

1 **Detecting context-dependence in the expression of demographic tradeoffs**

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Abstract

Life history tradeoffs are one of the central tenets of evolutionary demography. Tradeoffs, depicting negative phenotypic or genetic covariances between individuals' demographic rates, arise from a finite amount of resources that each individual has to allocate in a zero-sum game between somatic and reproductive functions. While theory predicts that tradeoffs are ubiquitous, empirical studies have often failed to detect such negative covariances in wild populations. One way to improve the detection of tradeoffs is by accounting for the environmental context, as tradeoff expression may depend on environmental conditions. However, current methodologies usually search for fixed covariances between traits, thereby ignoring their context dependence. Here, we present a hierarchical multivariate 'covariance reaction norm' model, adapted to help detect context dependence in the expression of demographic tradeoffs. The method allows continuous variation in the phenotypic correlation between traits. We validate the model on simulated data for both intraindividual and intergenerational tradeoffs. We then apply it to empirical datasets of yellow-bellied marmots (*Marmota flaviventer*) and Soay sheep (*Ovis aries*) as a proof-of-concept showing that new insights can be gained by applying our methodology, such as detecting tradeoffs only in specific environments. We discuss its potential for application to many of the existing long-term demographic datasets and how it could improve our understanding of tradeoff expression in particular, and life-history theory in general.

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Introduction

46 Demographic tradeoffs, which are characterized as negative covariances between fitness
47 components such as somatic or reproductive traits, are central to life-history theory (Stearns,
48 1989), and are thought to drive much of the life-history diversity that exists (Bielby et al., 2007;
49 Healy et al., 2019; Salguero-Gómez et al., 2016; Stearns, 1984). They originate from the basic fact
50 that the total amount of resources or energy acquired by any one individual is limited, and has
51 to be shared among several of the individual's fitness-related traits. In such a zero-sum game and
52 in the absence of change in the total amount of resources acquired, any increase in the allocation
53 of resources towards a specific fitness component will have to be at the expense of another
54 fitness component. If tradeoffs did not exist, selection would maximize all fitness-related traits
55 simultaneously and would lead to the impossible "darwinian demons" (Law, 1979). Therefore,
56 demographic tradeoffs should be faced by all organisms and are, in theory, ubiquitous (Stearns,
57 1989, 1992; Williams, 1966). They can come in several forms (Stearns, 1989), being either
58 intraindividual (traits involved relate to the fitness of the same individual) or intergenerational
59 (traits involved relate to the fitness of a parent-offspring pair; e.g., offspring quantity-quality
60 tradeoff). Despite their expected universality and being sought-after by evolutionary ecologists
61 and biodemographers alike, life-history tradeoffs have been surprisingly hard to detect in wild
62 populations (Metcalf, 2016), with successful probes too often confined to experimental
63 approaches.

64 Several reasons could explain why tradeoffs are hard to detect in wild populations. First,
65 while we often expect traits to covary in a simple bivariate manner by analyzing a single pair of
66 demographic rates (note that demographic or vital rates, phenotypes, traits, fitness components

67 can all be used interchangeably) at a time, tradeoff structures are often more complex. For
68 instance, many more than two traits are likely to be involved in the resource allocation process
69 (Cressler et al., 2017; de Jong, 1993; Pease & Bull, 1988), sometimes leading to complex
70 hierarchical allocation trees, potentially resulting in some pairs of traits not covarying negatively
71 (Gascoigne et al., 2022). Second, life-history traits can covary at different levels. While tradeoffs
72 result from individuals' resource allocation processes, biodemographers often study tradeoffs as
73 the temporal correlations among demographic rates at the population level (Compagnoni et al.,
74 2016; Fay et al., 2020; Fay, Hamel, et al., 2022; van Tienderen, 1995). Tradeoffs can occasionally
75 scale up to cause negative temporal covariances at the population level (van Tienderen, 1995),
76 but in most cases these covariances are the results of environmental stochasticity and
77 demographic reaction norms to shared ecological drivers (Fay, Hamel, et al., 2022; Knops et al.,
78 2007; Paniw et al., 2020). Third, even though tradeoffs might be intrinsically present, individual
79 heterogeneity can mask their presence among individuals. This specific ecological version of
80 Simpson's paradox (Simpson, 1951) has been demonstrated by van Noordwijk and de Jong
81 (1986): when the among-individual variance in resource acquisition is greater than the among-
82 individual variance in resource allocation, the tradeoff is not expressed among individuals — even
83 though it is theoretically present within individuals. In addition, expression of a tradeoff among
84 individuals can also be influenced if the allocation and acquisition processes are not independent
85 (Descamps et al., 2016; Fischer et al., 2009; Robinson & Beckerman, 2013). Altogether, this makes
86 the detection of tradeoffs in wild populations difficult.

87 How much individuals vary in acquisition and allocation of resources determines if a
88 tradeoff is detected among individuals (Metcalf, 2016; Reznick et al., 2000; van Noordwijk & de

89 Jong, 1986). Part of this variance might be fixed, stemming from genetic, developmental, or
90 consistent behavioral differences that constrain how much resources are acquired and allocated
91 to somatic vs. reproductive functions (Réale et al., 2007; Wilson & Nussey, 2010). The remaining
92 variance is likely to be plastic (Spigler & Woodard, 2019), where investment in acquisition vs.
93 allocation likely depends on the environmental context (Cohen et al., 2020; Sgrò & Hoffmann,
94 2004; Stearns et al., 1991). For instance, in several species, no tradeoffs were found among
95 captive animals fed *ad libitum* (Kengeri et al., 2013; Landes et al., 2019; Ricklefs & Cadena, 2007).
96 Similarly, controlled laboratory experiments on several species have shown that tradeoffs
97 detection and strength were dependent on resource abundance (Gebhardt & Stearns, 1988;
98 Messina & Fry, 2003; Messina & Slade, 1999; Spigler & Woodard, 2019). However, despite
99 evidence that tradeoff expression depends on the environmental context, statistical methods to
100 detect this context dependence in wild populations have, to date, rarely been applied.

101 Multivariate models are commonly employed to detect tradeoffs in wild populations
102 (Cam et al. 2002, 2013; Hamel et al. 2018; Paterson et al. 2018; Fay et al. 2022a). In quantitative
103 genetics, such models allow for the simultaneous analysis of multiple dependent variables like
104 fecundity, growth, and survival (Kruuk et al. 2008; Wilson et al. 2010). These variables each have
105 their own predictors, and the models estimate the correlated residual variances unaccounted for
106 by the primary predictors. These models can be used to study residual correlations between traits
107 at different levels, such as among-year or among-individual correlations. For example, after
108 accounting for primary predictors, such models quantify whether years with high survival in a
109 population are also years with high recruitment; or whether individuals with higher fecundity
110 have lower or higher growth rates. However, these correlations among residual variances are

111 estimated as fixed. Estimating fixed correlations might not necessarily be problematic in the case
112 of experimental work, in which environmental conditions can be held constant within each
113 treatment. However, wild populations are unlikely to experience fixed conditions, as the
114 environmental context will vary in a continuous fashion, hence influencing the expression of
115 tradeoffs. Therefore, there is a need to analyse and predict continuous variation of phenotypic
116 correlations.

117 Here, we repurpose a hierarchical multivariate ‘covariance reaction norm’ (hereafter
118 CRN) model recently developed by Martin (2023), which allows the incorporation of continuous
119 predictors directly on the covariance matrix, for application to sampling designs typical in
120 population ecology, enabling the study of the context-dependent expression of tradeoffs. As a
121 proof-of-concept, we first validate this model on two simulated datasets, respectively focusing
122 on an intergenerational tradeoff and an intraindividual tradeoff. We then apply our model on
123 two empirical datasets of wild populations of yellow-bellied marmots *Marmota flaviventris* and
124 Soay sheep *Ovis aries*. Prior studies have explored tradeoffs between demographic rates in both
125 species (Kroeger et al., 2020; Tavecchia et al., 2005). For instance, in yellow-bellied marmots, a
126 quality-quantity tradeoff in offspring has been observed for older mothers. In Soay sheep, the
127 costs of reproduction have been particularly evident for breeding ewes in high-density
128 populations or following harsh winters. However, the environmental context-dependence of
129 these tradeoffs has yet to be studied explicitly. In the marmots, which inhabit high-altitude, highly
130 seasonal environments, and the sheep, which face severe winter storms and fluctuating
131 population densities, we hypothesize that tradeoffs are more likely to manifest under
132 unfavorable ecological conditions (Cohen et al., 2020; Sgrò & Hoffmann, 2004).

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Methods

134 The model

135 In this study, we employ a newly introduced CRN model (Martin, 2023), which has been initially
136 developed to predict continuous changes in trait associations when genetic data or repeated
137 individual measurements are available. A key assumption of multivariate models thus far has
138 been that phenotypic correlations caused by demographic tradeoffs are fixed through time or
139 space (Cam et al., 2002; Hamel et al., 2018). The CRN approach provides a solution to this general
140 challenge, by allowing for phenotypic covariances to vary in response to variation in the
141 environment. In the present study, we extend application of this general CRN approach to the
142 detection of context-dependent demographic tradeoffs between life-history traits, with special
143 consideration to sampling conditions typical of long-term field research in population ecology.
144 Specifically, we examine the use of bivariate CRN models to test for the presence of tradeoffs
145 when genetic data or repeated individual measurements are lacking in a given environmental
146 context (e.g., during a specific sampling event such as a breeding season or a year). These are
147 typical situations in field research that motivate further refinement of the quantitative genetic
148 models proposed by Martin (2023).

149 Consider a CRN model investigating how environmental contexts \mathbf{X} (an $N \times P$ matrix of N
150 measurements of P predictors) affect the phenotypic means of β_μ and among-individual
151 correlations β_r between two Gaussian life history trait measures \mathbf{z}_1 and \mathbf{z}_2 with repeated
152 individual measurements in each environmental context. We begin by focusing on linear models
153 to simplify notation and aid comprehension, with generalized models for non-Gaussian

154 distributions discussed further below. Following Martin (2023) in the absence of genetic data,
 155 our bivariate phenotypic model is given by

$$\begin{aligned} \mathbf{z}_1 &= \mathbf{X} \boldsymbol{\beta}_{\mu 1} + \mathbf{Y} \boldsymbol{\alpha}_{1(X)} + \boldsymbol{\epsilon}_{1(X)} \\ \mathbf{z}_2 &= \mathbf{X} \boldsymbol{\beta}_{\mu 2} + \mathbf{Y} \boldsymbol{\alpha}_{2(X)} + \boldsymbol{\epsilon}_{2(X)} \end{aligned} \quad (1.1)$$

$$\begin{aligned} [\boldsymbol{\alpha}_{1(X)}, \boldsymbol{\alpha}_{2(X)}] &\sim N(\mathbf{0}, \mathbf{P}_{(X)}) \\ [\boldsymbol{\epsilon}_1, \boldsymbol{\epsilon}_2] &\sim N(\mathbf{0}, \boldsymbol{\Sigma}_{(X)}) \end{aligned}$$

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157 Trait values are expressed as a function of the average effects $\beta_{\mu 1}$ and $\beta_{\mu 2}$ of \mathbf{X} on each
 158 phenotype, as well as among-individual effects $\alpha_{1(X)}$ and $\alpha_{2(X)}$ that are repeatable across
 159 measurements and within-individual effects $\epsilon_{1(X)}$ and $\epsilon_{2(X)}$ that are stochastic across
 160 measurements. The model matrix \mathbf{Y} (an $N \times J$ matrix for J subjects) structures the among-
 161 individual effects $\alpha_{(X)}$ across repeated measurements. (Co)variances between independent
 162 among- and within-individual effects are respectively described by \mathbf{P} and $\boldsymbol{\Sigma}$ covariance matrices.
 163 To detect context-dependent tradeoff expression, we use environmental information in \mathbf{X} to
 164 predict the among-individual trait covariance matrix $\mathbf{P}_{(X)}$.

$$\mathbf{P}_{(X)} = \begin{bmatrix} \sigma_{\alpha_1(X)}^2 & r_{\alpha(X)} \sigma_{\alpha_1(X)} \sigma_{\alpha_2(X)} \\ r_{\alpha(X)} \sigma_{\alpha_2(X)} \sigma_{\alpha_1(X)} & \sigma_{\alpha_2(X)}^2 \end{bmatrix} \quad (1.2)$$

$$\text{atanh}(r_{\alpha(X)}) = \mathbf{X} \boldsymbol{\beta}_r$$

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166 where the inverse hyperbolic tangent function $\text{atanh}(r) = \text{logit}([r + 1]/2)/2$ is used as a link
 167 function to model additive environmental effects β_r on the logit scale while retaining the $[-1, 1]$
 168 scaling of the correlation coefficient r . The same approach can be taken to describe changes in
 169 within-individual variation across environmental contexts.

$$\mathbf{\Sigma}(X) = \begin{bmatrix} \sigma_{\epsilon_1(X)}^2 & r_{\epsilon(X)} \sigma_{\epsilon_1(X)} \sigma_{\alpha_2(X)} \\ r_{\epsilon(X)} \sigma_{\epsilon_2(X)} \sigma_{\epsilon_1(X)} & \sigma_{\epsilon_2(X)}^2 \end{bmatrix} \quad (1.3)$$

$$\operatorname{atanh}(r_{\epsilon(X)}) = \mathbf{X} \boldsymbol{\beta}_{r_{\epsilon}}$$

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Direct prediction of the transformed correlation coefficient is useful because we are principally interested in $r_{(X)}$ as an indicator of putative within- or among-individual tradeoffs, rather than the covariance $P_{1,2(X)} = r_{(X)}\sigma_1\sigma_2$ per se. Changes in the scale $\sigma_1\sigma_2$ of life history trait variation may occur independently of changes in positive or negative trait association among individuals, but these effects will be confounded together in the covariance $P_{1,2(X)}$. In contrast, the correlation coefficient $r_{(X)}$ is standardized relative to the scale of each phenotype, providing a more robust quantity for directly predicting and comparing estimates of life history tradeoffs across phenotypes and species. Our model also assumes that phenotypic variances can vary across environmental contexts, but no predictions are made on this variation. Greater plasticity is instead expected in the strength of tradeoff expression caused by fluctuating environmental factors (e.g., environmental harshness, resource availability, local predator density). See Martin (2023) for further details on relaxing these assumptions to model environmental effects on among- and within-individual variances.

Non-repeated measures

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Estimating **Eq 1** with empirical data requires multiple measurements of the same subjects across time to effectively partition trait correlations due to sources of among- $\mathbf{P}_{(X)}$ and within-individual $\mathbf{\Sigma}$ phenotypic variation, relative to a given window of sampling. Repeated individual measurements are often inconsistent or unavailable in a given environmental context (e.g., a single fecundity measurement for individuals in a given year) in long-term field studies, which

190 otherwise provide invaluable datasets for investigating context-specific tradeoffs in the wild.
 191 Fortunately, we can still take advantage of long-term environmental variation in such studies to
 192 detect variation in tradeoff expression without repeated measurements in a given environmental
 193 context. This requires simplifying the CRN model to predict observation-level phenotypic
 194 associations across environmental contexts.

$$\begin{aligned} \mathbf{z}_1 &= \mathbf{X} \boldsymbol{\beta}_{\mu 1} + \mathbf{o}_1(X) \\ \mathbf{z}_2 &= \mathbf{X} \boldsymbol{\beta}_{\mu 2} + \mathbf{o}_2(X) \end{aligned} \quad (2)$$

$$\begin{aligned} [\mathbf{o}_1(X), \mathbf{o}_2(X)] &\sim N(\mathbf{0}, \mathbf{P}_{o(X)}) \\ \mathbf{P}_{o(X)} &= \begin{bmatrix} \sigma_{o_1(X)}^2 & r_{o(X)} \sigma_{o_1(X)} \sigma_{o_2(X)} \\ r_{o(X)} \sigma_{o_2(X)} \sigma_{o_1(X)} & \sigma_{o_2(X)}^2 \end{bmatrix} \\ \text{atanh}(r_{o(X)}) &= \mathbf{X} \boldsymbol{\beta}_r \end{aligned}$$

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 196 Here $o_{1(X)} = a_{1(X)} + \epsilon_{1(X)}$ and $o_{2(X)} = a_{2(X)} + \epsilon_{2(X)}$ are observation-level random effects
 197 aggregating variation due to among- and within-individual differences across measurements,
 198 within a given environmental context defined by \mathbf{X} (e.g., a given year, position in space, level of
 199 resource abundance). Note that the \mathbf{Y} matrix from **Eq 1** is no longer necessary in the absence of
 200 repeated measurements. As a consequence, we expect that the observation-level correlation
 201 $r_{o(X)}$ between these random effects to reflect the combined effect of the among- and within-
 202 individual correlations between life history traits, weighted by their geometric mean
 203 repeatability R (Dingemanse & Dochtermann, 2013; Searle, 1961).

$$r_{o(X)} = r_{\alpha(X)} \sqrt{\frac{\sigma_{\alpha 1}^2 \sigma_{\alpha 2}^2}{\sigma_{z 1}^2 \sigma_{z 2}^2}} + r_{\epsilon(X)} \sqrt{\frac{\sigma_{\epsilon 1}^2 \sigma_{\epsilon 2}^2}{\sigma_{z 1}^2 \sigma_{z 2}^2}} = r_{\alpha(X)} \sqrt{R_1 R_2} + r_{\epsilon(X)} \sqrt{(1 - R_1)(1 - R_2)} \quad (3)$$

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205 Where phenotypic variances are adjusted for the mean effects of $X\beta_{\mu 1}$ and $X\beta_{\mu 2}$. We can see
206 that inferences about among-individual tradeoffs from the non-repeated measures model (Eq.
207 2) will be at greatest risk of bias when $\text{sign}(r_{\alpha}) \neq \text{sign}(r_{\epsilon})$ and $\sqrt{R_1 R_2} \ll \sqrt{(1 - R_1)(1 - R_2)}$.
208 Figure 1 shows these general relationships across correlation and repeatability ranges, identifying
209 regions of sign bias. Fortunately, researchers will generally be able to judge their risk of inferential
210 bias based on *a priori* knowledge about the repeatability of life history traits, which tends to be
211 medium to high (Dingemanse et al., 2021). For example, observation-level correlations of
212 behavioral traits will tend to be dominated by within-individual associations (Bell et al., 2009;
213 Cauchoux et al., 2018; Holtmann et al., 2017), while morphological associations will tend to be
214 dominated by among-individual variation (Dingemanse et al., 2021). Note that our model
215 considers no measurement errors, as we are not able to disentangle it from true within-individual
216 variation using non-repeated measures. Such considerations regarding trait repeatability and
217 measurement error should be explicit when interpreting results without repeated measures.

218 ***Hybrid scenarios***

219 Variation in repeated sampling is also likely to occur across phenotypes due to factors such as
220 difficulty of measurement and the rate of trait expression. While a single measure of age at first
221 reproduction or fecundity in a given environmental context may be available per individual,
222 multiple individual measures may be available for traits such as offspring quality. Such scenarios
223 require a hybrid modeling approach. For example, consider a model for an intergenerational
224 tradeoff between fecundity (e.g., clutch size) and offspring quality, but other traits could equally
225 be studied. The model structure for offspring quality (depicted as offspring body mass), a
226 gaussian trait, is given by

$$mass = X\beta_{\mu 1} + Y\alpha_{1(X)} + \epsilon_{1(X)} \quad (4.1)$$

227

228 The linear predictor for the mass of an offspring of a given mother in year X includes a year-
 229 specific mother random effect $a_{1(X)}$ and $\epsilon_{1(X)}$ being the within-brood/litter variance.

230 The model for fecundity follows the same basic structure, but using a Poisson distribution where
 231 we model the expected rate of offspring production using a log link function.

$$\log(fecundity) = X\beta_{\mu 2} + o_{2(X)} \quad (4.2)$$

232

233 Without repeated measures, the random effect $o_{2(X)}$ is specified at the observation-level,
 234 accounting for any overdispersion in the Poisson process across measurements of each female.

235 The context-dependent tradeoff will be estimated between the among-mother random effect in
 236 offspring quality and the observation-level random effect in fecundity.

$$[\alpha_{1(X)}, o_{2(X)}] \sim N(\mathbf{0}, \mathbf{P}_{(X)}) \quad (4.3)$$

$$\mathbf{P}_{(X)} = \begin{bmatrix} \sigma_{\alpha_{1(X)}}^2 & r_{(X)} \sigma_{\alpha_{1(X)}} \sigma_{o_{2(X)}} \\ r_{(X)} \sigma_{\alpha_{1(X)}} \sigma_{o_{2(X)}} & \sigma_{o_{2(X)}}^2 \end{bmatrix}$$

$$\text{atanh}(r_{(X)}) = X\beta_r$$

237

238 Reducing **Eq. 3**, the correlation r_X between the individual- $a_{1(X)}$ and observation-level $o_{2(X)}$
 239 effects will necessarily be proportional to the among-individual correlation across life history
 240 traits.

$$r_{o(X)} = r_{\alpha(X)} \sqrt{\frac{\sigma_{\alpha 2}^2}{\sigma_{z2}^2}} = r_{\alpha(X)} \sqrt{R_2} \quad (5)$$

241

242 Note that this method does not allow the inclusion of non-continuous traits (e.g., Bernoulli traits)
 243 in the absence of repeated measurements within a given environmental context (e.g., a given

244 year). In addition, another limitation is that this model does not partition context-invariant
245 differences between individuals (fixed among-individual heterogeneity across environmental
246 contexts) from context-dependent among-individual variation.

247

248 Validation on simulated datasets

249 We validated the CRN model on two different types of tradeoffs, an intergenerational tradeoff
250 between offspring quantity and quality (hybrid CRN model), and an intraindividual tradeoff
251 between fecundity and parental growth (non-repeated measures CRN model). See Martin (2023)
252 for a more extensive simulation-based calibration of CRN models over a broad range of
253 parameter values.

254 ***Intergenerational tradeoff (offspring quantity-quality)***

255 We first focused on an intergenerational tradeoff between offspring quantity and quality (hybrid
256 CRN model). This quantity-quality tradeoff has been the focus of numerous studies since Lack's
257 pioneering work on bird clutch sizes (Lack, 1947). We simulate 30 years of individual-based data
258 in which 25 new individuals enter the population each year, reproduce with an average
259 clutch/litter size of 2.5, and then have a probability to survive to next year of 0.6. This yielded a
260 final simulated dataset of 750 individuals, totaling 1578 reproductive events and 4783 offspring.
261 An observation-level correlation was included between offspring mass and clutch size, and this
262 correlation was made dependent on a single climatic predictor. The same climatic predictor was
263 also included to influence both mean clutch size and offspring mass.

264 ***Intraindividual tradeoff (fecundity-growth)***

265 We then simulated data for an intraindividual tradeoff between fecundity and growth (non-
266 repeated measures CRN model). This simulated dataset is also made of 30 years and 750
267 individuals, for a total of 1974 reproductive events, with a variable observation-level correlation
268 between individual growth and fecundity, which is itself dependent on a single climatic predictor.

269

270 Study systems and application on empirical datasets

271 ***Marmots***

272 We applied the hybrid CRN model (one trait with repeated individual measurements within a
273 given year and one trait without) on data from a yellow-bellied marmot population monitored at
274 the Rocky Mountain Biological Laboratory in Gothic, Colorado (38°57'N, 106°59'W) during the
275 summer season each year, whereby extensive individual-based data is collected (Armitage, 2014;
276 Blumstein, 2013). In Alpine marmots *Marmota marmota*, an offspring quality-quantity tradeoff
277 has been found (Berger et al., 2015), while it remained mostly elusive in yellow-bellied marmots,
278 being only found for older mothers (Kroeger et al., 2020), whereby within-cohort selection has
279 likely reduced the amount of among-individual variance in resource acquisition, thus making the
280 tradeoff visible (Kendall et al., 2011; van Noordwijk & de Jong, 1986). Therefore, we searched for
281 an intergenerational tradeoff between mothers' fecundity and offspring estimated mass
282 (offspring quality-quantity tradeoff). We used repeated measurements of offspring mass for each
283 mother (one mass estimate for each offspring in a given litter). The offspring weaning mass was
284 imputed based on the date of emergence for each litter and mass measurements from captures
285 later in the season, following the method of Ozgul et al. (2010). For the observation-level
286 correlation sub-model, we included measures quantifying environmental conditions for the year.

287 First, the total amount of snow during the preceding winter, with years of little overwinter snow
288 considered harsher for marmots as it offers limited thermal insulation during the hibernation
289 (Barash, 1973; Cordes et al., 2020; Wells et al., 2022). Second, the average daily maximum
290 temperature during the month of June, with warmer summer temperatures considered
291 unfavorable conditions for marmots as they are prone to overheating, hence limiting the time
292 that can be allocated to foraging (Cordes et al., 2020; Krajick, 2004; Melcher et al., 1990). Note
293 that we used temperature in June and not July as commonly used in this system (Cordes et al.,
294 2020), because this is more likely to represent the conditions experienced for most offspring
295 before emergence and weaning. For the mass sub-model, we included as covariates the total
296 amount of snow during the winter, June average maximum temperature, age of the mother, and
297 mother's estimated mass in early June. We also included these parameters as covariates for the
298 fecundity sub-model, except June average maximum temperature, since it cannot affect
299 fecundity as pregnancies mostly occur before this period. We expected tradeoffs to be more
300 strongly expressed among individuals in years with little overwinter snow or high summer
301 temperature. In total, we used 2540 offspring mass from 597 reproductive events, from 279
302 females across 42 years.

303 ***Soay sheep***

304 We applied the non-repeated measures CRN model on Soay sheep data, as we have no repeated
305 individual measurement within a given year available for neither of the traits studied. We used
306 data from an unmanaged population of feral sheep in the Village Bay area of the island of Hirta
307 (57°48'N, 8°37'W), which has been monitored since 1985 (Clutton-Brock & Pemberton, 2004). In
308 Soay sheep, survival costs of reproduction were found for breeding ewes, particularly in

309 populations at high densities or following stormy winters (Tavecchia et al., 2005). Therefore, we
310 searched for an intraindividual tradeoff between ewes' fecundity defined as the number of lambs
311 born in Spring (ranging from 0 to 2) and their log mass in the following summer, with both traits
312 conditional on ewes surviving the winter. Regarding the observation-level correlation sub-model,
313 we included two covariates characterizing the ecological harshness faced by the sheep:
314 population density and winter NAO (North Atlantic Oscillation), with high NAO values
315 corresponding to wet and stormy winters (Coulson et al., 2001; Regan et al., 2022). For the ewe's
316 fecundity sub-model, we included as covariates the individual's log mass preceding the
317 reproductive event, age (linear term), and population density. For the sub-model on the log mass
318 in the following summer, we included the same covariates, as well as winter NAO. We expected
319 tradeoffs to be more strongly expressed in years of high population density or high NAO. In total,
320 we used data from 2497 reproductive events across 37 years, for 861 ewes with known mass in
321 the summer preceding the reproductive event, as well as known mass in the following summer.

322

323 Model implementation

324 We implemented all multivariate models described above in a Bayesian framework using the Stan
325 statistical language (Carpenter et al., 2017), through the software R (R Core Team, 2021) using
326 the R package *CmdStanR* (Gabry & Češnovar, 2020). Stan was preferred for model
327 implementation because of its flexibility. Common regularizing priors were used for all model
328 parameters: normal distributions of mean 0 and standard deviation of 1 for intercepts and slopes
329 coefficients, and exponential distributions of rate 2 for variance parameters. Each model ran on
330 3 chains, with a burn-in period of 1000 iterations, sampling for 1000 iterations when analyzing

331 simulated data, and 3000 iterations when analyzing empirical data, keeping all the sampled
332 iterations (Link & Eaton, 2012). Convergence of parameter estimates was assessed visually and
333 using the Gelman-Rubin diagnostic (Gelman & Rubin, 1992). We report the full posterior
334 distributions, alongside their mean, 50%, and 89% credible intervals (McElreath, 2020). The Stan
335 code to implement all the CRN models presented in this study is archived on GitHub
336 (https://github.com/lbiard/detecting_tradeoffs_crn_models) and Zenodo (will be added upon
337 acceptance of the manuscript).

338

339

Results

340 The model validation performed on simulated datasets showed that parameters were correctly
341 recovered for both intergenerational tradeoffs (Figure 2) and intraindividual tradeoffs (Figure 3).
342 While these simulation examples do not quantify bias of estimations (more details from a
343 simulation-based calibration of CRN models are available in Martin (2023)), they still confirm that
344 the model presented in the methods is able to detect context-dependence in the expression of
345 tradeoffs.

346 The model applied to yellow-bellied marmot data shows trends towards tradeoffs being
347 more strongly expressed in years with harsh environmental conditions, albeit with high
348 uncertainty in the estimates (Figure 4). We found a positive mean effect of the amount of
349 overwinter snow on the correlation (Figure 4), meaning that the tradeoff between fecundity and
350 offspring quality was more strongly expressed after winters with little snow. We also found a
351 negative mean effect of the average maximum June temperature on the correlation (Figure 4),

352 where females with more offspring were more likely to have lighter offspring during warmer
353 summers.

354 Estimated effects of covariates on the correlation also had high uncertainty in the Soay
355 sheep dataset (Figure 5). Overall, we found that the correlation tended to be negative across all
356 environments, which means that ewe's growth was lower for the ones that weaned offspring
357 (Figure 5). Contrary to our expectations, while we hypothesized that the tradeoffs should be
358 more strongly expressed in wet and stormy winters (high NAO index), we found a positive effect
359 of winter NAO on the correlation between fecundity and growth (Figure 5). We did not find any
360 clear effect of population density on the expression of the tradeoff (Figure 5).

361

362

Discussion

363 Our proof-of-concept study demonstrates that the hierarchical multivariate CRN model (Martin,
364 2023) can be used successfully to detect and estimate context-dependent changes in tradeoff
365 expression. In agreement with theoretical predictions, we found that reproductive tradeoffs in
366 yellow-bellied marmots tend to be more strongly expressed under unfavorable climatic
367 conditions, even though uncertainty of the estimation was large. In Soay sheep, our proxies of
368 ecological harshness were not clearly linked with the expression of tradeoffs, and effect
369 directions were even opposite to our initial prediction. This hierarchical model has the potential
370 to be used on many long-term individual-based datasets and could help improve our
371 understanding of tradeoff expression and life-history theory.

372 Although the initial motivation to use this method partly rested on the observed difficulty
373 of finding tradeoffs in empirical datasets, we found that in both sheep and marmots, the

374 tradeoffs tend to be expressed across all environments, with mean phenotypic correlations being
375 negative regardless of the environmental conditions. Thus, ironically, in these two empirical
376 datasets, tradeoffs would have been detected using simpler multivariate methods without the
377 need for context dependence. However, this should not come as a surprise for Soay sheep, as
378 this negative correlation between growth and fecundity was already found on a smaller dataset
379 (Fung et al., 2022). Nonetheless, the results still highlight that context-dependence has the
380 potential to hinder our ability to detect tradeoffs in some cases. For instance, when marmots
381 experience favorable environmental conditions, the average correlation is closer to null with
382 credibility intervals nearing or overlapping zero (Figure 4), while the intergenerational tradeoff is
383 found to be more strongly expressed during harsh years. However, context dependence appears
384 to be limited or absent for the expression of the tradeoff in Soay sheep. Since ewes' mass is
385 measured in the following summer and not directly after parturition, harsh winter conditions are
386 expected to increase overwinter mortality (Milner et al., 1999), lowering spring population
387 density and reducing competition. This could potentially help surviving ewes to recover their
388 body condition between spring and summer, which is the period of greatest grass growth, hence
389 masking context dependence of the tradeoff. While this tradeoff might also just be invariant in
390 regard to the ecological context, we speculate that the result could have arisen from two
391 potential pitfalls due to idiosyncrasies of the Soay sheep data. First, among-individual variation
392 in fecundity is limited in sheep, ranging from no offspring to twins, potentially making it more
393 complicated for the model to estimate variances accurately (Fay, Authier, et al., 2022; Kain et al.,
394 2015). Second, both ewes' growth and fecundity are conditional on survival in the data, hence

395 individuals who suffered most from the cost of reproduction and did not survive are not present
396 in the analysis, potentially biasing the results.

397 Despite the potential of this modeling approach to study context-dependent tradeoffs, a
398 few methodological limitations are to be considered. A recent study conducted by Fay et al.
399 (2022) highlighted that multivariate models with correlated random effects for Bernoulli traits
400 performed rather poorly, resulting in a potentially large bias and imprecise estimates of variances
401 and covariances. This is in part because Bernoulli traits contain less information than continuous
402 variables, making estimations of variances complicated (Fay, Authier, et al., 2022), but also
403 because the data needed to estimate individual heterogeneity is usually scarce (Browne et al.,
404 2007). The model we present suffers from this limitation, and even more so when there is only a
405 single individual observation per individual per sampling occasion (e.g., parental survival). This
406 issue renders the model, as well as any other multilevel model, unable to meaningfully estimate
407 distinct mean and variance parameters for Bernoulli traits, due to the fact that the mean p of a
408 Bernoulli variable determines its variability $p(1-p)$ without scope for overdispersion. Therefore,
409 environmental effects on the mean of Bernoulli measures will necessarily change their variances
410 (Skrondal & Rabe-Hesketh, 2007). However, when repeated Bernoulli observations or a binomial
411 measure are available within each sampling occasion (e.g., survival of each offspring within a
412 litter), the CRN model can then be used to partition distinct environmental effects on trait means
413 and (co)variances. As we have shown in the present study, despite this limitation, the CRN
414 remains applicable to single measures of continuous traits and count measures (e.g., growth,
415 fecundity, phenology, behavioral traits), as well as proportions and various other kinds of non-
416 Gaussian measures. Another limitation of the proposed method is that sample sizes needed are

417 likely to be large, with enough individuals in each environmental context, and importantly
418 enough sampling occasions across which to estimate the context dependence of tradeoff
419 expression. Nonetheless, many long-term individual-based studies should have enough data to
420 fulfill these requirements (de Villemereuil et al., 2020).

421 Despite the abovementioned caveats and limitations of the methodology in the absence
422 of repeated measurements, this new model is a development that could be useful for many
423 datasets. Thanks to its implementation in a Bayesian framework using Stan (Carpenter et al.,
424 2017), it offers great flexibility and can be easily repurposed and modified to fit the idiosyncrasies
425 of a given dataset or species life history. It is also straightforward to extend the model by adding
426 a pedigree for quantitative genetic analysis (see Martin, 2023), even though phenotypic
427 correlations should suffice in most cases (Cheverud, 1988; Dochtermann, 2011; Roff, 1995).
428 While we presented a bivariate model, this model is not necessarily limited to two traits, and
429 more continuous traits and their covariances could also be analyzed simultaneously. We also
430 restricted our proof-of-concept study to the reaction norm of the correlation between traits, but
431 researchers interested in the canalization of traits variances as a response to the environmental
432 context could also benefit from this modeling approach (Péron et al., 2016).

433 Life-history tradeoffs have long been sought after, but difficult to detect in empirical data
434 due to individual heterogeneity (Metcalf, 2016; Reznick et al., 2000; van Noordwijk & de Jong,
435 1986). Previous studies have also highlighted that life-history tradeoffs could be expressed only
436 under unfavorable ecological conditions (Cohen et al., 2020; Stearns, 1989). Yet, despite our
437 knowledge of the issues hindering tradeoff detection, we still lacked a statistical framework that
438 permits the detection of context-dependence in tradeoff expression. Our proof-of-concept study

439 shows that this context dependence can be detected. This method has the potential to be applied
440 by demographers and evolutionary ecologists having long-term individual-based datasets at
441 hands, with many study systems having the required data (Culina et al., 2021; de Villemereuil et
442 al., 2020). Altogether, this method has the potential to help us improve our understanding of life
443 history theory, and in part resolve van Noordwijk and de Jong's (1986) conundrum of tradeoff
444 detection, by accounting for the context-dependence of their expression.

445

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451 D.T.B.).

452 **Authors contributions**

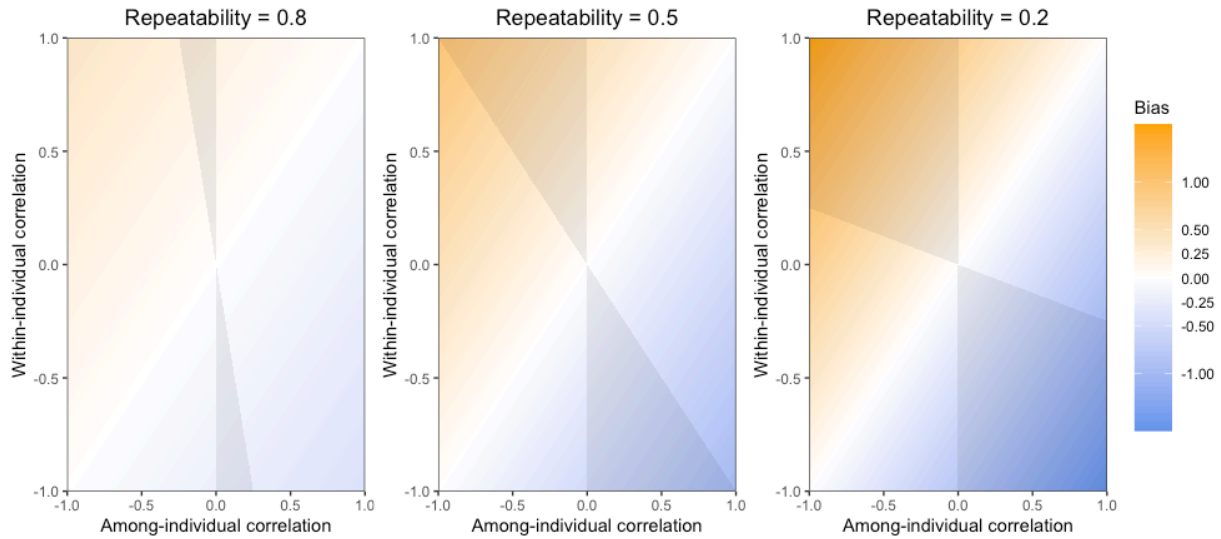
453 LB, JSM, AO, MP, DZC conceived the study. JSM designed the initial modeling framework and LB
454 analyzed the data. DTB, JGAM, JMP, DHN collected and curated the data. LB and JSM wrote the
455 first draft with input from AO, MP, DZC. All authors contributed to the editing of the manuscript.

456 **Data and code availability**

457 The data, as well as the R and Stan code necessary to reproduce the results are available on
458 GitHub (https://github.com/lbiard/detecting_tradeoffs_crn_models). They will be archived on
459 Zenodo upon acceptance.

460

461 **Figures**



462

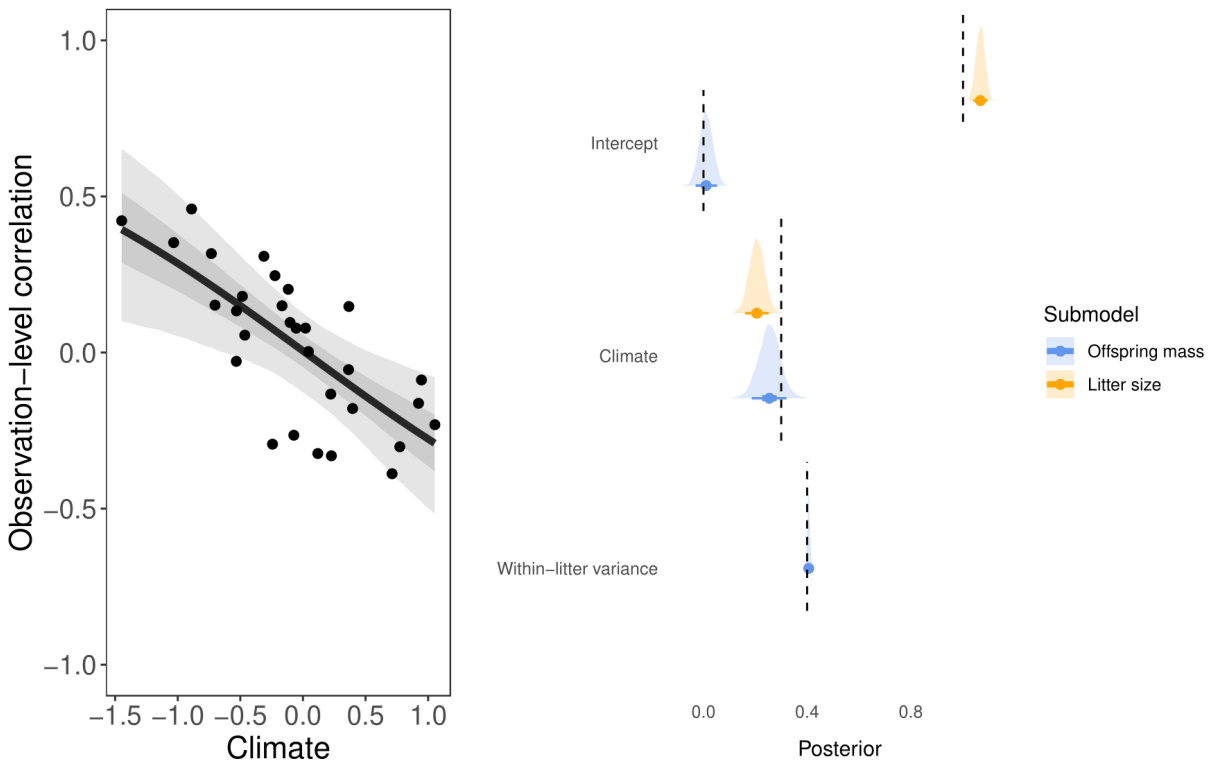
463 Figure 1: General relationships across correlations and repeatability ranges based on **Eq 3** for a
464 non-repeated measures CRN (model of **Eq 2**), identifying the magnitude of correlation bias and
465 the regions of sign bias. The bias is here defined as the difference between the observation-level
466 correlation and the among-individual correlation, using the latter as a reference. Parameter
467 spaces in gray represent the regions of sign bias, where the observation-level correlation has a
468 sign opposite to the among-individual correlation. This highlights that the observation-level
469 correlation is mostly influenced by the among-individual correlation for traits with high
470 repeatability, while it is mostly influenced by the within-individual correlation for traits with low
471 repeatability.

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477 Figure 2: Left panel: estimated vs. simulated observation-level correlation between litter size and

478 offspring mass as a function of climate, after accounting for the effect of climate on both traits.

479 The regression line indicates the mean effect of climate on the correlation, while the shaded

480 areas depict the 50% and 89% credible intervals predicted by the model. Each black dot

481 represents the simulated observation-level correlation between both traits in a given year

482 depending on climate. Right panel: estimated vs. simulated intercepts and slopes for the

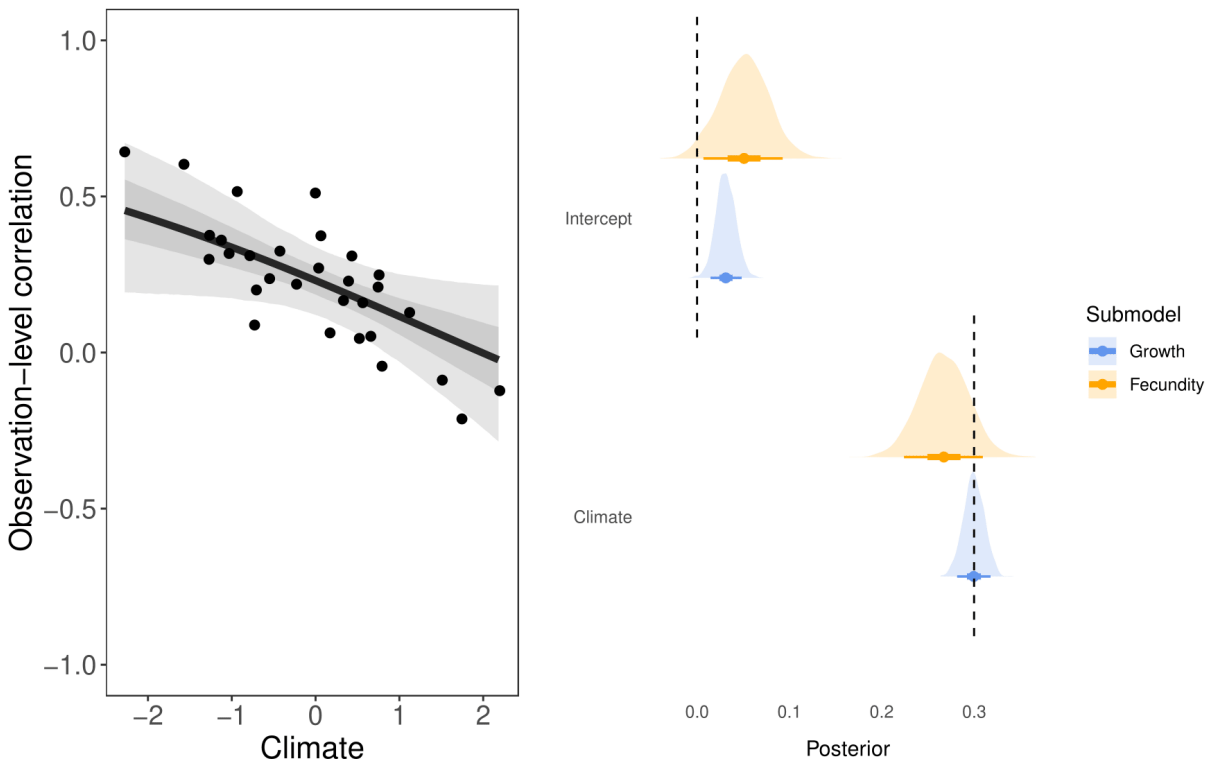
483 offspring mass and litter size sub-models. Dashed lines represent the value used to simulate the

484 data, while the distributions and intervals represent the posterior distributions estimated by the

485 model, alongside the mean, 50%, 89% credible intervals. Litter size estimates are presented on

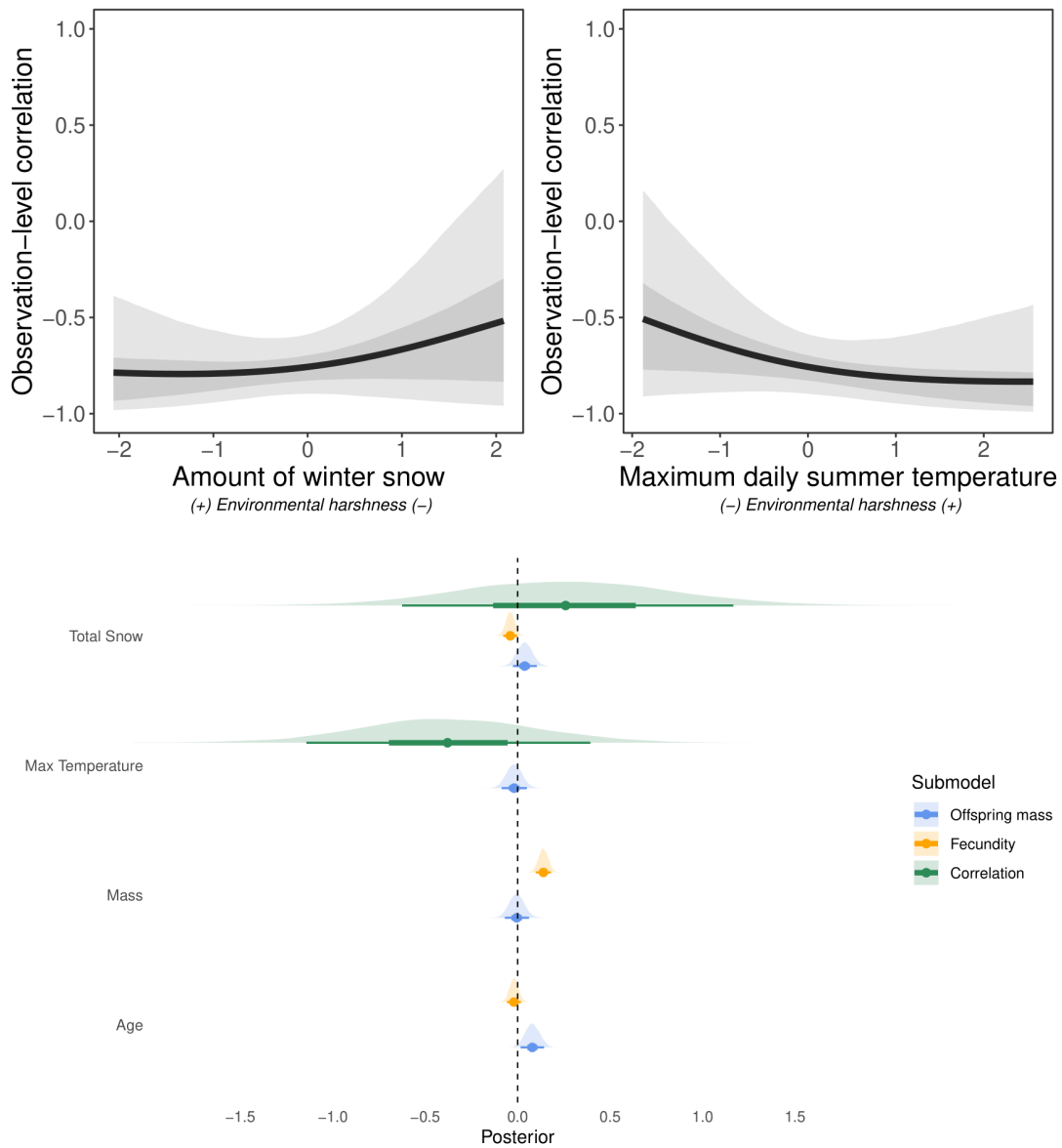
486 the log scale.

487



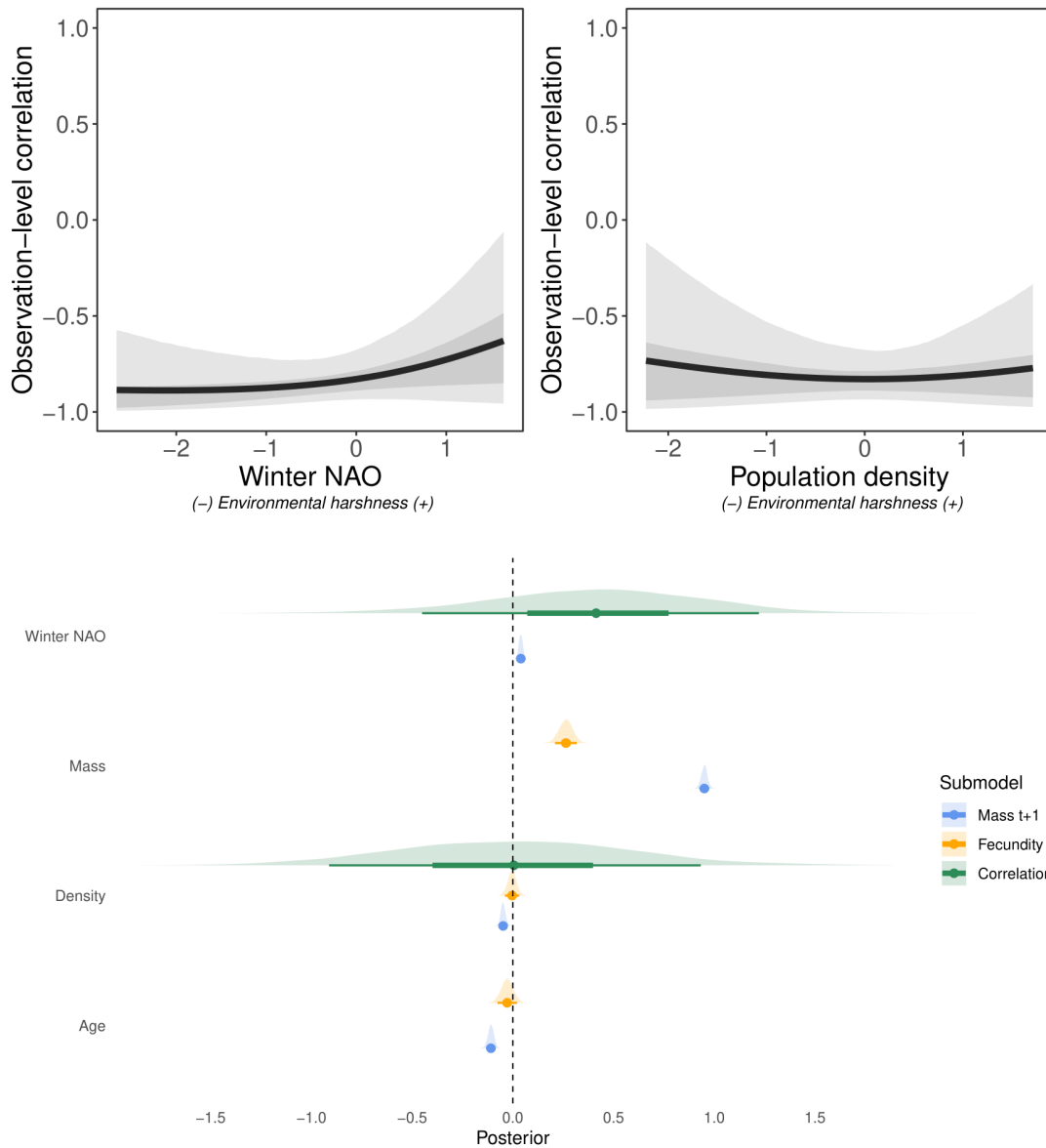
488

489 Figure 3: Left panel: estimated vs. simulated observation-level correlation between fecundity and
 490 growth as a function of climate, after accounting for the effect of climate on both traits. The
 491 regression line indicates the mean effect of climate on the correlation, while the shaded areas
 492 depict the 50% and 89% credible intervals predicted by the model. Each black dot represents the
 493 simulated observation-level correlation between both traits in a given year depending on climate.
 494 Right panel: estimated vs. simulated intercepts and slopes for the growth and fecundity sub-
 495 models. Dashed lines represent the value used to simulate the data, while the distributions and
 496 intervals represent the posterior distributions estimated by the model, alongside the mean, 50%,
 497 89% credible intervals. Fecundity estimates are presented on the log scale.



498

499 Figure 4: Observation-level correlation between litter size and offspring mass in marmots as a
 500 function of the total amount of snow in the preceding winter (top left panel) and the maximum
 501 daily June temperature of the year (top right panel). Estimated effects of standardized predictors
 502 (bottom panel) on offspring mass, fecundity, and the observation-level correlation between both
 503 traits in marmots. The regression line indicates the mean estimated effect, while the shaded
 504 areas depict the 50% and 89% credible intervals predicted by the model.



505

506 Figure 5: Observation-level correlation between fecundity and mothers' mass in the following

507 year in Soay sheep as a function of the winter NAO (top left panel) and the population density

508 (top right panel). Estimated effects (bottom panel) of standardized predictors on mother's mass

509 in the following year, fecundity, and the observation-level correlation between both traits in Soay

510 sheep. The figure displays the posterior distributions estimated by the model, alongside the

511 mean, 50%, and 89% credible intervals.

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