

Abstract

11
12 Ecosystems are linked by spatial flows of energy, nutrients, and organisms across ecosystem
13 boundaries, forming a meta-ecosystem. The study of spatial flows has typically focused on fluxes
14 of materials (e.g., nutrients or organic matter) and conspicuous organisms (e.g., fish or insects).
15 However, recent evidence from field studies has suggested that numerically significant spatial
16 flows of microorganisms occur across ecosystem boundaries, particularly across the terrestrial-
17 aquatic interface. These spatial microbial flows may be particularly relevant for meta-ecosystem
18 dynamics because the potentially high dispersal rates of microorganisms could exert spatial
19 controls on biodiversity and redistribute microbially mediated ecosystem functions. While
20 understudied empirically, the magnitudes of spatial microbial flows could be quite large in some
21 meta-ecosystems. Estimates of dispersal from terrestrial to aquatic systems have been on the
22 order of 10^{20} bacterial cells per year in some lakes, but microbial spatial flows likely vary widely
23 in strength and direction through time and space, raising questions about their ecological
24 significance. Much of this uncertainty stems from our lack of understanding of the fate and
25 activity of individual microbial dispersers, which are difficult to track and measure *in situ*.
26 Microbial spatial flows could sustain biodiversity and ecosystem functions in recipient
27 ecosystems if dispersers typically have high survival and metabolic activity (potentially
28 sustained by co-transport of available material flows); or they may serve as a spatial flow of
29 materials if survival is lower, primarily having indirect effects on diversity and functioning. The
30 fate of microbial dispersers depends on properties of donor and recipient ecosystems, the biotic
31 and abiotic flows that connect them, and the broader spatial and temporal variation across the
32 landscape. Meta-ecosystem ecology can help further integrate microbial movement into our
33 understanding of ecosystem dynamics at the landscape scale.

Introduction

Ecosystems are open to spatial flows of materials and organisms across their boundaries (Likens and Bormann 1974, Polis et al. 1997, 2004, Gounand et al. 2018b, Harms et al. 2021, Allen et al. 2024). Extensive exploration of the causes and consequences of these flows has coalesced into a set of related perspectives that includes the study of spatial subsidies, terrestrial-aquatic linkages, allochthony, eutrophication, and dispersal. For generality, I will refer to this set of topics by the term “spatial flows” to describe any translocation of nutrients, energy, or organisms from one ecosystem to another (while acknowledging that that these fields address different questions from different perspectives). Spatial flows can occur between any combination of ecosystem types and can lead to a number ecological responses across scales (Gounand et al. 2018a). Spatial flows can form linkages between nearby aquatic (e.g., riverine-to-marine) or terrestrial (e.g., forest and grassland exchanges) ecosystems (Fig. 1), but perhaps the most studied spatial flows are terrestrial-aquatic linkages (Fig. 2). Spatial flows may not necessarily lead to a subsidy if they do not enhance production (Kelly et al. 2014), possibly due to limits on the consumption of resource flows in recipient ecosystems that prevent subsidized production at higher trophic levels (Allen et al. 2024). Spatial flows can exhibit a wide range of variation across time, space, and ecosystem types, but the spatial flows discussed here are broadly applicable across collections of any ecosystem type linked by the movement of materials or organisms.

The collection of ecosystems linked by spatial flows of materials and organisms is known as a meta-ecosystem (Loreau et al. 2003, Gounand et al. 2018a, Guichard and Marleau 2021). Meta-ecosystem ecology has helped synthesize decades of empirical studies on spatial flows within a theoretical framework built on food web ecology, metacommunity theory, ecosystem ecology, and spatial ecology (Massol et al. 2011). In this paper, I argue that meta-ecosystem

57 ecology should build on recent field studies finding that microorganisms often disperse across
58 ecosystem boundaries in nature (Crump et al. 2012, Ruiz-González et al. 2015, Wisnoski et al.
59 2020, Stadler and del Giorgio 2022). Due to fact that microorganisms have high dispersal rates,
60 exhibit wide metabolic flexibility, and exert disproportionate control over biogeochemical cycles,
61 spatial microbial flows are highly relevant for meta-ecosystems in theory and practice. There
62 have been calls in meta-ecosystem ecology for a more careful consideration of the organisms that
63 consume detritus (Gounand et al. 2014), such as heterotrophic microorganisms. Incorporating
64 these additional spatial flows can generate new patterns and dynamics arising from feedbacks
65 between microbial metacommunity dynamics and fluxes of detritus and nutrients within and
66 between ecosystems.

67 **Microorganisms as mediators of spatial flows**

68 Studies of spatial flows have certainly included a focus on microorganisms, but this focus has
69 primarily been on their ability to stimulate primary and secondary production in the food web.
70 That is, microorganisms in these systems have typically been viewed as transformers of material
71 flows, either in donor systems prior to export or in recipient systems after import (Figs. 1-2,
72 dashed “t” arrows). In lakes, for example, carbon derived from the surrounding terrestrial
73 ecosystem often supports consumer biomass across multiple trophic levels (Carpenter et al.
74 2005), and can alter water chemistry and ecosystem functioning in ways that may reduce primary
75 production through browning (Solomon et al. 2015). An important pathway by which terrestrial-
76 derived carbon enters aquatic food webs is through bacterial biomass, whereby grazers consume
77 bacteria growing on labile carbon sources and rapidly transfer energy up to higher trophic levels
78 (Berggren et al. 2010). By stimulating bacterial production, spatial flows of carbon can also

79 influence ecosystem functioning through excess respiration (Lennon 2004). Autotrophic and
80 heterotrophic microorganisms can also transform spatial flows of other nutrients like nitrogen or
81 phosphorus, which can support recipient food webs even at great distances between donor and
82 recipient ecosystems (Rabalais et al. 2002). In general, studies from a variety of systems have
83 demonstrated that microbial pathways are critical for processing spatial material flows and
84 making them available to other organisms in the recipient ecosystem (Hall and Meyer 1998,
85 Moore et al. 2004, Jansson et al. 2007, Berggren et al. 2010, Benstead et al. 2021).

86 The impact of spatial organismal flows on recipient ecosystems may be mediated partly
87 by microorganisms. Dispersal between ecosystems is commonly observed (Leibold and Chase
88 2018, Guzman et al. 2019, Schlägel et al. 2020). For example, a recurrent, seasonal phenomenon
89 is the emergence of aquatic insects and their subsequent dispersal into nearby terrestrial
90 ecosystems, where they die due to abiotic constraints or are consumed by terrestrial predators
91 (Nakano and Murakami 2001, Gratton and Zanden 2009). These aquatic-derived prey then
92 contribute to terrestrial stocks of living consumer biomass, detritus (as uneaten insects or after
93 assimilation into terrestrial predators), and inorganic nutrients after decomposition by resident
94 microorganisms (Dreyer et al. 2012, 2015). Spatial flows of organisms may be recycled through
95 the food web (if they are consumed), eventually becoming detritus that is processed by
96 microorganisms (Hairston and Hairston 1993, Wetzel 1995). After microbial processing, these
97 resources can then be incorporated into local biomass or contribute to spatial flows of resources
98 across the meta-ecosystem (e.g., Fig. 2, dashed “t” arrows). In this scenario, microorganisms are
99 again playing a processing role, transforming materials before or after translocation across the
100 boundary.

101 Spatial material flows between ecosystems can be direct or indirect (Gravel et al. 2010),
102 with microbial metabolic pathways playing central roles in connecting different pools and fluxes
103 (Kayler et al. 2019). Direct flows reflect the translocation of inorganic materials between
104 ecosystems. In a direct flow, materials move from donor to recipient ecosystem without being
105 transformed. These translocated materials can then be processed in the recipient habitat by uptake
106 by primary producers. In contrast, indirect flows reflect the translocation of organic materials
107 between ecosystems. In an indirect flow, inorganic materials move between ecosystems bound up
108 in organic materials like organisms or detritus. After translocation, these organic materials must
109 be consumed or decomposed back to inorganic forms in the recipient ecosystem (Wetzel 1995).
110 Both direct and indirect flows are inefficient, resulting in the loss of materials due to sorption
111 (Groeneveld et al. 2020), uptake (Mineau et al. 2016, Peipoch et al. 2016, Tank et al. 2018), or
112 organismal death. Indirect flows may incur additional losses, such as respiration and excretion,
113 due to metabolic processes that consume energy during decomposition or secondary production.
114 Materials may be lost from the entire meta-ecosystem (e.g., respired carbon that is not fixed
115 again within any ecosystem), or they could be reintegrated back into the actively cycling channel
116 by physical or microbial processes.

117 In these scenarios, microorganisms feature prominently in their interactions with the
118 detrital component of the food webs, serving as an energy shunt up the food web bypassing
119 primary producers. It is often implicitly assumed that the microbial communities that will
120 process incoming spatial material flows are already established, but the composition and
121 functional capacity of microbial communities in recipient ecosystems can vary widely over time
122 and space (Martiny et al. 2006, Hanson et al. 2012). Understanding ecosystem functions then
123 may benefit from studying microbial community assembly processes, which include

124 environmental and biotic filters as well as stochastic processes and dispersal from the broader
125 metacommunity (Fukami 2010, Nemergut et al. 2013, Leibold and Chase 2018, Langenheder and
126 Lindström 2019). While the link between microbial community structure and ecosystem
127 functioning is often complex and difficult to predict (Allison and Martiny 2008, Krause et al.
128 2014, Louca et al. 2018, Sanchez et al. 2023), dispersal can be an important process in
129 facilitating or disrupting ecosystem functioning in microbial systems (Venail et al. 2008,
130 Lindström and Östman 2011, Székely and Langenheder 2017, Graham and Stegen 2017, Evans
131 et al. 2020, Walters et al. 2022).

132 **Spatial microbial flows: dispersal between ecosystems**

133 Dispersal is a central process guiding the assembly of microbial communities, and in many
134 habitat types, dispersing microorganisms may frequently arrive from regional pools beyond the
135 focal ecosystem (Lindström et al. 2006, Dodds et al. 2020). This cross-boundary movement of
136 microorganisms can be considered as a spatial microbial flow (Figs. 1-2, dashed “d” arrows).
137 Spatial microbial flows are a subset of other organismal flows but are worth distinguishing
138 because of their potentially large magnitudes and ability to translocate taxonomic and functional
139 diversity that can interact closely with material flows. Of course, not all terrestrial-derived
140 bacteria can persist in aquatic ecosystems, and microbial taxa with broader niche breadth or more
141 specialized persistence traits may be better able to survive across ecosystem boundaries than
142 others (Barcina et al. 1997, Wisnoski et al. 2020). Several recent studies have found an
143 abundance of terrestrial-derived bacteria in aquatic ecosystems, suggesting they may contribute
144 substantially to aquatic biodiversity (Crump et al. 2012, Ruiz-González et al. 2015, Hermans et

145 al. 2020, Wisnoski et al. 2020, Caillon and Schelker 2020, Caillon et al. 2021, Stadler and del
146 Giorgio 2022).

147 Some of the most compelling evidence for spatial microbial flows has come from
148 terrestrial-aquatic linkages in freshwater ecosystems. A series of studies from an Arctic landscape
149 showed that dispersal from terrestrial ecosystems can gradually shape aquatic microbial
150 community structure (Crump et al. 2007). Aquatic microbial diversity resembled the composition
151 of nearby soil communities from the surrounding watershed, and this compositional similarity
152 was highest in aquatic samples near the terrestrial-aquatic interface, suggesting a key role for
153 dispersal (Crump et al. 2012). The localization of terrestrial influence near the terrestrial-aquatic
154 interface may have resulted from the combination of high immigration rates of terrestrial-derived
155 bacteria and the concentrated influx of terrestrial-derived organic matter and nutrients near the
156 interface. Transplant experiments showed that aquatic communities were less productive when
157 grown in inlet water hypothesized to contain more terrestrial-derived materials (Adams et al.
158 2014). These hypotheses were further supported by more recent work from a temperate region,
159 suggesting spatial material flows interact with microbial dispersal (Bambakidis et al. 2024).

160 The evidence for spatial microbial flows can also be found at broader spatial extents.
161 Across a boreal landscape of hundreds of lakes, a similar pattern was detected: a gradual
162 transition from dominance by bacteria of terrestrial origin to bacteria of aquatic origin with
163 increased distance from the terrestrial-aquatic interface (Ruiz-González et al. 2015). Terrestrial-
164 derived bacteria were hypothesized to face increasingly strong environmental and biotic filtering
165 with increased residence time in the aquatic environment. This process would gradually select for
166 a bacterial community better suited to aquatic conditions along the hydrologic flowpath (Read et
167 al. 2015, Savio et al. 2015, Ruiz-González et al. 2015, Niño-García et al. 2016). These and other

168 studies have repeatedly found bacteria of suspected terrestrial origin in aquatic ecosystems at
169 local and regional scales, but the processes that regulate their arrival and persistence are still
170 unclear more generally.

171 A possible mediating factor in colonization success is the metabolic activity of
172 immigrating bacteria. The ability to enter dormancy, a reversible state of reduced metabolic
173 activity that can buffer against unfavorable environmental conditions, is widespread across the
174 microbial tree of life and is an especially common strategy employed by bacteria in terrestrial
175 soils (Lennon and Jones 2011, Lennon et al. 2021). If a large proportion of terrestrial soil
176 bacteria are inactive, their persistence in aquatic ecosystems may be aided by this inactivity. This
177 hypothesis was explored in a survey of bacterial diversity along a terrestrial-to-aquatic gradient
178 in a small, north-temperate reservoir (Wisnoski et al. 2020). Consistent with previous work,
179 many terrestrial-derived taxa were detected in the aquatic ecosystem. Aquatic richness and
180 compositional similarity to soils decreased with increased distance from the terrestrial-aquatic
181 interface, suggesting a role for dispersal. However, many terrestrial-derived taxa appeared to be
182 metabolically inactive, which may have enhanced their ability to persist in a less favorable
183 aquatic habitat (Wisnoski et al. 2020). Contemporaneous work from the same ecosystem has
184 shown that dormancy helps promote persistence of some taxa in the face of seasonal
185 environmental variation, particularly by preserving bacterial diversity in winter (Wisnoski and
186 Lennon 2021). In combination, these two studies suggest that terrestrial-derived bacteria disperse
187 into aquatic systems, survive initially through dormancy, reactivate later, with persistent taxa
188 affecting the recipient aquatic community and ecosystem functioning. Thus, dormancy may
189 alleviate some environmental constraints on successful establishment between donor and
190 recipient ecosystems, expanding the influence of spatial microbial flows.

191 A broader spatial survey across boreal lakes at higher latitudes found slightly different
192 results. In the boreal system, terrestrial-derived bacteria were not only present in aquatic
193 environments, but also appeared to be metabolically active across much of the region (Stadler
194 and del Giorgio 2022). In this study, terrestrial bacteria made up a dominant proportion of the
195 active bacterial communities in all aquatic ecosystems except for the estuaries. There was also a
196 temporal dimension to terrestrial-aquatic connectivity. High hydrological connectivity during
197 spring increased the relative importance of spatial microbial flows and helped maintain
198 terrestrial-derived bacteria in aquatic ecosystems; this connectivity was diminished during
199 summer months, when reduced spatial microbial flows allowed local environmental conditions to
200 shape aquatic bacterial communities still containing a large proportion of active bacteria of
201 terrestrial origin (Stadler and del Giorgio 2022). Thus, in this example, dispersal was important
202 for providing an initial source, but continued dispersal (and possibly dormancy) may have been
203 less necessary for allowing terrestrial-derived bacteria to persist in the aquatic ecosystem.

204 I focused on these spatial microbial flows between terrestrial and freshwater ecosystems
205 because they have been explicitly studied in the context of meta-ecosystems, but it is conceivable
206 that similar dynamics play out in other combinations of habitat types that have received less
207 attention (Dodds et al. 2020). When all these studies are considered together, it becomes clear
208 that spatial microbial flows occur in nature and can influence recipient ecosystems, at least in
209 terms of their apparent taxonomic diversity, but the effects of such flows may differ greatly
210 depending on their magnitude, timing, and activity levels. If microorganisms were previously
211 viewed as processors of materials exchanged across the interface, it is now becoming clear that
212 they are also exchanged between ecosystems. The degree to which allochthonous
213 microorganisms process allochthonous versus autochthonous materials remains to be seen,

214 however. A meta-ecosystem perspective could integrate microbial flows and their effects on
215 recipient and source ecosystems at different spatial scales.

216 **A meta-ecosystem view of spatial microbial flows**

217 Now consider spatial microbial flows alongside existing views of material and
218 (macro)organismal flows (Figs. 1-2). An important theme here is the ability of moving
219 microorganisms to not only influence patterns of microbial diversity, but to also perform
220 ecosystem functions that transform local materials, which can have meta-ecosystem scale effects
221 depending on how materials are transformed before or after translocation between ecosystems. A
222 recent framework describes three main ways that spatial flows can impact recipient ecosystem
223 functioning: through functioning by dispersers in the recipient ecosystem, by causing a change in
224 recipient ecosystem diversity and functioning, or by modifying donor ecosystem functions that
225 have consequences for recipient ecosystem functioning (Scherer-Lorenzen et al. 2022). Spatial
226 microbial flows can influence meta-ecosystems through all three pathways.

227 Microorganisms may play functionally different roles than other consumers typically
228 considered in meta-ecosystems. Microbial communities respond to the local environment,
229 regulate resource availability, and are consumed by higher trophic levels, but their controls on
230 resource availability via nutrient recycling can have different effects on the flow of energy than
231 the primarily consumptive effects of non-microbial members of the food web. Microorganisms
232 regulate many ecosystem functions, such as the decomposition of organic matter and the cycling
233 of nutrients (Falkowski et al. 2008, Krause et al. 2014). The presence of microbial taxa in an
234 ecosystem has the potential to influence internal cycling, but also play a direct role in controlling
235 the products leaving an ecosystem. Microbes may produce certain molecular compounds or

236 otherwise modify the composition of materials that flow into recipient ecosystems (Battin et al.
237 2008, Liang et al. 2017, Sokol et al. 2022). In addition, microbial necromass itself can constitute
238 a detrital component locally recycled or deposited in recipient ecosystems (Liang et al. 2019).
239 Modifications to the quality and quantity of resource flows can have effects on recipient
240 ecosystem functioning by mediating spatial functional complementarity (Osakpolor et al. 2023,
241 Pichon et al. 2023). When these local controls on material cycles are combined with microbial
242 dispersal, key microbial taxa may track favorable conditions across the landscape and
243 redistribute key ecosystem functions.

244 Compared to a simplified framework of two coupled ecosystems (e.g., Figs. 1-2), a meta-
245 ecosystem approach has greater generality and flexibility to explore more complex spatial
246 processes (Fig. 3). This is key because spatial flows span a range of spatial configurations
247 (Marleau et al. 2014). These can range from small aquatic ecosystems embedded in a terrestrial
248 landscape (Wisnoski et al. 2020, Caillon et al. 2021, Malazarte et al. 2022), to patchy
249 distributions of many aquatic and terrestrial ecosystems with heterogeneous spatial connectivity
250 across a larger spatial extent (Crump et al. 2007, Ruiz-González et al. 2015, Stadler and del
251 Giorgio 2022). A meta-ecosystem perspective can incorporate fluxes between neighboring
252 ecosystems as well as irregular connectivity networks across multiple ecosystems (Fig. 3). In
253 other words, a meta-ecosystem perspective is flexible enough to study unidirectional subsidies
254 between two ecosystems (e.g., terrestrial-aquatic linkages) as well as connections spanning
255 multiple ecosystem types across a landscape (Soininen et al. 2015, Gounand et al. 2018a, Collins
256 et al. 2020). The combination of organismal and material flows in meta-ecosystems can lead to
257 emergent feedbacks within and among ecosystems that affect diversity and ecosystem
258 functioning at different spatial scales (Marleau and Guichard 2019, Peller et al. 2022). The

259 congruence between microbial diversity and resources could influence functioning and the flow
260 of energy up through the food webs (Kayler et al. 2019). Spatial microbial flows can create
261 external controls on local nutrient recycling (or other functions) regulated by dispersal rates, with
262 microbial colonization dependent on spatial material flows.

263 **Meta-ecosystem controls on microbial fate**

264 One important factor regulating the impacts of spatial microbial flows is the fate of
265 microorganisms after dispersal into a recipient ecosystem (Székely et al. 2013, Wisnoski et al.
266 2020). If dispersed microorganisms survive, they constitute a spatial organismal flow that could
267 directly interact with the recipient community and modify its functioning (Fig. 4). If they die,
268 they instead serve as a spatial material flow that influences the recipient detrital pool and
269 indirectly shapes microbial and macrobial communities and the ecosystem functions they
270 perform. The fate of microorganisms, and thus the effective impact of spatial microbial flows in
271 the meta-ecosystem, can depend on how spatial material flows and local nutrient recycling
272 influence environmental conditions in the recipient ecosystem. Abiotic and biotic conditions in
273 the meta-ecosystem may inhibit establishment of microbial dispersers, support a persistent
274 population, or maintain a sink population via continued dispersal.

275 Microbial fate may be influenced by spatial material flows, which vary widely in the type
276 and magnitudes of their effects (Massol et al. 2017, Gounand et al. 2018a). Flows between
277 similar environments (e.g., a forest and a shrubland) may have weaker apparent effects on the
278 underlying spatial heterogeneity since both ecosystems share similar resources, detritus, and
279 possibly diversity, unless flux rates were strongly asymmetrical. Spatial flows between dissimilar
280 environments (e.g., an urban or agricultural region and a nearby estuary) could provide an influx

281 of new potential resources that may fuel productivity, shift community composition, and increase
282 niche opportunities in the recipient ecosystem (Stedmon et al. 2006, Soares et al. 2018,
283 Bambakidis et al. 2024, Crump and Bowen 2024). Consequently, they may have larger
284 consequences for microbial fate. Depending on the strength of these material flows at the meta-
285 ecosystem scale, the landscape could expand or contract in its suitability for dispersing
286 organisms, shifting patterns of diversity and function if material and organismal flows are at least
287 partially aligned in orientation.

288 Microbial fate also depends on whether dispersing microorganisms consume materials
289 derived primarily from the donor or recipient ecosystem. The composition of dissolved organic
290 matter can structure microbial communities (Muscarella et al. 2019, Bambakidis et al. 2024), and
291 may therefore influence the persistence of dispersing microorganisms. If dispersing microbes
292 primarily consume materials from their source ecosystem, whether they disperse along with
293 materials could determine not only microbial persistence but also energetic fluxes up the food
294 web. Regardless of their origin, bacteria can provide a relatively rapid transfer of terrestrial-
295 derived carbon to higher trophic levels primarily through the breakdown of low molecular
296 weight organic carbon sources (Berggren et al. 2010). More recalcitrant compounds tend to
297 accumulate due to their molecular complexity and because they gradually become too dilute to
298 find, both of which make them inaccessible to heterotrophic microorganisms (Zhang et al. 2018,
299 Kothawala et al. 2021). The supply of accessible compounds could subsidize microbial
300 heterotrophs and help stabilize their populations and functions. It is not clear whether the
301 microbial decomposers responsible for facilitating the transfer of terrestrial-derived organic
302 matter to aquatic food webs are maintained via dispersal from nearby terrestrial landscapes. If
303 dispersal were to contribute to decomposition of terrestrial-derived carbon, this would represent

304 external controls on both biodiversity and ecosystem functioning via coupled microbial and
305 material flows.

306 Coupling between microbial and material flows (e.g., co-transport) can enhance microbial
307 persistence and functioning in recipient ecosystems. For example, in the Pacific Northwest
308 (USA), red alder leaves fall into nearby rivers and are decomposed by aquatic organisms.
309 Experiments have shown that the bacterial communities involved in this process exhibit
310 consistent successional dynamics that contribute to faster decomposition of leaf litter from
311 nearby trees than conspecific litter from trees located further away (i.e., a home-field advantage),
312 partly owing to the presence of endophytic bacteria in the leaves prior to entering the aquatic
313 ecosystem (Jackrel et al. 2019). In other words, leaves were transported from terrestrial to
314 aquatic ecosystems with their intact microbiome, and this co-transport enhanced recipient
315 ecosystem function (i.e., decomposition). In addition to the co-transport of leaf litter and litter-
316 associated microorganisms (Marks 2019), there are likely many other scenarios where spatial
317 microbial flows are coupled with spatial material flows, but the implications of co-transport
318 across a broader array of meta-ecosystems have not been as well described yet.

319 Compared to macroorganisms, the fate of microorganisms in recipient ecosystems may
320 be more difficult to predict. Microorganisms may be somewhat flexible in the range of conditions
321 they can tolerate, which may be further complicated by their ability to switch into dormant states.
322 Dormancy could increase the lag time between arrival and growth stage in the recipient
323 ecosystem, which may be useful since dispersal can introduce organisms to environments that
324 are not yet favorable for growth. Interactions between dispersal and dormancy could allow
325 organisms to integrate across spatiotemporal variation in habitat suitability, promoting regional
326 persistence of important taxa (Buoro and Carlson 2014, Brendonck et al. 2017, Wisnoski et al.

327 2019, Wisnoski and Shoemaker 2022). Inputs of terrestrial-derived materials could further
328 interact with dispersal and dormancy to shape the composition of the metabolically active
329 portion of the bacterial community and overall resilience (Muscarella et al. 2016). In a meta-
330 ecosystem context, these impacts may become more entangled due to feedbacks with the abiotic
331 environment.

332 The magnitude of spatial microbial flows may also be relevant, as the number of
333 microorganisms exchanged between ecosystems could be high enough to generate persistence
334 (e.g., mass effects). Metacommunity theory has shown that dispersal rates can have strong
335 implications for community structure and compositional differences across the landscape, with
336 high rates of movement having a homogenizing effect on community structure and potentially a
337 detrimental effect on ecosystem functioning (Mouquet and Loreau 2003, Grainger and Gilbert
338 2016, Thompson et al. 2020). Empirical estimates of microbial flows across the terrestrial-aquatic
339 interface are rare, but evidence suggests they might be quite high, nearing 10^{20} cells/yr. in a large
340 Swedish lake (Bergström and Jansson 2000). This numerically important flux is likely to differ
341 across meta-ecosystems but indicates the potential for many possible colonization attempts
342 arising through microbial linkages.

343 **Landscape context of microbial flows**

344 Meta-ecosystems are distributed across landscapes that vary in their degrees of spatial
345 heterogeneity (Forman and Godron 1981, Pickett and Cadenasso 1995, Biswas and Wagner
346 2012). With increasing spatial heterogeneity, different ecosystem types are more likely to share
347 boundaries with one another along which spatial flows could occur. Depending on the differences
348 between donor and recipient ecosystems, ecosystem boundaries might represent gradual (e.g.,

349 grassland-to-forest) or steep (e.g., terrestrial-to-aquatic) environmental gradients between
350 ecosystems (Terborgh 1971). Landscapes with lower spatial autocorrelation may also have more
351 ecosystem boundaries relative to ecosystem area due to a higher density of interfaces. These
352 ecosystem interfaces may also be sites where the mixing of resources and organisms alleviates
353 resource limitation and increases productivity (Wetzel 1990, Naiman and Décamps 1997, Krause
354 et al. 2017). While not typically considered in patch-based perspectives on meta-ecosystems,
355 interface/ecotone habitats may be important sources of organismal and material flows in the
356 meta-ecosystem. Newer modeling techniques may offer potential ways to incorporate these
357 ecotones into the broader meta-ecosystem (Harvey et al. 2023). The overall spatial heterogeneity
358 in the meta-ecosystem (e.g., between all donor-recipient pairs) is likely to influence the degree of
359 spatial microbial flows at the landscape scale (Mony et al. 2020, 2022).

360 The mapping between spatial heterogeneity and meta-ecosystem dynamics depends on
361 the spatial configuration of ecosystems across the landscape. Spatial connectivity influences the
362 collection of donor-recipient ecosystem pairs in the meta-ecosystem because similar ecosystems
363 may have stronger or weaker connectivity depending on their distribution on the landscape and
364 the network of spatial flows that connect them (Little and Altermatt 2018, Peller et al. 2024).
365 Likewise, once the spatial network of connectivity is determined, spatial flows can vary in
366 direction (e.g., unidirectional, bidirectional, reciprocally alternating) and magnitude depending
367 on the processes that generate them (Nakano and Murakami 2001, Bartels et al. 2012, Leroux
368 and Loreau 2012, Fritz and Whiles 2021). Given the primary vectors of microbial dispersal
369 (Comte et al. 2017, Custer et al. 2022, Choudoir and DeAngelis 2022, Grupstra et al. 2022,
370 Chaudhary et al. 2022), the directionality of microbial flows may track the directionality of major

371 air currents, hydrological flowpaths, and animal/plant movement in relation to meta-ecosystem
372 structure on the landscape.

373 Ecosystems also vary through time in several properties that contribute to microbial
374 community dynamics. Fluctuations in temperature, resource inputs, salinity, or other factors can
375 drive fluctuations in the suitability of recipient ecosystems for successful colonization by
376 dispersing microorganisms. More spatially synchronous environmental fluctuations may promote
377 direct spatial microbial flows that contribute to diversity and ecosystem function in recipient
378 habitats, supported by dispersal from the donor system. Such spatial synchrony can promote
379 spatially synchronous fluctuations in ecosystem functioning, which could be a destabilizing
380 process for regional-scale ecosystem functioning (Wang and Loreau 2014, 2016, Lamy et al.
381 2021, Patrick et al. 2021). With respect to timing, spatial flows could be pulsed (Anderson et al.
382 2008, Piovia-Scott et al. 2019) or occur at a steady rate (Walters et al. 2018), spanning a range of
383 possible amplitudes, frequencies, and regularity. Depending on spatial flows and local
384 fluctuations, ecosystems could go through phases of spatial synchrony and asynchrony at
385 different timescales (Walter et al. 2017, Luo et al. 2021). The timing of exchanges relative to the
386 growth and survival rates of organisms that depend on these flows could be relevant because
387 pulses that are too short or infrequent to affect either donor or recipient ecosystem may have
388 minimal effects in the long run.

389 **Open questions**

390 Many questions remain about the viability, long-term persistence, and functional contributions of
391 externally derived microorganisms in recipient ecosystems. There are also opportunities to merge
392 ideas from the meta-ecosystem framework with complementary insights from the community

393 coalescence framework, which focuses on microbial and environmental mixing (Rillig et al.
394 2015, Mansour et al. 2018, Rocca et al. 2021, Custer et al. 2024). In addition, analytical
395 approaches from movement ecology may be helpful for considering how fluxes of individuals
396 scale across space and organizational hierarchies, aggregating individual movement into
397 implications at the community and ecosystem scales (Wisnoski and Lennon 2023). These
398 perspectives differ in their scope and range of processes but will likely yield new insights for a
399 more microbially compatible meta-ecosystem ecology.

400 A major outstanding question relates to whether microbial spatial flows are large enough
401 to have measurable impacts on recipient ecosystems. In the case of terrestrial-derived
402 microorganisms in aquatic ecosystems, one might ask whether bacteria maintained by dispersal
403 from the watershed help decompose complex forms of organic matter and cycle carbon back into
404 higher levels of the aquatic food web. Meta-ecosystem experiments (Scherer-Lorenzen et al.
405 2022, Châtillon et al. 2023) may help uncover the conditions under which eliminating microbial
406 spatial flows may reduce functioning locally and at the meta-ecosystem scale. To understand
407 these functional implications, it will be important to better characterize the fate of microbial
408 dispersers across a range of donor-recipient ecosystem pairs, investigating temporal dynamics,
409 microbial compositional shifts, and multiple ecosystem functions.

410 Another open question is how variable spatial flows are across space and over time and
411 how such variability affects microbial community dynamics and meta-ecosystem functioning.
412 Early descriptions are primarily snapshots in time and patchily distributed across space (but see
413 Hermans et al. 2020). There may be predictable sources of variation, such as the seasonal
414 variation in hydrological flow that regulates connectivity. For example, river flood plains go
415 through phases of connectivity with nearby ecosystems (Junk et al. 1989), stream networks

416 expand and contract within the watershed (Stanley et al. 1997, Benda et al. 2004), and
417 intermittent ecosystems exhibit phases of spatial isolation (Datry et al. 2014, Butterworth et al.
418 2023). At broad spatial extents, there may be geographic patterns in the degree of spatial flows
419 between ecosystems (Montagano et al. 2019). Anthropogenic modifications to the landscape may
420 also be introducing long-term shifts in spatial flows over decadal scales, for example, by
421 decreasing connectivity between adjacent ecosystems and severing spatial flows (Wesner et al.
422 2020). Data collected from a variety of ecosystems can be used to describe variation in the
423 magnitude and directionality of microbial flows, which can influence the range of phenomena
424 analyzed in meta-ecosystem models to generate novel predictions.

425 **Conclusions**

426 The aim of this paper is to describe a potentially overlooked organismal spatial flow in meta-
427 ecosystem ecology arising from cross-boundary microbial dispersal. Microbial spatial flows
428 seem to be common at the terrestrial-aquatic interface, where a substantial portion of aquatic
429 microbial diversity appears to be derived from surrounding terrestrial ecosystems. Across a more
430 general range of ecosystem types, the functional implications of microbial spatial flows are likely
431 to depend on the metabolic activity of allochthonous microorganisms in relation to the local
432 ecosystem conditions that regulate functioning. These can be further modified by other spatial
433 flows across the landscape, which likely depend on spatial configuration of ecosystems and their
434 connectivity, temporal variation within and between ecosystems, and the magnitude and
435 direction of dispersal vectors. Many open questions remain to understand the importance of
436 microbial spatial flows in different ecosystem types, and these questions may be addressed with
437 simulation models and creative lab and field experiments.

438

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446

Author contributions

447 NIW conceived of the project and wrote the paper. NIW declares no conflicts of interest.

448

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Figure captions

845 Figure 1 – A meta-ecosystem containing two similar ecosystem types with an explicit microbial
846 community. Within each ecosystem, inorganic nutrients (N) are consumed by primary producers
847 (P) and microorganisms (M), both of which are consumed by grazers (G), which are then eaten
848 by carnivores (C). All biotic components contribute to detritus (D). Detritus is converted back
849 into inorganic nutrients at a rate that depends on the composition of the microbial community
850 (dashed “t” arrow). Between ecosystems, spatial flows occur between all compartments (dashed
851 arrows), where microbial dispersal (dashed arrow “d”) relates to the movement of individual taxa
852 between each ecosystem, with each taxon contributing to ecosystem functioning (decomposition
853 rate, shown here).

854

855 Figure 2 – A meta-ecosystem containing two contrasting ecosystem types with an explicit
856 microbial community. Within each ecosystem, inorganic nutrients (N) are consumed by primary
857 producers (P) and microorganisms (M), both of which are consumed by grazers (G), which are
858 then eaten by carnivores (C). All biotic components contribute to detritus (D). Detritus is
859 converted back into inorganic nutrients at a rate that depends on the composition of the microbial
860 community (dashed “t” arrow). Between ecosystems, spatial flows (dashed arrows) directly link
861 non-living components (detritus and inorganic nutrients). However, due to dissimilarity between
862 ecosystem types, survival of primary producers, grazers, and carnivores is low, so spatial flows
863 contribute directly to the detrital pool in the other ecosystem. Microbial dispersal (dashed arrow
864 “d”) relates to the movement of individual taxa between each ecosystem, with each taxon
865 contributing to ecosystem functioning (decomposition rate, shown here). Some taxa can survive
866 dispersal across the ecosystem boundary, joining populations in the recipient ecosystem, but

867 other taxa may not survive and instead get shunted into the detrital pool, where they are then
868 recycled at rates that depend on microbial community composition (dashed “t” arrows).

869

870 Figure 3 – Complex spatial configurations of aquatic and terrestrial ecosystems linked by spatial
871 material flows (blue and green arrows) and spatial organismal flows (black arrows). Depicted
872 here is simply an example of the type of connectivity one might observe. (A) A meta-ecosystem
873 representation of the connectivity shown across the landscape in panel (B). Material flows may
874 be aligned with microbial flows or not. Biotic and abiotic components are linked by microbial
875 functions (dashed arrows) that transform materials within ecosystems and allow them to be
876 transported in organic or inorganic forms.

877

878 Figure 4 – The fate of microbial dispersers in recipient ecosystems relates to the functional
879 implications of spatial microbial fluxes in the meta-ecosystem. If dispersers survive and remain
880 metabolically active, then they are likely to contribute to contemporary ecosystem functioning. If
881 dispersers survive but remain metabolically inactive, then they exist with the potential to
882 contribute to functioning at some point in the future, should they survive until conditions are
883 more favorable. If dispersers die, they become detritus and do not contribute directly to
884 functioning (but may indirectly support ecosystem functioning as a substrate for decomposers).

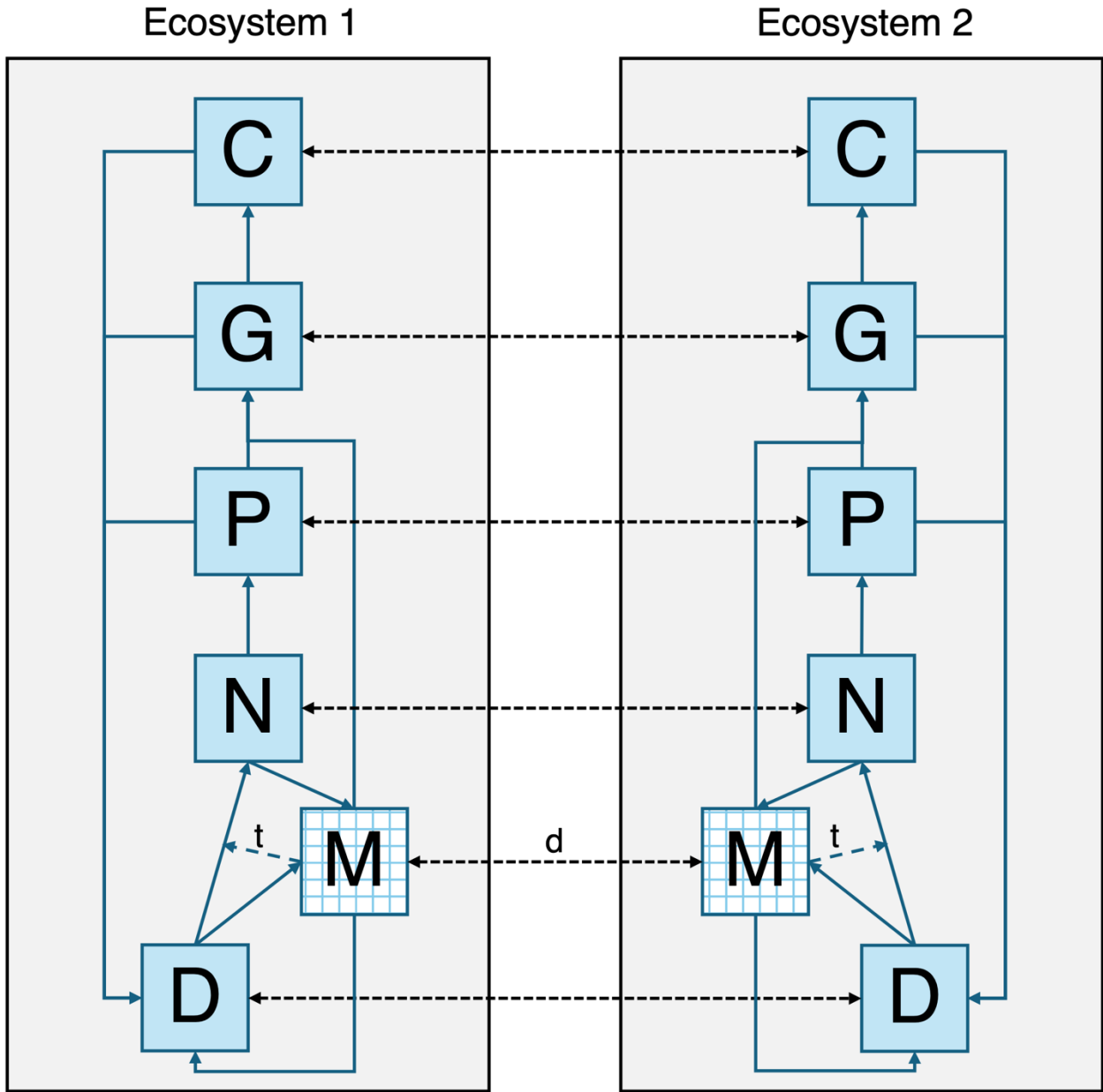
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Figures

888 Figure 1.

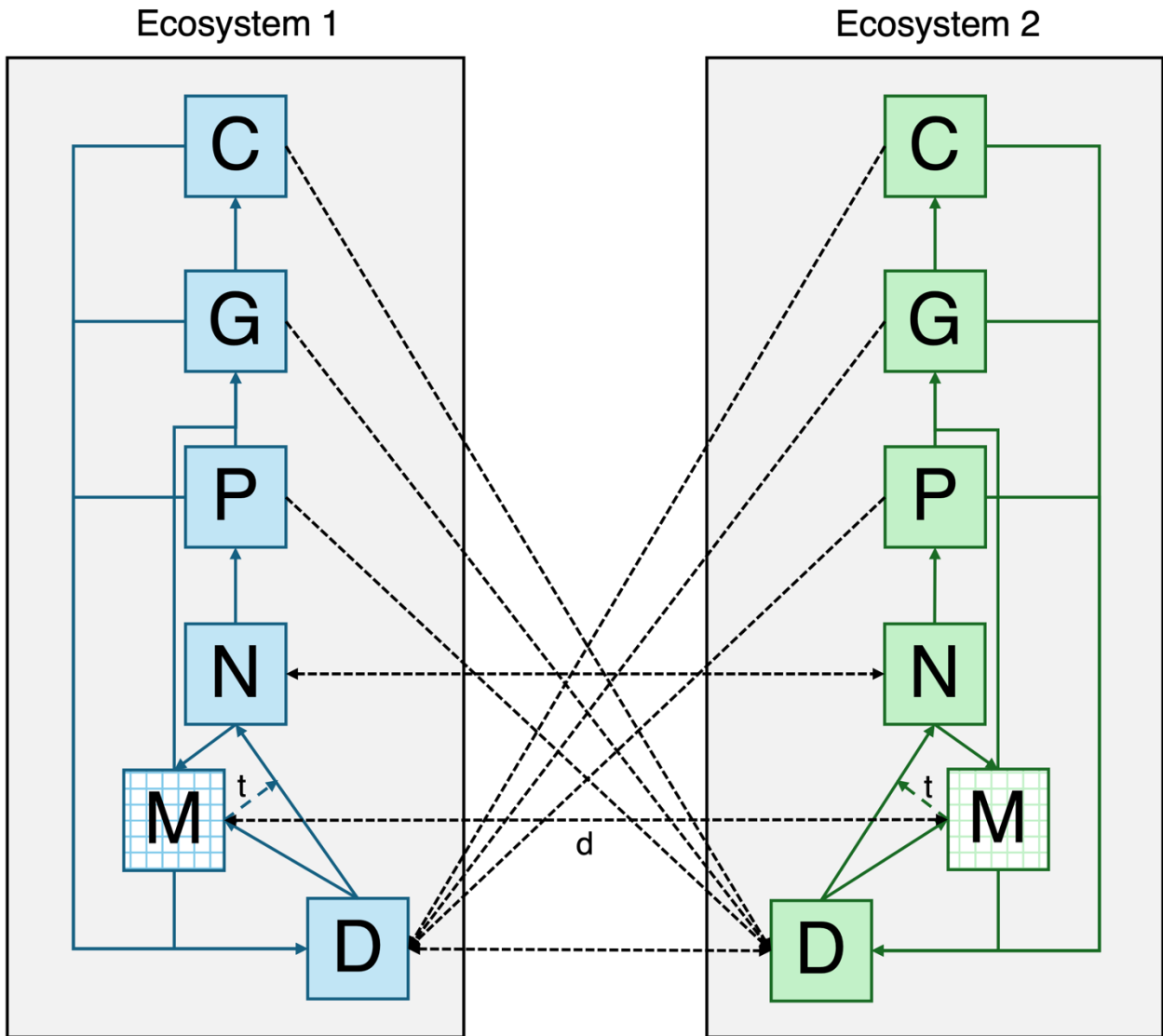


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892 Figure 2.

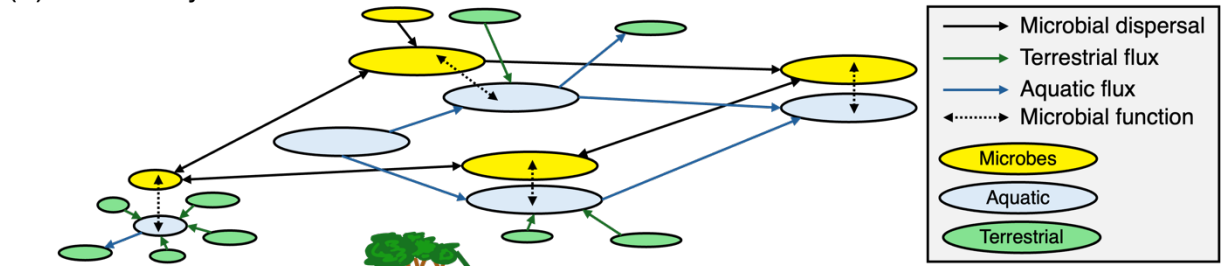


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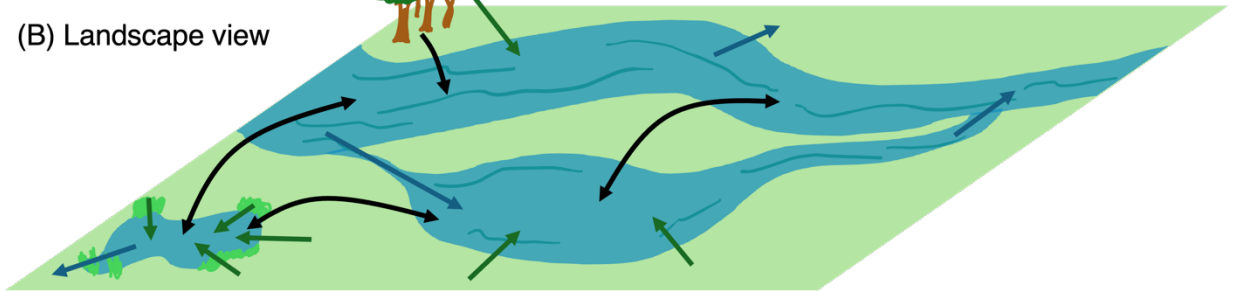
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895 Figure 3.

(A) Meta-ecosystem view



(B) Landscape view

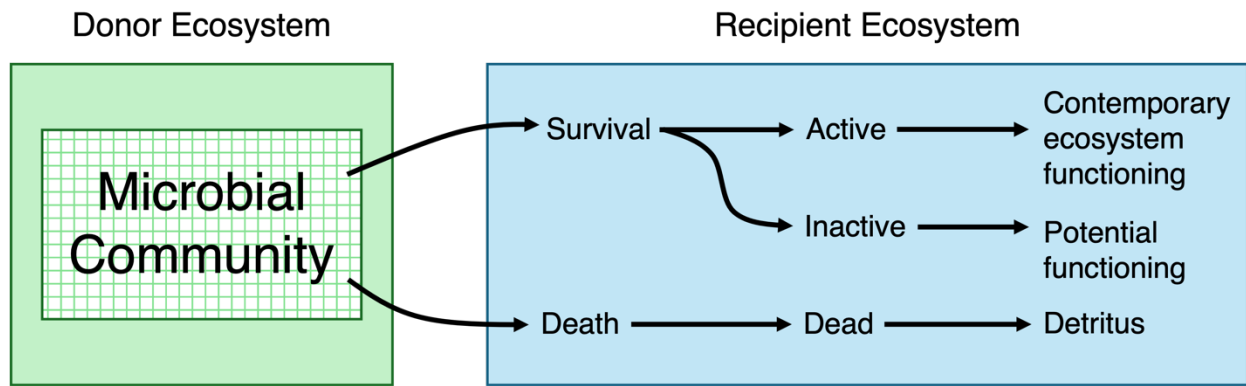


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899 Figure 4.



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