

1 **Detecting context-dependence in the expression of life history tradeoffs**

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3 Louis Bliard ^{1*}, Jordan S. Martin ^{2*}, Maria Paniw ^{1,3}, Daniel T. Blumstein ^{4,5}, Julien G.A. Martin ⁶,
4 Josephine M. Pemberton ⁷, Daniel H. Nussey ⁷, Dylan Z. Childs ⁸, Arpat Ozgul ¹

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6 1 Department of Evolutionary Biology and Environmental Studies, Zurich University, Zurich,
7 Switzerland

8 2 Institute of Evolutionary Medicine, Zurich University, Zurich, Switzerland

9 3 Department of Conservation Biology, Estación Biológica de Doñana (EBD-CSIC), Seville, Spain

10 4 Department of Ecology and Evolutionary Biology, University of California Los Angeles, Los
11 Angeles, California, USA

12 5 The Rocky Mountain Biological Laboratory, Crested Butte, Colorado, USA

13 6 Department of Biology, University of Ottawa, Ottawa, Ontario, Canada

14 7 Institute of Ecology and Evolution, University of Edinburgh, Edinburgh, UK

15 8 Department of Animal and Plant Sciences, University of Sheffield, Sheffield, UK

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17 * Shared first authorship

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Abstract

Life history tradeoffs are one of the central tenets of evolutionary demography. Tradeoffs, depicting negative covariances between individuals' life history traits, can arise from genetic constraints, or from a finite amount of resources that each individual has to allocate in a zero-sum game between somatic and reproductive functions. While theory predicts that tradeoffs are ubiquitous, empirical studies have often failed to detect such negative covariances in wild populations. One way to improve the detection of tradeoffs is by accounting for the environmental context, as tradeoff expression may depend on environmental conditions. However, current methodologies usually search for fixed covariances between traits, thereby ignoring their context dependence. Here, we present a hierarchical multivariate 'covariance reaction norm' model, adapted from Martin (2023), to help detect context dependence in the expression of life-history tradeoffs using demographic data. The method allows continuous variation in the phenotypic correlation between traits. We validate the model on simulated data for both intraindividual and intergenerational tradeoffs. We then apply it to empirical datasets of yellow-bellied marmots (*Marmota flaviventer*) and Soay sheep (*Ovis aries*) as a proof-of-concept showing that new insights can be gained by applying our methodology, such as detecting tradeoffs only in specific environments. We discuss its potential for application to many of the existing long-term demographic datasets and how it could improve our understanding of tradeoff expression in particular, and life history theory in general.

45

Introduction

46 Demographic tradeoffs, which are characterized as negative covariances between fitness
47 components such as somatic or reproductive traits, are central to life history theory (Stearns,
48 1989), and are thought to constrain and organize much of the life history diversity that exists
49 (Bielby et al., 2007; Healy et al., 2019; Salguero-Gómez et al., 2016; Stearns, 1984). They originate
50 from the basic fact that the total amount of resources or energy acquired by any one individual
51 is limited, and has to be shared among several of the individual's fitness-related traits. In such a
52 zero-sum game and in the absence of change in the total amount of resources acquired, any
53 increase in the allocation of resources towards a specific fitness component will have to be at the
54 expense of another fitness component. If tradeoffs did not exist, selection would maximize all
55 fitness-related traits simultaneously and would lead to the impossible "darwinian demons" (Law,
56 1979). Therefore, tradeoffs should be faced by all organisms and are, in theory, ubiquitous
57 (Stearns, 1989, 1992; Williams, 1966). They can come in several forms (Stearns, 1989), being
58 either intraindividual (traits involved relate to the fitness of the same individual) or
59 intergenerational (traits involved relate to the fitness of a parent-offspring pair; e.g., offspring
60 quantity-quality tradeoff). Despite their expected universality and being sought-after by
61 evolutionary ecologists and biodemographers alike, life-history tradeoffs have been surprisingly
62 hard to detect in wild populations (Chang et al., 2023; Metcalf, 2016; van Noordwijk & de Jong,
63 1986), with successful probes too often confined to experimental approaches.

64 Several reasons could explain why tradeoffs are hard to detect in wild populations. First,
65 we often expect traits to covary in a simple bivariate manner following the Y-model of resource
66 allocation, where any resources diverted from a trait will be allocated to the other one (de Jong

67 & van Noordwijk, 1992). Thus, while we are often analyzing a single pair of traits at a time,
68 tradeoff structures are often more complex. For instance, many more than two traits are likely
69 to be involved in the resource allocation process (Cressler et al., 2017; de Jong, 1993; Pease &
70 Bull, 1988), sometimes leading to complex hierarchical allocation trees, potentially resulting in
71 some pairs of traits not covarying negatively (Gascoigne et al., 2022). Second, life history traits
72 can covary at different levels. While tradeoffs result from individuals' resource allocation
73 processes, biodemographers often study tradeoffs as the temporal correlations among
74 demographic rates at the population level (Compagnoni et al., 2016; Fay et al., 2020; Fay, Hamel,
75 et al., 2022; van Tienderen, 1995). Tradeoffs can occasionally scale up to cause negative temporal
76 covariances at the population level (van Tienderen, 1995), but in most cases these covariances
77 are the results of environmental stochasticity and demographic reaction norms to shared
78 ecological drivers (Fay, Hamel, et al., 2022; Knops et al., 2007; Paniw et al., 2020). Third, even
79 though tradeoffs might be present, individual heterogeneity can mask their presence among
80 individuals. This specific ecological version of Simpson's paradox (Simpson, 1951) has been
81 demonstrated by van Noordwijk and de Jong (1986): when the among-individual variance in
82 resource acquisition is greater than the among-individual variance in resource allocation, the
83 tradeoff is not expressed among individuals — even though it is theoretically present within
84 individuals. In addition, expression of a tradeoff among individuals can also be influenced if the
85 allocation and acquisition processes are not independent (Descamps et al., 2016; Fischer et al.,
86 2009; Robinson & Beckerman, 2013). Altogether, this makes the detection of tradeoffs in wild
87 populations difficult.

88 How much individuals vary in acquisition and allocation of resources determines if a
89 tradeoff is detected among individuals (Metcalf, 2016; Reznick et al., 2000; van Noordwijk & de
90 Jong, 1986). Part of this variance might be fixed, stemming from genetic, developmental, or
91 consistent behavioral differences that constrain how much resources are acquired and allocated
92 to somatic vs. reproductive functions (Réale et al., 2007; Wilson & Nussey, 2010). The remaining
93 variance is likely to be plastic (Spigler & Woodard, 2019), where acquisition vs. allocation likely
94 depends on the environmental context (Cohen et al., 2020; Sgrò & Hoffmann, 2004; Stearns et
95 al., 1991). For instance, in several species, no tradeoffs were found among captive animals fed
96 *ad libitum* (Kengeri et al., 2013; Landes et al., 2019; Ricklefs & Cadena, 2007). Similarly, controlled
97 laboratory experiments on several species have shown that tradeoffs detection and strength
98 were dependent on resource abundance (Gebhardt & Stearns, 1988; Messina & Fry, 2003;
99 Messina & Slade, 1999; Spigler & Woodard, 2019). However, despite evidence that tradeoff
100 expression depends on the environmental context, statistical methods to detect this context
101 dependence in wild populations have, to date, rarely been applied.

102 Multivariate models are commonly employed to detect tradeoffs in wild populations
103 (Cam et al. 2002, 2013; Hamel et al. 2018; Paterson et al. 2018; Fay, Hamel, et al. 2022). In
104 quantitative genetics, such models allow for the simultaneous analysis of multiple dependent
105 variables like fecundity, growth, and survival (Kruuk et al. 2008; Wilson et al. 2010). These
106 variables each have their own predictors, and the models estimate the correlated residual
107 variances unaccounted for by the primary predictors. These models can be used to study residual
108 correlations between traits at different levels, such as among-year correlation and among-
109 individual correlation. For example, after accounting for primary predictors, such models quantify

110 whether years with high survival in a population are also years with high recruitment; or whether
111 individuals with higher fecundity have lower or higher growth rates. However, these correlations
112 among residual variances are estimated as fixed. Estimating fixed correlations might not
113 necessarily be problematic in the case of experimental work, in which environmental conditions
114 can be held constant within each treatment. However, wild populations are unlikely to
115 experience fixed conditions, as the environmental context will vary in a continuous fashion,
116 hence influencing the expression of tradeoffs. Therefore, there is a need to analyse and predict
117 continuous variation of phenotypic correlations.

118 Here, we repurpose a hierarchical multivariate ‘covariance reaction norm’ (hereafter
119 CRN) model recently developed by Martin (2023), which allows the incorporation of continuous
120 predictors directly on the covariance matrix, for application to sampling designs typical in
121 population ecology, enabling the study of the context-dependent expression of tradeoffs. As a
122 proof-of-concept, we first validate this model on two simulated datasets, respectively focusing
123 on an intergenerational tradeoff and an intraindividual tradeoff. We then apply our model on
124 two empirical datasets of wild populations of yellow-bellied marmots *Marmota flaviventer* and
125 Soay sheep *Ovis aries*. Prior studies have explored tradeoffs between vital rates in both species
126 (Kroeger et al., 2020; Tavecchia et al., 2005). For instance, in yellow-bellied marmots, a quality-
127 quantity tradeoff in offspring has been observed for older mothers. In Soay sheep, the costs of
128 reproduction have been particularly evident for breeding ewes in high-density populations or
129 following harsh winters. However, the environmental context-dependence of these tradeoffs has
130 yet to be studied explicitly. In the marmots, which inhabit high-altitude, highly seasonal
131 environments, and the sheep, which face severe winter storms and fluctuating population

132 densities, we hypothesize that tradeoffs are more likely to manifest under unfavorable ecological
133 conditions (Cohen et al., 2020; Sgrò & Hoffmann, 2004).

134 **Methods**

135 The model

136 In this study, we employ a newly introduced CRN model (Martin, 2023), which has been
137 developed as a quantitative genetic model to predict continuous changes in trait associations
138 when either genetic data or repeated individual measurements are available for all phenotypes
139 of interest. A key assumption of multivariate models thus far has been that phenotypic
140 correlations caused by tradeoffs are fixed through time or space (Cam et al., 2002; Hamel et al.,
141 2018). The CRN approach provides a solution to this general challenge, by allowing for phenotypic
142 covariances to vary in response to variation in the environment, that is, estimating under which
143 conditions among-individual variance in resources allocation is larger than among-individual
144 variance in acquisition (van Noordwijk & de Jong, 1986). In the present study, we extend
145 application of this general CRN approach to the detection of context-dependent tradeoffs (here
146 defined as among-individual correlations even though both are not always equivalent) between
147 life history traits, with special consideration to sampling conditions typical of long-term field
148 research in population ecology. Specifically, we examine the use of bivariate CRN models to test
149 for the presence of phenotypic tradeoffs when repeated individual measurements are lacking in
150 a given environmental context (e.g., during a specific sampling event such as a breeding season
151 or a year). These are typical situations in field research that motivate further development of the
152 quantitative genetic models proposed by Martin (2023).

153 Consider a CRN model investigating how environmental contexts C and individual factors
 154 affect the phenotypic means of $\beta_{\mu 1}$ and $\beta_{\mu 2}$ and among-individual correlations β_r between two
 155 Gaussian life history trait measures \mathbf{z}_1 and \mathbf{z}_2 with repeated individual measurements in each
 156 environmental context. \mathbf{X}_1 and \mathbf{X}_2 are $N \times P$ matrices of N measurements of P predictors. Note
 157 that in all the following models presented, measurements of the same individuals observed in
 158 different contexts are considered independent (see supplementary materials Section S1). We
 159 begin by focusing on linear models to simplify notation and aid comprehension, with generalized
 160 models for non-Gaussian distributions discussed further below. Following Martin (2023) in the
 161 absence of genetic data, our bivariate phenotypic model is given by

$$\mathbf{z}_1 = \mathbf{X}_1 \boldsymbol{\beta}_{\mu 1} + \mathbf{W} \boldsymbol{\alpha}_{1(C)} + \boldsymbol{\epsilon}_{1(C)} \quad (1.1)$$

$$\mathbf{z}_2 = \mathbf{X}_2 \boldsymbol{\beta}_{\mu 2} + \mathbf{W} \boldsymbol{\alpha}_{2(C)} + \boldsymbol{\epsilon}_{2(C)}$$

$$[\boldsymbol{\alpha}_{1(C)}, \boldsymbol{\alpha}_{2(C)}] \sim N(\mathbf{0}, \mathbf{P}_{(C)})$$

$$[\boldsymbol{\epsilon}_1, \boldsymbol{\epsilon}_2] \sim N(\mathbf{0}, \boldsymbol{\Sigma}_{(C)})$$

162

163 Trait values are expressed as a function of the average effects $\beta_{\mu 1}$ and $\beta_{\mu 2}$ of \mathbf{X}_1 and \mathbf{X}_2 on each
 164 phenotype, as well as among-individual effects $\alpha_{1(C)}$ and $\alpha_{2(C)}$ that are repeatable across
 165 measurements and within-individual effects $\epsilon_{1(C)}$ and $\epsilon_{2(C)}$ that are variable across
 166 measurements. The model matrix \mathbf{W} (an $N \times J$ matrix for J subjects) structures the among-
 167 individual effects $\alpha_{(C)}$ across repeated measurements. (Co)variances between independent
 168 among- and within-individual effects are respectively described by \mathbf{P} and $\boldsymbol{\Sigma}$ covariance matrices.
 169 To detect context-dependent tradeoff expression, we use environmental information in \mathbf{X}_3 (an C

170 x P matrix of C environmental contexts of P predictors) to predict the among-individual trait
 171 covariance matrix $\mathbf{P}_{(C)}$.

$$\mathbf{P}_{(C)} = \begin{bmatrix} \sigma_{\alpha_1(C)}^2 & r_{\alpha(C)} \sigma_{\alpha_1(C)} \sigma_{\alpha_2(C)} \\ r_{\alpha(C)} \sigma_{\alpha_2(C)} \sigma_{\alpha_1(C)} & \sigma_{\alpha_2(C)}^2 \end{bmatrix} \quad (1.2)$$

$$\text{atanh}(r_{\alpha(C)}) = \mathbf{X}_3 \boldsymbol{\beta}_r$$

172
 173 where the inverse hyperbolic tangent function $\text{atanh}(r) = \text{logit}([r + 1]/2)/2$ is used as a link
 174 function to model additive environmental effects β_r on the logit scale while retaining the [-1,1]
 175 scaling of the correlation coefficient r . This is akin to a logistic regression with bounds in [-1,1]
 176 instead of [0,1]. The same approach can be taken to describe changes in within-individual
 177 variation across environmental contexts.

$$\boldsymbol{\Sigma}_{(C)} = \begin{bmatrix} \sigma_{\epsilon_1(C)}^2 & r_{\epsilon(C)} \sigma_{\epsilon_1(C)} \sigma_{\epsilon_2(C)} \\ r_{\epsilon(C)} \sigma_{\epsilon_2(C)} \sigma_{\epsilon_1(C)} & \sigma_{\epsilon_2(C)}^2 \end{bmatrix} \quad (1.3)$$

$$\text{atanh}(r_{\epsilon(C)}) = \mathbf{X}_3 \boldsymbol{\beta}_{r_\epsilon}$$

178
 179 Direct prediction of the transformed correlation coefficient is useful because we are
 180 principally interested in $r_{(C)}$ as an indicator of putative within- or among-individual tradeoffs,
 181 rather than the covariance $P_{1,2(C)} = r_{(C)} \sigma_1 \sigma_2$ per se. Changes in the scale $\sigma_1 \sigma_2$ of life history trait
 182 variation may occur independently of changes in positive or negative trait association among
 183 individuals, but these effects will be confounded together in the covariance $P_{1,2(C)}$. In contrast,
 184 the correlation coefficient $r_{(C)}$ is standardized relative to the scale of each phenotype, providing
 185 a more robust quantity for directly predicting and comparing estimates of life history tradeoffs
 186 across phenotypes and species. Our model also assumes that phenotypic variances can vary
 187 across environmental contexts, but no predictions are made on this variation. Greater plasticity
 188 is instead expected in the strength of tradeoff expression caused by fluctuating environmental

189 factors (e.g., environmental harshness, resource availability, local predator density). See Martin
 190 (2023) for further details on relaxing these assumptions to model environmental effects on
 191 among- and within-individual variances.

192 ***Non-repeated measures***

193 Estimating **Eq 1** with empirical data requires multiple measurements of the same subjects
 194 to effectively partition trait correlations due to sources of among- $\mathbf{P}_{(C)}$ and within-individual $\mathbf{\Sigma}_{(C)}$
 195 phenotypic variation, relative to a given window of sampling (i.e., a given environmental context
 196 C). Repeated individual measurements are often inconsistent or unavailable in a given
 197 environmental context (e.g., a single fecundity measurement for individuals in a given year) in
 198 long-term field studies, which otherwise provide invaluable datasets for investigating context-
 199 specific tradeoffs in the wild. Fortunately, we can still take advantage of long-term environmental
 200 variation in such studies to detect variation in tradeoff expression without repeated
 201 measurements in a given environmental context. This requires simplifying the CRN model to
 202 predict observation-level phenotypic associations across environmental contexts.

$$\mathbf{z}_1 = \mathbf{X}_1 \boldsymbol{\beta}_{\mu 1} + \mathbf{o}_{1(C)} \quad (2)$$

$$\mathbf{z}_2 = \mathbf{X}_2 \boldsymbol{\beta}_{\mu 2} + \mathbf{o}_{2(C)}$$

$$[\mathbf{o}_{1(C)}, \mathbf{o}_{2(C)}] \sim \mathbf{N}(\mathbf{0}, \mathbf{P}_{\mathbf{o}(C)})$$

$$\mathbf{P}_{\mathbf{o}(C)} = \begin{bmatrix} \sigma_{o_1(C)}^2 & r_{o(C)} \sigma_{o_1(C)} \sigma_{o_2(C)} \\ r_{o(C)} \sigma_{o_2(C)} \sigma_{o_1(C)} & \sigma_{o_2(C)}^2 \end{bmatrix}$$

$$\text{atanh}(r_{o(C)}) = \mathbf{X}_3 \boldsymbol{\beta}_r$$

203

204 Here, the lack of repeated measurements mean that we cannot decompose the variance
 205 between among- and within-individual variation. Therefore, $o_{1(C)} = a_{1(C)} + \epsilon_{1(C)}$ and $o_{2(C)} =$
 206 $a_{2(C)} + \epsilon_{2(C)}$ are observation-level random effects aggregating variation due to among- and

207 within-individual differences across measurements, within a given environmental context
 208 defined by C (e.g., a given year, position in space, level of resource abundance). Note that the \mathbf{W}
 209 matrix from **Eq 1** is no longer necessary in **Eq 2** in the absence of repeated measurements. As a
 210 consequence, we expect that the observation-level correlation $r_{o(X)}$ between these random
 211 effects to reflect the combined effect of the among- and within-individual correlations between
 212 life history traits, weighted by the geometric mean of their repeatability R (Dingemanse &
 213 Dochtermann, 2013; Searle, 1961).

$$r_{o(C)} = r_{\alpha(C)} \sqrt{\frac{\sigma_{\alpha 1}^2 \sigma_{\alpha 2}^2}{\sigma_{z 1}^2 \sigma_{z 2}^2}} + r_{\epsilon(C)} \sqrt{\frac{\sigma_{\epsilon 1}^2 \sigma_{\epsilon 2}^2}{\sigma_{z 1}^2 \sigma_{z 2}^2}} = r_{\alpha(C)} \sqrt{R_1 R_2} + r_{\epsilon(C)} \sqrt{(1 - R_1)(1 - R_2)} \quad (3)$$

214
 215 Where phenotypic variances are adjusted for the mean effects of $X_1\beta_{\mu 1}$ and $X_2\beta_{\mu 2}$. We can see
 216 that inferences about among-individual tradeoffs from the non-repeated measures model (**Eq.**
 217 **2**) will be at greatest risk of bias when $\text{sign}(r_{\alpha}) \neq \text{sign}(r_{\epsilon})$ and $\sqrt{R_1 R_2} \ll \sqrt{(1 - R_1)(1 - R_2)}$.
 218 Figure 1 shows these general relationships across correlation and repeatability ranges, identifying
 219 regions of sign bias. Fortunately, researchers will generally be able to judge their risk of inferential
 220 bias based on *a priori* knowledge about the repeatability of life history traits, which tends to be
 221 medium to high (Dingemanse et al., 2021). For example, observation-level correlations of
 222 behavioral traits will tend to be dominated by within-individual associations (Bell et al., 2009;
 223 Cauchoux et al., 2018; Holtmann et al., 2017), while morphological associations will tend to be
 224 dominated by among-individual variation (Dingemanse et al., 2021). We reiterate that our
 225 models consider measurements of the same individuals observed in different contexts as
 226 independent (see supplementary materials Section S1). In addition, our model considers no
 227 measurement errors, as we are not able to disentangle it from true within-individual variation

228 using non-repeated measures. Such considerations regarding trait repeatability and
229 measurement error should be explicit when interpreting results without repeated measures.

230 ***Hybrid scenarios***

231 Variation in repeated sampling is also likely to occur across phenotypes due to factors such as
232 difficulty of measurement and the rate of trait expression. While a single measure of age at first
233 reproduction or fecundity in a given environmental context may be available per individual,
234 multiple individual measures may be available for traits such as offspring quality. Such scenarios
235 require a hybrid modeling approach. For example, consider a model with a single predictor for
236 an intergenerational tradeoff between fecundity (e.g., clutch size) and offspring quality, but other
237 traits could equally be studied. The model structure for offspring quality \mathbf{z}_1 (depicted as offspring
238 body mass), a gaussian trait, is given by

$$\mathbf{z}_1 = \mathbf{X}_1\boldsymbol{\beta}_{\mu 1} + \mathbf{W}\boldsymbol{\alpha}_{1(C)} + \boldsymbol{\epsilon}_{1(C)} \quad (4.1)$$

239
240 The linear predictor for \mathbf{z}_1 (mass of an offspring of a given mother) in year C includes a year-
241 specific mother random effect $\boldsymbol{\alpha}_{1(C)}$ and $\boldsymbol{\epsilon}_{1(C)}$ being the within-brood/litter variance.

242 The model for fecundity \mathbf{z}_2 follows the same basic structure, with a single fecundity measurement
243 per female per year. We can use a Poisson distribution where we model the expected rate of
244 offspring production using a log link function, but other distributions could equally be used.

$$\mathbf{z}_2 = \mathbf{X}_2\boldsymbol{\beta}_{\mu 2} + \boldsymbol{o}_{2(C)} \quad (4.2)$$

245
246 Without repeated measures, the random effect $\boldsymbol{o}_{2(C)}$ is specified at the observation-level,
247 accounting for any overdispersion in the Poisson process across measurements of each female.

248 The context-dependent tradeoff will be estimated between the among-mother random effect in
 249 offspring quality and the observation-level random effect in fecundity.

$$\begin{aligned}
 & [\boldsymbol{\alpha}_{1(C)}, \boldsymbol{o}_{2(C)}] \sim N(\mathbf{0}, \mathbf{P}_{(C)}) & (4.3) \\
 \mathbf{P}_{(C)} = & \begin{bmatrix} \sigma_{\alpha_{1(C)}}^2 & r_{(C)} \sigma_{\alpha_{1(C)}} \sigma_{o_{2(C)}} \\ r_{(C)} \sigma_{\alpha_{1(C)}} \sigma_{o_{2(C)}} & \sigma_{o_{2(C)}}^2 \end{bmatrix} \\
 & \operatorname{atanh}(r_{(C)}) = \mathbf{X}_3 \boldsymbol{\beta}_r
 \end{aligned}$$

250
 251 Reducing **Eq. 3**, the correlation r_C between the individual- $a_{1(C)}$ and observation-level $o_{2(C)}$
 252 effects will necessarily be proportional to the among-individual correlation across life history
 253 traits.

$$r_{o(C)} = r_{\alpha(C)} \sqrt{\frac{\sigma_{\alpha 2}^2}{\sigma_{z 2}^2}} = r_{\alpha(C)} \sqrt{R_2} \quad (5)$$

254
 255 Note that this method does not allow the inclusion of non-continuous traits (e.g., Bernoulli traits)
 256 in the absence of repeated measurements within a given environmental context C (e.g., a given
 257 year).

258

259 Validation on simulated datasets

260 We validated the CRN model on two different types of tradeoffs. First, we used the hybrid CRN
 261 model to study an intergenerational tradeoff between fecundity and quality. The hybrid model
 262 is well suited because fecundity (i.e., clutch/litter size) has a single measurement per mother per
 263 year, while offspring quality (i.e., offspring mass) has repeated measurements per mother per
 264 year (one measurement for each offspring produced). Second, we used the non-repeated
 265 measures CRN model to study an intraindividual tradeoff between fecundity (clutch/litter size)

266 and parental growth (the change of mass from a year to the next). The non-repeated measures
267 CRN model is well suited as both traits are expressed only a single time per year (one fecundity
268 and one parental growth measure per individual per year). Note that tradeoffs are described as
269 intergenerational or intraindividual depending on which traits are studied (as explained in
270 Stearns, 1989), and both type of tradeoff can be decomposed into among- and within-individual
271 covariation. We simulate data for these two tradeoffs using the individual-based simulation
272 described in Bliard et al. (2024), whereby the among-individual correlation between life history
273 traits can be made dependent on the environmental context. The code to generate data from
274 the individual-based simulation can be found on github
275 (https://github.com/lbiard/detecting_tradeoffs_crn_models). This model validation is only
276 intended to show that context-dependent among-individual correlations (i.e., context dependent
277 tradeoffs) can be successfully recovered. For a more extensive simulation-based calibration of
278 CRN models over a broad range of parameter values, see Martin (2023).

279 ***Intergenerational tradeoff (offspring quantity-quality)***

280 We first focused on an intergenerational tradeoff between offspring quantity and quality (hybrid
281 CRN model). This quantity-quality tradeoff has been the focus of numerous studies since Lack's
282 pioneering work on bird clutch sizes (Einum & Fleming, 2000; Fischer et al., 2011; Gillespie et al.,
283 2008; Lack, 1947; Williams, 1966). We simulate 30 years of individual-based data in which 25 new
284 individuals enter the population each year, reproduce with an average clutch/litter size of 2.5,
285 and then have a probability to survive to next year of 0.6. This yielded a final simulated dataset
286 of 750 individuals, totaling 1578 reproductive events and 4783 offspring. An observation-level
287 correlation was included between offspring mass and clutch size, and this correlation was made

288 dependent on a single climatic predictor. The same climatic predictor was also included to
289 influence both clutch size and offspring mass.

290 ***Intraindividual tradeoff (fecundity-growth)***

291 We then simulated data for an intraindividual tradeoff between fecundity and growth (non-
292 repeated measures CRN model). This simulated dataset is also made of 30 years and 750
293 individuals, for a total of 1974 reproductive events, with a variable observation-level correlation
294 between individual growth and fecundity, which is itself dependent on a single climatic predictor.

295

296 Study systems and application on empirical datasets

297 ***Marmots***

298 We applied the hybrid CRN model (one trait with repeated individual measurements within a
299 year and one trait without) on data from a yellow-bellied marmot population monitored at the
300 Rocky Mountain Biological Laboratory in Gothic, Colorado (38°57'N, 106°59'W) during the
301 summer season each year, whereby extensive individual-based data is collected (Armitage, 2014;
302 Blumstein, 2013). In Alpine marmots *Marmota marmota*, an offspring quality-quantity tradeoff
303 has been found (Berger et al., 2015), while it remained mostly elusive in yellow-bellied marmots,
304 being only found for older mothers (Kroeger et al., 2020), whereby within-cohort selection has
305 likely reduced the amount of among-individual variance in resource acquisition, thus making the
306 tradeoff visible (Kendall et al., 2011; van Noordwijk & de Jong, 1986). Therefore, we searched for
307 an intergenerational tradeoff between mothers' fecundity and offspring estimated mass
308 (offspring quality-quantity tradeoff). We used repeated measurements of offspring mass for each
309 mother (one mass estimate for each offspring in a given litter). The offspring weaning mass was

310 imputed based on the date of emergence for each litter and mass measurements from captures
311 later in the season, following the method of Ozgul et al. (2010). We considered two measures
312 quantifying environmental conditions for a given year. First, the total amount of snow during the
313 preceding winter, with years of little overwinter snow considered harsher for marmots as it offers
314 limited thermal insulation during the hibernation (Barash, 1973; Cordes et al., 2020; Wells et al.,
315 2022). Second, the average daily maximum temperature during the month of June, with warmer
316 summer temperatures considered unfavorable conditions for marmots as they are prone to
317 overheating, hence limiting the time that can be allocated to foraging (Cordes et al., 2020; Krajick,
318 2004; Melcher et al., 1990). Note that we used temperature in June and not July as commonly
319 used in this system (Cordes et al., 2020), because this is more likely to represent the conditions
320 experienced for most offspring before emergence and weaning, since most offspring emerge in
321 July. We expected tradeoffs to be more strongly expressed among individuals in years with little
322 overwinter snow or high summer temperature. In total, we used 2540 offspring mass from 597
323 reproductive events, from 279 females across 42 years.

324 We modeled offspring mass using a normal distribution (**Eq. 6.1**), and we included as
325 covariates (i.e., in \mathbf{X}_1) the total amount of snow during the winter, June average maximum
326 temperature, age of the mother and its quadratic effect, and mother's estimated mass in early
327 June. A year random effect δ_1 was also included.

$$\mathit{offspring\ mass} = \mathbf{X}_1\boldsymbol{\beta}_1 + \delta_1 + \mathbf{W}\boldsymbol{\alpha}_{1(Y)} + \boldsymbol{\epsilon}_{1(Y)} \quad (6.1)$$

328
329 With $\boldsymbol{\alpha}_{1(Y)}$ being a year-specific mother random effect and $\boldsymbol{\epsilon}_{1(Y)}$ the within-litter variance.

330 We modeled the second trait, fecundity (i.e., litter size), using a Poisson distribution (**Eq. 6.2**), as
 331 a function of the same covariates (\mathbf{X}_2), except June average maximum temperature, since it
 332 cannot affect fecundity as pregnancies mostly occur before this period. A year random effect δ_2
 333 was also included.

$$\log(\mathbf{litter\ size}) = \mathbf{X}_2\boldsymbol{\beta}_2 + \delta_2 + \mathbf{o}_{2(Y)} \quad (6.2)$$

334
 335 For the observation-level correlation (**Eq. 6.3**), the two environmental variables (winter snow and
 336 June temperature) were added as covariates (\mathbf{X}_3).

$$[\boldsymbol{\alpha}_{1(Y)}, \mathbf{o}_{2(Y)}] \sim N(\mathbf{0}, \mathbf{P}_{(Y)}) \quad (6.3)$$

$$\mathbf{P}_{(Y)} = \begin{bmatrix} \sigma_{\alpha_1(Y)}^2 & r_{(X)} \sigma_{\alpha_1(Y)} \sigma_{o_2(Y)} \\ r_{(Y)} \sigma_{\alpha_1(Y)} \sigma_{o_2(Y)} & \sigma_{o_2(Y)}^2 \end{bmatrix}$$

$$\text{atanh}(r_{(Y)}) = \mathbf{X}_3\boldsymbol{\beta}_3$$

337
 338 We performed posterior predictive checks, showing a good concordance between the litter size
 339 data, and data generated under the model (see Figure S3). However, the model slightly
 340 underestimates the variance in offspring mass. Overall, posterior predictive checks highlight that
 341 the use of a Normal distribution to model offspring mass, and a Poisson distribution with an
 342 observation random effect to model litter size, were appropriate in this system.

343

344 ***Soay sheep***

345 We applied the non-repeated measures CRN model on Soay sheep data, as we have no repeated
 346 individual measurement within a given year available for neither of the traits studied. We used
 347 data from an unmanaged population of feral sheep in the Village Bay area of the island of Hirta
 348 (57°48'N, 8°37'W), which has been monitored since 1985 (Clutton-Brock & Pemberton, 2004). In

349 Soay sheep, survival costs of reproduction were found for breeding ewes, particularly in
350 populations at high densities or following stormy winters (Tavecchia et al., 2005). Therefore, we
351 searched for an intraindividual tradeoff between ewes' fecundity defined as the number of lambs
352 born in Spring (ranging from 0 to 2) and their log mass in the following summer, with both traits
353 conditional on ewes surviving the winter. We considered two environmental variables to
354 characterize the ecological harshness faced by the sheep in a given year: population density and
355 NAO (North Atlantic Oscillation) in the winter preceding parturition, with high NAO values
356 corresponding to wet and stormy winters (Coulson et al., 2001; Regan et al., 2022). In total, we
357 used data from 2497 reproductive events across 37 years, for 861 ewes with known mass in the
358 summer preceding the reproductive event, as well as known mass in the following summer. We
359 expected tradeoffs to be more strongly expressed in years of high population density or high
360 NAO.

361 As ewes' fecundity in a given year is restricted to [0,2], we could not use a Poisson
362 regression. This is due to the count data being underdispersed relative to a Poisson distribution.
363 We therefore modeled the ewe's fecundity using an ordinal regression (also called cumulative
364 logistic regression; **Eq. 7.1**), and we included as covariates (\mathbf{X}_1) the individual's log mass preceding
365 the reproductive event, age and its quadratic effect, and population density.

$$\text{logit}(\Pr(\mathbf{N}_{\text{offspring}} \leq i)) = \theta_i - (\mathbf{X}_1 \boldsymbol{\beta}_1 + \boldsymbol{\delta}_1 + \boldsymbol{o}_{1(Y)}) \quad (7.1)$$

366
367 Where the cumulative probability of having at most i offspring is given as a function of the
368 threshold θ_i and the matrix of covariates \mathbf{X}_1 , as well as a year random effect $\boldsymbol{\delta}_1$ and a year specific
369 observation random effect $\boldsymbol{o}_{1(Y)}$.

370 We modeled the ewe's log mass in the following summer using a normal distribution (Eq. 7.2),
 371 and included in \mathbf{X}_2 the same covariates as in \mathbf{X}_1 , as well as NAO in the winter preceding parturition.
 372 A year random effect δ_2 was also included.

$$\mathbf{mass} = \mathbf{X}_2\boldsymbol{\beta}_2 + \delta_2 + \mathbf{o}_{2(Y)} \quad (7.2)$$

373
 374 For the observation-level correlation (Eq. 7.3), the two ecological variables (winter NAO and
 375 density) were added as covariates (\mathbf{X}_3).

$$[\boldsymbol{\alpha}_{1(Y)}, \mathbf{o}_{2(Y)}] \sim N(\mathbf{0}, \mathbf{P}_{(Y)}) \quad (7.3)$$

$$\mathbf{P}_{(Y)} = \begin{bmatrix} \sigma_{\alpha_1(Y)}^2 & r^{(Y)} \sigma_{\alpha_1(Y)} \sigma_{o_2(Y)} \\ r^{(Y)} \sigma_{\alpha_1(Y)} \sigma_{o_2(Y)} & \sigma_{o_2(Y)}^2 \end{bmatrix}$$

$$\text{atanh}(r^{(Y)}) = \mathbf{X}_3\boldsymbol{\beta}_3$$

376
 377 The posterior predictive checks we performed highlighted a good fit between the data and data
 378 generated under the model. This confirms that using a normal distribution to model ewe's mass,
 379 and using a cumulative logistic regression to model ewe's number of offspring, were appropriate
 380 (see Figure S4).

381

382 Model implementation

383 We implemented all multivariate models described above in a Bayesian framework using the Stan
 384 statistical language (Carpenter et al., 2017), through the software R (R Core Team, 2021) using
 385 the R package *CmdStanR* (Gabry & Češnovar, 2020). Stan was preferred for model
 386 implementation because of its flexibility. Common regularizing priors were used for all model
 387 parameters: normal distributions of mean 0 and standard deviation of 1 for intercepts and slopes
 388 coefficients, and exponential distributions of rate 2 for variance parameters. Each model ran on

389 3 chains, with a burn-in period of 1000 iterations, sampling for 3000 iterations, keeping all the
390 sampled iterations (Link & Eaton, 2012). Convergence of parameter estimates was assessed
391 visually and using the Gelman-Rubin diagnostic (Gelman & Rubin, 1992). We report the full
392 posterior distributions, alongside their mean, 50%, and 89% credible intervals (McElreath, 2020).
393 The Stan code to implement all the CRN models presented in this study is archived on GitHub
394 (https://github.com/lbiard/detecting_tradeoffs_crn_models) and Zenodo (will be added upon
395 acceptance of the manuscript).

396

397

Results

398 The model validation performed on simulated datasets showed that parameters were correctly
399 recovered for both intergenerational tradeoffs (Figure 2) and intraindividual tradeoffs (Figure 3).
400 While these simulation examples do not quantify bias of estimations (more details from a
401 simulation-based calibration of CRN models are available in Martin (2023)), they still confirm that
402 the model presented in the methods is able to detect context-dependence in the expression of
403 tradeoffs.

404 The model applied to yellow-bellied marmot data shows trends towards tradeoffs being
405 more strongly expressed in years with harsh environmental conditions, albeit with high
406 uncertainty in the estimates (Figure 4). We found a positive mean effect of the amount of
407 overwinter snow on the correlation (Figure 4), meaning that the tradeoff between fecundity and
408 offspring quality was more strongly expressed after winters with little snow. We also found a
409 negative mean effect of the average maximum June temperature on the correlation (Figure 4),
410 where females with more offspring were more likely to have lighter offspring during warmer

411 summers. Estimated effects of covariates on either fecundity or offspring mass can be found in
412 Figure 4, as well as in Figure S5.

413 Estimated effects of covariates on the correlation also had high uncertainty in the Soay
414 sheep dataset (Figure 5). Overall, we found that the correlation tended to be negative across all
415 environments, which means that ewe's growth was lower for the ones that weaned offspring
416 (Figure 5). Contrary to our expectations, while we hypothesized that the tradeoffs should be
417 more strongly expressed in wet and stormy winters (high NAO index), we found a positive effect
418 of winter NAO on the correlation between fecundity and growth (Figure 5). We did not find any
419 clear effect of population density on the expression of the tradeoff (Figure 5). Estimated effects
420 of covariates on either fecundity or ewe's mass can be found in Figure 5, as well as in Figure S6.

421

422 **Discussion**

423 Our proof-of-concept study demonstrates that hierarchical multivariate CRN models (Martin,
424 2023) can be used successfully to detect and estimate context-dependent changes in tradeoff
425 expression, though estimation uncertainty can be large. In agreement with theoretical
426 predictions and despite large uncertainty, we found that reproductive tradeoffs in yellow-bellied
427 marmots tend to be more strongly expressed under unfavorable climatic conditions. In Soay
428 sheep, we found some context-dependence in the expression of the tradeoff, but effect
429 directions were opposite to our initial prediction. This hierarchical model has the potential to be
430 used on many long-term individual-based datasets and could help improve our understanding of
431 tradeoff expression and life history theory.

432 Although the initial motivation to use this method partly rested on the observed difficulty
433 of finding tradeoffs in empirical datasets, we found that in both sheep and marmots, the
434 tradeoffs tend to be expressed across most environments, with mean phenotypic correlations
435 being negative overall. Thus, ironically, in these two empirical datasets, tradeoffs might have
436 been detected using simpler multivariate methods without the need for context dependence.
437 However, this should not come as a surprise for Soay sheep, as this negative correlation between
438 growth and fecundity was already found on a smaller dataset (Fung et al., 2022). Nonetheless,
439 the results still highlight that context-dependence has the potential to hinder our ability to detect
440 tradeoffs in some cases. For instance, when marmots experience favorable environmental
441 conditions, the average correlation is closer to null with credibility intervals nearing or
442 overlapping zero (Figure 4), while the intergenerational tradeoff is found to be more strongly
443 expressed during harsh years. In Soay sheep, context dependence appears to be marked for the
444 expression of the tradeoff, but opposite to our predictions. Indeed, we found a positive
445 correlation between growth and fecundity only under the harshest environmental conditions
446 (high population density and high winter NAO, Figure 5). Since ewes' mass is measured in the
447 following summer and not directly after parturition, harsh winter conditions are expected to
448 increase overwinter mortality (Milner et al., 1999), lowering spring population density and
449 reducing competition. This could potentially help surviving ewes to recover their body condition
450 between spring and summer, which is the period of greatest grass growth, hence potentially
451 explaining our counter-intuitive results. We can also speculate that the result could have arisen
452 from two potential pitfalls due to idiosyncrasies of the Soay sheep data. First, among-individual
453 variation in fecundity is limited in sheep, ranging from no offspring to twins, potentially making

454 it more complicated for the model to estimate variances accurately (Fay, Authier, et al., 2022;
455 Kain et al., 2015). Second, both ewes' growth and fecundity are conditional on survival in the
456 data, hence individuals who suffered most from the cost of reproduction and did not survive are
457 not present in the analysis, potentially biasing the results (Hadfield, 2008).

458 Despite the potential of this modeling approach to study context-dependent tradeoffs, a
459 few methodological limitations are to be considered. A recent study conducted by Fay et al.
460 (2022) highlighted that multivariate models with correlated random effects for Bernoulli traits
461 performed rather poorly, resulting in a potentially large bias and imprecise estimates of variances
462 and covariances. This is in part because Bernoulli traits contain less information than continuous
463 variables, making estimations of variances complicated (Fay, Authier, et al., 2022), but also
464 because the data available to estimate individual heterogeneity is usually scarce (Browne et al.,
465 2007). The model we present suffers from this limitation, and even more so when there is only a
466 single individual observation per individual per sampling occasion (e.g., parental survival), and
467 when the trait is not repeatable (death can only occur once). This issue renders the model, as
468 well as any other multilevel model, unable to meaningfully estimate distinct mean and variance
469 parameters for Bernoulli traits, due to the fact that the mean p of a Bernoulli variable determines
470 its variability $p(1-p)$ without scope for overdispersion. Therefore, environmental effects on the
471 mean of Bernoulli measures will necessarily change their variances (Skrondal & Rabe-Hesketh,
472 2007). However, when repeated Bernoulli observations or a binomial measure are available
473 within each sampling occasion (e.g., survival of each offspring within a litter), the CRN model can
474 then be used to partition distinct environmental effects on trait means and (co)variances. As we
475 have shown in the present study, despite this limitation, the CRN remains applicable to single

476 measures of continuous traits and count measures (e.g., growth, fecundity, phenology,
477 behavioral traits), as well as proportions and various other kinds of non-Gaussian measures.
478 Another limitation of the proposed method is that sample sizes needed are likely to be large,
479 with enough individuals in each environmental context, and importantly enough sampling
480 occasions across which to estimate the context dependence of tradeoff expression. Nonetheless,
481 many long-term individual-based studies should have enough data to fulfill these requirements
482 (de Villemereuil et al., 2020).

483 Despite the abovementioned caveats and limitations of the methodology in the absence
484 of repeated measurements, this new model is a development that could be useful for many
485 datasets. Thanks to its implementation in a Bayesian framework using Stan (Carpenter et al.,
486 2017), it offers great flexibility and can be easily repurposed and modified to fit the idiosyncrasies
487 of a given dataset or species life history. It is also straightforward to extend the model by adding
488 a pedigree for quantitative genetic analysis (see Martin, 2023), even though phenotypic
489 correlations should be good approximations of genetic correlations in most cases (Cheverud,
490 1988; Dochtermann, 2011; Roff, 1995). While we presented a bivariate model, this model is not
491 necessarily limited to two traits, and more continuous traits and their covariances could also be
492 analyzed simultaneously. We also restricted our proof-of-concept study to the reaction norm of
493 the correlation between traits, but researchers interested in the canalization of traits variances
494 as a response to the environmental context could also benefit from this modeling approach
495 (Péron et al., 2016).

496 Life history tradeoffs have long been sought after, but difficult to detect in observational
497 data due to individual heterogeneity (Metcalf, 2016; Reznick et al., 2000; van Noordwijk & de

498 Jong, 1986). Previous studies have also highlighted that life history tradeoffs could be expressed
499 only under unfavorable ecological conditions (Cohen et al., 2020; Stearns, 1989). Yet, despite our
500 knowledge of the issues hindering tradeoff detection, we still lacked a statistical framework that
501 permits the detection of context-dependence in tradeoff expression. Our proof-of-concept study
502 shows that this context dependence can be detected. This method has the potential to be applied
503 by demographers and evolutionary ecologists having long-term individual-based datasets at
504 hands, with many study systems having the required data (Culina et al., 2021; de Villemereuil et
505 al., 2020). Altogether, this method has the potential to help us improve our understanding of life
506 history theory, and in part resolve van Noordwijk and de Jong's (1986) conundrum of tradeoff
507 detection, by accounting for the context-dependence of their expression.

508

509 **Acknowledgements**

510 This work was supported by a Swiss National Science Foundation Grant (31003A_182286 to A.O).
511 We thank Billy Barr for the long-term collection of snow data at RMBL, and all the people who
512 helped collect data on marmots and sheep over the years. Most recently, marmots were studied
513 with support from the US National Science Foundation (D.E.B.-1119660 and 1557130 to D.T.B.).
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515 Research Council (NERC), and is indebted to many project members and volunteers; the National
516 Trust for Scotland for permission to work on St Kilda and QinetiQ and Kilda Cruises for logistical
517 support in the field. We thank the editor, an anonymous reviewer, and Remi Fay for useful
518 comments and suggestions.

519

520 **Authors contributions**

521 LB, JSM, AO, MP, DZC conceived the study. JSM designed the initial modeling framework and LB
522 analyzed the data. DTB, JGAM, JMP, DHN collected and curated the data. LB and JSM wrote the
523 first draft with input from AO, MP, DZC. All authors contributed to the editing of the manuscript.

524

525 **Data and code availability**

526 The data, as well as the R and Stan code necessary to reproduce the results are available on
527 GitHub (https://github.com/lbiard/detecting_tradeoffs_crn_models). They will be archived on
528 Zenodo upon acceptance.

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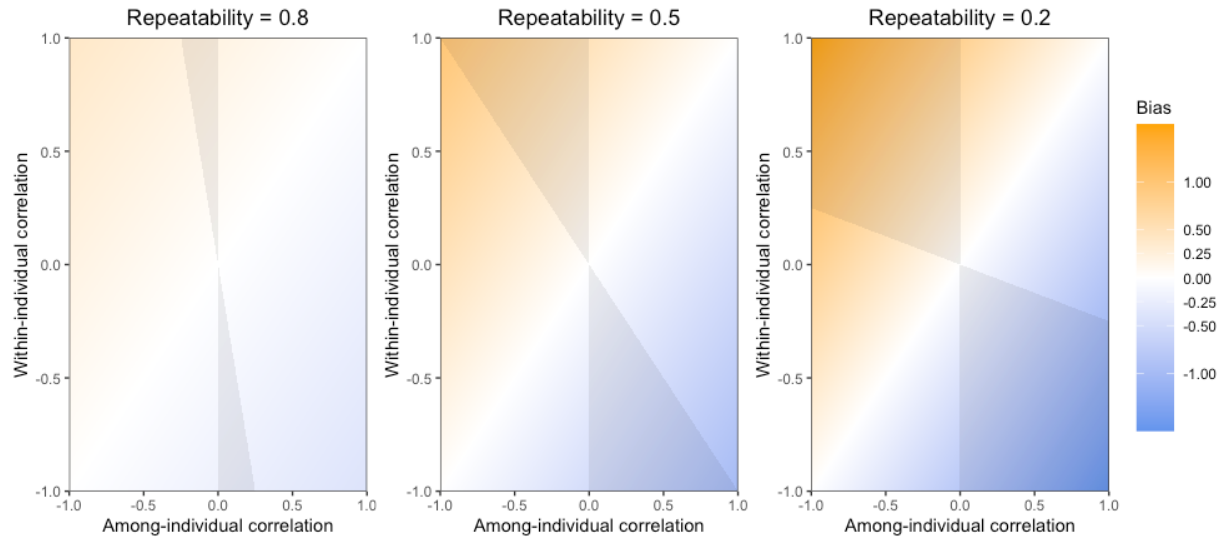
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542 **Figures**



543

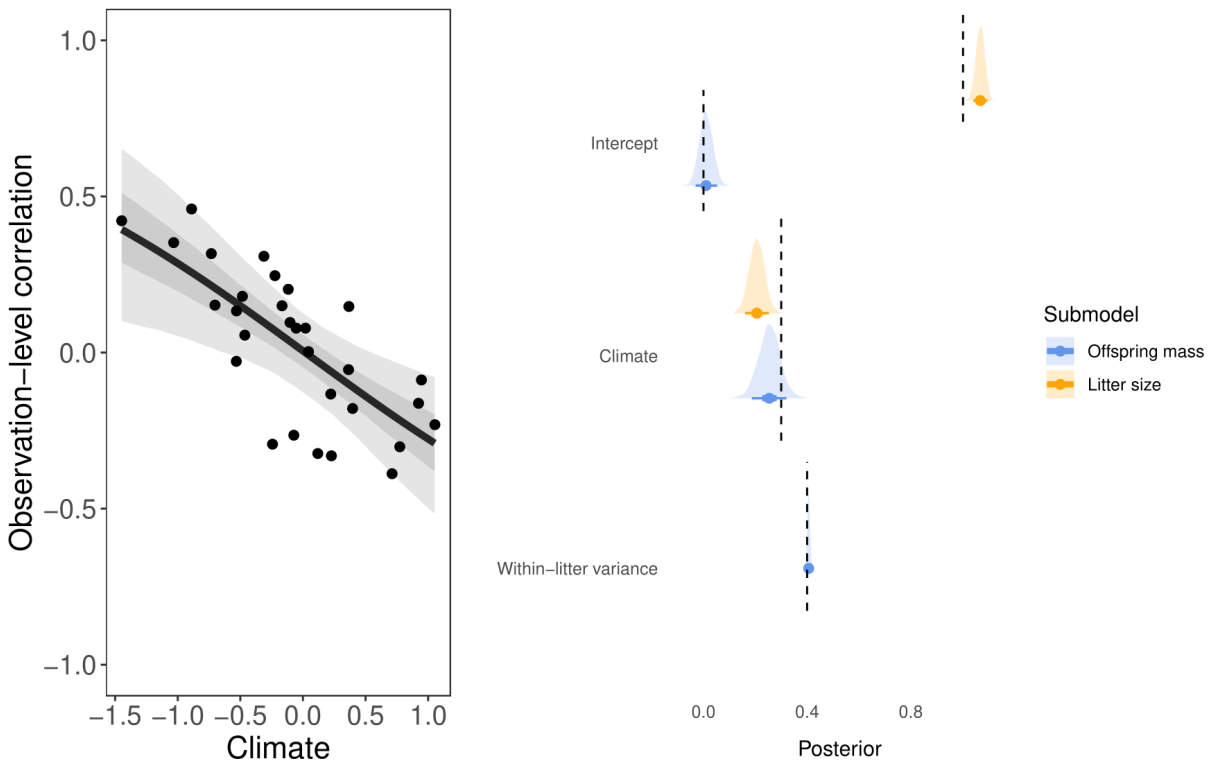
544 Figure 1: General relationships across correlations and repeatability ranges based on **Eq 3** for a
545 non-repeated measures CRN (model of **Eq 2**), identifying the magnitude of correlation bias and
546 the regions of sign bias. The bias is here defined as the difference between the observation-level
547 correlation and the among-individual correlation, using the latter as a reference. Parameter
548 spaces in gray represent the regions of sign bias, where the observation-level correlation has a
549 sign opposite to the among-individual correlation. This highlights that the observation-level
550 correlation is mostly influenced by the among-individual correlation for traits with high
551 repeatability, while it is mostly influenced by the within-individual correlation for traits with low
552 repeatability.

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558 Figure 2: Left panel: estimated vs. simulated observation-level correlation between litter size and

559 offspring mass as a function of climate, after accounting for the effect of climate on both traits.

560 The regression line indicates the mean effect of climate on the correlation, while the shaded

561 areas depict the 50% and 89% credible intervals predicted by the model. Each black dot

562 represents the simulated observation-level correlation between both traits in a given year

563 depending on climate. Right panel: estimated vs. simulated intercepts and slopes for the

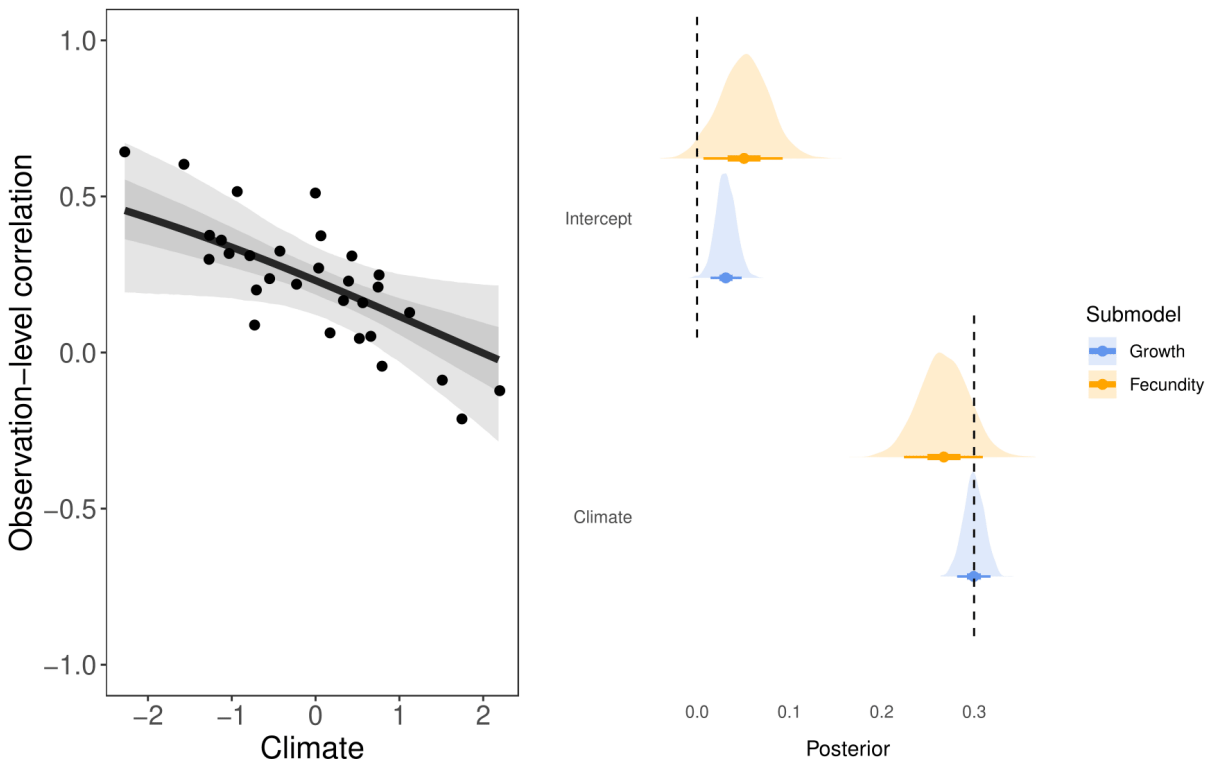
564 offspring mass and litter size sub-models. Dashed lines represent the value used to simulate the

565 data, while the distributions and intervals represent the posterior distributions estimated by the

566 model, alongside the median, 50%, 89% credible intervals. Litter size estimates are presented on

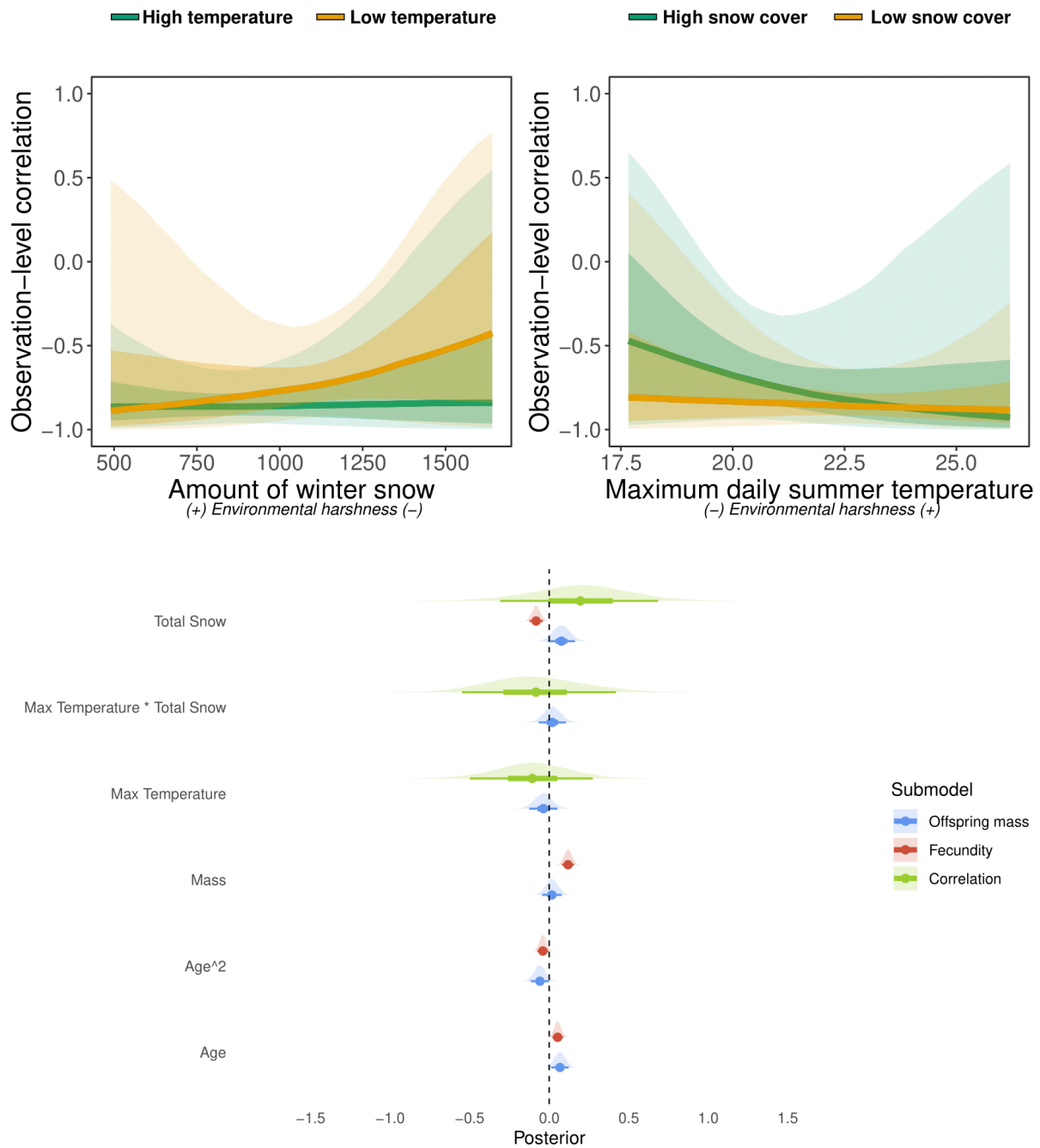
567 the log scale.

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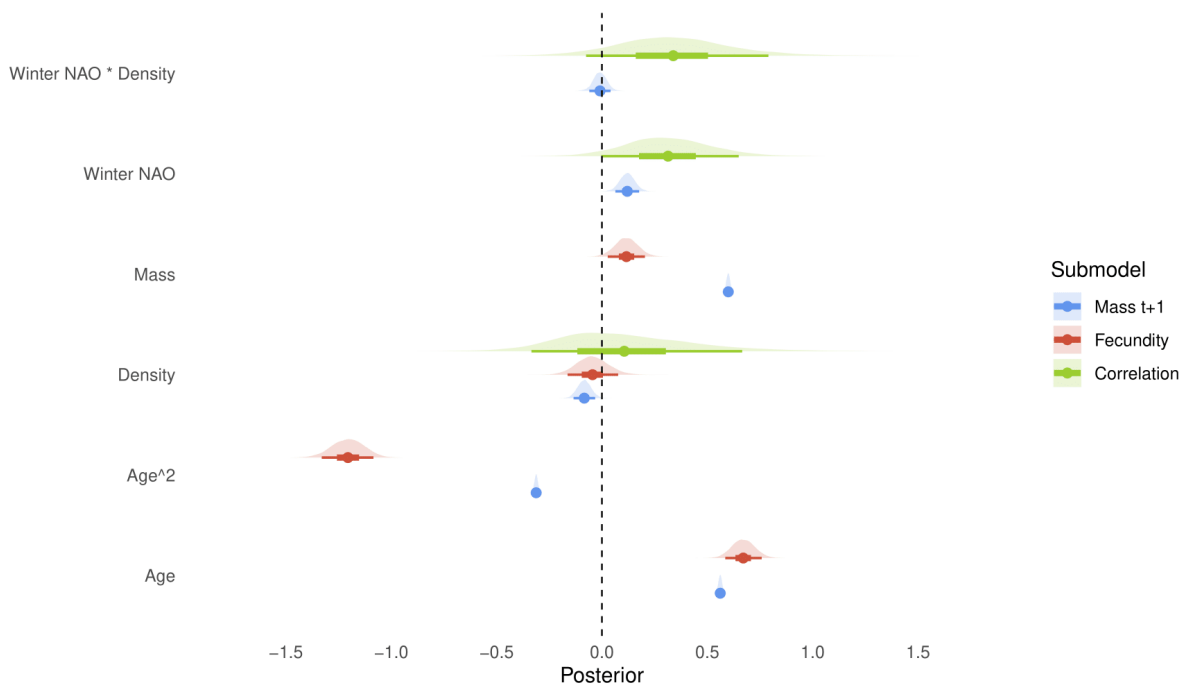
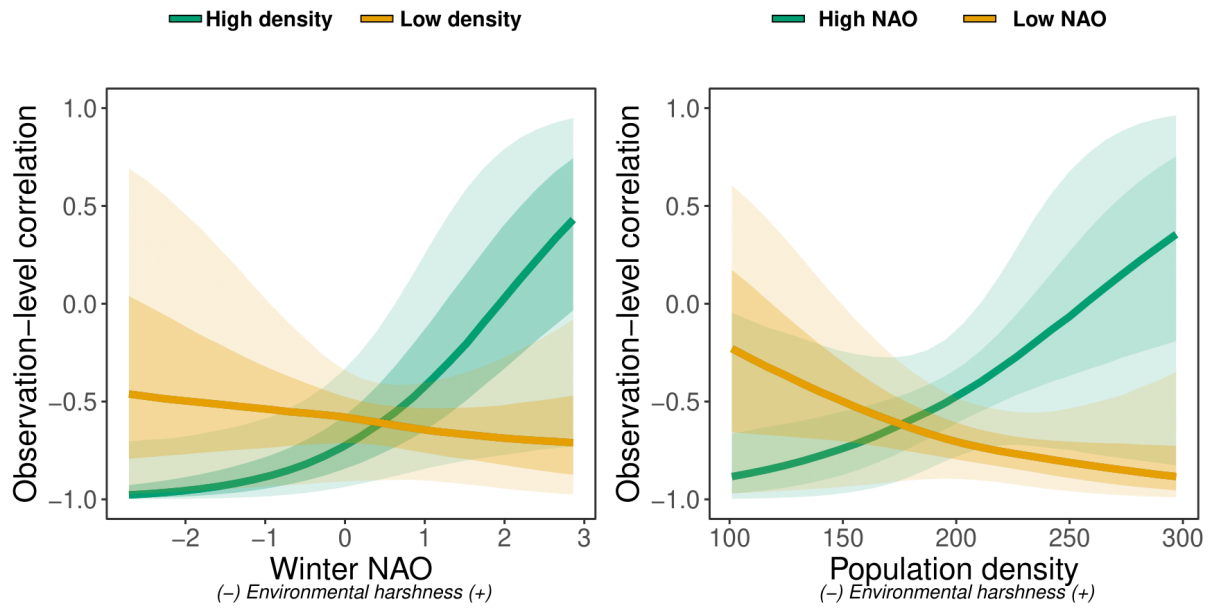
570 Figure 3: Left panel: estimated vs. simulated observation-level correlation between fecundity and
 571 growth as a function of climate, after accounting for the effect of climate on both traits. The
 572 regression line indicates the mean effect of climate on the correlation, while the shaded areas
 573 depict the 50% and 89% credible intervals predicted by the model. Each black dot represents the
 574 simulated observation-level correlation between both traits in a given year depending on climate.
 575 Right panel: estimated vs. simulated intercepts and slopes for the growth and fecundity sub-
 576 models. Dashed lines represent the value used to simulate the data, while the distributions and
 577 intervals represent the posterior distributions estimated by the model, alongside the median,
 578 50%, 89% credible intervals. Fecundity estimates are presented on the log scale.



579

580 Figure 4: Observation-level correlation between litter size and offspring mass in marmots as a
 581 function of the total amount of snow in the preceding winter at high and low temperature (top
 582 left panel) and the maximum daily June temperature of the year at high and low snow cover (top
 583 right panel). Estimated effects of standardized predictors (bottom panel) on offspring mass,

584 fecundity, and the observation-level correlation between both traits in marmots. The regression
 585 line indicates the median estimated effect, while the shaded areas depict the 50% and 89%
 586 credible intervals predicted by the model.



588 Figure 5: Observation-level correlation between fecundity and mothers' mass in the following
589 year in Soay sheep as a function of the winter NAO at high and low density (top left panel), and
590 as a function of the population density at high and low winter NAO values (top right panel).
591 Estimated effects (bottom panel) of standardized predictors on mother's mass in the following
592 year, fecundity, and the observation-level correlation between both traits in Soay sheep. The
593 figure displays the posterior distributions estimated by the model, alongside the median, 50%,
594 and 89% credible intervals.

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610 **Supplementary materials**

611 **Section S1: fixed individual heterogeneity**

612 The models presented in the main text do not include distinct parameters for fixed individual
613 heterogeneity across environmental contexts, hence considering observations from the same
614 individual but in different environmental contexts as independent. Here, we illustrate why this
615 limitation is needed to correctly estimate context-dependent covariation. Using simulated data,
616 we illustrate that it is not possible to estimate the among-individual variation across context,
617 while at the same time estimating among- and within- individual variation within context.
618 However, it is important to note that this should not have any consequences regarding the
619 accuracy of the estimation of the context-dependent correlations.

620 For this purpose, we simulate demographic data with a tradeoff between parental growth
621 and fecundity, suitable for the non-repeated measures CRN (model of equation 2). We include a
622 fixed heterogeneity component (context-independent individual random effect), as well as the
623 context-dependent component (context-dependent individual random effect) to make the
624 correlation vary across contexts. The data is similar to what is presented in the “validation on
625 simulated datasets” section, with the addition of the fixed heterogeneity component. We then
626 analyze this simulated dataset either with a model that estimates only the context-dependent
627 covariation (model presented in the manuscript, equation 2), or a model that does include fixed
628 (context-invariant) individual random effects in addition to the context-dependent covariation
629 (model of equation 2 with the addition of a context-independent individual random effect).

630 The figures for both scenarios are presented below, with Figure S1 highlighting that the
631 inclusion of a fixed among-individual random effect in the model leads to an erroneous

632 estimation of the context-dependent correlation. This is because the inclusion of this fixed
633 among-individual random effect captures part of the variation from the context-dependent
634 random effects, and therefore the context-dependent term will then only estimate deviations of
635 individuals from the fixed heterogeneity term. However, Figure S2 highlights that not including a
636 fixed among-individual random effect allows the model to properly recover the context-
637 dependent correlation.

638 Our results here reflect a more general theoretical point about the biological
639 interpretation of reaction norms. For any reaction norm model, there will not be a distinct
640 component of fixed individual heterogeneity separated from the process of phenotypic plasticity
641 shaping individual heterogeneity across environments. With simple linear reaction norms,
642 empiricists often conceptualize the intercept of the model as reflecting a fixed, environmentally
643 invariant component of the response, separate from the plastic effects described by reaction
644 norm slopes. However, while this can be heuristically useful for some purposes, it is in a strict
645 sense misleading, as the reaction norm intercept simply describes the variation expected when
646 the environmental variable defining the slope is fixed to 0 (e.g., in the average environment for
647 a mean-centered predictor or in the absence of an environmental exposure). Therefore, the value
648 of the intercept is no more fixed than the expected value at any other position along the slope
649 with respect to a fixed value of the environmental gradient. This thinking applies to the CRN and
650 any other reaction norm model. When sufficient data is available, individual random slopes could
651 also be estimated, which can be used to directly quantify the degree to which individuals' rank
652 order may shift across environments (Mitchell & Houslay, 2021). However, the depth of repeated
653 sampling required to fit such models for present purposes is unlikely to be achieved by many

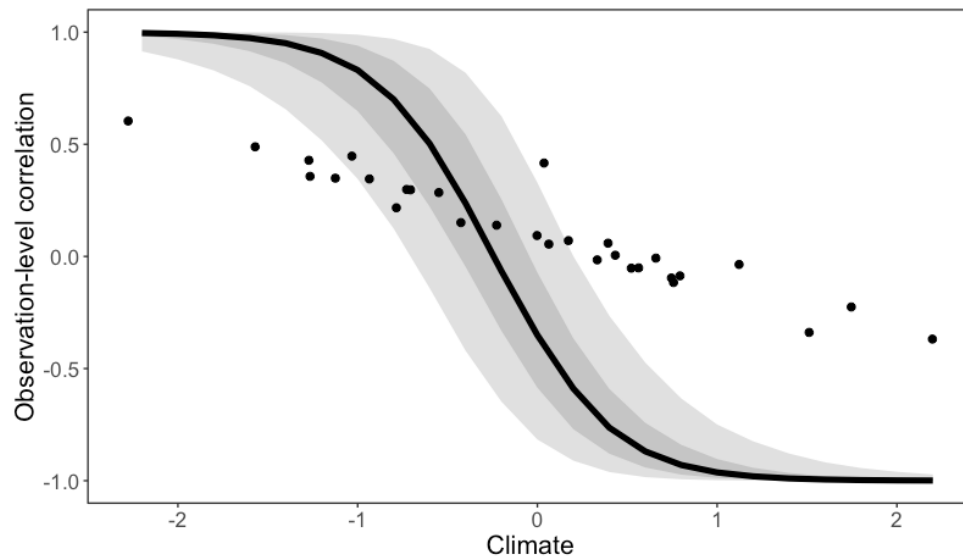
654 currently existing datasets, motivating our CRN approach. Moreover, these random individual
655 slopes will generally be of less interest for detecting demographic tradeoffs, as compared to the
656 average shift in among-individual trait covariance across the population as determined by the
657 fixed CRN slopes.

658 Taking a CRN approach to one's data thus requires taking seriously that there may not be
659 any biologically meaningful sense in which there is a fixed level of individual heterogeneity
660 irrespective of the environment (for traits that exhibit phenotypic plasticity). Rather, there is
661 simply the amount of individual heterogeneity given a particular environment, prior to exposure
662 to the environment, averaged across environments, and/or in the average environment. The
663 parameters from the CRN can always be used to predict any such quantities of interest. For
664 instance, applying the inverse link function to the intercept of the CRN (the first element of β_r)
665 will describe the expected trait correlation when the environmental covariates are fixed to zero.

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670 Figure S1: Estimated context-dependent correlation when a fixed individual-random effect is

671 included. The regression line indicates the mean effect of climate on the correlation, while the

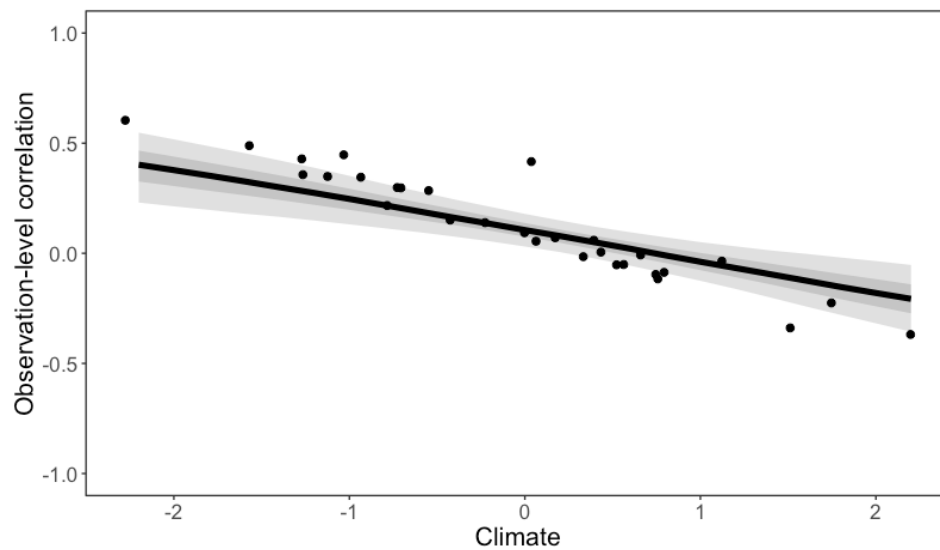
672 shaded areas depict the 50% and 89% credible intervals predicted by the model. Each black dot

673 represents the simulated observation-level correlation between both traits in a given year

674 depending on climate. This highlights that the inclusion of a fixed individual random effect leads

675 to a biased estimation of the context-dependent correlation.

676



677

678 Figure S2: Estimated context-dependent correlation without the inclusion of a fixed individual-
679 random effect in the model. The regression line indicates the mean effect of climate on the
680 correlation, while the shaded areas depict the 50% and 89% credible intervals predicted by the
681 model. Each black dot represents the simulated observation-level correlation between both traits
682 in a given year depending on climate. This highlights that not including a fixed individual random
683 effect leads to an appropriate estimation of the context-dependent correlation.

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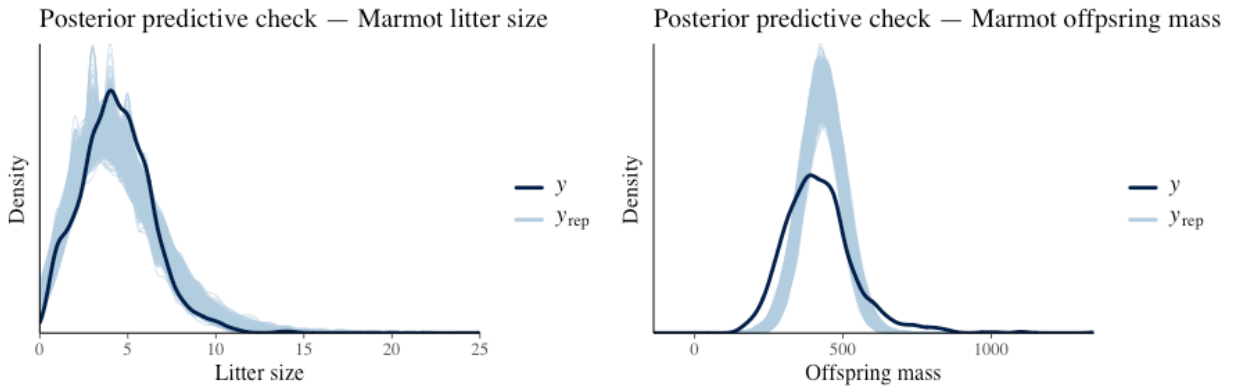
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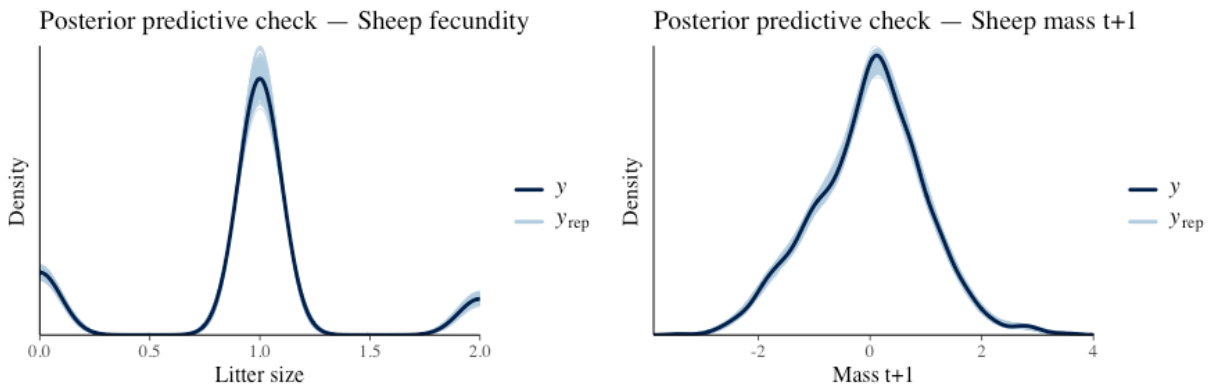
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692 **Section S2: posterior predictive checks**



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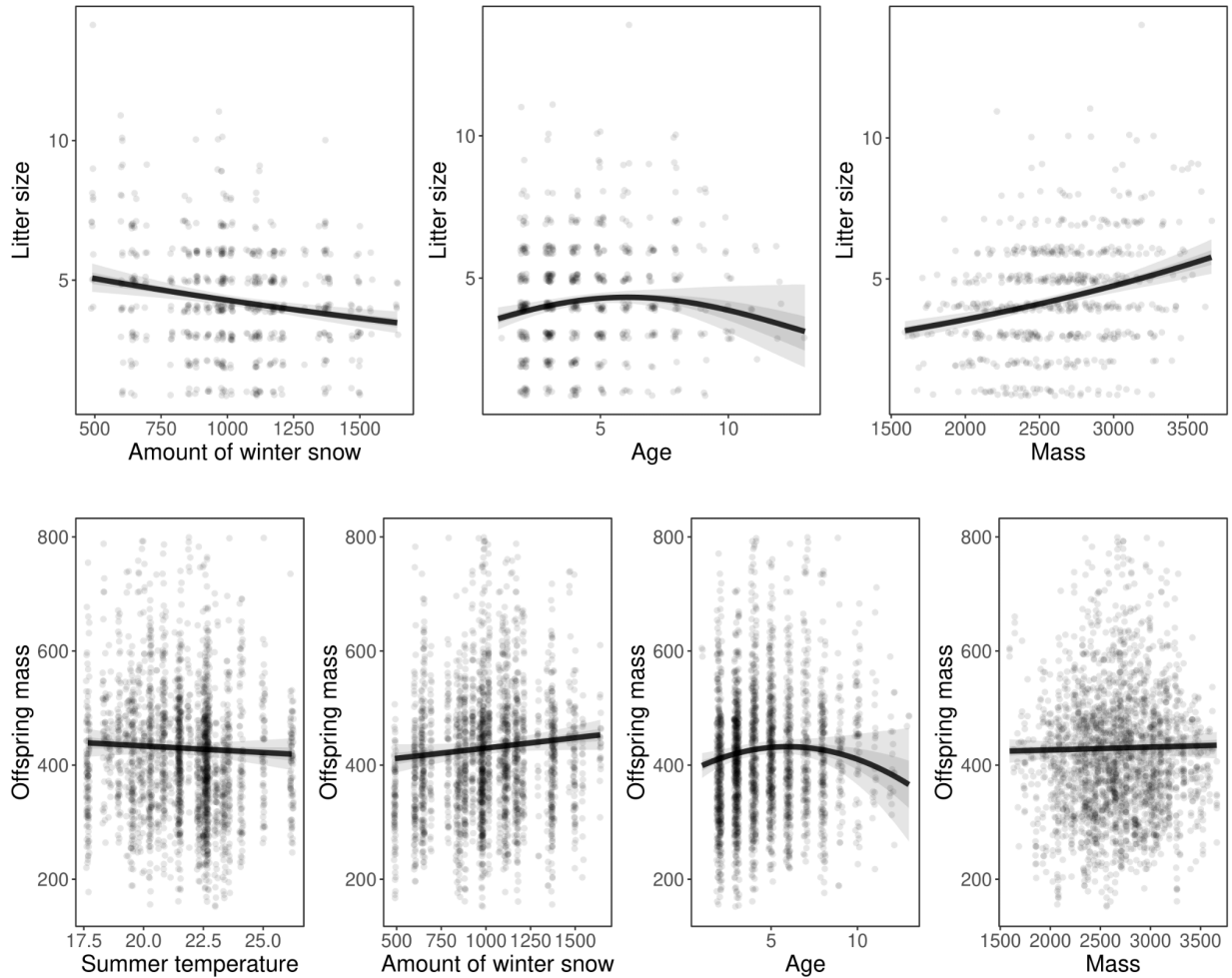
694 Figure S3: Posterior predictive checks showing the concordance between the distribution of the
695 data (y) and the distribution of data generated under the statistical model (y_{rep}), for litter size
696 (left panel) and offspring mass (right panel). This highlights a good fit for the litter size model. It
697 also highlights that there is a slight overdispersion in offspring mass that is not accounted for by
698 the model.



699

700 Figure S4: Posterior predictive checks showing the concordance between the distribution of the
701 data (y) and the distribution of data generated under the statistical model (y_{rep}), for number of
702 offspring (left panel) and ewe's mass in the following summer (right panel). This highlights a good
703 fit for both the litter size and mass models.

704 Section S3: associations between the covariates and traits studied



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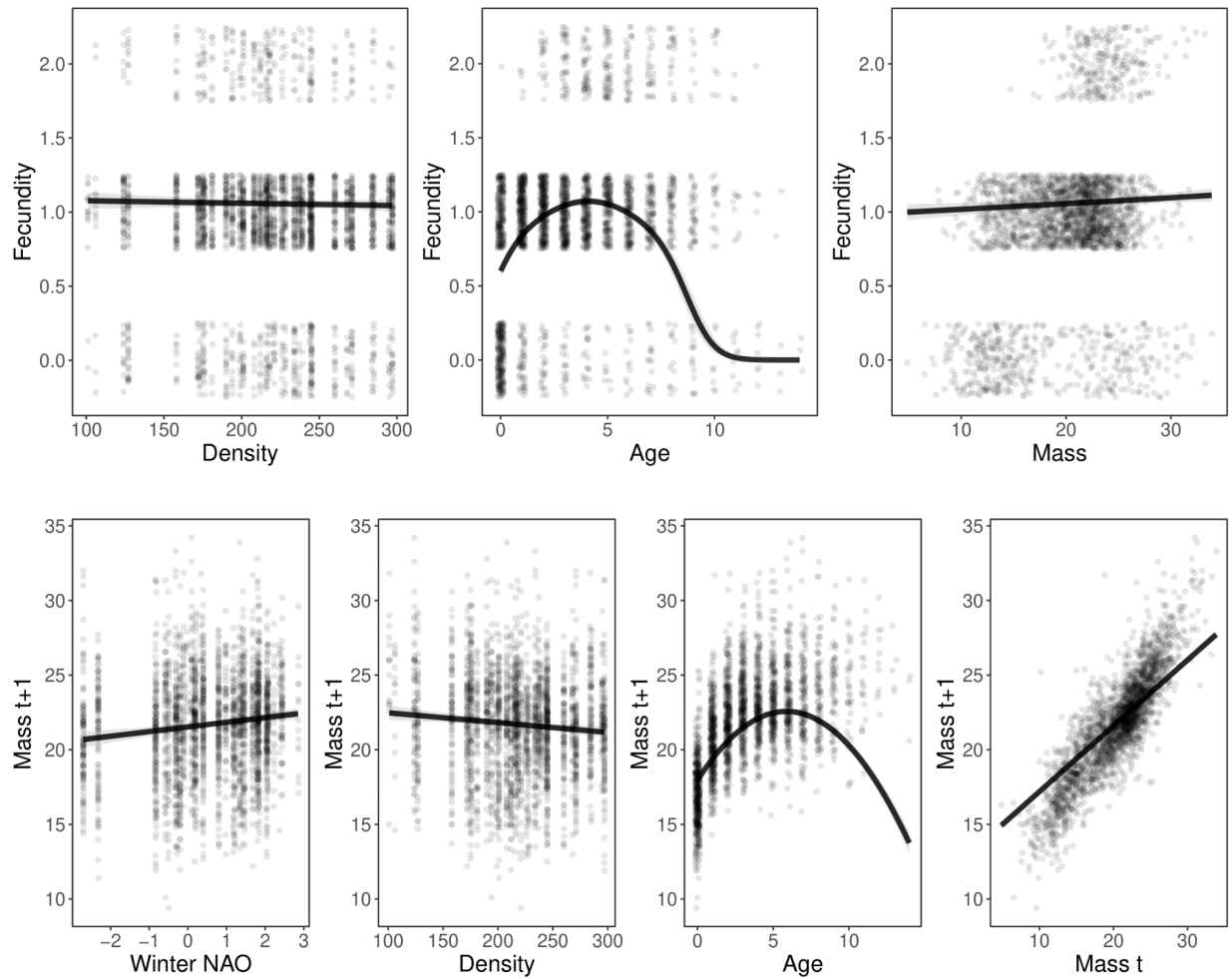
706 Figure S5: Top row: Association estimated by the model between the amount of winter snow,

707 age, and mass (panels from left to right) with litter size. Bottom row: Association estimated by

708 the model between summer temperature, the amount of winter snow, age, and mass (panels

709 from left to right) with offspring mass.

710



711

712 Figure S6: Top row: Association estimated by the model between population density, age, and

713 mass (panels from left to right) with fecundity. Bottom row: Association estimated by the model

714 between winter NAO, population density, age, and mass (panels from left to right) with mass at

715 t+1.

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