

Evidence of the Impacts of Pharmaceuticals on Aquatic Animal Behaviour (EIPAAB): a systematic map and open access database

Jake M. Martin^{1,2,3,4,*}, Marcus Michelangeli^{1,3,5}, Michael G. Bertram^{1,3,4}, Paul J. Blanchfield⁶, Jack A. Brand^{1,7}, Tomas Brodin¹, Bryan W. Brooks⁸, Daniel Cervený^{1,9}, Kate N. Fergusson³, Malgorzata Lagisz¹⁰, Lea M. Lovin^{1,8}, Isaac Y. Ligocki¹¹, Shinichi Nakagawa^{10,12}, Shiho Ozeki³, Natalia Sandoval-Herrera¹, Kendall R. Scarlett^{8,13}, Josefin Sundin¹⁴, Hung Tan^{3,15}, Eli S.J. Thoré^{1,16,17}, Bob B.M. Wong³, Erin S. McCallum^{1,*}

¹Department of Wildlife, Fish and Environmental Studies, Swedish University of Agricultural Sciences, Umeå, Sweden

²School of Life and Environmental Sciences, Deakin University, Waurn Ponds, Australia

³School of Biological Sciences, Monash University, Melbourne, Victoria, Australia

⁴Department of Zoology, Stockholm University, Stockholm, Sweden

⁵School of Environment and Science, Griffith University, Nathan 4111, Australia

⁶Fisheries and Oceans Canada, Freshwater Institute, Winnipeg, Manitoba, R3T 2N6, Canada

⁷Institute of Zoology, Zoological Society of London, London, United Kingdom

⁸Department of Environmental Science, Baylor University, Waco, Texas 76798 USA

⁹University of South Bohemia in Ceske Budejovice, Faculty of Fisheries and Protection of Waters, South Bohemian Research Center of Aquaculture and Biodiversity of Hydrocenoses, Zatisi 728/II, Vodnany, Czech Republic

¹⁰Evolution and Ecology Research Centre, School of Biological, Earth and Environmental Sciences, University of New South Wales, Sydney, NSW, Australia

¹¹Department of Biology, Millersville University, Millersville, Pennsylvania, USA.

¹²Department of Biological Sciences, University of Alberta, CW 405, Biological Sciences Building, Edmonton, AB T6G 2E9, Canada

27 ¹³Environment Protection Agency, EPA Office of Water, Office of Science and Technology

28 ¹⁴Department of Aquatic Resources, Swedish University of Agricultural Sciences,

29 Drottningholm, Sweden

30 ¹⁵Environment Protection Authority Victoria, EPA Science, Macleod, Victoria, Australia

31 ¹⁶TRANSfarm – Science, Engineering, & Technology Group, KU Leuven, Lovenjoel, Belgium

32 ¹⁷Laboratory of Adaptive Biodynamics, Research Unit of Environmental and Evolutionary

33 Biology, Institute of Life, Earth and Environment, University of Namur, Namur, Belgium

34 *corresponding author(s): Jake M Martin (JMM), jake.martin@slu.se and Erin S. McCallum

35 (ESM), erin.mccallum@slu.se

36

37 Jake M. Martin: jake.martin@slu.se, ORCID ID: 0000-0001-9544-9094

38 Marcus Michelangeli (MM): m.michelangeli@griffith.edu.au, ORCID ID: 0000-0002-0053-6759

39 Michael G. Bertram (MGB): michael.bertram@slu.se, ORCID ID: 0000-0001-5320-8444

40 Paul J. Blanchfield (PJB): paul.blanchfield@dfo-mpo.gc.ca, ORCID ID: 0000-0003-0886-5642

41 Jack A. Brand (JAB): jackbrand.work@gmail.com, ORCID ID: 0000-0003-3312-941X

42 Tomas Brodin (TB): tomas.brodin@slu.se ORCID ID: 0000-0003-1086-7567

43 Bryan W. Brooks (BWB): Bryan_Brooks@baylor.edu ORCID ID: 0000-0002-6277-9852

44 Daniel Cervený (DC): daniel.cervený@slu.se ORCID ID: 0000-0003-1491-309X

45 Kate N. Fergusson (KNF): kate.fergusson@monash.edu, ORCID ID: 0009-0008-1043-1871

46 Malgorzata Lagisz (ML): m.lagisz@unsw.edu.au ORCID ID: 0000-0002-3993-6127

47 Lea M. Lovin (LML): lea.lovin@slu.se, ORCID ID: 0000-0002-1056-2045

48 Isaac Y. Ligocki (IYL): Isaac.Ligocki@millersville.edu, ORCID ID: 0000-0002-2014-0479

49 Shinichi Nakagawa (SN): s.nakagawa@unsw.edu.au, ORCID ID: 0000-0002-7765-5182

50 Shiho Ozeki (SO): shiho.ozeki@monash.edu, ORCID ID: 0000-0002-4480-8891

51 Natalia Sandoval-Herrera (NSH): natalia.sandoval.herrera@slu.se ORCID ID: 0000-0002-2546-8983

52 Kendall R. Scarlett (KRS): scarlett.kendall@epa.gov ORCID ID: 0009-0005-4852-3614

53 Josefin Sundin (JS): josefin@teamsundin.se. ORCID ID: 0000-0003-1853-4046

54 Hung Tan (HT): Hung.Tan@monash.edu ORCID ID: 0000-0002-7500-8395

55 Eli Thoré (ESJT): eli.thore@slu.se, ORCID ID: 0000-0002-0029-8404

56 Bob B.M. Wong (BBMW): bob.wong@monash.edu ORCID ID: 0000-0001-9352-6500

57 Erin S. McCallum (ESM): erin.mccallum@slu.se ORCID ID: 0000-0001-5426-9652

58

59 **Keywords**

60 ecotoxicology, evidence synthesis, fitness, medicine, neurotoxicology, psychoactive

61 Abstract

62 Background

63 The global proliferation of pharmaceutical pollutants in aquatic ecosystems has emerged as
64 a pressing environmental concern. These contaminants—designed to modulate biological
65 functions at minute dosages—pose a unique threat to aquatic organisms, particularly through
66 behavioural alterations. Recent years have seen a surge in scientific interest in the use of
67 behavioural endpoints in chemical risk assessment and regulatory activities, underscoring
68 their importance for fitness and survival. In parallel, research on how pollution, particularly
69 pharmaceuticals, alters the behaviour of aquatic animals appears to have grown rapidly.
70 Despite this, there has been a notable absence of systematic efforts to consolidate and
71 summarise this field of study. To address this gap, our objectives are twofold: first, to
72 systematically identify, catalogue, and synthesise primary research articles on the effects
73 of pharmaceuticals on aquatic animal behaviour; and second, to organise the ‘Evidence of
74 the Impacts of Pharmaceuticals on Aquatic Animal Behaviour’ (EIPAAB) into a comprehensive
75 open-access database for scientists, policymakers, and environmental managers.

76

77 Methods

78 We systematically searched two electronic databases (Web of Science and Scopus) and
79 supplemented these with additional article sources. The search string followed a Population-
80 Exposure-Comparison-Outcome (PECO) framework to capture articles that used an aquatic
81 organism (population) to test the effects of a pharmaceutical (exposure) on behaviour
82 (outcome). Eligible articles also needed a control group (comparison). Articles were
83 screened in two stages: an initial screening of title and abstract, followed by full-text
84 screening alongside data extraction. Decision trees were designed *a priori* to appraise
85 eligibility at both stages. Information on study validity was collected but not used as a basis
86 for inclusion.

87

88 **Review findings**

89 We identified and screened 5,988 articles, of which 901 were included in the final EIPAAB
90 database, representing 1,739 species-by-compound combinations. The database includes
91 data collected over 48 years (1974-2022), with most articles having an environmental focus
92 (510) and fewer relating to medical and basic research topics (233 and 158, respectively).
93 The EIPAAB database includes 173 distinct species representing 8 phyla and 21 classes. Ray-
94 finned fishes were by far the most common clade (75% of the evidence base). The database
95 also includes 426 distinct pharmaceutical compounds; the most frequently investigated
96 groups were antidepressants (28%), antiepileptics (11%), and anxiolytics (10%). The impacts
97 of pharmaceuticals on locomotion and boldness/anxiety behaviours were most assessed out
98 of the 10 broad behavioural categories assigned in the database (62 sub-categories in total).
99 Generally, we detected poor reporting and/or compliance with several of our study validity
100 criteria, including the use of experimental blinding, randomisation, reporting of compound
101 details, and experimental treatment concentration verification.

102

103 **Conclusions:**

104 Our systematic map revealed a rapid increase in this research area over the past 15 years.
105 We highlight multiple areas now suitable for quantitative synthesis and areas where
106 evidence is lacking. We also highlight some obvious pitfalls in method reporting and
107 practice. More detailed reporting would facilitate the use of behavioural endpoints in
108 aquatic toxicology studies, chemical risk assessment, regulatory management activities, and
109 improve the overall replicability of this research area. The EIPAAB database can be used as
110 a tool for closing these knowledge and methodological gaps in the future.

111

112 Background

113 Pharmaceutical residues are ubiquitous in the environment and have been detected globally
114 on every continent [1,2]. Pharmaceuticals present a particular concern for aquatic animals,
115 with the discharge of human, veterinary, and livestock wastewater effluents being a primary
116 source of contamination. These contaminants can also enter the environment during
117 pharmaceutical manufacturing, through landfill leachates, and run-off from biosolids used
118 in agriculture [2,3]. Aquatic animals exposed to pharmaceuticals can directly or indirectly
119 bioconcentrate these compounds in their tissues [4,5]. Understandably, there are now
120 growing calls for the effective management of pharmaceutical pollution in aquatic
121 environments [6,7]. Yet, for many pharmaceuticals, empirical sublethal ecotoxicological
122 information is lacking, precluding robust ecological risk assessments for aquatic animals[8].
123 Where ecotoxicity data are available, they are often limited to standard toxicological
124 endpoints (i.e. morphometric endpoints), such as growth, reproductive output, and
125 mortality (reviewed in [8]). It is essential to consider that the effects of pharmaceutical
126 exposure on aquatic animals are likely to be subtle, given that pharmaceuticals are typically
127 detected at low concentrations (low ng/L - low µg/L), are specifically designed to have low-
128 dose effects in their target organisms, and many drug targets are conserved across
129 vertebrate taxa [9]. However, this does not discount adverse environmental impacts, as
130 wildlife may experience unintended, therapeutic-like or human side effects from
131 pharmaceutical exposure [10-12]. Consequently, a growing body of research is investigating
132 adverse outcomes of pharmaceutical exposure, specifically sub-lethal effects on processes
133 like endocrine signalling, development, bioenergetics, and behaviour (reviewed in [13-16]).

134 In recent years, behaviour has emerged as a key endpoint of interest for emerging
135 chemicals of environmental concern, including human pharmaceuticals and veterinary
136 medicines [13,17,18]. This is because behaviour is a tractable endpoint, as it is a particularly
137 sensitive indicator for measuring contaminant-induced effects on non-target species,
138 especially when compared to standard ecotoxicological endpoints [19,20]. Behaviour can

139 also bridge the gap among proximate, sub-organismal, individual-level processes, to
140 ultimate, ecologically relevant, population-level outcomes, which are important for
141 environmental protection goals [16,21]. However, behaviour is rarely used in a regulatory
142 context [17,18,22]. Recent recommendations have highlighted that integrating behavioural
143 endpoints with other adverse outcomes or standard endpoints (e.g. survival, growth) and
144 improving the reliability of behavioural studies will help improve the quality of scientific
145 contributions and utility in regulatory settings [17,22].

146 Alongside the increasing use of behavioural endpoints in ecotoxicology, there has
147 been growing awareness that pharmaceuticals specifically designed to modify behaviour are
148 present in the aquatic environment and the tissues of aquatic animals (e.g. antidepressants,
149 anxiolytics, antipsychotics [23-27]). Indeed, many pharmaceuticals are specifically designed
150 to alter behaviour as their primary therapeutic effect (e.g. antidepressants, anxiolytics,
151 antipsychotics), whereas others may inadvertently lead to behavioural changes (e.g.
152 analgesics, hormone therapies) [8,13]Widespread environmental contamination with
153 behaviour-modifying drugs, together with increased recognition of behaviour as a sensitive
154 endpoint for ecotoxicology, has culminated in an exponential growth of research focused on
155 the behavioural effects of a multitude of pharmaceuticals on aquatic organisms (e.g. [28-
156 32]). For this rapidly expanding field, it is now essential that we synthesise the data being
157 produced and identify focus areas, knowledge gaps, and opportunities for future research.

158 Here, we have conducted systematic mapping to identify, categorise, and visualise
159 research detailing the effects of pharmaceuticals on the behaviour of aquatic animals.
160 Systematic Evidence Maps (SEMs) help to identify research trends, show knowledge gaps
161 where further primary research is needed, and specify areas with enough data for targeted
162 evidence synthesis approaches (i.e. systematic review, meta-analysis) [33,34]. Importantly,
163 SEMs have recently been identified as an underutilised tool for chemical risk assessment and
164 decision-making because they can provide a comprehensive summary of literature relevant
165 for future policy while also minimising bias [35]. SEMs are especially valuable for connecting

166 heterogeneous interdisciplinary data, like those used in ecotoxicology and chemical risk
167 assessments, which are beyond the scope, and/or expertise of any one scientist [36].
168 Therefore, given the rapid expansion of behavioural ecotoxicology and growing interest in
169 behavioural endpoints for chemical risk assessment and management, a SEM is a timely
170 approach for understanding the behavioural effects of pharmaceuticals on aquatic animals.

171

172 Objective of the review

173 Primary objective

174 We aimed to identify and catalogue evidence on the effects of human and veterinary
175 pharmaceuticals on aquatic organism behaviour and present this evidence in an open-access
176 database. Our SEM has the following elements:

177

178 *Population:* Any aquatic animal that is a metazoan with at least one obligate aquatic phase
179 of its life (e.g. fish, amphibia, aquatic mammal, aquatic invertebrate).

180

181 *Exposure:* A human or veterinary pharmaceutical compound.

182

183 *Comparator:* A control (i.e. unexposed) or solvent control group of animals.

184

185 *Outcome:* A behavioural trait. We define behaviour as organismal kinematic responses, or
186 lack of kinematic responses (e.g. freezing, bursting), to an internal or external stimulus
187 (e.g. foraging in response to hunger [internal] or food [external] stimuli).

188

189 Secondary questions

190 In addition, our SEM addressed two secondary questions.

191 (1) Identify knowledge gaps, research priorities, and areas of research that have sufficient
192 data for further synthesis.

193

194 (2) Collate information on additional endpoints (e.g. sub-organismal, reproduction, growth,
195 survival) measured alongside behaviour in each article to facilitate connections across
196 domains that may be useful for future aquatic toxicology studies and environmental
197 management activities.

198

199 **Methods**

200 The reporting of the methodology followed MeRIT to improve author contributions'
201 granularity and accountability [37]. This systematic map is based on the methods described
202 in the previously published protocol [38]. It follows the reporting standards for Systematic
203 Evidence Syntheses in environmental research (ROSES [39]; see Supplementary File 1). Our
204 SEM has also been pre-registered using the Open Science Framework (OSF) online platform,
205 and the registration is freely available at: <https://doi.org/10.17605/OSF.IO/7N92E>.

206

207 **Deviations from the protocol**

208 Several deviations from the original published protocol [38] were made. These deviations
209 are summarised as follows:

- 210 1. The planned bibliometric analyses and the screening of academic theses were not
211 conducted because of changes to the initial search string during the protocol peer-
212 review process. This resulted in an increase in the total number of search returns
213 and, so too, the total amount of screening effort required for the project. The
214 additional workload meant that some elements of the project had to be reduced or
215 removed.
- 216 2. In the protocol, full-text screening was to be performed in duplicate. This was also
217 changed as a result of the increased number of search returns (i.e. 1,239 articles
218 underwent full-text screening). Instead, 10% of all articles at the full-text screening

219 stage ($n = 127$) underwent duplicate screening to estimate the consistency of
 220 eligibility decisions and meta-data extraction of the final EIPAAB database (see
 221 ‘Article screening and eligibility criteria’ and Table 1 for details). In addition, every
 222 article that was excluded at the full-text screening stage was subsequently cross-
 223 screened (i.e. subsequently screened in duplicate).

224 3. Some questions in the online full-text and data extraction form were removed and/or
 225 altered to decrease extraction workload and increase replicability. All changes were
 226 made before the full-text screening and data extraction began. The changes made
 227 to the data collection form are detailed in Table S1.

228 4. New authors were recruited to the project, and two original authors withdrew from
 229 the project (JTO and GCM). The new authors included were: SO, KNF, LML, KRS,
 230 ESJT, and NSH.

231

232 **Table 1.** Eligibility criteria associated question element (i.e. PECO element or other criteria
 233 such as language) and the screening stage at which it applies, title and abstract, full text
 234 or both.

235

Eligibility criteria	Question element	Screening stage
Uses an aquatic animal. <i>Animal that have at least one phase of their life as obligate aquatic (e.g. fish, amphibia, aquatic mammal, aquatic invertebrate)</i>	Population (P)	Both
Uses a wild type animal <i>An animal that is not genetically modified</i>	Population (P)	Full text
Uses at least one pharmaceutical compound <i>A decision tree will be used to assist screeners in deciding whether a compound qualifies as a pharmaceutical compound (Figure S1)</i>	Exposure (E)	Both
Has a control group	Comparator (C)	Both

A non-exposed group to which the exposed group is compared and is therefore not a review, meta-analysis, conference proceeding etc

Measures behaviour	Outcome (O)	Both
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An organism's kinematic response, or lack of kinematic response (e.g. freezing, resting), to an internal or external stimulus (e.g. foraging in response to hunger [internal] or food [external] stimuli)

Is in a language in which our review team is proficient: <i>English, Swedish, Norwegian, Czech, Slovak, Japanese, Polish, Russian</i>	Language	Both
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236

237 **Search for articles**

238 Search terms and strings

239 ESM and JMM designed the search string with assistance from ML for WoS and Scopus to
240 reflect our PECO framework. The aquatic organism search terms (i.e. population terms)
241 captured broad taxonomic groups for animals that have at least one phase of their life as
242 obligate aquatic (e.g. fish, amphibia, aquatic mammal, aquatic invertebrate), in addition
243 to the common aquatic model species or any species used in Organization for Economic
244 Cooperation and Development (OECD) Toxicity Testing Guidelines (e.g. guppy, medaka,
245 minnow, cladocerans; both common and genus names). Pharmaceutical compound terms
246 (i.e. exposure terms) included general synonyms for medications and specific
247 pharmaceutical classes (e.g. antidepressants, analgesics). Exposure environment terms
248 covered aspects of the experimental environment and the process of exposing animals to a
249 pharmaceutical (e.g. exposure, treatment, tank). Behaviour terms (i.e. outcome terms)
250 included variants of behaviours that could be measured in aquatic animals (e.g. movement,
251 cognition). No search terms were included addressing the comparator (i.e. a control group)
252 as these terms were unlikely to appear in bibliometric records. We instead covered this in
253 our screening process and eligibility criteria. The full search strings used in both WoS and
254 Scopus are reported in Table S2. The search strings were applied to all keywords, titles, and
255 abstracts in both databases. The searches in WoS and Scopus were initially performed on 17

256 November 2021 and were subsequently updated on 13 February 2024 to include the rest of
257 articles published 2021 and all of 2022.

258

259 Search filters

260 No filters for language or document type were used in WoS and Scopus. However, only
261 languages with which the co-authors are proficient were included (English, Swedish,
262 Norwegian, Czech, Slovak, Japanese, Polish, Russian). No limit was placed on publication
263 year (except up until 2022).

264

265 Search sources

266 Our map targeted experimental research articles (i.e. no reviews or meta-analyses). We
267 targeted this type of article because we wanted to build a database of articles where a
268 controlled pharmaceutical exposure has been conducted. We searched for articles in two
269 broad-coverage online databases: WoS (Core Collection) and Scopus.

270

271 Supplementary searches

272 We supplemented the database searches in two ways: First, we conducted reference
273 searches of key review articles published on the behavioural effects of pharmaceuticals in
274 aquatic animals. For this, JMM and ESM *a priori* selected six reviews, that focused on the
275 impacts of pharmaceuticals on aquatic organism behaviour (provided in [38]). Second, ESM
276 and the co-author team advertised on social media platforms and mailing lists (e.g. “X” and
277 the Society of Environmental Toxicology and Chemistry Pharmaceuticals Interest Group) that
278 we were seeking articles on this topic (including any well-documented reports from grey
279 literature). Any articles submitted were sent via a simple Google Form to collect basic
280 article information. We did not expect a large grey literature outside of academic or

281 government scientific research sources because aquatic environmental risk assessments
282 conducted for the approval of new pharmaceuticals do not include animal behaviour as an
283 endpoint (discussed in [8,17]).

284

285 Estimating comprehensiveness of the search

286 The details of how we estimated search comprehensiveness and sensitivity are detailed in
287 the published protocol [38]. Briefly, we tested the sensitivity using 83 benchmark articles
288 that were expected to be captured by the search string. Our search string recovered 95% of
289 the benchmark articles.

290

291 De-duplication of results

292 Search returns from WoS, Scopus, and the additional sources were combined and duplicates
293 were removed in Mendeley Desktop Software (Mendeley Ltd.) before being imported to
294 Rayyan [40], a software designed for article screening. Any remaining duplicates were
295 identified in Rayyan and removed before starting title and abstract screening.

296 **Article screening and study-eligibility criteria**

297 Articles were included at the title and abstract screening stage based on five eligibility
298 criteria (listed in Table 1). Title and abstract screening was performed using Rayyan, and
299 was completed in duplicate by two independent reviewers randomly assigned to each article
300 (12,094 total screenings [including duplicates]; JMM 27%, ESM 27%, KNF 12%, JS 12%, JAB
301 12%, DC 12%, IYL 12%, HT 12%, MM 12%, JTO 12%*, LML 12%, MGB 12%, SO 11%, KRS 11%, GCM
302 9%*; **left the project after title and abstract screening*). Both reviewers had to agree for
303 the article to be included before moving to the full-text screening and data extraction stage.
304 A list of all title and abstract screening decisions and reasons for exclusion are reported in
305 Supplementary File 2. The full-text screening was completed using Qualtrics Survey Software

306 (Qualtrics, Provo, UT) alongside data extraction. The inclusion decision at the full-text
307 screening stage was based on six eligibility criteria (listed in Table 1). Full-text screening
308 and data extraction were randomly assigned to screeners (1381 total screenings; JMM 10%,
309 ESM 8%, NSH 8%, ESJT 8%, MM 7%, JAB 7%, KNF 7%, LML 7% SO 7%, DC 6%, IYL 6%, KRS 6%, HT
310 6%, JS 6%, MGB 3%, ML <1%), as described above, a subset of full-text screening and data-
311 extraction was performed in duplicate (10%, $n = 127$ selected at random). This subset of
312 duplicate screened articles was used for consistency checks to estimate article inclusion
313 decision alignment. For the 127 articles screened in duplicate, there were 18 disagreements,
314 predominantly resulting from issues assessing the compound eligibility (see Supplementary
315 File 3 for a list of disagreements). In total, 10% of all duplicate-screened articles were
316 excluded incorrectly, while 4% were included incorrectly. As a result of a higher-than-
317 desired false exclusion rate, all articles that had been designated as 'excluded' were
318 subsequently cross-screened (by JMM and ESM). After cross-screening, 10% of articles that
319 were initially 'excluded', were subsequently changed to 'include' (38 of 373). Due to the
320 large number of articles considered in the systematic map, it was not feasible to cross-check
321 all 'included' articles at the full-text stage. Thus, we acknowledge a possible 4% false
322 inclusion rate in the project, which would result in approximately 50 articles being
323 incorrectly included in the final database. We highlight that the broader trends and field-
324 related insights gained from the EIPAAB database are likely robust to this small number of
325 false inclusions. Articles that were allocated as 'discuss' under the eligibility question
326 (indicating extractor uncertainty) were also cross-screened, and a final inclusion/exclusion
327 decision was made (by JMM). A list of all articles excluded at the full-text screening stage
328 and the reason for exclusion is reported in Supplementary File 4. For both screening stages,
329 screeners were not assigned articles in which they were listed as authors.

330

331 Study-validity assessment

332 We collected information on study validity from all included articles during data extraction;
333 however, articles were not excluded from the SEM based on any validity criteria. We
334 collected information on study validity guided by the Criteria for Reporting and Evaluating
335 Ecotoxicity Data (CRED [41]), extracting information directly relating to 10 of the 20 CRED
336 reliability criteria. Specifically, we extracted information relating to Criteria 1 (“Is a
337 guideline method [or modified guideline] used”), Criteria 2 (“Is the test performed under
338 GLP conditions”), Criteria 3 (“[A]re validity criteria fulfilled [control survival, growth]”),
339 Criteria 5 (“Is the test substance identified with name or CAS number...”), Criteria 6 (“Is the
340 purity of the test substance reported...”), Criteria 8 (“Are the organisms well described...”),
341 Criteria 9 (“Are the test organisms from a trustworthy source...”), Criteria 11 (“Is the
342 experimental system appropriate for the test organism...”), Criteria 14 (“Is the exposure
343 duration defined”), Criteria 15 (“Are chemical analyses adequate to verify concentrations
344 of the test substance...”). For a list of which metadata corresponded to each of the CRED
345 criteria, and details on why some of the criteria were not considered, see Table S3 (also
346 detailed in Supplementary File 6 ‘READ-ME’). In addition, we collected the following study
347 validity data not specific to ecotoxicity data: (1) whether animals were randomly assigned
348 to treatment groups, (2) whether behaviour was scored blind to treatment, (3) how
349 behaviour was scored (e.g. manual *versus* automated), (4) if any conflicts of interest were
350 stated.

351 In total, we had 19 metadata questions relating to study validity (detailed in
352 Supplementary File 6 ‘READ-ME’); we documented study validity via the CRED reliability
353 guidance and the above additional questions for three reasons. First, behavioural studies in
354 ecotoxicology have been criticised [42,43] for not following standardised methods or for
355 providing too little data for use in risk assessment procedures. These study validity
356 descriptors will allow us to identify common methodological gaps being overlooked by
357 scientists conducting behaviour-focused studies (e.g. not reporting CAS identifiers, not

358 reporting water quality parameters). Second, scoring behaviour blind to treatment is a
359 standard protocol in behavioural ecology to reduce experimental bias; however, this method
360 may be less prominent for researchers outside of behavioural ecology. Thus, we wanted to
361 identify the number of articles taking this key methodological consideration into account.
362 Third, we included study validity descriptors to improve the utility of the EIPAAB database
363 for future users.

364 **Data-coding strategy**

365 Data extraction protocol

366 All articles were assigned a numeric ‘article ID’ that identified the article throughout the
367 title and abstract screening, full-text screening, and the data extraction process. For full-
368 text screening and data extraction, the screening team was assigned a list of articles which
369 contained the article ID, article title, year of publication, journal, and authors (as a CSV
370 file). The screeners used this document to search for and download the articles. The data
371 extraction was coded using an online form (Qualtrics Survey Software; designed by ESM and
372 JMM with input from all co-authors). Before the allocation of full-text articles, all screeners
373 were first trained using a pilot screening with 10 randomly selected articles. This was done
374 to clarify uncertainty for extractors, and to test the efficacy and functionality of the full-
375 text screening and data collection form (as reported in [38]). The article metadata were
376 extracted in the following survey sections (full survey structure supplied in Supplementary
377 File 5):

- 378 1. Details about the screener and article: information on the screener and the article
379 being extracted (e.g. screener initials, article ID, DOI).
- 380 2. Inclusion criteria: data on the inclusion criteria (see Table 1). If the reviewer chooses
381 to exclude the article, they skip the remaining data extraction.
- 382 3. Study species: data on the aquatic organism(s) studied (e.g. species name, animal
383 source, sex, life stage).

- 384 4. Pharmaceutical compound(s): data on the pharmaceutical compound(s) being
385 studied and the exposure environment (e.g. compound name, route of exposure,
386 dosage, exposure duration).
- 387 5. Behavioural endpoints: data on which behaviours were measured. Behaviours are
388 first categorised into 10 broad categories (e.g. movement/activity, aggression,
389 foraging, boldness; see Table S4 for full list) and then into more specific
390 subcategories (2-12 per parent category; 62 total), to extract more detail on how
391 the behaviour was measured (e.g. within movement/activity: normal locomotor
392 activity, abnormal movements, dispersal/migration; see Table S4 for full list and
393 definitions).
- 394 6. Connecting across biological scales: data on whether the article also measured any
395 sub-organismal traits (e.g. hormone concentrations, mRNA transcription) and/or
396 endpoints capturing growth, reproduction, or survival. We included these questions
397 to increase the utility of the EIPAAB database.
- 398 7. Validity: data describing the study validity (see ‘Study validity assessment’ for
399 further details).
- 400 8. Research motivation: the primary scientific motivation of the article was allocated
401 to environmental (i.e. focus on predicting/measuring the effects of environmental
402 pollution on wildlife; ecotoxicology), medical (focus on improving human or
403 veterinary medical practice), or basic research (focus on understanding biological
404 phenomena or methodological development with no overt applicational claims for
405 medical or ecotoxicological purposes).

406 Data processing

407 The data collected by the online survey form were downloaded as CSV files and imported
408 into R (version 4.2.3, in the R studio environment, Build 463; [44]) for data processing (by
409 JMM). Errors with DOI and ‘article ID’ (i.e. unique project allocated IDs) were identified by
410 cross-referencing titles, DOIs, and article IDs with the article allocation list given to

411 extractors. The database was then re-shaped to a long format, where each article was given
412 a row for each tested chemical and each tested species, in other words, a row for each
413 unique species-by-compound combination. Compound names and species names were then
414 assessed for possible synonyms or typographical errors. For compounds, this was done by
415 searching compound names in the PubChem database [45], and collating PubChem CID,
416 PubChem name, CAS, and synonyms (Python script by JMM is provided on Github;
417 <https://github.com/JakeMartinResearch>). These identifier metadata were then used to
418 evaluate possible synonyms or typographical errors in the database (e.g. different compound
419 names that shared a CAS number). For species, this was done using the National Centre for
420 Biotechnology Information (NCBI) Taxonomy database [46]), with each species name
421 searched, and the taxonomy ID, current taxonomic name, and full lineage collated; these
422 species metadata were used to evaluate possible synonyms or typographical errors in the
423 database. For articles that had multiple species, the compound and behaviour data were
424 cross-checked to make sure that the answers given by extractors applied to all species, if
425 they did not, they were adjusted. This was necessary as the survey form did not allow
426 extractors to give separate answers for different species within the same article. All survey
427 questions with an 'Other' option to provide a free-text based alternate response (e.g. study
428 motivation, behavioural classification, methods used to score behaviour; see survey form
429 linked as Supplementary File 5) were then assessed by JMM and, where appropriate, were
430 re-assigned to existing categories or were grouped into new categories (see Table S4-5 list
431 of new categories).

432 Consistency estimates

433 In total, there were 84 duplicate screened articles included, which represented 305 rows of
434 data (i.e. each unique species-by-compound combination). To estimate the consistency of
435 metadata extraction, JMM calculated the alignment between each survey question within
436 each unique species-by-compound combination. When the answer from extractors matched

437 exactly, the data were assigned a '1', if it did not match they were assigned a '0'. The
438 median consistency across all metadata was $94.8\% \pm 8.8\%$, ranging from 60.8-100% (a list of
439 consistency for all metadata is reported in Table S6). Data that were implicitly consistent
440 (e.g. article ID, DOI, species name, compound name) or not consistent (e.g. screener name),
441 were not included in estimates of the median consistency. As a result of some of the specific
442 behavioural classifications having low consistency (median 95.8%, range 67.6-99.3%; see
443 Table S6), a Boolean value (1 or 0) for categorisation only at the broadest level of the
444 behavioural class was created, which had higher consistency (median 98.6%, range 75.6-
445 99.3%; see Table S6). The reason for low consistency for some of the metadata extraction
446 is discussed below in the limitations section. We have opted to maintain all metadata in the
447 database regardless of estimated extraction consistency, but we suggest that those using
448 the EIPAAB database check the level of consistency for the metadata they plan to use, and
449 decide whether it is appropriate for their individual usage.

450 Additional metadata to increase usability

451 To aid in cross-article comparison and to increase the usability of the database, the
452 following additional information was added to the EIPAAB database:

- 453 1. Standardised concentrations were added to the database which converted the
454 original concentration units reported by the authors to one of six standardised units
455 (original units and values were also maintained). Specifically, the following
456 conversions were made: mass/volume measures to $\mu\text{g/L}$, volume/volume measures
457 to $\mu\text{L/L}$, mass/mass measures to $\mu\text{g/g}$, mole units to μM , molarity (mole/volume)
458 units to $\mu\text{M/L}$, and dimensionless units of concentration to ppm.
- 459 2. Compounds were assigned to a therapeutic classification system, specifically the
460 Anatomical Therapeutic Chemical (ATC) classification tree (hereafter ATC; [47]). The
461 ATC classifies active ingredients of drugs according to the organ or system on which
462 they act and their therapeutic, pharmacological, and chemical properties. The ATC

463 classification was selected as it is widely used, covers many compounds in the EIPAAB
464 database (305 of 426 compounds), and has a simple classification structure. For
465 compounds that returned multiple ATC classification trees, the trees were collated.
466 ATCs were pulled from PubChem by JMM, by searching each compound name,
467 extracting the resulting PubChem substance ID (up to 150), and searching
468 classification information for each SID (Python scripts by JMM are provided on Github;
469 <https://github.com/JakeMartinResearch>). In addition to the full classification tree
470 (as a semicolon-separated list), the classifications are also provided at each level of
471 the tree separately (e.g. 5 ATC classification levels) to make the data more
472 accessible (see Supplementary File 6 for details).

473 3. Additional species metadata were added to the EIPAAB database from the
474 International Union for Conservation of Nature's (IUCN) Red List of Threatened
475 Species [48]. Specifically, JMM and MRM searched each species name in the IUCN Red
476 list, and for those with an associated IUCN Red List report, the IUCN report DOI, IUCN
477 Status, IUCN report publication year, geographic range, population trend, habitat
478 type, and movement patterns were collated (see Supplementary File 6 for details of
479 each data type).

480 4. Additional bibliometric metadata from WoS and Scopus were collected by JMM
481 (05/07/2024), using a search of the full DOIs list across both online databases ($n =$
482 894), or by searching the title if the article did not have a DOI ($n = 7$). A total of 879
483 articles were located on WoS, and the extracted metadata included: journal
484 abbreviation (ISO), author keywords, unique WoS ID, WoS Categories, WoS Research
485 Areas, number of cited references, and number of times the article was cited (across
486 all databases). A total of 888 articles were located on Scopus, and the extracted
487 metadata included: journal abbreviation, author keywords, Scopus EID, and number
488 of times the article was cited.

489 Data-mapping method

490 We summarise the available research at three levels: (1) the article level, represented as
491 'article_id' in the database; (2) the population level, represented as 'unique_population_id'
492 (i.e. article id + species name); and (3) the species-by-compound level, represented as
493 'unique_row_id' in the EIPAAB database (article id + species name + compound name). The
494 level at which our summaries were made depended on the level at which those metadata
495 were extracted and/or applied to the article. For example, metadata like the publication
496 year, conflict statements, and water quality were extracted and summarised at the article
497 level ($n = 901$). Metadata like species life stage, sex, and source were extracted and
498 summarised at the population level (i.e. unique_population_id; $n = 935$), because a single
499 article can have multiple species. Metadata like exposure duration, exposure concentration,
500 and category of behaviours measured were extracted and summarised at the species-by-
501 compound level (i.e. unique_row_id; $n = 1,739$) because in cases where multiple species
502 were used, different exposures and behaviours can be, and were, assessed. The level at
503 which metadata were extracted is listed within Supplementary File 6, and how this was
504 applied to summarise the data is illustrated in Supplementary File 7 (i.e. R script). We also
505 performed many of our summaries with respect to the motivation for the study. During
506 metadata extraction, we categorised each article based on its primary motivation, as either
507 environmental (i.e. focus on predicting/measuring the effects of environmental pollution
508 on wildlife; ecotoxicology), medical (focus on improving human or veterinary medical
509 practice), or basic research (focus on understanding biological phenomena or
510 methodological development with no overt applicational claims for medical or
511 ecotoxicological purposes). We did so because we predicted the motivation of the research
512 to strongly influence many aspects of the study design, such that some of our summary data
513 would be insightful only if applied within a given study motivation. For example, we would
514 expect the applied doses to be very different in an environmentally motivated study
515 compared to a medically motivated study. All data summary methods are explained in detail

516 in Supplementary File 7, which is also designed to act as a starting point for anyone who
517 wishes to use the EIPAAB database for their own projects.

518

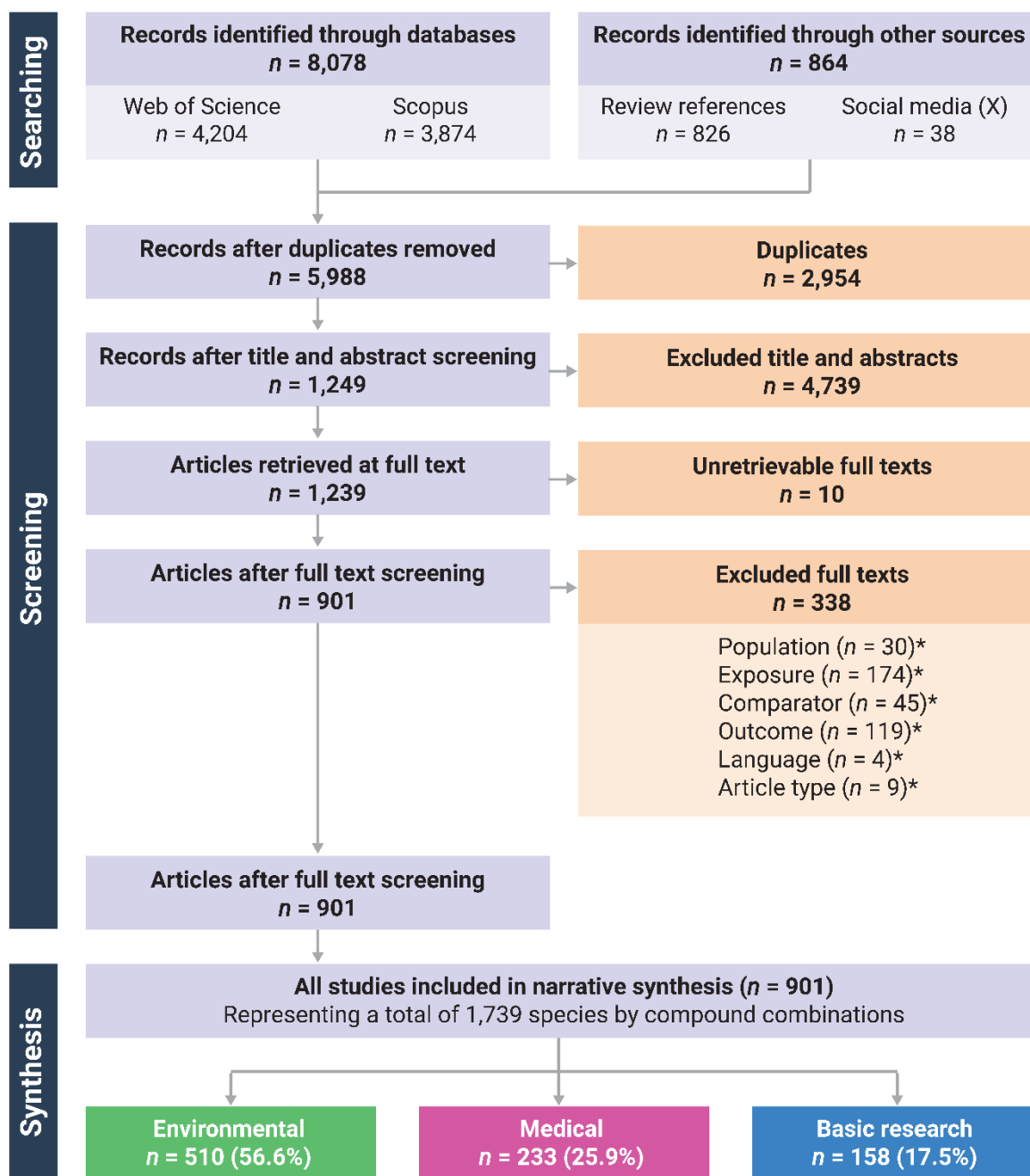
519 Results

520

521 Overview of the evidence base and temporal trends

522 In total, 901 articles—representing 1,739 unique species-by-compound combinations—were
523 included in the final EIPAAB database. After collating articles from all sources and de-
524 duplication, we screened a total of 5,988 unique articles for possible inclusion in the
525 systematic map and database (Fig 1). In brief, 4,739 articles were excluded after title and
526 abstract screening, 338 articles were excluded during full-text screening and data
527 extraction, and 10 articles were unretrievable for full-text screening (overall inclusion rate
528 of 21%; Fig 1). Most articles were excluded at the full-text screening stage for not having a
529 compound of interest (i.e. exposure: $n = 174$; Fig 1) or for not measuring a behaviour (i.e.
530 outcome: $n = 119$; Fig 1).

ROSES Flow Diagram for Systematic Reviews



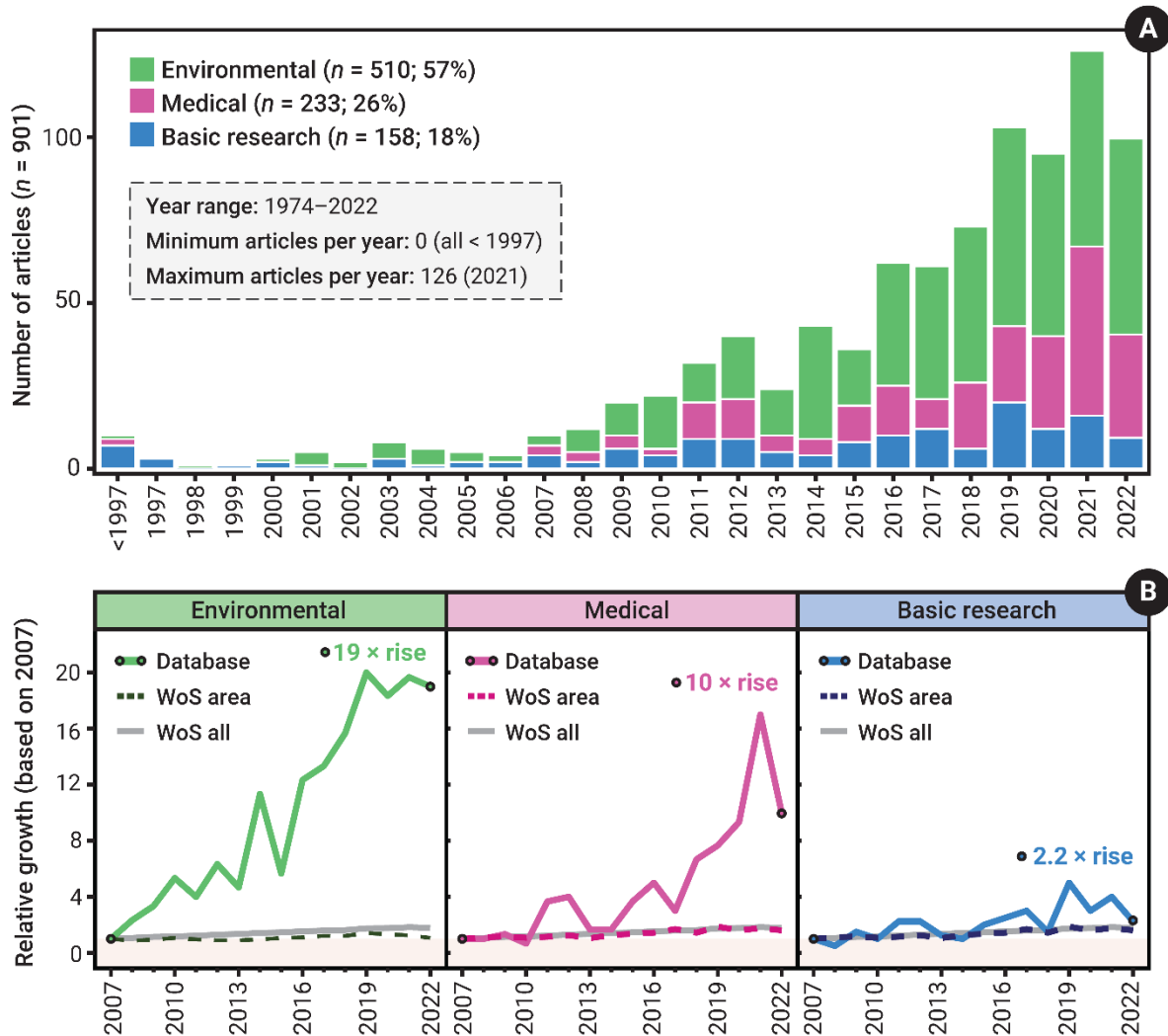
531

532 **Figure 1.** Flow diagram for the SMAP and EIPAAB database, showing the article numbers at
 533 each step of the process (i.e. searching, screening, and synthesis). This figure is based on
 534 the Reporting Standards for Systematic Evidence Syntheses (ROSES) flow diagram for
 535 systematic reviews, version 1.0 [49]. *The total number of articles for each full-text
 536 exclusion criterion includes multiple reasons allocated to a single article; we also expected

537 *that when articles failed to meet multiple exclusion criteria, screeners may not have*
538 *indicated every reason for exclusion (e.g. if the article was the wrong article type).*

539

540 Regarding study motivation, 510 articles had an environmental motivation (56.6%),
541 233 had a medical motivation (25.9%), and 158 had a basic research motivation (17.5%). The
542 included articles date from 1974-2022, with a steep rise in the number of articles around
543 2007 (Fig 2A). To specifically assess the growth of research on pharmaceutical impacts on
544 animal behaviour, we compared the relative increase in articles over the last 15 years in
545 the systematic map (2007-2022), against that of the most common WoS Research Area, as
546 well as all researcher areas in the WoS Core Collection (i.e. an overall publication trend).
547 This was done for each study motivation separately (see Supplementary File 6 for full details
548 and Supplementary File 8 for the search results). For articles allocated to the environmental
549 study motivation, the most common WoS Research area was ‘Environmental Sciences &
550 Ecology’ (65% fall within this research area); for those allocated to medical and basic
551 research, it was ‘Neurosciences & Neurology’ (47% and 39% fall within this research area,
552 respectively). The growth rate of research articles addressing the impacts of pharmaceutical
553 impacts on animal behaviour with an environmental focus far outpaces that of the broader
554 research area of ‘Environmental Sciences & Ecology’ and the overall publication trend from
555 2007-2022 (Fig 2B). The growth in research with a medical focus also outpaced the broader
556 research area of ‘Neurosciences & Neurology’ and overall publication trends, but this was
557 only evident from 2018-2022 (Fig 2B). The growth in research with a basic research focus
558 did not consistently deviate from the broader research area of ‘Neurosciences & Neurology’
559 or overall publication trends (Fig 2B).



560

561 **Figure 2.** (A) The total number of articles included in the EIPAAB database by publication
 562 year (articles published before 1997 were grouped; total range 1972-2022). Study motivation
 563 is represented by the stacked colours within the bar chart (Environmental = green, Medical
 564 = pink, Basic research = blue, stacked in that order). (B) The relative growth in the number
 565 of articles per year from 2007-2022 based on 2007, as compared to the respective WoS
 566 Research area (WoS area: 'Environmental Sciences & Ecology' or 'Neurosciences &
 567 Neurology'), and WoS global publication trends (WoS all), for Environmental, Medial and
 568 Basic research articles in the database.

569

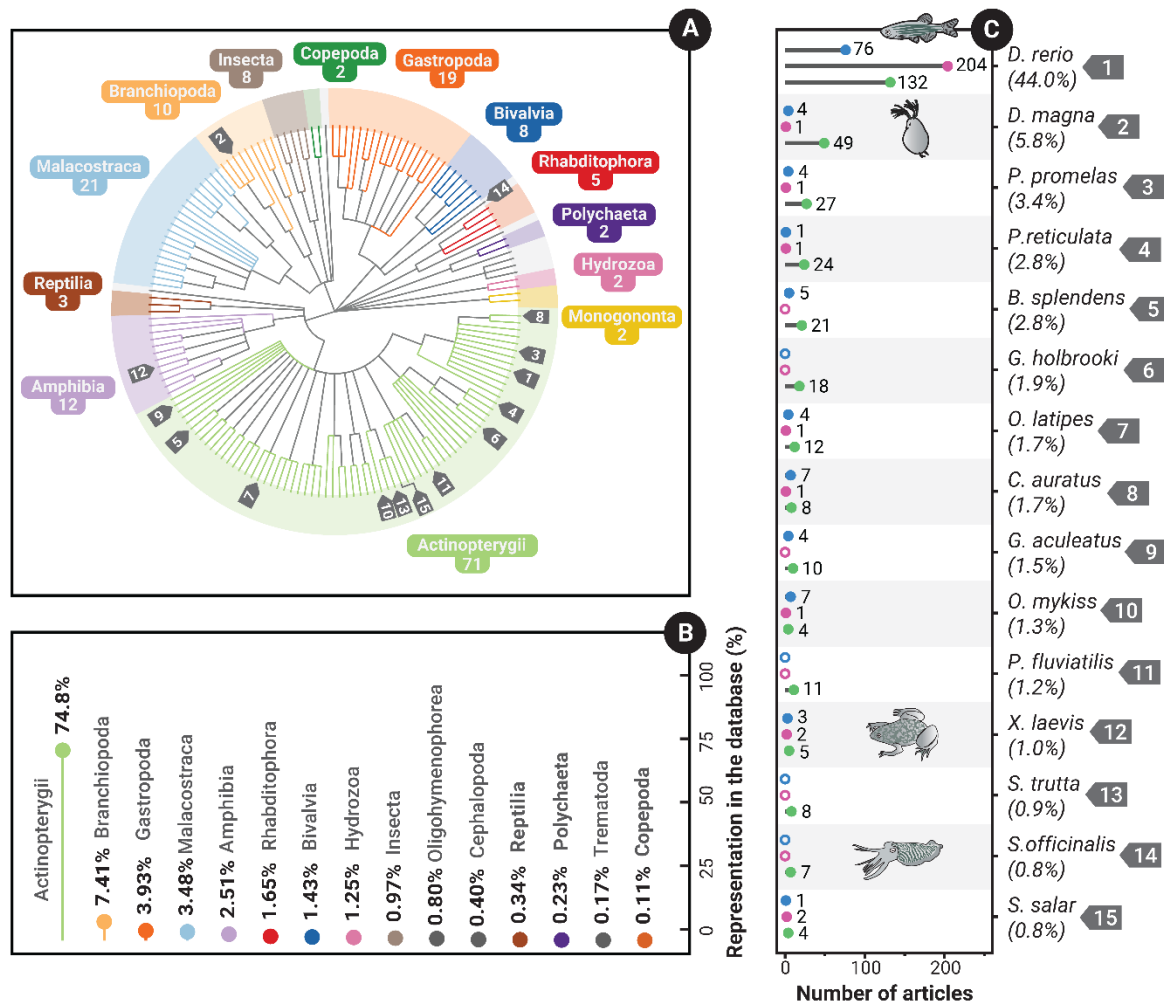
570 Mapping characteristics of the population, exposure, and outcomes

571 Study species (population)

572 Collectively, the database includes 173 different species from 21 classes (Fig 3A). In terms
573 of taxonomic diversity, 41.0% of the species present in the evidence database belonged to
574 the clades *Actinopterygii* (i.e. ray-finned fishes), 12.1% to *Malacostraca* (i.e. soft-shelled
575 crustaceans), 11.0% to *Gastropoda* (i.e. gastropods), 6.9% to *Amphibia* (i.e. amphibians),
576 and 5.8% to *Branchiopoda* (e.g. fairy shrimp, water fleas)—all other clades represent less
577 than 5% of the total distinct species (Fig 3A). Regarding the representation in the evidence
578 base (i.e. how often they were studied), *Actinopterygii* was by far the most common,
579 representing 75.4% of all data in the database; all other clades represented less than 10% of
580 the data included in the database (Fig 3B). The most common species in the database was
581 the zebrafish, *Danio rerio*, being included in 44.1% of all articles, which is almost a factor
582 of 10 higher than the next most common species, *Daphnia magna* (5.8%; the top 15 most
583 common species shown in Fig 3C). Interestingly, many species were only used in a single
584 article (103/173), with very few being used in more than 5 articles (17 species; Fig S2).

585 Taxonomic usage and representation also differed by study motivation; compared to
586 medical articles, those with an environmental and basic research motivation showed a more
587 even spread of taxa, although all had a very strong skew towards ray-finned fishes (Fig S3;
588 Figure 3C). Considering the total number of articles identified per study motivation,
589 environmental and basic research included substantially more species than medical research
590 (Environmental = 143:510; Medical = 26:233, Basic research = 43:158, species:articles).

591



592

593 **Figure 3.** (A) Cladogram showing all species included in the EIPAAB database. All classes
 594 with more than one species are shown in distinct colours (those with a single species are
 595 light grey). The numbered labels 1-15 represent each of the top 15 species represented in
 596 panel C. (B) The 15 most common taxonomic classes in the evidence database. The colours
 597 are unique to each phylum and apply across both plots A and B. (C) The 15 most common
 598 species used in articles within the evidence database. The percentage value given under the
 599 species name is the percentage of total articles, and the counts within the plot are the
 600 number of articles for each species by study motivation (Basic research = blue, Medical =
 601 pink, Environmental = green, in that order). The open circles are cases of zero articles. The
 602 accompanying species images indicate the first occurrence of a distinct taxonomic class in
 603 the top 15 species (i.e. Actinopterygii [1st], Branchiopoda [2nd], Amphibia [12th], and
 604 Cephalopoda [14th])

605

606 There was an overrepresentation for species from freshwater habitats compared to marine
607 (80.4% *versus* 19.6%), although this was less obvious in environmental and basic research
608 (Table S7). There was also an overrepresentation for studying animals at the adult life stage
609 (53.3%), compared to juveniles (14.8%), larvae (26.4%) and embryos/eggs (5.5%). This was
610 broadly consistent across all study motivations, although environmental articles had a more
611 balanced representation of life stages (Table S7). The use of female and male animals, when
612 reported, was roughly equal (44.9% *versus* 55.1%), and this was constant across all study
613 motivations (Table S7). Overall, the most common source of study animal was commercial
614 suppliers/fish farms (38.0%), followed by a lab stock with undisclosed origin (26.6%),
615 collection from the wild (24.4%), lab stock from a commercial supplier (6.9%), and lab stock
616 from a wild population (4.1%). The animal source did, however, vary by study motivation,
617 with environmental articles having the highest representation of wild-collected animals and
618 less sourced from commercial supplies or fish farms (Table S7).

619 Importantly, sex, life stage, or animal source were not obtained from all articles. In
620 some cases, these data were not reported at all, or were not reported in sufficient detail
621 (summary data reported below). The reporting of species-related metadata was considered
622 an aspect of study validity/quality and is discussed in more detail below. With that said, the
623 number of species with missing metadata is also important in interpreting the overall
624 population trends, so this information has been included in the summary table (Table S7).
625 IUCN data was also not available for all species (106 of 173 had IUCN data), which should be
626 considered when interpreting species IUCN red list metadata and habitat data.

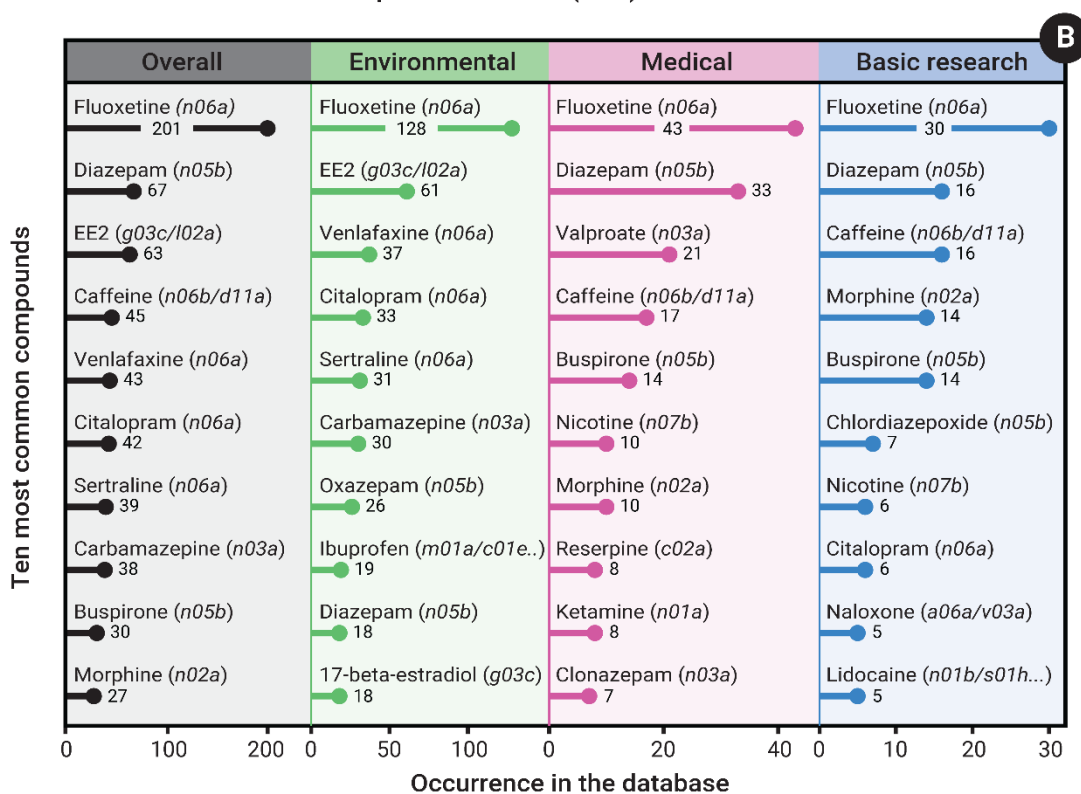
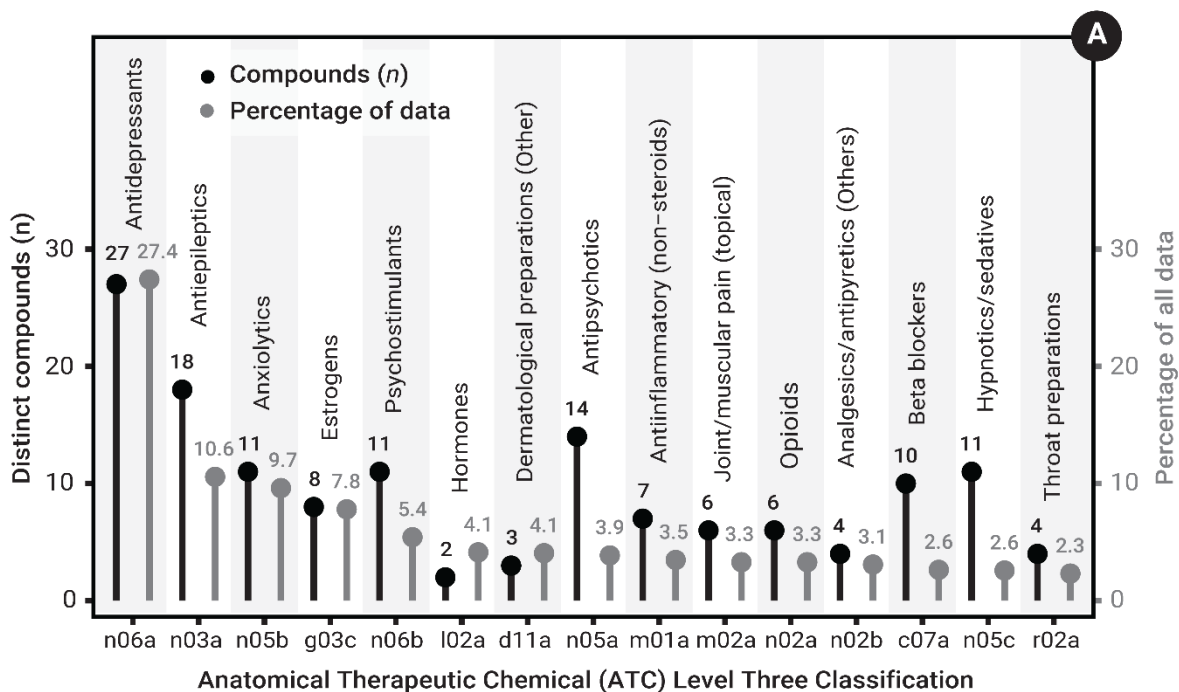
627

628 Pharmaceutical compounds and exposure (exposure)

629 Overall, 426 different pharmaceutical compounds were included in the evidence database.
630 The majority of articles used a single compound ($n = 624$, 69.3%), and very few used more
631 than 5 ($n = 38$, 3.9%), with a similar trend in the number of compounds used across study

632 motivations (Fig S4). We present the compound data in two ways, in terms of the diversity
633 of compounds (irrespective of the number of articles studying them in the EIPAAB database),
634 and their percentage overall representation in the EIPAAB database. In terms of compound
635 diversity—using the WHO Anatomical Therapeutic Chemical (ATC) classification tree—the
636 database includes compounds from all pharmacological groups at the broadest ATC level (14
637 groups). At this ATC level (i.e. 1st ATC level), the pharmacological group with the most
638 compounds was ‘nervous system’, with 43% of all classified compounds belonging to this
639 group, followed by ‘cardiovascular system’ and ‘alimentary tract and metabolism’ (Fig S5).
640 At the 3rd ATC classification level, antidepressants, antiepileptics, and antipsychotics have
641 the highest number of compounds, at 27, 18, and 11 distinct compounds, respectively (Fig
642 4A). In terms of overall percentage representation in the EIPAAB database, compounds
643 within the ATC level one group ‘nervous system’ made up 71.9% of all data, followed by
644 ‘genito urinary system and sex hormones’ (13.5%) and ‘cardiovascular system’ (10.6%). At
645 the 3rd ATC level, antidepressants (27.4%), antiepileptics (10.6%), and anxiolytics (9.7%)
646 were the most common (Fig 4A). Overall, the most common compound was fluoxetine
647 (antidepressants), which made up 11.5% of all data in the EIPAAB database (see Fig 4B for
648 the top 10 most common compounds). There were obvious differences in compound use
649 based on study motivation (Fig 4B). For example, 17- α -ethinylestradiol (EE2) was the
650 third most common compound overall (63 occurrences), but this was almost entirely driven
651 by environmental research (61 occurrences; Fig 4B). Medical and basic research shared a
652 more similar preference for compounds than they did for environmental research (Fig 4B).
653 It is important to highlight that not all articles had an assigned ATC classification (307 of
654 428 had an ATC classification; 72%); thus, all summaries based on ATC do not include all
655 available compounds within the database.

656



657

658

659 **Figure 4.** (A) The 15 most common level three ATC pharmacological groups, as shown by
 660 the number of distinct compounds within each group (black), and overall percentage of
 661 occurrence in the EIPAAAB database (grey). The x-axis lists the group's ATC code, while a
 662 simplified version of the ATC name is given inside the plot. Note that the total percentage

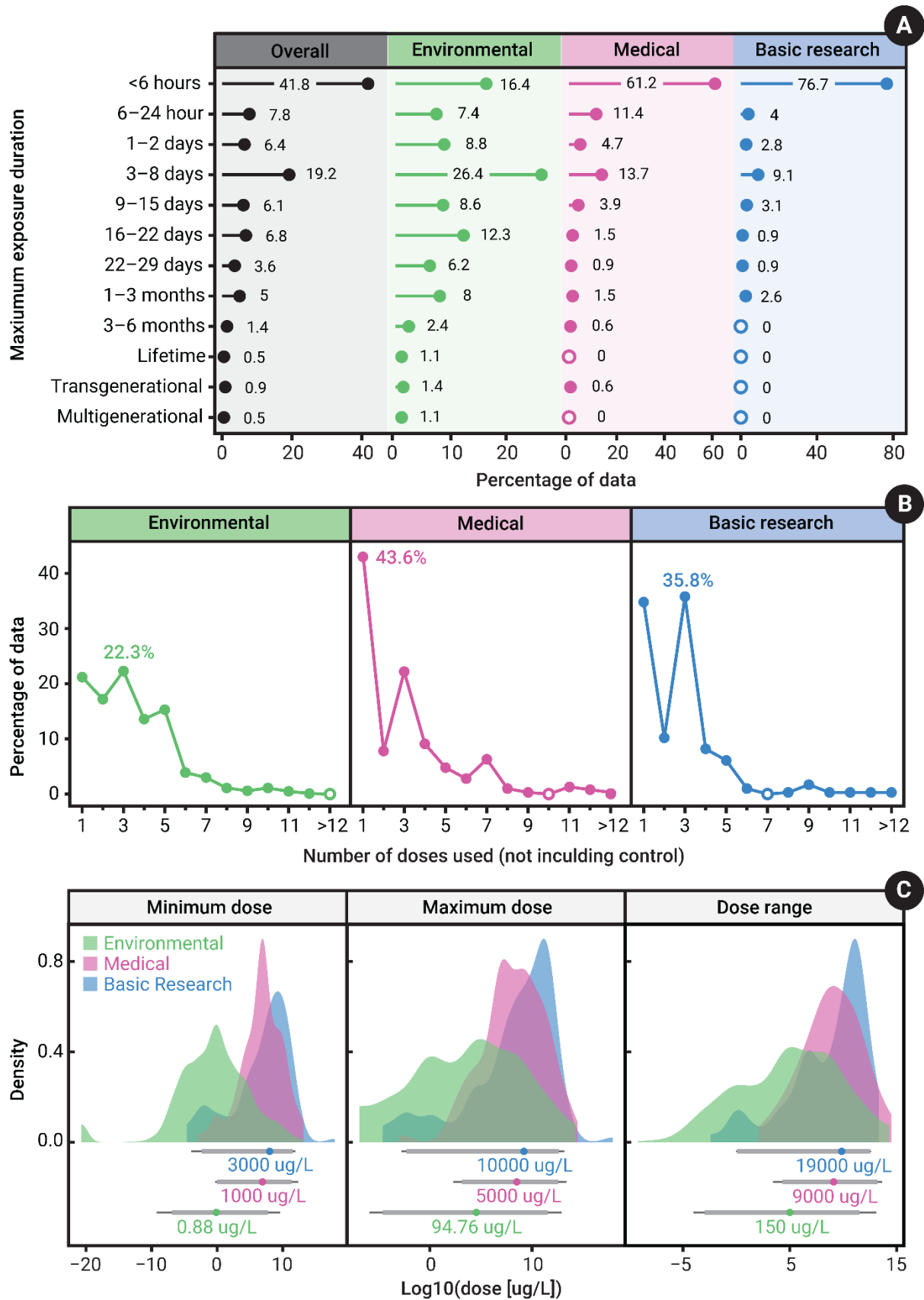
663 may exceed 100, as each compound may have multiple classifications. (B) The 10 most
664 common compounds in the database overall and for each study motivation (Environmental,
665 Medical, and Basic Research), the code in brackets following the compound name are the
666 level three ATC pharmacological groups associated with the compound.

667

668 Overall, 22.6% of articles included mixture exposures in addition to single compound
669 exposure. The use of mixture exposures differed substantially by study motivation.
670 Specifically, medical articles had a much higher rate of mixture exposure (48.4%) compared
671 to basic (25.4%) and particularly environmental research (12.8%). This is likely a result of
672 medical-based articles investigating potential treatments for various
673 psychological/neurological conditions (e.g. epilepsy), in which a phenotype for the
674 psychological/neurological condition of interest is induced using a compound exposure and
675 another compound is subsequently administered to alleviate the phenotype. Most exposures
676 were solely waterborne (86.8%), as compared to other exposure routes (e.g. injection,
677 dietary; 12.7%) or a combined exposure with multiple routes (0.9%). Exposure duration was
678 most often acute (i.e. < 96 hours), with very few studies using exposure durations over a
679 month (only 8.3%; Fig 6A). However, there were notable differences between the study
680 motivations. Medical and basic research articles typically employed exposures less than 6
681 hours (61.2 and 76.7%, respectively), and almost never over 3 months (0.6% and 0%,
682 respectively; Fig 5A). On the other hand, environmental articles had more variation in the
683 maximum exposure durations, with the most common being between 3-8 days (26.4%) and
684 more examples of exposures exceeding 3 months (6%; Fig 5A). Further, overall, most studies
685 exposed animals to a single dose of the compound (29.7%), and very few used more than 5
686 doses (only 9.8%; Fig 5B). For environmental research, there was a more even spread in the
687 percentage of articles that included up to 5 doses (15.3-22.3%; Fig 5B). Broadly speaking,
688 the concentrations used varied substantially, both within and across study motivation (Fig
689 5C). Generally, environmental studies used much lower concentrations (both the minimum

690 and maximum dose) and had a smaller within-study dose range (Fig 5C). Basic research
691 studies used the highest concentrations and had the highest within-study dose range (Fig
692 5C). With that said, there was still substantial overlap in the concentrations used between
693 study motivations, which could help facilitate across-discipline comparisons (although this
694 should be checked explicitly at the compound level). Almost all exposures were conducted
695 in indoor laboratory settings (99.4%) *versus* in a semi-controlled outdoor environment (0.3%)
696 or in the wild (0.2%).

697



698

699 **Figure 5.** (A) The duration of exposures used by articles in the database. The plot is split by
 700 the overall percentage breakdown and those for each study motivation. The percentage

701 values are calculated within each study motivation. (B) The number of different doses used
702 (excluding the control), as shown by study motivation. The percentage values are calculated
703 within each study motivation. (C) The distribution of minimum and maximum dose used, as
704 well as the within-study dose range (i.e. maximum - minimum). The x-axis (dose $\mu\text{g/L}$) is
705 plotted on a log₁₀ scale for the density plots and 'eye plots'. The eye plot shows the median,
706 89, and 95% intervals. The text with the eye plot shows the raw (untransformed) median
707 value and is used to aid in comparisons across study motivations.

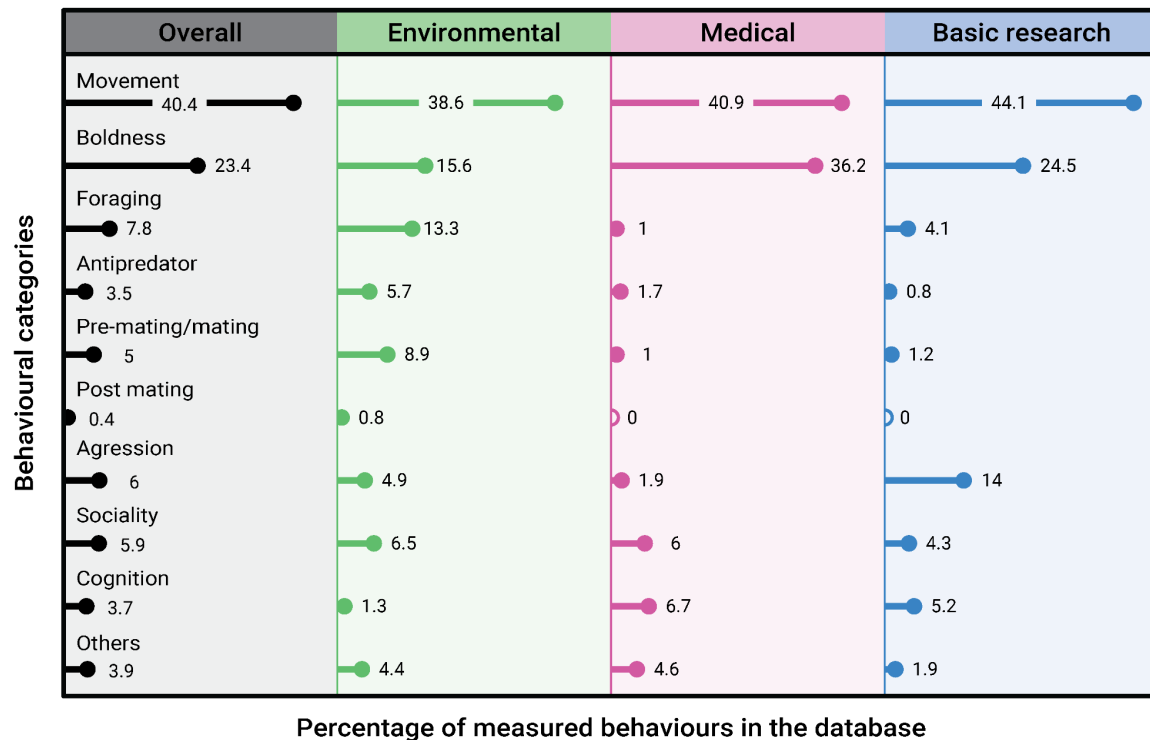
708

709 Behavioural endpoints (outcome)

710 We classified behaviour into 10 overarching categories and 62 sub-categories (2-12 sub-
711 categories within each parent category; a full list of sub-categories and descriptions is given
712 in Table S4). The 10 over-arching categories were: (1) movement and locomotion, (2)
713 anxiety and boldness, (3) foraging/feeding, (4) antipredator behaviour, (5) pre-mating and
714 mating behaviour, (6) post-mating behaviour, (7) aggression, (8) sociality, (9)
715 cognition/learning, and (10) other behaviours not categorised (see Table S4 for list).
716 Typically, only one of these behavioural categories was assessed following exposure (69.3%),
717 with few cases assessing more than 3 behavioural categories after exposure (7.8%); this
718 trend was seen within all study motivations. Overall, movement and locomotion behaviours
719 were the most common responses measured (40.4% of all recorded behaviours), followed by
720 boldness and anxiety-related behaviours (23.4%); all other overarching behavioural
721 categories each represented less than 10% of the data. The preference for
722 movement/locomotion and boldness/anxiety-related behaviours was present in all study
723 motivations, the preference for testing the other 7 categories was more variable (Fig 6).
724 Environmental research had a more even spread of research across the 10 behavioural
725 categories (Fig 6). Overall, the behavioural groups that have seen the least research
726 attention are post-mating behaviours (e.g. parental care; <1%), antipredator behaviours
727 (3.5%), and cognition and learning (3.7%). Within this manuscript, we will not detail the

728 specific breakdown of each behaviour sub-category, but this information is provided for
 729 each study motivation in Fig S6.

730



731

732 **Figure 6.** The percentage measurement of different behavioural categories. The plot is split
 733 into the overall percentage breakdown and those for each study motivation. For a list of all
 734 sub-categories of behaviours and definitions see Table S4 and Figure S7.

735

736 Almost all behaviours were assessed in a laboratory setting (99%), with less than 1%
 737 of measured behaviour being conducted in an outdoor natural setting (in an open natural
 738 setting or restricted natural setting). This almost complete preference for studies in
 739 laboratory settings was present regardless of study motivation (98.7%, 99.6%, 99.7%,
 740 environmental, medical, and basic research, respectively). Overall, only 22% of behavioural
 741 measures were conducted within a social context; in other words, behaviour was rarely
 742 tested in a setting in which multiple animals were able to interact freely. Automated
 743 behavioural scoring was the most common method for measuring behaviour (e.g. tools like
 744 Ethovision, ViewPoint, IDTracker), with 38.9% of articles using an automated quantification

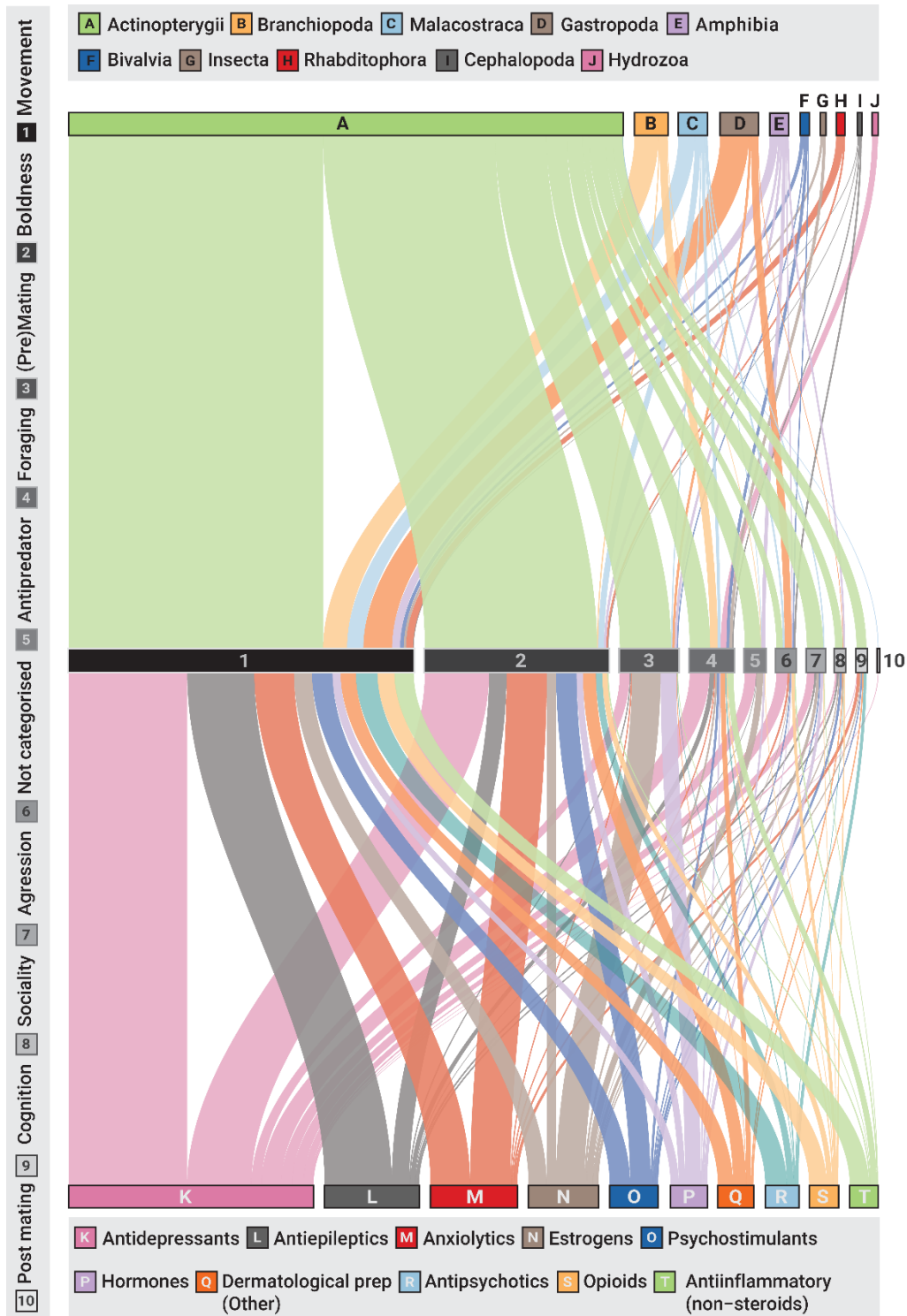
745 approach, 26.6% manually scoring behaviours from recordings, 21% using an indirect method
746 of counting food consumption (e.g. counting food items remaining), and 8.6% used live
747 scoring (all other methods were used in less than 1% of articles). It is important to highlight
748 that 22.7% of articles ($n = 221$) did not clearly specify the methods used to measure
749 behaviour; the information was considered as one of our validity indicators, and is also
750 presented below in the validity assessment.

751

752 Connecting population, exposure, and outcome (PEO)

753 Considering our population, exposure, and outcome elements (i.e. compounds, species, and
754 behaviours) in combination, we found that most articles addressed the effect of a single
755 pharmaceutical compound on a single species and measured a single behavioural category
756 (41.5% of all articles). The next most common study design was a single pharmaceutical
757 compound, a single species, and two behavioural categories (i.e. 17.7%), all other possible
758 combinations each made up less than 10% of the articles. As a broad overview of the
759 connections between compounds, species, and behaviours and how they varied, we
760 illustrate below the links between the 10 most common phylogenetic clades (class) and each
761 behavioural category, as well as the 10 most common therapeutic groups (ATC level 3; Fig
762 8). Broadly speaking, for most of the top 10 clades, movement and locomotion are the most
763 frequently measured behaviours, although there are clade-specific differences in the
764 remaining behaviour categories. For example, Actinopterygii has a relatively high
765 contribution to boldness behaviours, while Branchiopoda, Gastropoda, and Bivalvia are
766 seldom used in the investigation of boldness-related behaviour (Fig 7; see link in figure
767 caption for an interactive version of the figure). There is even more variation in selected
768 behavioural endpoints when looking at therapeutic groups. For example, antidepressants
769 (ATC n06a), anxiolytics (ATC n03a), and psychostimulants (ATC n06b) have high relative
770 contributions to measured boldness-related behaviour, while estrogens (ATC g03c) and
771 hormones (l02a) have a high relative contribution to measured pre-mating/mating behaviour

772 (Fig 7). In the supplementary material, we further illustrate the variability in the
 773 relationship between compound, species, and behaviour using fluoxetine, diazepam, and
 774 17-alpha-ethinylestradiol (the three most common compounds) as specific examples (see
 775 Fig S7).
 776



777

778 **Figure 7.** A broad overview of the link between population, exposure and outcome
779 elements. The Sankey plot shows the connection between all behavioural categories
780 (numbered 1-10; represented by the boxes in the middle of the plot), the top 10 most
781 common phylogenetic clades (Class; shown at the top of the plot), and the top 10
782 therapeutic groups (ATC level 3; shown at the bottom of the plot). The thickness of each
783 band that connects the population to behaviour, or exposure to behaviour element,
784 corresponds to the number of occurrences in the EIPAAB database. An interactive version of
785 the figure is available at <https://jakemartinresearch.github.io/EIPAAB-database/>
786

787 Additional ecotoxicological endpoints

788 A secondary goal of our SEM was to collate information on additional endpoints (e.g. sub-
789 organismal, reproduction, growth, survival) measured alongside behaviour to facilitate
790 connections across domains that may be useful for future chemical risk assessment and
791 management activities, including potential regulatory decision-making. We found that 51.7%
792 of articles (466/901) also included at least one additional sub-organismal physiological or
793 endocrine endpoint, such as hormone concentrations, biomarker expression, or mRNA
794 transcription. In addition, 39.7% of articles (358/901) measured at least one endpoint that
795 has been used in traditional ecotoxicity testing, such as survival, growth, reproductive
796 output, or developmental abnormalities.

797 Mapping the quality of the evidence base

798 Study validity was not used as an inclusion criterion; however, we did extract information
799 about study validity to enrich the database and to identify potential methodological
800 reporting gaps in the evidence base. We extracted information relevant to a subset of study
801 quality information from the CRED reporting guidelines [41] and several additional validity
802 metrics (see Table 2 and Table S3). To highlight key methodological and/or reporting gaps
803 identified: we observed a low percentage of studies employing (or reporting) experimenter

804 blinding during the scoring or analysing of behaviour (17.0%), randomly (or pseudo-
805 randomly) assigning organisms to exposure treatments (40.2%), providing key details about
806 the pharmaceutical compound used in the exposure (e.g. CAS registry number 24.8% or
807 purity 25.4%), employing exposure concentration verification (e.g. water verification 20.6%
808 or tissue verification 8.9%), following any type of guideline (or modified guideline; 15.0%),
809 or performed the test under Good Laboratory Practice (GLP) conditions (0.7%). In the
810 opposite direction, a high percentage of studies reported details related to the source of
811 the animals (84.4%), aspects of animal care and housing (e.g. animal feeding 79.5%; water
812 quality parameters 89.5%; dark-light cycle 83.9%), providing details about exposure duration
813 (minimum exposure duration 94.1%, maximum duration 94.5%), and describing methods for
814 scoring behavioural endpoints (77.3%; although we note lower levels of extractor
815 consistency with some of these metadata; see Table S6).

816 We should highlight that some of the species validity information may be implied or
817 assumed to those with expert knowledge of that species; for example, if a species is
818 hermaphroditic, sex may not have been reported; alternatively, for species that reach
819 adulthood within 14 days, a 14-day exposure may have implied an adult life stage. With that
820 said, we extracted these metadata based on the definitions given by the authors. Where
821 information was not supplied, it was not assumed or inferred by extractors.

822

823 **Table 2.** All extracted information that relates to study validity. If the validity metadata
824 are aligned with a CRED quality criteria [41], the associated CRED number is provided. The
825 percentage of articles meeting the validity criteria is shown overall, and for each study
826 motivation. NA indicated that the criterion was not part of CRED, but an additional criterion
827 we extracted information about.

Validity criteria	CRED	Overall	Environ	Medical	Basic
A guideline or modified guideline was followed	1	15%	21.8%	6.0%	6.4%
The test was performed under Good Laboratory Practice (GLP) conditions	2	0.7%	1.0%	0.4%	0.0%

Survival, growth and/or reproduction of the test organism(s) was reported	3	39.7%	53.5%	26.1%	15.3%
The test substance is identified with a CAS number	5	24.8%	36.7%	10/7%	7.0%
The purity of the test substance was reported	6	25.4%	38.8%	9.4%	5.7%
Organism(s) life stage is known and reported	8	83.4%	82.9%	91.0%	73.5%
Organism(s) sex is known and reported	8	53.5%	49.6%	59.0%	56.8%
Test organism source is reported	9	84.4%	86.9%	78.8%	84.5%
Information provided regarding feeding	11	79.5%	84.3%	68.2%	80.4 %
Information provided regarding water characteristics (e.g. temperature, pH, oxygen content)	11	89.5%	92.7%	88.5%	80.3 %
Information provided regarding light/dark conditions	11	83.9%	84.1%	85.9%	80.3 %
Exposure minimum duration is defined	14	94.1%	95.6%	90.0%	96.7%
Exposure maximum duration is defined	14	94.5%	96.3%	89.8%	97.0%
The concentration of the test substance is verified in the water (waterborne exposures only)	15	20.6%	35.8%	2.5%	2.7%
The concentration of the test substance is verified in the tissue of the organism (waterborne exposures only)	15	8.9%	13.4%	4.2%	4.7%
Employs randomisation (pseudo-randomisation) of treatment allocation	NA	40.2%	44.9%	32.2%	36.7%
Experimental blinding was performed	NA	17.0%	14.7%	18.9%	21.5%
Methods for scoring behavioural endpoints described	NA	77.3%	76.0%	78.5%	79.4%
Conflict of interest statement is made in the article (with or without conflict identified)	NA	54.8%	50.2%	72.1%	44.3%

828

829 Limitations of the systematic map

830 Two potential limitations of the evidence base to consider are the inherent complexity of
831 assigning therapeutic classes to pharmaceuticals and the complexity of defining animal
832 behavioural responses into discrete categories. First, we used Anatomical Therapeutic
833 Chemical (ATC) Classification to group our compounds, which assigns active ingredients of
834 drugs according to the organ or system on which they act and their therapeutic,
835 pharmacological, and chemical properties [47]. However, it is well recognized, even by the
836 World Health Organization (see “Classification Principles & Challenges” [50]), that

837 pharmaceuticals can be prescribed and used for treating non-target illnesses. For example,
838 beta-blockers (a family of blood-pressure regulating drugs) and certain antihistamines (used
839 for treating allergies), can also be prescribed for the treatment of anxiety [51]. As a result
840 of this complexity, we did not independently assign pharmaceuticals without an existing
841 ATC class to their own therapeutic class. Thus, we highlight that 121 drugs (28% of the total
842 database) are not included in summaries made at the pharmacological group level (e.g.
843 Figure 5A). Similarly, it can also be complex to categorise animal behaviour into discrete
844 overarching categories, as behaviour, and how scientists describe it, varies by species.
845 Moreover, behaviour is context-dependent, in that a given behaviour measured in one
846 context could represent a different underlying motivation in another context. For example,
847 affiliation with a group of conspecifics may represent social propensity in one context but
848 antipredator behaviour in another, if a perceived threat is present. We aimed to reduce
849 ambiguity in assigning behaviours to overarching classes (and the sub-categories within each
850 class) by following the author's definition of the behaviour in the article. This could lead to
851 inconsistencies where, for example, an animal solving a maze task could be defined as a
852 measure of "boldness and exploration" in one article, but the same task could be a measure
853 of "cognition" in another article. Moreover, authors can introduce inconsistencies even
854 within articles if they define or refer to behaviours in multiple ways throughout the text.
855 We note in the consistency section above that there was some extractor disagreement in
856 the assignment of behavioural measures to the overarching categories, ranging from 75.5 to
857 99.3% (median 98.6%, see Table S6), as well as the more specific subcategories with a range
858 of 67.6-99.3% (median 95.8%, see Table S6). We believe that this, in part, reflects the
859 inherent difficulty of assigning behavioural classes across a broad range of taxa and study
860 disciplines.

861 We also identified several potential limitations of the review search methods used.
862 Although we included articles written in all languages in which our review team was
863 proficient (8 different languages), the evidence is likely still biased towards research

864 published in English, because the search strings were written in English, and there is a higher
865 prevalence of English records in the databases used for the search. This is important to
866 highlight as it is well recognized that language can introduce bias in the evidence base [52].
867 With that said, only 4 articles were excluded from the EIPAAB database at the full-text
868 screening stage based on language. Another potential limitation in the review methods for
869 this map is a limited search of the grey literature. Although we allowed for grey literature
870 to be included from our database searches and we solicited grey literature submissions in
871 our supplementary article search advertising calls, we did not search any grey literature
872 databases and removed the planned screening of academic theses from the map. This
873 decision was taken in part due to time and resources needed to screen the evidence base,
874 but also because screening theses would require further quality checks and detailed
875 deduplication cross-checks to remove duplicated published thesis chapters. We suggest this
876 could be added for subsequent systematic review or meta-analytic projects using this
877 database that have a narrower research scope. Finally, we also screened only a subset of
878 articles at the full-text stage in duplicate, and we have discussed the implications of this
879 above regarding extraction consistency.

880

881 Conclusion

882

883 We sought to systematically synthesise all available Evidence for the Impacts of
884 Pharmaceuticals on Aquatic Animal Behaviour (EIPAAB). We report a considerable amount
885 of research on this topic, with 901 articles—representing 1,739 unique species-by-compound
886 combinations—being included in the EIPAAB database. Broadly, we see that the EIPAAB
887 database would be ideal in supporting future ecotoxicology studies and experiments focusing
888 on animal alternatives, identifying and incorporating evidence from behaviour endpoints
889 into chemical risk assessment and management, to highlight knowledge gaps for future
890 research, and to act as a launching pad for further targeted synthesis with more quantitative

891 meta-analytical methodologies. The implications of the collated evidence for
892 policy/management and research are discussed below.

893

894 *Implications for policy and management*

895 Increasingly, behavioural endpoints are being suggested as valuable tools in environmental
896 chemicals assessment and management (including regulatory activities for human and
897 veterinary pharmaceuticals) but are rarely included in such context [17,18]. There are
898 several possible reasons for this, including poor reporting of methodology, using non-
899 standard methods, and limited evidence in an ecotoxicological context of the links between
900 behaviour and adverse outcomes/standard endpoints [53]. The EIPAAB database provides
901 insights into all three of these potential barriers to inclusion in regulation.

902 Firstly, we have indeed identified several methodological and/or reporting pitfalls.
903 This includes a lack of studies employing (or reporting) experimenter blinding during the
904 scoring/analysing of behaviour, randomly (or pseudo-randomly) assigning organisms to
905 exposure treatments, providing key information about the pharmaceutical compound used
906 in the exposure (e.g. CAS registration number or purity), providing key information about
907 the study organism used in the exposure (e.g. sex), and validating exposure concentration
908 (e.g. water verification or tissue verification). Research on the effects of pharmaceuticals
909 on animal behaviour would benefit from addressing these aspects of methodical reporting
910 and study methodology, many of which require little additional effort from experimenters,
911 and we hope that this review can be a catalyst to improve these aspects in the field. With
912 that said, there are many articles that do not have these identified pitfalls in the evidence
913 base, and if required, those seeking to use this evidence for regulatory purposes (or likewise)
914 could filter the database to help identify those studies that meet relevant criteria. More
915 broadly, the field of behavioural ecotoxicology and toxicology studies with animal
916 alternatives (e.g. fish models) could benefit from the use of data reporting and reliability
917 guidelines specific to behavioural endpoints to increase the likelihood of these studies being

918 included in future chemical risk assessment and management, such as regulatory processes.
919 A recent set of such guidelines is provided in EthoCRED ([54]), a behavioural endpoint-
920 specific adaptation of the parent CRED guidelines. The use of such guidelines, like
921 EthoCRED, would improve reporting of important methodological information, guide
922 methodological decision-making for future studies, and increase the replicability of the
923 field.

924 Secondly, the database included a total of 63 different sub-categories of measured
925 behaviours and for which aquatic species they were measured. From these data, we suggest
926 that new standardised ecotoxicity test guidelines that include behaviour could be developed
927 by looking for the most common or most widely applicable testing parameters. As an
928 example, our SEM has revealed a wealth of studies focused on fishes (especially for
929 zebrafish) across multiple behavioural endpoints (particularly movement, anxiety/boldness,
930 and pre-copulatory/copulatory behaviours); by comparing such methods, one could arrive
931 at the most broadly suitable tests. We believe that the next step in achieving this would be
932 a focused review and meta-analysis, looking at the specific methods used for candidate
933 behavioural tests and the nature of the data they provide, followed by multi-lab validity
934 and repeatability tests once a candidate protocol is established.

935 Thirdly, within the EIPAAB database, we have identified which studies can provide
936 direct links between behaviour and other adverse outcomes/standard endpoints.
937 Specifically, we have identified studies that also measured sub-organismal
938 physiological/endocrine endpoints, as well as which studies assessed more traditional
939 endpoints like animal growth, survival, and/or reproduction alongside behaviour. We see
940 this as a starting point for future work to connect behavioural endpoints to molecular
941 initiating events and to endpoints currently being used in traditional risk assessments,
942 including integration with the adverse outcome pathway (AOP) concept [55]. Specifically,
943 we suggest targeted meta-analytic approaches focusing on articles that have measured
944 behaviour alongside additional morphometric endpoints (sub-organismal, growth, survival,

945 and/or reproduction endpoints), identifying potential correlations in the direction and
946 magnitude of observed effects.

947

948 *Implications for research*

949 Our SEM highlights that this rapidly growing research area has several knowledge clusters
950 appropriate for further quantitative synthesis. Specifically, future meta-analytical work
951 could focus on the behavioural impacts of antidepressants, antiepileptics, or estrogens,
952 particularly for endpoints like locomotion, boldness, and reproductive behaviours. We have
953 also identified that the evidence base is heavily skewed towards research on zebrafish,
954 which is perhaps unsurprising given that the zebrafish is a well-established model in
955 (eco)toxicological, medical, and basic research [56,57]. Therefore, future comparative
956 synthesis across behavioural categories or compounds using zebrafish may offer a suitably
957 homogenous prospect for detailed meta-analysis. Indeed, the available evidence on
958 zebrafish could be a valuable step towards disentangling and identifying quantitative
959 thresholds at which exposure to a given pharmaceutical affects behaviour. For instance,
960 how, and at which exposure concentration, the antidepressant fluoxetine impacts fish
961 behaviour has been disputed in the earlier literature [43].

962 We would also like to highlight gaps in the evidence base that require more primary
963 research. Firstly, there were relatively few studies using wild-caught animals. Wild-caught
964 *versus* lab-reared organisms can differ greatly in their behaviour and underlying physiology
965 traits [58-61], and thus, may also respond differently to pharmaceutical exposure. More
966 research using wild-caught organisms could help identify whether lab-reared model species
967 are equally sensitive to pharmaceutical exposure (e.g. [62]). Additionally, locomotion and
968 boldness were by far the most common behavioural endpoints measured. We argue that
969 measuring contaminant-induced impacts on a more diverse array of behavioural endpoints—
970 particularly those with obvious links to fitness (e.g. pre and post-copulatory, antipredator,
971 and foraging behaviours)—would give a more holistic understanding of potential impacts on

972 aquatic wildlife. However, we also acknowledge that the most commonly measured
973 behaviours, locomotion and boldness, are often the simplest to measure and offer the
974 highest throughput. There was also a distinct lack of studies measuring behaviour within a
975 social context (e.g. free-swimming groups) and employing exposure durations greater than
976 a week; it is reasonable to assume that for most animals, real-world exposures will occur in
977 social groups (animals rarely, if ever, exist in a social vacuum; [63]), and that many
978 pollutants would have environmental or biological half-lives exceeding seven days. Thus,
979 future research addressing the impacts of pharmaceutical pollutants on animals under a
980 social context and over chronic time scales would improve our understanding of real-world
981 impacts. Finally, we suggest that research is prioritised on pharmaceutical compounds that
982 are absent or infrequently represented in our database, yet are common in the environment
983 (i.e. what evidence are we currently missing). This could be done by cross-checking the
984 EIPAAB database against recent publications (e.g. [1]) and open databases reporting
985 environmental pharmaceutical concentrations around the world (e.g. AstraZeneca
986 EcoPharmacoVigilance Dashboard [64]; Umwelt Bundesamt “UBA-PHARMS” database [65];
987 NORMAN EMPODAT chemical occurrence database [66]).

988 We identified that many of the studies in our database have an environmental
989 motivation; however, we also identified a lot of available research in adjacent fields that
990 focus on medical research questions and basic research questions, particularly with fish
991 models employed as animal alternatives. Future work assessing the bibliometric connections
992 between the fields would be interesting to reveal how much crosstalk (if any) exists via the
993 use of co-author and co-citation networks [34].

994 We have already pointed out several gaps in study validity that should be considered
995 in future studies and noted that using standard reporting guidelines would increase their
996 utility in regulatory processes. We also advocate that the use of reporting guidelines (e.g.
997 EthoCRED) will more broadly increase the robustness and replicability of studies assessing
998 the effects of pharmaceuticals on aquatic animal behaviour. Importantly, we highlight that

999 disclosing details about how animals were housed, how they were assigned to treatments,
1000 how the behaviour was recorded and scored, and the use of blind scoring, is paramount to
1001 increasing transparency and reducing unintended experimenter bias.

1002

1003 **Declarations**

1004 *Ethics approval and consent to participate:* Not applicable

1005

1006 *Consent for publication:* Not applicable

1007

1008 *Availability of data and materials:* All supplementary files and the EIPAAB database can be
1009 accessed from the Open Sciences Framework (OSF) at
1010 <https://doi.org/10.17605/OSF.IO/ATWY6>. Below, we provide a list of all supplementary
1011 files and individual links. The R script used to generate the summary statistics and figures
1012 presented in this manuscript is available on GitHub
1013 (<https://github.com/JakeMartinResearch/EIPAAB-database>).

1014

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1016

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1028

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1030 first, second, and last authors. We have reported our methods in the manuscript text
1031 following the MeRIT guidelines to improve author contributions' granularity and
1032 accountability (see main text). Briefly, JMM and ESM conceived and designed this study,
1033 with guidance on study design from SN and ML. All authors participated in piloting the data
1034 extraction process and refining the study design. JMM, ESM, KNF, JS, JAB, DC, IYL, HT, MM,
1035 LML, MGB, SO, KRS, NSH, ESJT, and ML extracted the data to build the systematic evidence
1036 map. JMM, ESM, and MM extracted all supplementary metadata for compounds and species.
1037 JMM tidied and analysed the EIPAAB database (writing all R and Python Scripts used during
1038 these steps), with input from ESM. JMM and ESM wrote the manuscript with input from all
1039 authors. All authors read and approved the final manuscript.

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1044 America 40th Annual General Meeting in Toronto, Canada.

1045

1046 **Additional Files**

1047 **(1) EIPAAB database:** the Evidence of the Impacts of Pharmaceuticals on Aquatic
1048 Animal Behaviour (EIPAAB) database created from the systematic review and data
1049 extraction described in the manuscript (name: EIPAAB-database.csv; link:
1050 <https://osf.io/rzwv2> OR [https://github.com/JakeMartinResearch/EIPAAB-](https://github.com/JakeMartinResearch/EIPAAB-database)
1051 [database](https://github.com/JakeMartinResearch/EIPAAB-database))

- 1052 (2) **Supplementary file 1:** ROSES Form (name: martin-et-al-supp-file-1-ROSES.xlsx;
1053 link: <https://osf.io/g2h4x>)
- 1054 (3) **Supplementary file 2:** Title and abstract screening decisions (name: martin-et-al-
1055 supp-file-2-title-abstract-screen-decision.csv; link: <https://osf.io/hafs5>)
- 1056 (4) **Supplementary file 3:** List of eligibility disagreements for duplicate screenings at
1057 the full-text screening stage (name: martin-et-al-supp-file-3-eligibility-
1058 disagreements.csv; link: <https://osf.io/ajbs7>)
- 1059 (5) **Supplementary file 4:** Full-text screening excluded articles (name: martin-et-al-
1060 supp-file-4-full-text-exculded-atricles.csv; link: <https://osf.io/3ynrp>)
- 1061 (6) **Supplementary file 5:** Full-text extraction form from Qualtrics (name: martin-et-
1062 al-supp-file-5-full-text-screening-extraction-form.pdf; link: <https://osf.io/p57ev>)
- 1063 (7) **Supplementary file 6:** Read me file (name: READ-ME.csv; link: <https://osf.io/8x3rk>)
- 1064 (8) **Supplementary file 7:** R script used to summarise the EIPAAB Database interactive
1065 HTML (<https://jakemartinresearch.github.io/EIPAAB-database/>); a static version is
1066 also available on OSF (name: martin-et-al-supp-file-7-r-script.Rmd; link:
1067 <https://osf.io/z5tw4>)
- 1068 (9) **Supplementary file 8:** The Web of Science annual article counts for each of the
1069 most common research categories identified in the database (name: martin-et-al-
1070 supp-file-9-wos-research-areas-1992-2022.csv; link: <https://osf.io/35hgj>)

1071

1072 **References**

- 1073 1. Wilkinson JL, Boxall ABA, Kolpin DW, Leung KMY, Lai RWS, Wong D, et al.
1074 Pharmaceutical pollution of the world's rivers. Proceedings of the National Academy of
1075 Sciences. 2022;119:1-10.

1076

- 1077 2. Patel M, Kumar R, Kishor K, Mlsna T, Pittman CU, Mohan D. Pharmaceuticals of
1078 emerging concern in aquatic systems: Chemistry, occurrence, effects, and removal
1079 methods. *Chem Rev.* 2019;119:3510-673.
1080
- 1081 3. Pal A, Gin KYH, Lin AYC, Reinhard M. Impacts of emerging organic contaminants on
1082 freshwater resources: Review of recent occurrences, sources, fate and effects. *Science of*
1083 *the Total Environment.* 2010;408:6062-9.
1084
- 1085 4. Richmond EK, Rosi EJ, Walters DM, Fick J, Hamilton SK, Brodin T, et al. A diverse suite
1086 of pharmaceuticals contaminates stream and riparian food webs. *Nat Commun* [Internet].
1087 2018;9:4491. Available from: <http://www.nature.com/articles/s41467-018-06822-w>
1088
- 1089 5. Ramirez A, Brain R, Usenka S, Mottaleb M, O'Donnell J, Stahl L, et al. Occurrence of
1090 pharmaceuticals and personal care products in fish: Results of a national pilot study in the
1091 United States. *Environ Toxicol Chem.* 2009;28:2587-97.
1092
- 1093 6. Ågerstrand M, Berg C, Björlenius B, Breitholtz M, Brunström B, Fick J, et al. Improving
1094 environmental risk assessment of human pharmaceuticals. *Environ Sci Technol.*
1095 2015;49:5336-45.
1096
- 1097 7. OECD. *Pharmaceutical Residues in Freshwater: Hazards and Policy Responses.* 2019;17.
1098 Available from: [https://www.oecd.org/environment/resources/Pharmaceuticals-residues-](https://www.oecd.org/environment/resources/Pharmaceuticals-residues-in-freshwater-policy-highlights-preliminary-version.pdf)
1099 [in-freshwater-policy-highlights-preliminary-version.pdf](https://www.oecd.org/environment/resources/Pharmaceuticals-residues-in-freshwater-policy-highlights-preliminary-version.pdf)
1100
- 1101 8. Gunnarsson L, Snape JR, Verbruggen B, Owen SF, Kristiansson E, Margiotta-Casaluci L,
1102 et al. Pharmacology beyond the patient - The environmental risks of human drugs. *Environ*

1103 Int [Internet]. 2019;129:320-32. Available from:
1104 <https://doi.org/10.1016/j.envint.2019.04.075>
1105
1106 9. Gunnarsson L, Jauhiainen A, Kristiansson E, Nerman O, Larsson DGJ. Evolutionary
1107 Conservation of Human Drug Targets in Organisms used for Environmental Risk
1108 Assessments. Environ Sci Technol [Internet]. 2008;42:5807-13. Available from:
1109 <http://pubs.acs.org/doi/abs/10.1021/es8005173>
1110
1111 10. Berninger JP, Brooks BW. Leveraging mammalian pharmaceutical toxicology and
1112 pharmacology data to predict chronic fish responses to pharmaceuticals. Toxicol Lett
1113 [Internet]. 2010;193:69-78. Available from:
1114 <http://www.ncbi.nlm.nih.gov/pubmed/20025941>
1115
1116 11. Brown AR, Gunnarsson L, Kristiansson E, Tyler CR. Assessing variation in the potential
1117 susceptibility of fish to pharmaceuticals, considering evolutionary differences in their
1118 physiology and ecology. Philosophical Transactions of the Royal Society B: Biological
1119 Sciences. 2014;369.
1120
1121 12. Rand-Weaver M, Margiotta-casaluci L, Patel A, Panter GH, Owen SF, Sumpter JP. The
1122 Read-Across Hypothesis and Environmental Risk Assessment of Pharmaceuticals. Environ
1123 Sci Technol. 2013;47:11384-95.
1124
1125 13. Brodin T, Piovano S, Fick J, Klaminder J, Heynen M, Jonsson M. Ecological effects of
1126 pharmaceuticals in aquatic systems – impacts through behavioural alterations.
1127 Philosophical Transactions of the Royal Society B: Biological Sciences. 2014;369:20130580.
1128

- 1129 14. Corcoran J, Winter MJ, Tyler CR. Pharmaceuticals in the aquatic environment: a
1130 critical review of the evidence for health effects in fish. *Crit Rev Toxicol* [Internet]. 2010
1131 [cited 2012 Mar 30];40:287-304. Available from:
1132 <http://www.ncbi.nlm.nih.gov/pubmed/20225984>
1133
- 1134 15. Fabbri E. Pharmaceuticals in the environment: Expected and unexpected effects on
1135 aquatic fauna. *Ann N Y Acad Sci*. 2015;1340:20-8.
1136
- 1137 16. Saaristo M, Brodin T, Balshine S, Bertram MG, Brooks BW, Ehlman SM, et al. Direct and
1138 indirect effects of chemical contaminants on the behaviour, ecology and evolution of
1139 wildlife. *Proceedings of the Royal Society B: Biological Sciences* [Internet].
1140 2018;285:20181297. Available from:
1141 <http://rspb.royalsocietypublishing.org/lookup/doi/10.1098/rspb.2018.1297>
1142
- 1143 17. Ford AT, Ågerstrand M, Brooks BW, Allen J, Bertram MG, Brodin T, et al. The Role of
1144 Behavioral Ecotoxicology in Environmental Protection. *Environ Sci Technol*. 2021;
1145
- 1146 18. Ågerstrand M, Arnold K, Balshine S, Brodin T, Brooks BW, Maack G, et al. Use of
1147 behavioural endpoints in the regulation of chemicals. *Environ Sci Process Impacts*.
1148 2020;22:49-65.
1149
- 1150 19. Melvin SD, Wilson SP. The utility of behavioral studies for aquatic toxicology testing: A
1151 meta-analysis. *Chemosphere* [Internet]. 2013;93:2217-23. Available from:
1152 <http://dx.doi.org/10.1016/j.chemosphere.2013.07.036>
1153

- 1154 20. Robinson PD. Behavioural toxicity of organic chemical contaminants in fish: application
1155 to ecological risk assessments (ERAs). *Canadian Journal of Fisheries and Aquatic Sciences*.
1156 2009;66:1179-88.
- 1157
- 1158 21. Amiard-Triquet C. Behavioral Disturbances: The Missing Link between Sub-Organismal
1159 and Supra-Organismal Responses to Stress? Prospects Based on Aquatic Research. *Human
1160 and Ecological Risk Assessment*. 2009;15:87-110.
- 1161
- 1162 22. Legradi JB, Di Paolo C, Kraak MHS, van der Geest HG, Schymanski EL, Williams AJ, et
1163 al. An ecotoxicological view on neurotoxicity assessment. *Environ Sci Eur* [Internet].
1164 2018;30:1-34. Available from: <https://doi.org/10.1186/s12302-018-0173-x>
- 1165
- 1166 23. Brooks BW, Chambliss CK, Stanley JK, Ramirez A, Banks KE, Johnson RD, et al.
1167 Determination of select antidepressants in fish from an effluent-dominated stream.
1168 *Environmental toxicology and chemistry / SETAC* [Internet]. 2005;24:464-9. Available
1169 from: <http://www.ncbi.nlm.nih.gov/pubmed/15720009>
- 1170
- 1171 24. Arnnok P, Singh RR, Burakham R, Pérez-Fuentetaja A, Aga DS. Selective Uptake and
1172 Bioaccumulation of Antidepressants in Fish from Effluent-Impacted Niagara River. *Environ
1173 Sci Technol* [Internet]. 2017;51:10652-62. Available from:
1174 <http://pubs.acs.org/doi/abs/10.1021/acs.est.7b02912>
- 1175
- 1176 25. Calisto V, Esteves VI. Psychiatric pharmaceuticals in the environment. *Chemosphere*
1177 [Internet]. 2009 [cited 2013 Mar 7];77:1257-74. Available from:
1178 <http://www.ncbi.nlm.nih.gov/pubmed/19815251>
- 1179

- 1180 26. David A, Lange A, Tyler CR, Hill EM. Concentrating mixtures of neuroactive
1181 pharmaceuticals and altered neurotransmitter levels in the brain of fish exposed to a
1182 wastewater effluent. *Science of the Total Environment* [Internet]. 2018;621:782-90.
1183 Available from: <https://doi.org/10.1016/j.scitotenv.2017.11.265>
1184
- 1185 27. Grabicova K, Grabic R, Fedorova G, Fick J, Cerveny D, Kolarova J, et al.
1186 Bioaccumulation of psychoactive pharmaceuticals in fish in an effluent dominated stream.
1187 *Water Res* [Internet]. 2017;124:654-62. Available from:
1188 <http://www.sciencedirect.com/science/article/pii/S0043135417306711>
1189
- 1190 28. Bókony V, Verebélyi V, Ujhegyi N, Mikó Z, Nemesházi E, Szederkényi M, et al. Effects
1191 of two little-studied environmental pollutants on early development in anurans.
1192 *Environmental Pollution*. 2020;260.
1193
- 1194 29. Brodin T, Fick J, Jonsson M, Klaminder J. Dilute concentrations of a psychiatric drug
1195 alter behavior of fish from natural populations. *Science (1979)* [Internet]. 2013 [cited 2013
1196 Feb 27];339:814-5. Available from: <http://www.ncbi.nlm.nih.gov/pubmed/23413353>
1197
- 1198 30. Buřič M, Grabicová K, Kubec J, Kouba A, Kuklina I, Kozák P, et al. Environmentally
1199 relevant concentrations of tramadol and citalopram alter behaviour of an aquatic
1200 invertebrate. *Aquatic Toxicology*. 2018;200:226-32.
1201
- 1202 31. Martin JM, Bertram MG, Saaristo M, Ecker TE, Hannington SL, Tanner JL, et al. Impact
1203 of the widespread pharmaceutical pollutant fluoxetine on behaviour and sperm traits in a
1204 freshwater fish. *Science of the Total Environment* [Internet]. 2019;650:1771-8. Available
1205 from: <https://doi.org/10.1016/j.scitotenv.2018.09.294>
1206

- 1207 32. Rearick DC, Ward J, Venturelli P, Schoenfuss H. Environmental oestrogens cause
1208 predation-induced population decline in a freshwater fish. *R Soc Open Sci.* 2018;5.
1209
- 1210 33. James KL, Randall NP, Haddaway NR. A methodology for systematic mapping in
1211 environmental sciences. *Environ Evid.* 2016;5:1-13.
1212
- 1213 34. Nakagawa S, Samarasinghe G, Haddaway NR, Westgate MJ, O’Dea RE, Noble DWA, et
1214 al. Research Weaving: Visualizing the Future of Research Synthesis. *Trends Ecol Evol*
1215 [Internet]. 2019;34:224-38. Available from: <https://doi.org/10.1016/j.tree.2018.11.007>
1216
- 1217 35. Wolffe TAM, Whaley P, Halsall C, Rooney AA, Walker VR. Systematic evidence maps as
1218 a novel tool to support evidence-based decision-making in chemicals policy and risk
1219 management. *Environ Int* [Internet]. 2019;130:104871. Available from:
1220 <https://doi.org/10.1016/j.envint.2019.05.065>
1221
- 1222 36. Wolffe TAM, Vidler J, Halsall C, Hunt N, Whaley P. A Survey of Systematic Evidence
1223 Mapping Practice and the Case for Knowledge Graphs in Environmental Health and
1224 Toxicology. *Toxicological Sciences.* 2020;175:35-49.
1225
- 1226 37. Nakagawa S, Ivimey-Cook ER, Grainger MJ, O’Dea RE, Burke S, Drobniak SM, et al.
1227 Method Reporting with Initials for Transparency (MeRIT) promotes more granularity and
1228 accountability for author contributions. *Nat Commun.* 2023;14:1-5.
1229
- 1230 38. Martin JM, Bertram MG, Blanchfield PJ, Brand JA, Brodin T, Brooks BW, et al. Evidence
1231 of the impacts of pharmaceuticals on aquatic animal behaviour: a systematic map
1232 protocol. *Environ Evid* [Internet]. 2021;10:1-10. Available from:
1233 <https://doi.org/10.1186/s13750-021-00241-z>

1234

1235 39. Haddaway NR, Macura B, Whaley P, Pullin AS. ROSES Reporting standards for
1236 Systematic Evidence Syntheses: Pro forma, flow-diagram and descriptive summary of the
1237 plan and conduct of environmental systematic reviews and systematic maps. *Environ Evid*
1238 [Internet]. 2018;7:4-11. Available from: <https://doi.org/10.1186/s13750-018-0121-7>

1239

1240 40. Ouzzani M, Hammady H, Fedorowicz Z, Elmagarmid A. Rayyan-a web and mobile app
1241 for systematic reviews. *Syst Rev* [Internet]. 2016;5:1-10. Available from:
1242 <http://dx.doi.org/10.1186/s13643-016-0384-4>

1243

1244 41. Moermond CTA, Kase R, Korkaric M, Ågerstrand M. CRED: Criteria for reporting and
1245 evaluating ecotoxicity data. *Environ Toxicol Chem*. 2016;35:1297-309.

1246

1247 42. Tanoue R, Margiotta-Casaluci L, Huerta B, Runnalls TJ, Eguchi A, Nomiya K, et al.
1248 Protecting the environment from psychoactive drugs: Problems for regulators illustrated
1249 by the possible effects of tramadol on fish behaviour. *Science of the Total Environment*
1250 [Internet]. 2019;664:915-26. Available from:
1251 <https://doi.org/10.1016/j.scitotenv.2019.02.090>

1252

1253 43. Sumpter JP, Donnachie RL, Johnson AC. The apparently very variable potency of the
1254 anti-depressant fluoxetine. *Aquatic Toxicology* [Internet]. 2014;151:57-60. Available from:
1255 <http://www.ncbi.nlm.nih.gov/pubmed/24411166>

1256

1257 44. R Core Team. R: A Language and Environment for Statistical Computing [Internet].
1258 Vienna, Austria: R Foundation for Statistical Computing; 2023. Available from:
1259 <https://www.r-project.org/>

1260

1261 45. Kim S, Chen J, Cheng T, Gindulyte A, He J, He S, et al. PubChem 2023 update. *Nucleic*
1262 *Acids Res.* 2023;51:D1373-80.
1263
1264 46. Schoch CL, Ciufo S, Domrachev M, Hotton CL, Kannan S, Khovanskaya R, et al. NCBI
1265 *Taxonomy: A comprehensive update on curation, resources and tools. Database.*
1266 2020;2020:1-21.
1267
1268 47. WHO Collaborating Centre for Drug Statistics Methodology. ATC classification index
1269 with DDDs [Internet]. 2024 [cited 2024 Jun 1]. Available from:
1270 https://atcddd.fhi.no/atc_ddd_index_and_guidelines/atc_ddd_index/
1271
1272 48. IUCN. The IUCN Red List of Threatened Species [Internet]. Version 2023-1. 2023 [cited
1273 2024 Jun 1]. Available from: <https://www.iucnredlist.org/>
1274
1275 49. Haddaway N, Macura B, Whaley P, Pullin A. ROSES Flow Diagram for Systematic
1276 *Reviews* [Internet]. 2018. Available from:
1277 <https://doi.org/10.6084/m9.figshare.5897389.v3>
1278
1279 50. WHO. Anatomical Therapeutic Chemical (ATC) Classification - Classification principles
1280 and challenges [Internet]. 2024 [cited 2024 Jun 1]. Available from:
1281 <https://www.who.int/tools/atc-ddd-toolkit/atc-classification>
1282
1283 51. Garakani A, Murrough JW, Freire RC, Thom RP, Larkin K, Buono FD, et al.
1284 *Pharmacotherapy of Anxiety Disorders: Current and Emerging Treatment Options. Front*
1285 *Psychiatry.* 2020;11:1-21.
1286

1287 52. Livoreil B, Glanville J, Haddaway NR, Bayliss H, Bethel A, De Lachapelle FF, et al.
1288 Systematic searching for environmental evidence using multiple tools and sources. *Environ*
1289 *Evid.* 2017;6:1-14.
1290
1291 53. Ågerstrand M, Sobek A, Lilja K, Linderöth M, Wendt-Rasch L, Wernersson AS, et al. An
1292 academic researcher's guide to increased impact on regulatory assessment of chemicals.
1293 *Environ Sci Process Impacts.* 2017;19:644-55.
1294
1295 54. Bertram MG, Ågerstrand M, Thoré ESJ, Allen J, Balshine S, Brand JA, et al. EthoCRED:
1296 A framework to guide reporting and evaluation of the relevance and reliability of
1297 behavioural ecotoxicity studies. *Biological Reviews.* 2024;In press.
1298
1299 55. Ankley GT, Bennett RS, Erickson RJ, Hoff DJ, Hornung MW, Johnson RD, et al. Adverse
1300 outcome pathways: A conceptual framework to support ecotoxicology research and risk
1301 assessment. *Environ Toxicol Chem.* 2010;29:730-41.
1302
1303 56. Briggs JP. The zebrafish: a new model organism for integrative physiology. *Am J*
1304 *Physiol Regul Integr Comp Physiol.* 2002;282:3-9.
1305
1306 57. Dai YJ, Jia YF, Chen N, Bian WP, Li QK, Ma YB, et al. Zebrafish as a model system to
1307 study toxicology. *Environ Toxicol Chem.* 2014;33:11-7.
1308
1309 58. Burns JG, Saravanan A, Helen Rodd F. Rearing environment affects the brain size of
1310 guppies: Lab-reared guppies have smaller brains than wild-caught guppies. *Ethology.*
1311 2009;115:122-33.
1312

- 1313 59. Morgan R, Sundin J, Finnøen MH, Dresler G, Vendrell MM, Dey A, et al. Are model
1314 organisms representative for climate change research? Testing thermal tolerance in wild
1315 and laboratory zebrafish populations. *Conserv Physiol.* 2019;7:1-11.
1316
- 1317 60. Turko AJ, Firth BL, Craig PM, Eliason EJ, Raby GD, Borowiec BG. Physiological
1318 differences between wild and captive animals: a century-old dilemma. *J Exp Biol.*
1319 2023;226.
1320
- 1321 61. Näslund J. Reared to become wild-like: addressing behavioral and cognitive deficits in
1322 cultured aquatic animals destined for stocking into natural environments—a critical
1323 review. *Bull Mar Sci.* 2021;97:489-538.
1324
- 1325 62. Vossen LE, Červený D, Sen Sarma O, Thörnqvist PO, Jutfelt F, Fick J, et al. Low
1326 concentrations of the benzodiazepine drug oxazepam induce anxiolytic effects in wild-
1327 caught but not in laboratory zebrafish. *Science of the Total Environment.* 2020;703.
1328
- 1329 63. Martin JM, McCallum ES. Incorporating Animal Social Context in Ecotoxicology: Can a
1330 Single Individual Tell the Collective Story? *Environ. Sci. Tech.* 2021;55:1090810910
1331
- 1332 64. AstraZeneca. EcoPharmacoVigilance Dashboard [Internet]. Available from:
1333 [https://www.astrazeneca.com/sustainability/environmental-
1334 in-the-environment.html](https://www.astrazeneca.com/sustainability/environmental-protection/pharmaceuticals-in-the-environment.html)
1335
- 1336 65. Bundesamt U. PHARMS-UBA database [Internet]. Available from:
1337 [https://www.umweltbundesamt.de/en/database-pharmaceuticals-in-the-
1338 environment-0](https://www.umweltbundesamt.de/en/database-pharmaceuticals-in-the-environment-0)

1339 66. Network N. NORMAN EMPODAT Database - Chemical Occurrence Data [Internet].

1340 Available from: <https://www.norman-network.com/nds/empodat/>

1341

1342