Adaptive potential in the face of a transmissible cancer in Tasmanian devils

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Abstract 1

2 Emerging infectious diseases (EIDs) cause catastrophic declines in wildlife populations, but 3 also generate selective pressures that may result in rapid evolutionary responses. One such EID 4 is devil facial tumour disease (DFTD) in the Tasmanian devil. DFTD is almost always fatal, 5 which likely causes strong selection for traits that reduce susceptibility to the disease, but 6 population decline has also left Tasmanian devils vulnerable to inbreeding depression. We 7 analysed 22 years of data from an ongoing study of a population of Tasmanian devils on 8 Freycinet Peninsula, Tasmania, to (1) identify whether DFTD may be causing selection on 9 body size, by estimating phenotypic and genetic correlations between DFTD and size traits, (2) 10 estimate the additive genetic variance of susceptibility to DFTD, and (3) investigate whether size traits or susceptibility to DFTD were under inbreeding depression. We found a positive 11 12 phenotypic relationship between head width and susceptibility to DFTD, but this was not 13 underpinned by a genetic correlation. Conversely, we found a negative phenotypic relationship 14 between body weight and susceptibility to DFTD, and there was evidence for a negative genetic correlation between susceptibility to DFTD and body weight. There was additive genetic 15 16 variance in susceptibility to DFTD, head width and body weight, but there was no evidence for 17 inbreeding depression in any of these traits. These results suggest Tasmanian devils have the 18 potential to respond adaptively to DFTD, although the realised evolutionary response will 19 critically depend on the evolution of DFTD itself.

20 Keywords: transmissible cancer, wildlife disease, quantitative genetics, selection differential, 21 adaptive potential, inbreeding depression

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29 Introduction

30 Emerging infectious diseases (EIDs) are often critical drivers of population and evolutionary 31 dynamics in their host species (Daszak et al., 2000; Schrag & Wiener, 1995). In particular, 32 EIDs can induce rapid evolutionary responses in traits that determine hosts' exposure to 33 pathogens (Herrera & Nunn, 2019), pathogen load (disease resistance/susceptibility; (Rigby et 34 al., 2002)) and/or the costs of infection (disease tolerance; (Medzhitov et al., 2012)), especially 35 in cases where EIDs impact fertility or cause rapid mortality (Altizer et al., 2003; Cunningham 36 et al., 2021). However, whilst the ecological impacts of EIDs in natural populations are widely 37 reported, including rapid population decline and species range contractions (Fisher & Garner, 38 2020; C. Hoffmann et al., 2017; Hoyt et al., 2021), empirical evidence for evolutionary 39 consequences of the emergence of infectious diseases in wild populations has been more 40 limited, likely due to a lack of appropriate individual-based data.

41 EIDs should select for traits which improve host immune defences (Hayward et al., 2014; 42 Rarberg & Stjernman, 2003), but an adaptive evolutionary response in the susceptibility to disease is dependent on there being standing genetic variation in immune related traits (A. A. 43 44 Hoffmann et al., 2017). In wild populations, genetic variation in traits can be estimated by 45 combining individual-level phenotypic data with either a pedigree or genomic relatedness data 46 (Wilson et al., 2010), and although these data are hard to collect in natural populations, some 47 recent studies have used data from long-term field projects to estimate genetic variance in 48 susceptibility to disease. These studies have reported a range of estimates of genetic variance, 49 from a heritability of 0.12 for *Mycobacterium bovis* infection in European badgers (Marjamäki 50 et al., 2021) and 0.13 for Chlamydia pecorum infection in koalas (Cristescu et al., 2022), to a relatively high heritability of 0.55 for Mycoplasma ovipneumoniae infection in bighorn sheep 51 (Martin et al., 2021). Despite these recent studies, however, estimates of genetic variation in 52 susceptibility to pathogens in wild populations remain rare, limiting our understanding of the 53 54 potential for adaptive responses to EIDs in the wild.

55 Selection caused by EIDs should also indirectly impact traits that are correlated with 56 individuals' susceptibility to the disease. Body size, for instance, is an important fitness-related 57 trait that shapes individual variation in life-history traits (Healy et al., 2019) and is likely 58 correlated with disease traits as a result of trade-offs caused by differential allocation of 59 resources (Coltman et al., 2001; Gleeson et al., 2005; Silk & Hodgson, 2021; Valenzuela-60 Sánchez et al., 2021). A response to disease-induced selection in size traits would require a 61 genetic correlation between size and susceptibility to disease. However, phenotypic 62 correlations may also be driven by components of the environment that simultaneously impact 63 both traits, which may not result in evolutionary change (Falconer & Mackay, 1996). As such, 64 identifying whether phenotypic relationships are caused by genetic covariances is important in 65 predicting the observed response to selection in either trait (Lande & Arnold, 1983).

66 Whilst EIDs may induce selection for immune traits and those genetically correlated with them, 67 population declines following the emergence of disease can also cause a rapid decline in 68 genetic diversity concurrent with an increase in inbreeding (Hedrick & Kalinowski, 2000). 69 Increased inbreeding causes increased genome-wide homozygosity and, where this directly impacts fitness, will result in inbreeding depression (i.e., reduced fitness caused by inbreeding) 70 71 (O'Grady et al., 2006). Due to the tight association between disease traits and fitness 72 components (i.e., survival and/or reproduction), immune traits are likely to be depressed under 73 increased inbreeding (Spielman et al., 2004), which has been documented in a number of wild 74 animals (e.g. Reid et al., 2003; Ross-Gillespie et al., 2007; Trinkel et al., 2011). When assessing 75 the evolutionary impact of EIDs in declining populations, it is therefore necessary to test for 76 inbreeding depression in immune-related traits.

77 An EID currently imposing extreme selection in a wild animal population is the transmissible 78 cancer, devil facial tumour disease (DFTD), in Tasmanian Devils (Sarcophilus harrisii). 79 Tasmanian devils are the largest extant carnivorous marsupial, and are endemic to the island 80 of Tasmania, Australia. DFTD is a transmissible cancer that originated in a single Schwann 81 cell in the 1980s (Murchison et al., 2010; Patton et al., 2020), and has since spread across 82 almost the entirety of the species' range (Cunningham et al., 2021). Tumour cells are 83 transmitted between hosts by allograft, often during aggressive interactions in the mating 84 season or in competitive carrion feeding interactions when biting occurs (Hamede et al., 2013). 85 With very few exceptions (Pye et al., 2016), DFTD evades an immune response, becomes 86 malignant and causes mortality within 6 - 9 months of symptom onset (McCallum, 2008). As 87 a result, DFTD has caused local population declines of over 80% (Cunningham et al., 2021; 88 McCallum et al., 2007). Given the near 100% mortality associated with DFTD once infected, 89 it is likely that the emergence of the disease has generated strong selection for traits that reduce 90 susceptibility to the disease, and hence also on any genetically correlated traits. Accordingly, 91 several studies have provided evidence for phenotypic and genomic changes in devil 92 populations since the arrival of DFTD. First, allele frequencies at some immune-function genes

93 have changed since DFTD emerged, indicating that there might be contemporary selection on 94 immunity (Epstein et al., 2016; Stahlke et al., 2021). Second, the rate of females breeding at 95 one year's old increased sharply after DFTD was first detected (Jones et al., 2008), and while 96 reduced food competition associated with population decline may cause increased growth rates 97 (and hence higher chances of precocial breeding), selection may have played a role in the shift 98 to precocial breeding (Lachish et al., 2009). Further, a genome-wide association study has suggested that susceptibility to DFTD may have a genomic basis, indicating that there may be 99 100 genetic variance required for the population to mount an adaptive response (Margres et al., 101 2018). Finally, selection by DFTD appears to swamp out selection by local abiotic factors 102 (Fraik et al., 2020), indicating the ecological importance of the disease.

103 In this study, we applied quantitative genetic analyses to data collected from a closelymonitored wild population of Tasmanian devils on Freycinet Peninsula, on the east coast of 104 105 Tasmania, to estimate the potential for an evolutionary response following the emergence of 106 DFTD. In particular, we used genomic relatedness data to estimate the extent of genetic 107 variation and/or inbreeding depression in susceptibility to DFTD, as well as to measure the 108 phenotypic and genetic correlations between susceptibility to DFTD and body size. In 109 Tasmanian devils, body size may be subject to disease-induced selection as size commonly predicts social dominance, which in turn increases the frequency of the types of social 110 111 interaction which result in disease transmission (Hamede et al., 2008, 2009; Hamilton et al., 112 2020). If we assume susceptibility to DFTD directly correlates with a component of individual 113 fitness (i.e., survival), then the phenotypic and genetic correlations of individuals' DFTD 114 infection status with size traits should approximate the predicted change in size traits resulting 115 from selection induced by DFTD (Price, 1970; Robertson & Lewontin, 1968). Under these 116 general predictions, we specifically aimed to (1) identify phenotypic and genetic correlations 117 between susceptibility to DFTD and size, (2) estimate genetic variation in susceptibility to 118 DFTD, and (3) test for inbreeding depression by estimating the relationship between inbreeding 119 and susceptibility to DFTD or size traits (i.e., head width and body weight).

120 Materials and Methods

121 Tasmanian devil study site, trapping and phenotypic data

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We used data collected between January 1999 and May 2021 during an ongoing mark-recapture
study of Tasmanian devils on the Freycinet Peninsula, Tasmania, Australia. DFTD first

125 appeared at this site in 2001, resulting in two years' data pre-disease emergence followed by 20 years of data after disease arrival, as the population descended into long-term decline. 126 127 Tasmanian devils were trapped across the entire 160 km² peninsula up to four times a year using custom-built baited traps (Lachish et al., 2007), with trapping periods timed to coincide 128 129 with key stages in the breeding cycle: autumn (April/May), small pouch young; winter 130 (July/August), large pouch young; spring (October/November), females lactating with young in dens; summer (January/February), dependent young emerging from dens. At their first 131 132 capture, devils were sexed, individually tagged with an ear tattoo (from 1999 to 2004) or a 133 microchip (after 2004) and a 3mm biopsy sample of tissue taken from the outer edge of the ear 134 for genetic analysis (see below). At first capture and then at all subsequent recaptures their age, 135 head width (in mm) and body weight (in kg) were recorded as described in (Lachish et al., 136 2007). We use head width as a linear measure of body size because it is precise, as measured 137 across the bony jugal arches of the skull covered by skin with no muscle or fat deposits. Pouch-138 young of trapped females were sexed and measured, but ear tissue samples were generally not 139 taken because this would result in larger biopsy scars as the individual grew to adult size. A 140 small number of matched pouch-young and mothers were sampled between 2000 and 2003, 141 with 2mm biopsy tissue samples taken from N = 64 pouch young of N = 27 mothers and 142 subsequently sequenced (see below). This allowed us to use these known relatives to assess 143 accuracy and precision of genetic relatedness estimation (see below for details). Individuals 144 were aged using a combination of head width, molar eruption, molar tooth wear and canine over-eruption (Jones, 2023), and given a birthdate of April 1st for a given year, as per Lachish 145 146 et al. (2007). This method of aging is accurate up to two years of age, but most individuals 147 were first trapped as juveniles and were therefore of precisely known age. Disease status 148 (presence/absence) was determined for each capture by visual inspection for tumours and/or 149 histopathological examination of tumour biopsies (Hamede et al., 2015). The total number of 150 capture records across the 22 years was 2156 across 972 individuals, giving an average number of captures per individual was 2.31 (min = 1, max = 11), and DFTD was confirmed in 10% of 151 these captures with 17% of individuals caught with DFTD at least once. Average age at capture 152 153 was 22 months (inter-quartile range = 16 - 31 months), and average age at capture with 154 infection was 27 months.

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156 The analyses presented in this study used two different subsets of the full dataset collected 157 during the long-term project. For both datasets, we only included observations from 'adult' 158 individuals at least 14 months old. This was done (1) to minimize conflation of age and size 159 measurements, and (2) because this is the age at which female devils can be sexually mature 160 such that biting interactions begin and they can thus be at risk of contracting DFTD (Jones et 161 al., 2008). The first dataset was used for analyses of phenotypic relationships (see Statistical analysis section below). After removing observations of individuals younger than 14 months 162 163 and those where there were missing data, this dataset consisted of 1550 recaptures of 729 individuals (hereafter "phenotypic dataset"; 354 males and 375 females). The second subset 164 only included captures of individuals for which we also had genetic data in addition to 165 166 phenotypic data. Genetic data (described below) were used to estimate genetic relatedness and 167 inbreeding coefficients needed for quantitative genetic analyses to estimate additive genetic variance (V_A) and inbreeding depression, as well as genetic covariances between traits (see 168 169 Statistical analysis section below for details). This latter dataset used for quantitative genetic analyses comprised of 498 observations of 243 individuals (hereafter "genetic dataset"; 121 170 171 males and 122 females).

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173 DNA extraction and genotyping

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We extracted DNA from tissue samples and conducted genotyping as previously described 175 176 elsewhere (Epstein et al., 2016; Margres et al., 2018). Briefly, single-nucleotide polymorphism 177 (SNP) genotyping was achieved via single-digest RADcapture (i.e., "Rapture" (Ali et al., 2016)) of DNA extracted from tissue. All raw reads from sequencing were first aligned to to 178 179 the S. harrisii reference genome (Murchison et al., 2012). The first round of sequencing that 180 was conducted with this population resulted in data that were of low sequencing depth. As a 181 result, most samples were subsequently re-sequenced in another run to achieve deeper 182 sequencing depth, thereby improving genotyping accuracy. Reads generated from these 183 replicate runs of the same individuals were merged after aligning to the reference genome, and SNP calling was conducted using the merged "bam" files using the stacks pipeline as in 184 185 (Stahlke et al., 2021). PCR duplicates were removed and SNPs were discovered and called 186 using gstacks (Catchen et al., 2013). The function populations was then used to filter SNPs to 187 keep one random SNP per RAD locus and per 10Kb window, exclude SNPs with a minor allele 188 frequency (MAF) below 1%, remove individuals with more than 70% missing data, and remove 189 SNPs that were present in less than 50% of the samples. We then further filtered genotype calls 190 with a read depth of less than 4 in order to increase genotyping accuracy, before reapplying the 191 filtering parameters explained above. This resulted in a total of 2105 SNPs genotyped in a total of 584 individuals for the whole study population (which was further restricted for use in
further analyses – see below).

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195 Genomic relatedness estimates

197 Quantitative genetic analyses used to partition phenotypic variance into additive genetic and 198 environmental effects are often achieved via a pedigree (Kruuk, 2004; Wilson et al., 2010), 199 which can be based on field observations and/or constructed using genetic marker data. 200 Unfortunately, irrespective of the SNP filtering parameters we used, we were unable to 201 determine sufficient numbers of parentage assignments for a pedigree that could be used for 202 analyses: of 651 individuals, we were only able to assign maternities to 160 and paternities to 203 175 (N = 83 individuals with both parents assigned). We were also only able to match 40 of 64 204 known mother-offspring (pouch young) relationships, with the remaining 24 either not 205 assigned to a mother or mismatched. We therefore ran our quantitative genetics models using 206 a genomic relatedness matrix (GRM) instead of a pedigree (Bérénos et al., 2014; Gervais et al., 207 2019). It has been suggested that running these models with a GRM may in fact improve the 208 accuracy of quantitative genetics parameters estimated via these models, especially when 209 pedigree depth is relatively shallow (Gienapp et al., 2017).

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211 To estimate a GRM, we first filtered the set of SNP loci to improve the precision and accuracy 212 of the GRM, following Gervais et al. (2019). SNPs were filtered for a MAF of at least 10% and 213 missingness no greater than 50%, resulting in a set of 1811 SNPs. Prior to calculating the final 214 GRM, we first used the filtered SNP set to identify and remove possible duplicate pairs of individuals contained in the dataset. Duplicate pairs of individuals may occur in cases where, 215 216 for instance, an individual identification is lost (unreadable tattoo or failure to locate a 217 microchip) on recapture, and they are treated as a new individual and given a new 218 identification. Duplicate individuals were identified and removed from analyses using pairwise 219 relatedness and confirmed via matched life-history data. To do this, we first identified pairs of 220 sequenced samples that had extremely high estimated relatedness (threshold >0.8) and were 221 therefore likely duplicate samples of the same individual. This threshold was selected based on 222 the upper tail of the total distribution of relatedness estimates, assuming that there should be a 223 non-continuous distribution of relatedness values between (truly) highly related pairs and those 224 that are instead duplicates. For each putative duplicated pair, we then cross-referenced with 225 their estimated birth year and sex to ensure that they were indeed duplicates. This procedure

226 identified 44 pairs of samples, and for each duplicated pair, the sample with the best quality genotyping data was kept. After removing duplicated individuals and re-filtering the SNP 227 dataset according to parameters explained above, the final dataset for estimation of the GRM 228 included 540 individuals and 1808 SNPs. Note that not all of these 540 individuals had 229 230 phenotypic data for DFTD status associated with them, so therefore not all were included in 231 the statistical analyses below. However, we retained all these individuals for the estimation of the GRM so as to improve precision of allele frequencies of the population required for 232 233 estimating relatedness.

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235 We next assessed which relatedness estimate performed best at estimating known relatives in 236 this dataset (N = 51 mother – pouch-young pairs in which both individuals had genetic data). 237 Relatedness was estimated using six measures: Yang relatedness was estimated using GCTA 238 (Yang et al., 2011), and Wang, Queller and Goodnight, Dyad maximum likelihood, Lynch, and Ritland estimates were all estimated using COANCESTRY (Wang, 2010). Comparing pairwise 239 240 relatedness estimates for all mother-offspring pairs (detailed above), the Wang relatedness 241 estimate performed best, with an average R value for mother-offspring pairs of 0.47. We 242 therefore used the GRM calculated using *Wang* relatedness estimate in all further quantitative 243 genetic analyses. The variance in pairwise relatedness values using this estimate was 0.007, with approximately 518 pairs of first-degree relatives (i.e., parent-offspring pairs or full 244 245 siblings, r > 0.45) and 2414 pairs of second-degree relatives (e.g., half-siblings, r = 0.2 - 0.3) 246 (out of a total of 145,530 possible pairs) (Figure S1).

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248 Inbreeding coefficients

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We measured variation in inbreeding using genomic inbreeding coefficients estimated in 250 GCTA (Yang et al., 2011). We selected to use \hat{F} III (hereafter FGRM), which estimates the allelic 251 252 correlation between gametes, as this measure has been found to be most closely correlated with 253 runs of homozygosity on the genome (FROH), and is therefore likely a better measure of the 254 genomic consequences of inbreeding (Yang et al., 2011). We ensured that FGRM measures were 255 robust to SNP filtering by varying the MAF cut-off criterion (1%, 5% and 10%). F_{GRM} estimates 256 were all very highly correlated irrespective of which MAF cut-off we used (r > 0.99%). Our genomic measure of inbreeding, FGRM, ranged from -0.37 (indicating that the individual's 257 258 parents are not related to each other) to 0.36 (indicating that the individual's parents are highly 259 related to one another) (median $F_{GRM} = -0.04$, variance = 0.006, Fig S2).

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261 Statistical analyses

In all models, susceptibility to DFTD was fit as a case-control binary variable (1/0 262 case/control), where "cases" were any capture of a devil with a confirmed DFTD infection, and 263 264 "controls" were captures of an uninfected individual. Note that we used a dataset containing repeated measures of all individuals, which in some cases means that an individual may first 265 266 be considered a control before being diagnosed with DFTD at one or more subsequent 267 recaptures. All models were fit in stan via the brms R package (Bürkner, 2017) using default 268 flat priors on the fixed effects and half-Cauchy priors with 2 degrees of freedom on the random 269 effects. All models were run for 10 000 iterations with a warm-up period of 2000 across four 270 chains, and convergence was assessed by ensuring R-hat was below 1.01, effective sample 271 sizes for all parameters were at least 1000 and by visually ensuring chains had mixed well.

a. Selection on size via DFTD

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274 We estimated the phenotypic relationship between susceptibility to DFTD and size traits by 275 fitting a univariate mixed effects model of the effect of size traits on the probability of having 276 DFTD, using the phenotypic dataset. This model (Model 1; Table 1) fit DFTD occurrence on 277 a given capture with a logit link via the Bernoulli family and included the following fixed 278 effects: *age in months* to account for increased likelihood of contracting the disease as devils 279 age; sex to account for any potential sex differences in likelihood of contracting the disease; 280 the interaction between *age* and *sex*; *year* as a covariate to account for the increase in disease 281 presence in the population through time; *head width (mm)* and *body weight (kg)* measured at 282 the same capture. We also fit as multi-level random effects: year, to account for repeated 283 measures on multiple years and any non-linear variation between years in disease prevalence; trap ID, which described the location of the trap at which individuals were caught (trap 284 285 locations were consistent across years) and was used to account for spatial environmental 286 heterogeneity across the study area; and individual ID to account for repeated measures of 287 individuals.

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b. Additive genetic variance (V_A) and inbreeding depression

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To test whether there was evidence for variance in additive genetic effects (V_A) or inbreeding
depression in any of the phenotypic traits (susceptibility to DFTD, head width, weight), we ran

a suite of univariate animal models using the genetic dataset. Animal models extend linear
mixed effects models by incorporating relatedness information to partition phenotypic variance
into additive genetic and other sources of variance (Kruuk, 2004; Wilson et al., 2010). We ran
a single model for each trait, where DFTD occurrence was fit as a response variable with a
logit link via the Bernoulli family (Model 2; Table 1), and head width and body weight were
both fit as response variables as Gaussian traits (Model 3 and 4; Table 1).

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300 Animal models were fit with the following fixed effects: *age in months* to account for growth 301 and increased likelihood of contracting disease with age, year to account for phenotypic change 302 through time, FGRM to test for evidence for inbreeding depression, and the interaction between 303 age and FGRM to test whether the effect of inbreeding changed with age (Marjamäki et al., 304 2021). Animal models for head width and body weight further included sex, the quadratic effect of age (i.e. age²), and the interaction between sex and age, and sex and age². V_A was estimated 305 in animal models by fitting the genomic relatedness matrix as a covariance matrix. We 306 307 estimated permanent environment effects variance (VPE) by fitting repeated measures of 308 individuals via a random effect for individual ID. Animal models further included year as a random effect to account for non-linear variation across years (Vyear), as well as trap ID (VTrap). 309 310 Heritability (h²) for each trait was then estimated as the proportion of phenotypic variance (measured as the sum of all variance components) explained by VA. We present estimates of 311 312 heritability for DFTD on both the latent scale and observed data-scale, which was estimated by 313 converting latent-scale variance estimates to the data-scale using the QGGLMM package in R 314 (de Villemereuil et al., 2016). Latent scale heritability can be interpreted as the expected 315 heritability for a hypothetical (latent) trait reflecting overall susceptibility to DFTD, whereas 316 observed data-scale heritability can be interpreted as the heritability of the probability of being 317 diagnosed with DFTD in the population, which incorporates sampling variance in the observed 318 data.

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Estimates of V_A can be inflated by maternal effects that are unaccounted for in our models (Kruuk & Hadfield, 2007; Wilson et al., 2005). Unfortunately, in these data, maternities for most individuals were unknown because pedigree reconstruction was not possible with the available SNP dataset (see above for details). However, we explored several alternative methods to quantify maternal effects to examine whether our estimates of V_A were being inflated by maternal effects (see supplement). Estimates of V_A were not substantially inflated by not fitting maternal effects (estimated inflation of $h^2 = 1\%$ for DFTD, 5% for weight and 327 3% for head width, see supplementary Text S1 and Figure S3), and thus we present results328 without a maternal effects term fit.

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Finally, to ensure that the temporal trends in either head width or body weight estimated in their respective models in this section did not arise as an artifact of using the genetic dataset, we ran models with head width and body weight as response variables using the phenotypic dataset that included the same fixed and random effects structure as the animal models (Models 3 and 4; Table 1), but without FGRM or the relatedness matrix.

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c. Phenotypic, genetic and other covariances between traits

Phenotypic relationships may be causal if they are associated with a genetic covariance, but 338 339 may also arise when some component of the environment is affecting each trait in parallel (e.g. 340 (Hajduk et al., 2018)). As such, we next ran analyses to estimate the pairwise genetic 341 covariances between susceptibility to DFTD and each of the two size traits. To do this, we ran 342 a suite of bivariate animal models using the genetic dataset. These models used similar fixed and random effects structures to the univariate animal models explained in section b, but were 343 fit without year for head width and weight, and without the interaction between age and FGRM 344 345 for any trait because these effects were not different from zero and so we chose to remove these 346 terms in order to simplify the models. All were fit with two response traits at a time in order to estimate variance-covariance matrices for each random effect (i.e., VA, VPE, VYear, VTrap). 347 348 Specifically, we ran three bivariate models with the following combination of response variables: (1) body weight and head width (Model 5; Table 1); (2) susceptibility to DFTD and 349 350 head width (Model 6; Table 1); and (3) susceptibility to DFTD and body weight (Model 7; 351 Table 1) (note that a single trivariate model of all three traits had convergence problems). Re-352 fitting bivariate models with a 'body condition index' (i.e., body weight divided by head width) 353 did not qualitatively change the results presented. As explained above, bivariate models 354 including DFTD as a binary variable were fit with a logit link and therefore these models do 355 not estimate a residual covariance between the binary and Gaussian trait (Bürkner, 2021). 356 Therefore, we also fit bivariate models with 'relative DFTD' fit with Gaussian errors, where 357 relative DFTD was calculated by dividing observed DFTD at each observation by the mean 358 probability of having DFTD. These models have the added advantage of directly estimating 359 the selection differential (i.e. covariance) between susceptibility to DFTD and size (see (Price, 360 1970; Walsh & Lynch, 2018) for a detailed explanation). Although these models suggest that there was a negative residual covariance between susceptibility to DFTD and both body weight and head width, the overall qualitative inference of other covariance parameters did not change (Table S3 and S4). We therefore present parameter estimates derived from models where DFTD was fit with a logit link. All models estimated both covariances and correlations for each random effect and we present both parameters for comparison.

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367 Finally, the phenotypic relationships estimated in section *a* were estimated from the phenotypic 368 dataset which contained observations of individuals at least 14 months old for which there were complete phenotypic data (N = 1550 recaptures of N = 729 individuals). However, all 369 370 quantitative genetic analyses used to estimate genetic variances and covariances were run with 371 the genetic dataset which retained observations of individuals with genetic data (N = 498372 observations of N = 243 individuals). Therefore, to ensure any differences in the phenotypic 373 and genetic (or environmental) covariances were not artifacts that arose from the use of 374 different datasets, we re-ran the phenotypic model described in section a with the genetic data 375 to facilitate a more direct comparison with the estimated covariances.

376

377 **Results**

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379 Selection on size via DFTD

There was no evidence for sex differences in the probability of having DFTD (Table 2). However, the probability of an individual having DFTD increased over the study period, and also with individual age (Table 2). Devils with relatively larger heads had a greater probability of having DFTD, even after correcting for age (Table 2, Figure 1). Furthermore, devils with relatively lower body weight had a higher probability of having DFTD (Table 2, Figure 1).

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386 Additive genetic variance (V_A) and inbreeding depression in morphology and DFTD

In our animal models using the genetic dataset, we found effects of age and age^2 on both head width and body weight, indicating further growth in individuals older than 14 months old (see Table 3). There was also an effect of sex, reflecting sexual dimorphism in the species whereby adult males are larger than adult females (Table 3; average body weight: Males = 8.45 ± 2.02 kg, Females = 6.80 ± 1.48 kg), and an interaction between age and sex indicating greater rates of increase with age, even after 14 months. There was no evidence for any change over time in either head width or weight, as indicated by the 95% credible intervals for the linear effects of year overlapping zero (Table 3). Tests of temporal changes in either size trait using the larger phenotypic dataset yielded similar results, as both sets of analyses suggested that neither head width nor body weight was changing through time (see Table S2); these models also showed effectively the same sex and age effects as found in the genetic dataset.

398 Posterior distributions for estimates of additive genetic variance V_A from the animal models 399 were different from zero for all three traits (Table 3, Figure 2). Heritability was estimated at 400 0.14 (95% CI = 0.02 - 0.29) for head width, and 0.23 for body weight (95% CI = 0.09 - 0.38). 401 Heritability for susceptibility to DFTD was estimated at 0.40 on the latent scale (95% CI = 0.12402 -0.71) and 0.07 (95% CI = 0.02 -0.12) on the observed data-scale (Figure 2). All three traits 403 also showed permanent environment effects variance (VPE), but VPE was substantially lower 404 than V_A in susceptibility to DFTD (Figure 2, Table 3). Phenotypic variation associated with 405 spatial heterogeneity (as measured using Trap ID) was relatively small but non-zero for all 406 three traits (Table 3 and Figure 2).

407 There was no evidence for an effect of FGRM on either head width or susceptibility to DFTD: 408 the posterior distribution for the effect of FGRM on both traits centred close to zero (Table 3), 409 suggesting that there was no evidence of inbreeding depression in either head width or 410 susceptibility to DFTD. The 95% CI for the effect of FGRM on body weight also overlapped 411 zero (-3.79 - 0.39), suggesting no statistical support for inbreeding depression in body weight. 412 The posterior distribution did indicate that there was a 94% probability that the relationship 413 between FGRM and body weight was negative, although there remains a 6% probability that the 414 effect of FGRM is either positive or zero (Table 3, Figure 3). In identifying FGRM for genotyped 415 individuals, we found that there were approximately 8 individuals in the dataset that appeared 416 very outbred (i.e., $F_{GRM} < -0.3$). This may arise as an artifact of the dataset (e.g., excess heterozygosity caused by sequencing error), but there was nothing in the data of these 417 418 individuals that suggested that this was not a biological signal and this level of outbreeding may have emerged, for example, as a result of those individuals being immigrants to the study 419 420 site. Nonetheless, removing these very outbred individuals did not change our inferences about 421 inbreeding depression in this dataset.

422423 Phenotypic, genetic and other covariances between traits

424 *Head width and body weight:* The total phenotypic covariance between head width and body 425 weight, estimated as the sum of all covariances from the bivariate model, was positive (COV_P 426 = 3.40; 95% CI = 2.45 – 4.47). The permanent environment effects covariance between head 427 width and body weight was strongly positive (Table 4). There was no statistical support for a 428 positive genetic covariance between head width and body weight as posterior distributions 429 overlapped zero, although 91% of the posterior distribution was positive. The covariances for 430 both other terms (year and trap) were not different from zero (Table 4).

431 DFTD and head width: There was no evidence for an overall phenotypic covariance between 432 susceptibility to DFTD and head width, estimated as the sum of all covariances in a bivariate 433 model using the genetic dataset ($COV_P = 3.51$; 95% CI = -8.34 – 20.97). There was no statistical 434 support for either a genetic or a permanent environment covariance between the traits as the 435 posterior distributions for both were wide and overlapped zero (Table 4, Figure S4). Posterior 436 distributions for both other terms (year and trap) also overlapped zero. The results are in 437 contrast to the positive phenotypic association between susceptibility to DFTD and head width 438 estimated from the phenotypic dataset in section a, which may have been because the phenotypic associations between size traits and susceptibility to DFTD were estimated as 439 relative to each other (i.e., body weight relative to head width and vice versa). However, when 440 441 we re-ran the phenotypic selection model with the genetic dataset, we again found no 442 phenotypic association between susceptibility to DFTD and head width (see Table S1), 443 suggesting instead that the contrasting conclusions concerning the strength of statistical support 444 for the association between DFTD and head width likely occurred from differences between 445 the two datasets.

DFTD and body weight: When fitting susceptibility to DFTD as a binary variable, we found 446 that the total phenotypic covariance between susceptibility to DFTD and body weight, 447 448 estimated as the sum of all covariances in a bivariate model, was clearly negative ($COV_P = -$ 449 2.69; 95% CI = -7.77 - -0.71). The overall negative association was also confirmed when we 450 re-ran the phenotypic selection model with the genetic dataset, where we found a negative 451 phenotypic association between susceptibility to DFTD and body weight (see Table S1). We 452 found a negative genetic covariance between the two traits, estimated at -2.56 (posterior 453 median; 95% CI: -6.11 - -0.50). However, posterior distributions for the permanent

- 454 environmental effects covariance between susceptibility to DFTD and body weight, as well as
- the covariances for the year and trap terms, were wide and overlapped zero (Table 4, Figure
- 456 S4). Although the credible interval for the genetic covariance was different from zero, posterior
- 457 distributions for covariance estimates were all quite wide and uncertain.

458 **Discussion**

459 Our analyses of a long-term dataset of Tasmanian devils revealed evidence of additive genetic 460 variance in susceptibility to DFTD, suggesting that there is adaptive potential for Tasmanian devils to evolve resistance to DFTD. There was no statistical evidence for inbreeding 461 462 depression in susceptibility to DFTD, head width, or body weight. Finally, whilst there was 463 evidence for a positive phenotypic relationship between head width and susceptibility to 464 DFTD, this was not associated with a genetic covariance, whereas there was evidence that the 465 negative phenotypic relationship between weight and susceptibility to DFTD was underpinned 466 by a negative genetic covariance.

Additive genetic variance in a trait will determine the evolutionary response to selection on 467 that trait (Golas et al., 2021; Walsh & Lynch, 2018). Our estimates of V_A indicate a genetic 468 basis to susceptibility to DFTD in Tasmanian devils, which may result in the population 469 470 evolving resistance to the disease. This result aligns with those from a genome-wide association 471 study which suggested that major effect loci explain a significant proportion of variation in the 472 probability of having DFTD (Margres et al., 2018), and is also consistent with several previous 473 studies indicating rapid evolutionary responses of devils as evidenced by allele frequency 474 changes at some loci across the genome (Epstein et al., 2016; Fraik et al., 2020; Stahlke et al., 475 2021). Together with these previous studies, our results suggest there may be some potential 476 for the population to respond adaptively to DFTD. Strong directional selection on any fitness-477 related trait should eventually deplete additive genetic variance as alleles at causal loci move 478 towards fixation (Bulmer, 1971). We may therefore expect that additive genetic variance in 479 susceptibility to DFTD should decrease over time as the population evolves resistance. 480 Alternatively, additive genetic variance may be maintained as a result of the continued 481 evolution of DFTD, resulting in arms-race style host-pathogen coevolution (Best et al., 2008; 482 Boots et al., 2009; Stammitz et al., 2023). The realised evolutionary response in this 483 population will therefore be the product of selection acting on both devils and DFTD, as well 484 as the ecological environment in which devils live and are exposed to the disease. Additionally, 485 while our analyses focused on the resistance to DFTD, it is highly likely that tolerance to DFTD

486 is also evolving in the population (Hamede et al., 2020). In situations such as this, tolerance 487 could be assessed from an individual's survival following infection. However accurately 488 measuring disease tolerance in mark-recapture studies can be inhibited by recapture 489 probabilities, and future work could focus on incorporating data on survival post infection to 490 investigate how disease tolerance evolves in populations facing EIDs.

491 Inbreeding depression occurs when recessive deleterious mutations are expressed as 492 homozygotes as a result of inbreeding and negatively impact traits associated with fitness in a 493 population (Charlesworth & Willis, 2009; DeRose & Roff, 1999). Interestingly, despite DFTD 494 being a reliable predictor of survival, we did not find evidence for inbreeding depression in susceptibility to DFTD. Furthermore, whilst body weight is often directly related to fitness via 495 496 condition-related survival and reproduction, and has been found to be subject to inbreeding 497 depression in many wild animals (Hajduk et al., 2018; Huisman et al., 2016; Laikre & Ryman, 498 1991; Nielsen et al., 2012), we did not find statistical support for inbreeding depression in body 499 weight. Inbreeding depression in Tasmanian devils would be especially concerning considering 500 the repeated historical population bottlenecks and recent steep declines in population size (Brüniche-Olsen et al., 2013, 2014; Lachish et al., 2007; Patton et al., 2020), and so the overall 501 502 lack of evidence for inbreeding depression is positive when assessing the probability of the 503 population's persistence. This is an interesting finding given that inbreeding depression has 504 been found in other Tasmanian devil populations (R. M. Gooley et al., 2020), although studies 505 of captive Tasmanian devils have also found a lack of inbreeding depression (R. Gooley et al., 506 2017). One explanation for the overall lack of inbreeding depression could be that recessive, 507 deleterious alleles have already been purged from the population (Grossen et al., 2020; Hedrick 508 & Garcia-Dorado, 2016; Kirkpatrick & Jarne, 2000) either via inbreeding or during the 509 repeated population bottlenecks experienced across the species' range. Nonetheless, the 510 expression of inbreeding depression may be dependent on both environmental conditions and 511 genetic diversity within the population (Hedrick & Kalinowski, 2000), and whilst a lack of 512 inbreeding depression provides a positive outlook for the population now, it does not protect 513 against inbreeding depression in the future.

514 Phenotypic and genetic covariances between DFTD and size traits can be used to predict 515 whether either size trait will respond to selection caused by the disease (Price, 1970; Robertson 516 & Lewontin, 1968), on the assumption that DFTD is a strong predictor of survival and hence 517 fitness. We found that weight and susceptibility to DFTD were phenotypically and genetically 518 negatively correlated. It is important to note that our phenotypic analyses tested the effect of head width and body weight on DFTD concurrently and therefore our results reflect the effect 519 520 of relative measures of each size trait. This means that we found that individuals with *relatively* greater body weight for a given head width (i.e. skeletal size) were less likely to have DFTD. 521 522 The phenotypic covariance between these traits may reflect an immunocompetence - body-523 condition relationship, whereby (relatively) heavier individuals are in better condition and 524 consequentially have better resistance to disease (Gleeson et al., 2005). Alternatively, the 525 directionality of causality in the phenotypic covariance may be reversed whereby individuals 526 that have the disease subsequently lose weight (Sánchez et al., 2018). As the observed negative 527 phenotypic covariance was mirrored by a negative genetic covariance, this suggests that the 528 relationship is more likely an indirect measure of body-condition positively impacting immune 529 function (Gleeson et al., 2005).

530 We found that there was a positive phenotypic covariance between head width and susceptibility to DFTD at the phenotypic level, but we did not find evidence for this being 531 532 underpinned by a genetic covariance. The underlying mechanisms causing the phenotypic 533 relationship between susceptibility to DFTD and head width remain unclear, although one 534 possibility is that the association may reflect an indirect association with social dominance. For 535 instance, assuming that head width accurately predicts social dominance and males' access to 536 mates in the breeding season when much of the transmission-relevant injurious biting occurs, 537 the relationship between head width and susceptibility to DFTD may reflect a greater 538 probability of infection caused by increased rates of the interactions that cause disease 539 transmission that occur in socially dominant individuals (Hamede et al., 2008, 2009; Hamilton 540 et al., 2019). Interestingly, we found that this relationship was not associated with a genetic 541 covariance. However, re-running the phenotypic model with a smaller dataset did not indicate 542 the same phenotypic relationship between susceptibility to DFTD and head width, suggesting 543 that it is more likely that this dataset was limited in its statistical power to detect the phenotypic 544 relationship, and therefore presumably also any associated genetic or environmental 545 covariances.

546 In conclusion, EIDs are thought to dramatically alter the evolutionary dynamics of wild 547 populations (Rogalski et al., 2017), but empirical evidence of this process is rare. We show that 548 in an endangered marsupial facing an EID that has had a catastrophic impact on the species, 549 there is evolutionary potential in disease traits and current and ongoing selection acting on 550 correlated morphological traits. Critically, we show that susceptibility to DFTD and size traits 551 are all associated with underlying heritable genetic variance. We also show that these patterns 552 exist in the absence of inbreeding depression. These results therefore not only provide 553 important empirical evidence for how EIDs may shape future evolutionary dynamics of a 554 population, but critically suggest that the species may hold the adaptive potential required to 555 avoid extinction.

556

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Figure 1. Plot showing the relationship between head width and DFTD (a), and weight with DFTD (b). Points show observed data, and regression lines show the predicted relationship between size traits and DFTD derived from a mixed effects model which fits DFTD as a casecontrol response as a function of both size traits (see methods for full model structure). Solid dark line shows predictions derived from the median of the posterior and the lighter lines show 100 randomly selected draws from the posterior distribution.







Figure 2. Plot showing proportion of phenotypic variance in DFTD, head width and weight attributed to variance in additive genetic effects (V_A) (reflecting narrow-sense heritability (h²)); permanent environment effects (VPE); year (VYear) and spatial location (VTrap). Variances for DFTD shown on the observed data-scale (see Table 2 for estimates on latent-scale). Posterior median of estimates shown as point, with 75% CI's shown as heavy lines and 95% CI's as lighter line.



- **Figure 3.** Plot showing relationship between F_{GRM} (i.e., \hat{F} III) and body weight. Points show raw, 611 observed data, and regression lines show the predicted relationship between F_{GRM} and body 612 weight, where the solid dark line shows predictions derived from the median of the posterior 613 and the lighter lines show 100 randomly selected draws from the posterior distribution.



Table 1. Table outlining the structure of all linear mixed effects models outlined in *statistical analyses* section. All models were fit in stan via the brms package in R. *Model* refers to the model number referenced in text; *Response* refers to the response variable fit in the model; *Fixed effects* describes the fixed effects structure used in the model, where a colon represents an interaction term between two fixed effects; *Random effects* describes the random effects structure; *Family (link function)* describes the family with which the response variable was fit. Note that in bivariate models, the fixed effects structures varied between response variables and are shown on separate rows.

Model	Response	esponse Fixed effects		Family (link function)
	Univariate			
1	DFTD	Age + Sex + Year + Head width + Body weight	Year + Trap + ID	Bernoulli (logit)
2	DFTD	Age + Year + F _{GRM} + Age:F _{GRM}	Year + Trap + ID + a	Bernoulli (logit)
3	Head width	Age + Age ² + Sex + Year + F_{GRM} + Age: F_{GRM} + Age:Sex + Age ² :Sex	Year + Trap + ID + a	Gaussian
4	Body weight	Age + Age ² + Sex + Year + F_{GRM} + Age: F_{GRM} + Age:Sex + Age ² :Sex	Year + Trap + ID + a	Gaussian
	Bivariate			
5	Head width ; Body weight	Age + Age ² + Sex + Age:Sex + Age ² :Sex + F _{GRM} ; Age + Age ² + Sex + Age:Sex + Age ² :Sex + F _{GRM}	Year + Trap + ID + a	Gaussian Gaussian
6	Head width ; DFTD	Age + F _{GRM} + Year + Age ² + Sex + Age:Sex + Age ² :Sex ; Age + F _{GRM} + Year	Year + Trap + ID + a	Gaussian Bernoulli (logit)
7	Body weight ; DFTD	Age + F _{GRM} + Year + Age ² + Sex + Age:Sex + Age ² :Sex ; Age + F _{GRM} + Year	Year + Trap + ID + a	Gaussian Bernoulli (logit)

Age: linear covariate describing age of individual in months; *Sex*: two-level effect "Male" or "Female"; *Year*: year of observation: *Head width:* in mm; *Body weight*: in kg; F_{GRM:} individuals inbreeding coefficient; Age²: the quadratic of age in months: *Trap:* the name of the location the observation was taken; *ID*: individual microchip; *a*: additive genetic variance, estimated by fitting genomic relatedness matrix as a covariance matrix.

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Table 2. Table summarising results from a mixed effects model used to estimate phenotypic
 relationship between size traits (body weight and head width) and DFTD occurrence. Response 619 620 variable is the occurrence of DFTD at a given capture of an individual, fitted as a binary trait. 621 TrapID fitted the location of the trap where the individual was caught. Posterior medians of 622 linear coefficient estimate for fixed effects and variance estimates for random effects are 623 presented with 95% credible intervals of posterior distribution in parentheses. Fixed effect 624 estimates where the 95% CI's do not overlap with zero are given in bold. Parameter estimates are on the logit link scale. The dataset used is the phenotypic data set with N = 729 individuals 625 626 over N = 1550 captures, 22 years and 185 traps.

	Parameter	
	Sex _M	-1.30 (-4.39 - 1.14)
	Head width (mm)	0.32 (0.11 - 0.75)
Fixed Effects	Body Weight (kg)	-0.83 (-2.180.09)
	Age (months)	0.29 (0.10 - 0.83)
	Year (continuous variable)	1.16 (0.48 - 3.02)
	ID	7.31 (3.08 - 19.20)
Random effects	Year	4.87 (1.89 - 12.85)
	TrapID	1.87 (0.12 - 5.84)

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629 Table 3. The results of animal models estimating VA and the effect of FGRM on three traits: head width, body weight and probability of having DFTD (DFTD). Posterior medians of all effects 630 are presented with 95% credible intervals of posterior distributions in parentheses. Fixed effect 631 estimates where the 95% credible intervals of the posterior does not overlap with zero are in 632 633 bold. Variance components and proportion of phenotypic variance for susceptibility to DFTD are shown on the latent (logit link) scale (estimates on the data-scale can be found in Figure 2). 634 Estimates where posterior distribution does not overlap with zero in bold. The dataset used has 635 N = 243 individuals over N = 498 captures and 19 years and 128 traps. 636

		Head width	Body Weight	DFTD
	Age	1.19 (0.96 - 1.42)	0.22 (0.16 - 0.28)	0.35 (0.12 - 0.75)
	Age ²	-0.01 (-0.020.01)	-0.002 (-0.0030.001)	-
	Sex _M	-3.57 (-7.34 - 0.24)	-0.53 (-1.15 - 0.45)	-
Fixed Effects	F _{GRM}	-1.97 (-8.62 - 4.90)	-1.68 (-3.79 - 0.39)	-0.88 (-8.33 - 6.76)
	Year	-0.17 (-0.43 - 0.09)	-0.01 (-0.06 - 0.04)	1.68 (0.63 - 3.49)
	Age:Sex _M	0.57 (0.31 - 0.84)	0.12 (0.05 - 0.19)	-
	Age ² :Sex _M	-0.01 (-0.020.01)	-0.002 (-0.0030.002)	-
	VA	4.74 (0.76 - 10.11)	0.36 (0.14 - 0.61)	34.83 (6.17 - 220.55)
	V _{PE}	11.53 (6.92 - 16.96)	0.22 (0.03 - 0.45)	5.91 (0.05 - 60.41)
Random effects	V _{Year}	7.08 (3.66 - 14.34)	0.21 (0.09 - 0.47)	39.03 (8.23 - 244.43)
variance components	V _{Trap}	0.25 (0.002 - 1.30)	0.15 (0.06 - 0.27)	1.89 (0.02 - 19.63)
•	V _R	9.25 (7.89 - 10.93)	0.56 (0.47 - 0.66)	-
	h²	0.14 (0.02 - 0.29)	0.23 (0.09 - 0.38)	0.40 (0.12 - 0.71)
Proportion of				
phenotypic	ICCPE	0.34 (0.20 - 0.49)	0.15 (0.02 - 0.30)	0.07 (0.001 - 0.38)
variance				
	ICC ^{Year}	0.21 (0.12 - 0.36)	0.14 (0.07 - 0.27)	0.44 (0.19 - 0.72)
	ICC ^{Trap}	0.007 (0.0001 - 0.04)	0.38 (0.24 - 0.52)	0.02 (0.0002 - 0.13)

637 Linear coefficient estimates shown for fixed effects. Variance estimates shown for all random 638 effects: variance in additive genetic effects (V_A); permanent environment effects (V_{PE}); year 639 (V_{Year}); spatial location (V_{Trap}) and residual (V_R). Proportion of total phenotypic variance (i.e., 640 sum of all variance components) attributed to additive genetic effects, also known as narrow-641 sense heritability (h²); permanent environment effects (intraclass correlation, ICC_{PE}); year 642 (ICC_{Year}); and spatial location (ICC_{Trap}).

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646 Table 3. The results of the three bivariate models used to estimate covariances between head 647 width, body weight and DFTD. Models were fit with DFTD as a binary variable with a logit link. Posterior medians of all covariance estimates presented with 95% credible intervals of 648 posterior distribution in subscript parentheses. Covariances with DFTD given on the latent 649 650 scale. Full variance-covariance matrices from models can be found in supplementary material (Table S4). Covariance estimates where posterior distribution does not overlap with zero in 651 bold. The dataset used has N = 243 individuals over N = 498 captures and 19 years and 128 652 653 traps.

651				
054		Head width and	DETD and Head width	DFTD and Body
		Body Weight	DFID and nead width	Weight
655	COVA	0.94 (-0.02 - 2.49)	-2.63 (-13.34 - 5.91)	-2.56 (-6.110.50)
	COV _{PE}	2.03 (0.87 - 3.14)	0.26 (-5.40 - 6.18)	0.05 (-0.84 - 0.93)
	COV _{Year}	-0.31 (-0.99 - 0.33)	6.78 (-2.18 - 21.16)	-1.06 (-3.43 - 0.53)
	COV_{Trap}	0.02 (-0.08 - 0.17)	0.23 (-0.42 - 1.37)	0.29 (-0.12 - 1.02)
	COV _{Res}	0.74 (0.47 - 1.04)	-	-
	COVP	3.40 (2.45 - 4.47)	3.51 (-8.34 - 20.97)	-2.69 (-7.770.71)

656 Covariance estimates for additive genetic effects (COV_A), permanent environment effects 657 (COV_{PE}), year effects (COV_{Year}), location effects (COV_{Trap}) and residual effects (COV_{Res}). Total 658 phenotypic covariance between each pair of traits (COV_P) given as the sum of all covariances 659 estimated from bivariate models.

660	Supplementary materials
661 662	Figure S1. Distribution of pairwise relatedness values used in all quantitative genetics models
663	(see main text). Relatedness values estimated using Wang relatedness estimate in COANCESTRY
664	(Wang, 2010).
665	



Relatedness (Wang)

- **Figure S2**. Distribution of individual inbreeding coefficients as estimated by F_{GRM} values. F_{GRM}
- 674 was estimated using \hat{F}_{III} in GCTA (Yang et al., 2011), which approximates relatedness between
- 675 an individual's parents, averaged across all loci. More inbred individuals have higher values
- and more outbred individuals have negative values.



 $\mathsf{F}_{\mathsf{GRM}}$

Table S1. Results from phenotypic model run with the smaller subset of data as used in the quantitative genetic analyses, containing only individuals that had sufficient quality genetic data. *TrapID* is the location of the trap where the individual was caught. The values shown are the posterior medians of linear coefficient estimates for fixed effects and of variance estimates for random effects, with 95% credible intervals of posterior distributions in parentheses. Estimates where the 95% CI does not overlap zero are shown in bold.

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	Parameter	
	SexM	-0.45 (-6.70 - 5.82)
	Head width	0.26 (-0.49 - 1.12)
Fixed Effects	Body weight	-1.55 (-5.58 - 1.40)
	Age (months)	0.61 (0.11 - 1.94)
	Year	2.14 (0.50 - 5.88)
	ID	13.16 (3.69 - 37.71)
Random effects variance	Year	12.88 (3.11 - 40.60)
components	TrapID	3.97 (0.22 - 12.37)

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Table S2 Results of phenotypic models of head width and body weight used to investigate the temporal trends in each trait. *TrapID* is the location of the trap where the individual was caught. The values shown are the posterior medians of linear coefficient estimates for fixed effects and of variance estimates for random effects, with 95% credible intervals of posterior distributions in parentheses. Estimates where the 95% CI does not overlap zero are shown in bold.

		Head width	Body weight
	Age	0.96 (0.86 - 1.06)	0.20 (0.18 - 0.22)
	Age ²	-0.01 (-0.010.01)	-0.002 (-0.0020.001)
Fixed Effects	Sex _M	-2.21 (-4.390.04)	-0.15 (-0.59 - 0.30)
Fixed Effects	Year	-0.10 (-0.25 - 0.05)	0.004 (-0.02 - 0.03)
	$Age:Sex_M$	0.64 (0.49 - 0.79)	0.12 (0.09 - 0.15)
	$Age^2:Sex_M$	-0.01 (-0.010.001)	-0.001 (-0.0020.001)
Random effects	ID	3.82 (3.51 - 4.15)	0.82 (0.75 - 0.90)
variance	Year	2.22 (1.53 - 3.18)	0.45 (0.30 - 0.64)
components	TrapID	1.26 (0.83 - 1.70)	0.24 (0.13 - 0.35)

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Text S1. Estimating maternal effects without a pedigree.

707 Estimates of V_A may be inflated if the trait is affected by maternal effects that are not explicitly 708 modelled (Kruuk & Hadfield, 2007; Wilson et al., 2005). Usually, maternal effects are 709 estimated by fitting a vector of known maternities for all individuals in the dataset as a 710 random effect in the model, where maternities have been either observed in the field or 711 identified via a genetically reconstructed pedigree. This option was not possible in our 712 analyses because in our dataset maternities for most individuals were unknown because (1) 713 it is not possible to sample dependent young while with their mother (see main text for 714 details), and (2) pedigree reconstruction was not possible with the available SNP dataset. 715 Nevertheless, we attempted to examine whether estimates of V_A in DFTD or size traits may 716 be being inflated by maternal effects that we were unable to model by using the following 717 two alternative methods.

718

719 The first of these methods (*Ped+*) involved re-running our univariate animal models 720 (described in main text) with estimated maternities for all individuals included as an 721 additional random effect. We estimated maternities for all individuals using a combination of 722 maternities from the incomplete pedigree, relatedness estimates and life history information. 723 First, we assigned maternities to individuals if they had been successfully identified during 724 pedigree reconstruction (via the sequoia R package (Huisman, 2017). Then, we identified 725 further putative maternities using a combination of relatedness estimates, sex and age. 726 Specifically, we assigned pairs of individuals that putatively may be a mother-offspring pair if 727 they had a relatedness value of more than 0.45, and one of the individuals was a female and was at least 1 year older than the other. That female was then assigned as the possible mother 728 729 of the other individual. This resulted in a total of 116 putative mothers of 243 individuals 730 contained in the dataset. For the remaining individuals that we did not have a mother 731 estimated, we gave them a unique "dummy" mother.

732

The second of these methods (*ME Multi R*) was originally proposed by (Zaitlen et al., 2013), and has since been tested and applied in Soay Sheep (James, C et al., 2023). In this approach, maternal effects were estimated by fitting an additional matrix to our univariate animal models which was a modified version of the full GRM that aimed to group individuals that 737 likely had a shared maternal environment. To create this additional matrix, we truncated the 738 full GRM at a cut-off value such that everything below that value became zero. We did this 739 using two cut-off values of 0.5 (*ME Multi R 0.5*) and 0.25 (*ME Multi R 0.25*). A relatedness of 740 0.5 was used because this is the expected level of relatedness for both mother-offspring pairs 741 and full-siblings, both of which would have a common maternal environment. We then 742 repeated this with a relatedness of 0.25 in order to also capture maternal half-siblings in this 743 matrix.

744

We acknowledge that the methods we present here may be imprecise in their estimation of shared maternal environments: the first method may be missing some maternities, whereas the matrices used in the second method retain other types of relatives that do not share a maternal environment. However, we believe that, in combination, these methods are likely effective in assessing the extent that estimates of V_A may be being affected by maternal effects.

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752 We found that for all three traits (body weight, head width and DFTD) fitting a maternal effect 753 using either of the methods we tested reduced estimates of both VA and VPE. However, the 754 magnitude of difference in our estimates of V_A was very dependent on both the trait and the 755 method we used (Fig S3). Furthermore, the estimates of maternal effects variance were not 756 consistent between the different methods used (Fig S3) and, in all cases, fitting a maternal 757 effect reduced confidence in variance parameter estimates by generating wider posterior distributions. For DFTD, h² was estimated around 0.01 higher (on the data-scale) in a model 758 759 without a maternal effect fitted than when a maternal effect was fit (averaged between the 760 point estimates from the three different methods). For body weight, h² was estimated around 761 0.05 higher in a model without a maternal effect fitted than when a maternal effect was fit 762 (averaged between the point estimates from the three different methods). For head width, 763 h² was estimated around 0.03 higher in a model without a maternal effect fitted than when a maternal effect was fit (averaged between the point estimates from the three different 764 765 methods).

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768 Figure S3. Plot showing proportion of phenotypic variance in DFTD, head width and body 769 weight attributed to variance in additive genetic effects (V_A) (reflecting narrow-sense heritability (h²)); permanent environment effects (V_{PE}); year (V_{Year}) and spatial location (V_{Trap}) 770 and maternal effects (V_{ME}). Variances for DFTD shown on the observed data-scale (see Table 771 772 2 for estimates on latent-scale). Posterior median of estimates shown as point, with 75% Cl's 773 shown as heavy linen and 95% Cl's as lighter line. VME estimated using either approximated 774 maternities (Ped +), or by fitting an additional matrix which truncated the GRM at a cut-off value (ME Multi R, see text S1 for details). 775





Table S3. The results of the bivariate models used to estimate covariances between head width,
body weight and DFTD where models were fit with relative measures of DFTD, body weight
and head width (see Methods for details) and were all fit with Gaussian errors. Posterior
medians of all covariance estimates presented with 95% credible intervals of posterior
distribution in subscript parentheses. Full variance-covariance matrices from models can be
found in Table S4.

	DFTD and Head width	DFTD and Body Weight
COVA	-0.10 (-0.35 - 0.13)	-0.27 (-0.290.05)
COVPE	0.07 (-0.08 - 0.26)	0.06 (-0.08 - 0.20)
COV_{Year}	0.10 (-0.37 - 0.61)	-0.29 (-0.78 - 0.06)
	0.02 (-0.01 - 0.07)	0.04 (-0.01 - 0.12)
COV _{Res}	-0.13 (-0.230.04)	-0.11 (-0.210.001)
COVP	-0.03 (0.55 - 0.50)	-0.57 (-1.090.26)

785 Covariance estimates for additive genetic effects (COV_A), permanent environment effects 786 (COV_{PE}), year effects (COV_{Year}), location effects (COV_{Trap}) and residual effects (COV_{Res}). Total 787 phenotypic covariance between each pair of traits (COV_P) given as the sum of all covariances 788 estimated from bivariate models.

Table S4. Full variance-covariance matrices estimated from the three bivariate models used to estimate genetic covariances between DFTD, head width and body weight. Results from each bivariate model shown in turn. For each 2x2 matrix, variances are shown on the diagonal (and shaded in light grey), covariances below the diagonal and correlations above the diagonal. Each estimate is the median of the posterior distribution followed by the 95% CI in parentheses.

		Body Weight	Head width
Veer	Body Weight	0.24 (0.10 - 0.47)	-0.24 (-0.66 - 0.23)
Year	Head width	-0.30 (-0.99 - 0.33)	8.20 (3.86 - 15.03)
Additive constin	Body Weight	0.28 (0.03 - 0.62)	0.64 (-0.29 - 0.96)
Additive genetic	Head width	0.94 (-0.02 - 2.48)	5.29 (0.10 - 13.35)
Permanent	Body Weight	0.41 (0.15 - 0.66)	0.93 (0.82 - 0.99)
environment	Head width	2.03 (0.87 - 3.14)	11.71 (5.47 - 17.59)
Tree	Body Weight	0.07 (0.009 - 0.14)	0.05 (-0.84 - 0.84)
Irap	Head width	0.02 (-0.07 - 0.17)	0.34 (0.002 - 1.17)
		Head width	DFTD
Maraa	Head width	7.88 (3.65 - 14.45)	0.36 (-0.14 - 0.77)
Year	DFTD	6.78 (-2.18 - 21.16)	73.94 (6.39 - 212.14)
	Head width	5.85 (0.18 - 13.26)	-0.16 (-0.77 - 0.51)
Additive genetic	DFTD	-2.63 (-13.34 - 5.91)	78.39 (5.48 - 274.08)
Permanent	Head width	11.50 (5.84 - 17.71)	0.04 (-0.82 - 0.87)
environment	DFTD	0.26 (-5.40 - 6.18)	8.75 (0.02 - 34.69)
_	Head width	0.38 (0.002 - 1.26)	0.19 (-0.81 - 0.94)
Irap	DFTD	0.23 (-0.42 (1.37)	3.87 (0.01 - 14.73)
		Body Weight	DFTD
Mara	Body Weight	0.28 (0.09 - 0.45)	-0.37 (-0.83 - 0.22)
Year	DFTD	-1.06 (-3.43 - 0.53)	52.34 (5.64 - 173.06)
	Body Weight	0.38 (0.13 - 0.68)	-0.61 (-0.950.19)
Additive genetic	DFTD	-2.56 (-6.110.50)	75.40 (7.01 - 256.83)
Permanent	Body Weight	0.26 (0.05 - 0.48)	0.07 (-0.85 - 0.89)
environment	DFTD	0.05 (-0.84 - 0.93)	6.46 (0.02 - 26.36)
-	Body Weight	0.14 (0.05 - 0.25)	0.43 (-0.47 - 0.96)
Irap	DFTD	0.29 (-0.12 - 1.02)	4.73 (0.02-17.76)
		Relative Head width	Relative DFTD
Voor	Relative Head width	0.11 (0.05 - 0.22)	0.11 (-0.34 - 0.56)
Tedi	Relative DFTD	0.10 (-0.37 - 0.61)	8.98 (4.24 - 16.43)
Additive genetic	Relative Head width	0.08 (0.01 - 0.17)	-0.22 (-0.78 - 0.32)
Additive genetic	Relative DFTD	-0.09 (-0.34 - 0.13)	3.13 (1.47 - 4.96)

Permanent	Relative Head width	0.14 (0.07 - 0.21)	0.27 (-0.50 - 0.89)
environment	Relative DFTD	0.07 (-0.08 - 0.26)	0.72 (0.01 - 2.01)
Tran	Relative Head width	0.01 (0.0001 - 0.02)	0.41 (-0.57 - 0.96)
пар	Relative DFTD	0.02 (-0.01 - 0.07)	0.33 (0.01 - 0.84)
Posidual	Relative Head width	0.12 (0.10 - 0.14)	-0.16 (-0.270.04)
Residual	Relative DFTD	-0.13 (-0.220.04)	5.28 (4.44 - 6.19)
		Relative Body Weight	Relative DFTD
Voor	Relative Body Weight	0.07 (0.03 - 0.13)	-0.38 (-0.75 - 0.10)
Teal	Relative DFTD	-0.29 (-0.78 - 0.06)	8.98 (4.29 - 16.43)
	Relative Body Weight	0.11 (0.03 - 0.19)	-0.52 (-0.920.09)
Additive genetic	Relative DFTD	-0.27 (-0.490.05)	2.95 (1.36 - 4.81)
Permanent	Relative Body Weight	0.07 (0.01 - 0.13)	0.29 (-0.54 - 0.92)
environment	Relative DFTD	0.06 (-0.08 - 0.20)	0.82 (0.01 - 2.13)
Tran	Relative Body Weight	0.04 (0.01 - 0.07)	0.47 (-0.32 - 0.96)
пар	Relative DFTD	0.04 (-0.01 - 0.12)	0.28 (0.01 - 0.76)
Desidual	Relative Body Weight	0.15 (0.13 - 0.18)	-0.12 (-0.230.002)
Residual	Relative DFTD	-0.11 (-0.210.001)	5.31 (4.47 - 6.24)

- **Figure S4.** Genetic correlations (*COR*_A), permanent environment correlation (*COR*_{PE}), year correlation (*COR*_{Year}) and spatial correlation (*COR*_{TrapID}) between head width and body weight, DFTD and head width and DFTD and body weight estimated from (a) models fit with DFTD as a binary variable ("Original"), and (b) models fit with a relative measure of DFTD (DFTD relative to the average prevalence of DFTD in the population, see methods for details), which fit with gaussian errors ("Relative"). Covariance estimates can be found in Table S3. Posterior median of estimates shown as point, with 75% CIs shown as heavy lines
- and 95% CIs as lighter lines.



833 <u>References</u>

834

- Ali, O. A., O'Rourke, S. M., Amish, S. J., Meek, M. H., Luikart, G., Jeffres, C., & Miller, M. R.
- 836 (2016). RAD Capture (Rapture): Flexible and Efficient Sequence-Based Genotyping.
- 837 *Genetics*, 202(2), 389–400. https://doi.org/10.1534/genetics.115.183665
- 838 Altizer, S., Harvell, D., & Friedle, E. (2003). Rapid evolutionary dynamics and disease threats
- to biodiversity. *Trends in Ecology & Evolution*, *18*(11), 589–596.
- 840 https://doi.org/10.1016/j.tree.2003.08.013
- 841 Bérénos, C., Ellis, P. A., Pilkington, J. G., & Pemberton, J. M. (2014). Estimating quantitative
- 842 genetic parameters in wild populations: A comparison of pedigree and genomic
- 843 approaches. *Molecular Ecology*, *23*(14), 3434–3451.
- 844 https://doi.org/10.1111/mec.12827
- 845 Best, A., White, A., & Boots, M. (2008). Maintenance of host variation in tolerance to
- pathogens and parasites. *Proceedings of the National Academy of Sciences*, 105(52),
- 847 20786–20791. https://doi.org/10.1073/pnas.0809558105
- 848 Boots, M., Best, A., Miller, M. R., & White, A. (2009). The role of ecological feedbacks in the
- 849 evolution of host defence: What does theory tell us? *Philosophical Transactions of*

850 the Royal Society B: Biological Sciences, 364(1513), 27–36.

- 851 https://doi.org/10.1098/rstb.2008.0160
- 852 Brüniche-Olsen, A., Burridge, C. P., Austin, J. J., & Jones, M. E. (2013). Disease induced
- 853 changes in gene flow patterns among Tasmanian devil populations. *Biological*
- 854 *Conservation*, *165*, 69–78. https://doi.org/10.1016/j.biocon.2013.05.014
- 855 Brüniche-Olsen, A., Jones, M. E., Austin, J. J., Burridge, C. P., & Holland, B. R. (2014).
- 856 Extensive population decline in the Tasmanian devil predates European settlement

- and devil facial tumour disease. *Biology Letters*, *10*(11), 20140619.
- 858 https://doi.org/10.1098/rsbl.2014.0619
- 859 Bulmer, M. G. (1971). The Effect of Selection on Genetic Variability. The American

860 *Naturalist*, *105*(943), 201–211. https://doi.org/10.1086/282718

- Bürkner, P.-C. (2017). brms: An R package for Bayesian multilevel models using Stan. *Journal*of Statistical Software, 80(1), 1–28.
- Bürkner, P.-C. (2021). Bayesian Item Response Modeling in R with brms and Stan. *Journal of Statistical Software*, *100*(5), 1–54. https://doi.org/10.18637/jss.v100.i05
- 865 Catchen, J., Hohenlohe, P. A., Bassham, S., Amores, A., & Cresko, W. A. (2013). Stacks: An
- analysis tool set for population genomics. *Molecular Ecology*, 22(11), 3124–3140.
- 867 https://doi.org/10.1111/mec.12354
- Charlesworth, D., & Willis, J. H. (2009). The genetics of inbreeding depression. *Nature Reviews Genetics*, *10*(11), 783–796. https://doi.org/10.1038/nrg2664
- 870 Coltman, D. W., Pilkington, J., Kruuk, L. E. B., Wilson, K., & Pemberton, J. M. (2001). Positive
- 871 genetic correlation between parasite resistance and body size in a free-living
- ungulate population. *Evolution*, 55(10), 2116–2125. https://doi.org/10.1111/j.0014-
- 873 3820.2001.tb01326.x
- 874 Cristescu, R. H., Strickland, K., Schultz, A. J., Kruuk, L. E. B., de Villiers, D., & Frère, C. H.
- 875 (2022). Susceptibility to a sexually transmitted disease in a wild koala population
- shows heritable genetic variance but no inbreeding depression. *Molecular Ecology*,
- 877 31(21), 5455–5467. https://doi.org/10.1111/mec.16676
- 878 Cunningham, C. X., Comte, S., McCallum, H., Hamilton, D. G., Hamede, R., Storfer, A.,
- 879 Hollings, T., Ruiz-Aravena, M., Kerlin, D. H., Brook, B. W., Hocking, G., & Jones, M. E.
- 880 (2021). Quantifying 25 years of disease-caused declines in Tasmanian devil

- 881 populations: Host density drives spatial pathogen spread. *Ecology Letters*, 24(5),
- 882 958–969. https://doi.org/10.1111/ele.13703
- 883 Daszak, P., Cunningham, A., & Hyatt, A. D. (2000). Emerging Infectious Diseases of Wildlife-
- Threats to Biodiversity and Human Health. *Science*, *287*(5452), 443–449.
- 885 https://doi.org/10.1126/science.287.5452.443
- de Villemereuil, P., Schielzeth, H., Nakagawa, S., & Morrissey, M. (2016). General Methods
- 887 for Evolutionary Quantitative Genetic Inference from Generalized Mixed Models.
- 888 *Genetics*, 204(3), 1281–1294. https://doi.org/10.1534/genetics.115.186536 %J
- 889 Genetics
- B90 DeRose, M. A., & Roff, D. A. (1999). A COMPARISON OF INBREEDING DEPRESSION IN LIFE-
- HISTORY AND MORPHOLOGICAL TRAITS IN ANIMALS. *Evolution*, *53*(4), 1288–1292.
 https://doi.org/10.1111/j.1558-5646.1999.tb04541.x
- 893 Epstein, B., Jones, M., Hamede, R., Hendricks, S., McCallum, H., Murchison, E. P., Schönfeld,
- 894 B., Wiench, C., Hohenlohe, P., & Storfer, A. (2016). Rapid evolutionary response to a
- transmissible cancer in Tasmanian devils. *Nature Communications*, 7(1), 12684.
- 896 https://doi.org/10.1038/ncomms12684
- Falconer, D. S., & Mackay, T. F. C. (1996). Introduction to quantitative genetics. *Introduction to Quantitative Genetics.*, *Ed. 4*.
- Fisher, M. C., & Garner, T. W. J. (2020). Chytrid fungi and global amphibian declines. *Nature Reviews Microbiology*, *18*(6), 332–343. https://doi.org/10.1038/s41579-020-0335-x
- 901 Fraik, A. K., Margres, M. J., Epstein, B., Barbosa, S., Jones, M., Hendricks, S., Schönfeld, B.,
- 902 Stahlke, A. R., Veillet, A., Hamede, R., McCallum, H., Lopez-Contreras, E., Kallinen, S.
- 903 J., Hohenlohe, P. A., Kelley, J. L., & Storfer, A. (2020). Disease swamps molecular
- 904 signatures of genetic-environmental associations to abiotic factors in Tasmanian

- 905 devil (Sarcophilus harrisii) populations. *Evolution*, 74(7), 1392–1408.
- 906 https://doi.org/10.1111/evo.14023
- 907 Gervais, L., Perrier, C., Bernard, M., Merlet, J., Pemberton, J. M., Pujol, B., & Quéméré, E.
- 908 (2019). RAD-sequencing for estimating genomic relatedness matrix-based heritability
- 909 in the wild: A case study in roe deer. *Molecular Ecology Resources*, 19(5), 1205–1217.
- 910 https://doi.org/10.1111/1755-0998.13031
- 911 Gienapp, P., Fior, S., Guillaume, F., Lasky, J. R., Sork, V. L., & Csilléry, K. (2017). Genomic
- 912 Quantitative Genetics to Study Evolution in the Wild. Trends in Ecology & Evolution,
- 913 32(12), 897–908. https://doi.org/10.1016/j.tree.2017.09.004
- Gleeson, D. J., Blows, M. W., & Owens, I. P. (2005). Genetic covariance between indices of
- 915 body condition and immunocompetence in a passerine bird. *BMC Evolutionary*916 *Biology*, 5(1), 61. https://doi.org/10.1186/1471-2148-5-61
- 917 Golas, B. D., Goodell, B., & Webb, C. T. (2021). Host adaptation to novel pathogen
- 918 introduction: Predicting conditions that promote evolutionary rescue. *Ecology*
- 919 *Letters*, 24(10), 2238–2255.
- 920 Gooley, R., Hogg, C. J., Belov, K., & Grueber, C. E. (2017). No evidence of inbreeding
- 921 depression in a Tasmanian devil insurance population despite significant variation in
- 922 inbreeding. *Scientific Reports*, 7(1), 1830. https://doi.org/10.1038/s41598-017923 02000-y
- ----,
- 924 Gooley, R. M., Hogg, C. J., Fox, S., Pemberton, D., Belov, K., & Grueber, C. E. (2020).
- 925 Inbreeding depression in one of the last DFTD-free wild populations of Tasmanian
- 926 devils. *PeerJ*, e9220. https://doi.org/10.7717/peerj.9220

- 927 Grossen, C., Guillaume, F., Keller, L. F., & Croll, D. (2020). Purging of highly deleterious
- 928 mutations through severe bottlenecks in Alpine ibex. *Nature Communications*, *11*(1),
- 929 1001. https://doi.org/10.1038/s41467-020-14803-1
- Hajduk, G. K., Cockburn, A., Margraf, N., Osmond, H. L., Walling, C. A., & Kruuk, L. E. B.
- 931 (2018). Inbreeding, inbreeding depression, and infidelity in a cooperatively breeding
- 932 bird*. *Evolution*, 72(7), 1500–1514. https://doi.org/10.1111/evo.13496
- Hamede, R. K., Bashford, J., McCallum, H., & Jones, M. (2009). Contact networks in a wild
- 934 Tasmanian devil (Sarcophilus harrisii) population: Using social network analysis to
- 935 reveal seasonal variability in social behaviour and its implications for transmission of
- 936 devil facial tumour disease. *Ecology Letters*, *12*(11), 1147–1157.
- 937 https://doi.org/10.1111/j.1461-0248.2009.01370.x
- 938 Hamede, R. K., McCallum, H., & Jones, M. (2008). Seasonal, demographic and density-
- 939 related patterns of contact between Tasmanian devils (Sarcophilus harrisii):
- 940 Implications for transmission of devil facial tumour disease. *Austral Ecology*, 33(5),
- 941 614–622. https://doi.org/10.1111/j.1442-9993.2007.01827.x
- 942 Hamede, R. K., McCallum, H., & Jones, M. (2013). Biting injuries and transmission of
- 943 Tasmanian devil facial tumour disease. *Journal of Animal Ecology*, *82*(1), 182–190.
- 944 https://doi.org/10.1111/j.1365-2656.2012.02025.x
- Hamede, R. K., Owen, R., Siddle, H., Peck, S., Jones, M., Dujon, A. M., Giraudeau, M., Roche,
- 946 B., Ujvari, B., & Thomas, F. (2020). The ecology and evolution of wildlife cancers:
- 947 Applications for management and conservation. *Evolutionary Applications*, 13(7),
- 948 1719–1732. https://doi.org/10.1111/eva.12948
- 949 Hamede, R. K., Pearse, A.-M., Swift, K., Barmuta, L. A., Murchison, E. P., & Jones, M. E.
- 950 (2015). Transmissible cancer in Tasmanian devils: Localized lineage replacement and

- 951 host population response. Proceedings of the Royal Society B: Biological Sciences,
- 952 *282*(1814), 20151468. https://doi.org/10.1098/rspb.2015.1468
- 953 Hamilton, D. G., Jones, M. E., Cameron, E. Z., Kerlin, D. H., McCallum, H., Storfer, A.,
- 954 Hohenlohe, P. A., & Hamede, R. K. (2020). Infectious disease and sickness behaviour:
- 955 Tumour progression affects interaction patterns and social network structure in wild
- 956 Tasmanian devils. *Proceedings of the Royal Society B: Biological Sciences*, 287(1940),
- 957 20202454. https://doi.org/10.1098/rspb.2020.2454
- 958 Hamilton, D. G., Jones, M. E., Cameron, E. Z., McCallum, H., Storfer, A., Hohenlohe, P. A., &
- 959 Hamede, R. K. (2019). Rate of intersexual interactions affects injury likelihood in
- 960 Tasmanian devil contact networks. *Behavioral Ecology*, *30*(4), 1087–1095.
- 961 https://doi.org/10.1093/beheco/arz054
- 962 Hayward, A. D., Nussey, D. H., Wilson, A. J., Berenos, C., Pilkington, J. G., Watt, K. A.,
- 963 Pemberton, J. M., & Graham, A. L. (2014). Natural Selection on Individual Variation in
- 964 Tolerance of Gastrointestinal Nematode Infection. *PLOS Biology*, *12*(7), e1001917.
- 965 https://doi.org/10.1371/journal.pbio.1001917
- 966 Healy, K., Ezard, T. H. G., Jones, O. R., Salguero-Gómez, R., & Buckley, Y. M. (2019). Animal
- 967 life history is shaped by the pace of life and the distribution of age-specific mortality
- 968 and reproduction. *Nature Ecology & Evolution*, *3*(8), 1217–1224.
- 969 https://doi.org/10.1038/s41559-019-0938-7
- 970 Hedrick, P. W., & Garcia-Dorado, A. (2016). Understanding Inbreeding Depression, Purging,
- and Genetic Rescue. *Trends in Ecology & Evolution*, *31*(12), 940–952.
- 972 https://doi.org/10.1016/j.tree.2016.09.005

- 973 Hedrick, P. W., & Kalinowski, S. T. (2000). Inbreeding Depression in Conservation Biology.
- 974 Annual Review of Ecology and Systematics, 31(1), 139–162.
- 975 https://doi.org/10.1146/annurev.ecolsys.31.1.139
- 976 Herrera, J., & Nunn, C. L. (2019). Behavioural ecology and infectious disease: Implications for
- 977 conservation of biodiversity. *Philosophical Transactions of the Royal Society B:*
- 978 Biological Sciences, 374(1781), 20180054. https://doi.org/10.1098/rstb.2018.0054
- 979 Hoffmann, A. A., Sgrò, C. M., & Kristensen, T. N. (2017). Revisiting Adaptive Potential,
- 980 Population Size, and Conservation. *Trends in Ecology & Evolution*, *32*(7), 506–517.
- 981 https://doi.org/10.1016/j.tree.2017.03.012
- 982 Hoffmann, C., Zimmermann, F., Biek, R., Kuehl, H., Nowak, K., Mundry, R., Agbor, A.,
- 983 Angedakin, S., Arandjelovic, M., Blankenburg, A., Brazolla, G., Corogenes, K., Couacy-
- 984 Hymann, E., Deschner, T., Dieguez, P., Dierks, K., Düx, A., Dupke, S., Eshuis, H., ...
- 985 Leendertz, F. H. (2017). Persistent anthrax as a major driver of wildlife mortality in a
- 986 tropical rainforest. *Nature*, *548*(7665), 82–86. https://doi.org/10.1038/nature23309
- 987 Hoyt, J. R., Kilpatrick, A. M., & Langwig, K. E. (2021). Ecology and impacts of white-nose
- 988 syndrome on bats. *Nature Reviews Microbiology*, *19*(3), 196–210.
- 989 https://doi.org/10.1038/s41579-020-00493-5
- 990 Huisman, J. (2017). Pedigree reconstruction from SNP data: Parentage assignment, sibship
- 991 clustering and beyond. *Molecular Ecology Resources*, 17(5), 1009–1024.
- 992 https://doi.org/10.1111/1755-0998.12665
- Huisman, J., Kruuk, L. E. B., Ellis, P. A., Clutton-Brock, T., & Pemberton, J. M. (2016).
- 994 Inbreeding depression across the lifespan in a wild mammal population. *Proceedings*
- 995 *of the National Academy of Sciences*, *113*(13), 3585–3590.
- 996 https://doi.org/10.1073/pnas.1518046113

- 997 James, C, Pemberton, Josephine, Navarro, Pau, & Knott, Sara. (2023). Investigating pedigree-
- and SNP-associated components of heritability in a wild population of Soay sheep.

bioRxiv, 2023.06.02.543397. https://doi.org/10.1101/2023.06.02.543397

- 1000 Jones, M. E. (2023). Over-eruption in marsupial carnivore teeth: Compensation for a
- 1001 constraint. Proceedings of the Royal Society B: Biological Sciences, 290(2013),
- 1002 20230644. https://doi.org/10.1098/rspb.2023.0644
- Jones, M. E., Cockburn, A., Hamede, R., Hawkins, C., Hesterman, H., Lachish, S., Mann, D.,
- 1004 McCallum, H., & Pemberton, D. (2008). Life-history change in disease-ravaged
- 1005 Tasmanian devil populations. *Proceedings of the National Academy of Sciences*,
- 1006 *105*(29), 10023–10027. https://doi.org/10.1073/pnas.0711236105
- 1007 Kirkpatrick, M., & Jarne, P. (2000). The Effects of a Bottleneck on Inbreeding Depression and
 1008 the Genetic Load. *The American Naturalist*, 155(2), 154–167.
- 1009 https://doi.org/10.1086/303312
- 1010 Kruuk, L. E. B. (2004). Estimating genetic parameters in natural populations using the 'animal

1011 model'. *Proceedings of the Royal Society of London, Series B.*, 359, 873–890.

- 1012 Kruuk, L. E. B., & Hadfield, J. D. (2007). How to separate genetic and environmental causes
- 1013 of similarity between relatives. *Journal of Evolutionary Biology*, *20*(5), 1890–1903.
- 1014 https://doi.org/10.1111/j.1420-9101.2007.01377.x
- 1015 Lachish, S., Jones, M., & McCallum, H. (2007). The Impact of Disease on the Survival and
- 1016 Population Growth Rate of the Tasmanian Devil. *Journal of Animal Ecology*, 76(5),
- 1017 926–936. JSTOR.
- 1018 Lachish, S., McCallum, H., & Jones, M. (2009). Demography, Disease and the Devil: Life-
- 1019 History Changes in a Disease-Affected Population of Tasmanian Devils (Sarcophilus
- 1020 harrisii). *Journal of Animal Ecology*, 78(2), 427–436. JSTOR.

- 1021 Laikre, L., & Ryman, N. (1991). Inbreeding Depression in a Captive Wolf (Canis lupus)
- 1022 Population. *Conservation Biology*, *5*(1), 33–40. https://doi.org/10.1111/j.15231023 1739.1991.tb00385.x
- Lande, R., & Arnold, S. J. (1983). The Measurement of Selection on Correlated Characters.
- 1025 *Evolution*, *37*(6), 1210–1226. JSTOR. https://doi.org/10.2307/2408842
- 1026 Margres, M. J., Jones, M. E., Epstein, B., Kerlin, D. H., Comte, S., Fox, S., Fraik, A. K.,
- 1027 Hendricks, S. A., Huxtable, S., Lachish, S., Lazenby, B., O'Rourke, S. M., Stahlke, A. R.,
- 1028 Wiench, C. G., Hamede, R., Schönfeld, B., McCallum, H., Miller, M. R., Hohenlohe, P.
- 1029 A., & Storfer, A. (2018). Large-effect loci affect survival in Tasmanian devils
- 1030 (Sarcophilus harrisii) infected with a transmissible cancer. *Molecular Ecology*, 27(21),
- 1031 4189–4199. https://doi.org/10.1111/mec.14853
- 1032 Marjamäki, P. H., Dugdale, H. L., Delahay, R., McDonald, R. A., & Wilson, A. J. (2021).
- 1033 Genetic, social and maternal contributions to Mycobacterium bovis infection status
- in European badgers (Meles meles). *Journal of Evolutionary Biology*, *34*(4), 695–709.
- 1035 Martin, A. M., Cassirer, E. F., Waits, L. P., Plowright, R. K., Cross, P. C., & Andrews, K. R.
- 1036 (2021). Genomic association with pathogen carriage in bighorn sheep (Ovis
- 1037 canadensis). *Ecology and Evolution*, *11*(6), 2488–2502.
- 1038 McCallum, H. (2008). Tasmanian devil facial tumour disease: Lessons for conservation
- biology. *Trends in Ecology & Evolution*, 23(11), 631–637.
- 1040 https://doi.org/10.1016/j.tree.2008.07.001
- 1041 McCallum, H., Tompkins, D. M., Jones, M., Lachish, S., Marvanek, S., Lazenby, B., Hocking,
- 1042 G., Wiersma, J., & Hawkins, C. E. (2007). Distribution and Impacts of Tasmanian Devil
- 1043 Facial Tumor Disease. *EcoHealth*, 4(3), 318–325. https://doi.org/10.1007/s10393-

1044 007-0118-0

1045	Medzhitov, R., Schneider, D. S., & Soares, M. P. (2012). Disease Tolerance as a Defense
1046	Strategy. <i>Science</i> , <i>335</i> (6071), 936–941. https://doi.org/10.1126/science.1214935
1047	Murchison, E. P., Schulz-Trieglaff, O. B., Ning, Z., Alexandrov, L. B., Bauer, M. J., Fu, B., Hims,
1048	M., Ding, Z., Ivakhno, S., Stewart, C., Ng, B. L., Wong, W., Aken, B., White, S., Alsop,
1049	A., Becq, J., Bignell, G. R., Cheetham, R. K., Cheng, W., Stratton, M. R. (2012).
1050	Genome Sequencing and Analysis of the Tasmanian Devil and Its Transmissible
1051	Cancer. Cell, 148(4), 780–791. https://doi.org/10.1016/j.cell.2011.11.065
1052	Murchison, E. P., Tovar, C., Hsu, A., Bender, H. S., Kheradpour, P., Rebbeck, C. A., Obendorf,
1053	D., Conlan, C., Bahlo, M., Blizzard, C. A., Pyecroft, S., Kreiss, A., Kellis, M., Stark, A.,
1054	Harkins, T. T., Graves, J. A. M., Woods, G. M., Hannon, G. J., & Papenfuss, A. T.
1055	(2010). The Tasmanian Devil Transcriptome Reveals Schwann Cell Origins of a
1056	Clonally Transmissible Cancer. Science, 327(5961), 84–87.
1057	https://doi.org/10.1126/science.1180616
1058	Nielsen, J. F., English, S., Goodall-Copestake, W. P., Wang, J., Walling, C. A., Bateman, A. W.,
1059	Flower, T. P., Sutcliffe, R. L., Samson, J., Thavarajah, N. K., Kruuk, L. E. B., Clutton-
1060	Brock, T. H., & Pemberton, J. M. (2012). Inbreeding and inbreeding depression of
1061	early life traits in a cooperative mammal. <i>Molecular Ecology</i> , 21(11), 2788–2804.
1062	https://doi.org/10.1111/j.1365-294X.2012.05565.x
1063	O'Grady, J. J., Brook, B. W., Reed, D. H., Ballou, J. D., Tonkyn, D. W., & Frankham, R. (2006).
1064	Realistic levels of inbreeding depression strongly affect extinction risk in wild
1065	populations. <i>Biological Conservation</i> , 133(1), 42–51.
1066	https://doi.org/10.1016/i.biocon.2006.05.016

- 1067 Patton, A. H., Lawrance, M. F., Margres, M. J., Kozakiewicz, C. P., Hamede, R., Ruiz-Aravena,
- 1068 M., Hamilton, D. G., Comte, S., Ricci, L. E., Taylor, R. L., Stadler, T., Leaché, A.,

- 1069 McCallum, H., Jones, M. E., Hohenlohe, P. A., & Storfer, A. (2020). A transmissible
- 1070 cancer shifts from emergence to endemism in Tasmanian devils. *Science*, *370*(6522),
- 1071 eabb9772. https://doi.org/10.1126/science.abb9772
- 1072 Price, G. R. (1970). Selection and covariance. *Nature*, 227, 520–521. CABDirect.
- 1073 https://doi.org/10.1038/227520a0
- 1074 Pye, R., Hamede, R., Siddle, H. V., Caldwell, A., Knowles, G. W., Swift, K., Kreiss, A., Jones, M.
- 1075 E., Lyons, A. B., & Woods, G. M. (2016). Demonstration of immune responses against
- 1076 devil facial tumour disease in wild Tasmanian devils. *Biology Letters*, 12(10),
- 1077 20160553. https://doi.org/10.1098/rsbl.2016.0553
- 1078 Rarberg, L., & Stjernman, M. (2003). Natural selection on immune responsiveness in blue tits
- 1079 Parus caeruleus. *Evolution*, *57*(7), 1670–1678. https://doi.org/10.1111/j.0014-
- 1080 3820.2003.tb00372.x
- 1081 Reid, J. M., Arcese, P., & Keller, L. F. (2003). Inbreeding depresses immune response in song
- 1082 sparrows (Melospiza melodia): Direct and inter–generational effects. *Proceedings of*
- 1083 the Royal Society of London. Series B: Biological Sciences, 270(1529), 2151–2157.
- 1084 https://doi.org/10.1098/rspb.2003.2480
- 1085 Rigby, M. C., Hechinger, R. F., & Stevens, L. (2002). Why should parasite resistance be
- 1086 costly? *Trends in Parasitology*, *18*(3), 116–120. https://doi.org/10.1016/S1471-
- 1087 4922(01)02203-6
- Robertson, A., & Lewontin, R. (1968). Population biology and evolution. *The Spectrum of Genetic Variation. Syracuse Univ. Press, New York*, 5–16.
- 1090 Rogalski, M. A., Gowler, C. D., Shaw, C. L., Hufbauer, R. A., & Duffy, M. A. (2017). Human
- 1091 drivers of ecological and evolutionary dynamics in emerging and disappearing

- 1092 infectious disease systems. *Philosophical Transactions of the Royal Society B:*
- 1093 *Biological Sciences*, *372*(1712), 20160043. https://doi.org/10.1098/rstb.2016.0043
- 1094 Ross-Gillespie, A., O'Riain, M. J., & Keller, L. F. (2007). VIRAL EPIZOOTIC REVEALS
- 1095 INBREEDING DEPRESSION IN A HABITUALLY INBREEDING MAMMAL. *Evolution*, 61(9),
- 1096 2268–2273. https://doi.org/10.1111/j.1558-5646.2007.00177.x
- 1097 Sánchez, C. A., Becker, D. J., Teitelbaum, C. S., Barriga, P., Brown, L. M., Majewska, A. A.,
- 1098 Hall, R. J., & Altizer, S. (2018). On the relationship between body condition and
- 1099 parasite infection in wildlife: A review and meta-analysis. *Ecology Letters*, 21(12),
- 1100 1869–1884. https://doi.org/10.1111/ele.13160
- 1101 Schrag, S. J., & Wiener, P. (1995). Emerging infectious disease: What are the relative roles of
- ecology and evolution? *Trends in Ecology & Evolution*, *10*(8), 319–324.
- 1103 https://doi.org/10.1016/S0169-5347(00)89118-1
- 1104 Silk, M. J., & Hodgson, D. J. (2021). Life history and population regulation shape
- 1105 demographic competence and influence the maintenance of endemic disease.
- 1106 *Nature Ecology & Evolution*, *5*(1), 82–91. https://doi.org/10.1038/s41559-020-
- 1107 01333-8
- 1108 Spielman, D., Brook, B. W., Briscoe, D. A., & Frankham, R. (2004). Does Inbreeding and Loss

1109 of Genetic Diversity Decrease Disease Resistance? *Conservation Genetics*, 5(4), 439–

1110 448. https://doi.org/10.1023/B:COGE.0000041030.76598.cd

- 1111 Stahlke, A. R., Epstein, B., Barbosa, S., Margres, M. J., Patton, A. H., Hendricks, S. A., Veillet,
- 1112 A., Fraik, A. K., Schönfeld, B., McCallum, H. I., Hamede, R., Jones, M. E., Storfer, A., &
- 1113 Hohenlohe, P. A. (2021). Contemporary and historical selection in Tasmanian devils
- 1114 (Sarcophilus harrisii) support novel, polygenic response to transmissible cancer.

- 1115 Proceedings of the Royal Society B: Biological Sciences, 288(1951), 20210577.
- 1116 https://doi.org/10.1098/rspb.2021.0577
- 1117 Stammnitz, M. R., Gori, K., Kwon, Y. M., Harry, E., Martin, F. J., Billis, K., Cheng, Y., Baez-
- 1118 Ortega, A., Chow, W., Comte, S., Eggertsson, H., Fox, S., Hamede, R., Jones, M.,
- 1119 Lazenby, B., Peck, S., Pye, R., Quail, M. A., Swift, K., ... Murchison, E. P. (2023). The
- evolution of two transmissible cancers in Tasmanian devils. Science, 380(6642), 283–
- 1121 293. https://doi.org/10.1126/science.abq6453
- 1122 Trinkel, M., Cooper, D., Packer, C., & Slotow, R. (2011). INBREEDING DEPRESSION INCREASES
- 1123 SUSCEPTIBILITY TO BOVINE TUBERCULOSIS IN LIONS: AN EXPERIMENTAL TEST USING
- 1124 AN INBRED–OUTBRED CONTRAST THROUGH TRANSLOCATION. Journal of Wildlife
- 1125 *Diseases*, 47(3), 494–500. https://doi.org/10.7589/0090-3558-47.3.494
- 1126 Valenzuela-Sánchez, A., Wilber, M. Q., Canessa, S., Bacigalupe, L. D., Muths, E., Schmidt, B.
- 1127 R., Cunningham, A. A., Ozgul, A., Johnson, P. T. J., & Cayuela, H. (2021). Why disease
- ecology needs life-history theory: A host perspective. *Ecology Letters*, 24(4), 876–
- 1129 890. https://doi.org/10.1111/ele.13681
- 1130 Walsh, B., & Lynch, M. (2018). *Evolution and selection of quantitative traits*. Oxford
- 1131 University Press.
- 1132 Wang, J. (2010). coancestry: A program for simulating, estimating and analysing relatedness
- and inbreeding coefficients. *Molecular Ecology Resources*, *11*(1), 141–145.
- 1134 https://doi.org/10.1111/j.1755-0998.2010.02885.x
- 1135 Wilson, A. J., Coltman, D. W., Pemberton, J. M., Overall, A. D. J., Byrne, K. A., & Kruuk, L. E. B.
- 1136 (2005). Maternal genetic effects set the potential for evolution in a free-living
- 1137 vertebrate population. *Journal of Evolutionary Biology*, *18*(2), 405–414.
- 1138 https://doi.org/10.1111/j.1420-9101.2004.00824.x

1139	Wilson, A. J., Réale,	D., Clements,	M. N., Morrissev	y, M. M.,	Postma, E.	, Walling,	C. A.,	Kruuk,
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1140 L. E. B., & Nussey, D. H. (2010). An ecologist's guide to the animal model. *Journal of*

1141 *Animal Ecology*, *79*(1), 13–26. https://doi.org/10.1111/j.1365-2656.2009.01639.x

- 1142 Yang, J., Lee, S. H., Goddard, M. E., & Visscher, P. M. (2011). GCTA: a tool for genome-wide
- 1143 complex trait analysis. *American Journal of Human Genetics*, 88(1), 76–82.
- 1144 https://doi.org/10.1016/j.ajhg.2010.11.011
- 1145 Zaitlen, N., Kraft, P., Patterson, N., Pasaniuc, B., Bhatia, G., Pollack, S., & Price, A. L. (2013).
- 1146 Using Extended Genealogy to Estimate Components of Heritability for 23
- 1147 Quantitative and Dichotomous Traits. *PLOS Genetics*, *9*(5), e1003520.
- 1148 https://doi.org/10.1371/journal.pgen.1003520