

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23

Satellite derived trait data slightly improves tropical forest biomass, NPP and GPP estimates

Christopher E. Doughty*¹, Camille Gaillard¹, Patrick Burns¹, Yadvinder Malhi², Alexander Shenkin¹, David Minor³, Laura Duncanson³, Jesus Aguirre-Gutierrez², Scott Goetz¹, Hao Tang⁴

¹School of Informatics, Computing, and Cyber Systems, Northern Arizona University, Flagstaff, AZ, USA

²Environmental Change Institute, School of Geography and the Environment, University of Oxford, Oxford, UK

³Geographical Sciences, University of Maryland College Park, Maryland , USA

⁴Department of Geography, National University of Singapore, Singapore

*to whom correspondence should be addressed – chris.doughty@nau.edu

Keywords –GEDI, tropical forests, traits, LMA, biomass

24

25 **Abstract**

26 Improving tropical forest current biomass estimates can help more accurately evaluate ecosystem
27 services in tropical forests. The Global Ecosystem Dynamics Investigation (GEDI) lidar provides
28 detailed 3D forest structure and height data, which can be used to improve above-ground
29 biomass estimates. However, there is still debate on how best to predict tropical forest biomass
30 using GEDI data. Here we compare stand biomass predicted by GEDI data with the observed
31 data of 2,102 inventory plots in tropical forests and find that adding a remotely sensed (RS) trait
32 map of LMA (Leaf Mass Area) significantly ($P < 0.001$) improves field biomass predictions, but
33 by only a small amount ($r^2 = 0.01$). However, it may also help reduce the bias of the residuals
34 because there was a negative relationship between both LMA (r^2 of 0.34) and percentage of
35 phosphorus (%P, $r^2 = 0.31$) and residuals. Leaf spectral data (400-1075 nm) from 523 individual
36 trees along a Peruvian tropical forest elevation gradient predicted Diameter at Breast height
37 (DBH) (the critical measurement underlying plot biomass) with an $r^2 = 0.01$ and LMA predicts
38 DBH with an $r^2 = 0.04$. Other datasets may offer further improvements and max temperature
39 (T_{\max}) predicts Amazonian biomass residuals with an r^2 of 0.76 ($N = 66$). Finally, for a network
40 of net primary production (NPP) and gross primary production (GPP) plots ($N = 21$), leaf traits
41 predicted with remote sensing are better at predicting fluxes than structure variables. Overall,
42 trait maps, especially future improved ones produced by Surface Biology Geology (SBG), may
43 improve biomass and carbon flux predictions by a small but significant amount.

44

45

46 **Introduction**

47 In an era of rapid climate change, accurately predicting forest carbon stocks is
48 increasingly important because carbon stored in forests can potentially offset anthropogenic
49 emissions that cause climate change. For this reason, international climate agreements such as
50 REDD+ (Reducing Emissions from Deforestation and Degradation) have been developed to
51 encourage countries to conserve their forests (Goetz et al., 2015). Using forests as natural
52 climate change solutions, by incentivizing carbon trading and offset schemes, requires accurate
53 and repeatable measurements of forest aboveground biomass (AGB) (CEOS, 2014, Goetz et al.,
54 2015). Earth observation satellite remote sensing (RS), coupled with ground-based
55 measurements, have the potential to provide systematic estimates of AGB over vast spatial
56 extents. Therefore, much effort has been put into developing such maps of AGB, albeit with
57 mixed results. For instance, two remotely sensed biomass maps showed markedly different
58 biomass trends from each other and from 413 ground plots (Avitabile et al., 2016; Baccini et al.,
59 2012; Mitchard et al., 2014; Saatchi et al., 2011). Mitchard et al (2013 and 2014) found the
60 uncertainties were actually > 25% more than those listed in the RS maps of Baccini et al (2012)
61 and Saatchi et al (2011) (Mitchard et al., 2013, 2014). They advise to incorporate basal area-
62 weighted wood density estimates and note that depending only on the relationships between tree
63 height and biomass may lead to large, spatially correlated errors. Partially in response to such
64 difficulties in predicting biomass with optical RS, the Global Ecosystem Dynamics Investigation
65 (GEDI) Lidar mission was launched and installed on the International Space Station (ISS) in late
66 2018 and operational products started in March 2019 (R. Dubayah et al., 2020). GEDI is the first
67 spaceborne lidar designed for terrestrial ecosystem research and the first specifically developed
68 to accurately measure forest canopy 3D structure. However, converting from laser energy
69 returns to accurate biomass predictions is not trivial.

70 GEDI covers most land areas below 52 degrees latitude, but it does not provide wall to
71 wall coverage and gaps between GEDI tracks are greatest at tropical latitudes owing to the
72 orbital configuration of the ISS (R. Dubayah et al., 2022). To develop pre-launch calibrated
73 models of AGB, ground biomass plots were combined with coincident aircraft lidar data using a
74 waveform simulator (Hancock et al., 2019) to produce the GEDI Level-4A (footprint level)
75 algorithm (Duncanson et al., 2022). Currently the L4A product for tropical forests uses relative
76 height (RH -the height that a certain quantile of energy is returned relative to the ground) 98 and
77 RH 50 to predict a median Above Ground Biomass (AGB) of 300 Mg Ha⁻¹ for tropical forests
78 (0.66 r² and RMSE of 10.4). Duncanson et al. (2022) compares these results to previous studies.
79 For instance, Asner and Mascaro (2014) used a network of 804 field inventory plots and aircraft
80 discrete return lidar in 5 tropical countries to estimate biomass with a R² = 0.92 and RMSE =
81 17.1 Mg/ha. Saatchi et al. (2011) combined several datasets with a Maximum Entropy modelling
82 framework across the Tropics to get an r² of 0.80 and RMSE= 23.8. Baccini et al. (2012) used
83 GLAS (Global Laser Altimetry System) on IceSat-1 together with image data from MODIS
84 (MODerate resolution Imaging Sensor) to estimate biomass across the Tropics in a modelling
85 framework of ordinary least squares regression and random forest machine learning algorithms
86 with predictors of HOME (Height of Median Energy), other Height Metrics, and total Canopy
87 returned energy to get an r² of 0.83 and RMSE= 22.6. These early studies exemplify the wide
88 variety of techniques and accuracies used to predict biomass in tropical forests. Forest structure
89 data products derived from GEDI are also related to AGB. For instance, Doughty et al (2023)

90 found forest stratification (% of forests with only one peak in PAVD (Plant Area Volume
91 Density) versus those with several peaks) correlated with biomass more strongly than tree height
92 (Doughty et al., 2023). Duncanson et al (2022) used algorithms stratified by 4 plant functional
93 types and 6 world regions but did not include other remotely sensed (e.g. optical image) data as
94 predictor variables for biomass. Here we explore the extent to which incorporating external
95 datasets and having more regional calibrations can improve GEDI biomass predictions across
96 tropical forests.

97 Environment (e.g., soils and climate) influences the community assembly of tropical
98 forests and knowing species composition could improve biomass estimates since different
99 species have different wood density and structure. For instance, Amazonian plant biogeography
100 may follow a south-west/north-east soil fertility gradient and a north-west/south-east
101 precipitation gradient (ter Steege et al., 2006). Soil cation concentrations are the primary driver
102 of floristic variation for Amazonian trees (Tuomisto et al., 2019) with climate being of secondary
103 importance. However, in central African forests, climate is considered to be the driving factor of
104 floristic patterns (Réjou-Méchain et al., 2021). Therefore, inclusion of soils or forest floristic
105 maps could improve biomass predictions, since both are key variables in different biogeographic
106 zones. Floristics could also determine the relationship between leaf traits and biomass.

107 Leaf traits may also improve tropical forest biomass predictions. One global study of
108 plant traits found that three-quarters of trait variation is captured in a two-dimensional global
109 spectrum of plant form and function (Díaz et al., 2016). One major dimension within this plane
110 reflects the size of whole plants and their parts; the other represents the leaf economics spectrum,
111 which balances leaf construction costs against growth potential (Díaz et al., 2016). Since the size
112 of whole plants may reflect their biomass, ideally there are other traits correlated with plant size
113 and structure that may prove predictive. Traits, such as foliar chemical content, like nitrogen (N),
114 and morphological traits, like leaf mass area (LMA), can be predicted remotely using high-
115 resolution leaf (Asner & Martin, 2008; Homolová et al., 2013) and canopy (Asner et al., 2016;
116 Cawse-Nicholson et al., 2021) spectroscopy (400-2500nm) and algorithms based on partial least
117 squares (PLS) regression or other machine learning statistical techniques. Spectral properties can
118 even predict chemicals not directly expressed in the spectrum, such as base cations or
119 phosphorus (P) because these chemicals have stoichiometric relationships with chemicals that
120 are expressed spectrally (Ustin et al., 2006). Other tree traits such as wood density can be
121 predicted with spectroscopy, i.e. traits that are not directly expressed in leaf spectra but that are
122 instead correlated with leaf traits such as LMA (Doughty et al., 2017). Wall to wall trait maps for
123 leaf chemistry, leaf thickness ($r^2 = 0.52$) leaf carbon content ($r^2 = 0.70$) and maximum rates of
124 photosynthesis ($r^2 = 0.67$) have recently been created using Sentinel-2 spectral data, soils and
125 environmental data (Aguirre-Gutiérrez et al., 2021).

126 Gross primary production (GPP) and Net Primary Production (NPP) are also important
127 fluxes to calculate, but currently are not accurately predicted for tropical forests. For instance,
128 Cleveland et al (2015) compared tropical NPP estimates from field-based methods, RS methods
129 (like MODIS) and mechanistic model-based methods (like the Community Land Model -CLM).
130 The three methods had similar estimates of NPP (i.e., $\sim 10 \text{ Mg C yr}^{-1}$), but displayed differing
131 patterns of NPP through space and through time. The RS based methods to predict NPP made
132 limited use of RS spectral data and relied more on climate based inputs. We are approaching the
133 era of Surface Biology and Geology (SBG, an upcoming wall to wall hyperspectral satellite)

134 (Cawse-Nicholson 2021; Schimel & Poulter, 2022) with hopes for accurate wall to wall trait
135 maps of tropical forests.

136 For this paper we focus on the extent to which plant trait data may help to improve
137 predictions of tropical forest biomass and fluxes. We start by using a large trait database to
138 explore whether traits can predict individual tree DBH, since DBH is always included in
139 allometric equations predicting biomass, while other variables like tree height and species (to get
140 wood density) are only sometimes included (Feldpausch et al., 2011). Next, we compare GEDI
141 predicted biomass to field plot biomass and examine how well RS derived trait maps predict
142 field and RS biomass. Finally, we determine the extent to which structure and traits can improve
143 predictions of tropical forest carbon fluxes (NPP and GPP). To the best of our knowledge, this
144 the first paper directly combining GEDI and satellite derived trait data to predict biomass. We
145 test the following hypotheses:

146 *H1 - Leaf spectral and trait data are correlated with (and not orthogonal to) tree diameter*
147 *(DBH), an important variable for predicting biomass.*

148 *H2 - Leaf traits and environmental data will improve predictions of both field and GEDI*
149 *biomass.*

150 *H3 - GEDI structure or RS trait maps will improve NPP or GPP predictions.*

151

152

153 **Materials and Methods**

154 **Field leaf trait and spectroscopy data** - We used leaf trait and spectral data from an extensive
155 field campaign along an elevation gradient (from 3500 m to 220 m elevation) in the Peruvian
156 Amazon where leaf traits for 60-80% of basal area of trees >10cm DBH were measured within a
157 well-studied 1ha plot network from April to November 2013 (Enquist et al., 2017). In each 1ha
158 plot (N=10 plots), we sampled the most abundant species as determined through basal area
159 weighting (enough species generally to cover ~80% of the plot's basal area). For each species,
160 we sampled the five (three in the lowlands) largest trees (based on diameter at breast height
161 (DBH)) and sampled one sun and one shade branch. On each of these branches, leaf chemistry
162 and leaf mass area (LMA) was measured with methodology detailed in Asner et al. (2014). On
163 five randomly selected leaves for each branch, we measured hemispherical reflectance with an
164 ASD Fieldspec Handheld 2 with fiber optic cable, contact probe which has its own calibrated
165 light source and a leaf clip (Analytical Spectral Devices High Intensity Contact Probe and Leaf
166 Clip, Boulder, Colorado, USA) following (Doughty et al., 2017). We measured leaf spectroscopy
167 (400-1075 nm) on the same branches where the leaf traits were collected. Both LMA and
168 Chlorophyll A had previously been shown with this dataset to have a correlation with leaf
169 spectroscopy (Doughty et al., 2017). However, we had not previously tried to compare leaf
170 spectral data with DBH directly.

171 **Plot data** –

172 *Aboveground biomass* - We used 2,102 of 19,160 total AGB field plots that are between +30°
173 and -30° latitude classified as broadleaf evergreen trees by MODIS PFT using public data
174 (Duncanson et al 2022) that was organized and publicly available through ORNL DAAC as an
175 RDS (R data serialization) file. Distribution of plots are shown in Fig S1 (AGB) and S2
176 (residuals).

177 *NPP and GPP* - We also used 21, 1ha plots where NPP and sometimes GPP were measured
178 following the GEM protocol (Araujo-Murakami et al., 2014; Malhi et al., 2021). We focused on
179 two regions: a Peruvian elevation transect with both NPP & GPP (n= 10, RAINFOR plot codes
180 are ALP11, ALP30, SPD02, SPD01, TRU03, TRU08, TRU07, ESP01, WAY01, ACJ01 (Malhi
181 et al., 2017)) and a Bornean logging transect with only NPP (n= 11 RAINFOR plot codes are
182 DAN-04, DAN-05, LAM-01, LAM-02, MLA-01, MLA-02, SAF-01, SAF-02, SAF-03, SAF-04,
183 SAF-05 (Riutta et al., 2018). These plots were chosen because there are large changes in
184 NPP/GPP across the elevation or logging gradient.

185 **GEDI data** – We used the vertical forest structure (L2A and L2B, Version 2) and biomass (L4a)
186 products from the GEDI instrument (R. Dubayah et al., 2020) from April 2019 to December
187 2022 for tropical forest regions (R. O. Dubayah et al., 2023). We used a quality filtering recipe
188 developed in collaboration with GEDI Science Team members from University of Maryland and
189 NASA Goddard to identify the highest quality GEDI vegetation shots (R. Dubayah et al.,
190 2022). A data layer that this iterative local outlier detection algorithm uses to exclude data is
191 publicly available at (R. O. Dubayah et al., 2023). For instance, some of the key data filters we
192 applied were: included degrade flags of 0, 3, 8, 10, 13, 18, 20, 23, 28, 30, 33, 38, 40, 43, 48, 60,
193 63, 68, L2A and L2B quality flags = 1 (only use highest quality data), sensitivity >= 0.98. With

194 the GEDI data we used canopy height, height of median energy (HOME), and the number of
195 canopy layers following Doughty et al (2023).

196 Across all tropical forests, we created 300 by 300m pixels containing all averaged (mean)
197 GEDI data between 2019 and 2022. Using the centroid coordinates from each of the 2,102 plots,
198 we found the 300 by 300 m averaged GEDI pixel that encompassed the plot. If the plot was not
199 encompassed by the GEDI data, we searched a wider area by incrementally averaging a
200 gradually increasing area of 1, 3, 5, and 10 pixels. In other words, if no 300 by 300m pixel
201 encompassed the plot, then we averaged all GEDI data an area one pixel out (4 by 4 = 1200 by
202 1200m, 6 by 6 = 1800 by 1800m, 11 by 11 = 3300m by 3300m), gradually increasing the square
203 until it encompassed an area with GEDI data. To compare with the NPP/GPP plots we compared
204 RS trait and GEDI data for individual footprints within a 0.03km radius of the plot coordinates.

205

206 **Remotely sensed leaf trait data** - Based on a broader set of field campaigns, Aguirre-Gutiérrez
207 et al., (2021) used Sentinel-2, climate, topographic and soil data to create remotely sensed
208 canopy trait maps for %P=phosphorus % leaf concentration, WD = wood density g cm^{-3} , and
209 LMA=Leaf mass area g m^{-2} .

210 **Other data layers** – We compared percentage of one peak (% one peak hereafter) (an estimate
211 of canopy stratification with 1 = more than one vertical peak in PAVD and 0 = one vertical peak
212 in PAVD) to several other climate, soils, leaf traits, and ecoregion maps listed below for the
213 Amazon basin. Each dataset had its own resolution, which we standardized to 0.1 by 0.1 degrees.
214 We used total cation exchange capacity (CEC) from soil grids (Batjes et al., 2020) in the top 0-
215 5cm layer in units of mmol(c) kg^{-1} . We averaged TerraClimate (Abatzoglou et al., 2018) data
216 between 2000 and 2018 for Vapor Pressure Deficit (VPD in kPa), Mean Monthly Precipitation
217 (MMP) (mm/month), potential evapotranspiration (PET) and maximum and minimum
218 temperature ($^{\circ}\text{C}$).

219 **Statistical analysis** – We did a principal component analysis (PCA) for the tree level trait data of
220 DBH, % of nitrogen in leaf of leaf nitrogen concentration (% N), LMA, and chl A using the
221 matlab (Matlab, MathWorks Inc., Natick, MA, USA) function `pca`. We used the matlab function
222 “`fitlm`” to fit linear models to examine the relationship between the variables, such as soils data,
223 environmental data, leaf trait data (at 0.1° resolution) and GEDI structure data (300m and bigger
224 resolution), and field biomass and NPP/GPP estimates. The P values listed are for the *t*-statistic
225 of the two-sided hypothesis test. We used R to create a linear model to predict the best model
226 ranked by Akaike Information Criterion AIC and parsimony using the dredge function from the
227 MuMIn library (Bartoń, 2009). We also used CAR package (Fox J & S, 2019) and the VIF
228 command to test for multi-collinearity between variables. To account for spatial autocorrelation,
229 we used Simultaneous Auto-Regressive (SAR^{err}) models (F. Dormann et al., 2007) using the R
230 library ‘`spdep`’ (Bivand, Hauke, & Kossowski, 2013). We tested different neighborhood
231 distances from 10km to 300km and found that AIC was minimized at 80km (Fig S3) and the
232 corresponding correlogram showed reduced spatial autocorrelation (Fig S4). To predict leaf
233 traits with the spectral information, we used the Partial Least Squares Regression (PLSR) (Geladi
234 & Kowalski, 1986) using the `PLSregress` command in Matlab (Matlab, MathWorks Inc., Natick,
235 MA, USA). To avoid over-fitting the number of latent factors we minimized the mean square
236 error with K-fold cross validation. We use 70% of our data to calibrate our model and then the
237 remaining 30% to test the accuracy of our model using r^2 . We use adjusted r^2 , which penalizes
238 for small sample sizes throughout the manuscript.

239 Results

240 We examined the relationship between averaged trait values collected from cut branches and
241 DBH of that tree for 3695 leaves from 523 trees (Doughty et al., 2017) along a Peruvian
242 elevation gradient and found a low correlation ($r^2 < 0.01$) between leaf chemistry (%N and %P)
243 and DBH. However, LMA showed a significant ($P < 0.0001$) positive correlation with DBH and
244 Chl A showed a significant ($P < 0.0001$) negative correlation but with relatively low variance
245 explained ($r^2 = \sim 0.04$ and 0.06 respectively) (Figure 1). LMA had a significant ($P < 0.0001$)
246 negative correlation with tree height ($r^2 = \sim 0.17$). We did a PCA for the tree level trait data of
247 DBH, %N, LMA, and chl A and found the 1st principal component axis explained 94% of the
248 variance and was dominated by DBH, while LMA dominated the second PCA axis and
249 explained $\sim 6\%$ in the orthogonal direction (Fig S5). This may be why traits explain little of the
250 variance in our dataset. We then compared tree averaged leaf spectral data (400 to 1075 nm) to
251 DBH using the PLSR technique and found only a weak correlation (Figure 2, $r^2 = 0.01$). LMA is
252 predictable with spectroscopy ($r^2 = 0.63$) and DBH is weakly predictable with LMA ($r^2 = 0.04$),
253 and this translated into spectra being able to predict DBH with an $r^2 = 0.01$ in this dataset.

254 We then compared predictions of GEDI biomass to 2,102, 25m (although some plots are 1 ha)
255 biomass plots across *all tropical forests* (not just Peru) (Fig 3). These plot data were used to
256 create GEDI's Level 4 footprint-level AGB product using simulated waveforms from ALS
257 collocated with field plots. In contrast, we created 300 by 300m pixels containing all averaged
258 (mean) GEDI data between 2019 and 2022 across all tropical forests. We acknowledge a degree
259 of circularity in our analysis, but the comparison is different than Duncanson et al. (2022).
260 Because of the variable nature of GEDI data collection and the variable ISS orbital tracks, only
261 247 ($\sim 45\%$) of the plots had plot data within the 300 by 300m pixel and $\sim 2.5\%$ of the plots
262 needed an area of 3300m by 3300m. We therefore are not aligning field and GEDI data but are
263 instead assessing regional correlations among variables of interest, thus our expected correlations
264 will be much lower than where GEDI and field plots are geolocated and temporally aligned. We
265 then subtracted GEDI regional averages of predicted biomass from field derived biomass
266 (henceforth referred to as residuals) for 2102 plots across the tropics and showed both their
267 location, AGB, and the average difference from the GEDI predicted value (Fig 3). There are
268 spatial patterns with the residuals with, for instance, GEDI overestimating AGB in the Yucatan
269 Peninsula and underestimating in the Eastern Amazon. Overall, the residuals have two modes at
270 ~ -100 and 100 Mg ha^{-1} . Next, our goal is to determine whether the bias can be reduced by
271 incorporating RS leaf traits or other external datasets.

272 For these 2,102 plots, there was a significant ($P < 0.0001$) negative correlation between the
273 remotely sensed trait of LMA for both GEDI biomass ($r^2 = 0.38$) and GEDI measured forest
274 height ($r^2 = \sim 0.43$) (Fig 4). There was a significant ($P < 0.0001$) negative correlation between
275 remotely sensed % P and biomass and height ($r^2 = 0.31$ and $r^2 = 0.36$ respectively). However,
276 LMA predicted field derived biomass poorly ($r^2 = \sim 0.01$) and % P was not correlated with field
277 derived biomass ($P > 0.05$). LMA was always a stronger predictor than %P, for height, RS
278 biomass and field derived biomass. We correlated field vs remotely sensed biomass ($r^2 = \sim 0.05$)
279 and remotely sensed tree height ($r^2 = \sim 0.03$) (note again that these are regional correlations and
280 not exact geolocated comparisons).

281 We then compared LMA, %P, GEDI height and percentage one peak (an estimate of canopy
282 stratification with 1 = more than one vertical peak in PAVD and 0 = one vertical peak in PAVD)

283 to biomass residuals and found a negative relationship between LMA and residuals (r^2 of 0.34,
284 $N=66$) and a negative relationship with %P ($r^2=0.31$). Of GEDI structure variables, % one peak
285 did poorly, only predicted 4 % of the variance but tree height predicted biomass strongly with an
286 r^2 of 0.74 (Figure 5). We then subset the AGB field plots for the Amazon basin ($N=66$ of 2102
287 total) to match our climate and soils datasets. We compared climate data (VPD, T_{max} , PET) and
288 soils data (cation exchange capacity - CEC) to biomass residuals and found T_{max} was best in
289 predicting residuals with an r^2 of 0.79 followed by PET ($r^2=0.70$) and VPD ($r^2=0.28$) (Figure 6).
290 We did not find a significant relationship ($P>0.05$) between CEC and biomass residuals.

291 We tested for spatial autocorrelation and found that averaging around a radius of 80km (this
292 large radius may incorporate broader climate trends) minimized AIC (Figure S3) which reduced
293 spatial autocorrelation according to the correlogram (Figure S4). There was some collinearity
294 between the trait variables and structure variables ($VIF>3$), so we removed %P and HOME and
295 this reduced all collinearity scores to under ~ 1.5 . To predict RS biomass, the best model by AIC
296 included LMA, height, and % one peak, but LMA was only marginally significant (Table 1). For
297 field biomass, the best model by AIC again included all three variables but % one peak was not
298 significant. After controlling for spatial autocorrelation by grouping the plot data into
299 neighborhoods of 80km, the statistical models changed. Adding LMA (but not %P, HOME, or %
300 one peak) significantly ($P<0.0001$) improved field biomass predictions. Adding traits (neither
301 LMA or P) did not significantly improve RS biomass but both % one peak and HOME did
302 ($P<0.0001$). We \log_{10} transformed the data (biomass, LMA and height) which improved
303 predictions of field biomass (from r^2 of 0.03 to an r^2 of 0.16) but did not improve predictions of
304 RS biomass. Overall, canopy height was always by far the most important predictor of AGB but
305 adding RS LMA did improve predictions of field biomass by $\sim 0.01 r^2$.

306 We then estimated NPP and GPP data with traits (LMA and % P) and structure (biomass, tree
307 height, and % one peak). LMA showed the strongest correlation with both NPP ($r^2=0.38$) and
308 GPP ($r^2=0.41$) (Figure 7). Tree height and % one peak were not significantly correlated with the
309 NPP/GPP plot data. For the logging gradient in Borneo there was a significant correlation with
310 both tree height and LMA to NPP with LMA stronger than other traits. However, when we
311 combined the Borneo and Amazonia data sets together, only LMA remained significantly
312 correlated with NPP (Figure 8).

313

314

315

316 Discussion

317 After controlling for spatial autocorrelation, adding RS derived LMA trait data significantly
318 improves predictions of field measured (but not GEDI estimated) tropical forest biomass, only by
319 a small amount (improving r^2 by ~ 0.01) but information criteria (AIC) suggest LMA should be
320 added. An important caveat is that we are not comparing geolocated field plot data to GEDI and
321 trait data for the same exact area, but instead for the broader region (i.e. only 45% of the ABG
322 plots have GEDI data within a 300 by 300m area). This differs from the study by Duncanson et
323 al (2022) where airborne lidar data were used to simulate GEDI data for each plot, therefore
324 comparing predicted GEDI structure for the same area as the field plots. Since there is much
325 regional variation in biomass, our predictions of field measured biomass have a low r^2 ($r^2 \sim 0.03$)
326 but were significantly improved with RS LMA data. LMA also directly predicts field biomass
327 with an $r^2 \sim 0.01$ (Figure 4). At the individual tree scale (Figure 1), we show similar results with
328 LMA predicting 4% of DBH variance (highly correlated with biomass) and spectral properties
329 predicting 1% of DBH variance (Figure 2). However, predicting biomass at the canopy scale
330 may have more success than at the leaf scale, because canopies incorporate more spectral
331 information with higher LAI (Baret et al., 1994). Therefore, we estimate that adding RS trait data
332 to GEDI results in a real, but very small improvement in field biomass predictability. But is this
333 meaningful? The GEDI L4A product for tropical forests currently has an accuracy of 0.66 r^2
334 (Duncanson et al., 2022), so any real improvement is welcome, if real. However, adding non-
335 GEDI data to biomass predictions could also introduce error which could cancel out the 1%
336 improvement.

337 Some of our results tentatively suggest that adding traits could lead to a greater improvement in
338 AGB prediction than suggested above by reducing bias in the residuals. For instance, we found
339 the remotely sensed trait of LMA was correlated with both GEDI biomass ($r^2=0.38$) and GEDI
340 measured forest height ($r^2 \sim 0.43$) (Fig 4). We also found both LMA (r^2 of 0.34) and %P
341 ($r^2=0.31$) correlated with the biomass residuals. This suggests that traits could potentially correct
342 for bias in current GEDI predictions, which could be more useful than a 0.01 improvement in r^2 .
343 However, because the leaf traits maps use predictors of soils and climate data in addition to
344 Sentinel 2 spectral data, the improvements to biomass prediction may be due to the influence of
345 the underlying climate variables as shown in Fig 6. LMA and %P correlated more with RS AGB
346 than field AGB possibly for this reason as well. We focused on using trait data in tropical forests
347 because remotely sensed species detection is difficult (Ferret & Asner, 2013; Mulatu et al., 2017),
348 but similar approaches could potentially be used in lower diversity temperate and boreal forests
349 as well. There is optimism for future improvements in predictability because our leaf spectral
350 data only extends through 1075 nm, and there is likely important spectral information at longer
351 wavelengths (e.g. in the shortwave infrared). The current RS trait maps (Aguirre-Gutiérrez et al.,
352 2021) use a few Sentinel 2 spectral bands but future satellites like Surface Biology Geology
353 (SBG) (Cawse-Nicholson 2021; Schimel & Poulter, 2022) or the Plankton, Aerosol, Cloud,
354 ocean Ecosystem (PACE) mission (Gorman et al., 2019) will have improved or wall to wall
355 hyperspectral data and therefore future, more accurate trait maps may improve biomass estimates
356 by a greater amount or reduce uncertainties.

357 Our strongest (non-GEDI) predictor of biomass residuals was T_{\max} with an r^2 of 0.79, but we
358 note that this is based on a much smaller Amazon only dataset (N=66) (Fig 6). The negative
359 correlation suggests that GEDI underpredicts biomass in regions where VPD or T_{\max} is on

360 average higher. Stressful temperature or aridity may reduce tree biomass and height from their
361 maximum potential or select for smaller species with more conservative strategies. This result is
362 supported by literature showing higher temperatures reducing tropical forest growth rates (Clark
363 et al., 2003). Soil cation concentration was not a strong predictor of biomass residuals in our
364 dataset which is surprising because soil cation concentrations are the primary driver of floristic
365 variation for Amazonian trees (Tuomisto et al., 2019) with climate being of secondary
366 importance.

367 In a previous paper, we had hypothesized that forest stratification (% one peak or the number of
368 single stratum forests as a percentage of total) might improve biomass predictions better than a
369 simple metric like rh50 (Doughty et al., 2023) because in that paper, % one peak predicted
370 biomass better than tree height. Ecological theory suggests that a stratified forest with more
371 large emergent trees is indicative of an older forest (Halle et al., 1980), which generally has
372 higher biomass and carbon content. However, in our study, % one peak was a fairly poor
373 predictor of the residuals explaining only 4% of the variance. This compares with other traits
374 that predicted more variance such as 75% with tree height, 16% with rh50 and 36% with HOME.
375 When we added % one peak to our overall model it did not improve the AIC, and therefore
376 seems a poor predictor of biomass across tropical forests. We also found a high correlation
377 between GEDI height and biomass residuals (Fig 5c), which may be due to the transformation
378 (log or square root) of biomass in GEDI L4A models such that error increases as biomass/height
379 increases. Further, a recent paper found that GEDI accurately predict redwood tree heights but
380 still underestimates AGB because tree height may not be an accurate predictor for high biomass
381 forests (Sillett et al., 2024). Moving forward, terrestrial lidar can expand our understanding of
382 tree structure and possibly create improved biomass estimates beyond DBH (Stovall & Shugart,
383 2018).

384 Remotely sensed MODIS NPP and GPP is a commonly used input to many global models
385 (Zhang et al., 2012) but previous studies have found that MODIS NPP does not match ground
386 based estimates of NPP seasonality and therefore, there is a need for improved remote sensed
387 NPP estimates (Cleveland et al., 2015). Our results (Fig 7 and 8) suggest that adding trait maps
388 to predictions of GPP and NPP could potentially improve accuracy, but GEDI structure metrics
389 did not improve predictability. For instance, remotely sensed LMA predicted GPP ($r^2=0.4$) and
390 NPP ($r^2=0.35$) better than GEDI height in an Andean elevation gradient (Fig 7). When we
391 combined both datasets, only LMA continued to predict NPP (Fig 8). However, although we
392 used the biggest NPP and GPP dataset in the tropics, our sample size ($N=21$) was small. More
393 ground based NPP/GPP networks are necessary for validation before we would have confidence
394 in this result.

395 Overall, we show several lines of evidence (tree DBH versus leaf traits, tree DBH versus
396 spectroscopy, RS traits versus field biomass, RS traits versus field NPP/GPP) that traits can
397 slightly improve estimates of tropical forest biomass and fluxes and possibly may be further
398 improved in the future with data from new satellite missions like SBG. Other potential
399 improvements in remote biomass estimates might come from integrating dynamic vegetation
400 models that have trait data with GEDI observations (Ma et al., 2023).

401

402

403 **Code and data availability** - Description of the Type(s) of data and/or software -
404 Data - Data and its descriptions to create all figures and tables in this paper are available
405 (Doughty 2024).
406 Software – All code and its descriptions to create all figures and tables in this paper are available
407 (Doughty 2024).
408
409

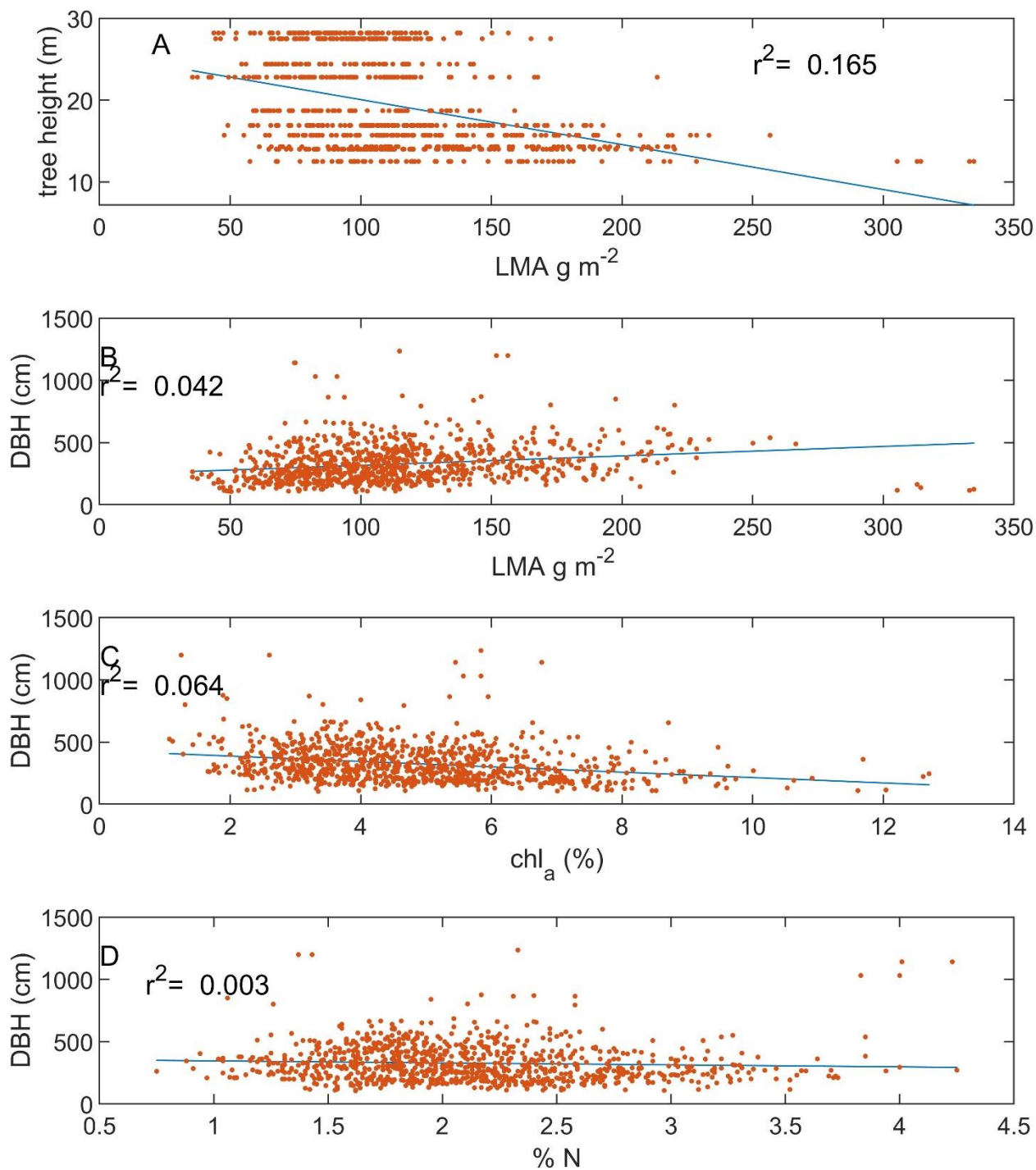
410 **Table 1** – Model results (ΔAIC and adjusted R^2) for field derived biomass, and GEDI predicted
 411 biomass using GEDI measured forest height, GEDI measured maximum PAVD height, % one
 412 peak, and leaf traits of LMA and % P. For ΔAIC we give the change in ΔAIC between the best
 413 model and the second-best model. The best model column gives the best model according to
 414 AIC and the variable removed (bolded and italicized) for the next best model.

field derived biomass				RS biomass		
Variables	ΔAIC	Best model	Adj r^2	ΔAIC	Best model	Adj r^2
height, peak, P	1	height, P, <i>PEAK</i>	0.0356	1.5	height, peak, <i>P</i>	0.799
height, peak, LMA		height, peak	0.0281		height, peak	0.799
height, HOME, P	3	height, P, <i>HOME</i>	0.0368	22	height, HOME, <i>P</i>	0.795
height, HOME, LMA	2	height,HOME, <i>LMA</i>	0.0326	7	height, HOME, <i>LMA</i>	0.793
height	-		0.0272	-		0.787

415

416

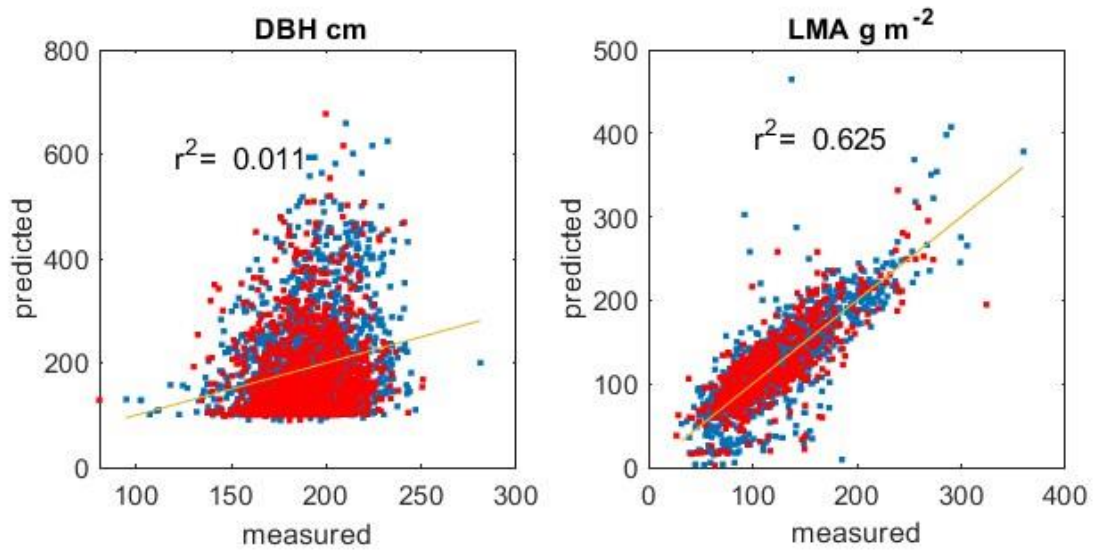
417 **Figures**



418
419 **Fig 1** –Individual tree height compared with leaf LMA g m^{-2} (A), DBH compared with leaf LMA
420 g m^{-2} (B), % Chlorophyll A (C) and % N (D), averaged on ~3 branches and 5 leaves per branch.

421

422

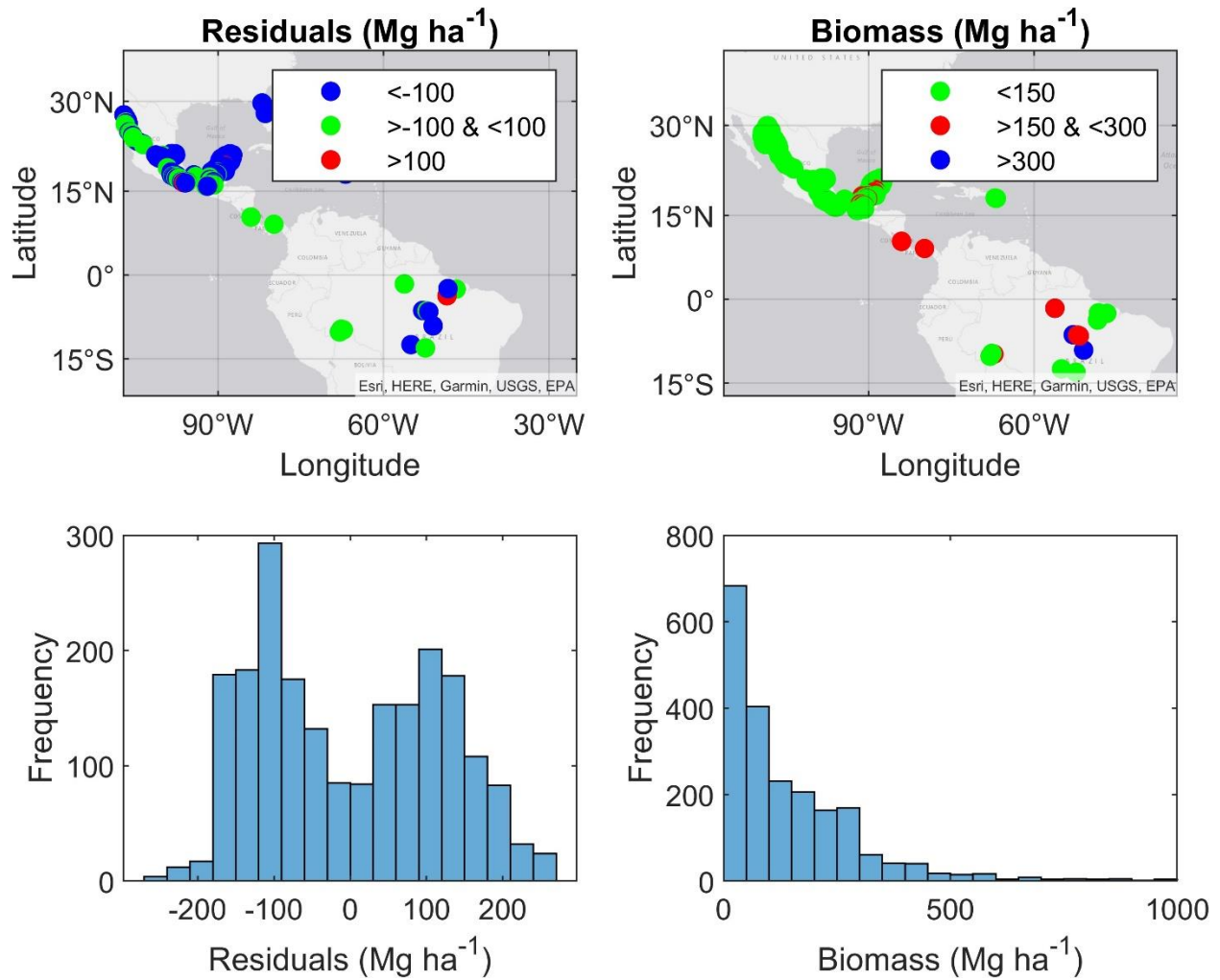


423

424 **Fig 2** –Leaf spectral (400-1075 nm) (N= 4690 individual leaves) averaged on ~3 branches and 5
 425 leaves per branch versus their diameter at Breast Height (DBH) (left) or Leaf Mass Area (LMA)
 426 (right) using the PLSR technique (blue is training data and red is the validation data).

427

428

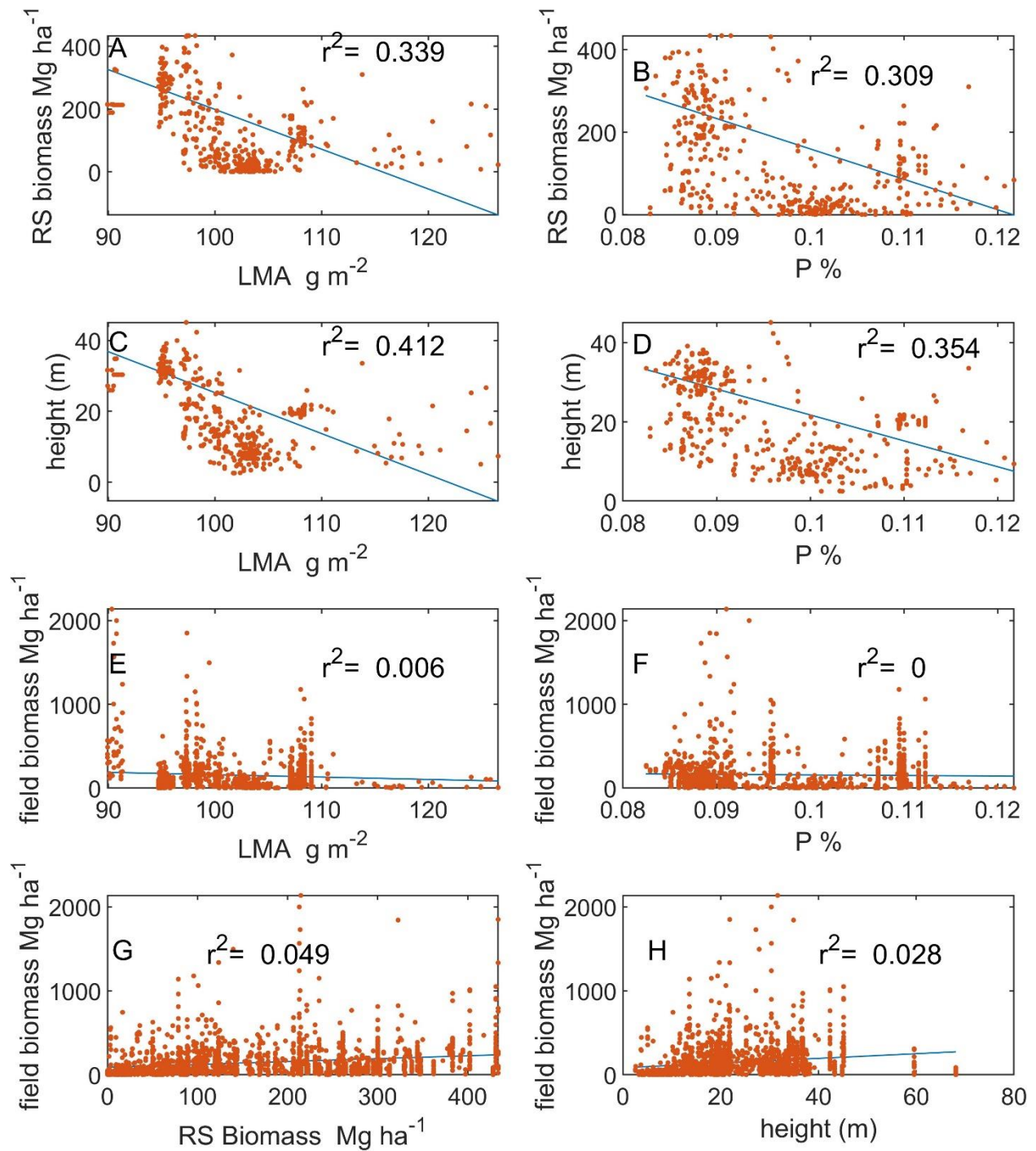


429

430 **Fig 3**—GEDI predicted biomass minus field biomass (residuals) (left) and field biomass (right)
 431 where (top) the color dots represents the value (residuals Mg ha^{-1} between 100 and -100 = green,
 432 >100 = red, and <-100 blue and $\text{AGB Mg ha}^{-1} < 150$ =green, between 150 and 300 = red and
 433 >300 = blue). For the maps we show a subset of the data for visual clarity. The full maps are
 434 shown in fig S1 and S2. On the bottom, we show a histogram of the residuals (left) and field
 435 biomass (right). All comparisons were aggregated to 300 by 300 m areas.

436

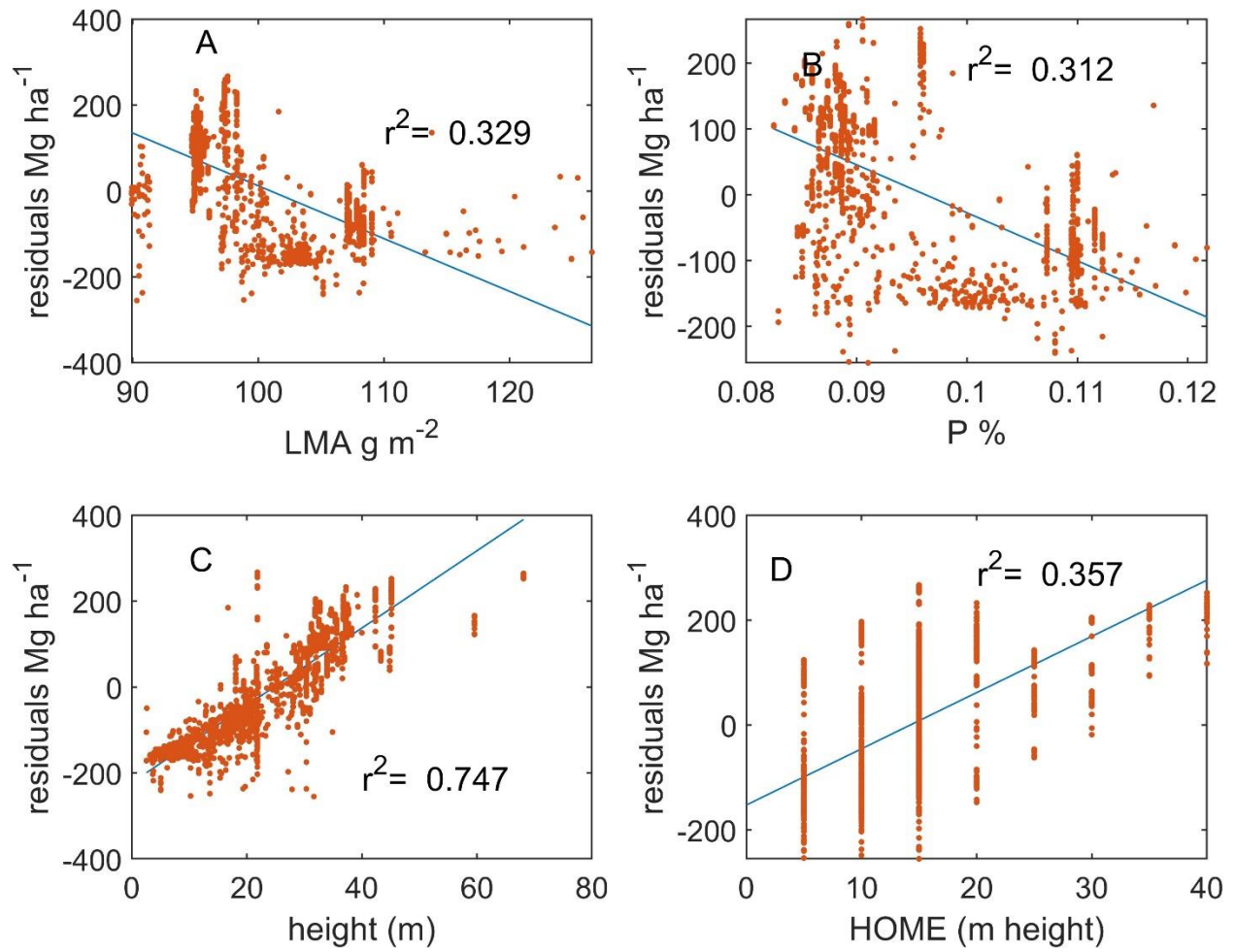
437



438

439 **Fig 4** –LMA (g m^{-2}) versus RS biomass (A) Mg ha^{-1} , tree height (C) (m), and field derived
 440 biomass (E) Mg ha^{-1} . P (%) versus RS biomass (B) Mg ha^{-1} , tree height (D) (m), and field
 441 derived biomass (F) Mg ha^{-1} . RS biomass (G) Mg ha^{-1} versus field derived biomass and (H) tree
 442 height (m) versus Mg ha^{-1} versus field derived biomass.

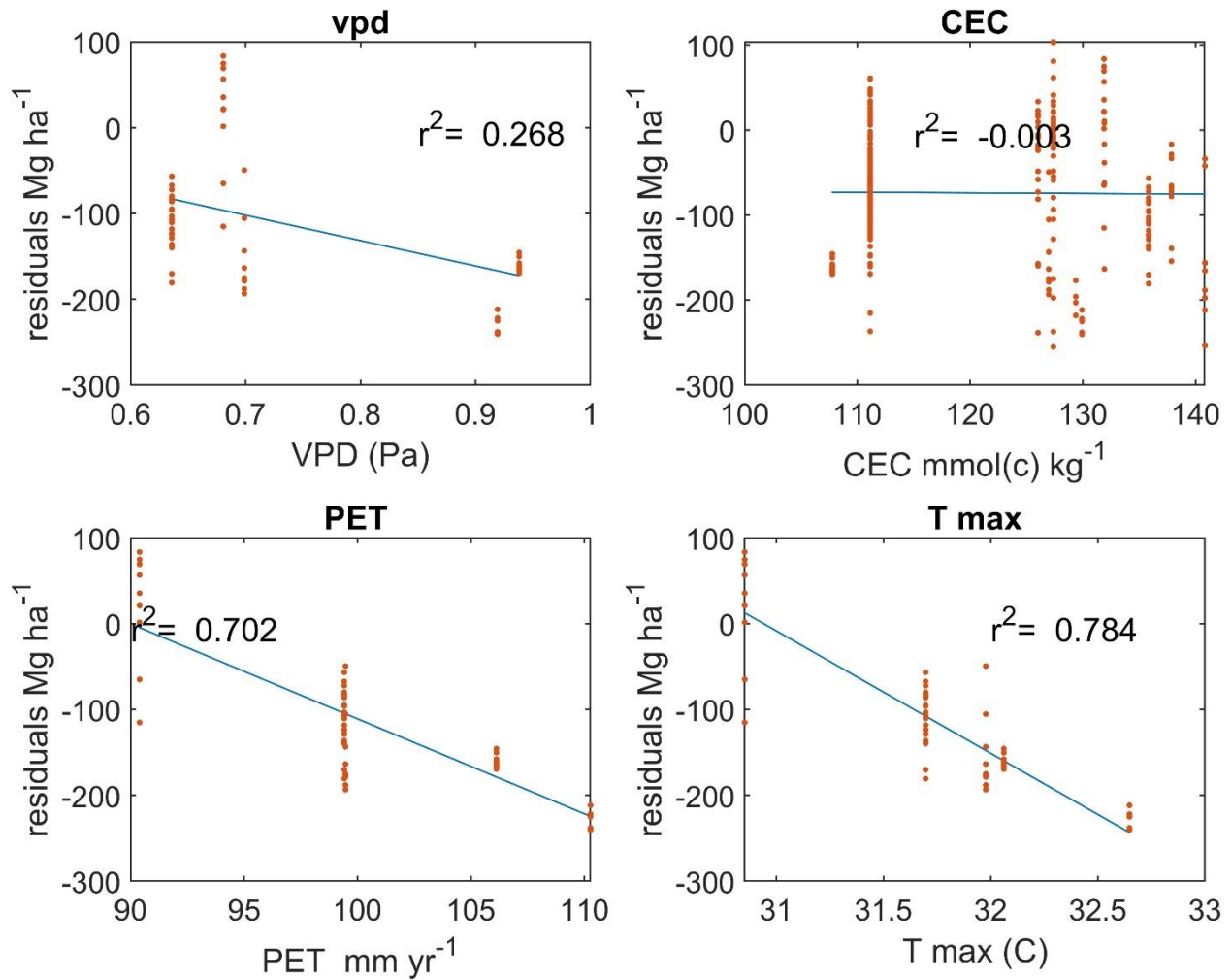
443



444

445 **Fig 5** – Biomass residuals (plot biomass minus GEDI predicted biomass) versus remotely sensed
 446 LMA (A), % P (B), and GEDI predicted structural variables (height (C) and HOME(D)).

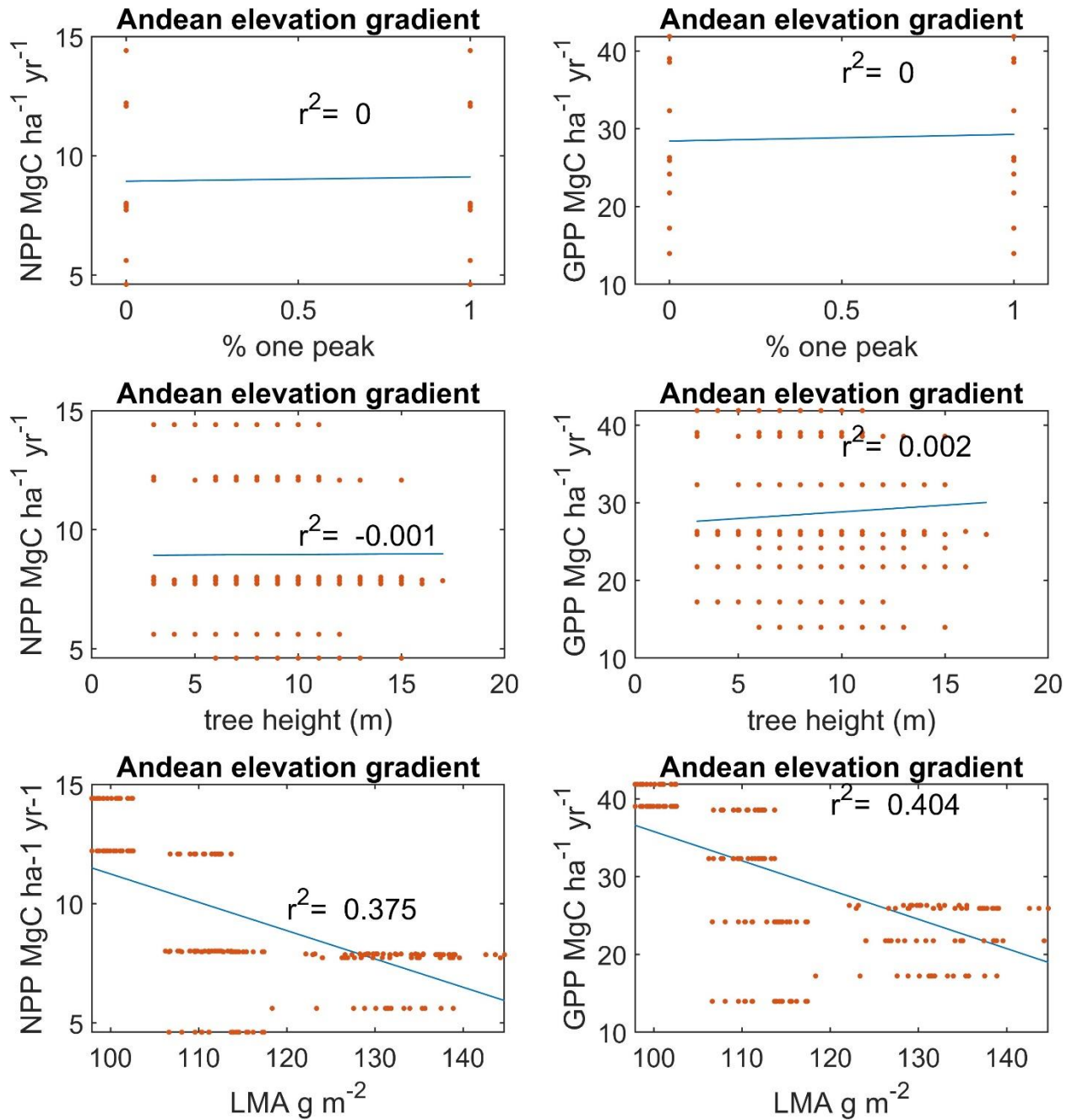
447



448

449 **Fig 6** –Biomass residuals (plot biomass minus GEDI predicted biomass) versus soils (cation
 450 exchange capacity - CEC) and climate data (vapor pressure deficit (VPD), potential
 451 evapotranspiration (PET), and maximum temperature (T_{\max})).

452

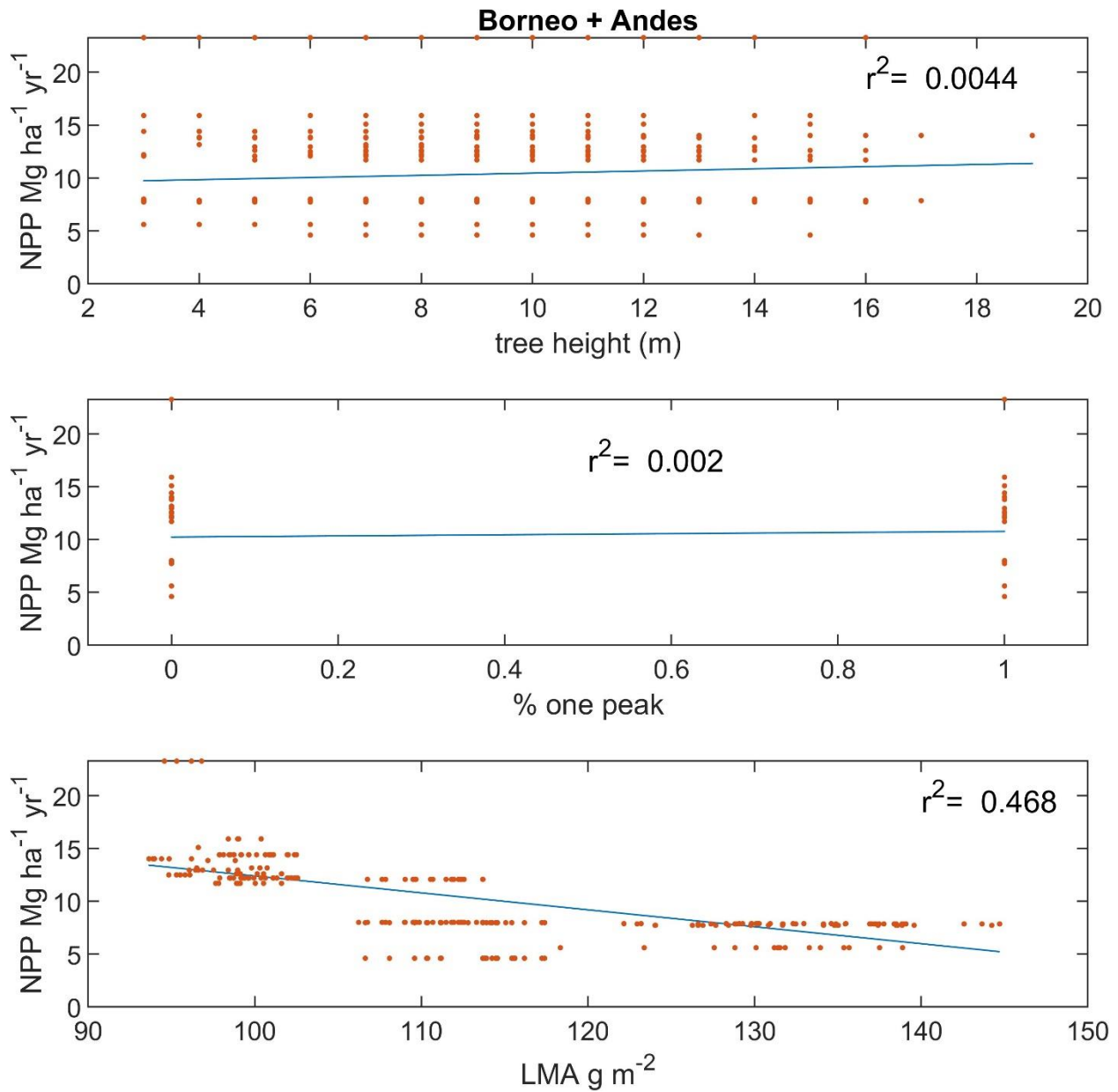


453

454 **Fig 7** – Net Primary Production (left) and Gross primary production (right) data from South
 455 America compared to % one peak (top) (an estimate of canopy stratification with 1 = more than
 456 one vertical peak in PAVD and 0 = one vertical peak in PAVD), GEDI calculated tree height
 457 (middle), and remote sensed LMA (bottom). GEDI data are from the nearest 0.03 degrees pixel.

458

459



460

461 **Fig 8** – Net Primary Production data from Borneo and South America compared to GEDI
 462 calculated tree height (top), % one peak (middle) (an estimate of canopy stratification with 1 =
 463 more than one vertical peak in PAVD and 0 = one vertical peak in PAVD) and remote sensed
 464 LMA (bottom).

465

466

467

468 **References**

- 469 Abatzoglou, J. T., Dobrowski, S. Z., Parks, S. A., & Hegewisch, K. C. (2018). TerraClimate, a
470 high-resolution global dataset of monthly climate and climatic water balance from 1958–
471 2015. *Scientific Data*, 5(1), 170191. <https://doi.org/10.1038/sdata.2017.191>
- 472 Aguirre-Gutiérrez, J., Rifai, S., Shenkin, A., Oliveras, I., Bentley, L. P., Svátek, M., et al. (2021).
473 Pantropical modelling of canopy functional traits using Sentinel-2 remote sensing data.
474 *Remote Sensing of Environment*, 252, 112122.
475 <https://doi.org/https://doi.org/10.1016/j.rse.2020.112122>
- 476 Araujo-Murakami, A., Doughty, C. E., Metcalfe, D. B., Silva-Espejo, J. E., Arroyo, L., Heredia,
477 J. P., et al. (2014). The productivity, allocation and cycling of carbon in forests at the dry
478 margin of the Amazon forest in Bolivia. *Plant Ecology & Diversity*, 7(1–2), 55–69.
479 <https://doi.org/10.1080/17550874.2013.798364>
- 480 Asner, G. P., & Martin, R. E. (2008). Spectral and chemical analysis of tropical forests: Scaling
481 from leaf to canopy levels. *Remote Sensing of Environment*.
482 <https://doi.org/10.1016/j.rse.2008.07.003>
- 483 Asner, G. P., Knapp, D. E., Anderson, C. B., Martin, R. E., & Vaughn, N. (2016). Large-scale
484 climatic and geophysical controls on the leaf economics spectrum. *Proceedings of the
485 National Academy of Sciences of the United States of America*.
486 <https://doi.org/10.1073/pnas.1604863113>
- 487 Avitabile, V., Herold, M., Heuvelink, G. B. M., Lewis, S. L., Phillips, O. L., Asner, G. P., et al.
488 (2016). An integrated pan-tropical biomass map using multiple reference datasets. *Global
489 Change Biology*, 22(4), 1406–1420. <https://doi.org/https://doi.org/10.1111/gcb.13139>
- 490 Baccini, A., Goetz, S. J., Walker, W. S., Laporte, N. T., Sun, M., Sulla-Menashe, D., et al.
491 (2012). Estimated carbon dioxide emissions from tropical deforestation improved by
492 carbon-density maps. *Nature Climate Change*, 2(3), 182–185.
493 <https://doi.org/10.1038/nclimate1354>
- 494 Baret, F., Vanderbilt, V. C., Steven, M. D., & Jacquemoud, S. (1994). Use of spectral analogy to
495 evaluate canopy reflectance sensitivity to leaf optical properties. *Remote Sensing of
496 Environment*, 48(2), 253–260. [https://doi.org/https://doi.org/10.1016/0034-4257\(94\)90146-
497 5](https://doi.org/https://doi.org/10.1016/0034-4257(94)90146-5)
- 498 Bartoń, K. (2009). MuMIn: Multi-model inference. *R Package Version 0.12.2* .
- 499 Batjes, N. H., Ribeiro, E., & van Oostrum, A. (2020). Standardised soil profile data to support
500 global mapping and modelling (WoSIS snapshot 2019). *Earth System Science Data*, 12(1),
501 299–320. <https://doi.org/10.5194/essd-12-299-2020>
- 502 Cawse-Nicholson, K., Townsend, P. A., Schimel, D., Assiri, A. M., Blake, P. L., Buongiorno, M.
503 F., et al. (2021). NASA’s surface biology and geology designated observable: A perspective
504 on surface imaging algorithms. *Remote Sensing of Environment*, 257, 112349.
505 <https://doi.org/https://doi.org/10.1016/j.rse.2021.112349>

- 506 Clark, D. A., Piper, S. C., Keeling, C. D., & Clark, D. B. (2003). Tropical rain forest tree growth
507 and atmospheric carbon dynamics linked to interannual temperature variation during 1984–
508 2000. *Proceedings of the National Academy of Sciences*, *100*(10), 5852–5857.
509 <https://doi.org/10.1073/pnas.0935903100>
- 510 Cleveland, C. C., Taylor, P., Chadwick, K. D., Dahlin, K., Doughty, C. E., Malhi, Y., et al.
511 (2015). A comparison of plot-based satellite and Earth system model estimates of tropical
512 forest net primary production. *Global Biogeochemical Cycles*.
513 <https://doi.org/10.1002/2014GB005022>
- 514 Díaz, S., Kattge, J., Cornelissen, J. H. C., Wright, I. J., Lavorel, S., Dray, S., et al. (2016). The
515 global spectrum of plant form and function. *Nature*. <https://doi.org/10.1038/nature16489>
- 516 Doughty, C. E., Santos-Andrade, P. E., Goldsmith, G. R., Blonder, B., Shenkin, A., Bentley, L.
517 P., et al. (2017). Can Leaf Spectroscopy Predict Leaf and Forest Traits Along a Peruvian
518 Tropical Forest Elevation Gradient? *Journal of Geophysical Research: Biogeosciences*.
519 <https://doi.org/10.1002/2017JG003883>
- 520 Doughty, C. E., Gaillard, C., Burns, P., Keany, J. M., Abraham, A. J., Malhi, Y., et al. (2023).
521 Tropical forests are mainly unstratified especially in Amazonia and regions with lower
522 fertility or higher temperatures. *Environmental Research: Ecology*, *2*(3), 35002.
523 <https://doi.org/10.1088/2752-664X/ace723>
- 524 Dubayah, R., Blair, J. B., Goetz, S., Fatoyinbo, L., Hansen, M., Healey, S., et al. (2020). The
525 Global Ecosystem Dynamics Investigation: High-resolution laser ranging of the Earth's
526 forests and topography. *Science of Remote Sensing*.
527 <https://doi.org/10.1016/j.srs.2020.100002>
- 528 Dubayah, R., Armston, J., Healey, S. P., Bruening, J. M., Patterson, P. L., Kellner, J. R., et al.
529 (2022). GEDI launches a new era of biomass inference from space. *Environmental*
530 *Research Letters*, *17*(9), 95001. <https://doi.org/10.1088/1748-9326/ac8694>
- 531 Dubayah, R. O., Armston, J., Healey, S. P., Yang, Z., Patterson, P. L., Saarela, S., et al. (2023).
532 GEDI L4B Gridded Aboveground Biomass Density, Version 2.1. ORNL Distributed Active
533 Archive Center. <https://doi.org/10.3334/ORNLDAAC/2299>
- 534 Duncanson, L., Kellner, J. R., Armston, J., Dubayah, R., Minor, D. M., Hancock, S., et al.
535 (2022). Aboveground biomass density models for NASA's Global Ecosystem Dynamics
536 Investigation (GEDI) lidar mission. *Remote Sensing of Environment*, *270*, 112845.
537 <https://doi.org/https://doi.org/10.1016/j.rse.2021.112845>
- 538 Enquist, B. J., Bentley, L. P., Shenkin, A., Maitner, B., Savage, V., Michaletz, S., et al. (2017).
539 Assessing trait-based scaling theory in tropical forests spanning a broad temperature
540 gradient. *Global Ecology and Biogeography*, *26*(12), 1357–1373.
541 <https://doi.org/https://doi.org/10.1111/geb.12645>
- 542 F. Dormann, C., M. McPherson, J., B. Araújo, M., Bivand, R., Bolliger, J., Carl, G., et al. (2007).
543 Methods to account for spatial autocorrelation in the analysis of species distributional data:
544 a review. *Ecography*, *30*(5), 609–628. [https://doi.org/https://doi.org/10.1111/j.2007.0906-
545 7590.05171.x](https://doi.org/https://doi.org/10.1111/j.2007.0906-7590.05171.x)

- 546 Feldpausch, T. R., Banin, L., Phillips, O. L., Baker, T. R., Lewis, S. L., Quesada, C. A., et al.
547 (2011). Height-diameter allometry of tropical forest trees. *Biogeosciences*, 8(5), 1081–1106.
548 <https://doi.org/10.5194/bg-8-1081-2011>
- 549 Feret, J.-B., & Asner, G. P. (2013). Tree Species Discrimination in Tropical Forests Using
550 Airborne Imaging Spectroscopy. *IEEE Transactions on Geoscience and Remote Sensing*,
551 51(1), 73–84. <https://doi.org/10.1109/TGRS.2012.2199323>
- 552 Fox J, & S, W. (2019). An R Companion to Applied Regression.
- 553 Geladi, P., & Kowalski, B. R. (1986). Partial least-squares regression: a tutorial. *Analytica*
554 *Chimica Acta*. [https://doi.org/10.1016/0003-2670\(86\)80028-9](https://doi.org/10.1016/0003-2670(86)80028-9)
- 555 Goetz, S. J., Hansen, M., Houghton, R. A., Walker, W., Laporte, N., & Busch, J. (2015).
556 Measurement and monitoring needs, capabilities and potential for addressing reduced
557 emissions from deforestation and forest degradation under REDD+. *Environmental*
558 *Research Letters*, 10(12), 123001. <https://doi.org/10.1088/1748-9326/10/12/123001>
- 559 Gorman, E. T., Kubalak, D. A., Patel, D., Dress, A., Mott, D. B., Meister, G., & Werdell, P. J.
560 (2019). The NASA Plankton, Aerosol, Cloud, ocean Ecosystem (PACE) mission: an
561 emerging era of global, hyperspectral Earth system remote sensing. In *Proc.SPIE* (Vol.
562 11151, p. 111510G). <https://doi.org/10.1117/12.2537146>
- 563 Halle, F., Oldeman, R., & Tomlinson, P. (1980). *Tropical Trees and Forests: An Architectural*
564 *Analysis*. New York: Springer.
- 565 Hancock, S., Armston, J., Hofton, M., Sun, X., Tang, H., Duncanson, L. I., et al. (2019). The
566 GEDI Simulator: A Large-Footprint Waveform Lidar Simulator for Calibration and
567 Validation of Spaceborne Missions. *Earth and Space Science*, 6(2), 294–310.
568 <https://doi.org/https://doi.org/10.1029/2018EA000506>
- 569 Homolová, L., Malenovský, Z., Clevers, J. G. P. W., García-Santos, G., & Schaepman, M. E.
570 (2013). Review of optical-based remote sensing for plant trait mapping. *Ecological*
571 *Complexity*, 15, 1–16. <https://doi.org/https://doi.org/10.1016/j.ecocom.2013.06.003>
- 572 Ma, L., Hurtt, G., Tang, H., Lamb, R., Lister, A., Chini, L., et al. (2023). Spatial heterogeneity of
573 global forest aboveground carbon stocks and fluxes constrained by spaceborne lidar data
574 and mechanistic modeling. *Global Change Biology*, 29(12), 3378–3394.
575 <https://doi.org/https://doi.org/10.1111/gcb.16682>
- 576 Malhi, Y., Girardin, C. A. J., Goldsmith, G. R., Doughty, C. E., Salinas, N., Metcalfe, D. B., et
577 al. (2017). The variation of productivity and its allocation along a tropical elevation
578 gradient: a whole carbon budget perspective. *New Phytologist*, 214(3), 1019–1032.
579 <https://doi.org/https://doi.org/10.1111/nph.14189>
- 580 Malhi, Y., Girardin, C., Metcalfe, D. B., Doughty, C. E., Aragão, L. E. O. C., Rifai, S. W., et al.
581 (2021). The Global Ecosystems Monitoring network: Monitoring ecosystem productivity
582 and carbon cycling across the tropics. *Biological Conservation*, 253, 108889.
583 <https://doi.org/https://doi.org/10.1016/j.biocon.2020.108889>
- 584 Mitchard, E. T. A., Saatchi, S. S., Baccini, A., Asner, G. P., Goetz, S. J., Harris, N. L., & Brown,
585 S. (2013). Uncertainty in the spatial distribution of tropical forest biomass: a comparison of

- 586 pan-tropical maps. *Carbon Balance and Management*, 8(1), 10.
587 <https://doi.org/10.1186/1750-0680-8-10>
- 588 Mitchard, E. T. A., Feldpausch, T. R., Brienen, R. J. W., Lopez-Gonzalez, G., Monteagudo, A.,
589 Baker, T. R., et al. (2014). Markedly divergent estimates of Amazon forest carbon density
590 from ground plots and satellites. *Global Ecology and Biogeography*, 23(8), 935–946.
591 <https://doi.org/https://doi.org/10.1111/geb.12168>
- 592 Mulatu, K. A., Mora, B., Kooistra, L., & Herold, M. (2017). Biodiversity Monitoring in
593 Changing Tropical Forests: A Review of Approaches and New Opportunities. *Remote*
594 *Sensing*. <https://doi.org/10.3390/rs9101059>
- 595 Réjou-Méchain, M., Mortier, F., Bastin, J.-F., Cornu, G., Barbier, N., Bayol, N., et al. (2021).
596 Unveiling African rainforest composition and vulnerability to global change. *Nature*,
597 593(7857), 90–94. <https://doi.org/10.1038/s41586-021-03483-6>
- 598 Riutta, T., Malhi, Y., Kho, L. K., Marthews, T. R., Huaraca Huasco, W., Khoo, M., et al. (2018).
599 Logging disturbance shifts net primary productivity and its allocation in Bornean tropical
600 forests. *Global Change Biology*, 24(7), 2913–2928. <https://doi.org/10.1111/gcb.14068>
- 601 Saatchi, S. S., Harris, N. L., Brown, S., Lefsky, M., Mitchard, E. T. A., Salas, W., et al. (2011).
602 Benchmark map of forest carbon stocks in tropical regions across three continents.
603 *Proceedings of the National Academy of Sciences*, 108(24), 9899–9904.
604 <https://doi.org/10.1073/pnas.1019576108>
- 605 Schimel, D. S., & Poulter, B. (2022). The Earth in Living Color - NASA's Surface Biology and
606 Geology Designated Observable. In *2022 IEEE Aerospace Conference (AERO)* (pp. 1–6).
607 <https://doi.org/10.1109/AERO53065.2022.9843640>
- 608 Sillett, S. C., Graham, M. E., Montague, J. P., Antoine, M. E., & Koch, G. W. (2024). Ground-
609 based calibration for remote sensing of biomass in the tallest forests. *Forest Ecology and*
610 *Management*, 561, 121879. <https://doi.org/https://doi.org/10.1016/j.foreco.2024.121879>
- 611 ter Steege, H., Pitman, N. C. A., Phillips, O. L., Chave, J., Sabatier, D., Duque, A., et al. (2006).
612 Continental-scale patterns of canopy tree composition and function across Amazonia.
613 *Nature*, 443(7110), 444–447. <https://doi.org/10.1038/nature05134>
- 614 Stovall, A. E. L., & Shugart, H. H. (2018). Improved Biomass Calibration and Validation With
615 Terrestrial LiDAR: Implications for Future LiDAR and SAR Missions. *IEEE Journal of*
616 *Selected Topics in Applied Earth Observations and Remote Sensing*, 11(10), 3527–3537.
617 <https://doi.org/10.1109/JSTARS.2018.2803110>
- 618 Tuomisto, H., Van doninck, J., Ruokolainen, K., Moulatlet, G. M., Figueiredo, F. O. G., Sirén,
619 A., et al. (2019). Discovering floristic and geocological gradients across Amazonia.
620 *Journal of Biogeography*, 46(8), 1734–1748.
621 <https://doi.org/https://doi.org/10.1111/jbi.13627>
- 622 Ustin, S. L., Asner, G. P., Gamon, J. A., Fred Huemrich, K., Jacquemoud, S., Schaepman, M.,
623 & Zarco-Tejada, P. (2006). Retrieval of quantitative and qualitative information about plant
624 pigment systems from high resolution spectroscopy. In *International Geoscience and*
625 *Remote Sensing Symposium (IGARSS)*. <https://doi.org/10.1109/IGARSS.2006.517>

626 Zhang, F., Chen, J. M., Chen, J., Gough, C. M., Martin, T. A., & Dragoni, D. (2012). Evaluating
627 spatial and temporal patterns of MODIS GPP over the conterminous U.S. against flux
628 measurements and a process model. *Remote Sensing of Environment*, 124, 717–729.
629 <https://doi.org/https://doi.org/10.1016/j.rse.2012.06.023>

630