

1 **Temporal changes in streamflow can predict parasitism levels in freshwater prawns**
2 **better than host traits**

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24
25 **Keywords:** Cymothoid; *Macrobrachium*; rainfall; stream discharge; *Telotha henselii*.

27 **Abstract**

28 1. Understanding how changes in the hydrological regime drive parasite loads and dynamics
29 remains a challenging issue in ecological parasitology. Temporal changes in streamflow and
30 rainfall are key hydrological factors that could alter interactions between the parasite and host
31 and affect parasitism levels. However, to investigate the effect of streamflow, rainfall, and its
32 mechanisms, it is important to control host traits that can also influence parasitism levels in
33 freshwater systems.

34 2. Here, we used a cymothoid-palaemonid prawn model to test the combined effects of
35 streamflow and rainfall accounting for host traits to predict variations in parasite loads in
36 stream ecosystems. We collected the palaemonid prawns monthly for two years and measured
37 the variation of rainfall and streamflow (i.e. stream discharge) at the time of prawn sampling.

38 3. Our best model showed that streamflow can predict parasitism levels in prawns better than
39 host traits. We found a higher prevalence and abundance of parasites in reduced streamflow
40 compared to increased ones. We also found higher ectoparasite loads in females rather than
41 males and occurring in autumn and winter rather than in spring and summer. Our results also
42 showed that ectoparasite loads also suffer an effect of host body size, sex, and molt stage, but
43 not host age.

44 4. Our findings show that reduced streamflow can facilitate host finding that favors parasite
45 transmission in dry seasons and increase parasitism levels in stream systems. Temporal
46 variability in streamflow can have a strong influence on parasite loads and dynamics compared
47 to host traits in stream ecosystems.

48 5. Identifying how interactions between aquatic invertebrates and parasites respond to
49 variability in the hydrological regime can help to better understand and predict disease
50 outbreaks as habitat reduction and disturbance continue in stream ecosystems.

51

52 1. INTRODUCTION

53 Parasites represent a significant component of biodiversity, with key roles in most ecosystems
54 of the world (Poulin, 2007; Hechinger et al., 2011; Timi & Poulin, 2020). Although the
55 relevance of parasites in ecosystem functions and conservation has increased in recent years
56 (Poulin, 2021; Gagne et al., 2022), this topic has been neglected in studies on crustacean
57 biology and ecology, particularly in the neotropical freshwater environment (Wunderlich et al.,
58 2011; Smit et al., 2014; Smit et al., 2019). Moreover, freshwater ecosystems are among the
59 most diverse parasite species (Poulin, 2016) and face increasing high anthropogenic pressure
60 and hydrological disturbance (Poff, 2018; Boos et al., 2019; Palmer & Ruhí, 2019; Zanghi et
61 al., 2023; Rumschlag et al., 2023), which calls for more studies to understand how these
62 environmental changes interact with the properties of hosts to affect parasite infestations and
63 dynamics.

64 Environmental changes in hydrological regime act as a strong driver to sustain and alter
65 the dynamics of macroinvertebrates and fish communities in freshwater ecosystems (Poff &
66 Zimmerman, 2010; Jellyman et al., 2013; Jellyman et al., 2014; Northington & Webster, 2017;
67 McIntosh et al., 2018; Herbst et al., 2019; Mendes et al., 2021). In rivers and stream habitats,
68 rainfall and/or precipitation play a central role in affecting the interaction between the biota
69 and the physical factors of the stream (Palmer & Ruhí, 2019; McIntosh et al., 2024). Changes
70 in river flow are driven by variation in precipitation, that can change the flow of stream many
71 times (Paull et al., 2012; Botter et al., 2013; Olsson et al., 2024), and influence the dispersal
72 (Miyazono & Taylor, 2013), the predation (Jellyman et al., 2014; McIntosh et al., 2018),
73 community stability (Greig et al., 2022) and food webs (McIntosh et al., 2024). These changes
74 in hydrological regime have also been suggested to affect host-parasite interactions
75 (Marcogliese, 2016; Shearer & Ezenwa, 2020), through parasite transmission mechanisms and
76 host susceptibility to parasites (Blasco-Costa et al., 2013; Penczykowski et al., 2014), which

77 can increase or decrease parasite loads (Altizer et al., 2006; Groner et al., 2018; Vestbo et al.,
78 2019; Shearer & Ezenwa, 2020; Tadiri & Ebert, 2023). For example, some studies have
79 suggested that the reduction in streamflow (i.e., stream discharge) might increase host
80 aggregation and facilitate host finding by the parasite that also is more clumped into the stream
81 (Brunner & Ostfeld, 2008; Paull et al., 2012; Shearer & Ezenwa, 2020). However, most studies
82 have focused on the effects of increasing temperature rather than variation in streamflow and
83 their mechanisms (Marcos-López et al., 2010; Löhmus & Björklund, 2015; Marcogliese, 2016;
84 Lymbery et al., 2020). Therefore, it is fundamental in aquatic parasitology to disentangle the
85 contribution of these sources of variation, particularly in the hydrological regime, to better
86 understand how these changes in streamflow and rainfall correlate with host traits to affect
87 parasite loads over time (Poulin, 2007; Kamiya et al., 2014; Poulin, 2021; Tadiri & Ebert,
88 2023).

89 Furthermore, the dynamic of host-parasite systems is also driven by host traits that can
90 influence the likelihood of infection on several scales of observation (Poulin, 2007; Castrillo,
91 2018; Schmid-Hempel, 2021). It is important in ecological studies, particularly in stream
92 ecosystems, to control for host trait effects such as host body, sex, age and molt that can affect
93 host exposure and susceptibility to parasite infections commonly found in the field (Altizer et
94 al., 2006; Cattadori et al., 2019; Izhar et al., 2020). For example, host body size often correlates
95 positively with parasite abundance (Poulin, 2007; Kamiya et al., 2014). Other host traits such
96 as host age, sex, and molt are also correlated with parasite variations between individuals
97 within populations (Christe et al., 2007; Ben-Ami, 2019, Izhar et al., 2020), which can be age-
98 dependent (Wunderlich et al., 2022) or sex-biased (Poulin & Forbes, 2012). In crustaceans, the
99 molting stage is a key growth factor, while it can expose hosts to parasite infections (Duneau
100 & Ebert, 2012) or reduce the likelihood of parasite penetration (Izhar et al., 2020). Thus, it is
101 important to account for the role of host traits when investigating the effects of streamflow and

102 rainfall on parasite dynamics, because these traits could confound the results and interpretations
103 of the focal predictors that freshwater ecologists seek to uncover (Timi & Poulin, 2020).

104 Here, we investigate the influence of the hydrological regime, such as streamflow and
105 rainfall, on the accounting for host traits on parasitism levels in a species of palaemonid prawn
106 from freshwater streams in the Brazilian Southern. To test whether the hydrological regime can
107 affect parasite infestation, in this study we examined the effects of streamflow (Q , m^3/s),
108 rainfall, seasons, host size, sex, age, molt stage, and their interactions on prevalence and
109 abundances of a cymothoid isopod, *Telotha henselii* (Von Martens, 1869) parasitizing the
110 palaemonid prawn *Macrobrachium brasiliense* (Heller, 1862), as our model. We predict that a
111 reduction in streamflow in dry seasons could indirectly facilitate the chances of parasite
112 encounters and transmission due to greater aggregation of hosts and free-swimming parasites
113 during this season. Therefore, we also expect that prawns collected in dry seasons (i.e., low
114 rainfall and reduced streamflow) would have higher levels of cymothoid isopod infestation
115 than in wet seasons (i.e., high rainfall and normal streamflow). Finally, we also expected that
116 the host traits may have a potential additive effect on parasitism levels. Since host traits are
117 crucial to driving parasite dynamics, we expect a potential additive effect of body size, sex,
118 age, and molt stage on cymothoid isopod infestations in *M. brasiliense* populations.

119

120 **2. METHODS**

121 **2.1. Host-parasite model system**

122 Cymothoid isopods are a diverse group of parasitic crustaceans with a broad distribution
123 throughout the world (Williams & Boyko 2012). Most cymothoids are parasites of marine,
124 freshwater, and brackish teleosts (Smit et al., 2014), but can also parasitize freshwater
125 palaemonid prawns in tropical South American river systems (Wunderlich et al., 2011;
126 Andrade et al., 2020). Cymothoids can also cause damage and affect the growth of their hosts

127 (Smit et al., 2019; Rose et al., 2020). Adults of *Telotha henselii* are well-known hematophagous
128 cymothoid ectoparasites on fish, mostly found on skin and gills, and juvenile males use
129 palaemonid prawns (e.g., *Macrobrachium brasiliense*) as optional intermediate hosts besides
130 final hosts (Trilles & Öktener 2004; Wunderlich et al., 2011; Andrade et al., 2020; Anaguano-
131 Yancha & Pilatasig, 2022; Fujita et al., 2023). These cymothoids also exhibit protandrous
132 hermaphroditism in which male reproductive organs develop first (i.e. that can attach to the
133 prawn), and when parasitizing a fish host, they transform into a sexually mature female (Cook
134 & Munguia, 2015; Rose et al., 2020). Therefore, the cymothoid-palaemonid system is a
135 tractable model that could facilitate our understanding of how changes in streamflow and host
136 traits can influence host-ectoparasite interactions and the presence of juvenile cymothoids in
137 freshwater ecosystems.

138

139 **2.2. Study area and sampling design**

140 As part of a long-term study of ecological parasitology, palaemonid prawns were collected for
141 two years (i.e., 1995 and 1998), in a stream of first and second order (Águas Claras stream)
142 from the Pardo River basin (21° 18' 30'' S - 47° 36' 16'' W), in southwest Brazil (Figure 1).
143 The Águas Claras stream is a tributary of the Pardo River that has a drainage basin of 10,694
144 km² and a course of approximately 550 km, with the dry season occurring from April to
145 September and the wet season from October to March (Alves et al., 2018). The bottom of the
146 stream bottom was mainly composed of sand and the depth rarely exceeds 1 m. We collected
147 prawns monthly from three sites at the same time, for 2 hours each by two people from 10:00
148 am to 12:00 noon. At each site, the prawns were manually collected by sweeping sieves (50 cm
149 in diameter and mesh size of 4 mm mesh size) near the riverbank, in places with marginal
150 aquatic vegetation consisting mainly of species of Hydrocharitaceae (aquatic herbs), Juncaceae
151 (reeds), and Pontederiaceae (water hyacinths) from the stream (Mantelatto & Barbosa, 2005).

152 Prawns were individualized in plastic bags while still in the field and kept in thermic box on
153 crushed ice until they arrived in the laboratory, where the specimens were euthanized by
154 thermal shock (at 3°C, for 30 min), fixed, and preserved in 70% ethanol.

155

156 **2.3. Measures of host traits**

157 In the laboratory, the prawn specimens were identified according to Melo (2003). Posteriorly,
158 prawns were submitted to biometry based on body measures (CL, carapace length – from the
159 base of the rostral spine to the opposite region of the carapace; and TL, total length –
160 comprising total body size, from the base of the rostral spine to the extremity of the telson),
161 with a 0.05 mm precision caliper; and weigh (W, total wet weight – with a digital balance of
162 0.01 g), and also characterized according to the age, sex, and molt stage. In this study, we
163 consider age as a function of the developmental phase (juvenile and adult), using as proxies:
164 1) the size at onset of morphological maturity for males and females (9 and 8 mm CL,
165 respectively, according to Nogueira et al., 2019); 2) sex of each exemplar, by checking the
166 presence of the appendix masculine in the endopod of the second pair of pleopods (Nogueira
167 et al., 2020); and 3) molt stage (intermolt, premolt, and molting or ecdysis) determined in each
168 specimen by exoskeleton hardness inspection, especially the carapace (Drach &
169 Tchernigovtzeff 1967).

170

171 **2.4. Streamflow and rainfall measurements**

172 The monthly rainfall, streamflow (i.e. stream discharge), and habitat area data were obtained
173 from an archival hydrological data set from the Department of Waters and Electric Energy of
174 the State of São Paulo (DAEE, 2023). We used a rainfall station near the sampling sites to
175 obtain total rainfall (mm), average monthly stream discharge (Q, m³/s), velocity (m/s), depth
176 (m), width (m) and area (m²). We used the streamflow (Q) based on the velocity-area method

177 at the time of prawn sampling. We used three rainfall variables; total rainfall (mm) for the
178 month in which the sample was collected (R_t), total rainfall for the month before sample
179 collection (R_{t-1}), and total rainfall two months before sample collection (R_{t-2}). Three different
180 measures were used to test whether current (immediate effect) or previous (delayed effect)
181 rainfall was more strongly associated with ectoparasite loads. Together, changes in the rainfall
182 regime are associated with changes in streamflow.

183

184 **2.5. Parasitological screening**

185 The palaemonid prawns were examined and the ectoparasites found were removed from each
186 animal and measured to obtain the total length (TL) under a stereomicroscope (Zeiss® Stemi®
187 SV-6) attached to a computer image analysis system using Zeiss® KS-100 3.0 software (0.01
188 mm precision). Posteriorly, the ectoparasites were fixed in 70% ethanol and subsequently
189 identified through specific characteristics (e.g., 2nd pereopod showing five spines in the inner
190 margin of the propodus and the 6th pereopod showing four spines in the propodus and one in
191 merus), according to Lemos de Castro (1985) and Andrade et al. (2020). The abundance of
192 ectoparasites was considered as the total number of cymothoid isopods in each host individual,
193 which can be categorized in terms of infestation levels as follows: 0 = no infestation; 1 = low
194 infestation; and 2-3 = high infestation (Horton & Okamura, 2001; Kawanishi et al., 2016). The
195 prevalence was calculated as the proportion of prawns infected by cymothoid isopods with the
196 total number of hosts analyzed, and the mean intensity was calculated as the total number of
197 cymothoids in a sample of infected host species divided by the number of infected individuals
198 of the host species in the sample (Bush et al., 1997).

199

200

201 **2.6. Data analysis**

202 We used generalized linear models (GLM) to test the relationship between streamflow (i.e.
203 stream discharge, Q) and parasitological indices (i.e., prevalence and parasite abundances), and
204 whether this pattern can vary between sexes. To estimate the effects of streamflow, rainfall,
205 habitat area, and host traits (i.e., total length, sex, age, and molt stage) on ectoparasite
206 abundances (i.e., response variable), we fit generalized linear mixed models (GLMMs), using
207 Poisson and negative binomial distributions (Zuur et al., 2013). Streamflow, habitat area,
208 rainfall, year, host traits, and their interaction were the main predictors, while month and year
209 were included as random intercepts. Individual IDs (i.e., prawn ID) nested within each month
210 were used as random factors in all models. We used prawn ID as an observation-level random
211 effect to control overdispersion in our parasite count data (Zuur & Ieno, 2016). Before fitting
212 the model, we evaluated the multicollinearity of the predictors using the variance inflation
213 factor (VIFs) with the R performance packages (Lüdecke et al., 2021), and excluded those with
214 VIFs >10, which represent high multicollinearity among predictors (Borcard et al., 2011). We
215 also used diagnostic plots to test for normal distribution and equal variance of residuals with
216 the R package DHARMA (Hartig, 2022).

217 We used the Akaike Information Criterion (AIC) and Bayesian Information Criterion
218 (BIC) to select the best competitive model with R performance packages (Lüdecke et al., 2021).
219 After choosing the best model, we used the semipartial coefficients of determination (partR2)
220 to check the variation to individual predictors of this model (Stoffel et al., 2021). The partR2
221 can decompose the variance of R^2 into components uniquely explained by individual predictors
222 (Stoffel et al., 2021). GLM and GLMM analyses were performed in R v. 4.2.1 (R Development
223 Core Team, 2022) with packages lme4 (Bates et al., 2015) and glmmTMB (Magnusson et al.,
224 2020), respectively.

225

226 3. RESULTS

227 We found that streamflow can predict parasitism levels in prawns better than host traits (Table
228 1). Of the models that we compared, the one that included streamflow (Q), concurrent rainfall
229 (R_t), and habitat area received the most support (Table 1). We found a higher prevalence
230 (GLMM: estimate \pm se: -0.756 ± 0.167 , $p < 0.001$, $R^2_{adj} = 0.72$; Figure 2A) and abundance
231 (GLMM: estimate \pm se: -0.625 ± 0.203 , $p < 0.002$, $R^2_{adj} = 0.77$; Figure 2B) of parasites in
232 reduced streamflow (i.e., $Q < 2$) compared to increased ones (i.e., $Q > 2$; Figure 2). We also
233 found a significant difference between males and females in total abundance to streamflow
234 (GLMM: estimate \pm se: 0.951 ± 0.307 , $p = 0.002$; Figure 2B), but not in prevalence (GLMM:
235 estimate \pm se: -0.101 ± 0.345 , $p = 0.769$; Figure 2A). The higher parasite infestations (i.e., two
236 to three parasites per infested prawn) were observed in lower rainfall (dry season) rather than
237 in higher rainfall (wet season) (Figure 3).

238 Overall, we found a prevalence of 16.4 % of a total of 1,012 prawns collected over two
239 years (Table 3). In 1995, the prevalence of ectoparasites was higher in winter and autumn
240 (18.9% and 16.6%, respectively) rather than in spring and summer (11.9 and 9.2%, respectively
241 - Table 3). For 1998, the prevalence in autumn was also higher (34.1%) than in the summer
242 (14.2%; Table 3). The mean intensity of the infection was slightly higher in 1998 than in 1995
243 (Table 3). The abundance of ectoparasites varied between seasons, with a higher infestation
244 (i.e., two to three parasites per infested prawn) in the autumn of both years (Figure 4a, Table
245 3), but there are no differences between the two years in our model (Wald = -1.849 ; $p = 0.064$;
246 Table 1, Figure 4b). We also found higher ectoparasite loads in females than in males in
247 autumn/winter than in spring/summer.

248 Our results also showed a significant effect of host traits on ectoparasite abundances
249 (Figure 5, Table 1), with the size of the host's body (one to three parasites in prawns size ≤ 35
250 mm TL – Figure 5a; Table 1); host sex (parasitism in female hosts more abundant than in males

251 – Figure 5c; Table 1); and host molt stage (parasite abundance following this hierarchical
252 sequence of stages of molt: intermolt > ecdysis > premolt – Figure 5d; Table 1). However, no
253 effect of age on parasitism was found (the abundance of parasites did not differ between
254 juveniles and adults – Figure 5b; Table 1).

255 Of all fixed variables included in the GLMM models, season was excluded because it
256 showed high multicollinearity with $VIF > 10$ (Table 2). The relative importance of the
257 predictors used in the GLMM was decomposed to check the variation with the individual
258 predictors of the best model (Figure 6). Of three hydrological factors, streamflow seems to be
259 the best predictor of the ectoparasite infestation model, presenting the largest inclusive R^2
260 (Figure 6A), negative structure coefficient (Figure 6B) and negative beta weight (Figure 6C),
261 while host length, sex (female), and molt (ecdysis) contributed positively to the predicted
262 model (Figure 6A-C; Table 1).

263

264 **4. DISCUSSION**

265 Our results reveal that the effect of streamflow on parasitism levels surpasses those of host
266 traits. Changes in hydrological regime act as a strong drive to sustain and alter the dynamic of
267 benthic invertebrate communities (Poff et al., 2018; McIntosh et al., 2018; Herbst et al., 2019;
268 Palmer & Ruhi, 2019), as well as infectious disease and parasite dynamics in freshwater
269 ecosystems (Altizer et al., 2006; Paull et al., 2012; Penczykowski et al., 2014; Groner et al.,
270 2018; Tadiri & Ebert, 2023). Freshwater organisms are exposed to variable rainfall regimes
271 that lead to changes in streamflow, and can affect their behaviour and the interaction between
272 host and parasite in stream ecosystems (Paull et al., 2012; Penczykowski et al., 2014; Reynolds
273 et al., 2019).

274 The effect of variation in rainfall regime that leads to a temporal reduction in
275 streamflow can facilitate host finding by cymothoid parasites and, consequently, increase

276 parasitism levels in a more confined environment (Paull et al., 2012; Penczykowski et al., 2014;
277 Shearer & Ezenwa, 2020). Some mechanisms have been suggested to explain the influence of
278 hydrological regime such as streamflow on successful transmission of parasites during this
279 period (Sousa & Grosholz, 1991; Penczykowski et al., 2014; Smit et al., 2019). This reduction
280 in streamflow can affect host behaviour and aggregation, increasing the probability of contact
281 between parasite and host (Paull et al., 2012; Reynolds et al., 2019; Schmid-Hempel, 2021).
282 Studies have shown that a high density of hosts in a more confined stream, as also shown by
283 our results, can accelerate contact between those most susceptible, as well as intensify the risk
284 of infection for host individuals (Lafferty et al., 2015; Buck et al., 2017). The same can occur
285 with parasites and their transmission stages because they are also more clumped together across
286 the environment, resulting in hosts experiencing a higher infection level (Brunner & Ostfeld,
287 2008; Paull et al., 2012; Davenport et al., 2024). Furthermore, a recent study showed that
288 disease outbreaks in a freshwater zooplankton are closely linked to the concentration of the
289 transmission stage in the environment (i.e., mean spore concentration increased prior to disease
290 outbreaks in hosts; Davenport et al., 2024), reinforcing our prediction that high aggregation of
291 free-swimming parasites in reduced streamflow can favor the higher infestation of isopods in
292 the prawns. With the reduction in depth/width in the stream with a lower flow/discharge, the
293 incidence of light is high (Bernhardt et al., 2022), which can facilitate the attack of the manca
294 juveniles (i.e., free-swimming parasites) on the prawn hosts. Manca juveniles are positively
295 phototactic (i.e., they spend the day on the surface) and possess large eyes that help the parasite
296 locate the shadows of potential prawns at the bottom of the sand where they are foraging (Smit
297 et al., 2019). However, depending on the parasite species and system, the increase in light can
298 also reduce the infectivity of parasites that use spores to transmit (Shaw et al. 2020). Studies
299 have shown that parasites respond to host and environmental cues (e.g. light) in ways that bring
300 them into the habitat (i.e., a few centimeters of an appropriate host's "active space") commonly

301 frequented by their potential hosts (Rea & Irwin, 1994; Goater et al., 2014; Cook & Munguia
302 2015; Shaw et al. 2020). An interesting study also showed that males in parasitic crustaceans
303 were more active in light with reduced activity in darkness (Hunt et al., 2021), with light being
304 their dominant stimulus (Bandilla et al., 2007). This makes sense because in low-flow habitat,
305 manca juveniles are probably closer to prawns, which might increase their visual capacity to
306 detect the hosts.

307 In addition, streamflow velocity can be another mechanism that drives parasite
308 transmission success (Sousa & Grosholz, 1991; Reynolds et al., 2019). Fast streamflow might
309 decrease the capacity of the parasite (i.e., manca juveniles) to detect and attach to the hosts
310 (Smit et al., 2019). For example, in an interesting experiment with fish, the parasite
311 transmission rate was higher in shoals exposed to interrupted flow (Reynolds et al., 2019),
312 resulting in a higher parasitic prevalence compared to a continuous water flux system.
313 Furthermore, some studies have shown that slow water flow downstream can accumulate more
314 infective stages of free-living, which can influence the probability of host and parasite
315 encounter, resulting in high parasitism in this part of the stream (Blasco-Costa et al., 2013;
316 Resataris & Byers, 2023). In contrast, during the wet seasons, excessive rainfall and streamflow
317 can drag away and reduce the abundance of the free-living infective stages of the parasites due
318 to higher water volume, a fact verified in several helminth and crustacean parasites, especially
319 in mancae of cymothoids (Stromberg, 1997; Altizer et al., 2006). Our results are consistent
320 with the above pattern, showing that the higher prevalences and abundances of ectoparasites
321 were found in dry (slow-flowing) rather than wet (fast-flowing) seasons.

322 Few published studies have shown strong evidence for the role of rainfall and/or
323 precipitation as a key driver of seasonality in parasitism in freshwater and terrestrial ecosystems
324 (Aneesh et al., 2013; Hiestand et al., 2014; Gagne & Blum, 2016; Jemi et al., 2020; Shearer &
325 Ezenwa, 2020). Although these studies have shown the influence of rainfall on parasite

326 infections, the mechanisms for this were not well elucidated. Some authors recorded higher
327 parasite loads associated with low rainfall, suggesting a gradual increase in host susceptibility
328 and exposure to helminth infections during the dry season (Gagne & Blum, 2016; Shearer &
329 Ezenwa, 2020). In aquatic ecosystems, cymothoids parasitizing fish were also more abundant
330 during the autumn season (i.e., dry season in India), suggesting that this season can facilitate
331 the infestation of ectoparasites (Aneesh et al., 2013; Jemi et al., 2020). Another interesting
332 study showed that drought in stream fish communities can intensify parasitic crustacean
333 infestations (Medeiros & Maltchik, 1999), which could support our study, with a more intense
334 infestation of prawns by these juvenile cymothoid isopods during the dry seasons. Precipitation
335 also has a profound effect on the free-living stages of terrestrial mammal helminths because
336 some parasites use aquatic invertebrates as their intermediate hosts (Hiestand et al., 2014;
337 Carlson et al., 2017). This point also suggests a more complex interaction and certain
338 dependence that some terrestrial animals have with freshwater ecosystems and their changes.
339 Furthermore, changes associated with rainfall and/or precipitation in parasitism can vary
340 depending on key host traits (Shearer & Ezenwa, 2020).

341 We also found an effect of host traits on parasite infestation. Previous studies have
342 found a positive relationship between ectoparasite infestation and host traits, suggesting
343 heterogeneity in transmission, susceptibility, tolerance, and resistance to determine parasite
344 infections (Johnson & Hoverman, 2014; Izhar et al., 2020; Clark et al., 2021; Schmid-Hempel,
345 2021). We found that body length, sex, and molt stage (but not host age), influenced the
346 abundance of ectoparasites. The size of the host body is a critical characteristic that influences
347 several host-parasite systems (Poulin, 2007; Schmid-Hempel, 2021). For example, parasite
348 abundance (or loads) often correlates positively with host body size (Kamiya et al., 2014;
349 Schmid-Hempel, 2021). Furthermore, body size correlates with other host traits that affect the
350 likelihood of acquiring infections (Cooper et al., 2012; Blasco-Costa & Poulin, 2013; Schmid-

351 Hempel, 2021), such as longevity. The host sex has also been an important factor in explaining
352 sex-biased parasitism (Poulin & Forbes, 2012). In many parasite-host systems, males are more
353 parasitized than females (Cozzarolo et al., 2019), however, in our study female prawns have a
354 higher abundance of ectoparasites than males. In general, the prevalence or intensity of
355 parasites in arthropods is also higher in males than females (Klein et al., 2004), but in some
356 studies it was not observed (see Sheridan et al., 2000). A potential explanation of a female bias
357 might be caused by differences in immunity and parasite exposure between the sexes (Poulin,
358 2007, Cozzarolo et al., 2019; Schmid-Hempel, 2021). Males and females are likely to exhibit
359 differences in behaviour and spatial aggregation in the host population (Nogueira et al., 2019;
360 Bardera et al., 2020), which can influence parasite exposure and susceptibility and favor sex-
361 biased parasitism (Poulin, 2007; Christe et al., 2007).

362 Another key factor that can interfere with infestation levels is the molting process
363 (Groner et al., 2018; Izhar et al., 2020). Our results suggest that the abundance of parasites was
364 positively associated with the ecdysis stage. In crustaceans, the molting stage is a key growth
365 factor, while it can expose hosts to parasite infections (Duneau & Ebert, 2012) or reduce the
366 likelihood of parasite infestations (Izhar et al., 2020). Although we did not find an effect of
367 host age on the abundance of ectoparasites, age has been correlated with parasite variation
368 between individuals within invertebrates (Ben-Ami, 2019, Izhar et al., 2020) and vertebrates
369 host populations (Wunderlich et al., 2022). For example, Izhar et al. (2020) found that older
370 adults have better immune responses than younger adults and juveniles of *Daphnia magna*.
371 Finally, our host trait data suggested an additive effect on parasite infestation and underscores
372 the importance of controlling these traits in ecological studies of host-parasite systems.

373 In addition, the shortage in resource availability during the dry season (i.e. in a more
374 confined environment), could also lead to poor host body conditions and immune function,
375 with an increase in parasite prevalence and intensity (Marshall et al., 2008; Shearer & Ezenwa,

376 2020). Another important gap that could be considered is a potential combined effect of
377 streamflow and habitat size on the variation of aquatic populations and their parasites. Habitat
378 size has been suggested to drive many important ecological processes (McIntosh et al., 2018;
379 McIntosh et al., 2024). However, it is unclear how this combined effect could influence host-
380 parasite dynamics. Furthermore, a study that integrates different stream habitats, local
381 environmental factors, and resource availability in ecological parasitology could produce new
382 broadly applicable strategies to help parasitism management and conservation in freshwater
383 ecosystems (Civitello et al., 2018; Hasik & Siepielski, 2022). Although our study was not
384 designed to address the influence of the variation in resource availability and habitat size, we
385 suggest the need for additional experiments to test the hypothesis that host body conditions
386 should change between smaller and larger habitats, thus modifying the ectoparasite-host
387 dynamics.

388 In conclusion, our study demonstrated that changes in streamflow combined with host
389 traits can influence host-parasite interactions and affect the risk of parasitism in stream
390 ecosystems. More studies are needed to understand how variation in streamflow and habitat
391 structure as a result of climate change and hydrological disturbances can impact the natural
392 flow regime and affect how benthic macroinvertebrates communities will overcome the risk of
393 parasitism in a stream ecosystem that changes. Our findings also revealed that additive effects
394 of streamflow and host traits in predicting the parasitism levels are relevant and should be
395 considered in future studies in freshwater ecosystems.

396

397 **AUTHOR CONTRIBUTIONS**

398 Conceptualization, data interpretation, writing: AW, EM, and MP. Conducting the research,
399 data analysis, and preparation of figures and tables: AW, MP.

400

401 **ACKNOWLEDGEMENTS**

402 We thank members of the CRUSTA's research group (Gustavo Y. Hattori, Lucimari M. Seto,
403 and Fabiano G. Taddei), during field expeditions to Águas Claras stream (Serra Azul
404 municipality, SP), and help with their biometry. MAAP thanks to 'Fundação de Amparo à
405 Pesquisa do Estado de São Paulo' – FAPESP, which indirectly funded this work during the
406 execution of the *Arenaeus* Project (# 1995/09495-2) and to the 'Conselho Nacional de
407 Desenvolvimento Científico e Tecnológico' – CNPq, due to the research fellowship grant (#
408 303286/2016-4 and # 305957/2019-8). ACW also thanks São Paulo State University
409 (UNESP/PROPe) for a postdoctoral research fellowship (Edital 13/2022 - Programa Unesp de
410 Pós-Doutorado- Chamada Pública para Seleção de Bolsistas - Apoio a Grupos de Pesquisa)
411 and indirectly to FAPESP (# 2017/16650–5) due to previous financial support to statistical
412 courses (e.g., Mixed effects modelling) that makes possible the data analyses in this study. We
413 also thank the Associate Editor and two anonymous referees for their helpful comments that
414 improved the quality of the manuscript.

415

416 **CONFLICT OF INTEREST STATEMENT**

417 The authors declare that they have no conflicts of interest.

418

419 **DATA AVAILABILITY STATEMENT**

420 Data supporting the findings of this study are available upon request from the corresponding
421 author.

422

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689 **Table 1.** Results of generalized linear mixed models describing the total abundance of
 690 ectoparasites (i.e., response variables) as a function of streamflow (Q, m³/s), rainfall (mm),
 691 habitat area, years, and host traits (host body size, host sex, host age, and host molt stage).
 692 Parameter estimates, standard error (SE), Wald z-values, p-value, and 95% confidence interval
 693 for the parameters of explanatory variables describing variation in ectoparasite abundances are
 694 provided. Bold values indicate significant variables. The confidence intervals (CI) that exclude
 695 zero are also reported in bold.

Parameter	Parameter estimate	SE	z-value	p-value	95% CI	
					Lower	Upper
Intercept	-3.354	0.610	-5.499	< 0.001	-4.550	-2.159
Streamflow (Q)	-0.985	0.300	-3.275	0.001	-1.575	-3.957
Rainfall (Rt)	-0.003	0.001	-2.887	0.003	-0.005	-0.001
Habitat area	0.409	0.175	2.328	0.019	0.064	0.754
Year	-0.244	0.132	-1.849	0.064	-0.503	0.014
Host body size	0.051	0.007	6.893	< 0.001	0.036	0.065
Host sex (Female)	0.546	0.186	2.925	0.003	0.180	0.912
Host age (Adult)	0.459	0.608	0.755	0.755	-0.732	1.651
Host molt stage (Premolt)	-0.333	0.424	-0.786	0.432	-1.165	0.498
Host molt stage (Ecdysis)	0.511	0.200	2.551	0.010	0.118	0.904

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702 **Table 2.** Variable inflation scores used to select variables included in the GLMM models.
 703 Predictor terms (term), variance inflation factor (VIF) and 95% confidence interval for VIF
 704 (VIF 95% CI). Bold values indicate exclusion of VIF > 10

705				
706	Term	VIF	VIF 95% CI	
707			Lower	Upper
708	Streamflow (Q)	9.51	8.11	11.24
709	Rainfall (Rt)	5.68	5.09	6.35
710	Habitat area	8.58	7.66	9.63
711	Season	98.92	87.71	111.59
	Year	6.48	5.80	7.26
	Host sex	1.04	1.01	1.20
	Host body size	1.06	1.02	1.19
	Host age	1.07	1.03	1.19
	Host molt stage	1.04	1.01	1.22

712 **Table 3.** Variation of the parasitism indices of the *Telotha henselii* cymothoid isopod on *Macrobrachium brasiliense* prawn with seasons,
 713 years, and host traits. Where: Mean \pm SD = average \pm standard deviation.

Season/Year	Total length		Total length		Prevalence	Mean	Intensity
	Male		Female			Intensity	range
	n	Mean \pm SD	n	Mean \pm SD	%	Mean \pm SD	Min-Max
1995							
Winter	81	15.3 \pm 4.8	167	15.1 \pm 3.8	18.9	1.04 \pm 0.20	1-2
Spring	82	17.5 \pm 4.3	120	15.7 \pm 3.6	11.9	1	1
Summer	59	15.3 \pm 4.5	71	15.7 \pm 3.6	9.2	1.08 \pm 0.29	1-2
Autumn	99	14.2 \pm 3.8	100	15.5 \pm 3.8	16.6	1.12 \pm 0.41	1-3
Total	321	15.5 \pm 4.5	458	15.4 \pm 3.7	14.9	1.06 \pm 0.27	1-3
1998							
Winter	-	-	-	-	-	-	-
Spring	-	-	-	-	-	-	-
Summer	43	28.3 \pm 17.1	105	19.7 \pm 9.5	14.2	1.23 \pm 0.44	1-2
Autumn	16	31.7 \pm 11.8	69	23.9 \pm 8.2	34.1	1.24 \pm 0.51	1-3
Total	59	29.2 \pm 15.7	174	21.3 \pm 9.4	21.5	1.24 \pm 0.48	1-3
TOTAL	380	17.7 \pm 8.9	632	17.1 \pm 6.5	16.4	1.11 \pm 0.35	1-3

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717 **Figure Captions**

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719 **Figure 1.** Map of the study area showing the geographic location of the Águas Claras stream
720 (below square), the Pardo River basin (above square, São Paulo State, Brazil), where the
721 freshwater prawns *Macrobrachium brasiliense* (Heller, 1862) and the cymothoid *Telotha*
722 *henselli* (Von Mertens, 1869) were captured at three sites simultaneously over two years (1995
723 and 1998).

724

725 **Figure 2.** Influence of streamflow (m³/s) on prevalence (a) and total abundance (b) per month
726 of parasitism on freshwater prawns *Macrobrachium brasiliense* (Heller, 1862) between male
727 and female in the Águas Claras stream during the two years. Rug marks represent the
728 distribution of the data along the y and x axes. R² and significance values are given in the text.

729

730 **Figure 3.** Variation in the total abundance of ectoparasites in *Macrobrachium brasiliense*
731 (Heller, 1862) between the dry and wet seasons based on the variation of the rainfall in the
732 month (mm). On top, a diagram representing our prediction about the effect of dry and wet
733 seasons on the variation in parasitism levels with streamflow (m³/s).

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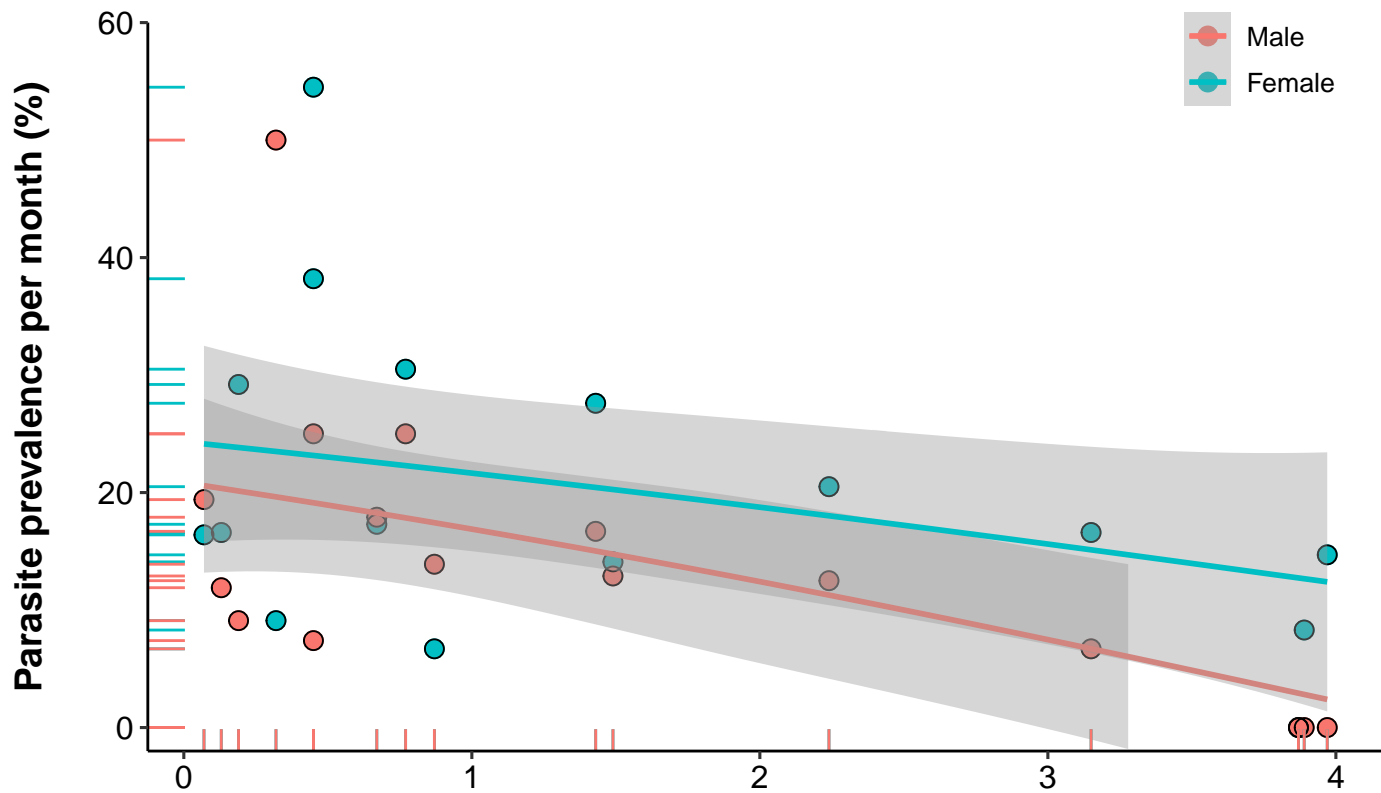
735 **Figure 4.** Variation in total ectoparasite abundance on *Macrobrachium brasiliense* (Heller,
736 1862) between seasons (a) and years (b). Violin plots show the distribution and probability
737 density for each season and year.

738

739 **Figure 5.** Variation in the total abundance of ectoparasites on *Macrobrachium brasiliense*
740 (Heller, 1862) between host size (a), host age (b), host sex (c), and host molt stage (d). Violin
741 plots show the distribution and probability density for each host trait.

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743 **Figure 6.** Forest plots for the comparison of part R^2 (coefficient of determination) for the model
744 inclusive R^2 (A), structure coefficients (B) and beta weights (C) including confidence intervals
745 (CI) for the ectoparasite infestation model. Total rainfall (Rt) and streamflow (Q).

A**B**