

1 **Harnessing social media data to track species range shifts**

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32 **Abstract**

33 Biodiversity monitoring programs and citizen science data remain heavily biased towards the Global
34 North. Incorporating social media data can complement existing gaps, especially in megadiverse
35 countries with limited records, but whether such data can significantly improve our understanding of
36 range-shifting species is unknown. Here, we collated locality data from Flickr and Facebook, in addition
37 to occurrence data from the Global Biodiversity Information Facility (GBIF). We tested whether social
38 media data improved our knowledge of the range dynamics of a rapid range-shifting butterfly, tawny
39 coster (*Acraea terpsicore*), when compared to GBIF-only data. Social media data increased occurrence
40 records by 35%. The proportion of social media data was higher in countries poorly represented on
41 GBIF; however, we also obtained new distributional information from well-represented countries (e.g.,
42 Australia and Malaysia). We constructed ecological niche models (ENM) with data from both sources.
43 ENMs based on the social media data highlighted greater expansion rates to higher latitudes and
44 elevations compared to ENMs based on GBIF data only. Our results highlight the potential of harnessing
45 social media data to track biodiversity redistribution in response to climate change.

46

47 **Main**

48 The current era is marked by a pressing biodiversity crisis (Dirzo et al., 2014; Pimm et al., 2014; Diaz et
49 al., 2019), driven by various factors such as agricultural intensification and expansion, habitat loss and
50 fragmentation, biological invasions, and climate change (Wilson, 1989; Butchart et al., 2010; Joppa et al.,
51 2016; Maxwell et al., 2016; Jaureguiberry et al., 2022). In response to these combined pressures on the
52 environment, many species have shifted their distributions (Chen et al., 2011; Yackulic et al., 2011;
53 Lenoir & Svenning, 2015; Lenoir et al., 2020; Chan et al., 2024). While range-shifting can be feasible for
54 migratory and highly mobile species, which can easily expand towards new areas (Chowdhury et al.,
55 2021a, b), it poses a significant challenge to narrow-range and low-mobility species that may be unable
56 to reach new suitable habitats, and thus may be doomed to range contraction and ultimately extinction
57 (Pound et al., 2006; Freeman et al., 2018; Chowdhury, 2023). Many taxonomic groups contain species
58 that have recently shifted range size, including insects (McCain & Garfinkel, 2021), mammals (Santos et
59 al., 2017), birds (Rushing et al., 2020), amphibians (Nowakowski et al., 2017), plants (Auffret & Svenning,
60 2022; Iseli et al. 2023), and marine organisms (Poloczanska et al., 2016). When a species migrates to a
61 new habitat, it may be restricted to similar climatic conditions (niche-conserving species; Wiens et al.,
62 2010) or adapt to different ones (niche-shifting species; Guisan et al., 2014; Di Marco et al., 2021). For
63 instance, a native range shifter may track shifting isotherms altitudinally or latitudinally to remain within
64 the same climatic space, while a non-native invader may spread from its introduction point to exploit
65 new climatic spaces. The colonisation or extirpation rates of a range-shifting species (being native or
66 non-native) depends largely on prevailing landscape conditions: if only a small amount of landscape is
67 suitable, the establishment of a founding population might be uncertain or slow, or the founding
68 population may go extinct due to Allee effects (Kuussaari et al., 1998; Hodgson et al., 2012; Blackburn et
69 al., 2016).

70 The issue of biodiversity redistribution in response to human activities is a global concern (Lenoir &
71 Svenning, 2015; Lenoir et al., 2020). However, our understanding of the issue is limited and biased, both
72 geographically and taxonomically (Lenoir & Svenning, 2015; Feeley et al., 2017). For instance, a
73 systematic review of 258 peer-reviewed studies reporting species range shifts revealed that nearly 40%

74 of the identified range shifts (12,415 species) were from flowering plants, and 22%, 4%, 12% and 0.5%
75 were from insects, fish, birds and mammals, respectively (Lenoir et al., 2020). For plants, the data is
76 strongly biased towards flowering plants with very few records for ferns, mosses, etc. while for animals,
77 the data is strongly biased towards birds and fish. Likewise, in insects, there is a strong bias towards
78 specific groups, including orthoptera and dragonflies (Lenoir et al., 2020). This taxonomic bias is
79 compounded by a severe geographical bias, with the majority of studies coming from countries of the
80 Global North and only a limited number from the tropics and the Global South (Lenoir et al., 2020;
81 Lawlor et al., 2024; Parker et al., 2024), including many megadiverse countries. This supports the claim
82 that researchers' efforts to document species range shifts have been heavily biased towards certain
83 regions and taxa, meaning none of the 'global reviews' are truly global (Feeley et al., 2017). Assessing
84 species range shifts requires detailed species distribution data, typically unavailable for species from the
85 Global South (Hortal et al., 2015; Hughes et al., 2021). This, in turn, biases our global overview of species
86 undergoing climate-induced range shifts. To improve our understanding of species redistribution at a
87 global scale, it is essential that we compile data from all available sources to better test how species are
88 responding to global change drivers.

89 With the rapid technological advances involving mobile phones, digital cameras, and fast internet has
90 revolutionised data collection (Van Klink et al., 2022; Sheard et al., 2024), but also anyone from
91 anywhere in the world can now share their biodiversity observations on a variety of digital platforms
92 (e.g., apps, web pages, and social media; Chandler et al., 2017; Pocock et al., 2018; Toivonen et al., 2019;
93 Caley & Cassey, 2023). If appropriately harvested, such data can be crucial in filling knowledge gaps in
94 biodiversity distribution and monitoring and thus help us answer questions about how species respond
95 to global changes (Jarić et al., 2020; Soriano-Redondo et al., 2024). This is especially relevant for
96 countries where systematic biodiversity monitoring programs are uncommon and which are poorly
97 represented in global biodiversity repositories (Chowdhury et al., 2023a; Marcenò et al., 2021; Mota et
98 al., 2022). Indeed, recent studies have demonstrated that biodiversity data posted on Facebook can
99 sometimes be much more comprehensive than the data available from the Global Biodiversity
100 Information Facility (GBIF) (Chowdhury et al., 2023a). Using a comprehensive set of biodiversity data in
101 ecological research is key in assessing species redistribution in the context of global change. For
102 example, combining social media data with GBIF data can substantially improve conservation
103 assessments (Chowdhury et al., 2024a), which can be useful in identifying priority conservation areas
104 (Chowdhury et al., 2024b). Nevertheless, biodiversity observation data from social media have rarely
105 been used in conservation assessment studies (Di Minin et al., 2015) and for tracking range-shifting
106 species (Sbragaglia et al., 2024). Here, as a proof of concept, we aim to address this important
107 knowledge gap by harnessing social media data to evaluate the distribution pattern of an ecologically
108 important species that is undergoing a rapid range expansion in response to global changes: the tawny
109 coster butterfly (*Acraea terpsicore*).

110 We compiled the distribution data of the tawny coster butterfly from social media (Facebook and Flickr),
111 and combined these with GBIF records to assess the extent to which social media data adds new locality
112 information where it occurs. We further analysed the data into two groups: GBIF-only and combined
113 (GBIF and social media) data. We organised the data in five time intervals of four years each and fitted
114 ecological niche models to identify if the geographic range distribution of this species would differ from
115 the known range extent (Chowdhury et al., 2021a) and how this would impact the calculation of range
116 expansion rates. Finally, we ran a niche assessment analysis to identify whether the additional social

117 media data identified novel combinations of the environmental niche space. Building on our findings, we
118 provide recommendations on how to use social media data to answer general ecological questions
119 relating to species' biogeography and macroecology.

120

121 **Results**

122 Our cleaned and compiled dataset included 6459 occurrence records, of which 65% (4206) were from
123 GBIF and 35% (2253) were from social media (Flickr: 5%, Facebook: 30%). We noticed marked
124 differences between data sources when we compared the number of occurrence records across
125 countries (Figure 1). For most countries, the number of occurrence records substantially increased after
126 including social media data (e.g., data increased from 10 to 224 for Bangladesh, 262 to 468 for Malaysia;
127 Figure 1A, B). The distribution of the tawny coster butterfly is known from 17 countries (Chowdhury et
128 al., 2021a), and we obtained a higher percentage of data than GBIF from social media for five of these
129 countries (Figure 1A, B; Supplementary Data S1). In countries with the most occurrence records, the
130 percentage of the data coming from social media was generally lower, but >10% in all cases (range 10-
131 44%). For example, we obtained 3096 occurrence records from India, of which 64% (1968) were from
132 GBIF, and 36% (1128) were from social media (Figure 1B). The percentages of species occurrence
133 records from social media were higher for countries with lower number of total occurrence records
134 (Figure 1B).

135 When we analysed the temporal distribution of occurrence records, we noticed substantial differences
136 over the years (Figure 1C). While the initial period, from 2005 to 2007, had a larger percentage of
137 occurrence records from GBIF, subsequent years (2008-2018) were characterized by a higher percentage
138 of occurrence records from social media, except for 2013. Following a substantial decline during 2017-
139 2022, the proportion of social media data stabilized recently (Figure 1C).

140 The addition of social media data in ecological niche models contributed substantially to the
141 identification of potentially new suitable areas for the tawny coster. The total surface area predicted to
142 be suitable for the tawny coster is in general larger when combining social media data with GBIF data
143 than when relying on GBIF-data only, at least during the periods 2005-2008, 2017-2020 and 2021-2024
144 (Figure 2). For the period 2009-2012, we found the opposite pattern. New suitable areas identified with
145 the addition of social media data were mostly distributed in South Asia (especially towards higher
146 elevations in the Indo-Himalaya region) during 2005-2020, while new suitable areas identified with the
147 addition of social media data during 2021-2024 were distributed throughout the entire region and
148 especially towards higher latitudes (Figure 2). When we combined all the suitability maps over the five
149 time intervals, we found that predictions from the models relying on GBIF-only data missed many areas
150 at higher latitudes and at mid- to high elevations (Figure 2G). In terms of range expansion, the combined
151 data captured a larger expansion area initially; however, it slightly declined afterwards and increased
152 again recently (Figure 2H).

153 To understand the benefit of adding social media data to fit our models, from the perspective of the
154 environmental niche space, we conducted a Principal Component Analysis (PCA) at each of the five
155 studied time intervals separately for both GBIF and combined data. The PC1 and PC2 axes explained
156 from 59-63% of the total variance. Across all 5 consecutive time periods, the model combining GBIF data
157 with social media data captured a broader niche space than the model relying on GBIF-only data (Figure

158 3). While the niche overlap between the model relying on GBIF-only data and the model combining
159 social media data with GBIF data was fairly large, reaching 65% and 76% during 2005-2008 and 2009-
160 2012, respectively, the overlap was much smaller during the three subsequent periods (49%, 37%, and
161 39% during 2013-2016, 2017-2020, and 2021-2024, respectively). The PCA identified precipitation
162 (monthly total), maximum temperature, and elevation to be the most important environmental
163 variables determining the differences in the covered environmental niche space between the model
164 relying on GBIF-only data and the model combining both GBIF and social media data. Across time
165 intervals, the GBIF-only data failed to capture regions with lower maximum temperature, lower
166 precipitation (monthly total), and higher elevation.

167

168 **Discussion**

169 Social media data can help reduce the global biodiversity data shortfall (Di Minin et al., 2015;
170 Chowdhury et al., 2023a) and improve our understanding of biodiversity (re)distribution for
171 conservation purposes (Chowdhury et al., 2024a). Yet, such data have rarely been used in large-scale
172 studies (Di Minin et al., 2015). Using standardised protocols to scrape data from social media, we
173 harvested more occurrence data for modelling the potential redistribution of a rapidly expanding
174 species than would typically be used in traditional distribution modelling studies that rely on GBIF data
175 only. We demonstrated that social media data can identify new distribution: occurrence records
176 increased by 53% (4206 to 6459). For at least some of the recent years, the percentages of social media
177 data decreased slightly, which could be explained by COVID pandemic and lockdowns (Chowdhury et al.,
178 2024b), people travelled less and reduced tourism could have resulted in less biodiversity records
179 posted. When combining occurrence records from social media with those from GBIF and fitting models
180 to project habitat suitability maps, the suitable area of potential occupancy increased for three of the
181 five studied time intervals, and the total amount of additional areas increased over time compared to a
182 model relying on GBIF data only.

183 Systematic biodiversity monitoring programs and citizen science provide important biodiversity data
184 sources for scientists and conservation biologists (Mesaglio & Callaghan, 2021). In addition, social media
185 data has a key potential to improve our basic understanding of species' distribution and spread, even in
186 better surveyed (e.g., developed) countries. For example, biodiversity data from Australia is
187 comparatively well-represented in global biodiversity repositories like GBIF, but we still obtained many
188 new localities from social media that represented uncharted conditions, from a GBIF perspective, within
189 the climatic space. The total number of occurrence records increased by 12% (440 to 493) and the
190 suitable area of potential occupancy increased by 9% (1.64 million km² to 1.79 million km²). Such gains
191 were even more pronounced for meagdiverse countries of the Global South, such as Bangladesh, where
192 the total number of occurrence records retrieved from social media was 22.4 times higher than from
193 GBIF. This illustrates the enormous potential of social media data to reduce the global biodiversity data
194 shortfall as a means to better track range-shifting species. By doing so, we managed to identify many
195 new localities (at higher latitudes and at higher elevations, chiefly representing climate conditions from
196 colder environments with lower maximum temperature and lower precipitation) that the tawny coster
197 butterfly might colonise in the future in response to climate warming. We also showed that social media
198 data helped to capture a broader niche space exploited by the tawny coster butterfly, some of which
199 were not captured by the model relying on GBIF-only data.

200 We used two social media channels – Facebook and Flickr – to harvest more occurrence data for the
201 tawny coster butterfly. While the additional data substantially improved the performances of our
202 species distribution model and niche assessments, we faced several obstacles. First, we used machine
203 learning to automatically scrape data from Flickr (following Hausmann et al., 2018) while we had to
204 manually extract data from Facebook (following Chowdhury et al., 2024b), which was a time-consuming
205 task compared to an approach that relies solely on artificial intelligence (AI). In the future, it should be
206 possible to develop an automated approach to extract species' occurrences from Facebook, which
207 would save a substantial amount of time (Jarić et al., 2020; Correia et al., 2021; Chowdhury et al.,
208 2024b). For example, Castro et al. (2024) showed that the success rate of AI models in extracting
209 information from unstructured text is quite high, making them valuable tools for managing ecological
210 data efficiently. Second, we faced two major data issues when using Flickr: photographs with no location
211 data and photographs erroneously flagged as the tawny coster. Because of that, we could only use 5% of
212 the data we initially scraped from Flickr. To handle this issue, it is important to improve the Flickr data
213 extraction process, by carefully checking individual photographs, and validating whether or not they
214 represent what we are looking for. Finally, many photographs shared on social media might not be the
215 species the photographers assume to be. To handle this issue, having someone in the group with
216 taxonomic expertise is essential, especially the people who are extracting records from social media and
217 validating species information.

218 It is important to think strategically to get the maximum value from social media data. We only used
219 Facebook and Flickr as social media platforms, which tend to be less popular in some countries.
220 Including other popular platforms for particular countries (e.g., Weibo in China or possibly Instagram in
221 other countries) could provide many more new records. We recommend future studies assessing data
222 quality performance across several social media channels. Furthermore, we had to remove many
223 records due to data quality issues (e.g., the locations were unspecified, and photographs were unclear).
224 To solve these issues, group moderators are needed and should maintain strict rules about sharing
225 biodiversity observations so that everyone knows the species' details.

226 Although open data would revolutionise scientific research, it is important to think differently in the
227 case of threatened species, as such data can increase threats (e.g., poaching, disturbance; Bergman et
228 al., 2022; Di Minin et al., 2015, 2022). Group moderators and regional legal authorities should deal with
229 such issues. When using social media data to extract biodiversity data, personal information should be
230 carefully handled and potential intentional and unintentional physical and mental harm to the
231 photographers should also be carefully considered (Di Minin et al., 2021). To mitigate these risks and
232 ensure user safety, adequate practices such as data minimisation, anonymisation, and strict data
233 management protocols should be adopted (Di Minin et al., 2021).

234 The Kunming-Montreal Global Biodiversity Framework (CBD, 2022) aims to ensure the best data
235 available for conservation assessments (Target 21), and to protect 30% of the Earth by 2030 (Target 3).
236 Here, by comparing data distribution between the most comprehensive global biodiversity repository
237 (GBIF) and social media, we showed that biodiversity data shared on social media can improve scientific
238 knowledge on species distributions, even in countries that are well represented in global biodiversity
239 repositories like GBIF or iNaturalist. Due to environmental changes, range-shifting species (including
240 invasive species) are expanding rapidly, and social media data are especially powerful in this situation as
241 they allow for almost real-time monitoring, which is not typically possible when relying on GBIF data
242 solely. This makes social media data especially useful to set up 'early warning' systems of species

243 colonisation (Soriano-Redondo et al., 2024). There is potentially even more data available if we develop
244 more powerful digital tools. In addition to other platforms that are currently difficult to access (e.g.,
245 Instagram), there is also incidental (or secondary) biodiversity data (e.g., posted photographs of flowers
246 that, by chance, have a butterfly on them; Pernat et al., 2024). These data might come into play with
247 improvements in automated species recognition tools. Such approaches can help better understand and
248 track ongoing species' movements and future biological invasions (Capinha et al., 2024; Cardoso et al.,
249 2024). The current conservation literature is highly biased, chiefly stemming from North America and
250 Western Europe (Di Marco et al., 2017; Dawson et al., 2024), because of significant and long-standing
251 human capacity limitations in the tropics. Our findings suggest that combining data from multiple
252 sources can eventually help answer key ecological questions, especially for countries with limited
253 biodiversity observation records currently registered in global biodiversity repositories.

254

255 **Methods**

256 **Tawny coster**

257 The tawny coster has a well-documented geographic range area (Chowdhury et al., 2021a), and its
258 charismatic status, like many butterfly species, attracts high public attention on social media. This
259 butterfly is native to the Indian subcontinent (India, Bangladesh, and Sri Lanka) (Braby et al., 2014;
260 Chowdhury et al., 2021a) and since the 1980s it has rapidly expanded its range to other parts of South
261 Asia (e.g., Bhutan, Nepal, and Pakistan) and Southeast Asia (e.g., Malaysia, Singapore, and Thailand),
262 eventually entering Australia in April 2012 (Braby et al., 2014). It was first recorded in Australia near
263 Darwin in the Northern Territory (Sanderson et al., 2012). In subsequent years, the species started to
264 spread towards Western Australia (till Kimberly), before the spread abruptly shifted towards
265 Queensland. Since its arrival in Australia, the tawny coster has expanded within the country at a rate of
266 approximately 135 km/year, while remaining within its native climatic niche (Chowdhury et al., 2021a).
267 The species can cover a wide range of habitats and can migrate long distances (Chowdhury et al.,
268 2021b), which might have facilitated its documented expansion.

269

270 **Data**

271 We collated occurrence records and locality data for tawny coster's sightings from three different
272 sources: GBIF, Flickr, and Facebook. For all the data sources, we selected data from January 2005 to May
273 2024 to maintain a comparable sampling period. We removed potential duplicates in occurrence records
274 using two approaches: (I) exact duplicates (i.e., rows containing the exact same values for all the
275 columns), that we used to compare the distribution of occurrence data, and (II) occurrence records
276 falling within the same 4.65×4.65 km² grid cells (only using longitude and latitude coordinates), which
277 we used for niche modelling and the niche overlap analysis.

278 We downloaded GBIF data manually from the website (<https://www.gbif.org/>; GBIF, 2024). The GBIF
279 portal is a collection of hundreds of citizen science applications (Heberling et al., 2021), so we did not
280 look for citizen science data from other potential sources (e.g., iNaturalist). When downloading
281 occurrence records from GBIF for the tawny coster, we kept only the presence data with coordinate
282 uncertainty below 10 km. It should be noted that many GBIF occurrence records lack information on

283 coordinate uncertainty, and our choice to ignore these records resulted in a reduced (but more reliable)
284 sample.

285 We used a Python (<https://www.python.org/>) script, which uses the Flickr's application programming
286 interface (API) (<https://www.flickr.com/services/developer>) and its keyword search, to collect all
287 publicly available Flickr posts related to the tawny coster. We used the scientific name and the English
288 common name of the species as a set of keywords for the searches. We then deduplicated the data and
289 removed any posts not containing a geotag. Using the URLs of the posts, we manually double-checked
290 all the photographs of the tawny coster.

291 For Facebook data, we followed the protocol developed by Chowdhury et al. (2024b). Specifically, the
292 entire data extraction process was divided into three steps. First, we searched for butterfly groups using
293 a combination of taxon and country names (Supplementary Table S1). Here, we collected all the known
294 distribution (17 country names used as keywords) of the tawny coster from Chowdhury et al. (2021a).
295 When searching for Facebook groups, we included ten more countries from the surrounding area. With
296 these 27 keywords corresponding to 27 countries, we identified 41 Facebook groups from 17 countries
297 (Supplementary Table S1). Second, in each Facebook group, we searched twice using both the scientific
298 name (*Acraea terpsicore*) and the common name (tawny coster). We carefully went through each
299 photograph and validated the species' information. From each photograph, we extracted date (day,
300 month, year), location, and photographer's information. We excluded photographs if their quality was
301 unsuitable for identification up to the species level, if a specific date and location were not provided,
302 and if the location provided in the photographs was unspecific (> 10 km uncertainty). Finally, we used
303 Google Maps (<https://www.google.com/maps>) to georeference the location information and get the
304 longitude and latitude coordinates.

305 During the initial data cleaning process, we removed all duplicate records (same coordinates) and only
306 kept records between January 2005 and May 2024. Our compiled dataset included 6459 occurrence
307 records (GBIF: 4206; Flickr: 325; Facebook: 1928). We provide the Facebook and Flickr data in the
308 supplementary material (Supplementary Data S1), while the GBIF data is publicly available (GBIF, 2024).

309 We used the TerraClimate database (<https://www.climatologylab.org/terraclimate.html>; Abatzoglou et
310 al., 2018) to obtain climatic predictor variables at a yearly resolution (2005-2023) at 21.625 km²
311 resolution. The climatic data for 2024 is yet to be published. We downloaded ten climatic predictor
312 variables from TerraClimate: actual evapotranspiration, climate water deficit, potential
313 evapotranspiration, precipitation (monthly total), soil moisture, maximum temperature, minimum
314 temperature, wind speed, and the Palmer Drought Severity Index. We downloaded the elevation data
315 from the WorldClim (<https://www.worldclim.org/data/worldclim21.html>) database at the same
316 resolution, which corresponds to 21.625 km² resolution at the equator (4.65 km × 4.65 km).

317

318 **Data preparation**

319 We analysed range-shift dynamics of the tawny coster by splitting the 2005-2024 period into five
320 intervals of four years each (2005-2008; 2009-2012; 2013-2016; 2017-2020; and 2021-2024) and
321 assigned the occurrence and climatic data, available at a yearly resolution, accordingly to each of those 5
322 periods. For the occurrence data, separately for the GBIF and social media datasets, we grouped the

323 observations into year intervals and kept a single occurrence record per grid cell if several occurrences
324 from subsequent years of the same period were available at 21.625 km². Afterwards, we compared the
325 number of occurrence records between the GBIF-only dataset and the social media dataset and kept the
326 exact same number of records between both datasets, using a randomized sampling procedure in R. For
327 example, if the GBIF and social media data contained X and Y occurrences, respectively, for a given
328 period t , with $X > Y$ at t , then we subsampled X into a smaller subset x such that $x = Y$. By doing so, we
329 managed to balance the sampling effort between the GBIF and social media data, thus limiting any
330 improvement in model accuracy between both datasets that would be due to sample size. Finally, we
331 merged the GBIF and social media datasets to create the combined dataset, so we had two datasets for
332 the subsequent analyses GBIF-only and combined datasets.

333 For the environmental data, we cropped the layers to the study extent ($x_{min} = 60.875$, $x_{max} = 158.9583$,
334 $y_{min} = -54.75$, $y_{max} = 53.54167$) and calculated the mean climatic conditions over the four years of each
335 time period and did that for each of the ten climatic variables (there was no yearly elevation data).
336 Given that the climatic data for 2024 is unpublished, for the last time period (2021-2024) we computed
337 mean climatic conditions over three years, instead of four years. We checked for multicollinearity issues
338 among the predictor variables and removed highly correlated ones ($|r| > 0.75$). Consequently, we
339 removed four variables and kept seven remaining variables for the final analysis: climate water deficit,
340 precipitation (monthly total), soil moisture, maximum temperature, wind speed, Palmer Drought
341 Severity Index, and elevation.

342

343 **Habitat suitability maps**

344 To obtain habitat suitability maps for the tawny coster butterfly, we fitted MaxEnt species distribution
345 models (SDMs) (Elith et al., 2010; Phillips et al., 2017) in R (R Core Team, 2024), using the ENMeval
346 package (version 2.0.4; Kass et al., 2021). We ran the model twice for each of the five periods, once for
347 the GBIF-only dataset and a second time using the combined datasets.

348 We fitted SDMs using the following settings: seven predictor variables (i.e., the ones selected after
349 removing highly correlated ones, see the previous section entitled 'Data preparation'), and 10,000
350 randomly generated background points at 21.625 km² resolution (2.5 arc minute). For all five time
351 periods and for both the model relying on GBIF-data only and the model relying on the combined
352 dataset, we used the exact same set of pseudo-absences by randomly selecting 10,000 background
353 points across the entire study extent (Supplementary Figure S2). We did that to avoid the background
354 selection strategy to affect the model outputs when comparing model performances over time and
355 between data sources (GBIF-only vs. combined data). Before fitting the model, we removed duplicate
356 values in each raster pixel and created a 500 km buffer around the spatial records. We cropped the
357 environmental variables to the buffered region to limit model overfitting. We assigned the records to
358 grid cells and then randomly assigned grid cells to particular folds (Kass et al., 2021). We used the
359 'checkerboard2' evaluation method (with the presence and background points), which handles
360 overinflation of model performance, at least from biased sampling. This evaluation method partitions
361 geospatial records and background points into evaluation bins to reduce spatial autocorrelation
362 between points in the testing and training bins. To improve MaxEnt's modelling performances, we
363 performed a calibration procedure by fitting the model under different combinations of parameters and
364 hyperparameters. Specifically, we fitted the model under six feature class combinations (L, LQ, H, LQH,

365 LQHP, and LQHPT, where L is linear, Q is quadratic, H is hinge, P is product, and T is threshold) and eight
366 different regularisation multipliers (0.5 to 4 at 0.5 intervals). While the feature class allows MaxEnt to
367 develop composite models to ensure a good fit to the data, regularisation multiplier values control
368 model overfitting (Kass et al., 2021).

369 Overall, there were 48 models (6 (feature class) x 8 (regularisation multiplier) for each data group (GBIF-
370 only vs. combined data) in each time interval. We chose the best model with the lowest Akaike
371 Information Criterion (AICc) (Kass et al., 2021). Using the raster package in R (Hijmans, 2023). We used
372 the 10% omission rate threshold value and transformed the suitability map into binary classes based on
373 the threshold value (1 for presence with suitability value \geq threshold value; and 0 for absence with
374 suitability value $<$ threshold value). We calculated the centroid position of these binary maps and used
375 the geosphere (Hijmans, 2022) package in R to calculate the range expansion rate.

376

377 **Niche assessment**

378 We used the ecospat R package (Broennimann et al., 2023) to evaluate, for each time period separately,
379 whether the use of additional data from social media led to significant differences in the realised niche
380 space occupied by the species. We quantified the niche overlap between the GBIF-only and combined
381 dataset, for each of the five time periods separately. We used the same seven environmental variables
382 that were used earlier for fitting SDMs. We extracted the environmental data corresponding to the
383 occurrence records and ran the Principal Component Analysis (PCA) to reduce dimensionality. We
384 transformed the first two components (PC1 and PC2) into density by kernel smoothers. The PC1 and PC2
385 explained from 59-63% of the total variance. We quantified niche overlap using Schoener's D metric and
386 assessed the statistical significance through niche equivalency and similarity tests following the methods
387 described by Warren et al. (2008) and Broennimann et al. (2012). These tests were implemented to
388 determine whether the observed niche overlap was greater than expected by chance, providing insight
389 into niche dynamics over time.

390

391 **Data and code availability**

392 We have attached the social media data in the supplementary section (Supplementary Data S1) and the
393 GBIF data are publicly available (GBIF, 2024).

394 All the R scripts are available in the following public GitHub repository:
395 https://github.com/ShawanChowdhury/SocialMedia_RangeChange_TC.

396

397 **Acknowledgments**

398 We are grateful to the numerous volunteers for data collection and sharing their records on GBIF and
399 social media. We collated the species distribution data from GBIF, Flickr and Facebook. The GBIF data
400 (<https://doi.org/10.15468/dl.wq7282>) are publicly available. We have provided the Flickr and Facebook
401 data in the supplementary data (Supplementary Data S1).

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405

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610

611 **List of Figures**

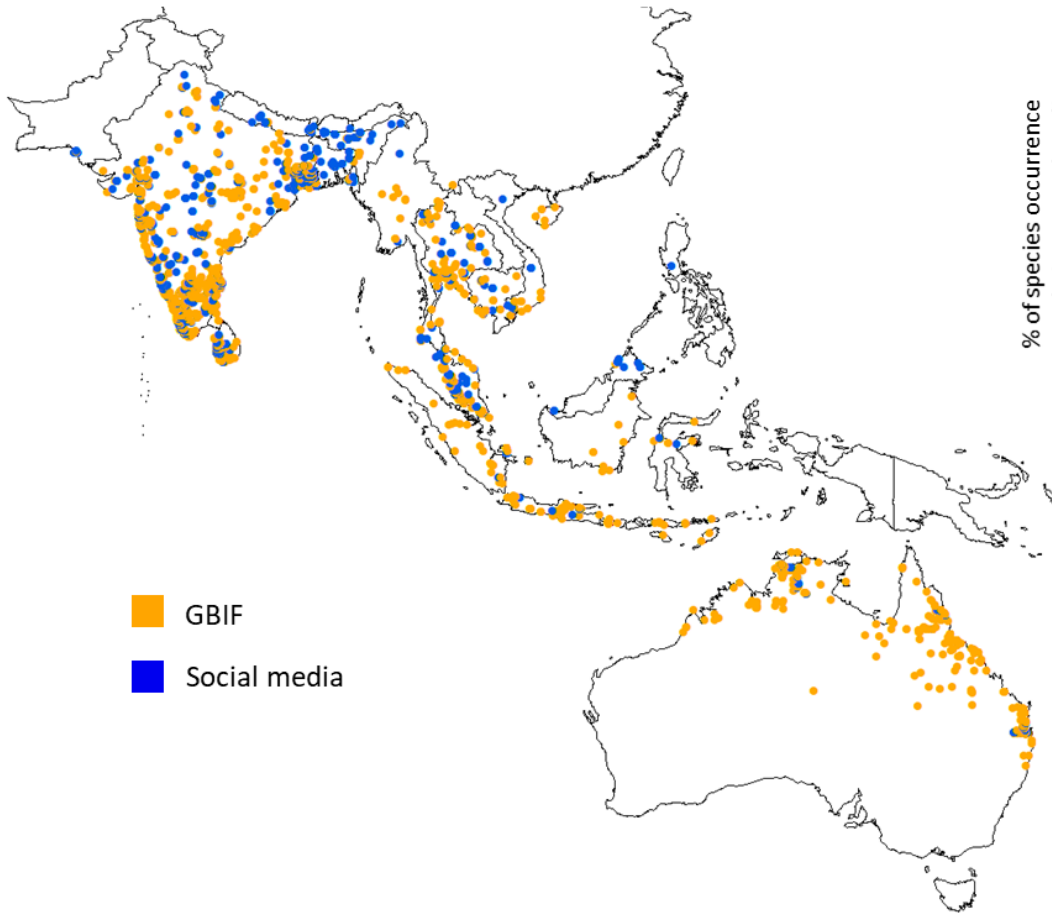
612 **Figure 1.** Distribution of occurrence records by source (GBIF and social media, where social media
613 contains records from both Facebook and Flickr) with (A) the entire known distribution of tawny coster
614 (*Acraea terpsicore*) (each point represents one occurrence record); (B) the association between the total
615 number of occurrence records and the percentages of occurrences records obtained from social media
616 data; and (C) temporal trends in occurrence records by data source (GBIF vs. social media).

617 **Figure 2.** Tracking range shift dynamics of the tawny coster butterfly (*Acraea terpsicore*) during 2005-
618 2024. (A-E) Maps of predicted habitat suitability in five-year intervals, highlighting areas where model
619 predictions altered after adding the social media data (i.e., contribution from social media data in model
620 predictions is highlighted in purple); (F) the differences in the surface area obtained from binarized
621 model predictions; (G) latitudinal and elevational distribution of the suitable areas (for all year intervals);
622 and (H) the differences in estimated range expansion.

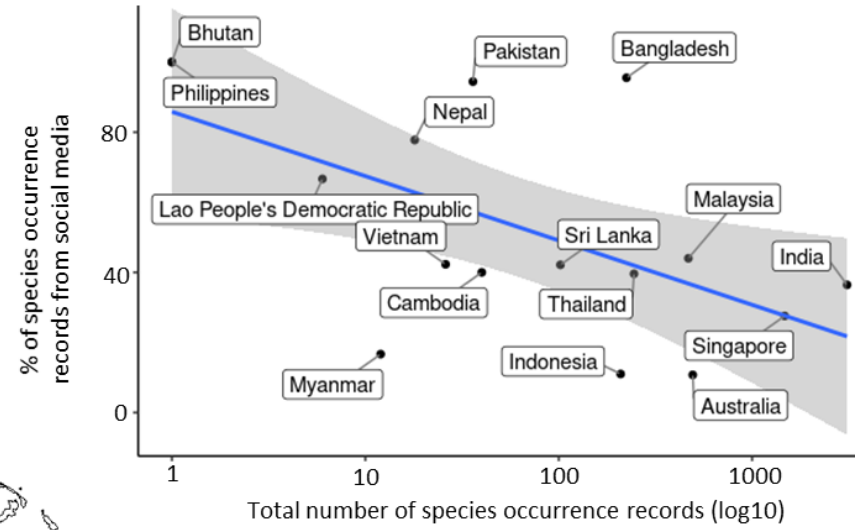
623 **Figure 3.** The differences in identified niche space with adding social media data in different year
624 intervals. The inset figures (correlation plot) show the importance and direction of impact of each
625 predictor variable. Here, def = climate water deficit, ppt = precipitation (monthly total), soil = soil
626 moisture, tmax = maximum temperature, ws = wind speed, PDSI = Palmer Drought Severity Index, and
627 elev = elevation.

628

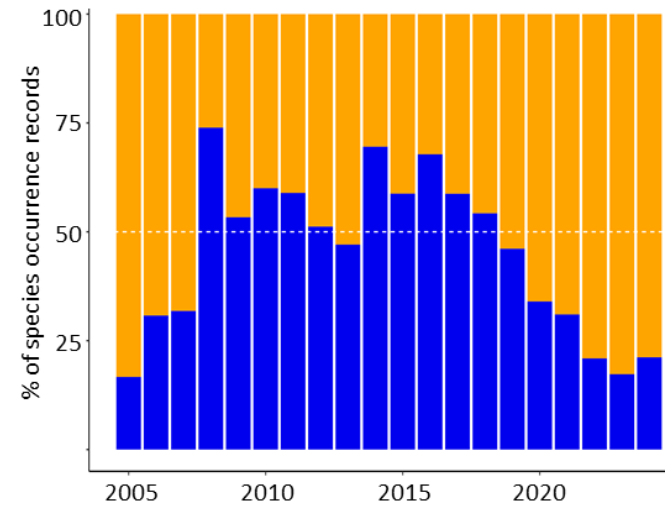
A)

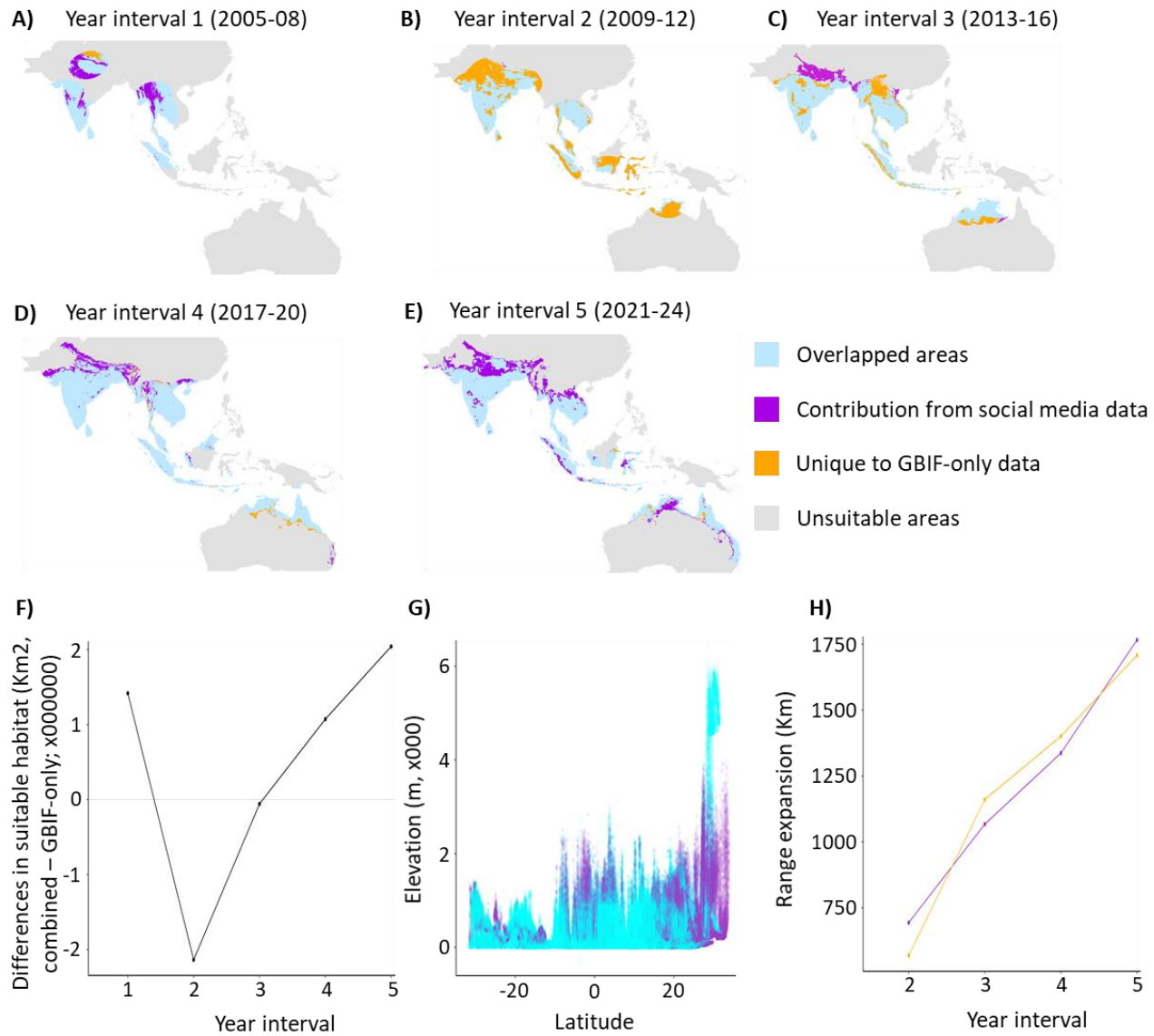


B)

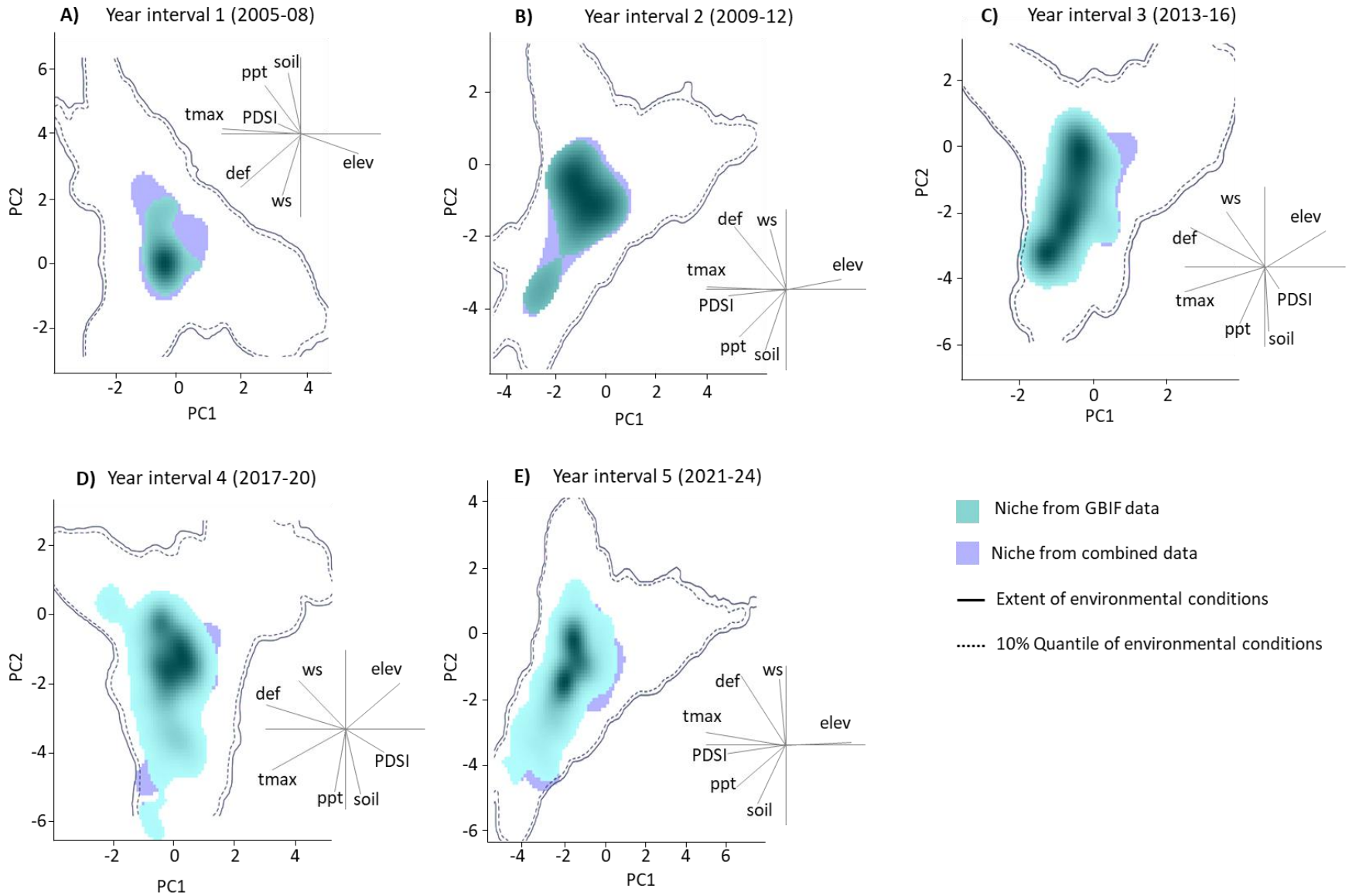


C)





633 **Figure 3.**



Supplementary Table S1: Details of the Facebook group, showing the keywords used, country information, group names, and URL. The country information NA means that the group is not country-specific.

Keywords	Country	Group Name	Group URL
Butterfly Bangladesh	Bangladesh	Butterfly Bangladesh	https://www.facebook.com/groups/butterflybangladesh/
Butterfly India	India	ButterflyIndia	https://www.facebook.com/groups/ButterflyIndia/
Butterfly India	NA	World of Butterflies	https://www.facebook.com/groups/208726699864915/
Butterfly India	India	Indian Butterflies	https://www.facebook.com/groups/indianbutterflies/
Butterflies India	India	Moths and Butterflies of Northeast India	https://www.facebook.com/groups/654169375175548/
Butterflies India	India	Butterflies Of West Bengal	https://www.facebook.com/groups/224547761077777/
Butterflies India	India	Butterflies of North-Eastern India	https://www.facebook.com/groups/butterfliesofnortheastindia/
Butterflies India	India	Insects and butterflies of India	https://www.facebook.com/groups/1609109409365987/
Butterflies India	India	Moths and Butterflies of Northeast India	https://www.facebook.com/groups/654169375175548/?hoisted_section_header_type=recently seen&multi_permalinks=1496482467610897
Butterflies India	India	Butterflies From India	https://www.facebook.com/groups/butterfliesfromindia/
Butterflies India	India	Butterflies of India	https://www.facebook.com/groups/246930848714418/?hoisted_section_header_type=recently seen&multi_permalinks=8239610469446376

Butterfly Australia	Australia	Butterflies, Moths and other Invertebrates of Australia	https://www.facebook.com/groups/170745013686340/
Butterfly Australia	Australia	Australian Butterflies	https://www.facebook.com/groups/1642134862743621/
Butterfly Australia	Australia	Australian butterflies and moths	https://www.facebook.com/groups/799465170167144/
Butterfly India	India	InsectIndia	https://www.facebook.com/groups/insectIndia/
Butterfly India	India	Ask IDs of Indian Butterflies	https://www.facebook.com/groups/275650439625884/
Butterfly India	NA	Australasian Butterflies and Moths	https://www.facebook.com/groups/305768640902911/
Butterfly Pakistan	Pakistan	Butterflies of Pakistan	https://www.facebook.com/groups/131718433700946/
Butterfly Bhutan	Bhutan	Butterflies Society of Bhutan	https://www.facebook.com/groups/2287652334879957/
Butterfly Bhutan	Bhutan	Butterfly and Moths of Bhutan	https://www.facebook.com/groups/bhutanmoths/
Butterfly Nepal	Nepal	Butterfly Conservation Nepal.NP	https://www.facebook.com/groups/himalayanbirder/
Butterfly Nepal	Nepal	Butterflies Diversity of Nepal	https://www.facebook.com/groups/1979680945491783/
Butterfly Nepal	Nepal	Butterflies of Nepal	https://www.facebook.com/groups/butterfliesnepal/
Butterfly China	NA	Butterflying Around the World 寰宇蝶影	https://www.facebook.com/groups/649414648579768/
Butterfly Australia	Australia	South-east Queensland Butterfly Watching	https://www.facebook.com/groups/seqbutterflies/

Butterflies Bangladesh	Bangladesh	Butterflies of Bangladesh	https://www.facebook.com/groups/129902820822634/
Butterflies Bangladesh	Bangladesh	ButterflyBengal	https://www.facebook.com/groups/333424654047804/
Butterfly SriLanka	SriLanka	Butterfly Conservation & Research Group of Sri Lanka	https://www.facebook.com/groups/bcrgsl/
Butterflies Myanmar	Myanmar	Butterflies and Moths of Myanmar	https://www.facebook.com/groups/624495890937290/
Butterflies Thailand	Thailand	Butterflies of Thailand	https://www.facebook.com/groups/365316596902175/?hoisted_section_header_type=recently seen&multi_permalinks=2977081182392357
Butterfly Cambodia	Cambodia	Natural Cambodia	https://www.facebook.com/groups/naturalcambodia/
Butterfly Singapore	Singapore	Butterfly & Macro Singapore	https://www.facebook.com/groups/1885942558332507/
Butterfly Singapore	Singapore	ButterflyCircle (Butterflies of Singapore)	https://www.facebook.com/groups/240038746511844/
Butterfly Singapore	Singapore	Singapore (SG) Insect ID and Records	https://www.facebook.com/groups/sginsectid/
Butterflies Singapore	Singapore	Butterflies of Singapore and Malaysia	https://www.facebook.com/groups/255047171183480/?hoisted_section_header_type=recently seen&multi_permalinks=7776253999062722
Butterfly Taiwan	Taiwan	台灣超微距暨昆蟲攝影交流會 Taiwan Super Macro	https://www.facebook.com/groups/1416552225029363/?hoisted_section_header_type=recently seen&multi_permalinks=25562340933357155

Butterfly Indonesia	Indonesia	Indonesia Photography Family	https://www.facebook.com/groups/950185375659760/?hoisted_section_header_type=recently seen&multi-permalinks=1346355336042760
Butterflies Indonesia	Indonesia	Butterflies of Borneo	https://www.facebook.com/groups/1637283956484389/
Butterflies Philippines	Philippines	Philippine Lepidoptera	https://www.facebook.com/groups/488909304537513/
Butterflies Philippines	Philippines	PARUPAROZZIS: Butterfly Watchers Philippines	https://www.facebook.com/groups/paruparozzi/
Butterflies Papua New Guinea	Papua New Guinea	Butterflies of Papua	https://www.facebook.com/groups/2185597161721882/

Supplementary Figure S2: The 10000 background maps that were randomly generated using all the occurrence records of the tawny coster butterfly.

