

1 **Title:** Measuring the edges of species' geographic ranges

2 **Author name:** Alexa L. Fredston

3 **Institutions and addresses:** Department of Ocean Sciences, University of California, Santa
4 Cruz, 1156 High Street, Santa Cruz, CA, 95064

5 **Acknowledgments:** Zoë Kitchel, Owen R. Liu, and Monte Neate-Clegg provided valuable
6 feedback on an earlier draft of the manuscript.

7 **Author contributions:** ALF conceived of, drafted, and revised the manuscript, and conducted
8 all analyses.

9 **Data availability statement:** All data and code are available at
10 <https://github.com/afredston/range-edge-methods> and will be archived in a repository with a DOI
11 when the manuscript is accepted for publication.

12 Abstract

- 13 1. The fundamental unit of spatial ecology is a species range: the geographic area that it
14 occupies. Species ranges are delineated by range edges (also known as boundaries or
15 limits). Why range edges occur where they do and not elsewhere, and what makes them
16 move, has been an active area of research since the 19th century. In the present day,
17 range edge dynamics are an important metric of biodiversity's response to climate
18 change, as species shift toward the poles to track their climatic niches. Yet methods for
19 measuring range edges and quantifying their displacement have never been formalized.
- 20 2. Here I described common methods for describing range edge positions and applied
21 them to example data for a bird species and a fish species, using some of the most
22 popular datasets in climate biogeography: the Audubon Society Christmas Bird Count
23 and a National Oceanic and Atmospheric Administration bottom trawl survey.
- 24 3. I showed that the choice of range edge metric influences where range edge positions
25 are estimated to occur; whether they are estimated to be shifting over time; and the
26 estimated rate of shift. The lack of universal metrics for range edges has likely shaped
27 statistics reported in synthesis studies that measured overall biodiversity responses to
28 climate change and global rates of range shifts. Through simulation, I found that reliably
29 detecting range edge shifts may require decades of data or more, suggesting that many
30 global change studies in this field are underpowered.
- 31 4. Pairing metrics to research questions, sharing raw data and code, and conducting power
32 analyses before reporting statistically significant results will all help to minimize this
33 issue. Going forward, the field of biogeography should confront the degree to which *ad*
34 *hoc* methods have influenced our understanding of range edge dynamics, and move
35 toward universally accepted metrics.

36 Data/code

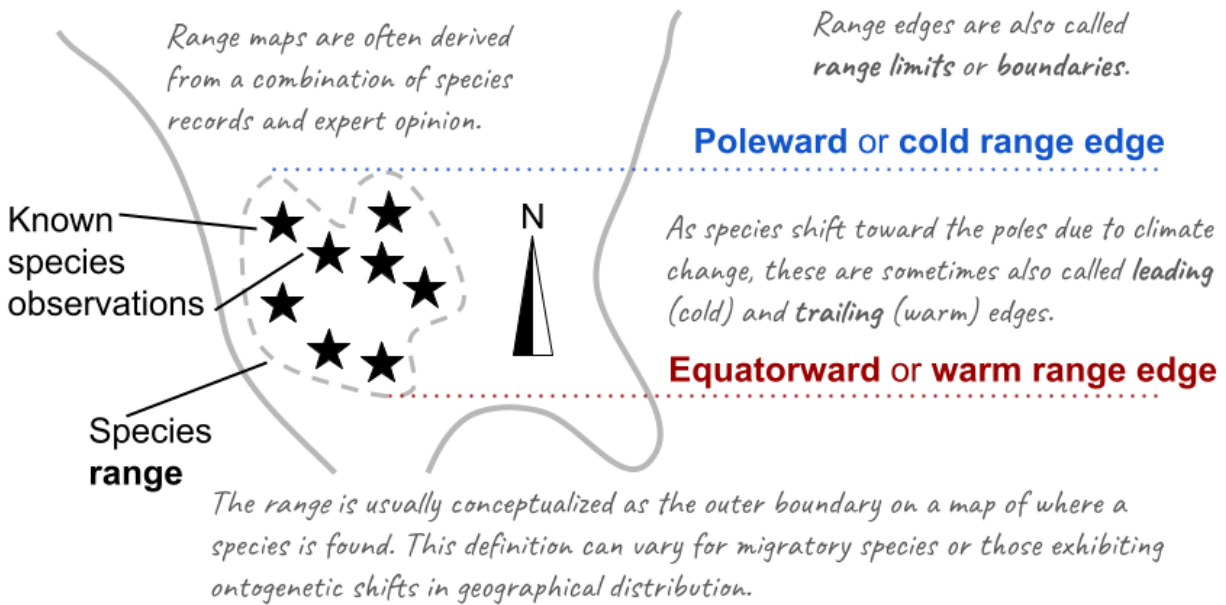
37 The data and code for this project are available on GitHub. An anonymized link to the repository
38 can be found here: <https://anonymous.4open.science/r/range-edge-methods-48BD/>.

39 Keywords

40 Range edges; range boundaries; range limits; biogeography; range shifts; non-random
41 subsampling; time-series; power analysis

42 Introduction

43 Range edges (or limits, or boundaries) are a fundamental unit of spatial ecology. Range
44 edges are commonly understood to delineate where a species' geographic range ends on a
45 map (Gaston 2003). They have fascinated ecologists and evolutionary biologists since the
46 origins of our field (MacArthur 1972; Wallace 1876). Range edges are dynamic, over short and
47 long time scales (Boisvert-Marsh *et al.* 2014; Williams & Blois 2018). In recent decades, many
48 species' range edges have begun to shift polewards in concert with global climate change
49 (Hickling *et al.* 2006; Parmesan & Yohe 2003). These recent shifts have underscored how range
50 edges are part of humanity's relationship with nature; they influence our management
51 boundaries and conservation plans, and are embedded in our culture and history (Pecl *et al.*
52 2017). Range edges are also where we often test foundational hypotheses in ecology and
53 evolutionary biology—about dispersal, population genetics, species interactions, species
54 invasions, and of course biogeography (Sexton *et al.* 2009; Willi & Van Buskirk 2019). Here I
55 define range edges as geographical limits outlining the boundaries of a species' range (Figure
56 1).



57

58 **Figure 1.** Illustration of terms and concepts relevant to species range edges, using a Northern
 59 Hemisphere orientation (*i.e.*, the poleward edge is toward the north). Note that the present study
 60 focused only on geographical range edges (not depth or elevational edges).

61

62 To measure the position of a species' range, studies commonly use some version of an
 63 abundance-weighted geographical center, *i.e.*, a range centroid (La Sorte & Thompson, III 2007;
 64 Perry *et al.* 2005). No such universal metric exists for range edges, and their quantification is far
 65 more variable in literature (Lawlor *et al.* 2024). Further, because centroid metrics are mean
 66 statistics and range edge metrics are extreme value statistics, greater variation may be
 67 expected in the latter. Range edges have been characterized using expert opinion, species
 68 distribution models, or through metrics based on species' abundance or presence, in one, two,
 69 or three spatial dimensions—each using bespoke methods that are not standardized within
 70 fields or among studies (Gaston 2003; Guisan & Thuiller 2005). The effect of these
 71 methodological choices on analytical conclusions has been little-explored, although a recent

72 study illuminated how user decisions influence the delineation of species range maps (Hughes
73 *et al.* n.d.).

74 While biogeography in the 20th century focused on describing species ranges (Gaston
75 2000; MacArthur 1972), the field is increasingly focused on range *shifts* in light of human
76 impacts on the planet (Parmesan & Yohe 2003). A large and growing literature has documented
77 recent shifts in species' ranges and range edges, often tracking local climate velocities and/or
78 moving toward the poles (Lenoir *et al.* 2020). Many of these studies have compared a historic
79 dataset to a modern "resurvey" to measure range edge shifts. Unfortunately, this approach may
80 lead to misleading inferences about global change impacts (Stuble *et al.* 2021), and longitudinal
81 studies with more time points for studying range edge dynamics are rare. Inconsistencies in
82 global change metrics used by individual studies and researchers have ripple effects in
83 synthesis science, making it difficult to accurately estimate large-scale patterns and trends
84 (Brown *et al.* 2016; Lenoir *et al.* 2020). Further, simulation testing found that many years of data
85 are often required to accurately estimate population trends (White 2019); if the same is true of
86 range trends, many range edge shift studies are possibly underpowered.

87 The challenge of measuring range edge dynamics is evident in my own work. Using a
88 50-year annual time-series of marine bottom trawl surveys with hundreds of samples per year,
89 colleagues and I found that it was still difficult to separate the signal of a putative range edge
90 shift from noise [CITATION REDACTED FOR DOUBLE-BLIND PEER REVIEW]. A follow-up
91 study that used alternative methods to measure range edge dynamics came to very different
92 conclusions about which species' range edges had shifted and how far [CITATION REDACTED
93 FOR DOUBLE-BLIND PEER REVIEW].

94 Here I applied a suite of range edge metrics that are common in the biogeography
95 literature (Table 1) to standardized biodiversity data. I found substantial heterogeneity in the
96 edge positions measured and the resulting estimates of range edge shifts. I showed that

97 methodological choices in measuring range edges can influence not only the rate of inferred
 98 shifts, but whether shifts are identified at all. I also simulated range edge data to demonstrate
 99 that for realistic values of shift rates and interannual variability, many decades of data are likely
 100 to be needed to reliably quantify edge shifts.

101

102

Method	Description	Examples
Most distal point	Edge is identified at the most distal point observation or occupied grid cell along a single dimension (typically latitude)	(La Sorte & Thompson, III 2007; Schoeman <i>et al.</i> 2015)
Mean of distal points or grid cells	Mean of n most distal point observations or occupied grid cells along a single dimension (typically latitude)	(Hällfors <i>et al.</i> 2021; Hickling <i>et al.</i> 2006; Kerr <i>et al.</i> 2015)
Abundance-weighted mean of distal points or grid cells	Abundance-weighted mean of n most distal point observations or occupied grid cells along a single dimension (typically latitude)	(Perry <i>et al.</i> 2005)
Quantile of observations	Quantile of observations in a single dimension (usually latitude), often 0.01, 0.05, or 0.1 for equatorward edges, and 0.9, 0.95, or 0.99 for poleward edges	(Fredston-Hermann <i>et al.</i> 2020; Sittaro <i>et al.</i> 2017; Zhu <i>et al.</i> 2012)
Quantile of modeled distribution	Species distribution model is fitted to data, and then quantiles of presence or abundance are extracted from the models	(Fredston <i>et al.</i> 2021)

103 **Table 1.** Summary of common methods for measuring range edge positions, typically for use in
 104 studying climate-related range shifts. This does not represent an exhaustive review of the
 105 literature. Many other examples of these approaches exist, as do many other bespoke methods.

106 Methods

107 This manuscript uses two species as focal examples: *Urophycis tenuis* (white hake), a
108 marine fish whose distribution in the Northwest Atlantic includes an equatorward range edge
109 that falls within the Northeast US continental shelf, and *Coragyps atratus* (black vulture), a bird
110 species with a poleward range edge in the continental US. Both of these range edges—the
111 equatorward edge of the white hake and the poleward edge of the black vulture—have been
112 described as shifting northward by previous studies (Fredston-Hermann *et al.* 2020; La Sorte &
113 Thompson, III 2007).

114 Data on the white hake was downloaded from FISHGLOB, an international effort to
115 collate, harmonize, and publish scientific bottom trawl surveys (“FishGlob_data” 2023;
116 Maureaud *et al.* 2024). I used data from the Northeast US, where a biannual survey has been
117 operated since the 1960s by the National Oceanic and Atmospheric Administration, NOAA
118 (Azarovitz 1981; Despres-Patanjo *et al.* 1988). Consistent with other publications using this
119 dataset, I used data from the fall survey from 1972 onward (Fredston-Hermann *et al.* 2020;
120 Pinsky *et al.* 2013). Data on the black vulture was downloaded from the National Audubon
121 Society Christmas Bird Count, or CBC (National Audubon Society 2023). I downloaded
122 historical data on the black vulture from the CBC online data portal. Consistent with other
123 publications using this dataset, I used data from 1975 onward (La Sorte & Thompson, III 2007).

124 For each of these species, I calculated annual range edge positions using a subset of
125 the methods in Table 1: the three methods based on distal points, and quantiles of observations
126 (0.9, 0.95, and 0.99 for a poleward edge, and 0.01, 0.05, and 0.10 for an equatorward edge).
127 For the metrics based on a mean of points, I used $n=3$, i.e., a mean of the three most distal
128 points. I estimated whether edges had shifted over time using linear regression of edge position

129 (in degrees latitude) on year with a significance threshold of 0.05. P-values less than 0.001 are
130 reported as “< 0.001” in-text.

131 Regardless of which edge metric is chosen, researchers must also ensure that their
132 study has sufficient power to detect a “true trend”. I conducted a set of simulations to evaluate
133 which combinations of edge shift rates and time-series lengths have sufficient power to detect
134 significant shifts when they occur, using non-random resampling (White & Bahlai 2021). I
135 simulated the position of a range edge over a century given some shift rate and error:

$$136 \text{ edge position } _t = \text{shift rate} \times \text{year} + e$$

137 Where t represented year and e represented an error term. The error term was drawn
138 each year from a normal distribution with a mean of 0 and a standard deviation of 0.65, which
139 was derived from the mean conditional standard deviations of the linear regressions previously
140 fitted to white hake and black vulture data. Using realistic values from literature (Lenoir *et al.*
141 2020), I tested range edge shift rates from 0.01 – 0.1 °lat/year in 0.01 °lat/year steps.

142 I then subsetted the simulated data into different time-series lengths, as if an observer
143 were only able to record n continuous years of data, where n ranged from three to 100. For
144 each time-series length, I fitted linear regressions of position on time to every possible subset of
145 the simulated data. For example, for a time-series length of four years, I fitted a regression to
146 the simulated data from years one, two, three, and four; another to years two, three, four, and
147 five; another to years three, four, five, and six; and so on. Then for each time-series length, I
148 estimated its power as how often the regressions found the “correct” result (a significant positive
149 shift, with a significance threshold of 0.05). I repeated the workflow above 100 times to generate
150 a distribution of power values for every combination of time-series length and edge shift rate. In
151 interpreting the results, I considered a power value of 0.8 or higher to mean that the analysis
152 accurately captured the “true trend” of the simulated data.

153 All data analysis was conducted in R version 4.1.1 (R Core Team 2021). Code and data
154 to reproduce the results can be found online at [*GitHub link to be inserted after peer review*].

155 Results

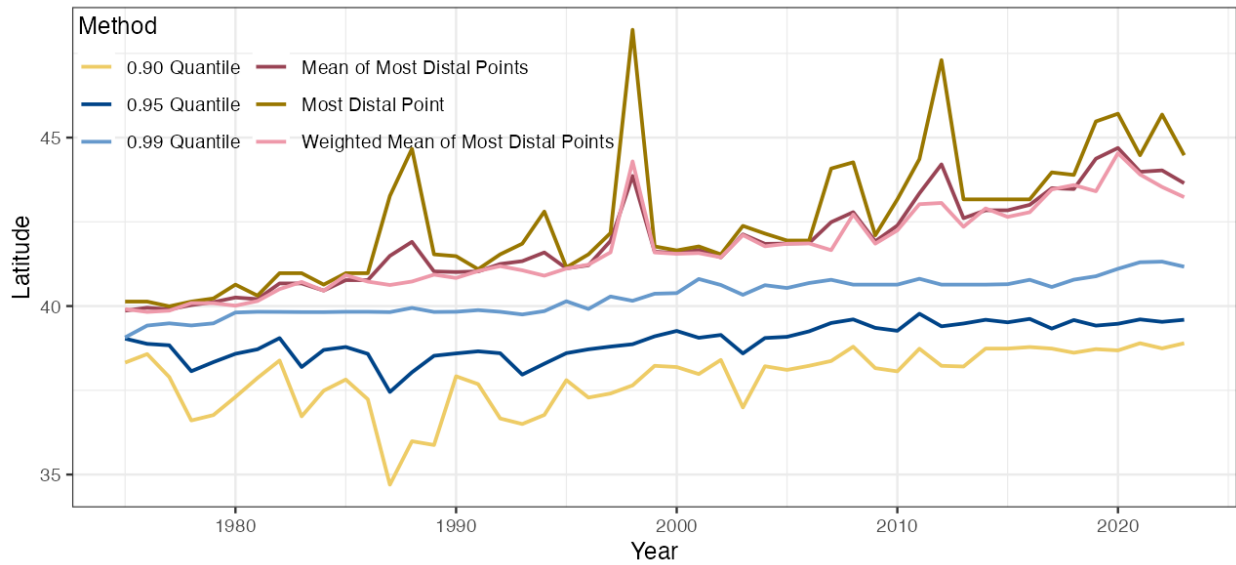
156 For both the poleward edge of the black vulture and the equatorward edge of the white
157 hake, different methods estimated dramatically different range edge positions—in some years,
158 more than five degrees of latitude apart (Figure 2). Interannual variability was also lower for
159 quantile-based approaches than for methods using most distal points (Figure 2).

160 All methods tested found a significant shift in the poleward range edge position of the
161 black vulture from 1975-2023 (Figure 2). However, the methods differed markedly in the
162 estimated shift rates. Shift rates using edge metrics that were based on most distal points were
163 all within one standard deviation of one another (most distal point: 0.097 ± 0.013 °lat/year, $p <$
164 0.001 ; mean of most distal points: 0.085 ± 0.0052 °lat/year, $p < 0.001$; abundance-weighted
165 mean of most distal points: 0.081 ± 0.0051 °lat/year, $p < 0.001$). By contrast, the shift rates
166 using quantile edge metrics were much slower, especially the 0.95 quantile method (0.9
167 quantile: 0.038 ± 0.0075 °lat/year, $p < 0.001$; 0.95 quantile: 0.028 ± 0.0035 °lat/year, $p < 0.001$;
168 0.99 quantile: 0.036 ± 0.0017 °lat/year, $p < 0.001$).

169 For the equatorward edge of the white hake, the three methods using abundance-
170 weighted quantiles found a significant northward shift in edge position from 1972-2019, and the
171 methods calculated using most distal points did not (Figure 2). The rates of shift estimated by
172 the three quantile methods all fell within one standard error of each other (0.1 quantile: $0.0086 \pm$
173 0.0036 °lat/year, $p = 0.02$; 0.05 quantile: 0.015 ± 0.0046 °lat/year, $p = 0.003$; 0.01 quantile:
174 0.016 ± 0.0054 °lat/year, $p = 0.004$).

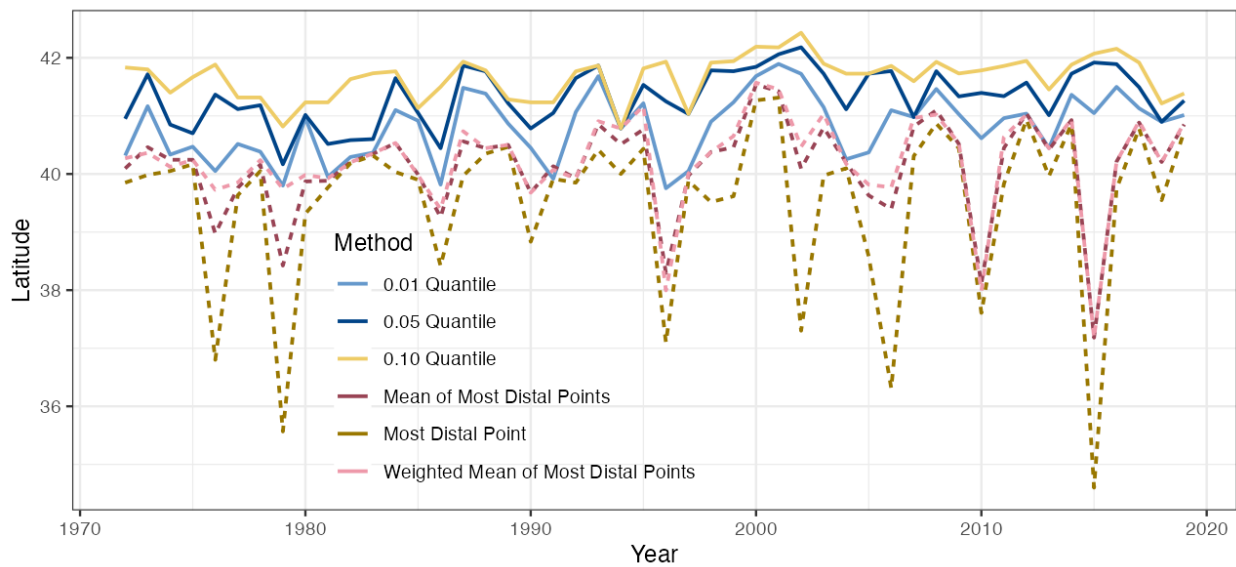
175

Coragyps atratus poleward edge



176

Urophycis tenuis equatorward edge



177

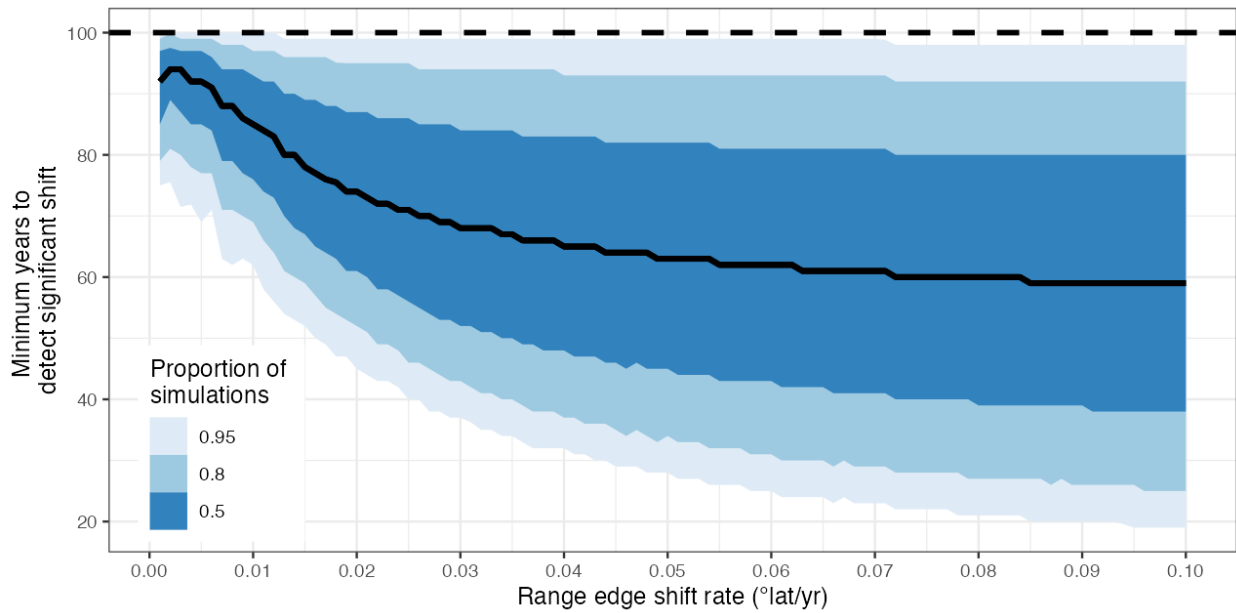
178 **Figure 2.** Positions of the poleward edge of the black vulture (*Coragyps atratus*) estimated from
 179 the Christmas Bird Count (top) and the equatorward edge of the white hake (*Urophycis tenuis*)
 180 in the Northwest Atlantic estimated from an annual NOAA survey (bottom) using various
 181 methods that are common in the biogeography literature. Solid lines indicate that a linear

182 regression of latitude on year found a significant shift ($p \leq 0.05$) over the time-series; dashed
183 lines indicate no significant shift.

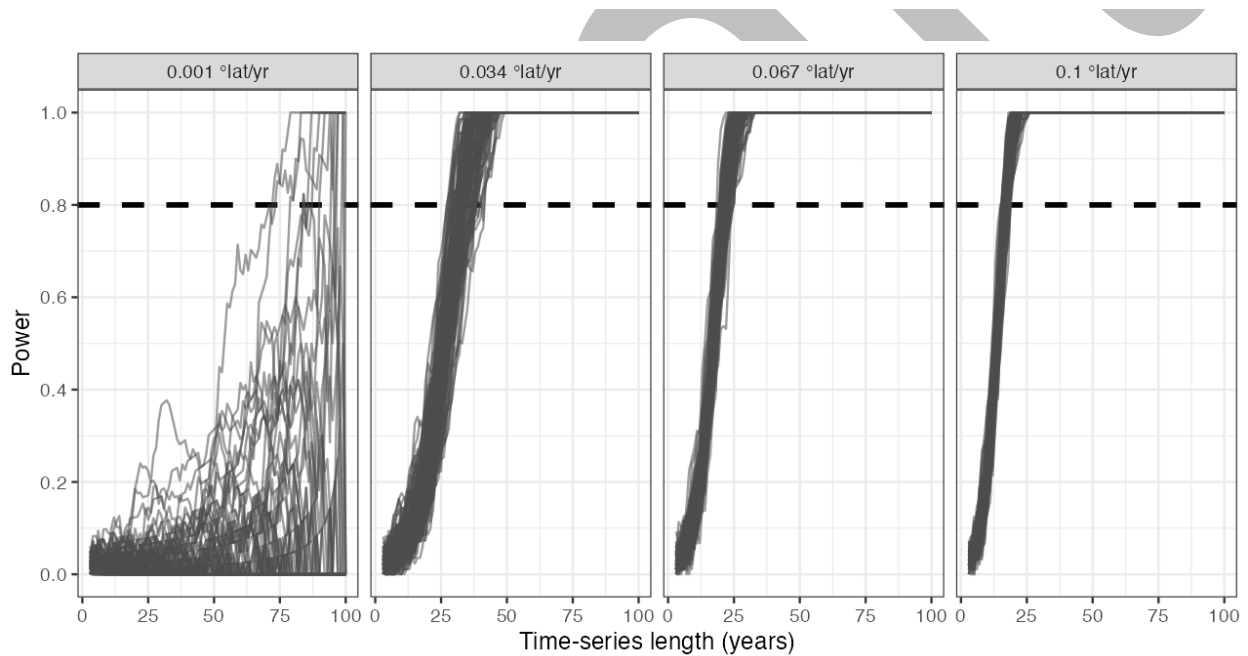
184 Using realistic time-series lengths, shift rates, and error terms, I simulated that many
185 decades of data would often be needed to reliably detect range edge shifts (Figure 3). The
186 median minimum time-series length required to accurately detect a significant shift ranged from
187 94 years (in the third-slowest range edge shift rate tested, $0.003^\circ\text{lat}/\text{year}$) to 59 years (for edge
188 shift rates of $0.085 - 0.1^\circ\text{lat}/\text{year}$). Across all time-series subsets and simulation iterations, the
189 shortest time-series to reach the statistical power threshold of 0.8 was 14 years long, in a
190 simulation where the range edge shift rate was $0.099^\circ\text{lat}/\text{year}$ —the second-fastest shift rate I
191 tested. Conversely, for the slowest rates of range edge shifts tested ($0.001 - 0.011^\circ\text{lat}/\text{year}$), at
192 least one iteration of the simulation never reached the statistical power threshold of 0.8 after
193 100 years (Figure 3).

194

195



196



197

198 **Figure 3.** Results of a non-random subsampling analysis exploring statistical power in analyses
199 of simulated range edge data, showing the number of years required to accurately identify
200 significant shifts in the simulation, i.e., when the simulated analysis reached a power level of 0.8
201 (top) and the relationship between time-series length and statistical power for four example
202 rates of range edge shift (bottom). The dashed line represents the end of the simulation after

203 100 years (top) and the power threshold of 0.8 (bottom). Both plots show results for simulations
204 run 100 times. In the top plot, the color fill represents what proportion of those iterations falls
205 within each band, and the black line is the median value. In the bottom plot, each line is an
206 iteration of the simulation.

207 Discussion

208 Here I used two prominent long-term monitoring datasets—each of which encompasses
209 decades of warming and has been previously found to reveal climate-related range edge shifts
210 (Fredston-Hermann *et al.* 2020; La Sorte & Thompson, III 2007)—to show that estimated edge
211 positions, shift rates, and statistical significance depend on methodological choices by
212 researchers (Figure 2). Quantile-based methods showed less interannual variability in edge
213 position than distal points-based methods, and different approaches estimated notably disparate
214 edge positions—sometimes differing by up to five degrees of latitude (Figure 2). For the
215 equatorward edge species, only quantile-based methods found a significant northward shift.
216 While all methods found a significant northward shift in the poleward edge species, their
217 estimated magnitudes were very different (Figure 2).

218 It is generally understood that individual researcher decisions affect science (Gould *et al.*
219 2023), and previous work in global change biology found that methodological decisions explain
220 a large proportion of variation in estimated range shifts (Brown *et al.* 2016; Lenoir *et al.* 2020).
221 This issue is especially salient for range edges, which are extreme value statistics that may be
222 highly sensitive to seemingly small differences in methodology. The results presented here only
223 represent a subset of common approaches to estimating range edge positions (Table 1), and
224 many other approaches exist in literature. Thus, the effect of *ad hoc* methods on range edge
225 measurements and estimated shift rates may be even greater than demonstrated here with a

226 handful of approaches. Observer error, low and/or variable probability of detection, and changes
227 in sampling over time can further confound range edge measurements.

228 Studies often pool results about poleward and equatorward edge shifts as if they were
229 comparable processes. Especially when using presence-only metrics, this approach may have
230 the unintended consequence of biasing results toward recording range edge extensions over
231 range edge retractions. This is because in order for a range edge retraction to be found using
232 presence-only data, every individual in a population has to vanish from the most distal site(s);
233 but for a range edge extension to occur just one individual has to spread beyond the historical
234 range (Thomas *et al.* 2006). Indeed, a recent synthesis of evidence for range shifts revealed
235 that poleward edge shifts are much more commonly recorded than equatorward edge shifts
236 (Lawlor *et al.* 2024). Whether this phenomenon is considered a “feature” or a “bug” in analysis
237 depends on the research question (and note that range edge extensions can also be difficult to
238 detect if populations are small and/or monitoring effort is low; see Robinson *et al.* (2015)).
239 Researchers aiming to identify the full geographical area occupied by a species—i.e., a range
240 map—may prefer to use presence-only metrics for precisely the reason that they capture the
241 most distal occupied sites. However, if the goal is to calculate rates of range edge shifts or to
242 understand where the vast majority of a species’ density is found, abundance-weighted metrics
243 may provide a more accurate picture of equatorward edge dynamics. Alternatively, researchers
244 may want to combine these metrics in a single study to capture both distal colonization or
245 extinction events and changing abundance patterns.

246 Even when range edges are measured with abundance-weighted metrics, it must be
247 emphasized that colonization and extirpation are distinct processes that happen at very different
248 rates. The relatively slow rate of extirpation, which may cause “extinction debt”, means that
249 equatorward-edge populations could be declining even if their geographical shifts are slower
250 than at the poleward edge (Hampe & Petit 2005; Jackson & Sax 2010; Tilman *et al.* 1994). This

251 is evident in the white hake equatorward edge example: the statistically significant shift rates I
252 calculated were less than 2 km/year, which is far slower than literature syntheses on marine
253 range shift rates overall (Lenoir *et al.* 2020; Poloczanska *et al.* 2013). It may be that measuring
254 demographic parameters in range edge populations is a more reliable predictor of their
255 trajectories than abundance (Gaston *et al.* 2009). Researchers may also consider more
256 biologically relevant metrics that exclude adult migration and foraging from range measurement.
257 For example, range expansion in California sea otters has been modeled using only abundance
258 data on females and pups (Krkošek *et al.* 2007), since males are mobile foragers and their
259 geographical distributions are not reflective of the species' year-round range.

260 This study focused on geographical range edges measured primarily in one dimension:
261 latitude. In the context of global change, species are expected to track climate velocities, which
262 do not always point north (Burrows *et al.* 2014). Range edge dynamics in response to climate
263 change may be better captured by measuring shifts relative to these local climate velocities
264 (Pinsky *et al.* 2013), or along custom “rulers” that correspond to coastlines, isotherms, contours,
265 or other features that range-shifting species might follow (Fredston *et al.* 2021). While not
266 discussed here, depth and elevational range edges are typically studied in an *ad hoc* fashion
267 similar to geographical range edges, and their reported positions and shifts may be similarly
268 sensitive to researchers' methodological choices.

269 Methods matter in measuring range edge dynamics, but so does statistical power. Non-
270 random resampling of time-series datasets can reveal how much data is necessary to detect a
271 “true trend” (White 2019; White & Bahlai 2021). The simulation results presented here
272 underscore the long time-series that may be needed to reliably detect the signal of a range edge
273 shift (Figure 3). This analysis was not intended to provide quantitative rules of thumb, and did
274 not explore the effects of different range edge metrics on estimated shift rates and statistical
275 significance. Rather, it is the type of simulation that could be conducted by researchers

276 analyzing novel empirical data to test whether their study is underpowered. By adapting the shift
277 rate, time-series length, and error rate parameters to their own study systems, researchers can
278 simulate how likely they are to detect a significant shift should one exist. The example values I
279 used suggest that decades of data may be needed to reliably detect modern range edge shifts
280 (Figure 3).

281 While the exact threshold for statistical power varies by system and study, the
282 implication remains that studies with just a few time points are unlikely to reflect true rates of
283 edge shifts (Stuble *et al.* 2021). Publishing under-powered interpretations of ecological time-
284 series is not uncommon, despite statistical evidence that doing so may lead to misleading
285 inferences about trajectories of populations and ecosystems with complex dynamics (Bahlai *et*
286 *al.* 2021; McCain *et al.* 2016; White 2019). With high interannual variability being a common
287 feature of range edges, and few large-scale, long-term, and high-resolution datasets such as
288 the CBC and NOAA surveys available (Thomas *et al.* 2006), it remains an open question if and
289 how the field is ready to accurately quantify range edge dynamics.

290 The simulation presented here did not address additional sources of uncertainty, such as
291 observer error and the abundance and detectability of a species. These can be quantified and
292 partitioned from biological processes using species distribution models (Brodie *et al.* 2022).
293 Another advantage of using species distribution models is that they can quantify uncertainty
294 associated with range edge positions (Fredston *et al.* 2021). However, the degree to which
295 species distribution models accurately predict biological parameters (e.g., presence,
296 abundance) is widely debated (Lee-Yaw *et al.* 2022), and their usefulness in projecting future
297 species distributions—one primary focus of modern biogeography—is unclear (Fourcade *et al.*
298 2018; Sofaer *et al.* 2018). Mechanistic models may provide an alternative path toward
299 forecasting range edge dynamics while accounting for observation error (Briscoe *et al.* 2019).

300 Syntheses of edge shift rates across studies, systems, and taxa is facilitated by
301 researchers sharing their raw data, which allows future re-analysis or meta-analysis with
302 different methods and motivating questions (Carpenter *et al.* 2009). Ecological synthesis is
303 especially important in the context of global change biology, because the field-collected
304 datasets researchers use to study biogeography in a warming world are each irreplaceable
305 (Wolkovich *et al.* 2012). In addition to open data, open code will make it possible to investigate
306 the effect of methodological choices on reported measurements of range edges. Openly sharing
307 data and code has many other advantages that are well-described (Cheruvilil & Soranno 2018;
308 Fredston & Lowndes 2024; McKiernan *et al.* 2016; O’Dea *et al.* 2021).

309 Species range edges are a core concept in biogeography, and range edge shifts are
310 among the most common metrics of “species on the move” in a warming world (Lenoir *et al.*
311 2020). Predicted future range edge shifts are already influential in biodiversity conservation and
312 natural resource management, as models project when species will shift out of protected areas
313 or across jurisdictional boundaries (Pinsky *et al.* 2018). Yet range edges lack a universally
314 accepted measurement, and the *ad hoc* approach taken by global change studies to date
315 makes it impossible to accurately synthesize—let alone forecast—the true rates of range edge
316 shifts in the Anthropocene. Improving measurements of range edge positions and shifts would
317 enable global change researchers to better describe changes that are occurring and to more
318 effectively relate them to important environmental variables. Several recommended best
319 practices for quantifying range edge dynamics emerge from this study:

- 320 1. Range edge metrics should be chosen to match a research question, and care
321 should be taken when comparing poleward and equatorward edge dynamics—
322 especially when using presence-based approaches. Sensitivity analyses can
323 reveal whether results are driven by the choice of metric.

- 324 2. Researchers should conduct simulations to ensure that their studies are not
325 under-powered, especially when testing for shifts in range edge positions.
- 326 3. Demographically relevant parameters, such as population growth rates or
327 abundance of reproductive females, may lead to more informative metrics of
328 range edge dynamics than simple inventories of presence or abundance.
- 329 4. Publishing raw data and code means that range edge studies can be re-analyzed
330 in the future to assess the effect of methodological choices on results and to
331 conduct syntheses and meta-analyses about range edge dynamics.

332 The methods that I tested here (Table 1) are simple statistics, and it is likely that none of
333 them are an ideal metric for range edges. Far more sophisticated approaches exist that could
334 be applied to measuring the boundaries of species' ranges. For example, kernel densities,
335 contour estimation, and two- and three-dimensional range edge metrics all merit much greater
336 attention in the biogeography and global change biology fields. How to measure other
337 foundational properties of species—e.g., abundance, growth rates, population structure—has
338 received extensive research attention in quantitative ecology for many years. By contrast,
339 biogeography has yet to grapple with to what degree our understanding of range edge dynamics
340 is driven by methods rather than biological processes, let alone to move toward universally
341 accepted metrics. It is time for range edges to get their due.

342 References

- 343 Azarovitz, T.R. (1981). A brief historical review of the Woods Hole Laboratory trawl survey time
344 series. *Can. Spec. Publ. Fish. Aquat. Sci.*, 58, 62–7.
- 345 Bahlai, C.A., White, E.R., Perrone, J.D., Cusser, S. & Stack Whitney, K. (2021). The broken
346 window: An algorithm for quantifying and characterizing misleading trajectories in
347 ecological processes. *Ecol. Inform.*, 64, 101336.
- 348 Boisvert-Marsh, L., Périé, C. & de Blois, S. (2014). Shifting with climate? Evidence for recent
349 changes in tree species distribution at high latitudes. *Ecosphere*, 5, 1–33.
- 350 Briscoe, N.J., Elith, J., Salguero-Gómez, R., Lahoz-Monfort, J.J., Camac, J.S., Giljohann, K.M.,
351 *et al.* (2019). Forecasting species range dynamics with process-explicit models:

352 matching methods to applications. *Ecol. Lett.*, 22, 1940–1956.

353 Brodie, S., Smith, J.A., Muhling, B.A., Barnett, L.A.K., Carroll, G., Fiedler, P., *et al.* (2022).

354 Recommendations for quantifying and reducing uncertainty in climate projections of

355 species distributions. *Glob. Change Biol.*, 28, 6586–6601.

356 Brown, C.J., O'Connor, M.I., Poloczanska, E.S., Schoeman, D.S., Buckley, L.B., Burrows, M.T.,

357 *et al.* (2016). Ecological and methodological drivers of species' distribution and

358 phenology responses to climate change. *Glob. Change Biol.*, 22, 1548–1560.

359 Burrows, M.T., Schoeman, D.S., Richardson, A.J., Molinos, J.G., Hoffmann, A., Buckley, L.B., *et*

360 *al.* (2014). Geographical limits to species-range shifts are suggested by climate velocity.

361 *Nature*, 507, 492–495.

362 Carpenter, S.R., Armbrust, E.V., Arzberger, P.W., Chapin, F.S., III, Elser, J.J., Hackett, E.J., *et*

363 *al.* (2009). Accelerate Synthesis in Ecology and Environmental Sciences. *BioScience*,

364 59, 699–701.

365 Cheruvilil, K.S. & Soranno, P.A. (2018). Data-Intensive Ecological Research Is Catalyzed by

366 Open Science and Team Science. *BioScience*, 68, 813–822.

367 Despres-Patanjo, L.I., Azarovitz, T.R. & Byrne, C.J. (1988). Twenty-five Years of Fish Surveys

368 in the Northwest Atlantic: The NMFS Northeast Fisheries Center's Bottom Trawl Survey

369 Program. *Mar. Fish. Rev.*, 50, 69–71.

370 FishGlob_data. (2023). .

371 Fourcade, Y., Besnard, A.G. & Secondi, J. (2018). Paintings predict the distribution of species,

372 or the challenge of selecting environmental predictors and evaluation statistics. *Glob.*

373 *Ecol. Biogeogr.*, 27, 245–256.

374 Fredston, A., Pinsky, M., Selden, R.L., Szuwalski, C., Thorson, J.T., Gaines, S.D., *et al.* (2021).

375 Range edges of North American marine species are tracking temperature over decades.

376 *Glob. Change Biol.*, 27, 3145–3156.

377 Fredston, A.L. & Lowndes, J.S.S. (2024). Welcoming More Participation in Open Data Science

378 for the Oceans. *Annu. Rev. Mar. Sci.*, 16, 537–549.

379 Fredston-Hermann, A., Selden, R., Pinsky, M., Gaines, S.D. & Halpern, B.S. (2020). Cold range

380 edges of marine fishes track climate change better than warm edges. *Glob. Change*

381 *Biol.*, 26, 2908–2922.

382 Gaston, K.J. (2000). Global patterns in biodiversity. *Nature*, 405, 220–227.

383 Gaston, K.J. (2003). *The structure and dynamics of geographic ranges*. Oxford series in ecology

384 and evolution. University Press, Oxford.

385 Gaston, K.J., Chown, S.L., Calosi, P., Bernardo, J., Bilton, D.T., Clarke, A., *et al.* (2009).

386 Macrophysiology: A Conceptual Reunification. *Am. Nat.*, 174, 595–612.

387 Gould, E., Fraser, H.S., Parker, T.H., Nakagawa, S., Griffith, S.C., Vesk, P.A., *et al.* (2023).

388 Same data, different analysts: variation in effect sizes due to analytical decisions in

389 ecology and evolutionary biology.

390 Guisan, A. & Thuiller, W. (2005). Predicting species distribution: offering more than simple

391 habitat models. *Ecol. Lett.*, 8, 993–1009.

392 Hällfors, M.H., Pöyry, J., Heijölä, J., Kohonen, I., Kuussaari, M., Leinonen, R., *et al.* (2021).

393 Combining range and phenology shifts offers a winning strategy for boreal Lepidoptera.

394 *Ecol. Lett.*, 24, 1619–1632.

395 Hampe, A. & Petit, R.J. (2005). Conserving biodiversity under climate change: the rear edge

396 matters. *Ecol. Lett.*, 8, 461–467.

397 Hickling, R., Roy, D.B., Hill, J.K., Fox, R. & Thomas, C.D. (2006). The distributions of a wide

398 range of taxonomic groups are expanding polewards. *Glob. Change Biol.*, 12, 450–455.

399 Hughes, A.C., Dorey, J.B., Bossert, S., Qiao, H. & Orr, M.C. (n.d.). Big data, big problems? How

400 to circumvent problems in biodiversity mapping and ensure meaningful results.

401 *Ecography*, n/a, e07115.

402 Jackson, S.T. & Sax, D.F. (2010). Balancing biodiversity in a changing environment: extinction
403 debt, immigration credit and species turnover. *Trends Ecol. Evol.*, 25, 153–160.

404 Kerr, J.T., Pindar, A., Galpern, P., Packer, L., Potts, S.G., Roberts, S.M., *et al.* (2015). Climate
405 change impacts on bumblebees converge across continents. *Science*, 349, 177–180.

406 Krkošek, M., Lauzon-Guay, J.-S. & Lewis, M.A. (2007). Relating dispersal and range expansion
407 of California sea otters. *Theor. Popul. Biol.*, 71, 401–407.

408 La Sorte, F.A. & Thompson, III, F.R. (2007). Poleward Shifts in Winter Ranges of North
409 American Birds. *Ecology*, 88, 1803–1812.

410 Lawlor, J.A., Comte, L., Grenouillet, G., Lenoir, J., Baecher, J.A., Bandara, R.M.W.J., *et al.*
411 (2024). Mechanisms, detection and impacts of species redistributions under climate
412 change. *Nat. Rev. Earth Environ.*, 5, 351–368.

413 Lee-Yaw, J.A., L. McCune, J., Pironon, S. & N. Sheth, S. (2022). Species distribution models
414 rarely predict the biology of real populations. *Ecography*, 2022, e05877.

415 Lenoir, J., Bertrand, R., Comte, L., Bourgeaud, L., Hattab, T., Muriene, J., *et al.* (2020).
416 Species better track climate warming in the oceans than on land. *Nat. Ecol. Evol.*, 1–16.

417 MacArthur, R.H. (1972). *Geographical Ecology: Patterns in the Distribution of Species*.
418 Princeton University Press.

419 Maureaud, A.A., Palacios-Abrantes, J., Kitchel, Z., Mannocci, L., Pinsky, M.L., Fredston, A., *et*
420 *al.* (2024). FISHGLOB_data: an integrated dataset of fish biodiversity sampled with
421 scientific bottom-trawl surveys. *Sci. Data*, 11, 24.

422 McCain, C., Szewczyk, T. & Knight, K.B. (2016). Population variability complicates the accurate
423 detection of climate change responses. *Glob. Change Biol.*, 22, 2081–2093.

424 McKiernan, E.C., Bourne, P.E., Brown, C.T., Buck, S., Kenall, A., Lin, J., *et al.* (2016). How
425 open science helps researchers succeed. *eLife*, 5, e16800.

426 National Audubon Society. (2023). The Christmas Bird Count Historical Results [Online].

427 O’Dea, R.E., Parker, T.H., Chee, Y.E., Culina, A., Drobniak, S.M., Duncan, D.H., *et al.* (2021).
428 Towards open, reliable, and transparent ecology and evolutionary biology. *BMC Biol.*,
429 19, 68.

430 Parmesan, C. & Yohe, G. (2003). A globally coherent fingerprint of climate change impacts
431 across natural systems. *Nature*, 421, 37–42.

432 Pecl, G.T., Araújo, M.B., Bell, J.D., Blanchard, J., Bonebrake, T.C., Chen, I.-C., *et al.* (2017).
433 Biodiversity redistribution under climate change: Impacts on ecosystems and human
434 well-being. *Science*, 355, eaai9214.

435 Perry, A.L., Low, P.J., Ellis, J.R. & Reynolds, J.D. (2005). Climate Change and Distribution
436 Shifts in Marine Fishes. *Science*, 308, 1912–1915.

437 Pinsky, M.L., Reygondeau, G., Caddell, R., Palacios-Abrantes, J., Spijkers, J. & Cheung,
438 W.W.L. (2018). Preparing ocean governance for species on the move. *Science*, 360,
439 1189–1191.

440 Pinsky, M.L., Worm, B., Fogarty, M.J., Sarmiento, J.L. & Levin, S.A. (2013). Marine taxa track
441 local climate velocities. *Science*, 341, 1239–1242.

442 Poloczanska, E.S., Brown, C.J., Sydeman, W.J., Kiessling, W., Schoeman, D.S., Moore, P.J., *et*
443 *al.* (2013). Global imprint of climate change on marine life. *Nat. Clim. Change*, 3, 919–
444 925.

445 R Core Team. (2021). R: A language and environment for statistical computing.

446 Robinson, L.M., Gledhill, D.C., Moltschanivskyj, N.A., Hobday, A.J., Frusher, S., N.Barrett, *et al.*
447 (2015). Rapid assessment of an ocean warming hotspot reveals “high” confidence in
448 potential species’ range extensions. *Glob. Environ. Change*, 31, 28–37.

449 Schoeman, D.S., Schlacher, T.A., Jones, A.R., Murray, A., Huijbers, C.M., Olds, A.D., *et al.*
450 (2015). Edging along a Warming Coast: A Range Extension for a Common Sandy Beach
451 Crab. *PLoS ONE*, 10, e0141976.

452 Sexton, J.P., McIntyre, P.J., Angert, A.L. & Rice, K.J. (2009). Evolution and Ecology of Species
453 Range Limits. *Annu. Rev. Ecol. Evol. Syst.*, 40, 415–436.

454 Sittaro, F., Paquette, A., Messier, C. & Nock, C.A. (2017). Tree range expansion in eastern
455 North America fails to keep pace with climate warming at northern range limits. *Glob.*
456 *Change Biol.*, 23, 3292–3301.

457 Sofaer, H.R., Jarnevich, C.S. & Flather, C.H. (2018). Misleading prioritizations from modelling
458 range shifts under climate change. *Glob. Ecol. Biogeogr.*, 27, 658–666.

459 Stuble, K.L., Bewick, S., Fisher, M., Forister, M.L., Harrison, S.P., Shapiro, A.M., *et al.* (2021).
460 The promise and the perils of resurveying to understand global change impacts. *Ecol.*
461 *Monogr.*, 91.

462 Thomas, C.D., Franco, A.M.A. & Hill, J.K. (2006). Range retractions and extinction in the face of
463 climate warming. *Trends Ecol. Evol.*, 21, 415–416.

464 Tilman, D., May, R.M., Lehman, C.L. & Nowak, M.A. (1994). Habitat destruction and the
465 extinction debt. *Nature*, 371, 65–66.

466 Wallace, A.R. (1876). *The Geographical Distribution of Animals: With a Study of the Relations of*
467 *Living and Extinct Faunas as Elucidating the Past Changes of the Earth's Surface.*
468 Cambridge University Press, Cambridge.

469 White, E.R. (2019). Minimum Time Required to Detect Population Trends: The Need for Long-
470 Term Monitoring Programs. *BioScience*, 69, 40–46.

471 White, E.R. & Bahlai, C.A. (2021). Experimenting With the Past to Improve Environmental
472 Monitoring. *Front. Ecol. Evol.*, 8.

473 Willi, Y. & Van Buskirk, J. (2019). A Practical Guide to the Study of Distribution Limits. *Am. Nat.*,
474 193, 773–785.

475 Williams, J.E. & Blois, J.L. (2018). Range shifts in response to past and future climate change:
476 Can climate velocities and species' dispersal capabilities explain variation in mammalian
477 range shifts? *J. Biogeogr.*, 45, 2175–2189.

478 Wolkovich, E.M., Regetz, J. & O'Connor, M.I. (2012). Advances in global change research
479 require open science by individual researchers. *Glob. Change Biol.*, 18, 2102–2110.

480 Zhu, K., Woodall, C.W. & Clark, J.S. (2012). Failure to migrate: lack of tree range expansion in
481 response to climate change. *Glob. Change Biol.*, 18, 1042–1052.