Topoclimate buffers floristic diversity from macroclimate in temperate mountain forests.

3 Jeremy Borderieux¹, Emiel De Lombaerde², Karen De Pauw², Pieter Sanczuk², Pieter

4 Vangansbeke², Thomas Vanneste², Pieter De Frenne², Jean-Claude Gégout¹, Josep M. Serra-5 $Diaz$ ^{$1,3$}.

- 1. Université de Lorraine, AgroParisTech, INRAE, UMR Silva, 54000 Nancy, France
- 2. Forest & Nature Lab, Department of Environment, Ghent University, Geraardsbergsesteenweg 267, 9090 Gontrode, Belgium
- 3. Eversource Energy Center and Department of Ecology and Evolutionary Biology, University of Connecticut, Storrs, CT, United States of America
- *Orcid ID:*
- Jeremy Borderieux : 0000-0003-3993-1067
- Emiel De Lombaerde : 0000-0002-0050-2735
- Karen De Pauw : 0000-0001-8369-2679
- Pieter Sanczuk : 0000-0003-1107-4905
- Pieter Vangansbeke : 0000-0002-6356-2858
- Thomas Vanneste : [0000-0001-5296-917X](https://orcid.org/0000-0001-5296-917X)
- Pieter De Frenne : 0000-0002-8613-0943
- Jean-Claude Gégout : 0000-0002-5760-9920
- Josep M. Serra-Diaz: 0000-0003-1988-1154
- Corresponding author: Jeremy Borderieux: jeremy.borderieux@agroparistech.fr

Abstract

 Microclimates strongly influence the composition and diversity of forest plant communities. Recent studies have highlighted the role of tree canopies in shaping understory 26 thermal conditions at small spatial scales, especially in lowland forests. In mountain forests, however, the influence of topography in environmental conditions (e.g. topoclimate) is ought to also influence plants' perceived temperature. Understanding how topography and canopies interactively affect understory temperature is key to identifying stable refugia that could shelter cold-adapted forest specialist species under climate change.

 Here we report on growing season understory temperatures using 48 loggers in contrasting topographic features of a mid-range mountain valley spanning from 475 m.a.s.l. to 1203 m.a.s.l. in the Vosges Mountains (NE France). We disentangle the relative importance and the effects of topography vs. canopies in determining local temperatures. We then evaluate how topography and canopy-induced variation in temperature drive plant community composition and richness in 306 floristic surveys distributed across the studied mountain valley.

 Our results show that topography outweighed canopy cover in explaining growing season understory temperatures. Regardless of canopy, the daily mean temperature of the 40 growing season in south-facing ridges was 1.5 °C (CI: \pm 0.88 °C) warmer than shaded valley 41 bottoms, while dense canopies cooled temperatures by 0.5 °C (CI: \pm 0.48 °C) compared to open canopies. Topoclimate explained community composition as much as elevation and was the only significant predictor of species richness. Cold topoclimates harbored 30% more species than the average species richness across our plots. This increase in species richness was explained by an increase of cold-adapted species, both forest specialist and generalist species.

 Our findings highlight a stronger role of topography compared to canopy cover on community composition in mountain forests via topoclimatic cooling of north-facing slopes and valley bottoms. The importance of topographic features to explain temperature cooling and diversity underpins their role as present and future microrefugia.

Keywords

 Community ecology, forest, topoclimate, microclimate, topography, temperature, climatic refugia, diversity, understory vegetation.

1.Introduction

 The study of topography influences on vegetation has fascinated ecologists for more than 150 years (Johnston et al., 1848), and has further gained relevance in the context of 58 the 21st century climate warming (Ashcroft, 2010; Dobrowski, 2011; IPCC, 2021; Lenoir *et al.*, 2017). Species distribution and climatic conditions are often modeled at a coarse resolution (typically 1 km or coarser), and thereby fail to capture local variation of climate at fine grains (Franklin *et al.*, 2013) : for instance, the topoclimate shaped by topography and the microclimate shaped by forest canopy (Bramer *et al.*, 2018; De Frenne *et al.*, 2021; Kemppinen *et al.*, 2023). Given that these factors can attenuate warm macroclimate temperatures, the study of the effects and interactions between topography and forest canopy are key to identify areas of climate stability in a warmer future (Ashcroft, 2010; De Frenne *et al.*, 2021; Haesen *et al.*, 2023; Hannah *et al.*, 2014).

 Variations in aspect can create contrasting topoclimates as slopes oriented to the equator receive more solar radiation. As a result, southwest-facing slopes in northern hemisphere mountains display warmer mean temperatures, longer growing seasons and shorter snow cover durations (Ashcroft *et al.*, 2008; Davis *et al.*, 2019; Rita *et al.*, 2021). The physical properties of air also interact with topographic features such as hydrological basins (McLaughlin *et al.*, 2017), valley bottoms and sinks, and create local areas of cold and dense air pooling that decouple local conditions from the regional climate (Gudiksen *et al.*, 1992; Pastore *et al.*, 2022). , thus creating topographic refugia (Dobrowski, 2011). The topoclimate created by these terrain features interacts with the microclimate induced by forest canopies and jointly determines the understory temperature experienced by forest organisms. Canopy shading and evapotranspiration lead to an overall decrease of temperature throughout the year, compounded by a buffering of high summer temperatures compared to open-air, and an increase in winter temperatures due to insulation, (De Frenne *et al.*, 2021; Zellweger, Coomes, *et al.*, 2019). These buffering effects are apparent and well documented in temperate lowland forests, but their relative importance in contrast to 82 elevation and topography is less known, and current evidence has not reached consensus (Macek *et al.*, 2019; Vandewiele *et al.*, 2023).

 Canopy cover cooling of understory temperature has strong effects on forest communities. This is evidenced by the increases in the average thermal optimum of the species present (a proxy of species' affinity to climate) in forests where tree canopy was removed (De Frenne *et al.*, 2013; Dietz *et al.*, 2020; Richard *et al.*, 2021) and where warmer understory temperatures are predicted (Zellweger *et al.*, 2020). This sheltering of cold-89 adapted species by a dense canopy needs to be compared with the sheltering provided by topography in mountain forests, as topographical refugia are likely to offer longer-term buffering of temperature. Topographic refugia also harbor cold-adapted flora and host 92 populations of species outside their expected climatic range (Ellis & Eaton, 2021; Finocchiaro *et al.*, 2023; Macek *et al.*, 2019). In addition, understanding the characteristics of the

 sheltered species can also bring new insights, an increase of forest generalists for example demonstrates that topoclimate can mimic understory conditions of dense forests.

 Here we assessed the effects and relative importance of elevation, topography and canopy cover on *in situ* measured understory temperatures and plant community composition and richness. After accounting for the elevation gradient, we specifically asked: (1) Does topography (aspect and topographic position) outweigh canopy in explaining understory temperature, (2) does topography and canopy-induced variation in microclimate determine community richness and mean species thermal optimum? (3) Are plant habitat preference and climatic affinity related to specific temperatures?

2.Materials and Methods

2.1. Study Area

 Our study region (221 km²) is delineated by the basin of the Thur River, located in one of the southmost valleys of the Vosges Mountain range in France [\(Figure 1\)](#page-4-0). The Vosges 107 are characterized by a continental climate with harsh winters and short and stormy summers. 108 Its mean annual temperature ranges from 6 °C to 10 °C and precipitation ranges from 800 to 109 2,000 mm year⁻¹ (period 1970-2000, Météo France weather stations IGN, 2013). The Thur River basin is on the warm and dry end gradient of the Vosges Mountains (IGN, 2013). Forests cover 76% of the Vosges, which transitions from mixed oak stands and monospecific *Picea abies* stands to mixtures of *Picea abies*, *Abies alba* and *Fagus sylvatica* as elevation increases IGN, 2013). The soil of our study region is mostly shallow loam and sand with coarse elements. The most acidic soils are found at higher altitudes because of the dominance of needles in the humus and the lower temperature at mountaintops (IGN, 2013; Piqué *et al.*, 1994; Thomas *et al.*, 1999). The topography is highly variable, with an elevation ranging from 400 to 1424 m.a.s.l (but forest occurrence stops past 1250 m.a.s.l) with high topographic heterogeneity [\(Figure 1\)](#page-4-0).

 Figure 1: Map of the study area (black outline) with the location of the temperature loggers (white circles) and the floristic surveys (black crosses). The colored scale represents elevation above sea level, in meters, obtained from a 25-m spatial resolution digital elevation model. Hillshade effects have been added to visualize the terrain. The inset shows the Vosges Mountain range (grey) and the location of the studied valley (black point) in western Europe.

2.2. Temperature Predictors

 We used 25-meter resolution digital elevation models to extract elevation (m.a.s.l.), slope and aspect and to calculate topographical indices, heat load and topographic position (IGN, 2017). We handled spatial data with the '*raster*' and '*sf*' packages (Hijmans, 2020; Pebesma, 2018), and all the later analyses were carried on with R.4.2.2 (R Core Team, 2019). We used '*ggplot2'* and '*ggspatial'* packages for data visualization (Dunnington & Thorne, 2020; Wickham, 2011). We computed the heat load index (McCune & Keon, 2002) using the 'spatialEco' R package (Evans & Murphy, 2021). The heat load index ranges from 0 to 1 (least to most incoming solar radiation) contingent on the slope orientation and shading from nearby topographic features. The topographic position index is the relative position of the 137 cell in the shortest trajectory between a ridge and a drainage basin end, ranging from 0 (valley bottom) to 1 (ridge, Piedallu *et al.*, 2023).

 We obtained the 'tree cover density' from the 2018 product of the Copernicus monitoring service as a proxy for local canopy closure (Copernicus, 2018; Sannier *et al.*, 2023). This product consists of a 10-meter resolution percentage of canopy presence within the pixel (ranging from 0 to 100%) and was successfully used before to model microclimate buffering by canopy (Haesen *et al.*, 2021). This product was correlated with our field measurements of canopy closure (see below, 2.3 Temperature sampling). We rescaled this product to match the 25-m resolution of our other maps using bilinear interpolation (Hijmans, 2020). We rasterized (25-meter resolution) a 20-meter precision map of French forest to create a mask of the forested area of our study region, in order to limit our analysis and temperature projection to the forest of the region (IGN, 2019).

2.3. Temperature Sampling

 We created a stratified sampling scheme to capture forest understory microclimate variability (Lembrechts *et al.*, 2021; Schweiger *et al.*, 2016). We created 8 elevation strata (spanning 20 m intervals) separated by 102 m ranging from [468 - 488] to [1184 - 1204] m a.s.l., aimed to control for the lapse rate (Lembrechts *et al.*, 2021), as it is the main driver of temperature in the study area. Inside each of these strata, we defined 8 types of plots: 4 plots of below and above the median canopy cover of our study area (90% canopy cover) with a south or a north-facing slope (defined as lower or higher than 0.75 heat load index). These 4 plots had moderate topographic position indices (between 0.2 and 0.8) and slope (10° < slope < 25°), to avoid confounding their effects with the canopy cover and heat load effects. Additionally, we defined 2 plots with contrasting topographic position indices (lower than 0.2 and higher than 0.8) under high canopy cover and moderate slope. Lastly, we defined 2 161 plots with contrasting slopes: one flat (slope \leq 10°) and one steep (slope $>$ 25°) under high canopy cover and moderate topographic position.

 Of the initial 64 theoretical plots spanning the 8 strata, only 59 of the defined situations occurred, mostly because we lacked low topographic position indices (valley bottom) in high elevation classes. We randomly selected one pixel for each plot and stratum located in public forests. We repeated this random drawing 10,000 times and kept the set of plots that maximized the mean minimum distance between plots to reduce spatial autocorrelation.

 We established the 59 temperature loggers in May 2021 and recorded their location 170 with a GNSS receiver (Trimble TDC600, accuracy= ± 2 m undercover). We placed every logger in public forests to avoid legal constraints (public forest makes up 80% of the forested area in our study region), with no constraints regarding accessibility. We measured canopy closure (0-100%) by visual observation in a 25-meter radius around the logger. We also estimated canopy cover (0-100%) with a planar picture of the canopy using a smartphone placed on top of the logger and the '*Glama'* application (Tichý, 2016). Plots tagged as low canopy cover were placed accordingly by selecting sites with less than 50% canopy closure as computed by '*Glama*'. The visual estimation of canopy closure (25-meter radius) was significantly 178 correlated with the remote sensed tree density (R^2 of the linear relationship = 30.0%, Figure S1), but a weak and non-significant correlation was found with the picture analyzed by '*Glama*' (Figure S1).

 We recorded air and soil temperatures with TMS-4 loggers (resolution= 0.0625 °C, accuracy= ±0.5 °C) protected with a radiation shield (Wild *et al.*, 2019). The loggers recorded temperature every 15 minutes until August 2022. We used air temperature 15 cm above the soil surface because it is the most representative temperature experienced by understory plants. We cleaned the time series with the '*myClim*' R package (Man *et al.*, 186 2023). We calibrated the loggers beforehand for a range of -20 \degree C to +40 \degree C by placing them 187 in a freezer and drying oven along with a T-type thermocouple (accuracy= \pm 0.2 °C). From the recorded period, we focused on the growing season, from 01/04/2023 to 15/08/2023, as it is the most critical period for plant growth. Out of the 59 loggers, 11 were either malfunctioning, stolen, destroyed by animals or displayed erroneous values and were discarded. We checked the capacity of our final sample to cover the variability of our study region following the PCA-based approach of Lembrechts *et al.*, (2021). Our final sampling was able to cover the variability of the valley, except for extreme values of low canopy cover and the unusual valley bottoms of high elevations. The loss of loggers was evenly distributed over plot types, except for the low canopy cover that suffered the most losses (Figure S2).

2.4. Floristic and Species Characteristic Dataset

 We compiled floristic surveys performed (during the growing season) by students and professors covering soil and climatic transects of the region between 2009 and 2022 (average year= 2015.6). All plots were surveyed for all vascular plant species in the herb layer (smaller than 1 m) and their percentage ground cover was visually estimated. We had 306 floristics 202 surveys in total across the study region. Floristic surveys were performed in 20 x 20 m squares 203 (400 m²) with the GPS position (recorded with built-in tablet GPS; accuracy= \pm 10 m) as the 204 center. We used this position to extract elevation, heat load index, topographic position index and canopy cover for every survey. We harmonized taxonomy to the TaxRef V13 standard (Gargominy, 2022). We focused on herbaceous species in the analysis to focus on 207 community dynamics that may reflect shorter-term climate and are less influenced by 208 management than trees or shrubs.

 We used the thermal optimum species' value from ClimPlant V.1.2 (Vangansbeke *et al.*, 2021). These thermal optima are computed from the mean annual temperature within the range of species obtained from Europe-extent distribution atlases. Out of the 348 unique recorded species, 30 were assigned a thermal optimum value, covering 90.0% of the occurrences of the whole floristic dataset. We averaged the thermal optimum of every species (without weighting for abundance) of a given survey to obtain the Community Thermal Index (hereafter CTI), which quantifies the thermal preference of the whole community (Borderieux *et al.*, 2023; Vangansbeke *et al.*, 2021). We calculated species 217 richness of a plot as the number of recorded species whether they had an associated thermal 218 optimum in the database or not. By doing so, we wanted to include rare species that were not included in ClimPlant so that our specific richness is representative of the species pool of our study region. We also assigned a pH optimum value obtained from a bioindication database to each species (Gégout *et al.*, 2005), and averaged (not weighted by abundance) 222 it to obtain a bioindicated pH per plot.

 We used the EuForPlant regional list of forest plant species (Heinken *et al.*, 2022) to assess species habitat affinity. We assigned to each species one of the following affinities: (1.1) species of closed forest (1.2) species that occur in forest edges and openings (2.1) Species that primarily occur in forests but also found in cultural landscapes and forest 227 remnants (2.2) species of open habitats that occur in forest exclusively through opening and 228 early succession. We excluded species of open vegetation (classified "O") because of their low number of occurrences (42). In total, 274 species were assigned to an affinity class, covering 85.7% of the occurrences.

2.5. Understory Temperature Modeling

 We aggregated the 15-minute frequency time series of the recorded temperature of the growing season 2022 to daily mean and maximum temperature. First, we removed values 234 Iower than the $5th$ centile of the day and values higher than the 95th centile to avoid biasing results due to logger malfunction or a brief burst of sunshine on a logger. We then averaged 236 the mean or maximum daily temperature to obtain one unique value per logger, the mean 237 daily and maximum daily temperature of the growing season.

 We used a linear model to predict mean and maximum daily temperature of the growing season with elevation, heat load index, topographic position index and remote sensed canopy density as explanatory variables. We preferred remote-sensed canopy cover over the *in-situ* measurements which allowed us to map the temperature models over the 242 entire study area, and thus infer the understory temperature of floristic surveys (mostly without canopy closure records). We additionally fitted two linear models with the field measured canopy closure (25 m radius observation and planar photography) instead of the remotely sensed measurement to test different methods of canopy closure estimations (Table S2, Table S3). The exceed in warming due to radiation can be amplified when canopy 247 cannot intercept light, thus, we tested an interaction between heat load index and canopy closure and retained the interaction in the final model if found significant (Davis *et al.*, 2019).

250 The mean understory temperature model $(R^2 = 92.2%)$ allowed us to map the contribution of elevation (i.e., lapse rate), map the topoclimate (heat load index and topographic position) and the microclimate (canopy density) separately to the mean understory temperature [\(Figure 2\)](#page-10-0). We mapped the lapse rate by using only the intercept and the elevation parameter. We mapped the contribution of topography cooling compared to the warmest situation (heat load index and topographic position index equal to 1) assuming a median canopy cover (90%) and using the two topographic indices. We mapped 257 the contribution of canopy cover by multiplying its parameter by the tree density product, this projection is however extrapolated for the 20% of pixels with a canopy closure lower than 79%.

2.6. Floristic Composition Analyses

 The soil of our study region can display very different nutrition status and acidity, 262 which can impact both the richness and composition of a community (Degen et al., 2005; Koerner et al., 1997; Zellweger et al., 2015). In addition, soil pH is also negatively correlated with elevation. To account for soil acidity, we first fitted a linear model to predict species 265 richness and CTI with bioindicated pH as the only predictor. These models had a significant R² of 32.6% and 21.5%, respectively. We then summed the mean species richness or CTI to the residual of the corresponding bioindicated pH model to obtain the corrected value. The corrected values allow comparison between communities with bioindicated pH considered equal.

 We used a linear model to predict the corrected species richness and CTI with the 271 contribution to mean understory temperature of elevation, topoclimate and microclimate 272 as predictors (the unit of every predictor is thus \degree C). The parameters of these two models (species richness and CTI) are shown in [Table 2.](#page-11-0) We discretized our results to better illustrate the control of the significant predictors of the model. We split the 306 surveys into three classes with an equal number of surveys, distributed in "cold", "intermediate" and "warm" classes. We tested the difference in species richness and CTI between these classes with Wilcoxon rank-sum tests (Rey & Neuhäuser, 2011).

 We tested the assumption of normality and homoscedasticity of the residuals of the 279 microclimatic model, the species richness and the CTI model following (Zuur & Ieno, 2016), and we tested the significant difference from 0 of the estimated parameters with a Wald test. We partitioned the variance of the predictors of all the models with the '*modEvA'* package (Barbosa *et al.*, 2013).

3.Results

3.1. Environmental determinant of the understory microclimate

 The growing season (GS) temperature of 2022 was above average (mean GS temperature of the period 2005-2020=13.2°C, mean GS temperature=11.6 °C, Markestein weather station (1,184 m a.s.l), (Météo France, 2024)), as a result, the mean daily 288 temperature of the understory (15 cm above the soil surface) was 14.6 °C and spanned 289 between 11.9 °C to 17.5 °C for the higher (1203 m a.s.l) and lower (475 m a.s.l) elevation 290 sensors, respectively. The mean daily maximum temperature of the GS was 19.3 \degree C and 291 reached a maximum of 24.7 \degree C for the lowest elevation plots. Elevation was the primary 292 driver of mean temperature variability, with a lapse rate estimated at -0.68 \degree C 100m \degree 1 (Table [1\)](#page-9-0). The heat load index- contingent on aspect and slope - was the second driver of mean 294 temperature, which can vary up to 1°C between low and high radiation slopes [\(Table 1\)](#page-9-0). 295 Topographic position had a lesser effect on temperature: the mean temperature was 0.56° C lower in the bottom of a valley compared to ridges [\(Table 1\)](#page-9-0). Lastly, canopy closure (remotely sensed) cooled understory temperatures. An increase of 20% of total canopy cover 298 resulted in a decrease of 0.57 °C [\(Table 1\)](#page-9-0). The lapse rate explained 87.4% of the variability 299 in mean temperature, the topographic factors (heat load and topographic position index) 300 3.95%, and canopy cover accounted for 0.82%. The R^2 of the linear model was 92.2%.

 The same predictors except for topographic position were significant in the mean daily maximum temperature model. The heat load index had a higher contribution (21.5%) in the maximum temperature compared to the mean temperature model, daily maxima varied for 3.3°C between low and high heat load indices (Table S1).

 Canopy cover visually estimated in a 25-meter radius was not significant in predicting mean temperature (Table S2). Immediate canopy cover (smartphone photography) above the logger was significant in explaining mean temperature with an interaction with heat load index, low immediate canopy cover in high radiation slopes displayed warmer mean temperature (Table S3).

 *Table 1: Estimated parameters, their standard error and p-values of the predictors included in models of the daily mean growing season temperature. The range of the predictors in the calibration dataset and their effect size on the temperature (range * estimate) are displayed. The percentage of explained variation per type of predictor is included. P-values were obtained with a Wald test on parameters. Heat load and topographic position have no units, refer to the methods for their calculation.*

 The spatial variation of elevation, topography and canopy closure reveals complex and fine-grained contributions to the forest understory climate (Fig. 2). We mapped the individual contributions of elevation [\(Figure 2.](#page-10-0)a), topoclimate (heat load index and topographic position summed; [Figure 2.](#page-10-0)b) and canopy cover (i.e., microclimate; [Figure 2.](#page-10-0)c) in the study area. We represented topoclimate as a cooling effect compared to a baseline location of a south-facing valley top (heat load index =1, topographic position=1) [\(Figure](#page-10-0) [2.](#page-10-0)b). The baseline for canopy cooling of temperature was 0% canopy closure (as pixels displayed the whole 0-100% range), however, the range of microclimatic cooling from our model is 79% to 100% (80% of the pixel, [Figure 2.](#page-10-0)c). We observed strong effects on understory temperatures caused by steep spatial differences in elevation, topography and fine-grained canopy cover [\(Figure 2.](#page-10-0)d). We used this map and model to predict the mean understory temperature and the contribution of the three components described above for further community composition analyses.

 Figure 2: a) Elevation induced change in mean growing season understory temperature of the growing season (lapse of -0.68°C 100 m-1), assuming a canopy closure of 90% and no effect from topography. b) mean understory temperature cooling induced by topography (heat load and topographic position, i.e. topography) assuming an average canopy cover (90%), compared to the warmest situation (south-facing ridges). c) mean understory temperature cooling induced by canopy closure (i.e. microclimate) assuming no effect from topography. We restrained the minimal cooling to -1.5°C, however some pixels displayed lower values up to 0°C due to low to no canopy closure. d) 2 km per 2 km zoomed inset of the red square of the other panels, their color gradient corresponds to the color scale presented in the other panels a-c, respectively. Blank pixels represent land covers other than forests or forests outside of the study region. Linear model R²: 92.2%.

343 *3.2. Microclimatic Determinants of the Floristic Composition*

 Floristic surveys harbored on average 19 herbaceous species (s.d. 10.7), and the mean community thermal index (CTI) was 7.8 (s.d. 0.55). pH was strongly correlated with CTI 346 (R^2 =28.3%) and species richness (R^2 =32.6%). More acidic soils had less diverse and cold- adapted communities. We accounted for this relationship by extracting the residual of a linear model predicting CTI or species richness with pH as the sole predictor (see methods). After accounting for soil effects, elevation-induced change in temperature was the main predictor of CTI, but it did not significantly explain species richness [\(Table 2\)](#page-11-0). The microclimate was not a significant predictor in any of the two models [\(Table 2\)](#page-11-0). Topoclimate was the sole significant predictor of species richness, and it significantly explained CTI. The contribution of topoclimatic cooling to the explained variability of CTI (4.64%) was comparable to the explained variability by elevation (4.6%). We focused the subsequent community analysis around topoclimate cooling effects, as canopy cooling did not significantly explain the species richness nor CTI.

 *Table 2: Estimated parameters, their standard error and p-values of the predictors of the specific richness and community thermal index (CTI) linear models. The range of the predictors and their effect size on the community predicted variable (range * estimate) are displayed. Both species richness and CTI have previously been corrected for their correlation*

361 *with soil pH. The P-value is obtained by a Wald test on the parameter.*

362

 We divided the 306 floristic surveys into cold, intermediate and warm topoclimatic classes each comprised of 102 surveys based on topography-induced cooling. The cold topoclimatic class displayed 23 species on average, while the two other classes displayed 18.5 species on average [\(Figure 3.](#page-12-0)a). This difference of approximately 5 species was 367 significantly different [\(Figure 3.](#page-12-0)a). The mean CTI of the cold topoclimatic class was 7.7 \degree C, which is significantly lower by 0.19°C than the CTI of the two other classes [\(Figure 3.](#page-12-0)b). No such differences were found when using microclimatic (canopy) cooling was used to create the classes (Figure S3). This discretization of the dataset displayed similar patterns as those observed in an alternative analysis using the continuous predictors of the linear model [\(Table](#page-11-0) [2,](#page-11-0) Figure S4).

373
374 *Figure 3: Species richness (a) and community thermal index (b), corrected for bioindicated pH, of 306 floristic surveys evenly spread into three topoclimatic cooling classes. The correction consists of extracting the residuals of a linear model with pH as a sole predictor, this process could thus lead to negative specific richness. The p-value significance of a Wilcoxon test between two classes is displayed as follows: (ns): p>0.05 (*): p<0.05(**): p<0.01 (***): P<0.001.*

 The decreases in CTI and the increase in species richness in the cold topoclimatic class were explained by a surplus of relatively cold-adapted species (i.e. with a species 382 thermal optimum of 9 \degree C or less) [\(Figure 4.](#page-13-0)a). The plots (n=102) in cold topoclimates displayed in total more than 50 to 100 more occurrences of relatively cold-adapted species 384 per thermal optimum classes $(1^{\circ}C)$ than the other two categories [\(Figure 4.](#page-13-0)a). The intermediate topoclimatic class (n=102) also had a higher number of cold-adapted species compared to the warm topoclimatic class (n=102) [\(Figure 4.](#page-13-0)a). The cold topoclimatic class displayed 300 more forest-specialist species occurrences (Heinken *et al.*, 2022) than the other warmer topoclimatic classes, whereas the occurrences of generalist species increased by 200 in total [\(Figure 4.](#page-13-0)b). We recorded a total of 246, 242 and 223 species (i.e., species pool) in the cold, intermediate and warm topoclimatic classes, respectively. A total of 58, 41, and 33 species were unique to the cold, intermediate and warm topoclimatic classes, respectively. This means that there is nestedness of species between communities, as shown in Figure S5.

 Figure 4: Occurrences of species in the three topoclimatic classes as a function of a) their thermal optimum (°C) and b) their habitat affinity defined by the EuForPlant list as follows:

 1,1: closed forest mainly 1,2: forest edges and opening 2,1: forest and open vegetation 2,2: mainly in open vegetation (Heinken et al.*, 2022) The plot-scale occurrence of species is also shown (e.g., 400 occurrences corresponds to approximately 4 species per plots).*

4.Discussion

 We found that both canopy cover and topographic factors strongly influenced understory temperature during the growing season. We disentangled the elevation gradient 404 from the topoclimatic and microclimatic factors by estimating the lapse rate separately, which was expectably the main driver of understory temperature [\(Figure 2\)](#page-10-0). After controlling for the lapse and pH, the temperature cooling by topographic factors, namely topoclimate, was the only significant driver of community composition and richness.

 The positive correlation found between temperature and heat load can be attributed to the higher radiation an equator-facing slope receives, which increases both the mean and daily maximum temperature of the growing season in closed forests. This contrasts with a previous study which only found an effect of heat load on maximum temperature (Macek *et al.*, 2019). Alongside heat load, we found that topographic position influenced mean temperature so that ridges were warmer, and valley bottoms were cooler but had no effect on maximum temperature. We attribute this decrease in temperature to cold air pooling that occurs during nighttime, thus influencing mean daily temperature but with a minimal effect during the hottest hour of the day, when air temperature is homogeneously warm (Smith *et al.*, 2010; Vosper & Brown, 2008). The cooling effect of understory temperature by canopy cover was most apparent for maximum temperature but was also significant for mean temperature. These observations concur with studies with comparable sampling (Davis *et al.*, 2019; Macek *et al.*, 2019). We showed a strong effect of topoclimatic factors on community composition and richness but no contribution of microclimatic factors. Our 422 microclimatic model allowed us to separately predict the lapse rate, topoclimatic cooling and canopy cover cooling with mean temperature as a unit. This allows inferring direct links between temperature variation and communities, a necessary step to advance correlative studies.

 The lack of correlation between species richness or community composition (climatic affinity) with microclimatic cooling is surprising as a majority of studies conducted in lowland forests concluded that dense canopy cover (or closure) explains both the assembly of communities and their slow temporal response to climate change (De Frenne *et al.*, 2013, 2019; Maclean *et al.*, 2015; Richard *et al.*, 2021; Zellweger *et al.*, 2020). In mountain forests, however, the contribution of canopy cover to understory temperature is still under scrutiny (Davis *et al.*, 2019; Macek *et al.*, 2019; Zellweger, Coomes, *et al.*, 2019). We found that topoclimatic factors outweighed canopy closure in explaining understory temperature in our study area (that harbors limited canopy closure variation and high topographic variation), 435 which may explain the absence of a link between canopy cover and communities. This finding adds to the current divergent results from Macek *et al.*, (2019) who found no effect of canopy and Vandewiele *et al.*, (2023) who found a predominance of canopy control on temperature in mountain forests. These apparent contrasting results illustrate the complexity and interactions of factors in mountain forest microclimates, potentially depending on site- specific variations in topography and canopy cover, alongside with synoptic conditions leading to difference in transmittance.

 Part of the challenge to determine canopy cover controls in mountain forests stems from the myriads of methods that are used to estimate canopy cover, ranging from hemispheric photographs, and terrestrial lidar-derived metrics to remotely sensed canopy cover estimations (Ma *et al.*, 2017; Zellweger, De Frenne, *et al.*, 2019). We used Copernicus tree density 2018 satellite images to calibrate the microclimatic model and predict its buffering effect on communities. Remote sensed tree closure density does not account for the vertical profile of trees, which have profound influence on sunlight interception and consequently on understory temperatures (Gril *et al.*, 2023; Zellweger, Coomes, *et al.*, 2019). Remotely sensed canopy cover was significantly but poorly correlated with our field measures (visual estimation and photography), and the year of remote sensing (2018) does not match the average year of a floristic survey (2015.6). These inaccuracies and the missing link of the forest vertical profile could partly explain the lack of a significant relationship between community compositions and cooling induced by canopy cover. We fitted additional understory temperature models with *in situ* measurements of canopy cover to conservatively reject canopy cover as a prominent driver of microclimate and consequently community composition. These models showed no correlation between understory temperatures and canopy closure except for the interaction between immediate canopy closure (photography) in equator-facing slopes (Table S2, Table S3). This demonstrates the need to simultaneously study multiple microclimatic drivers and their interactions in mountain ranges (Davis *et al.*, 2019; Greiser *et al.*, 2020).

 We found that temperature variation owing to topography was equally important in shaping a community's affinity to climate compared to that of the elevational gradient [\(Table 2,](#page-11-0) after soil pH has been controlled for). This is understandably a consequence of community assembly dictated in part by the environment. Lower temperatures at higher altitudes or in topographically shaded slopes can exert a selection pressure on species not adapted to cold whereas lower elevation and high radiation slopes select species not sensitive to late freezing and adapted to warmer temperatures [\(Figure 3\)](#page-12-0). Our prediction of both elevation and topography control on mean temperature are quantified the same unit, Celsius degrees °C, but topography-induced temperature effect on community composition is fourfold compared to that of elevation [\(Table 2\)](#page-11-0). This implies that mean temperature alone cannot drive the difference in community composition, and other biophysical factors correlated with topography-induced temperature should be at play. Maximum temperature could be a better predictor of the crossing of physiological thresholds dictating species selection (Macek *et al.*, 2019; Pérez‐Navarro *et al.*, 2021). However, this hypothesis could not be tested with our dataset as mean and maximum understory temperature were highly correlated. Soil moisture and vapor pressure deficit can also explain the important contribution of topography to communities (Davis *et al.*, 2019).

 Our topographic position metric relies on hydrography, demonstrating that cold air pooling could occur alongside wetter soils and synergistically favor cold-adapted species not tolerant to drought (Bénichou & Le Breton, 1987; Finocchiaro *et al.*, 2023; Raduła *et al.*, 2018). Conversely, ridges and south-facing slopes exacerbate the effect of warmer temperatures by desiccation, via stronger winds and evaporation, respectively (Davis *et al.*, 2019; Piedallu *et al.*, 2023; Rita *et al.*, 2021). These factors altogether and the differences we found in contribution to community composition [\(Table 2\)](#page-11-0) challenge the use of a single microclimate variable (e.g., mean temperature) to predict community patterns and species distribution. Explicitly considering other microscale biophysical factors in a multivariate fashion (Pérez‐Navarro *et al.*, 2021), the improvement of mechanistic modeling of microclimate (Maclean, 2020) could improve predictions of present and future community composition.

 The cold-adapted communities we observed in cold topoclimates are the result of an increase in relatively cold-adapted species occurrences rather than of a decrease in relatively warm-adapted species [\(Figure 3\)](#page-12-0). This hints that the constraints on community assembly, in our study region, are a result of temperature becoming too warm for cold- adapted species, rather than otherwise. This increase in occurrences explains the higher specific richness in cold topoclimates [\(Figure 3\)](#page-12-0). Canopy cover has been identified as the driver of the diversity of many taxa in lowland forests due to its buffering of microclimate and light interception (Tinya *et al.*, 2021; Zellweger *et al.*, 2015). Its lower contribution to microclimate variation in mountain forests and the limitation in its measurement mentioned earlier may explain why we do not detect this pattern. Further to an understory cooling, colder topoclimates could also increase moisture, thus alleviating competition for water during summer and allowing more species to co-occur (Raduła *et al.*, 2018; Sanczuk *et al.*, 2022).

 How these local cooler and wetter conditions are decoupled from the climate warming trend is of utmost importance as they allow for the persistence of cold-adapted species (Greiser *et al.*, 2020; Lenoir *et al.*, 2017), or provide opportunities to facilitate colonization thus facilitating range shifts (Serra-Diaz *et al.*, 2015). The thermal heterogeneity topoclimate produced in mountain ranges [\(Figure 2\)](#page-10-0) should also be considered as a driver of landscape-scale diversity (Stein *et al.*, 2014) and a potential source of community adaptation because species of diverging climatic adaptation coexist in a relatively small area (Hylander *et al.*, 2022; Lenoir *et al.*, 2013). More specifically, our results support the "identifying and protecting microrefugia" section highlighted by Hylander *et al.*, (2022), as north-facing slopes and topographic depressions are easily identifiable from maps, and their cooling capacities and cold-adapted communities confirmed by visits to the field. Although we didn't find a significant canopy variation contribution, canopy is essential to create the ultimate understory condition and should be preserved to take advantage of the topoclimate. This could be achieved through selective logging and continuous cover silviculture and the reduction of edge effects thanks to buffer zones around the microrefugia. Conservation targeting cold topoclimates is more robust because of the increase in redundancy and biodiversity those locations provide (Table S5). Additionally, maintaining a connected forest will foster the benefits of the thermal heterogeneity created by topography (Hylander *et al.*, 2022). Indeed, warm topoclimates will serve as source populations of species adapted to the current climate, and cold topoclimates will maintain cold-adapted populations, resulting in a heterogenous landscape.

 In summary, we show that elevation, topography, and to a lesser extent, canopy closure determines growing season understory temperature in the Vosges mountains in France. Besides elevation, the contribution of topoclimate was the main predictor of community composition and diversity. Understory plant communities of cold topoclimates (north-facing slopes and valley bottoms) harbored a higher number of generalist and forest specialist cold-adapted species. Our results place topography as a prominent driver of forest temperature and a key factor to consider for protecting forest cold-adapted species in the context of accelerated global warming.

5.Data availability

 The spatial, microclimatic, and floristic data used for this analysis can be found in the repository: https://github.com/Jeremy-borderieux/Article_microclim_vosges along with the R script that can be used to reproduce the analyses and the figures, under the DOI [https://doi.org/10.5281/zenodo.12626861.](https://doi.org/10.5281/zenodo.12626861)

6.Acknowledgment

 The authors are grateful to the Grand Ventron naturel reserve and its director Laurent Domergue for the permission to access the core of the protected forest. The authors acknowledge the National Office for Forests (ONF) for permission to place loggers in public forests. The authors thank the AgroParisTech students and professors involved in the collection of floristic data. The authors thank the funding from a PHC Tournesol mobility grant N° 47550SB. JB Acknowledge the funding from a joint funding from Region Grand Est and AgroParisTech (19_GE8_01020p05035). JMSD was funded by the ANR-JCJC (Agence Nationale de la Recherche, jeunes chercheuses et jeunes chercheurs) SEEDFOR (ANR-21- CE32-0003). JMSD acknowledges the support from NASA for UConn's Ecological Modelling Institute (#80NSSC 22K0883).

7.References

- Ashcroft, M. B. (2010). Identifying refugia from climate change : Identifying refugia from climate change. *Journal of Biogeography*. https://doi.org/10.1111/j.1365- 2699.2010.02300.x
- Ashcroft, M., Chisholm, L., & French, K. (2008). The effect of exposure on landscape scale soil surface temperatures and species distribution models. *Faculty of Science - Papers (Archive)*, 211‑225. https://doi.org/10.1007/s10980-007-9181-8
- Barbosa, A. M., Real, R., Munoz, A. R., & Brown, J. A. (2013). New measures for assessing model equilibrium and prediction mismatch in species distribution models. *Diversity and Distributions*, *19*(10), 1333‑1338. https://doi.org/10.1111/ddi.12100
- Bénichou, P., & Le Breton, O. (1987). Prise en compte de la topographie pour la cartographie de champs pluviométriques statistiques : La méthode Aurelhy. *Colloques de l'INRA*, *39*(51‑69).
- Borderieux, J., Gégout, J.-C., & Serra-Diaz, J. M. (2023). High landscape-scale forest cover favours cold-adapted plant communities in agriculture–forest mosaics. *Global Ecology and Biogeography*, *32*(6), 893‑903. https://doi.org/10.1111/geb.13676
- Bramer, I., Anderson, B. J., Bennie, J., Bladon, A. J., De Frenne, P., Hemming, D., Hill, R. 566 A., Kearney, M. R., Körner, C., Korstjens, A. H., Lenoir, J., Maclean, I. M. D., Marsh, C. D., Morecroft, M. D., Ohlemüller, R., Slater, H. D., Suggitt, A. J., Zellweger, F., & Gillingham, P. K. (2018). Chapter Three—Advances in Monitoring and Modelling Climate at Ecologically Relevant Scales. In D. A. Bohan, A. J. Dumbrell, G. Woodward, & M. Jackson (Éds.), *Advances in Ecological Research* (Vol. 58, p. 101‑161). Academic Press. https://doi.org/10.1016/bs.aecr.2017.12.005
- Copernicus. (2018). *High Resolution Layer Tree Cover Density* [Data set]. https://land.copernicus.eu/en/products/high-resolution-layer-tree-cover-density
- Davis, F. W., Synes, N. W., Fricker, G. A., McCullough, I. M., Serra-Diaz, J. M., Franklin, J., & Flint, A. L. (2019). LiDAR-derived topography and forest structure predict fine- scale variation in daily surface temperatures in oak savanna and conifer forest landscapes. *Agricultural and Forest Meteorology*, *269*‑*270*, 192‑202. https://doi.org/10.1016/j.agrformet.2019.02.015
- De Frenne, P., Lenoir, J., Luoto, M., Scheffers, B. R., Zellweger, F., Aalto, J., Ashcroft, M. B., Christiansen, D. M., Decocq, G., De Pauw, K., Govaert, S., Greiser, C., Gril, E., Hampe, A., Jucker, T., Klinges, D. H., Koelemeijer, I. A., Lembrechts, J. J., Marrec, R., … Hylander, K. (2021). Forest microclimates and climate change : Importance, drivers and future research agenda. *Global Change Biology*. https://doi.org/10.1111/gcb.15569
- De Frenne, P., Rodriguez-Sanchez, F., Coomes, D. A., Baeten, L., Verstraeten, G., Vellend, M., Bernhardt-Romermann, M., Brown, C. D., Brunet, J., Cornelis, J., Decocq, G. M., Dierschke, H., Eriksson, O., Gilliam, F. S., Hedl, R., Heinken, T., Hermy, M., Hommel, P., Jenkins, M. A., … Verheyen, K. (2013). Microclimate moderates plant responses to macroclimate warming. *Proceedings of the National Academy of Sciences*, *110*(46), 18561‑18565. https://doi.org/10.1073/pnas.1311190110
- De Frenne, P., Zellweger, F., Rodríguez-Sánchez, F., Scheffers, B. R., Hylander, K., Luoto, M., Vellend, M., Verheyen, K., & Lenoir, J. (2019). Global buffering of temperatures under forest canopies. *Nature Ecology & Evolution*, *3*(5), 744‑749. https://doi.org/10.1038/s41559-019-0842-1
- Degen, T., Devillez, F., & Jacquemart, A.-L. (2005). Gaps promote plant diversity in beech forests (Luzulo-Fagetum), North Vosges, France. *Annals of Forest Science*, *62*(5), 429‑440. https://doi.org/10.1051/forest:2005039
- Dietz, L., Collet, C., Dupouey, J.-L., Lacombe, E., Laurent, L., & Gégout, J.-C. (2020). Windstorm-induced canopy openings accelerate temperate forest adaptation to global warming. *Global Ecology and Biogeography*. https://doi.org/10.1111/geb.13177
- Dobrowski, S. Z. (2011). A climatic basis for microrefugia : The influence of terrain on climate. *Global Change Biology*, *17*(2), 1022‑1035. https://doi.org/10.1111/j.1365- 2486.2010.02263.x
- Dunnington, D., & Thorne, B. (2020). ggspatial : Spatial Data Framework for ggplot2. *R package version1*, *1*.
- Ellis, C. J., & Eaton, S. (2021). Climate change refugia : Landscape, stand and tree-scale microclimates in epiphyte community composition. *The Lichenologist*, *53*(1), 135‑148. https://doi.org/10.1017/S0024282920000523
- Evans, J. S., & Murphy, M. A. (2021). *spatialEco*. https://github.com/jeffreyevans/spatialEco
- Finocchiaro, M., Médail, F., Saatkamp, A., Diadema, K., Pavon, D., & Meineri, E. (2023). Bridging the gap between microclimate and microrefugia : A bottom-up approach reveals strong climatic and biological offsets. *Global Change Biology*, *29*(4), 1024‑1036. https://doi.org/10.1111/gcb.16526
- Franklin, J., Davis, F. W., Ikegami, M., Syphard, A. D., Flint, L. E., Flint, A. L., & Hannah, L. (2013). Modeling plant species distributions under future climates : How fine scale do climate projections need to be? *Global Change Biology*, *19*(2), 473‑483. https://doi.org/10.1111/gcb.12051
- Franklin, J., Serra-Diaz, J. M., Syphard, A. D., & Regan, H. M. (2016). Global change and terrestrial plant community dynamics. *Proceedings of the National Academy of Sciences*, *113*(14), 3725‑3734. https://doi.org/10.1073/pnas.1519911113
- Gargominy, O. (2022). *TAXREF v13.0, référentiel taxonomique pour la France.* [Data set]. UMS PatriNat (OFB-CNRS-MNHN), Paris. https://doi.org/10.15468/VQUEAM
- Gégout, J.-C., Coudun, C., Bailly, G., & Jabiol, B. (2005). EcoPlant: A forest site database linking floristic data with soil and climate variables. *Journal of Vegetation Science*, *16*(2), 257‑260. https://doi.org/10.1111/j.1654-1103.2005.tb02363.x
- Greiser, C., Ehrlén, J., Meineri, E., & Hylander, K. (2020). Hiding from the climate : Characterizing microrefugia for boreal forest understory species. *Global Change Biology*, *26*(2), 471‑483. https://doi.org/10.1111/gcb.14874
- Gril, E., Laslier, M., Gallet-Moron, E., Durrieu, S., Spicher, F., Le Roux, V., Brasseur, B., Haesen, S., Van Meerbeek, K., Decocq, G., Marrec, R., & Lenoir, J. (2023). Using airborne LiDAR to map forest microclimate temperature buffering or amplification. *Remote Sensing of Environment*, *298*, 113820. https://doi.org/10.1016/j.rse.2023.113820
- Gudiksen, P. H., Leone, J. M., King, C. W., Ruffieux, D., & Neff, W. D. (1992). Measurements and Modeling of the Effects of Ambient Meteorology on Nocturnal Drainage Flows. *Journal of Applied Meteorology and Climatology*, *31*(9), 1023‑1032. https://doi.org/10.1175/1520-0450(1992)031<1023:MAMOTE>2.0.CO;2
- Haesen, S., Lembrechts, J. J., De Frenne, P., Lenoir, J., Aalto, J., Ashcroft, M. B., Kopecký, M., Luoto, M., Maclean, I., Nijs, I., Niittynen, P., van den Hoogen, J., Arriga, N., Brůna, J., Buchmann, N., Čiliak, M., Collalti, A., De Lombaerde, E., Descombes, P., … Van Meerbeek, K. (2021). ForestTemp – Sub-canopy microclimate temperatures of European forests. *Global Change Biology*, *27*(23), 6307‑6319. https://doi.org/10.1111/gcb.15892
- Haesen, S., Lenoir, J., Gril, E., Frenne, P. D., Lembrechts, J., Kopecký, M., Macek, M., Man, M., Wild, J., & Meerbeek, K. V. (2023). *Uncovering the hidden niche : Incorporating microclimate temperature into species distribution models*. https://ecoevorxiv.org/repository/view/5364/
- Hannah, L., Flint, L., Syphard, A. D., Moritz, M. A., Buckley, L. B., & McCullough, I. M. (2014). Fine-grain modeling of species' response to climate change : Holdouts, stepping-stones, and microrefugia. *Trends in Ecology & Evolution*, *29*(7), 390‑397. https://doi.org/10.1016/j.tree.2014.04.006
- Heinken, T., Diekmann, M., Liira, J., Orczewska, A., Schmidt, M., Brunet, J., Chytrý, M., Chabrerie, O., Decocq, G., De Frenne, P., Dřevojan, P., Dzwonko, Z., Ewald, J., Feilberg, J., Graae, B. J., Grytnes, J.-A., Hermy, M., Kriebitzsch, W.-U., Laiviņš, M., … Vanneste, T. (2022). The European Forest Plant Species List (EuForPlant) : Concept and applications. *Journal of Vegetation Science*, *33*(3), e13132. https://doi.org/10.1111/jvs.13132
- Hijmans, R. J. (2020). *raster : Geographic Data Analysis and Modeling*. https://CRAN.R-project.org/package=raster
- Hylander, K., Greiser, C., Christiansen, D. M., & Koelemeijer, I. A. (2022). Climate adaptation of biodiversity conservation in managed forest landscapes. *Conservation Biology*, *36*(3), e13847. https://doi.org/10.1111/cobi.13847
- IGN. (2013). *Fiches descriptives des grandes régions écologiques (GRECO) et des sylvoécorégions (SER)*. https://inventaire-forestier.ign.fr/spip.php?article773
- IGN. (2017). *BD ALTI Le modèle numérique de terrain (MNT) maillé qui décrit le relief du territoire français à moyenne échelle* [Data set]. https://geoservices.ign.fr/documentation/donnees/alti/bdalti
- IGN. (2019). *BD Forêt version 2*. Institut National de l'Information Géographique et Forestière. https://inventaire-forestier.ign.fr/spip.php?article646
- IPCC. (2021). Summary for Policymakers. In V. Masson-Delmotte, P. Zhai, A. Pirani, S. L. Connors, C. Péan, S. Berger, N. Caud, Y. Chen, L. Goldfarb, M. I. Gomis, M. Huang, K. Leitzell, E. Lonnoy, J. B. R. Matthews, T. K. Maycock, T. Waterfield, O. Yelekçi, R. Yu, & B. Zhou (Éds.), *Climate Change 2021 : The Physical Science Basis. Contribution of Working Group I to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change* (p. 3−32). Cambridge University Press. https://doi.org/10.1017/9781009157896.001
- Johnston, A. K., Brewster, D., & Berghaus, H. K. W. (1848). *The physical atlas : A series of maps & notes illustrating the geographical distribution of natural phenomena* [Map]. William Blackwood & Sons.
- Kemppinen, J., Lembrechts, J. J., Van Meerbeek, K., Carnicer, J., Chardon, N. I., Kardol, P., Lenoir, J., Liu, D., Maclean, I., Pergl, J., Saccone, P., Senior, R. A., Shen, T., Słowińska, S., Vandvik, V., von Oppen, J., Aalto, J., Ayalew, B., Bates, O., … De Frenne, P. (2023). *Microclimate, an inseparable part of ecology and biogeography*. Zenodo. https://doi.org/10.5281/zenodo.7973314
- Koerner, W., Dupouey, J. L., Dambrine, E., & Benoit, M. (1997). Influence of Past Land Use on the Vegetation and Soils of Present Day Forest in the Vosges Mountains, France. *Journal of Ecology*, *85*(3), 351‑358. https://doi.org/10.2307/2960507
- Lembrechts, J. J., Lenoir, J., Scheffers, B., & De Frenne, P. (2021). Designing countrywide and regional microclimate networks. *Global Ecology and Biogeography*. https://doi.org/10.1111/geb.13290
- Lenoir, J., Graae, B. J., Aarrestad, P. A., Alsos, I. G., Armbruster, W. S., Austrheim, G., Bergendorff, C., Birks, H. J. B., Bråthen, K. A., Brunet, J., Bruun, H. H., Dahlberg, C. J., Decocq, G., Diekmann, M., Dynesius, M., Ejrnæs, R., Grytnes, J.-A., Hylander, K., Klanderud, K., … Svenning, J.-C. (2013). Local temperatures inferred from plant communities suggest strong spatial buffering of climate warming across Northern Europe. *Global Change Biology*, *19*(5), 1470‑1481. https://doi.org/10.1111/gcb.12129
- Lenoir, J., Hattab, T., & Pierre, G. (2017). Climatic microrefugia under anthropogenic climate change : Implications for species redistribution. *Ecography*, *40*(2), 253‑266. https://doi.org/10.1111/ecog.02788
- Ma, Q., Su, Y., & Guo, Q. (2017). Comparison of Canopy Cover Estimations From Airborne LiDAR, Aerial Imagery, and Satellite Imagery. *IEEE Journal of Selected Topics in Applied Earth Observations and Remote Sensing*, *10*(9), 4225‑4236. https://doi.org/10.1109/JSTARS.2017.2711482
- Macek, M., Kopecký, M., & Wild, J. (2019). Maximum air temperature controlled by landscape topography affects plant species composition in temperate forests. *Landscape Ecology*, *34*(11), 2541‑2556. https://doi.org/10.1007/s10980-019-00903-x
- Maclean, I. M. D. (2020). Predicting future climate at high spatial and temporal resolution. *Global Change Biology*, *26*(2), 1003‑1011. https://doi.org/10.1111/gcb.14876
- Maclean, I. M. D., Hopkins, J. J., Bennie, J., Lawson, C. R., & Wilson, R. J. (2015). Microclimates buffer the responses of plant communities to climate change. *Global Ecology and Biogeography*, *24*(11), 1340‑1350. https://doi.org/10.1111/geb.12359
- Man, M., Kalčík, V., Macek, M., Brůna, J., Hederová, L., Wild, J., & Kopecký, M. (2023). myClim : Microclimate data handling and standardised analyses in R. *Methods in Ecology and Evolution*, *14*(9). https://doi.org/10.1111/2041-210X.14192
- McCune, B., & Keon, D. (2002). Equations for potential annual direct incident radiation and heat load. *Journal of Vegetation Science*, *13*(4), 603‑606. https://doi.org/10.1111/j.1654-1103.2002.tb02087.x
- McLaughlin, B. C., Ackerly, D. D., Klos, P. Z., Natali, J., Dawson, T. E., & Thompson, S. E. (2017). Hydrologic refugia, plants, and climate change. *Global Change Biology*, *23*(8), 2941‑2961. https://doi.org/10.1111/gcb.13629
- Météo France. (2024). *Meteo.data.gouv.fr*. https://meteo.data.gouv.fr/datasets
- Pastore, M. A., Classen, A. T., D'Amato, A. W., Foster, J. R., & Adair, E. C. (2022). Cold-air pools as microrefugia for ecosystem functions in the face of climate change. *Ecology*, *103*(8), e3717. https://doi.org/10.1002/ecy.3717
- Pebesma, E. (2018). Simple Features for R: Standardized Support for Spatial Vector Data. *The R Journal*, *10*(1), 439‑446. https://doi.org/10.32614/RJ-2018-009
- Pérez‐Navarro, M. Á., Serra‐Diaz, J. M., Svenning, J., Esteve‐Selma, M. Á., Hernández‐ Bastida, J., & Lloret, F. (2021). Extreme drought reduces climatic disequilibrium in dryland plant communities. *Oikos*. https://doi.org/10.1111/oik.07882
- Piedallu, C., Dallery, D., Bresson, C., Legay, M., Gégout, J.-C., & Pierrat, R. (2023). Spatial vulnerability assessment of silver fir and Norway spruce dieback driven by climate warming. *Landscape Ecology*, *38*(2), 341‑361. https://doi.org/10.1007/s10980-022- 01570-1
- Piqué, A., Pluck, P., Schneider, J.-L., & Whitechurch, H. (1994). The Vosges Massif. In J. Chantraine, J. Rolet, D. S. Santallier, A. Piqué, & J. D. Keppie (Éds.), *Pre-Mesozoic Geology in France and Related Areas* (p. 416‑425). Springer. https://doi.org/10.1007/978-3-642-84915-2_32
- R Core Team. (2019). *R: A Language and Environment for Statistical Computing*. R Foundation for Statistical Computing. https://www.R-project.org/
- Raduła, M. W., Szymura, T. H., & Szymura, M. (2018). Topographic wetness index explains soil moisture better than bioindication with Ellenberg's indicator values. *Ecological Indicators*, *85*, 172‑179. https://doi.org/10.1016/j.ecolind.2017.10.011
- Rey, D., & Neuhäuser, M. (2011). Wilcoxon-Signed-Rank Test. In M. Lovric (Éd.), *International Encyclopedia of Statistical Science* (p. 1658‑1659). Springer. https://doi.org/10.1007/978-3-642-04898-2_616
- Richard, B., Dupouey, J.-L., Corcket, E., Alard, D., Archaux, F., Aubert, M., Boulanger, V., Gillet, F., Langlois, E., Macé, S., Montpied, P., Beaufils, T., Begeot, C., Behr, P., Boissier, J.-M., Camaret, S., Chevalier, R., Decocq, G., Dumas, Y., … Lenoir, J. (2021). The climatic debt is growing in the understorey of temperate forests : Stand characteristics matter. *Global Ecology and Biogeography*, *30*(7). https://doi.org/10.1111/geb.13312
- Rita, A., Bonanomi, G., Allevato, E., Borghetti, M., Cesarano, G., Mogavero, V., Rossi, S., Saulino, L., Zotti, M., & Saracino, A. (2021). Topography modulates near-ground microclimate in the Mediterranean Fagus sylvatica treeline. *Scientific Reports*, *11*(1), 8122. https://doi.org/10.1038/s41598-021-87661-6
- Sala, O. E., Chapin, F. S., Armesto, J. J., Berlow, E., Bloomfield, J., Dirzo, R., Huber- Sanwald, E., Huenneke, L. F., Jackson, R. B., Kinzig, A., Leemans, R., Lodge, D. M., Mooney, H. A., Oesterheld, M., Poff, N. L., Sykes, M. T., Walker, B. H., Walker, M., & Wall, D. H. (2000). Global biodiversity scenarios for the year 2100. *Science (New York, N.Y.)*, *287*(5459), 1770‑1774. https://doi.org/10.1126/science.287.5459.1770
- Sanczuk, P., De Lombaerde, E., Haesen, S., Van Meerbeek, K., Luoto, M., Van der Veken, B., Van Beek, E., Hermy, M., Verheyen, K., Vangansbeke, P., & De Frenne, P. (2022). Competition mediates understorey species range shifts under climate change. *Journal of Ecology*, *110*(8), 1813‑1825. https://doi.org/10.1111/1365-2745.13907
- Sannier, C., Gallego, J., Langanke, T., Donezar, U., & Pennec, A. (2023). Tree cover area estimation in europe based on the combination of in situ reference data and the copernicus high resolution layer on tree cover density. *The International Archives of the Photogrammetry, Remote Sensing and Spatial Information Sciences*, *XLVIII-M-1*‑*2023*, 277‑284. https://doi.org/10.5194/isprs-archives-XLVIII-M-1-2023-277-2023
- 773 Schweiger, A. H., Irl, S. D. H., Steinbauer, M. J., Dengler, J., & Beierkuhnlein, C. (2016).
774 **Chang Destimizing sampling approaches along ecological gradients.** *Methods in Ecology and* Optimizing sampling approaches along ecological gradients. *Methods in Ecology and Evolution*, *7*(4), 463‑471. https://doi.org/10.1111/2041-210X.12495
- Serra-Diaz, J. M., Scheller, R. M., Syphard, A. D., & Franklin, J. (2015). Disturbance and climate microrefugia mediate tree range shifts during climate change. *Landscape Ecology*, *30*(6), 1039‑1053. https://doi.org/10.1007/s10980-015-0173-9
- 779 Smith, S. A., Brown, A. R., Vosper, S. B., Murkin, P. A., & Veal, A. T. (2010). Observations
780 and Simulations of Cold Air Pooling in Valleys. Boundary-Laver Meteorology. 134(1). and Simulations of Cold Air Pooling in Valleys. *Boundary-Layer Meteorology*, *134*(1), 85‑108. https://doi.org/10.1007/s10546-009-9436-9
- Stein, A., Gerstner, K., & Kreft, H. (2014). Environmental heterogeneity as a universal driver of species richness across taxa, biomes and spatial scales. *Ecology Letters*, *17*(7), 866‑880. https://doi.org/10.1111/ele.12277
- Thomas, A. L., Dambrine, E., King, D., Party, J. P., & Probst, A. (1999). A spatial study of the relationships between streamwater acidity and geology, soils and relief (Vosges, northeastern France). *Journal of Hydrology*, *217*(1), 35‑45. https://doi.org/10.1016/S0022-1694(99)00014-1
- Tichý, L. (2016). Field test of canopy cover estimation by hemispherical photographs taken with a smartphone. *Journal of Vegetation Science*, *27*(2), 427‑435. https://doi.org/10.1111/jvs.12350
- Tinya, F., Kovács, B., Bidló, A., Dima, B., Király, I., Kutszegi, G., Lakatos, F., Mag, Z., Márialigeti, S., Nascimbene, J., Samu, F., Siller, I., Szél, G., & Ódor, P. (2021). Environmental drivers of forest biodiversity in temperate mixed forests – A multi- taxon approach. *Science of The Total Environment*, *795*, 148720. https://doi.org/10.1016/j.scitotenv.2021.148720
- Vandewiele, M., Geres, L., Lotz, A., Mandl, L., Richter, T., Seibold, S., Seidl, R., & Senf, C. (2023). Mapping spatial microclimate patterns in mountain forests from LiDAR. *Agricultural and Forest Meteorology*, *341*, 109662. https://doi.org/10.1016/j.agrformet.2023.109662
- Vangansbeke, P., Máliš, F., Hédl, R., Chudomelová, M., Vild, O., Wulf, M., Jahn, U., Welk, E., Rodríguez-Sánchez, F., & Frenne, P. D. (2021). ClimPlant : Realized climatic niches of vascular plants in European forest understoreys. *Global Ecology and Biogeography*, *30*(6), 1183‑1190. https://doi.org/10.1111/geb.13303
- Vosper, S. B., & Brown, A. R. (2008). Numerical Simulations of Sheltering in Valleys : The Formation of Nighttime Cold-Air Pools. *Boundary-Layer Meteorology*, *127*(3), 429‑448. https://doi.org/10.1007/s10546-008-9272-3
- Wickham, H. (2011). Ggplot2. *WIREs Computational Statistics*, *3*(2), 180‑185. https://doi.org/10.1002/wics.147
- Wiens, J. J. (2016). Climate-Related Local Extinctions Are Already Widespread among Plant and Animal Species. *PLOS Biology*, *14*(12), e2001104. https://doi.org/10.1371/journal.pbio.2001104
- Wild, J., Kopecký, M., Macek, M., Šanda, M., Jankovec, J., & Haase, T. (2019). Climate at ecologically relevant scales : A new temperature and soil moisture logger for long- term microclimate measurement. *Agricultural and Forest Meteorology*. https://doi.org/10.1016/j.agrformet.2018.12.018
- Zellweger, F., Braunisch, V., Morsdorf, F., Baltensweiler, A., Abegg, M., Roth, T., Bugmann, H., & Bollmann, K. (2015). Disentangling the effects of climate, topography, soil and vegetation on stand-scale species richness in temperate forests. *Forest Ecology and Management*, *349*, 36‑44. https://doi.org/10.1016/j.foreco.2015.04.008
- 821 Zellweger, F., Coomes, D., Lenoir, J., Depauw, L., Maes, S. L., Wulf, M., Kirby, K. J., Brunet, J., Kopecký, M., Máliš, F., Schmidt, W., Heinrichs, S., den Ouden, J., Jaroszewicz, 823 B., Buyse, G., Spicher, F., Verheyen, K., & De Frenne, P. (2019). Seasonal drivers of understorey temperature buffering in temperate deciduous forests across Europe. *Global Ecology and Biogeography*, *28*(12), 1774‑1786. https://doi.org/10.1111/geb.12991
- Zellweger, F., De Frenne, P., Lenoir, J., Rocchini, D., & Coomes, D. (2019). Advances in Microclimate Ecology Arising from Remote Sensing. *Trends in Ecology & Evolution*, *34*(4), 327‑341. https://doi.org/10.1016/j.tree.2018.12.012
- Zellweger, F., De Frenne, P., Lenoir, J., Vangansbeke, P., Verheyen, K., Bernhardt- Römermann, M., Baeten, L., Hédl, R., Berki, I., Brunet, J., Van Calster, H., Chudomelová, M., Decocq, G., Dirnböck, T., Durak, T., Heinken, T., Jaroszewicz, B., Kopecký, M., Máliš, F., … Coomes, D. (2020). Forest microclimate dynamics drive plant responses to warming. *Science*, *368*(6492), 772‑775. https://doi.org/10.1126/science.aba6880
- Zuur, A. F., & Ieno, E. N. (2016). A protocol for conducting and presenting results of regression-type analyses. *Methods in Ecology and Evolution*, *7*(6), 636‑645. https://doi.org/10.1111/2041-210X.12577
-
-