

1 Effects of dissolved organic matter and ecocorona formation on
2 the toxicity of micro- and nanoplastic particles to *Daphnia* - A
3 meta-analysis

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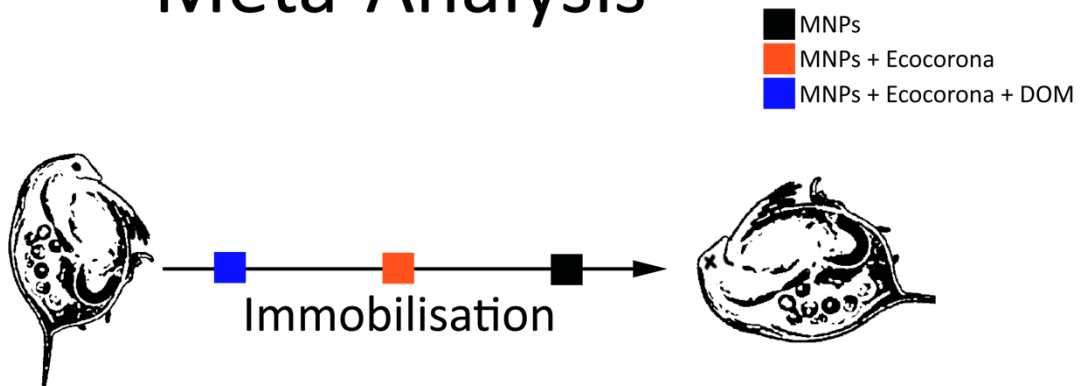
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Meta-Analysis



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24 Highlights

- 25 • Micro- and nanoplastic particles are known to increase *Daphnia* mortality.
- 26 • Dissolved organic matter and ecocorona formation may moderate these
- 27 effects.
- 28 • Meta-analysis shows attenuating effects of ecocorona and DOM on
- 29 microplastic toxicity.
- 30 • Moderating effects depend on DOM type and experimental setup.
- 31 • Moderating effects are stronger in co-exposure than when incubated before
- 32 exposure.

33

34 1. Abstract

35 Significant quantities of micro- and nanoplastic particles (MNP) end up in the
36 environment, either due to larger plastic debris breaking down or by entering directly
37 as MNPs. Effects of MNPs on organisms have been increasingly reported in recent
38 years, with a large number of studies conducted on water fleas of the genus *Daphnia*.
39 Most of the available studies used pristine particles that have not been exposed to the
40 environment or to organic substances. In natural environments, however, proteins,
41 organic substances and, if the particles are large enough, bacteria attach to MNP,
42 forming an ecocorona or biofilm on the particles' surface. How the formation of an
43 ecocorona influences MNP toxicity is still uncertain. While some studies suggest that
44 ecocorona formation can mitigate the negative effects of MNP on organisms, other
45 studies did not find such associations. In addition, it is unclear whether the ecocorona
46 itself is attenuating the effects of MNP or whether dissolved organic matter (DOM)
47 affects toxicity indirectly such as by increasing *Daphnia's* resilience to stressors in
48 general. To draw more solid conclusions about the direction and size of the mediating
49 effect of DOM and ecocorona formation on MNP-associated immobilization in *Daphnia*
50 spp., we synthesized evidence from the published literature and compiled 305 data
51 points from 13 independent studies. The results of our meta-analysis show that the
52 toxic effects of MNP are likely reduced in the presence of certain types of DOM. We
53 observed similar mediating effects when MNP were incubated in media containing
54 DOM before the exposure experiments, although to a lesser extent. Future studies
55 designed to disentangle the effects of the ecocorona itself from the general effects of
56 DOM will contribute to a deeper mechanistic understanding of MNP toxicity in nature
57 and enhance the reliability of MNP risk assessment.

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64 2. Introduction

65 Plastics are widely used in industry and are present in almost all areas of our daily
66 lives. Plastic materials are used for food packaging, in transportation and construction,
67 for recreation activities, clothing and personal care products, research and agriculture
68 (Andrady & Neal, 2009). In 2021, global plastic production was estimated to be 390.7
69 million tones (Plastics Europe, 2022). Through inappropriate disposal, large amounts
70 of plastic end up in the environment (Barnes et al., 2009). Over time, they break down
71 into smaller fragments, eventually degrading into microplastic (100 nm - 5 mm) and
72 nanoplastic (< 100 nm) particles (MNP; GESAMP, 2016). In addition, plastic particles
73 enter the environment directly, for example in the form of intentionally added
74 microbeads, textile fibers and tire abrasion. Once in the environment, MNPs are
75 transported between environmental compartments, enter water bodies through surface
76 runoff and wastewater, and accumulate in oceans, rivers and lakes (Koutnik et al.,
77 2021; Nava et al., 2023; Petersen & Hubbart, 2021).

78 The number of studies investigating the potential harm of MNPs to organisms has
79 increased substantially over the last years (Barbosa et al., 2020). MNPs are ingested
80 by animals and can block their digestive tract, injure gut epithelia and change the gut
81 microbiome (Qiao et al., 2019; Susanti et al., 2020). If small enough, particles can
82 translocate into tissues and cells (Hara et al., 2020; Von Moos et al., 2012) and induce
83 inflammatory responses and cellular damage (Solomando et al., 2020; M. Zhang et al.,
84 2022). On an organismic level, effects on body growth, reproduction and survival have
85 been reported, among others (Lahive et al., 2019; Sussarellu et al., 2016; Ziajahromi
86 et al., 2017). In addition to effects elicited by the particles themselves, additives such
87 as UV-stabilizers (Song et al., 2021) or plasticizers (Y. Yan et al., 2021) added to the
88 polymers may leach from the particles and induce toxic responses (Zimmermann et
89 al., 2020). Pollutants or pathogens may attach to the particle surfaces and induce toxic
90 effects upon ingestion (vector effect; Rochman et al. 2013; Gkoutselis et al., 2021; but
91 see also do Prado Leite et al., 2022 and Koelmans et al., 2016).

92 Due to their complexity, the effects of MNP cannot be easily generalized. Unlike
93 chemicals which have distinct molecular structures and stable properties and can
94 usually be assigned unique identifiers (e.g., CAS number), MNP are more diverse.
95 Each single particle possesses its own set of chemical and physical properties. These

96 properties include simple characteristics such as polymer type, size and shape, as well
97 as more complex properties like mixtures of plastic-associated chemicals, surface
98 structure and charge, and substances attached to the particles' surface including
99 organic substances from the environment (ecocorona; Nasser et al., 2020), proteins
100 (protein corona; Fadare et al., 2020; Kihara et al., 2020) and bacteria (biofilm; Barros
101 et al., 2020; Shi et al., 2023). In addition, all these properties can change over time.
102 This complexity makes extensive testing necessary if we want to understand how
103 specific MNP properties relate to specific toxicity outcomes and how strong effects are
104 to be expected in natural environments.

105 Water fleas of the genus *Daphnia* are important standard test organisms for aquatic
106 systems that occur in most stagnant and slowly flowing freshwater habitats. As part of
107 the zooplankton, *Daphnia* spp. play an essential role in the aquatic food web. They
108 form a link from producer plankton which contains essential fatty acids to higher trophic
109 levels of the food web and are considered sensitive keystone species. As highly
110 effective filter feeders they are exposed to substances and matter contained in the
111 surrounding water. The non-selective feeding method makes them particularly
112 vulnerable to accidental ingestion of particulate pollutants (Giovio et al., 2020).
113 *Daphnia* spp. are easy to maintain in the lab, as they have short generation times and
114 typically reproduce parthenogenetically, if not stressed. Consequently, *Daphnia* spp.
115 are widely used in acute and chronic standard toxicity tests (OECD, 2004, 2012).
116 *Daphnia magna* is the most tested aquatic invertebrate species in studies investigating
117 the ecotoxicological effects of MNP on organisms (see ToMEx database: Hampton et
118 al., 2022; and Brehm et al., 2023).

119 Several studies have shown effects of MNP on *Daphnia* spp. including a variety of
120 endpoints and particle characteristics (for an overview see Brehm et al., 2023). For
121 example, Eltemsah & Bøhn (2019) observed increased mortality rates, decreased
122 growth, and stimulation of early reproduction at the expense of later reproduction in
123 *Daphnia* exposed to polystyrene microbeads; P. Zhang et al. (2019) observed changes
124 in levels of radical oxygen species (ROS); and Lin et al. (2019) observed changes in
125 the daphnids' swimming activity. While earlier studies often tested only one specific
126 type of MNP, later studies more frequently demonstrated that effects are not uniform
127 across all types of MNP but instead depend on the MNPs' properties. Effects have
128 been shown to depend, among others, on polymer type, particle size and shape,

129 surface charge and the presence of additives (Lin et al., 2019; Saavedra et al., 2019;
130 Schrank et al., 2019; Schwarzer et al., 2022; Zimmermann et al., 2020). Most studies
131 so far have worked with pristine particles (Brehm et al., 2023), i.e., particles that have
132 not had contact with natural environments. Only a few studies have attempted to
133 investigate effects under more realistic circumstances. Whether effects observed in the
134 lab are representative of true effects expected in nature is therefore still under debate
135 (Nasser et al., 2020; Nasser & Lynch, 2016; see also Petersen et al., 2022)

136 In natural environments, an ecocorona forms on the particles within seconds when
137 organic molecules attach to the MNPs' surface (e.g., Rummel et al., 2017). This
138 ecocorona alters the particles' surface structure and charge (Natarajan et al., 2021;
139 Shi et al., 2023; Witzmann et al., 2022), influences their behavior in the water column
140 (Elagami et al., 2022; Fischer et al., 2022) and modifies their attachment rate to cellular
141 surfaces (Ramsperger et al., 2020). These alterations may in consequence also affect
142 outcomes on the organismic level, for instance mediated by changes in uptake and
143 tissue translocation rates (Raftis & Miller, 2019; Triebkorn et al., 2019). In *Daphnia*, it
144 has been demonstrated that the presence of ecocorona influences the uptake rates for
145 MNPs and their retention time in the gut (Nasser & Lynch, 2016). In exposed
146 organisms, the presence of an ecocorona has for instance been shown to alter MNP
147 effects on mobility/survival (e.g., Fadare et al., 2019; Schür et al., 2021), feeding
148 behavior (Nasser & Lynch, 2016) and molecular effects (e.g., Fadare et al., 2020).
149 However, deriving a general pattern for the direction of these mediating effects of
150 ecocorona formation is still challenging. One reason for this is that results in the
151 published literature are not entirely consistent: while ameliorating effects of ecocorona
152 formation were found for some MNPs (e.g., Amariei et al., 2022; Saavedra et al., 2019;
153 Wu et al., 2019), this was not the case for other MNPs (e.g., Pochelon et al., 2021; F.
154 Zhang et al., 2019), and also increased toxicities have been observed (Nasser &
155 Lynch, 2016). Another factor complicating generalized conclusions is the way
156 experiments are conducted. While ecocorona formation in some experiments was
157 induced by co-exposing organisms to MNP and dissolved organic matter (DOM) or
158 organic substances simultaneously (e.g., F. Zhang et al., 2019), other studies
159 investigated mediating effects of the ecocorona alone by incubating MNP in media
160 containing DOM prior to being transferred to the exposure medium (i.e., no additional
161 DOM in the exposure medium; e.g. Schür et al., 2021). It is thus unclear, whether it was
162 the ecocorona itself or the DOM in the media that led to the observed differences.

163 Furthermore, the type of DOM used in experiments affects the composition and
164 thickness of the formed ecocorona and may in turn influence observed outcomes
165 (Reilly et al., 2022; Schefer et al., 2023).

166 Meta-analyses are a tool for addressing exactly these kinds of questions, where the
167 presence, direction and size of effects are unclear (Field & Gillett, 2010). By
168 aggregating data from several studies in a quantitative way, meta-analyses aim to
169 derive effect size estimates with reduced bias and greater precision (lower uncertainty)
170 than estimates from single studies (Gurevitch et al., 2001; Harrison, 2011).

171 We performed a meta-analysis to answer the question of how the presence of DOM
172 and ecocorona alter the effects of MNP on *Daphnia* immobilization rates. Through a
173 systematic literature search of experimental studies, we compared the effects of MNP
174 with ecocorona/DOM to effects of the exact same particles without ecocorona/DOM.
175 Based on the gathered data, we discuss the strength of evidence regarding mediating
176 effects of ecocorona formation on MNP toxicity. In addition, we investigate whether the
177 mediating effects depend on the type of DOM and the type of experimental approach
178 used (either co-exposure with DOM or incubation prior to exposure).

179

180 3. Materials and Methods

181 3.1 Literature search

182 We conducted a literature search for studies that investigated effects of MNP on
183 immobilization rates (including mortality) in water fleas of the genus *Daphnia*. The aim
184 was to compile data from studies that met all of the following inclusion criteria: (1)
185 experimental research (excluding books and reviews) published in English, (2)
186 investigation of alterations of MNP effects due to ecocorona formation (i.e. studies
187 contained at least one MNP treatment with ecocorona and one treatment with the same
188 particle properties without ecocorona), (3) testing of water fleas of the genus *Daphnia*,
189 and (4) absence of additional stressors during exposure (e.g., chemicals). The final
190 search in December on Web of Science (WoS) and PubMed using the search string
191 “((micro* OR nano*) AND (plastic* OR particle*) OR microplastic OR nanoplastic) AND
192 *Daphnia* AND (eco-corona OR ecocorona OR biofilm OR humic acid OR DOC OR
193 DOM OR fulvic acid OR lake water OR protein corona OR protein-corona OR

194 proteincorona OR incub*)". In parallel, we searched for review articles addressing
195 effects of MNPs on *Daphnia* or freshwater organisms as additional sources of
196 literature. After removing duplicates, all titles and abstracts were screened. Studies
197 that clearly did not follow the inclusion criteria were removed. All remaining studies
198 were subjected to full text screening and only those studies that met all selection
199 criteria were kept for data extraction (see full text screening list in the supplementary
200 online material).

201

202 3.2 Data extraction

203 The extracted data consists of immobilization and mortality measurements, information
204 on added DOM, characteristics of the used MNP, information on the test organisms
205 and experimental parameters. As studies frequently did not distinguish between
206 immobilization (absence of movement after agitation; OECD, 2004) and mortality
207 (absence of heartbeat in addition to the absence of movement), we will refer to both
208 as immobilization.

209 We extracted immobilization rates (i.e., the number of immobile/mobile individuals or
210 the proportion of immobile individuals) for both the MNP treatment with
211 ecocorona/DOM and the control treatment without ecocorona/DOM. Whenever
212 possible, we extracted the rates directly from the text, data tables or raw data files
213 provided in the supplementary online material. In cases where the data was instead
214 presented in figures, the rates were extracted from the plots using the R package
215 *metaDigitise* (Pick et al., 2019). If daphnids were observed repeatedly over time,
216 immobilization rates were extracted for the latest time point (one of either 24, 48, 72 or
217 96 hours). If none of these time points were measured, we used the latest reported
218 time point in the study. Immobilization rates for all other time points were neglected.
219 Additionally, we extracted the number of replicates (i.e., the number of independent
220 test vessels) and the number of individuals per replicate (i.e., the number of daphnids
221 per test vessel). These numbers were used to calculate the number of immobile and
222 mobile individuals if immobilization rates were reported as proportions.

223 The pairing of treatments using identical experimental setups and particles that differed
224 only in the presence or absence of an ecocorona/DOM allowed us to directly control
225 for confounding by other MNP properties, different experimental parameters, and

226 characteristics of the test organisms. As the main explanatory variables, we thus only
227 noted (1) the type of DOM added in experiments (*DOM_type*; e.g., humic or fulvic acid,
228 different types of lake water, etc.), (2) whether DOM was added during exposure
229 (*DOM_conditioned*: no) or whether the particles had instead been incubated in DOM-
230 containing media prior to their use in experiments (i.e. they were removed from the
231 DOM-containing media prior to transfer to the exposure vessels without DOM;
232 *DOM_incubated*: yes).

233 For completeness and to enable extended use of the data in the future, we extracted
234 the following additional information from the studies if available: particle concentration
235 (in mg per ml and particles per ml), MNP properties including polymer type, particle
236 size (mean \pm standard deviation), particle shape (spherical, fragment, fiber), the
237 particles' chemical surface modification (e.g. carboxylation, amination) and surface
238 charge (either positive or negative); characteristics of the test organisms including
239 species, clone and age at the start of exposure; experimental conditions including
240 temperature, pH of the test medium, whether food was provided during exposure and
241 its concentration, and the concentration of DOM added during MNP incubation or
242 during exposure.

243

244 3.3 Statistical analysis

245 As a measure of effect size, we calculated log risk ratios (log (RR)) for immobilization.
246 A multiple mixed meta regression model without intercept was fit to the data including
247 the two factors *DOM_type* and *DOM_conditioned* (either yes or no) as moderators and
248 a random intercept for the publication identifier (*Publication ID*):

249

250 Moderator effects, i.e., the effects of the two fixed factors, were investigated by fitting
251 reduced models including only one of the moderators at a time and comparing these
252 reduced models to the full model through likelihood ratio tests. In addition, the variance
253 component attributed to *Publication ID* (sigma squared) was checked using profile
254 likelihood plots to ensure that the component was successfully identified, indicated by
255 the curve peaking at the maximum likelihood estimate. To see whether the two factors
256 in the model adequately accounted for the heterogeneity among data points, a test for

257 residual heterogeneity was conducted and the variance attributable to among-sample
258 heterogeneity rather than sampling variance (I^2) was calculated.

259 Based on the full model, an orchard plot was generated to visualize the average
260 marginal effect estimates for each combination of moderator levels. In contrast to forest
261 plots, orchard plots include individual effect size estimates, confidence intervals (CI)
262 and 95% prediction intervals.

263 To investigate potential publication bias visually, standard errors were plotted against
264 residuals of the full model in a funnel plot. Selective publishing in funnel plots is
265 indicated by a clear shift of data to higher effect sizes in studies with high standard
266 errors, suggesting that studies with low power are preferentially published when
267 observed effects are large.

268 All analyses were done in R version 4.3.1 (R Core Team 2023). The *metafor* package
269 version 4.2 (Viechtbauer, 2010) was used for effect size calculation, fitting meta
270 regression models and investigating publication bias. The *orchaRd 2.0* package
271 (Nakagawa et al., 2023) was used for creating the orchard plot and calculating I^2 .

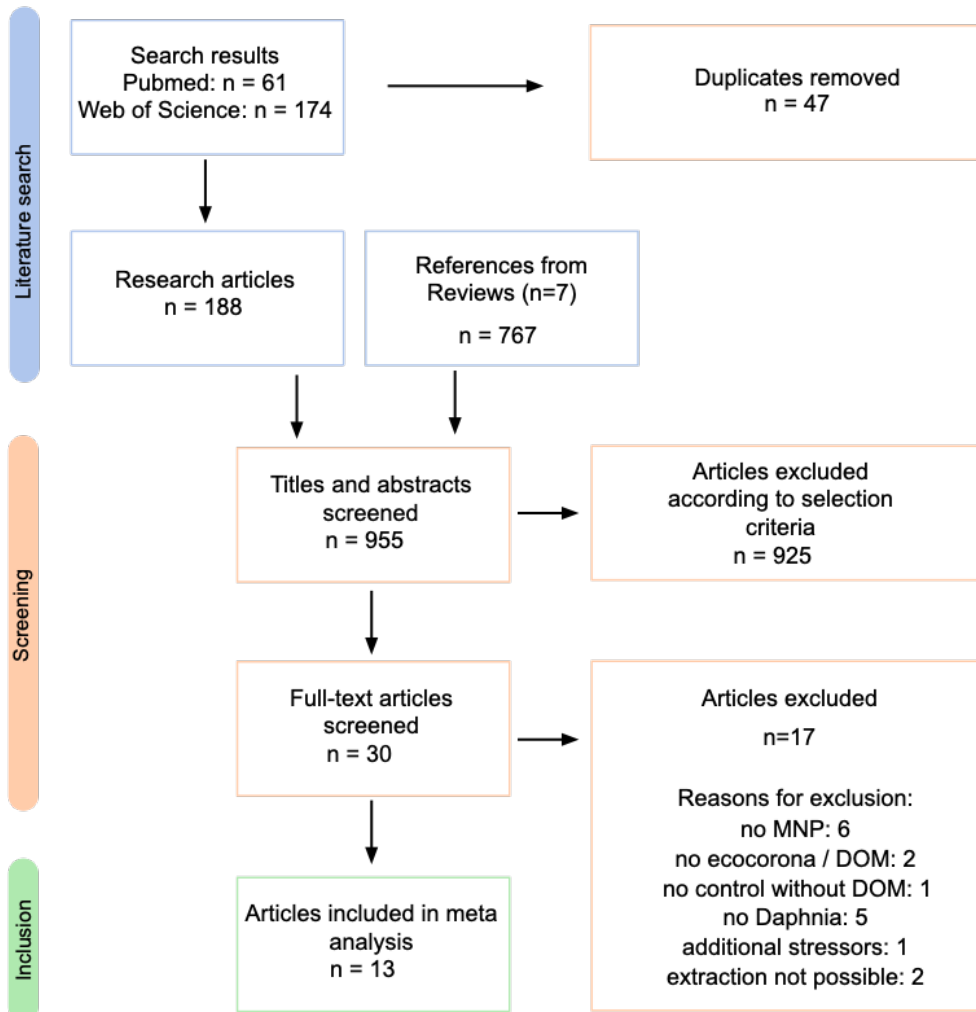
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273 4. Results

274 4.1 Literature search and data extraction

275 The literature search resulted in 955 publications, of which 925 failed to meet at least
276 one of the selection criteria. After full text screening of the remaining 30 publications,
277 17 studies were excluded either because they did not fulfill all the selection criteria or
278 data extraction was not possible. The remaining 13 publications were used for data
279 extraction. In total, we extracted 305 data points (Fig. 1). We grouped the types of
280 DOM used in the studies into seven main categories: metabolites excreted from
281 *Daphnia* (2 studies, 21 data points), humic acid (4 studies, 43 data points), fulvic acid
282 (2 studies, 24 data points), commercially bought natural organic matter (1 study, 6 data
283 points), stream water (3 studies, 34 data points), lake water (3 studies, 128 data points)
284 and wastewater (4 studies, 49 data points). In seven of the studies, the MNPs were
285 incubated in the respective DOM type prior to exposure (105 data points). In the other
286 six studies, DOM was added during exposure or exposure was conducted in water
287 sampled from natural environments (200 data points). All included studies tested the

288 species *Daphnia magna* (for a full overview of experimental parameters covered in the
 289 studies see raw data in the supplementary online material).



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291

292 **Fig. 1:** PRISMA flow diagram illustrating the results from the systematic literature search and stepwise
 293 exclusion of studies not fulfilling defined selection criteria. MNP: micro- and nanoplastic particles.

294 DOM: dissolved organic matter.

295

296 4.2 Effects of DOM type and conditioning on immobilization rates

297 In the full model, moderators had a significant effect on immobilization risk (test of
 298 moderators: $Q_M = 17.33$, $df = 8$, $p = 0.023$) indicating that at least one of them
 299 accounted for differences in effects across samples. A comparison of full and reduced
 300 models showed that DOM type was more important for the model fit (comparison of
 301 the full versus the reduced model without factor *DOM_type*: $LR = 12.47$, $p = 0.052$)

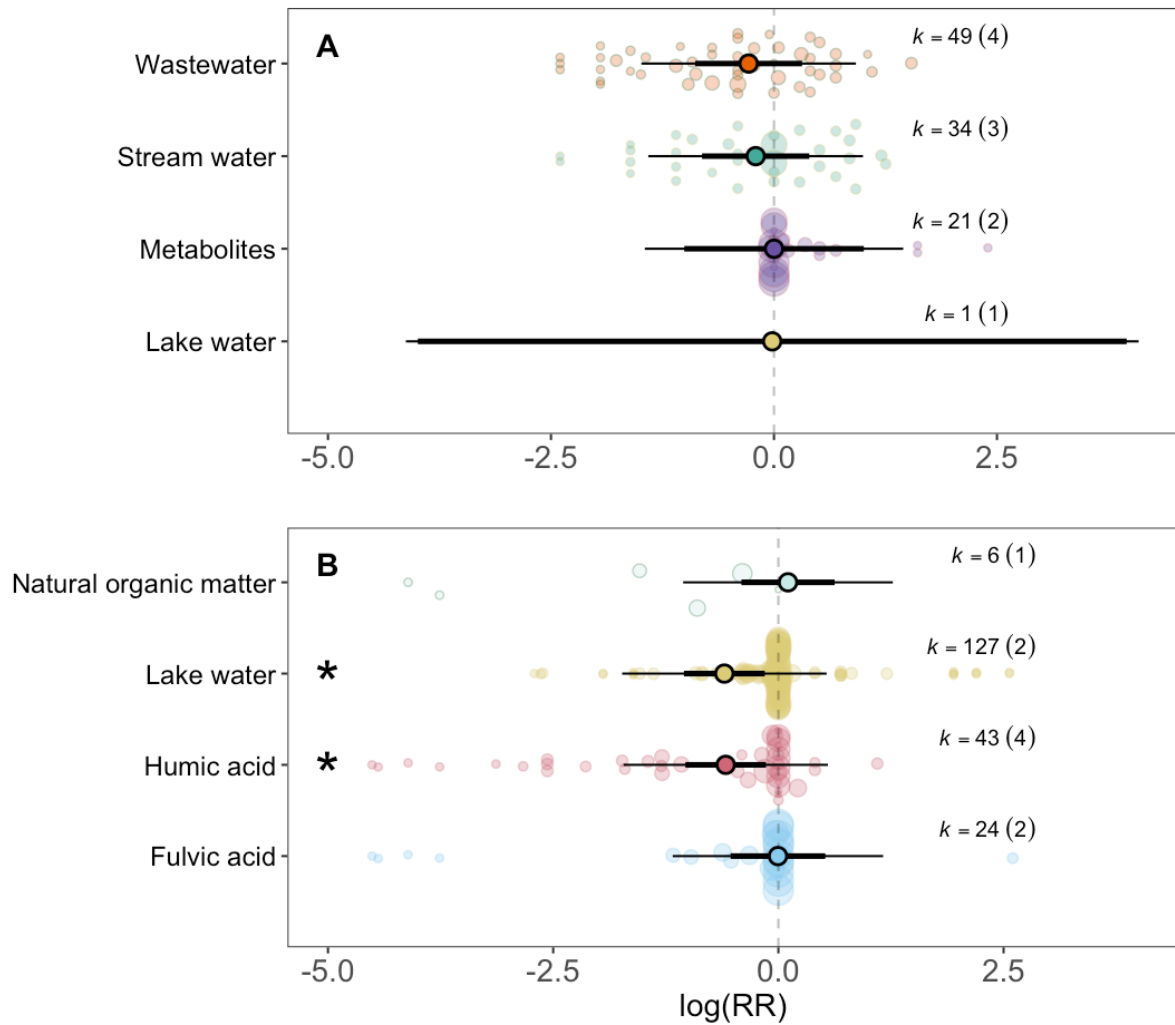
302 than whether the particles had been conditioned prior to their use in exposure
303 experiments (comparison of the full versus the reduced model without factor
304 *DOM_conditioned*: LR = 0.08, p = 0.78). A reduction in immobilization risk was
305 observed when DOM was added in the form of humic acid (log(RR): -0.58 (CI: -1.03, -
306 0.14), RR: 0.56 (CI: 0.36, 0.87), z = -2.56, p = 0.01) or lake water during exposure
307 (log(RR): -0.60 (CI: -1.05, -0.15), RR: 0.55 (CI: 0.35, 0.86), z = -2.62, p = 0.009), and
308 to a lesser degree when MNP were incubated in wastewater (log(RR): -0.28 (CI: -0.88,
309 0.32), RR: 0.76 (CI: 0.41, 1.37), z = -0.93, p = 0.35) or stream water (log(RR): -0.21
310 (CI: -0.81, 0.40), RR: 0.81 (CI: 0.44, 1.49), z = -0.67, p = 0.50) prior to exposure (Fig.
311 2). Changes in immobilization risks due to DOM were below 15% in all other
312 treatments.

313 In general, we found high variance in the data in all groups. In addition, while DOM
314 type and experimental approach (*DOM_conditioned*) explained part of the variance in
315 the data, the model left significant residual heterogeneity ($Q_M = 680.25$, df = 297, p <
316 0.0001) indicating that other factors not included in our analysis additionally contribute
317 to the differences. In accordance with this, 97% of the variance among data points can
318 be attributed to sample heterogeneity ($I^2 = 0.97$) and only 3% are estimated to result
319 from sampling variance.

320

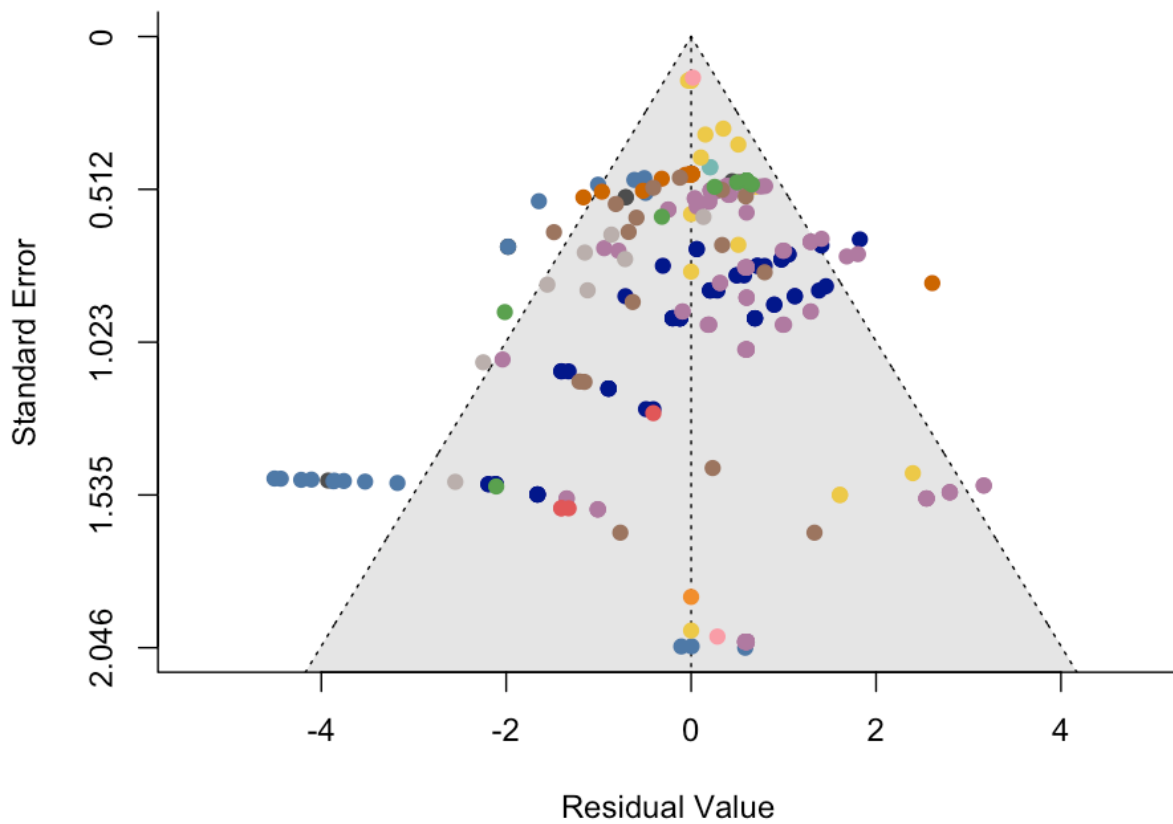
321 4.3 Publication bias

322 A visual inspection of the funnel plot (Fig. 3) did not show severe deviations from
323 symmetry, except for a small gap in data points on the right for medium powered
324 studies (see gap on the right at medium standard errors in Fig. 3).



326

327 **Fig. 2:** Mean treatment effects of different dissolved organic matter (DOM) types from the selected
 328 studies. A: effects shown for micro- and nanoplastic particles (MNP) conditioned in DOM containing
 329 medium; B: effects of MNP with DOM added to the medium during exposure; log [RR]: Log risk ratio;
 330 thicker black lines show 95% confidence intervals; narrow lines show prediction intervals; asterisks
 331 indicate significant moderation of MNP effects ($p < 0.05$). Point sizes reflect inverse standard errors.



333

334 **Fig. 3:** Standard errors of data points from the studies included in the meta-analysis plotted against
 335 residual log (RR). Different colors represent data points extracted from different publications and the
 336 grey shaded area represents the 95 % pseudo confidence interval.

337

338 5. Discussion

339 Our meta-analysis shows that immobilization of *Daphnia* by MNPs can be alleviated
 340 by DOM. While it has been argued that the reduction in negative effects in the presence
 341 of DOM might be attributed to the formation of an ecocorona on the MNP surface alone
 342 (e.g., Fadare et al., 2020; Junaid & Wang, 2021), our data indicate that DOM present
 343 in the media may contribute additionally to the observed mitigating effects.
 344 Furthermore, the compiled data suggest that moderating effects of DOM depend on
 345 the type of DOM used.

346 DOM is known to alleviate effects of various pollutants in *Daphnia* spp. For example,
 347 humic acid has been shown to attenuate negative effects of soluble substances (e.g.,

348 Oris et al., 1990; Paulauskis & Winner, 1988; Y. Zhang et al., 2019), pesticides (e.g.,
349 Day, 1991) and various particulate pollutants (e.g., Y. Zhang et al., 2019), and natural
350 organic matter has been shown to reduce the toxicity of perfluorooctane sulfonate
351 (PFOS; Kovacevic et al., 2019) and heavy metals (e.g., De Schamphelaere et al.,
352 2004; Penttinen et al., 1998), among others. In addition, beneficial impacts of DOM on
353 effects imposed by MNP and other pollutants have also been demonstrated in several
354 other aquatic organisms. Saavedra et al. (2019) for example tested the toxicity of
355 MNPs to the rotifer *Brachionus calyciflorus* and larvae of *Themnocephalus platyurus*
356 and found lowered effects for ecocorona-coated compared to pristine MNP. Similarly,
357 ameliorating effects of DOM have, among others, been found for MNP-induced
358 oxidative stress responses (ROS production) in algae and fish (Liu et al., 2019;
359 Natarajan et al., 2021), copper-induced mortality in freshwater mussels (Gillis et al.,
360 2008), and for pesticide-induced mortality in the freshwater mysid shrimp
361 *Americamysis bahia* (Mézin & Hale, 2004)

362 One mechanism by which DOM can reduce toxic effects is its ability to bind pollutants,
363 alter the physico-chemical properties of particles and thus decrease their bioavailability
364 (Kukkonen & Oikari, 1991). For instance, humic acid and, to a lesser extent, fulvic acid
365 bind hydrophobic organic pollutants including pesticides (Chianese et al., 2020; De
366 Paolis & Kukkonen, 1997; Landrum et al., 1985), and DOM binds pesticides and metal
367 ions (Aiken et al., 2011; He et al., 2020; Reuter & Perdue, 1977). In contact with
368 particulate pollutants, DOM leads to ecocorona formation (e.g., Elagami et al., 2022;
369 Rummel et al., 2017) altering the particles' surface characteristics (Natarajan et al.,
370 2021; Shi et al., 2023; Witzmann et al., 2022) and changing interactions with tissues
371 and cells (Ramsperger et al., 2020). Changes in physico-chemical properties also lead
372 to altered colloidal interactions (Witzmann et al., 2022) and altered aggregation in the
373 test media (Aiken et al., 2011; Meng et al., 2023; M. Yan et al., 2021). Particle
374 aggregation in turn affects the particles' transport behavior, can lead to increased
375 sinking velocities and consequently lower the availability of particulate pollutants in the
376 water column (Elagami et al., 2022; Karakaş et al., 2009; Petosa et al., 2010; Y. Yan et
377 al., 2021). In addition, particle aggregation itself can alter MNP toxicity (Albanese &
378 Chan, 2011; Meng et al., 2023; Wu et al., 2019). Furthermore, DOM attached to the
379 particles may add nutritional value to the particles and thus partly reduce food dilution
380 effects (Arruda et al., 1983). Nevertheless, all these effects of DOM on particle
381 behavior and properties cannot explain sufficiently why moderating effects in our

382 dataset were stronger in co-exposure as compared to MNP incubation setups, as these
383 effects should show up in both setups similarly.

384 In general, stressed *Daphnia* are more sensitive towards additional stressors (Serra et
385 al., 2020; Yin et al., 2011) while beneficial environments help daphnids become more
386 resilient (Lye Koh et al., 1997; Vandenbrouck et al., 2011; for a conceptual discussion
387 of stress addition, see Liess et al., 2016). A potential alternative explanation for the
388 strong attenuating effects of humic acid and lake water in co-exposure experiments
389 may be that DOM in the media generally contributes to the well-being of the daphnids,
390 thus making them more resilient to stressors. Supplementary DOM for instance may
391 serve as a food and nutrient source for phytoplankton, indirectly leading to better food
392 supply for the daphnids (Thomas, 1997) or increase food supply through the microbial
393 loop when algal food becomes limited (e.g., Hiltunen et al., 2017; McMeans et al.,
394 2015). Second, similar to other organisms, daphnids may benefit directly from DOM in
395 the test media via increased intestinal health, increased reproduction or increased
396 growth induced by mild chemical stress responses (Gao et al., 2017; Steinberg et al.,
397 2009). In contrast to these findings however, other studies have demonstrated adverse
398 effects of natural organic matter (Wenzel et al., 2021) and humic acid (Euent et al.,
399 2008; Saebelfeld et al., 2017) on *Daphnia* and on other invertebrate freshwater species
400 (e.g., Timofeyev et al., 2006), indicating that the processes and mechanisms in natural
401 waters are likely more complex and not understood well enough yet.

402 Among the types of DOM used in the screened literature, humic acid and lake water
403 added during exposure had the strongest mitigating effects, decreasing the risk of
404 immobilization caused by MNP by almost 50% (RR = 0.56 and 0.55, respectively).
405 Whether these effect sizes are comparable to effects in natural environments depends
406 among other factors on how realistic the concentrations of DOM applied in experiments
407 were. In the studies included in our meta-analysis that reported DOM concentrations,
408 test concentrations ranged from 1 to 50 mg L⁻¹ (see raw data in the supplementary
409 online material). Similar ranges have been reported for natural aquatic systems,
410 spanning for instance from 0.1 to 322 mg L⁻¹ in a dataset of measurements of dissolved
411 organic carbon (DOC) from 7,500 lakes (Sobek et al., 2007; see also Thomas, 1997).
412 It is therefore likely that attenuating effects of DOM on MNP toxicity can occur in a
413 similar way in natural habitats.

414 Although the compiled dataset indicates that the mere presence of an ecocorona has
415 lower moderating effects on immobilization risks than co-exposure with DOM, and that
416 effect sizes differ between different DOM types, the dataset also has some important
417 limitations. The types of DOM tested in conditioning experiments were different from
418 the types tested in co-exposure experiments (except for one data point from an
419 experiment using lake water-conditioned MNP). Due to this limited overlap, it is difficult
420 to disentangle the effects of DOM type and experimental approach at this point.
421 However, the observed patterns can serve as valuable hypotheses that can be easily
422 validated (or disproved) in experiments or when more studies become available in the
423 future.

424 Publication bias can lead to wrong effect size estimates derived from meta-analyses.
425 Bias arises when some effects sizes are published selectively, e.g., when only
426 significant outcomes are published or when confirmatory results are preferentially
427 published (Sterling, 1959). A second factor that can lead to wrong effect size estimates
428 is study (or sample) heterogeneity (Higgins & Thompson, 2002; Kenny & Judd, 2019).
429 Significant heterogeneity indicates that the variance among data points cannot be
430 sufficiently explained by sampling variance, but instead likely results from measured
431 effects not being derived from a true common effect. The funnel plot from our meta-
432 analysis shows a slight gap of data points on the side of increased immobilization risk
433 in the presence of DOM and the full model showed significant residual heterogeneity.
434 A likely reason for both these patterns is that the effect sizes in our dataset are
435 moderated by additional factors not accounted for in our analysis (Ioannidis &
436 Trikalinos, 2007). For example, it is possible that the effect of DOM on immobilization
437 caused by MNP is further moderated by experimental temperature, food availability
438 during exposure, different MNP concentrations, different concentrations of DOM or
439 other experimental parameters and MNP properties. Although we accounted for
440 confounding by pairing measurements from treatments that differed solely in DOM
441 presence/absence while keeping all other parameters the same, we cannot rule out
442 interaction effects. For example, it might be possible that the strength of the mediating
443 effect of DOM on MNP toxicity differs for different polymer types, experimental
444 temperatures, or any other parameter.

445 In general, we think that further research is needed to disentangle the effects of
446 ecocorona formation and the presence of DOM in the media on MNP toxicity in

447 *Daphnia*. This includes addressing in particular the mechanisms that lead to observed
448 differences among different approaches and DOM types. Although *Daphnia* is among
449 the most frequently used organisms in ecotoxicological on MNP effects (Brehm et al.,
450 2023; Hampton et al., 2022), we found only 13 studies (published until December 2022)
451 that met our selection criteria and allowed for the extraction of effect size data. Further
452 experiments investigating the mechanisms of MNP toxicity are thus needed to enhance
453 our understanding of MNP effects in the presence of DOM in natural environments.

454

455 Conclusions

456 In the present meta-analysis, we synthesized data from 13 studies that investigated
457 the effects of ecocorona formation and DOM on MNP induced immobilization risk in
458 *Daphnia* spp. We showed that the mere presence of an ecocorona on the particles can
459 moderate MNP toxicity, but the presence of DOM in the test media during exposure
460 appears to be another important predictor for the observed attenuation of negative
461 outcomes. Based on our results and evidence from the literature on other stressors,
462 we hypothesize that DOM and in particular humic acid mitigates negative effects of
463 MNP by either (1) reducing bioavailability or (2) making daphnids more resilient to
464 stressors in general. Additional experiments are needed to challenge these hypotheses
465 and disentangle the effects of ecocorona formation and the presence of DOM in the
466 media, and to understand how effects of DOM on particle behavior in the medium
467 translate into reduced effects on an organismic level. Such experiments could for
468 example directly compare the impact of ecocorona formation during the incubation of
469 MNP with the impact of adding the same type of DOM to the media during exposure.
470 Comparisons could also be made regarding the attenuating effect of DOM on negative
471 effects of different additional stressors including chemical or particulate pollutants, as
472 well as other stressors such as heat stress.

473

474 Author contributions

475 Conceptualization - MMM. Data extraction - SS, EG, PK, EN, AR, SW, FB, MMM
476 Data validation - ALAV. Statistical analysis - PK, ALAV, MMM. Visualization - EG,

477 ALAV, MS, MMM. Writing - original draft - SS, EG, PK, EN, AR, SW, FB. Writing -
478 review & editing - SS, EG, EN, AR, SW, ALAV, MS, MMM. All authors approved for
479 the final version of the manuscript.

480

481 Open data statement

482 All raw data are provided in the supplemental material. All data and code are also
483 available on github (XXto be added upon acceptanceXX) and Zenodo (XXto be
484 added upon acceptanceXX).

485

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491

492 Conflicts of interest

493 The authors declare that they have no conflicts of interest.

494

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