

1 A State and Transition framework to guide riparian woodland vegetation 2 management and environmental water decisions

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9 **Keywords:** Environmental flows, macrophytes, waterway, river regulation, STM, decision
10 trees, lowland rivers

11 **Abstract**

12 River regulation and water extraction are major threats to the health and persistence of
13 water-dependent ecosystems, such as riparian woodlands and forests. In heavily
14 modified agricultural landscapes, riparian vegetation is also impacted by site-level
15 stressors like livestock grazing, tree clearing, and weed invasions. Complex interactions
16 among spatial and temporal drivers in water-dependent ecosystems can result in poorly
17 articulated conservation objectives and inefficient or siloed management decisions.
18 Where restoration funds and environmental water allocations are limited, these
19 inefficiencies are magnified. We propose a management-focused State and Transition
20 Model to describe the expected interactions among management at different spatial
21 scales, develop measurable objectives and implement targeted monitoring. Derived from
22 a multi-community eucalypt woodland model, the riparian State and Transition Model

23 further refines the states to better describe key indicators of riparian condition and
24 provides a catalogue of transitions describing the key site-level interventions, biotic
25 processes and changes to flows that are expected to drive changes from one state to
26 another. This resource can be used to support spatially explicit strategies and
27 prioritisation of environmental flows or other management actions to improve vegetation
28 condition along regulated waterways. Additionally, we demonstrate how the riparian State
29 and Transition Model can be used for structured decision making, targeted monitoring and
30 adaptive management by land and waterway managers.

31 **Introduction**

32 Water extraction supports human communities and global industries, but changes to river
33 flows and catchment-scale degradation are a threat to global riverine ecosystem integrity
34 and biodiversity (Bunn & Arthington 2002; Millennium Ecosystem Assessment 2005;
35 Richardson et al. 2007; Bernhardt & Palmer 2011). Waterway management occurs at local
36 and catchment scales, often at great expense due to the high cost of infrastructure, water,
37 or intensive restoration (Richardson et al. 2007). Effects of management and restoration
38 efforts in water dependant ecosystems are difficult to predict accurately due to the
39 complex interactions among spatial and temporal variables and regimes (Campbell et al.,
40 2023; Overton et al., 2014). For instance, at the catchment scale, river regulation
41 influences the entire flow regime (e.g. frequency, magnitude, timing and duration of
42 flows), causing local and catchment-wide impacts on riverine biota. These flow impacts
43 interact with local site conditions, such as livestock grazing, channel morphology and
44 land clearing, resulting in site-level responses that may be exacerbated by facilitative
45 impacts or offset by opposing impacts. As such, most attempts to understand and

46 manage riparian and floodplain ecosystems focus on hydrology or land management
47 separately (González et al., 2015).

48 The riparian zone occupies the ecotone between aquatic and terrestrial vegetation from
49 the high-water mark of a watercourse to the terrestrial dominated floodplain zone (Good
50 et al. 2017; Riis et al. 2020). The unique composition of species in riparian ecosystems
51 contributes disproportionately to landscape biodiversity (Sabo et al. 2005; Bennett et al.
52 2014; Hansen et al. 2019), provides significant human cultural resources (Humphries
53 2007) and can act as a refuge for fauna during climatic extremes (Nimmo et al. 2016).
54 Furthermore, riparian vegetation influences the structure, function and composition of the
55 in-stream environment, both locally and downstream (Gurnell et al. 2012; Paice et al.
56 2017). However, their ecotonal nature makes them particularly sensitive to changes in
57 hydrological conditions (Riis & Biggs 2003; Poff & Zimmerman 2010). Degradation and
58 loss of riparian vegetation is especially severe along lowland waterways (Feld et al., 2011;
59 Fraaije et al., 2019; Tonkin et al., 2020) due to the intensification and expansion of river
60 regulation for agriculture and this impacts the overall health of river systems and
61 landscape-level diversity. The effects of climate change on precipitation patterns,
62 temperatures, and water cycles will likely exacerbate these impacts (Palmer et al. 2009;
63 Rivaes et al. 2022).

64 Various frameworks and tools have been developed to support the management and
65 restoration of water-dependent ecosystems globally. These frameworks typically focus on
66 understanding and modelling hydrological changes to guide flow management (Bunn et
67 al. 2014; Swirepik et al. 2016; Arthington et al. 2023) or site-level management such as
68 livestock removal, weed control, or revegetation (Holmes et al. 2008; Omidvar et al. 2021;

69 Jones et al. 2022). However, there are limitations to these approaches given the known
70 interactions between flow and site interventions (Richardson et al. 2007). The apparent
71 disconnect between site-scale restoration efforts in riparian zones and the management
72 of flows in waterways may also reflect organisational challenges, such as different funding
73 sources, or different staff or organisational responsibilities for flow and site management.
74 However, identifying and highlighting opportunities and risks for flow and site interactions
75 is crucial, given the importance of riparian vegetation condition to overall waterway
76 health, and the importance of management on riparian vegetation condition (Merritt et al.
77 2010; Rivaes et al. 2015; Campbell et al. 2023).

78 State and transition models (STMs) can be used to describe and predict how ecosystems
79 change over time due to natural processes or human activities. STMs delineate and
80 describe common states (groups of structural and/or compositional expressions that a
81 given ecosystem can exist in) and the transitions between them, aiding in ecosystem
82 management and conservation planning. They are commonly used as a framework to
83 describe ecosystem state changes and they offer a distinct approach to ecosystem
84 management by emphasizing the importance of the current condition of a site or system
85 on its response to management interventions and/or resilience to disturbance events
86 (Westoby et al. 1989; Briske et al. 2008; Yates & Hobbs 1997; Standish et al. 2008;
87 Sinclair et al. 2019; Sato & Lindenmayer 2021). Cataloguing and describing common
88 condition states and drivers of transitions can allow for planning and prioritisation at
89 multiple spatial and temporal scales (Bestelmeyer et al. 2009; Sinclair et al 2019).
90 Further, STMs can integrate different types of knowledge, including experimental data,
91 field survey data, expert knowledge, and practitioner experiences into one framework
92 (Knapp et al. 2011; Bestelmeyer et al. 2017). They are effective for highlighting knowledge

93 gaps and uncertainties, turning them into testable hypotheses that can drive an increasing
94 understanding of the system and supporting adaptive management (Rumpff et al. 2011).

95 The use of system states enables clear articulation of changes to multiple univariate
96 measures simultaneously (such as tree density and native groundcover) to reach multiple
97 system endpoints (Rumpff et al. 2011; Jones et al. 2023). Additionally, they allow for more
98 meaningful articulation and assessment of goals that apply to target systems (as opposed
99 to individual variables), which are key elements of effective conservation planning
100 (Margoluis et al. 2009; Biggs et al. 2011).

101 Understanding and accounting for state-dependent trajectories of ecosystems in
102 response to managed environmental river flows is vital to the restoration and
103 maintenance of highly modified and regulated river basins in agricultural regions. Bond et
104 al. (2018) demonstrated the importance of understanding the current condition of riparian
105 and floodplain ecosystems when making predictions about the impacts of environmental
106 flows on future ecosystem condition. In floodplain areas with extensive remnant
107 vegetation, historical flow regimes are likely to be good predictors of the current condition
108 (Bond et al. 2018). However, for riparian vegetation in highly modified agricultural
109 landscapes, multiple additional stressors acting at the site level are just as likely to have
110 resulted in the current vegetation condition (Campbell et al. 2023). Therefore, the recovery
111 pathways for these systems will require a combination of site-level and river- or reach-
112 level interventions (Campbell et al. 2023), especially in river basins where regulated
113 waterways intersect highly modified agricultural landscapes.

114 Globally, significant resources have been invested into improving ecological health in
115 highly modified and economically significant river basins, by allocating environmental

116 water and reducing river regulation influences (Overton et al. 2014; Marshall & Alexandra
117 2016; Hart 2020), but inconsistent attempts at implementing effective and targeted
118 monitoring to demonstrate the impacts of these allocations are notable (Swirepik et al.
119 2016). Additionally, while broad objectives for vegetation management and monitoring
120 exist, quantified targets that align with feasible and desirable ecosystem condition states
121 are uncommon (but see Richards et al., 2020). Further, while there are excellent
122 resources and research into the management and effects of inundation and
123 environmental water in the more expansive areas of floodplain and wetland vegetation
124 (Bino et al., 2015; I. Overton et al., 2018), the specific combination of site scale
125 management and river flows that best support the recovery of the narrow and dynamic
126 riparian zone have not been well articulated to date. We propose a framework that
127 ensures transparency around: the decision-making process undertaken by water
128 managers; the expected outcomes from environmental flows; how flows might interact
129 with site-level management; and the best variables to measure progress towards or away
130 from those outcomes. Such a framework can support the implementation of structured,
131 consistent and explicit adaptive management.

132 In this paper, we describe a State and Transition Model that synthesizes current
133 knowledge about common riparian woodland condition states and the factors driving
134 inland riparian woodland condition in southeastern Australia. Our model is based on the
135 expert-elicited multi-community STM for eucalypt woodlands of southern Australia (Good
136 et al. 2024) which synthesized the knowledge and expertise of Australian woodland
137 ecologists to provide a framework for conservation planning and decision making. Using
138 the 'Floodplain and Riparian' sub-group from the General Woodland Model, we refined
139 this template model with a focus on lowland riparian woodlands, aiming to summarise the

140 states, transitions, thresholds, and management actions that are relevant to waterway
141 managers in southeastern Australia. Our model provides a framework to guide
142 management at multiple spatial and temporal scales via the classification of site-level
143 riparian condition and reasonable desired target states achievable within relevant
144 management timeframes. We illustrate how the model can be used to guide both water
145 allocation and vegetation management decisions and discuss how this approach is more
146 likely to achieve and detect a transition towards the target state.

147 **Methods**

148 *Study region*

149 We refined the multi-community State and Transition Model for eucalypt woodlands of
150 southern Australia (Good et al. 2024; hereafter ‘General Woodland Model’) to
151 demonstrate how it can be applied to decision making for lowland regulated rivers of
152 temperate southeastern Australia. This includes most lowland areas within the Australian
153 States of South Australia, Victoria and New South Wales, which includes waterways
154 throughout the Murray Darling Basin that are dominated by woodland vegetation
155 communities. While this vast area occupies many different ecosystem types, the lowland
156 riparian woodland communities are similar, mostly dominated by a canopy of *Eucalyptus*
157 *camaldulensis* (river red gum), variable shrub layer, and herbaceous understorey. This
158 similarity enables a ‘Riparian Woodland Model’ to be relevant to a very large spatial area,
159 although the specific model details will require understanding the local reference
160 vegetation types. The Riparian Woodland Model does not apply to upland areas that
161 generally have different flow conditions, regulation actions, disturbances, and vegetation

162 communities, but it will be an approximate analogue for lowland regulated waterways
163 with woodland vegetation communities globally.

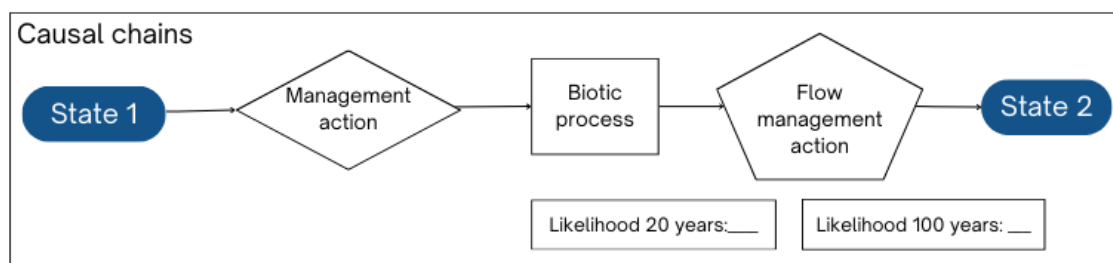
164 *Refining the General Woodland Model for lowland riparian management*

165 We adapted and refined the General Woodland Model to increase its applicability to
166 temperate lowland riparian vegetation in southeastern Australia. We first considered and
167 revised the existing set of eight general condition states and all existing plausible direct
168 transitions between each unique pair of states, then revised the drivers of these
169 transitions using simple causal chains (e.g. Figure 2). We aimed to increase the specificity
170 for riparian ecosystems, especially in relation to the interaction between changes to flow
171 management and changes to site management, while trying to maintain simplicity and
172 generality across waterways.

173 We extracted all data associated with the Floodplain and Riparian sub-group of the
174 General Woodland Model (available publicly [here](#)). Given we altered the number of
175 condition states, we revised the plausible transitions that were presented in the General
176 Woodland Model to better reflect lowland regulated waterways within southeastern
177 Australia. We applied the logic described in Good et al. (2024), when determining
178 plausible direct transitions, specifically ‘direct transitions are those that would plausibly
179 occur over 20 or 100-year timeframes without passing through any of the other states,
180 assuming that resources and effort are not limited.’ Plausibility of the riparian transitions
181 was estimated by the authors. We also removed drivers that were not directly relevant to
182 riparian condition.

183 For each of the plausible riparian woodland transitions, we selected the relevant site
184 management drivers, flow management drivers, and biotic processes required for the

185 transition to occur, as well as the likelihood that (if all these factors are present) the
 186 transition would happen within 20 and 100-year timescales. Likelihoods were qualitative
 187 categories (Almost no chance; Very unlikely; Unlikely; Neither likely nor unlikely; Likely;
 188 Very likely) and were estimated by the authors based on the assumption that the specified
 189 management was implemented. These pathways (including all drivers, processes, and
 190 measurable attributes that might indicate the transition has occurred and their associated
 191 likelihoods) were given a unique identifier and are herein referred to as ‘causal chains’
 192 (Niemeijer & De Groot 2008).



193
 194 **Figure 1.** An example of a causal chain indicating the drivers required to cause a state transition to
 195 occur. The chain may be simple or complex with one or many actions or processes that drive the
 196 transition.
 197

198 We didn't include drivers that are external to the system such as climatic events, intense
 199 bushfires, disease or outbreaks of insects. However, these factors do influence the
 200 condition of riparian vegetation and so we've included them as 'hazards' which will need
 201 to be considered as modifiers of the likelihood of transitions being successful.

202 *Incorporating flow management into the model*

203 We describe flows only in their relative change from the flow regime that supported the
 204 starting state. The reason we chose not to describe flows in absolute terms is because the
 205 impact on the riparian vegetation is closely associated with site-scale variables that
 206 cannot be accommodated in this model. For example, the riparian zone is a function of

207 bank steepness, channel depth, water movement (which can be influenced by in-stream
208 structures) and therefore absolute values for flows would be impossible to estimate at the
209 site or reach level across such a large region. Therefore, we describe the change in overall
210 flow regime (increase, decrease or status quo) that would be required for the transition to
211 occur. While these categories are a simplification of the detailed flow regime
212 requirements of riparian communities, there is extensive literature and management
213 plans for environmental watering that would guide decisions within system constraints for
214 increased or decreased flow regimes. For example, an increased flow regime is an
215 increase in flow volume above the current conditions, which can be achieved through the
216 increase in volume and duration of baseflows, and/or the increase in volume, duration or
217 frequency of high flows, but the specific actions will differ between waterways and years.
218 To investigate potential dependencies among drivers that occur at different spatial scales,
219 we compared the frequency of transitions that require combinations of site or flow
220 management interventions. Each transition was summarised by the presence or absence
221 of drivers relating to changed flows, changed site management and/or other biotic
222 processes. We removed those containing any mention of 'status quo' and cross-checked
223 to ensure all drivers included in this part of the analysis represented significant
224 interventions. To visualise the intersections among different types of drivers, we used
225 UpSetR (Conway et al 2017) which is a package that visualises intersecting sets
226 (observations) and their properties (variables). In this study, each transition represents a
227 set, and each driver group is either present or absent in the causal chain for that
228 transition. To visualise the intersections of driver groups, the number of transitions
229 involving each combination of driver groups were tallied and presented in an UpSetR plot.

230 *Applying the state and transition model to real-world riparian management*

231 We demonstrate how the Riparian Woodland Model could be incorporated into different
232 levels of management, monitoring and decision making in riparian and river systems.

233 Application of a STM requires some structured breakdown of transitions into their
234 constituent parts and the strategic application of those parts to the most relevant use-
235 case. We take the Riparian Woodland Model presented in this paper and provide some
236 practical examples of how to convert the information provided in the model to real-world
237 management pathways. In doing so, we show how the causal chains describing a given
238 transition (from a starting state to an end or target state) can be used for setting targets,
239 developing management plans, communicating with different stakeholders, measuring
240 progress and impact on target variables and progress reporting.

241 **Results**

242 *Riparian woodland states*

243 We reviewed the eight General Woodland Model vegetation states described in Good et al.
244 (2024) as a starting point and systematically refined them to better reflect common
245 categories of riparian woodland vegetation condition in temperate southeastern Australia.

246 The resultant set of states shares most of the original eight-state structure (Table 1), but
247 includes nested sub-states for the Highly Modified Woodlands to account for the strong
248 influence of functional-group dominance (riparian vs terrestrial) in the groundstorey (Table
249 1). This distinction between riparian and terrestrial plant dominance was essential
250 because of the close links between these groups and flow regimes, as well as being a
251 common indicator of ecological objectives for riparian woodlands (Tonkin et al. 2020). We
252 also added an 'Intermediate Restoration' state which is not a stable state but represents a

253 longer transition that will likely span years to decades, depending on the growth and
 254 development of a mature eucalypt woodland structure.

255 **Table 1.** Comparison of the condition states described in the General Woodland Model and the refined
 256 condition states used in the Riparian Woodland Model described in this paper.

General woodland states	Riparian woodland states	Description of riparian model state	Rationale for change
Exemplar	Exemplar	All vegetation strata are intact; native riparian species richness is high in all strata and includes disturbance sensitive species; low weed cover; soil is stable and has a natural nutrient balance.	No change
Modified woodland	Modified Woodland	Overstorey is mostly intact; mid/understorey are depleted in both richness and cover; understorey flora is primarily native and riparian; soil nutrient levels and stability close to reference.	No change
Highly modified woodland	Highly Modified Woodland A	Overstorey and midstorey reduced; understorey is depleted in richness and is dominated by terrestrial species; possible altered soil processes and degraded structure (bank erosion).	A switch from riparian to terrestrial understorey requires flow management that wasn't captured in original model
	Highly Modified Woodland B	Overstorey and midstorey reduced; understorey is dominated by exotic riparian species; possible altered soil processes and degraded structure (bank erosion).	A switch from native riparian species to exotic riparian species signals a significant functional and compositional change that requires different site and flow interventions
Diverse derived grassland	Diverse Derived Understorey	Overstorey and midstorey mostly absent; understorey mostly intact and dominated by native riparian species; soil nutrient levels and stability close to reference.	A name change, as riparian zones where trees have been removed aren't typical 'grasslands'
Depleted derived grassland	Depleted Derived Understorey	Overstorey and midstorey mostly absent; understorey is depleted and dominated by exotic riparian species; soil may have degraded structure (bank erosion).	
Thicket	Tree Thicket	Mature trees mostly absent; overly dense native overstorey; understorey cover suppressed, and may be dominated by native or exotic species; soil stability may be compromised	A name change to specify difference between shrub and tree thickets (which are more common in riparian zones)
Overstorey and midstorey thicket	Shrub Thicket	Mature trees mostly absent; overly dense native shrubs; understorey cover suppressed, and may be dominated by native or exotic species; soil stability may be compromised	
Transformed	Transformed	Mature native trees and shrubs mostly absent or dominated by exotic species (e.g. Willow or Blackberry); understorey dominated by exotic or native terrestrial species; bare or low groundcover; erosion may be severe; may have high cover of exotic trees and/or shrubs	No change
N/A	Intermediate Restoration	Tree and shrub species composition similar to Exemplar but not in size or basal area. Understorey and soil may range from slightly depleted to degraded, and may need further management.	An added transition state to capture the need for long time-scales to allow mature woodland development, and the possibility for requiring different management interventions based on the likely trajectory.

257 *Transitions among riparian states*

258 There were 39 plausible direct transitions among the ten riparian woodland condition
259 states (Appendix S1: Table S1). The management interventions and biotic processes
260 expected to drive each transition included 21 site-scale drivers, of which 16 were
261 common to both the General Woodland Model and the Riparian Woodland Model (Table
262 2). Riparian-specific biotic processes were added to account for the recruitment of
263 riparian weeds, including exotic trees and shrubs, the removal of exotic trees and shrubs,
264 control of pest animals and thinning of shrubs (Table 2). We removed drivers that aren't
265 common in riparian zones such as cropping and soil nutrification. An additional five flow-
266 management drivers were included to specify the direction of the change in flows
267 replacing the less specific mention of flood and flow regimes from the General Woodland
268 Model.

269 Site management drivers include common restoration activities such as revegetation,
270 livestock access controls and exotic plant removal, as well as activities that are
271 associated with degradation of riparian zones such as inappropriate livestock grazing and
272 removal of native vegetation. Some management interventions could be used for
273 restoration purposes or could be associated with site degradation depending on the
274 transition being described (for example thinning of midstorey might be required for
275 restoration of shrub thickets but could be associated with degrading transitions if applied
276 to Exemplar riparian woodlands). For this reason, we didn't group management
277 interventions as negative or positive in terms of their potential effect on riparian condition.

278 **Table 2.** Drivers used to describe transitions in the Riparian Woodland Model compared with those used in
 279 the Floodplain and Riparian causal chains in the General Woodland Model. Hazards were added but not
 280 included in the individual transitions because they are likely to have unpredictable but generally negative
 281 impacts on the condition of Riparian vegetation. Any transitions involving an increase in condition, therefore,
 282 would be less likely to occur if any of these hazards are present.

Drivers of transitions		
General and Riparian Models	Riparian Model only	General Model only
<i>Site management drivers</i>	<i>Site management drivers</i>	<i>Site management drivers</i>
Inappropriate grazing pressure	Thin shrubs	Cease land use
Thin mature trees	Exotic tree and shrub removal	Status quo
Degrade soil structure	Manage pest animals	Cropping
Remove all vegetation		Soil nutrient increase
Remove understorey		
Remove overstorey	<i>Flow drivers</i>	<i>Flow drivers</i>
Remove midstorey	Reduce flows	Flow or flood regime altered
Thin saplings	Increase flows	Reinstate appropriate flow or flood regime
	Status quo*	
Manage total grazing pressure	Status quo or reduce flows*	<i>Biotic processes</i>
Control erosion	Status quo or increase flows*	Self-thinning
Weed control		Weed recruitment
Plant understorey	<i>Biotic processes</i>	
Plant midstorey	Riparian weed recruitment	
Plant overstorey	Exotic tree and shrub recruitment	
<i>Biotic processes</i>		
Native recruitment dense woody	<i>Hazards</i>	
Mature tree death	High intensity fire	
	Ongoing drought	
	Extreme heat wave	
	Defoliating insect outbreak	
	Pathogen outbreak	

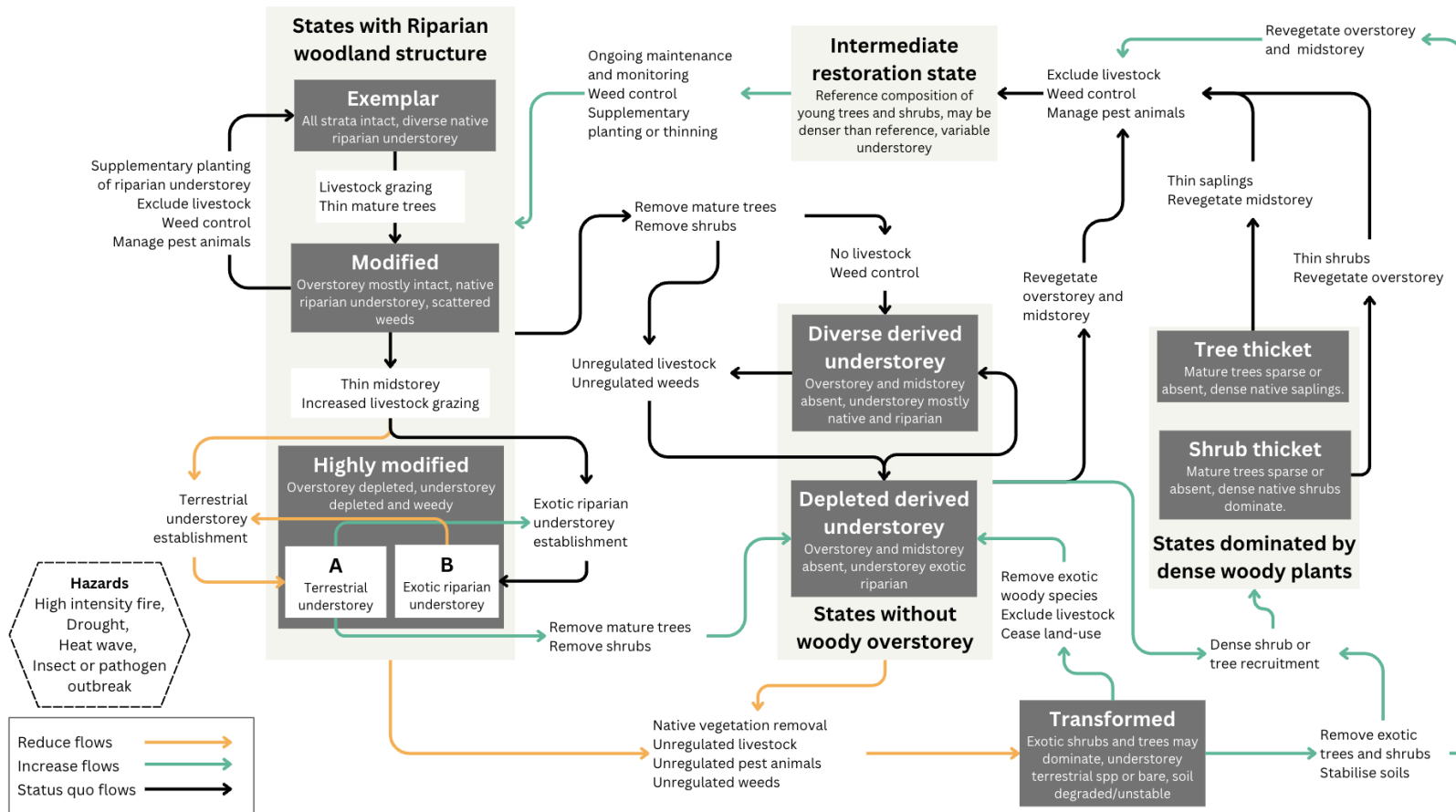
283 *denotes drivers that were not included in the driver interaction analysis

284 Flow management interventions were categorized by the direction of change from current
 285 flows (increase, decrease, status quo), rather than individual flow components, to avoid
 286 overcomplicating the model. Similarly, the direction of change in flows could be
 287 associated with restoration or degradation pathways depending on the starting and
 288 ending states.

289 Causal chains describing the 39 transitions among states were described by identifying
 290 which drivers (site or flow) and biotic processes are required to shift from one state to

291 another. To simplify the visualization of the model we grouped states with similar
292 structure: mature woodlands, no overstorey, and dense woody (Figure 3) and if there was
293 overlap in the drivers associated with transitions among these grouped states, we
294 represented these as one arrow in the box and arrow plot. For example, the two causal
295 chains describing the transitions from Diverse Derived Understorey and Depleted Derived
296 Understorey to the Intermediate Restoration state are summarised by the same arrow
297 pathway (Figure 3) and the same corresponding drivers (Appendix S1: Table S1).

298 In general, reductions in flow regimes resulted in shifts to more degraded states (those
299 with a terrestrial understorey), and subsequently, increased regimes were required for
300 state improvements (Figure 3). However, increased flows were associated with some
301 'negative' transitions when they were coupled with negative drivers (e.g. from the Highly
302 Modified Woodland A to Depleted Derived Understorey), which highlights the importance
303 of considering the influence of current flows, other drivers, current site condition and the
304 desired transition when making decisions about changes to flows. At the site scale,
305 inappropriate grazing pressure and vegetation removal were the most common drivers for
306 transitions towards the more degraded riparian states, while weed control, revegetation
307 and livestock removal were common drivers of transitions to less degraded states.



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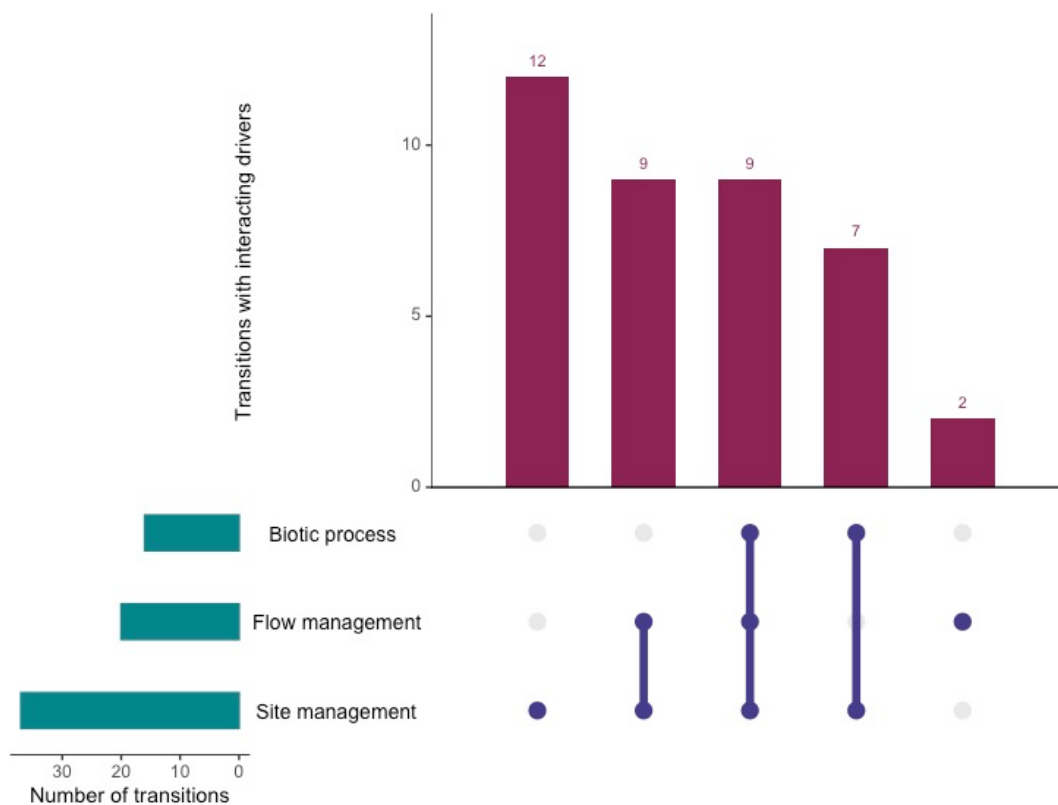
309 **Figure 2.** Diagram showing riparian states and transitions among states for the Riparian Woodland Model. Where arrows originate or end at one of the
 310 shaded boxes (containing several states), the transition represents the combination of the states within the shaded box (thus one arrow may represent
 311 several distinct transitions). For example, the transition from the Diverse Derived Understorey and the Depleted Derived Understorey states to the
 312 Intermediate Restoration state involve the same drivers (perhaps with slightly different levels of effort required to ‘control weeds’). Arrows are coloured
 313 where a decrease or increase in flows is required for the transition to occur. For example, from Highly Modified Woodland A (which has a terrestrial
 314 understorey), an increase in flows, along with removal of woody species is required to get to the Depleted Derived Understorey state. The full suite of
 315 transitions, drivers, indicators and likelihoods for each unique transition can be found in Appendix Table A1.

316 *Interactions between flow and site management*

317 We grouped the drivers that influence transitions into three broad groups: 1) changes to
318 site management; 2) changes to flows; and 3) biotic processes; Table 2). We found that
319 most transitions required some kind of management at the site level (37 of the 39
320 transitions) and 12 of the transitions required only site interventions (Figure 4). Changes to
321 flows without any site management (including the biotic processes) only accounted for
322 two transitions, although flow management had some role in over 20 of the 39 transitions
323 (most of which were positive). Both site management and changes to flows (+/- biotic
324 processes) were required for 18 of the 39 transitions.

325 The most frequent monitoring indicators that may be used to assess transitions between
326 condition states were midstorey density and mature tree density, however measuring the
327 cover of native and exotic riparian groundcover and terrestrial groundcover were also
328 associated with more than half of the transitions (Appendix S1: Table S1). For all
329 transitions to or from states with mature woodland structure, or from thickets to other
330 states, monitoring of mature tree density is an essential variable. Monitoring of
331 understorey was more important for transitions between states within the structural
332 groups e.g., the Derived Understorey states; Appendix S1: Table S1). The density of tree
333 saplings and shrubs was an important distinguishing variable for many transitions but
334 usually alongside the mature tree or understorey variables, except for thicket transitions.
335 The Intermediate Restoration state contains a wide range of understorey vegetation
336 conditions but a specific (and unstable) density of immature and mature trees and
337 shrubs. This means that transition thresholds to or from the Intermediate Restoration
338 state cannot be determined, but rather, vegetation conditions of this state will represent

339 an intermediate level somewhere between the starting state prior and one or more
 340 possible end states. While this defies the typical definition of states, it is a useful
 341 component of the model due to the prolonged period that vegetation communities can
 342 occupy this transition, and the common occurrence of this ‘state’ in degraded
 343 landscapes.



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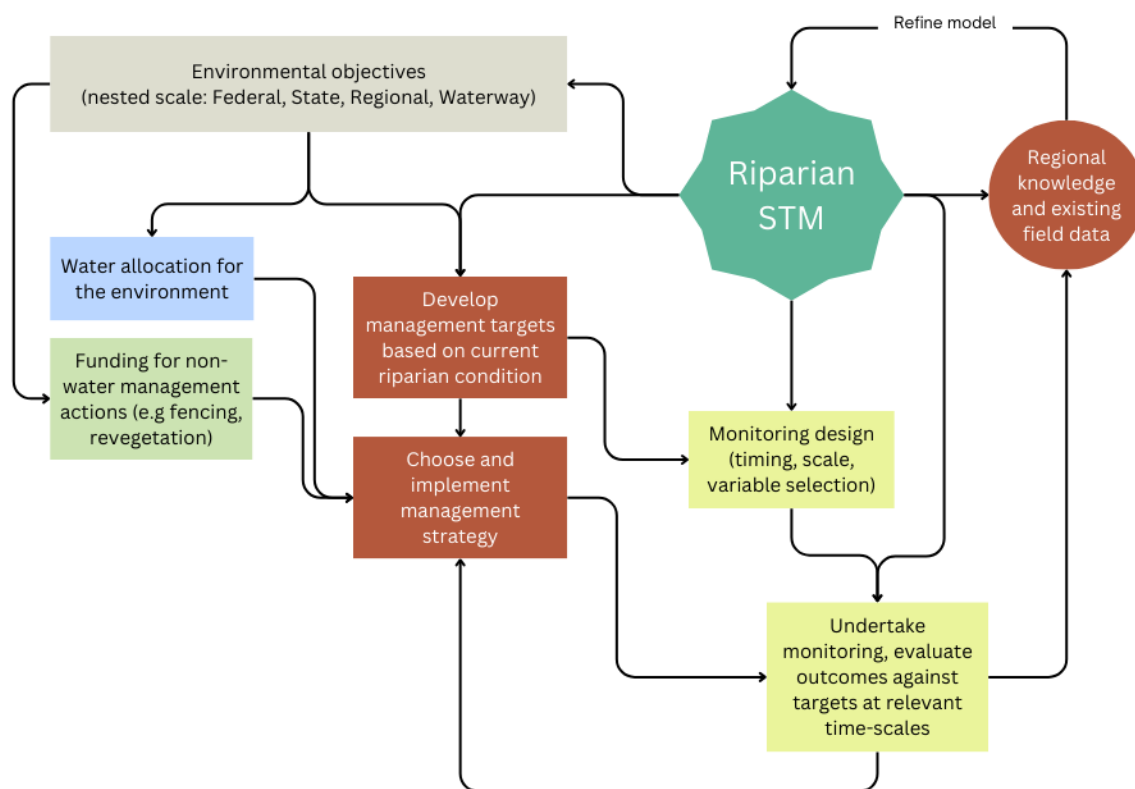
345 **Figure 3.** An UpSetR plot showing the number of transitions that required different combinations
 346 of site management, flow management and various biotic processes. Blue filled circles represent
 347 driver groups that are associated with transitions, vertical bars represent the number of transitions
 348 in which the driver group (or combination of driver groups if there are multiple filled circles
 349 connected by a vertical line) are involved. The horizontal bars represent the total number of
 350 transitions that involve the corresponding driver group (irrespective of intersecting driver groups).
 351 Drivers, indicators and likelihoods of each unique transition can be found in Appendix Table A1.

352 *Applying the Riparian Woodland Model to management scenarios*

353 The Riparian Woodland Model described above can be used by managers to inform and
354 structure decision making related to vegetation management in regulated rivers. This STM
355 feeds into all levels of the existing management hierarchy common to most regulated
356 waterways in southeastern Australia, and will be relevant to many systems internationally
357 (Figure 5). Inputs into each component of the hierarchy are iterative and are updated as
358 new information is obtained or constraints are altered, e.g. funding or infrastructure.
359 Objectives can be altered to align with the STM either explicitly or indirectly. For example,
360 objectives for a waterway or region can be made to maintain or improve the state of
361 existing vegetation communities. Following this, management targets can then specify a
362 desirable proportion of the waterway (or representative sites) within a particular state, or
363 alternatively, to increase the levels of vegetation attributes (e.g. tree density, plant cover
364 or richness) to be above thresholds that correspond to desirable states. The management
365 strategy is then designed to alter relevant vegetation attributes in specific locations to
366 meet these objectives and targets. Importantly, this process provides quantitative targets
367 and thresholds that enable targeted monitoring data and analysis to evaluate progress
368 towards targets over time – a process that is poorly achieved in many management
369 programs.

370 Here we demonstrate how to extract relevant information provided in the Riparian
371 Woodland Model and apply this to different levels of management decision making,
372 restoration implementation, monitoring, adaptive management and progress reporting
373 (Figure 4). A key step in applying the STM to management and monitoring is to develop
374 plans that incorporate components of the STM, including likelihoods of transitions with

375 specific time-frames, as well as the most likely indicators for transitions in order to
 376 objectively evaluate impacts of interventions. Once the monitoring plan is developed, a
 377 process for feeding new monitoring data into the management framework needs to be
 378 determined, via an adaptive management framework. This can follow simple or complex
 379 processes, such as to continue a management action until a threshold is crossed (i.e.
 380 transition to a new state), or to refine and adapt management actions from a suite of
 381 approaches to reach desired targets.



382
 383 **Figure 4.** An illustrated suggestion for incorporating the Riparian Woodland Model ‘Riparian STM’
 384 into multiple levels of waterway and vegetation management. For example, the Riparian Woodland
 385 Model can be used to help to identify objectives (i.e. preferred states), develop management
 386 targets and assist in monitoring design via selection of variables. Finally, the STM provides a
 387 means of synthesising and storing systems understanding that is easy to communicate and is
 388 amenable to updating over time.

389

390 The broad steps to integrate the Riparian Woodland Model into management at the State
391 or regional scale are:

- 392 1. Review objectives for the region and integrate the STM into these at the highest
393 level. For example, objectives ‘to maintain vegetation condition’ (a common but
394 imprecise objective) can be modified to speak to the STM and improve clarity, such
395 as ‘maintain condition of healthy states and improve degraded states’ – as defined
396 by the model application.
- 397 2. Assess current riparian vegetation condition (or tree cover from aerial imagery if
398 condition information is not available) and categorise deviation of flows relative to
399 reference for all waterways – then use this to estimate proportions of each
400 waterway in each state;
- 401 3. Develop targets (at the regional scale) for changes in the proportion of condition
402 states, e.g. all waterways to have >20% of the linear extent of riparian communities
403 in XX state and <5% in XX state in 10 years;
- 404 4. Develop high-level management plans for each transition relating to the targets
405 (including any requirements for changes in flows) and include high-level
406 monitoring plans using suggested indicators of transitions;
- 407 5. Evaluate success based on targets and monitoring indicators and incorporate
408 lessons back into the STM.

409 The broad steps to integrate the Riparian Woodland Model into management at the
410 waterway or reach scale are:

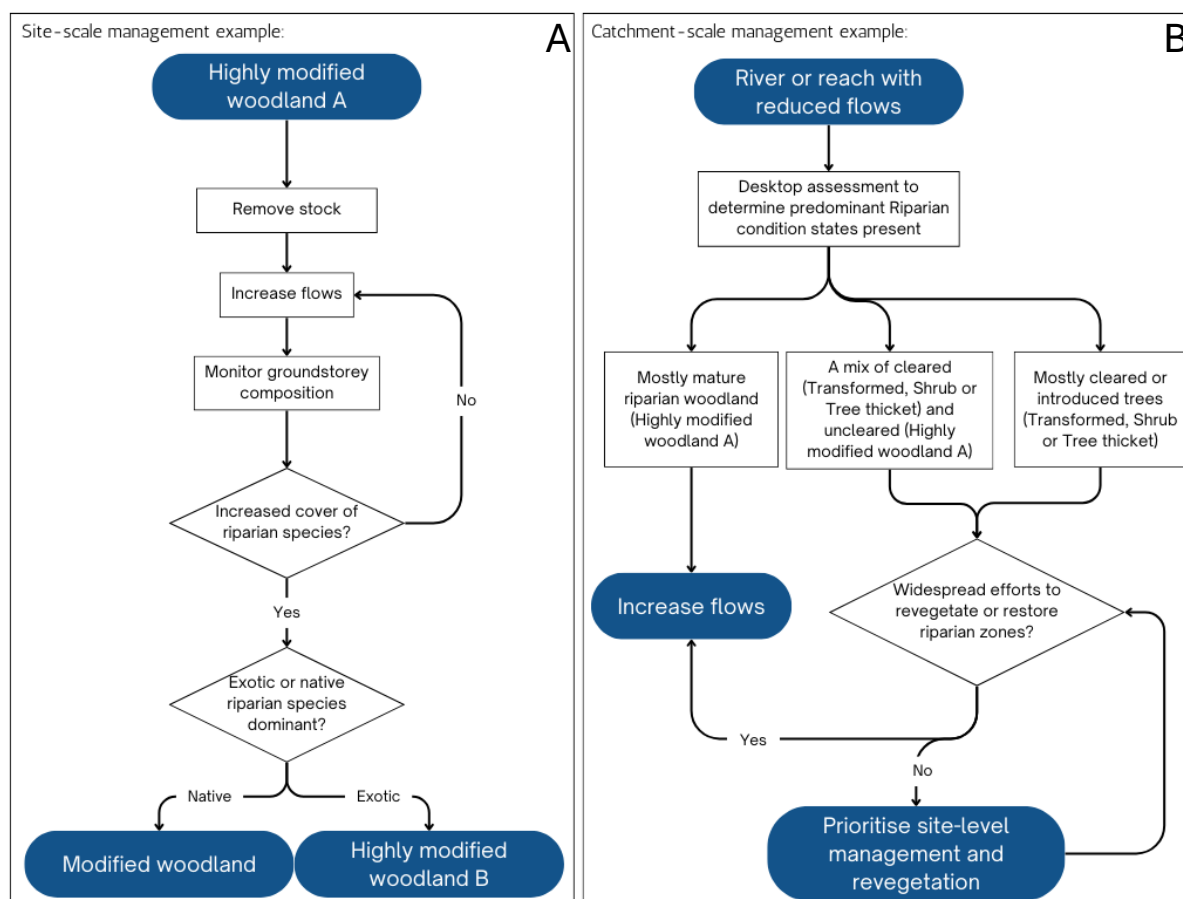
- 411 1. If regional objectives do not refer to STM, review waterway or reach scale
412 objectives and modify to incorporate locally appropriate objectives that refer to the
413 STM.
- 414 2. Assess current riparian vegetation condition (or tree cover if condition information
415 is not available) and categorise deviation of flows relative to reference for the
416 waterway – then use this to estimate proportions of the waterway in each state;
- 417 3. Develop targets (at the waterway scale) for changes in the proportion of condition
418 states, e.g. >40% of the linear extent of XX waterway riparian communities in either
419 XX or XX state in 10 years;
- 420 4. Develop detailed management plans for each transition relating to the targets
421 (including any requirements for changes in flows) and include detailed monitoring
422 plans using suggested indicators of transitions;
- 423 5. Evaluate success based on targets and monitoring indicators and incorporate
424 lessons back into the STM.

425 *Using STMs to structure decision making*

426 The Riparian Woodland Model can be used to support structured decision making to help
427 achieve the best outcomes for riparian vegetation within a complex decision context. In
428 this section we describe examples of how decisions at the site and flow management
429 scale could be informed by the Riparian Woodland Model. In Figure 5A we show how a
430 target transition from Highly Modified Woodland A to Modified Woodland would involve
431 reduced grazing pressure at the site level (if currently grazed) and increased flows to aid
432 the recruitment and establishment of native riparian species. However, if exotic riparian
433 species arrive and establish in place of native species, this represents a transition to

434 Highly Modified Woodland B and would need to be managed accordingly. The monitoring
435 of understorey composition and cover to ensure that 1) flows are adequate for riparian
436 establishment and survival and 2) native species are dominant, is therefore key to
437 measuring success for this transition. In this example, we used the condition state
438 descriptions, management drivers (grazing and flows) and indicator variables to develop a
439 decision tree based around the starting condition state (Highly Modified Woodland A).

440 Another approach is to create a decision framework around catchment scale targets
441 (Figure 5B). Here we started with an example of a river or reach with reduced flows. In this
442 situation, areas with a mature overstorey are most likely to have a terrestrial dominated
443 understorey (due to the lack of flow dynamics) and are therefore would most likely be in
444 the Highly Modified Woodland A state. Areas without a mature overstorey structure are
445 most likely to be either Transformed or Thicket states which require some site-level
446 management. All states require increased flows, however, increasing the flows without
447 first implementing site-level management is unlikely to result in improved condition,
448 unless there is already a mature overstorey layer. Therefore, the order in which site and
449 flow management are applied is dependent on the proportion of the river or reach in
450 different condition states.



451

452 **Figure 5.** Two decision trees using information from the causal chains developed in the Riparian
 453 Woollands Model to guide management at the site-scale (A) and an alternative approach at the
 454 river-, reach- or catchment-scale (B).

455

456 **Discussion**

457 In this study we refined the temperate Australian General Woodland Model (Good et al.
 458 2024) for a specific ecosystem – riparian woodlands in southeastern Australia. The
 459 process of refining and articulating changes to the General Woodland Model captured the
 460 importance of plant functional groups in riparian systems and the complexity of multi-
 461 scale management that may be required to achieve particular transitions. While site
 462 interventions were required for most transitions among riparian woodland condition
 463 states, changes to flow alone were only associated with two transitions, highlighting that a
 464 focus on flow management in isolation of other management actions will most likely not

465 result in achieving ecological objectives particularly in highly modified landscapes.
466 Indeed, our refined model demonstrates that riparian ecosystems cannot be restored at
467 the catchment scale without some site-scale interventions in addition to increased flows.
468 We suggest that the outcomes and impacts of environmental flows could be greatly
469 improved with a coordinated decision-making approach to allocating resources to site-
470 scale interventions and environmental flows.

471 In general, the set of condition states described in the General Woodland Model (Good et
472 al. 2024) were largely applicable to riparian woodlands, and the changes we made were
473 based on the different management required when groundstorey composition shifts from
474 riparian to terrestrial species, and from native riparian to exotic riparian species. While
475 this reinforces the usefulness of the General Woodland Model as a base STM for
476 temperate Australian woodlands, it highlights the importance of refining the general
477 model to suit the spatial and temporal scope of a particular use. We expect that the
478 General and Riparian Woodland Models would be well suited as baseline models of
479 temperate woodlands, forests, and grasslands globally.

480 An important addition to our STM is the inclusion of an 'Intermediate Restoration' state
481 because it provides a way to track the slow transition from treeless to mature woodland
482 states. Given the amount of time needed for the development of a mature woodland
483 structure from previously cleared states, it is important to identify the variables that might
484 indicate a site is on a recovery pathway within management time-scales. This state can
485 act as a placeholder to encourage monitoring in older revegetation sites that appear to be
486 developing characteristics of the Exemplar state. These attributes could act as indicators
487 of progress towards a longer-term goal (for example when moving from treeless states to

488 states with a woodland structure it's unlikely that you will be able to measure 'success'
489 until many decades after the management is implemented). Indeed, the inclusion of this
490 state reflects the dearth of published studies demonstrating successful restoration from
491 degraded or cleared states to the Exemplar state (in terms of structure, composition and
492 function; Atkinson et al. 2022). Studies of riparian restoration trajectories demonstrate the
493 importance of choosing appropriate indicators and planning for long-term monitoring
494 (Tonkin et al. 2020). Some of these gaps in our understanding of riparian restoration
495 trajectories could be filled by revisiting older restoration sites and comparing their
496 structure and composition to reference conditions.

497 Another key difference in the Riparian Woodland Model compared to the General
498 Woodland Model is the importance of the understorey functional groups (riparian versus
499 terrestrial), rather than exotic versus native. This approach aligns more closely with the
500 framework by Richardson et al. (2007) that describes management actions under three
501 scenarios dictated by the prevalence of invasive plants. While these three scenarios are
502 useful, they fall short of describing the important states that occur and are desired within
503 a riparian woodland context.

504 Transitions to other states require changes to the strength of drivers, their presence, or
505 their reversal; this reflects a resilience-based idea that is important to capture in the
506 development of STMs (Briske et al 2008). Management interventions associated with
507 transitions from treeless states to treed states generally involve planting trees and shrubs,
508 managing or removing stock and controlling weeds. In most cases, riparian revegetation
509 or natural recruitment requires additional water resources in the first year to enable
510 establishment, either via elevated flows (e.g. Deng et al. 2024) or through on-ground

511 watering of tubestock, which highlights a potential benefit of incorporating site and flow
512 decisions to improve outcomes for plantings and increase efficiency in restoration effort
513 and investment. While some transitions were possible with either site or flow
514 management only, many desirable transitions from degraded states required both flow
515 and site management to occur. Managers of waterways with stable and unchangeable
516 flow regimes may rightfully consider site management actions only, but in regulated
517 systems where flow manipulation is possible, such as the provision of environmental
518 flows, managers will benefit from considering the interactions between site and flow
519 management.

520 For the Riparian Woodland Model, the monitoring variables expected to be the most
521 beneficial were selected based on differences in vegetation attributes between states.
522 While this process is a very quick and efficient way of identifying vegetation variables that
523 will indicate a transition between states, these variables may not be most effective if the
524 threshold is uncertain due to high variability in variables. A more robust alternative is to
525 use data collected from each state of the target system and quantify threshold values to
526 identify variables that provide discrete or more confident thresholds (Jones et al. 2023).
527 The three vegetation variables that are expected to discriminate most state transitions in
528 our model align well with the findings of Jones et al. (2023) who identified the cover of
529 exotic understorey plants and the density of immature trees as the two most frequently
530 important variables to distinguish transitions in non-riparian woodlands. A similar formal
531 analysis of the relative importance of variables and determination of the quantitative
532 transition thresholds between states could be conducted for states in our Riparian
533 Woodland Model when sufficient data are collected from a representative sample of
534 vegetation communities within each state.

535 Monitoring should be undertaken to identify and manage unintended transitions (towards
536 the two Highly Modified Woodland states, instead of the Modified Woodland or Exemplar
537 states). This framework demonstrates the role of frequent and targeted monitoring in
538 situations where flows have been altered and the understorey composition is in flux to
539 ensure that highly competitive riparian exotic species do not establish and become
540 dominant (Tonkin et al. 2020). If appropriate flow management has not been undertaken,
541 the primary risk will be a transition towards woodlands with a terrestrial understorey,
542 therefore the relative abundance of terrestrial vegetation is a good indicator which could
543 be used to assess the appropriateness of flows to support riparian vegetation.

544 This STM framework provides a very high-level recommendation for flow management, i.e.
545 transitions via increased or decreased flow regimes. This is a necessary simplification of
546 the detailed flow management decisions that are required (and are used) to achieve
547 multiple objectives. This context specificity means that it is unrealistic and inappropriate
548 to incorporate detailed flow recommendations into the broad Riparian Woodland Model
549 but we believe that this first step is an important contribution to the management
550 framework. The high-level guidance provided via the STM will indicate what states
551 comprise the vegetation communities of a waterway, what states are desired, what
552 attributes need to be changed using the suite of flow and non-flow factors, and what
553 attributes to monitor to detect transitions. If a change in flow regime is required, managers
554 will need to determine the most effective way to deliver the flow regime within the
555 constraints of their system, such as water availability, physical infrastructure, legal or
556 policy constraints, consumptive water uses, competing objectives, and social
557 constraints.

558 The influence of different flow components on vegetation is expected to vary, for example,
559 baseflows will help to sustain aquatic vegetation communities (Tonkin et al. 2020) and
560 may provide groundwater resources to deeper rooted plants (Deng et al. 2024). Elevated
561 flow periods in spring and summer will provide water resources and propagules higher up
562 the bank and stimulate germination and enhance recruitment of riparian species (Tonkin
563 et al. 2020, Pereira et al. 2021, Deng et al. 2024). Additionally, each of these flow
564 components that increase soil moisture resources, disperse propagules, and subject
565 plants to periods of inundation will shift the competitive dominance of plants from
566 terrestrial groups to riparian groups (Miller et al. 2013, Main et al. 2022). Regardless of the
567 specific change in the regime, water availability within the soil will increase, propagules
568 will be dispersed to different areas, and competitive interactions between terrestrial and
569 riparian plants will shift in ways that promote condition of riparian vegetation
570 communities, provided that survival thresholds are not exceeded (Vivian et al. 2020,
571 Gower et al. 2024).

572 We created example decision trees to demonstrate how information can be extracted
573 from the Riparian Woodland Model to operationalize and standardize decisions in a
574 simple and transparent way. These decision trees can be modified to meet the specific
575 conditions of a system for the most common or important state transitions. Even if the
576 process is not fully embedded within a larger management framework of the system,
577 these decision trees can be used to guide management at the local level to improve
578 efficiency and outcomes.

579 Future work building on this study is hoped to include the development of an interactive
580 and updateable online-interface-tool for managers, researchers and governments (and

581 other river management stakeholders). Such a tool would provide spatially explicit layers
582 to aid in the assessment of current riparian condition states, options for site-scale
583 information to be added (for example where there is on-ground data available),
584 management options based on current and target condition states and a reporting tool for
585 uploading on-ground interventions (such as tree planting). This would allow for multi-
586 scale management decisions to be made, for example where there's a decision about
587 where to allocate flows, having access to the locations and timing of any on-ground
588 projects could help to achieve better outcomes for both environmental flows and on-
589 ground management. Importantly, for the Australian context, this process would include
590 work with Traditional Owners within a region to ensure that Cultural values and
591 Indigenous-led management are incorporated into management decisions (Goolmeer et
592 al 2021).

593 **Conclusions**

594 We adapted a General Woodland Model to riparian woodlands. Using this refined Riparian
595 Woodland Model we were able to assess which drivers were needed to facilitate
596 transitions between states. This analysis revealed that alterations in flows alone will, in
597 most cases, not cause transitions (to better or worse states) without biotic or site level
598 drivers also occurring. Likewise, beneficial transitions are not likely to occur from many
599 degraded states without increasing flows. This study highlights the utility of the General
600 Woodland Model which was straightforward to adapt to a specific vegetation type.
601 Furthermore, we show how the STM can be embedded in a broader framework to guide
602 decision making and management planning at a site-scale and catchment scale.

603 This framework is scalable and can be linked to broad national objectives while also
604 having the potential to guide on-ground works and management planning. Each condition
605 state is a relative departure from the reference for a given location or vegetation type, thus
606 there is room to specify fine-scale detail such as species or community requirements
607 within the framework, while maintaining a clear link to the broader context.

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614 **Conflict of Interest statement**

615 The authors have no conflicts of interest to declare

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