* linked with fluctuations of reduced
634 and oxidized glutathione, catalase activity and gene (haemoglobin) expression. *Biol. Cell* **103**:
635 351–363. doi:10.1042/BC20100145

636 Bennett, L., B. Melchers, and B. Proppe. 2020. Curta: A General-purpose High-Performance
637 Computer at ZEDAT, Freie Universität Berlin. doi:10.17169/refubium-26754

638 Bradshaw, W. E., and C. M. Holzapfel. 2006. Evolutionary Response to Rapid Climate Change. *Science*
639 **312**: 1477–1478. doi:10.1126/science.1127000

640 Broad Institute. 2024. Picard toolkit.

641 Burge, D. R. L., M. B. Edlund, and D. Frisch. 2018. Paleolimnology and resurrection ecology: The
642 future of reconstructing the past. *Evol. Appl.* **11**: 42–59. doi:10.1111/eva.12556

643 Bürger, R., and M. Lynch. 1995. Evolution and Extinction in a Changing Environment: A Quantitative-
644 Genetic Analysis. *Evolution* **49**: 151–163. doi:10.1111/j.1558-5646.1995.tb05967.x

645 Burton, T., B. Zeis, and S. Einum. 2018. Automated measurement of upper thermal limits in small
646 aquatic animals. 1-5. doi:10.1242/jeb.182386

647 Calvin, K. and others. 2023. IPCC, 2023: Climate Change 2023: Synthesis Report. Contribution of
648 Working Groups I, II and III to the Sixth Assessment Report of the Intergovernmental Panel
649 on Climate Change [Core Writing Team, H. Lee and J. Romero (eds.)]. IPCC, Geneva,
650 Switzerland. First. Intergovernmental Panel on Climate Change (IPCC).

651 Catullo, R. A., J. Llewelyn, B. L. Phillips, and C. C. Moritz. 2019. The Potential for Rapid Evolution
652 under Anthropogenic Climate Change. *Curr. Biol.* **29**: R996–R1007.
653 doi:10.1016/j.cub.2019.08.028

654 Chaturvedi, A. and others. 2021. Extensive standing genetic variation from a small number of
655 founders enables rapid adaptation in *Daphnia*. *Nat. Commun.* **12**: 4306. doi:10.1038/s41467-
656 021-24581-z

657 Chevin, L.-M., R. Lande, and G. M. Mace. 2010. Adaptation, Plasticity, and Extinction in a Changing
658 Environment: Towards a Predictive Theory. *PLOS Biol.* **8**: e1000357.
659 doi:10.1371/journal.pbio.1000357

660 Cousyn, C., L. De Meester, J. K. Colbourne, L. Brendonck, D. Verschuren, and F. Volckaert. 2001.
661 Rapid, local adaptation of zooplankton behavior to changes in predation pressure in the
662 absence of neutral genetic changes. *Proc. Natl. Acad. Sci.* **98**: 6256–6260.
663 doi:10.1073/PNAS.111606798

664 Crow, J. F. 1994. Advantages of sexual reproduction. *Dev. Genet.* **15**: 205–213.
665 doi:10.1002/dvg.1020150303

666 Dane, M., N. J. Anderson, C. L. Osburn, J. K. Colbourne, and D. Frisch. 2020. Centennial clonal
667 stability of asexual *Daphnia* in Greenland lakes despite climate variability. *Ecol. Evol.* **10**:
668 14178–14188. doi:10.1002/ece3.7012

669 Danecek, P. and others. 2021. Twelve years of SAMtools and BCFtools. *GigaScience* **10**: giab008.
670 doi:10.1093/gigascience/giab008

671 Decaestecker, E., L. De Meester, and J. Mergeay. 2009. Cyclical Parthenogenesis in *Daphnia*: Sexual
672 Versus Asexual Reproduction, p. 295–316. *In* *Lost Sex*. Springer Netherlands.

673 Decaestecker, E., S. Gaba, J. A. M. Raeymaekers, R. Stoks, L. Van Kerckhoven, D. Ebert, and L. De
674 Meester. 2007. Host–parasite ‘Red Queen’ dynamics archived in pond sediment. *Nat.* 2007
675 4507171 **450**: 870–873. doi:10.1038/nature06291

676 Dufresne, F., and P. D. N. Hebert. 1997. Pleistocene glaciations and polyphyletic origins of polyploidy
677 in an arctic cladoceran. *Proc. R. Soc. Lond. B Biol. Sci.* **264**: 201–206.
678 doi:10.1098/rspb.1997.0028

679 Finn, C., F. Grattarola, and D. Pincheira-Donoso. 2023. More losers than winners: investigating
680 Anthropocene defaunation through the diversity of population trends. *Biol. Rev.* **98**: 1732–
681 1748. doi:10.1111/brv.12974

682 Fox, H. M., and E. A. Phear. 1953. Factors influencing haemoglobin synthesis by *Daphnia*. *Proc. R.*
683 *Soc. Lond. Ser. B - Biol. Sci.* **141**: 179–189. doi:10.1098/rspb.1953.0034

684 Frisch, D., P. K. Morton, P. R. Chowdhury, B. W. Culver, J. K. Colbourne, L. J. Weider, and P. D.
685 Jeyasingh. 2014. A millennial-scale chronicle of evolutionary responses to cultural
686 eutrophication in *Daphnia*. *Ecol. Lett.* **17**: 360–368. doi:10.1111/ele.12237

687 Garant, D. 2020. Natural and human-induced environmental changes and their effects on adaptive
688 potential of wild animal populations. *Evol. Appl.* **13**: 1117. doi:10.1111/eva.12928

689 Garrison, E., Z. N. Kronenberg, E. T. Dawson, B. S. Pedersen, and P. Prins. 2022. A spectrum of free
690 software tools for processing the VCF variant call format: vcflib, bio-vcf, cyvcf2, hts-nim and
691 slivar. *PLOS Comput. Biol.* **18**: 1–15. doi:10.1371/journal.pcbi.1009123

692 Garrison, E., and G. Marth. 2012. Haplotype-based variant detection from short-read
693 sequencing. doi:10.48550/arXiv.1207.3907

694 Geerts, A. N. and others. 2015. Rapid evolution of thermal tolerance in the water flea *Daphnia*. *Nat.*
695 *Clim. Change* **5**: 665–668. doi:10.1038/nclimate2628

696 Haileselasie, T. H. and others. 2016. Environment not dispersal limitation drives clonal composition
697 of Arctic *Daphnia* in a recently deglaciated area. *Mol. Ecol.* **25**: 5830–5842.
698 doi:10.1111/mec.13843

699 Hairston, N. G. and others. 1999. Rapid evolution revealed by dormant eggs. *Nature* **401**: 446–446.
700 doi:10.1038/46731

701 Harianto, J., N. Carey, and M. Byrne. 2019. respR — An R package for the manipulation and analysis
702 of respirometry data. *Methods Ecol. Evol.* **10**: 912–920. doi:10.1111/2041-210X.13162

703 Hazuková, V., B. T. Burpee, R. M. Northington, N. J. Anderson, and J. E. Saros. 2024. Earlier ice melt
704 increases hypolimnetic oxygen despite regional warming in small Arctic lakes. *Limnol.*
705 *Oceanogr. Lett.* **9**: 258–267. doi:10.1002/lo2.10386

706 Herrando-Pérez, S., R. Tobler, and C. D. Huber. 2021. smartSNP, an R package for fast multivariate
707 analyses of big genomic data. *Methods Ecol. Evol.* **12**: 2084–2093. doi:10.1111/2041-
708 210X.13684

709 Hoffmann, A. A., and C. M. Sgrò. 2011. Climate change and evolutionary adaptation. *Nature* **470**:
710 479–485. doi:10.1038/nature09670

711 Hothorn, T., F. Bretz, and P. Westfall. 2008. Simultaneous Inference in General Parametric Models.
712 *Biom. J.* **50**: 346–363. doi:10.1002/bimj.200810425

713 Jaron, K. S., J. Bast, R. W. Nowell, T. R. Ranallo-Benavidez, M. Robinson-Rechavi, and T. Schwander.
714 2021. Genomic Features of Parthenogenetic Animals. *J. Hered.* **112**: 19–33.
715 doi:10.1093/jhered/esaa031

716 Jeppesen, E. and others. 2017. The structuring role of fish in Greenland lakes: an overview based on
717 contemporary and paleoecological studies of 87 lakes from the low and the high Arctic.
718 *Hydrobiologia* **800**: 99–113. doi:10.1007/s10750-017-3279-z

719 Jürgens, K., J. M. Gasol, R. Massana, and C. Pedròs-Aliò. 1994. Control of heterotrophic bacteria and
720 protozoans by *Daphnia pulex* in the epilimnion of Lake Cisó. *Arch. Für Hydrobiol.* **131**: 55–78.

721 Kankaala, P., S. Taipale, L. Li, and R. I. Jones. 2010. Diets of crustacean zooplankton, inferred from
722 stable carbon and nitrogen isotope analyses, in lakes with varying allochthonous dissolved
723 organic carbon content. *Aquat. Ecol.* **44**: 781–795. doi:10.1007/s10452-010-9316-x

724 Karapli-Petritsopoulou, A., J. J. Heckelmann, N. J. Anderson, S. Franzenburg, and D. Frisch. 2024.
725 Phenotypic divergence associated with genomic changes suggest local adaptation in obligate
726 asexuals. 2024.08.15.608098. doi:10.1101/2024.08.15.608098

727 Kearney, M. R., M. E. Jasper, V. L. White, I. J. Aitkenhead, M. J. Blacket, J. D. Kong, S. L. Chown, and A.
728 A. Hoffmann. 2022. Parthenogenesis without costs in a grasshopper with hybrid origins.
729 Science **376**: 1110–1114. doi:10.1126/science.abm1072

730 Kimura, M. 1967. On the evolutionary adjustment of spontaneous mutation rates. Genet. Res. **9**: 23–
731 34. doi:10.1017/S0016672300010284

732 Kočí, J. and others. 2020. No evidence for accumulation of deleterious mutations and fitness
733 degradation in clonal fish hybrids: Abandoning sex without regrets. Mol. Ecol. **29**: 3038–
734 3055. doi:10.1111/mec.15539

735 Krueger, F., F. James, P. Ewels, E. Afyounian, M. Weinstein, B. Schuster-Boeckler, G. Hulselmans, and
736 sclamons. 2023. FelixKrueger/TrimGalore: v0.6.10 - add default decompression
737 path.doi:10.5281/zenodo.7598955

738 Kuznetsova, A., P. B. Brockhoff, and R. H. B. Christensen. 2017. lmerTest package: Tests in linear
739 mixed effects models. J. Stat. Softw. **82**: 1–26. doi:10.18637/jss.v082.i13

740 Lamkemeyer, T., B. Zeis, and R. J. Paul. 2003. Temperature acclimation influences temperature-
741 related behaviour as well as oxygen-transport physiology and biochemistry in the water flea
742 *Daphnia magna*. Can. J. Zool. **81**: 237–249. doi:10.1139/z03-001

743 Larsson, P., and W. Lampert. 2011. Experimental evidence of a low-oxygen refuge for large
744 zooplankton. Limnol. Oceanogr. **56**: 1682–1688. doi:10.4319/lo.2011.56.5.1682

745 Li, H., and R. Durbin. 2009. Fast and accurate short read alignment with Burrows-Wheeler transform.
746 Bioinforma. Oxf. Engl. **25**: 1754–1760. doi:10.1093/bioinformatics/btp324

747 Lynch, M., R. Bürger, D. Butcher, and W. Gabriel. 1993. The Mutational Meltdown in Asexual
748 Populations. J. Hered. **84**: 339–344. doi:10.1093/oxfordjournals.jhered.a111354

749 MacColl, A. D. C. 2011. The ecological causes of evolution. *Trends Ecol. Evol.* **26**: 514–522.
750 doi:10.1016/j.tree.2011.06.009

751 Massana, R., J. M. Gasol, K. Jürgens, and C. Pedròs-Aliò. 1994. Impact of *Daphnia pulex* on a
752 metalimnetic microbial community. *J. Plankton Res.* **16**: 1379–1399.
753 doi:10.1093/plankt/16.10.1379

754 McGowan, S., R. K. Juhler, and N. J. Anderson. 2008. Autotrophic response to lake age, conductivity
755 and temperature in two West Greenland lakes. *J. Paleolimnol.* **39**: 301–317.
756 doi:10.1007/s10933-007-9105-2

757 Meirmans, P. G., S. Liu, and P. H. van Tienderen. 2018. The Analysis of Polyploid Genetic Data. *J.*
758 *Hered.* **109**: 283–296. doi:10.1093/jhered/esy006

759 Ohtsuki, H., H. Norimatsu, T. Makino, and J. Urabe. 2022. Invasions of an obligate asexual daphnid
760 species support the nearly neutral theory. *Sci. Rep.* **12**: 7305. doi:10.1038/s41598-022-
761 11218-4

762 Overmann, J. 2008. Ecology of Phototrophic Sulfur Bacteria, p. 375–396. *In* R. Hell, C. Dahl, D. Knaff,
763 and T. Leustek [eds.], *Sulfur Metabolism in Phototrophic Organisms. Advances in*
764 *Photosynthesis and Respiration*, vol 27.

765 Overmann, J., K. J. Hall, T. G. Northcote, and J. T. Beatty. 1999. Grazing of the copepod *Diatomus*
766 *connexus* on purple sulphur bacteria in a meromictic salt lake. *Environ. Microbiol.* **1**: 213–
767 221. doi:10.1046/j.1462-2920.1999.00026.x

768 Parmesan, C. 2006. Ecological and Evolutionary Responses to Recent Climate Change. *Annu. Rev.*
769 *Ecol. Evol. Syst.* **37**: 637–669. doi:10.1146/annurev.ecolsys.37.091305.110100

770 Paul, R. J., B. Zeis, T. Lamkemeyer, M. Seidl, and R. Pirow. 2004. Control of oxygen transport in the
771 microcrustacean *Daphnia*: regulation of haemoglobin expression as central mechanism of
772 adaptation to different oxygen and temperature conditions. *Acta Physiol. Scand.* **182**: 259–
773 275. doi:10.1111/j.1365-201X.2004.01362.x

774 Pedersen, T. L. 2024. patchwork: The composer of plots. manual.

775 Pirow, R., C. Bäumer, and R. J. Paul. 2001. Benefits of haemoglobin in the cladoceran crustacean
776 *Daphnia magna*. *J. Exp. Biol.* **204**: 3425–3441. doi:10.1242/jeb.204.20.3425

777 Pörtner, H. O. 2002. Climate variations and the physiological basis of temperature dependent
778 biogeography: systemic to molecular hierarchy of thermal tolerance in animals. *Comp.*
779 *Biochem. Physiol. A. Mol. Integr. Physiol.* **132**: 739–761. doi:10.1016/S1095-6433(02)00045-
780 4

781 R Core Team. 2023. R: a language and environment for statistical computing. manual R Foundation
782 for Statistical Computing.

783 Saebelfeld, M., L. Minguez, J. Griebel, M. O. Gessner, and J. Wolinska. 2017. Humic dissolved organic
784 carbon drives oxidative stress and severe fitness impairments in *Daphnia*. *Aquat. Toxicol.*
785 **182**: 31–38. doi:10.1016/j.aquatox.2016.11.006

786 Saros, J. E. and others. 2019. Arctic climate shifts drive rapid ecosystem responses across the West
787 Greenland landscape. *Environ. Res. Lett.* **14**: 074027. doi:10.1088/1748-9326/ab2928

788 Seidl, M. D., R. J. Paul, and R. Pirow. 2005. Effects of hypoxia acclimation on morpho-physiological
789 traits over three generations of *Daphnia magna*. *J. Exp. Biol.* **208**: 2165–2175.
790 doi:10.1242/jeb.01614

791 Šmejkalová, T., M. E. Edwards, and J. Dash. 2016. Arctic lakes show strong decadal trend in earlier
792 spring ice-out. *Sci. Rep.* **6**: 38449. doi:10.1038/srep38449

793 Storelli, N., O. S. Steiner, F. D. Nezio, S. Roman, A. Buetti-Dinh, and D. Bouffard. 2024. Ecological
794 dynamics of anoxygenic microorganisms in stably redox-stratified waters: Intra and inter-
795 seasonal variability of Lake Cadagno. doi:10.21203/rs.3.rs-3358324/v1

796 Tonolla, M., S. Peduzzi, D. Hahn, and R. Peduzzi. 2003. Spatio-temporal distribution of phototrophic
797 sulfur bacteria in the chemocline of meromictic Lake Cadagno (Switzerland). *FEMS*
798 *Microbiol. Ecol.* **43**: 89–98. doi:10.1111/j.1574-6941.2003.tb01048.x

799 Tucker, A. E., M. S. Ackerman, B. D. Eads, S. Xu, and M. Lynch. 2013. Population-genomic insights
800 into the evolutionary origin and fate of obligately asexual *Daphnia pulex*. *Proc. Natl. Acad.*
801 *Sci.* **110**: 15740–15745. doi:10.1073/pnas.1313388110

802 Verberk, W. C. E. P., J. Overgaard, R. Ern, M. Bayley, T. Wang, L. Boardman, and J. S. Terblanche.
803 2016. Does oxygen limit thermal tolerance in arthropods? A critical review of current
804 evidence. *Comp. Biochem. Physiol. A. Mol. Integr. Physiol.* **192**: 64–78.
805 doi:10.1016/j.cbpa.2015.10.020

806 Visser, M. E. 2008. Keeping up with a warming world; assessing the rate of adaptation to climate
807 change. *Proc. R. Soc. B Biol. Sci.* **275**: 649–659. doi:10.1098/rspb.2007.0997

808 Weider, L. J. 1987. Life history variation among low-arctic clones of obligately parthenogenetic
809 *Daphnia pulex*: a diploid-polyploid complex. *Oecologia* **73**: 251–256.
810 doi:10.1007/BF00377515

811 Weider, L. J., A. Hobæk, J. K. Colbourne, T. J. Crease, F. Dufresne, and P. D. N. Hebert. 1999. Holarctic
812 Phylogeography of an asexual species complex I. Mitochondrial DNA variation in arctic
813 *Daphnia*. *Evolution* **53**: 777–792. doi:10.1111/j.1558-5646.1999.tb05372.x

814 Weider, L. J., P. D. Jeyasingh, and D. Frisch. 2018. Evolutionary aspects of resurrection ecology:
815 Progress, scope, and applications—An overview. *Evol. Appl.* **11**: 3–10.
816 doi:10.1111/eva.12563

817 Wersebe, M. J., and L. J. Weider. 2023. Resurrection genomics provides molecular and phenotypic
818 evidence of rapid adaptation to salinization in a keystone aquatic species. *Proc. Natl. Acad.*
819 *Sci.* **120**: e2217276120. doi:10.1073/pnas.2217276120

820 Wickham, H. 2016. *ggplot2: Elegant graphics for data analysis*, Springer-Verlag New York.

821 Xu, S., A. R. Omilian, and M. E. Cristescu. 2011. High Rate of Large-Scale Hemizygous Deletions in
822 Asexually Propagating *Daphnia*: Implications for the Evolution of Sex. *Mol. Biol. Evol.* **28**:
823 335–342. doi:10.1093/molbev/msq199

824 Yampolsky, L. Y., T. M. M. Schaer, and D. Ebert. 2014. Adaptive phenotypic plasticity and local
825 adaptation for temperature tolerance in freshwater zooplankton. *Proc. R. Soc. B Biol. Sci.*
826 **281**: 20132744. doi:10.1098/rspb.2013.2744

827 Yousey, A. M., P. R. Chowdhury, N. Biddinger, J. H. Shaw, P. D. Jeyasingh, and L. J. Weider. 2018.
828 Resurrected 'ancient' *Daphnia* genotypes show reduced thermal stress tolerance compared
829 to modern descendants. *R. Soc. Open Sci.* **5**: 172193. doi:10.1098/rsos.172193

830 Zeis, B. and others. 2009. Acclimatory responses of the *Daphnia pulex* proteome to environmental
831 changes. I. Chronic exposure to hypoxia affects the oxygen transport system and
832 carbohydrate metabolism. *BMC Physiol.* **9**. doi:10.1186/1472-6793-9-7

833 Zeis, B. 2020. Hemoglobin in Arthropods—*Daphnia* as a Model, p. 163–194. *In* U. Hoeger and J.R.
834 Harris [eds.], *Vertebrate and Invertebrate Respiratory Proteins, Lipoproteins and other Body*
835 *Fluid Proteins*. Springer International Publishing.

836 Zheng, X., S. M. Gogarten, M. Lawrence, A. Stilp, M. P. Conomos, B. S. Weir, C. Laurie, and D. Levine.
837 2017. SeqArray—a storage-efficient high-performance data format for WGS variant calls.
838 *Bioinformatics* **33**: 2251–2257. doi:10.1093/bioinformatics/btx145

839 Zheng, X., D. Levine, J. Shen, S. M. Gogarten, C. Laurie, and B. S. Weir. 2012. A high-performance
840 computing toolset for relatedness and principal component analysis of SNP data.
841 *Bioinformatics* **28**: 3326–3328. doi:10.1093/bioinformatics/bts606

842

Supplementary Material

Altered phenotypic responses of asexual Arctic *Daphnia* after 10 years of rapid climate change

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Supplementary Figures

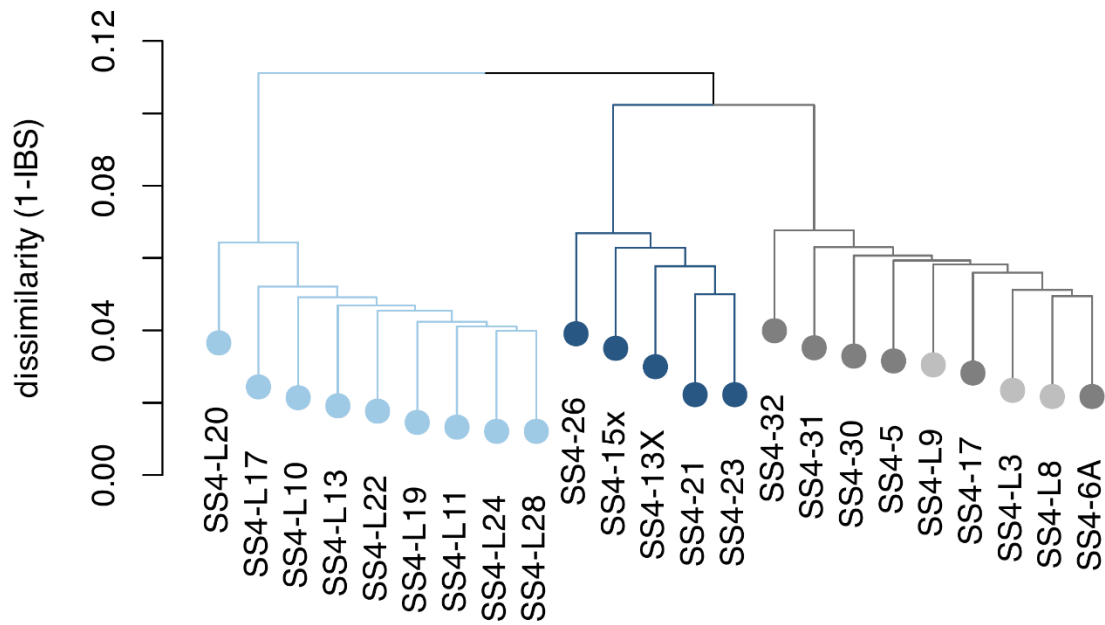


Figure S1. Dendrogram of genetic clusters of clonal lineages based on identity-by-state (IBS) labeled for clone. Dark and light shades represent historical and modern clones respectively. The historical cluster is shown in dark blue, the modern cluster in light blue and the mixed cluster in grey.

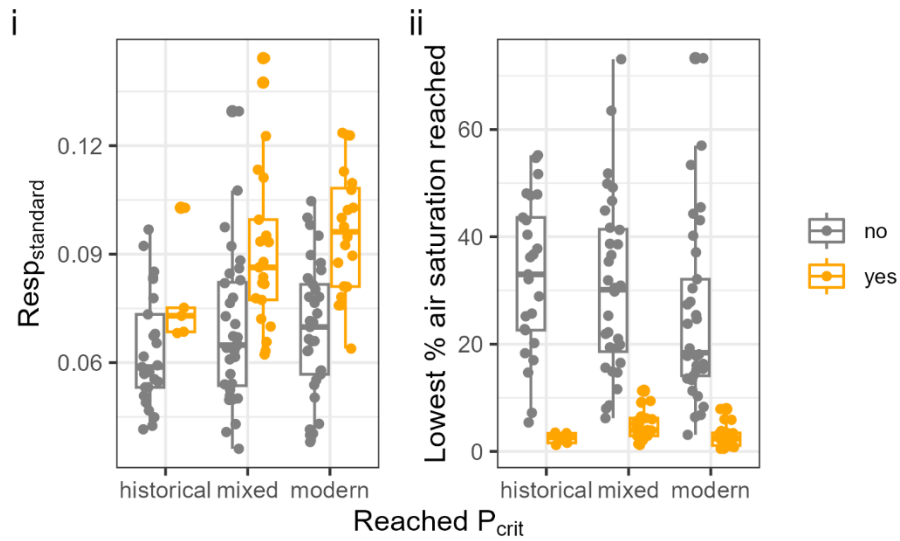


Figure S2. $Resp_{standard}$ (i) and lowest % air saturation value reached at the end of six hours (ii) per genetic cluster according to whether individuals reached P_{crit} . Individuals where P_{crit} was successfully measured are coloured with orange, while grey points show non-obtained P_{crit} data.

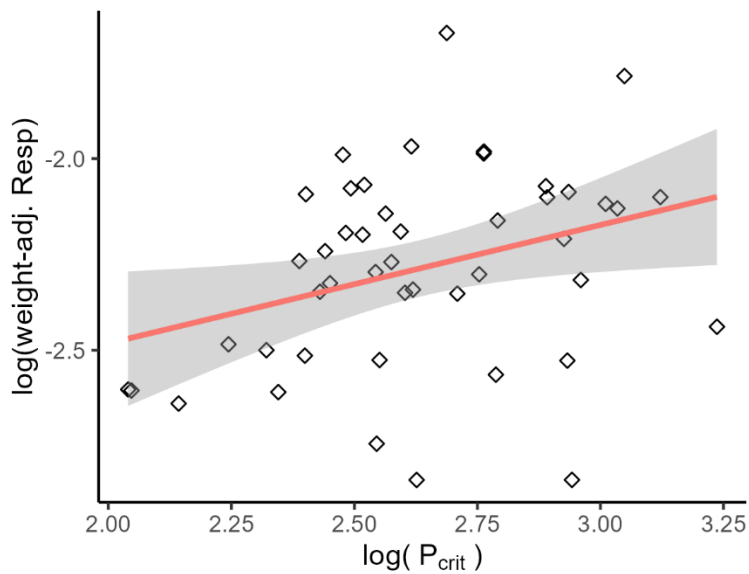


Figure S3. Linear regression between weight-adjusted respiration rates and P_{crit} (model: $\log(P_{crit}) \sim \log(\text{weight_adj_rates})$).

Supplementary Tables

Table S1. Lead-210 dating of sediment core where *Daphnia* were hatched from.

Top of Interval	Base of Interval	Mid of Interval	Total 210Pb	Error of Total Pb	Cum. Dry Mass	Unsup. 210Pb	Error of Unsup Pb	Age: Base of Int.	Error of Age	Date: Base A.D.	Date: Mid A.D.	Sediment DMAR	Error of DMAR
(cm)	(cm)	(cm)	(pCi/g)	(±s.d.)	(g/cm ²)	(pCi/g)	(±s.d.)	(yr)	(±s.d.)			(g/cm ² yr)	(±s.d.)
0	0.5	0.25	10.520	0.425	0.032	10.294	0.425	3.96	0.76	2018.4	2020.4	0.0081	0.0003
0.5	1	0.75	9.311	0.274	0.056	9.085	0.274	6.92	0.80	2015.4	2016.9	0.0082	0.0003
1	1.5	1.25	8.395	0.309	0.096	8.169	0.309	11.85	0.86	2010.5	2013.0	0.0081	0.0003
1.5	2	1.75	9.362	0.257	0.137	9.137	0.257	18.71	0.97	2003.6	2007.1	0.0060	0.0002
2	2.5	2.25	5.849	0.189	0.191	5.624	0.189	25.54	1.10	1996.8	2000.2	0.0079	0.0003
2.5	3	2.75	3.602	0.124	0.287	3.377	0.125	34.92	1.34	1987.4	1992.1	0.0102	0.0005
3	3.5	3.25	3.066	0.111	0.366	2.841	0.112	43.52	1.66	1978.8	1983.1	0.0092	0.0005
3.5	4	3.75	2.094	0.047	0.437	1.869	0.049	49.90	1.98	1972.4	1975.6	0.0111	0.0007
4	4.5	4.25	1.570	0.052	0.562	1.345	0.054	60.54	2.66	1961.8	1967.1	0.0118	0.0009
4.5	5	4.75	1.019	0.029	0.735	0.794	0.032	72.92	3.83	1949.4	1955.6	0.0140	0.0015
5	5.5	5.25	0.711	0.022	0.893	0.486	0.026	82.63	5.12	1939.7	1944.6	0.0162	0.0023
5.5	6	5.75	0.541	0.018	1.039	0.316	0.023	90.35	6.47	1932.0	1935.9	0.0190	0.0036
6	6.5	6.25	0.514	0.017	1.195	0.288	0.022	100.23	8.74	1922.1	1927.1	0.0158	0.0038
6.5	7	6.75	0.346	0.011	1.351	0.121	0.018	105.47	10.25	1916.9	1919.5	0.0297	0.0096
7	7.5	7.25	0.443	0.016	1.511	0.218	0.021	118.33	15.21	1904.0	1910.5	0.0125	0.0050
7.5	8	7.75	0.380	0.012	1.662	0.154	0.019	131.01	22.50	1891.3	1897.7	0.0119	0.0070
8	8.5	8.25	0.336	0.010	1.799	0.111	0.017	143.27	32.86	1879.1	1885.2	0.0112	0.0096

Table S2 Pair-wise dissimilarity matrix (1-IBS) between clones. Clone labels are coloured according to genetic cluster and subpopulation membership (see colour code in Figure S1).

	SS4-L20	SS4-L17	SS4-L10	SS4-L13	SS4-L22	SS4-L19	SS4-L11	SS4-L24	SS4-L28	SS4-32	SS4-31	SS4-30	SS4-5	SS4-L9	SS4-17	SS4-L3	SS4-L8	SS4-6A	SS4-26	SS4-15x	SS4-13X	SS4-21	SS4-23
SS4-L20	0	0.06922	0.06734	0.06567	0.06503	0.06257	0.06241	0.06068	0.06173	0.13333	0.13007	0.12857	0.12787	0.1279	0.12673	0.12343	0.12319	0.12162	0.12488	0.12346	0.12126	0.1092	0.11942
SS4-L17	0.06922	0	0.05577	0.05462	0.05374	0.05127	0.05055	0.04894	0.05017	0.12342	0.12053	0.11869	0.11801	0.11794	0.11699	0.11325	0.11279	0.11117	0.11521	0.11351	0.11101	0.09848	0.10937
SS4-L10	0.06734	0.05577	0	0.05206	0.05113	0.04877	0.04854	0.04669	0.0478	0.12136	0.11798	0.11655	0.11574	0.11576	0.11494	0.11101	0.11066	0.10901	0.11276	0.11114	0.10909	0.09633	0.10701
SS4-L13	0.06567	0.05462	0.05206	0	0.04993	0.04707	0.0468	0.04483	0.04601	0.12054	0.11709	0.11521	0.11447	0.11506	0.11346	0.10976	0.10932	0.10788	0.11201	0.10998	0.10782	0.09482	0.10566
SS4-L22	0.06503	0.05374	0.05113	0.04993	0	0.04622	0.0461	0.04411	0.04535	0.11982	0.11657	0.11475	0.11386	0.11404	0.11279	0.10905	0.10861	0.10697	0.1114	0.1093	0.10731	0.09432	0.10504
SS4-L19	0.06257	0.05127	0.04877	0.04707	0.04622	0	0.04331	0.04128	0.04247	0.1176	0.11433	0.11243	0.11178	0.11179	0.11084	0.10671	0.10645	0.10455	0.10927	0.10712	0.10467	0.0914	0.10268
SS4-L11	0.06241	0.05055	0.04854	0.0468	0.0461	0.04331	0	0.04055	0.04152	0.11741	0.1137	0.11177	0.11111	0.11092	0.10984	0.10646	0.10577	0.10403	0.10889	0.10711	0.10457	0.09125	0.10283
SS4-L24	0.06068	0.04894	0.04669	0.04483	0.04411	0.04128	0.04055	0	0.03987	0.11574	0.11195	0.10981	0.10936	0.10925	0.10807	0.10445	0.10396	0.10235	0.10723	0.10563	0.10306	0.08912	0.10113
SS4-L28	0.06173	0.05017	0.0478	0.04601	0.04535	0.04247	0.04152	0.03987	0	0.11675	0.113	0.111	0.11019	0.11023	0.10916	0.10569	0.10538	0.10355	0.10813	0.1063	0.10415	0.0902	0.10214
SS4-32	0.13333	0.12342	0.12136	0.12054	0.11982	0.1176	0.11741	0.11574	0.11675	0	0.07247	0.07044	0.06928	0.06963	0.06891	0.06432	0.06402	0.06242	0.11486	0.11241	0.11064	0.09915	0.10835
SS4-31	0.13007	0.12053	0.11798	0.11709	0.11657	0.11433	0.1137	0.11195	0.113	0.07247	0	0.06638	0.06544	0.06529	0.06433	0.06067	0.06009	0.05872	0.11105	0.10958	0.10741	0.09498	0.1056
SS4-30	0.12857	0.11869	0.11655	0.11521	0.11475	0.11243	0.11177	0.10981	0.111	0.07044	0.06638	0	0.06363	0.06384	0.06265	0.05887	0.05818	0.05691	0.1096	0.10792	0.10567	0.09304	0.10398
SS4-5	0.12787	0.11801	0.11574	0.11447	0.11386	0.11178	0.11111	0.10936	0.11019	0.06928	0.06544	0.06363	0	0.06304	0.06208	0.05812	0.05748	0.05595	0.10855	0.10689	0.10466	0.0923	0.10244
SS4-L9	0.1279	0.11794	0.11576	0.11506	0.11404	0.11179	0.11092	0.10925	0.11023	0.06963	0.06529	0.06384	0.06304	0	0.06183	0.05784	0.05747	0.05589	0.10901	0.10717	0.10492	0.09223	0.10315
SS4-17	0.12673	0.11699	0.11494	0.11346	0.11279	0.11084	0.10984	0.10807	0.10916	0.06891	0.06433	0.06265	0.06208	0.06183	0	0.05688	0.05646	0.05465	0.10796	0.10597	0.104	0.09149	0.10206
SS4-L3	0.12343	0.11325	0.11101	0.10976	0.10905	0.10671	0.10646	0.10445	0.10569	0.06432	0.06067	0.05887	0.05812	0.05784	0.05688	0	0.0522	0.05021	0.10441	0.10233	0.10037	0.08752	0.09811
SS4-L8	0.12319	0.11279	0.11066	0.10932	0.10861	0.10645	0.10577	0.10396	0.10538	0.06402	0.06009	0.05818	0.05748	0.05747	0.05646	0.0522	0	0.0495	0.10402	0.10196	0.09987	0.08698	0.09781
SS4-6A	0.12162	0.11117	0.10901	0.10788	0.10697	0.10455	0.10403	0.10235	0.10355	0.06242	0.05872	0.05691	0.05595	0.05589	0.05465	0.05021	0.0495	0	0.10208	0.10045	0.09823	0.08538	0.09626
SS4-26	0.12488	0.11521	0.11276	0.11201	0.1114	0.10927	0.10889	0.10723	0.10813	0.11486	0.11105	0.1096	0.10855	0.10901	0.10796	0.10441	0.10402	0.10208	0	0.07249	0.06994	0.05697	0.06804
SS4-15x	0.12346	0.11351	0.11114	0.10998	0.1093	0.10712	0.10711	0.10563	0.1063	0.11241	0.10958	0.10792	0.10689	0.10717	0.10597	0.10233	0.10196	0.10045	0.07249	0	0.06813	0.05514	0.06521
SS4-13X	0.12126	0.11101	0.10909	0.10782	0.10731	0.10467	0.10457	0.10306	0.10415	0.11064	0.10741	0.10567	0.10466	0.10492	0.104	0.10037	0.09987	0.09823	0.06994	0.06813	0	0.0523	0.06319
SS4-21	0.1092	0.09848	0.09633	0.09482	0.09432	0.0914	0.09125	0.08912	0.0902	0.09915	0.09498	0.09304	0.0923	0.09223	0.09149	0.08752	0.08698	0.08538	0.05697	0.05514	0.0523	0	0.05002
SS4-23	0.11942	0.10937	0.10701	0.10566	0.10504	0.10268	0.10283	0.10113	0.10214	0.10835	0.1056	0.10398	0.10244	0.10315	0.10206	0.09811	0.09781	0.09626	0.06804	0.06521	0.06319	0.05002	0

Table S3. Post-hoc test among genetic clusters for $\text{gen1}(T_{\text{imm}} \sim \text{Genetic Cluster} + (1 | \text{Run}))$ for T_{imm} data using Holm method.

Gen. cluster comparison	Estimate	Std. Error	z value	Pr(> z)
mixed - historical	-158.3	250.5	-0.63	0.527
modern -historical	-1271.1	255.7	-4.97	1.33E-06
modern - mixed	-1112.8	62.8	-6.84	2.45E-11

Table S4. Post-hoc test among genetic clusters (for ANOVA model $P_{\text{crit}} \sim \text{genetic cluster}$) for P_{crit} using Holm method.

Gen. cluster comparison	Estimate	Std. Error	t value	Pr(> t)
mixed - historical	1.72	1.86	0.93	0.621
modern -historical	4.65	1.87	2.49	0.042
modern - mixed	2.93	1.17	2.51	0.040

