

1 **Multifaceted density dependence: Social structure and seasonality**
2 **effects on Serengeti lion demography**

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56 **Conflict of interest statement**

57

58 The authors declare no conflict of interest.

59

60 **Author Contributions**

61

62 **Eva Conquet:** Conceptualization, Methodology, Software, Validation, Formal
63 analysis, Data curation, Writing – original draft, Writing – review and editing,
64 Visualization.

65 **Maria Paniw:** Conceptualization, Writing – review and editing, Supervision, Funding
66 acquisition.

67 **Natalia Borrego:** Conceptualization, Investigation, Resources, Data curation,
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71 **Arpat Ozgul:** Conceptualization, Resources, Writing – review and editing,

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73

74 **Statement on inclusion**

75

76 Our study does not include scientists based in the country where the study was
77 carried out. We recognise that it is paramount to include the local scientific
78 community in our research and are planning to address these caveats in future
79 research wherever possible.

80

81 **Data and Code Availability Statement**

82

83 The processed data and MCMC samples necessary for reproducing results and
84 graphs presented in the study will be available on Zenodo [link placeholder] [citation
85 placeholder]. Original data can be requested from Craig Packer (packer@umn.edu).
86 Data and code for implementing and running models and analyses, and plotting
87 results is available on GitHub: <https://github.com/EvaCnqt/LionsDensity>. The version
88 of code used for this study will be archived on Zenodo [link placeholder] [citation
89 placeholder].

1 Abstract

2

3 1. Interactions between density and environmental conditions have important
4 effects on vital rates and consequently on population dynamics and can take
5 complex pathways in species whose demography is strongly influenced by
6 social context, such as the African lion, *Panthera leo*. In populations of such
7 species, the response of vital rates to density can vary depending on the
8 social structure (e.g., effects of group size or composition).

9 2. However, studies assessing density dependence in populations of lions and
10 other social species have seldom considered the effects of multiple socially-
11 explicit measures of density, and—more particularly for lions—of nomadic
12 males. Additionally, vital-rate responses to interactions between the
13 environment and various measures of density remain largely uninvestigated.

14 3. To fill these knowledge gaps, we aimed to understand how a socially- and
15 spatially-explicit consideration of density (i.e., at the local scale) and its
16 interaction with environmental seasonality affect vital rates of lions in the
17 Serengeti National Park, Tanzania. We used a Bayesian multistate capture-
18 recapture model and Bayesian GLMMs to estimate lion stage-specific survival
19 and between-stage transition rates, as well as reproduction probability and
20 recruitment, while testing for season-specific effects of density measures at
21 the group and home-range levels.

22 4. We found evidence for several such effects. For example, resident-male
23 survival increased more strongly with coalition size in the dry season
24 compared to the wet season and adult-female abundance affected subadult
25 survival negatively in the wet season, but positively in the dry season.

26 Additionally, while our models showed no effect of nomadic males on adult-
27 female survival, they revealed strong effects of nomads on key processes
28 such as reproduction and takeover dynamics.

29 5. Therefore, our results highlight the importance of accounting for seasonality
30 and social context when assessing the effects of density on vital rates of
31 Serengeti lions and of social species more generally.

32

33 Keywords: density dependence, density-environment interactions, sociality,
34 Bayesian models, multistate capture-recapture models, demographic rates

35

36 **Introduction**

37

38 Population dynamics are shaped by vital-rate responses to both density-dependent
39 and -independent (e.g., environmental) factors. Interactions between density and
40 environmental variables (hereafter environment-density interactions) occur across
41 many systems, with important consequences on populations dynamics (Coulson et
42 al. 2001; Gamelon et al. 2017). For example, density dependence can mediate the
43 effects of environmental factors through compensatory density feedbacks that can
44 buffer adverse environmental effects (e.g., through an increase in offspring survival
45 due to a lower competition following a decline in recruitment under reduced food
46 availability; Reed et al. 2013). By capping population abundances to a certain upper
47 threshold, density feedbacks can also exacerbate detrimental environmental effects
48 by exposing populations to demographic stochasticity (e.g., Jaatinen et al. 2021), or
49 even dampen positive effects of beneficial environmental conditions by constraining

50 populations to remain under that threshold even when the environment has strong
51 positive effects on vital rates (e.g., Layton-Matthews et al. 2020). In specific cases,
52 such as social species, density feedbacks can affect populations through complex
53 pathways, as vital rates can show strong responses to both intra- (e.g., number of
54 reproducing adults) and extra-group density factors (e.g., home range of a focal
55 group; Packer & Pusey 1983a; Maag et al. 2018; Behr et al. 2020), with contrasting
56 effects of such factors on vital rates of different social statuses (e.g., Paniw et al.
57 2019). While studies commonly assess the role of environment-density interactions
58 (e.g., Coulson et al. 2001; Gamelon et al. 2017), assessing vital-rate responses to
59 interactions between environmental conditions and several measures of density at
60 different scales could help obtain better insights on the role of density feedbacks in
61 shaping population demography.

62
63 Socially structured populations are often more susceptible to multifaceted density
64 effects (e.g., Behr et al. 2020; Ausband et al. 2021). Such complexity in density
65 effects on vital rates is likely to be at play in African lions, for which sociality has
66 strong effects on demography (Bygott et al. 1979; Packer & Pusey 1987; Elliot et al.
67 2014; Borrego et al. 2018). The African lion is therefore an ideal case study for
68 investigating the response of vital rates to density measures at different scales and
69 their interactions with the environment. Lion sociality is characterized by fission-
70 fusion dynamics with an egalitarian social structure represented by prides
71 (permanent, stable groups of females) and coalitions (permanent, stable groups of
72 males) (Schaller 1972; Packer 2023). Young males in the Serengeti system disperse
73 from their natal pride by four years of age and enter a nomadic phase during which
74 they band together with related or unrelated males to form coalitions of 1–9

75 individuals with no defined territory that can travel very long distances (Bygott et al.
76 1979; Packer & Pusey 1982; Hanby & Bygott 1987; Packer & Pusey 1987). Nomadic
77 males play a key role in shaping lion demography (Whitman et al. 2004; Borrego et
78 al. 2018). Male coalitions compete for access to prides; coalitions successfully taking
79 over a pride from a rival coalition gain reproductive benefits by killing the ousted
80 coalition's cubs (infanticide; Packer & Pusey 1983a; 1983b) and subsequently
81 mating with its females. Additionally, the newly resident males oust any subadult
82 males, who are sometimes too young to survive this forced dispersal (Elliot et al.
83 2014). Once they become resident, male coalitions typically remain with a pride for
84 2–3 years and often father only a single cohort. Takeover dynamics thus greatly
85 affect young survival (Bertram 1975; Elliot et al. 2014; Borrego et al. 2018) and
86 largely depend on the size of the coalition of resident males—who are the primary
87 defenders of a pride against rival males (Schaller 1972)—and on the size of the
88 challenging nomadic coalition (Bygott et al. 1979; Packer & Pusey 1987). Females
89 also take part in defending a pride against nomadic males, consequently decreasing
90 both young mortality and the probability of a successful takeover of a pride with
91 females living in groups compared to singletons (Grinnell & McComb 1996).
92 Successful takeovers also affect the reproductive status of females, who come into
93 oestrous and subsequently give birth synchronously (Bertram 1975). This synchrony
94 allows them to raise their cubs in crèches (Schaller 1972; Packer et al. 1990), where
95 cubs are better protected and have a higher survival rate (Bertram 1975). These
96 studies show the importance of socially-explicit density dependence in lion
97 populations but often focus on a single density measure (e.g., male coalition size or
98 number of females in a pride). However, we lack a comprehensive analysis of the
99 relative effects of various density measures on lion vital rates. Despite the decline in

100 the overall African lion population (Trinkel & Angelici 2016), the Serengeti population
101 is one of the few to remain apparently stable (Bauer et al. 2015; but see Riggio et al.
102 2016). A better understanding of the density-dependent drivers of vital-rate variation
103 in the Serengeti population could therefore benefit other lion populations as well as
104 social species beyond the African lion.

105

106 The effects of density on vital rates are typically mediated by environmental factors
107 (Courchamp et al. 1999; Paniw et al. 2019) but little is known about the response of
108 lion vital rates to interactions between density and environmental variables, such as
109 seasonal climatic patterns. Serengeti lions experience strong environmental
110 seasonality due to seasonal rainfall patterns driving prey availability (Norton-Griffiths
111 et al. 1975; Sinclair et al. 2013). These seasonal patterns in turn affect vital rates and
112 population dynamics. For instance, in wetter years, the increase in prey availability
113 favours recruitment through higher cub survival, leading to increases in the lion
114 population size (Packer et al. 2005). Additionally, Serengeti lions live in two distinct
115 habitats: the plains and the woodland, which are characterized by differences in
116 seasonal patterns of prey availability (Packer et al. 2005). Lions in the plains
117 experience strong decreases in prey availability during the dry season—when
118 migrating herds leave for the north. In the woodland, prey abundance (but not
119 composition) is relatively constant throughout the year, leading to higher lion density
120 (Hanby & Bygott 1979) and hence less opportunities for plain lions to settle in the
121 woodland. In a context of strong environmental seasonality, and under the predicted
122 important changes in seasonal patterns (IPCC 2014), understanding how season-
123 density interactions affect the vital rates of lions would provide more insights on how
124 density-dependent processes affect lion demography (Conquet et al. 2023) and

125 could ultimately benefit other social species living under strong environmental
126 periodicity.

127

128 To understand how different density-dependent variables affect seasonal lion vital
129 rates, we fitted a Bayesian multistate capture-recapture model and Bayesian
130 generalized linear mixed models (GLMMs) to data from a uniquely long monitoring
131 (30 years) of a population of African lions in the Serengeti to estimate season-
132 specific local density effects (as opposed to density at the population scale) by
133 assessing the response of lion survival, between-stage transition, and reproductive
134 rates (i.e., reproduction probability and recruitment) to socially- and spatially-explicit
135 density measures and to the habitat (plains or woodland). We used socially-explicit
136 density measures taken at the group level, more specifically the number of females
137 in a pride and the size of a resident or nomadic male coalition. For the spatially-
138 explicit effect of density, corresponding to density at the home-range level, we tested
139 for the effect of the number of nomadic coalitions in the home range of a pride or
140 resident coalition. Notably, ours is the first analysis to include multiple density
141 measures, including from nomadic males, in a multi-state African lion population
142 model. Considering the strong responses of vital rates of young lions to both season
143 and density, we expected the strongest seasonal effects of socially-explicit density
144 measures on young survival.

145 **Methods**

146

147 Study species

148

149 *Demographic data*

150

151 We used individual-based life-history data of 1347 lions (65 prides and 242 male
152 coalitions ranging size from 1–8 individuals), collected between 1984 and 2014 during
153 a consistent monitoring in a 2000-km² area located in the Serengeti National Park
154 (SNP), Tanzania (-2°27' N, 34°48' E) (Packer & Pusey 1987; Appendix S1). Starting
155 in 1984, one or two females per pride were equipped with VHF collars (VanderWaal
156 et al. 2009; Packer 2023). Each pride was then visited at least once every two weeks
157 by locating the collared females (VanderWaal et al. 2009; Borrego et al. 2018).
158 Additionally, lions or groups of lions away from their pride, as well as nomadic males,
159 were observed and recorded opportunistically during the monitoring. Lions were
160 identified by eye based on photographs of features such as scars and individual-
161 specific whisker spots recorded at the first sighting (Pennycuick & Rudnai 1970;
162 Packer & Pusey 1993). The age of individuals not observed as cubs was determined
163 from nose coloration, coat condition, and tooth wear (Whitman et al. 2004). Using
164 these natural markings allowed tracking of each individual from its birth (or entry into
165 the study area) until its death (or permanent emigration from the study area).
166 Additionally, while the death of most individuals could not be observed, we used dead-
167 recovery data available for 105 lions found dead from natural causes—i.e., not killed
168 by humans—opportunistically during the regular pride surveys to provide the model
169 with additional insights on the difference between mortality and lack of observation,

170 thereby better informing the survival process and obtaining more accurate survival
171 estimates.

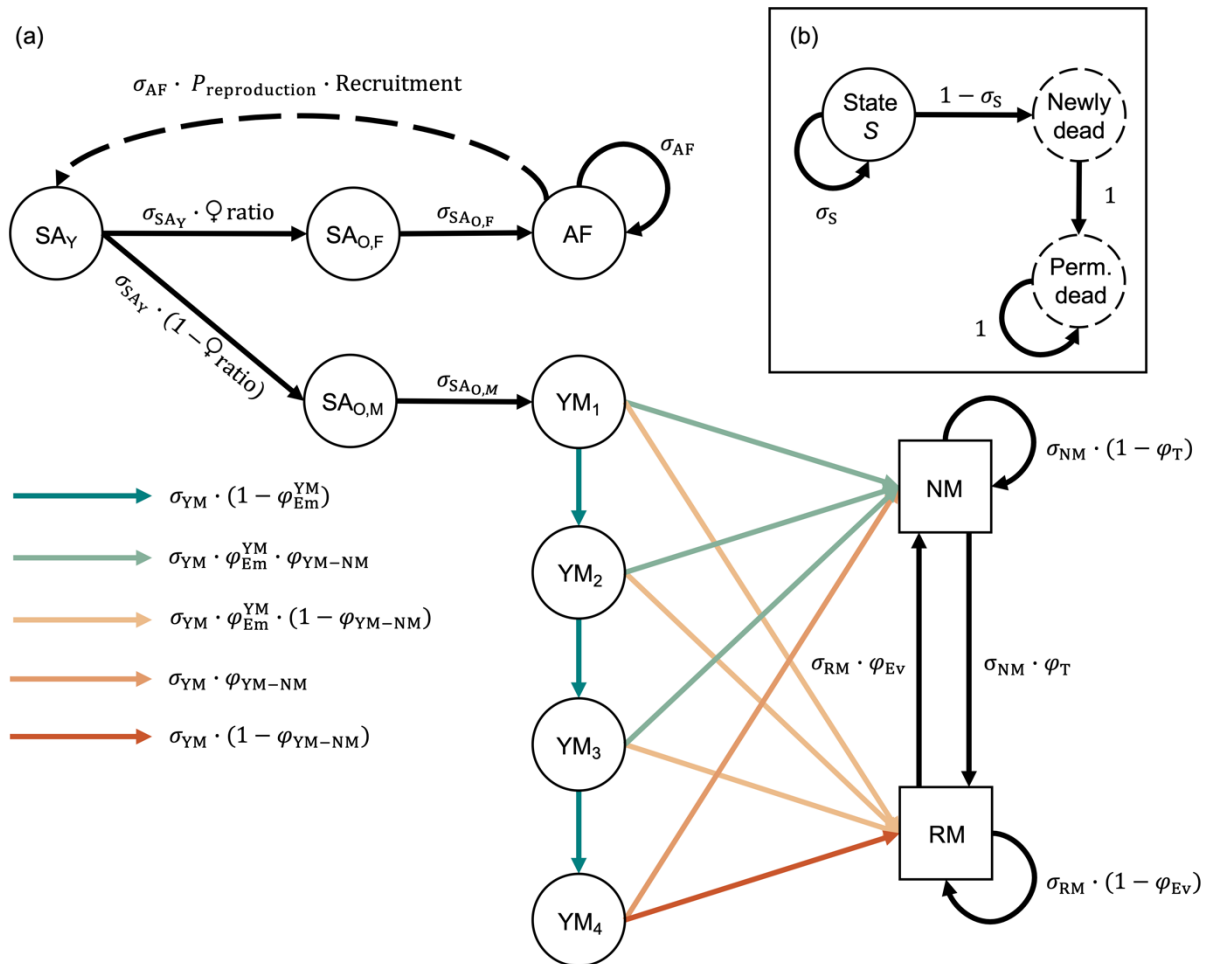
172

173 *Life history*

174

175 We divided the lion life history into 10 stages based on age, sex, and social status
176 (Fig. 1a). Subadults were divided into young subadults (SA_Y ; 1–1.5 years), and old
177 subadults (1.5–2 years), separated into females ($SA_{O,F}$) and males ($SA_{O,M}$). Female
178 subadults then become adult females (AF; > 2 years) in their natal pride. We
179 considered females to become adults at 2 years old; although females do not
180 necessarily reproduce at that age, their contribution to the pride is similar as that of
181 older females. In contrast, males could leave their natal pride as early as 2 years of
182 age but could also remain up to 4 years of age; males were considered as adults at
183 their departure from their natal pride. To represent males older than 2 years and still
184 in their natal pride and ensure they automatically left their natal pride after 4 years, we
185 used four young-male stages: YM_1 (2–2.5 years), YM_2 (2.5–3 years), YM_3 (3–3.5
186 years), and YM_4 (3.5–4 years). Finally, we divided males outside their natal pride
187 between two stages: nomadic male (NM; > 2 years and nomadic), and resident male
188 (RM; > 2 years and resident in a different pride). In the resulting life cycle (Figure 1a),
189 transitions between stages are all conditional on survival (σ). Additionally, transitions
190 from young subadult to female or male old subadult assume a fixed female-to-male
191 sex ratio of 0.55, representing a conservative value of the observed female-biased sex
192 ratio in the population (~ 0.60). Young males in stages YM_1 to YM_3 can leave their
193 natal pride conditional on emigration probability ϕ_{Em}^{YM} , while young males in YM_4
194 automatically leave their natal pride to become adult males. An emigrated young-male

195 can transition to either of the two adult-male stages (nomadic or resident) conditional
 196 on the probability of becoming nomadic (φ_{YM}). Nomadic and resident males then
 197 transition to the other adult male stage when respectively gaining (φ_T) or losing tenure
 198 of a pride (φ_{Ev}). Adult females recruit cubs conditional on their survival and
 199 reproduction probability ($P_{\text{reproduction}}$), and on the per-female number of cubs born in a
 200 given season that survived until their first birthday (Recruitment). Therefore, in our
 201 analysis, reproduction probability is not a component of recruitment and is estimated
 202 separately, with recruitment being conditional on reproducing.



203 **Figure 1 – Lion life cycle.** (a) The life cycle represents seasonal transitions
 204 between stages (solid arrows) and reproduction (dashed arrow); all transitions are
 205 conditional on survival (σ). The first stage, young subadult (SA_Y ; 12–18 months), is
 206 sex-independent. Young subadults transition to female ($SA_{O,F}$) or male ($SA_{O,M}$) old

207 subadults (18–24 months) depending on the sex ratio (0.55). Female old subadults
208 then transition to adult females (AF; >2 years), and male old subadults to the first
209 young-male stage (YM₁; 2–2.5 years in their natal pride). Young males (YM₁, YM₂,
210 YM₃, and YM₄; 2–4 years in their natal pride) transition to nomadic (NM; >2 years
211 nomadic) or resident males (RM; >2 years in another pride) conditional on emigration
212 (φ_{Em}^{YM} ; except for YM₄) and probability of transitioning to nomadic male (φ_{YM}).
213 Nomadic and resident males transition to the other adult male stage conditional
214 respectively on takeover (φ_T) and eviction (φ_{Ev}). Cubs are recruited by adult females
215 conditional on adult-female survival and reproduction probability ($P_{reproduction}$) as well
216 as on recruitment (Recruitment), which corresponds to the number of cubs born in a
217 given season that survived their first year per female. Circles and squares respectively
218 represent stages inside and outside their natal pride (in another pride for resident
219 males and in no pride for nomadic males). (b) To take advantage of the dead-recovery
220 data available for 105 lions, we included two dead stages: Newly and permanently
221 dead. Any alive state can transition to the newly dead state conditional on survival.
222 Newly dead individuals then transition to the absorbing permanently dead state. The
223 solid circle represents any alive state, dashed circles represent dead states.

224

225 Estimation of lion vital rates

226

227 *Survival and transition rates*

228

229 We estimated stage-specific survival and transitions, as well as detection
230 probabilities of pride individuals and nomadic males for the Serengeti lion population
231 using a Bayesian multistate capture-recapture model (MSCR; Lebreton & Pradel

232 2002; Schaub et al. 2004). In addition to the life stages described above, we also
233 included two more states, an observable newly dead and unobservable permanently
234 dead state (Gauthier & Lebreton, 2008), which allowed us to take advantage of the
235 dead-recovery data available for 105 individuals (i.e., lions found dead, as opposed
236 to lions with unknown fates) (Fig. 1b). Overall, we estimated the following
237 parameters: state-specific survival (σ_s), young-male emigration and transition to
238 nomadic male (φ_{Em}^{YM} and φ_{YM}), resident-male eviction (φ_{Ev}), and nomadic-male
239 takeover (φ_T). Lion prides are stable, territorial social groups (Schaller 1972); we can
240 thus expect that all pride members are in a fixed area in the vicinity of the collared
241 female in the pride. Consequently, we assumed all lions belonging to a pride to have
242 the same detection probability (p_{pride}) but estimated a separate parameter for
243 nomadic males (p_{NM}). In addition, we estimated the probability to observe a dead lion
244 (p_{dead}). Details on the multistate capture-recapture model can be found in Appendix
245 S2.

246

247 *Reproductive rates*

248

249 We estimated female reproduction probability and recruitment (i.e., number of cubs
250 born in a given season that reached their first birthday per female, conditional on
251 survival and reproduction) using a Bayesian generalized linear mixed model.

252 Following previous studies on the Serengeti lion, we defined recruitment as the
253 number of cubs reaching their first birthday (Packer et al. 2001). Because females
254 raise their cubs in crèches, the true mother of a given cub can be unknown.

255 Therefore, we first used data on cubs with known mothers to assign the total number
256 of cubs with a unique ID—i.e., the initial litter size regardless of whether they survived

257 their first year—to the right females. From the obtained number of identified cubs per
258 female, we created an initial litter-size distribution and used it to assign the cubs left
259 to their true mother among several potential females. We assigned each cub born in
260 a given season to a female among those available in the pride (i.e., with no more
261 than the maximum number of cubs observed; see details in Appendix S3).

262

263 We treated reproduction probability as a binary variable (i.e., 1 to females who
264 reproduced, 0 to females who did not). Based on the assignment of mothers to cubs
265 described above, we assigned 1 to females with cubs in the birth season of the cubs
266 (wet or dry) if the cubs were born more than 105 days after the beginning of the
267 season (i.e. the average gestation period; Schaller 1972) , or in the previous season
268 otherwise. We also assigned 1 to females identified as having lost their litter. In
269 addition, we assigned 0 to females without dependent offspring—young < 2 years
270 old—who could reproduce and NA to females with dependent offspring. We
271 modelled reproduction probability with a binomial distribution and recruitment with a
272 Poisson distribution using a generalized linear mixed model (GLMM) fitted in a
273 Bayesian framework (Kéry & Royle 2016).

274

275 Effects of density, season, and habitat on vital rates

276

277 *Density dependence*

278

279 To understand how socially- and spatially-explicit density measures affect lion
280 vital rates, we investigated vital-rate responses to various density-dependent factors
281 at the group (i.e. pride or male coalition) and home-range level (Table 1). To assess

282 the effect of density at the group level, we used the number of females in a pride and
283 the size of a resident or nomadic male coalition as density measures at the group
284 level. Both measures corresponded to the observed number of individuals in a given
285 group in each season. For the home-range level, we tested for the effect of the
286 number of nomadic coalitions in the home range of a pride or of a resident male
287 coalition using the overlap between that home range and the GPS location points of
288 a nomadic coalition (see Appendix S4 for details on the computation of home ranges
289 and of the number of nomadic coalitions in a home range). As nomadic coalitions do
290 not have assigned home ranges, we only tested for the effect of nomadic coalitions
291 on the vital rates of pride individuals. We only investigated the response of nomad
292 vital rates (i.e. survival and takeover probabilities) to coalition size and habitat.

293

294 Table 1 compiles the covariates included in the different vital-rate models and the
295 justification for their inclusion. While we estimated separate intercepts for female and
296 male old-subadult survival (Appendix S5: Fig. S5), we did not test for sex-specific
297 effects of density to avoid increasing model complexity. Due to methodological
298 constraints on the complexity of the model, we focused on assessing lion vital-rate
299 responses to density at the group and home-range level and did not explicitly test for
300 the effects of density at the higher population level. However, we investigated the
301 presence of signals of such effects by evaluating the correlation between time-
302 varying overall population size and season-specific yearly random effects (Appendix
303 S5: Fig. S3). Constraints on model complexity also prevented us from properly
304 testing for senescence in survival and reproduction—for which we only included a
305 quadratic age effect—which could have been done using a threshold model (e.g.,
306 Lemaître et al. 2020; Moullec et al. 2023).

307 *Seasonality*

308

309 Lions in the Serengeti experience strong seasonal patterns in rainfall (Norton-
310 Griffiths et al. 1975; Sinclair et al. 2000; 2013), and variability in such patterns can
311 have important consequences on food availability and thereby on lion demography
312 (Packer et al. 2005; Borrego et al. 2018). To understand whether seasonal
313 environmental patterns lead to seasonal density feedbacks, we estimated season-
314 specific vital rates—i.e., we estimated season-specific coefficients in all vital-rate
315 models described above—, with the wet season starting mid-November and the dry
316 season mid-May. However, due to a lack of data, we could not estimate a season-
317 specific effect of the number of nomadic coalitions on old-subadult survival and thus
318 only estimated the mean effect across seasons. Although we did not include the
319 effect of rainfall in our models, we investigated signals of potential effects of rainfall
320 on vital rates by assessing the correlation between rainfall and coefficients of
321 random effects (Appendix S5: Fig. S3).

322

323 **Table 1 – Socially- and spatially-explicit density covariates included in**
324 **the various vital-rate models.** We tested for the effect of density measures at the
325 group- (number of adult females in the pride and male coalition size) and home-
326 range level (number of nomadic coalitions in the home range) on lion survival,
327 transition, and reproductive rates. In addition, we tested for the effect of age on
328 adult-female survival and reproduction probability, and of its quadratic term on
329 reproduction probability. Each covariate (Covariate) is associated to the
330 corresponding vital rates (Vital rate) according to previous studies or assumptions
331 that have previously not been investigated (Motivation).

Covariate	Vital rate	Motivation
Number of adult females in the pride	Young subadult survival	<p>Takeovers can be prevented by females protecting their offspring, thus reducing the probability of a successful takeover in groups of females compared to singletons (Grinnell & McComb 1996), and consequently the mortality of young individuals (Packer et al. 1990). However, small and large prides can attract nomadic coalitions more, leading to a higher takeover rate in these prides and thereby a higher mortality of young through infanticide or forced dispersal (Packer & Pusey 1987; Pusey & Packer 1994; Elliot et al. 2014), with potentially severe consequences at the population level (Whitman et al. 2004).</p> <p>Moreover, the survival of adult females can be affected by the size of the pride: Females in small prides have lower survival rates, probably due to encounters with infanticidal males or females of other prides competing for the territory (Pusey & Packer 1994; Packer & Pusey 1997).</p>
	Old subadult survival	
	Young male survival	
	Adult female survival	
	Reproduction probability	
	Recruitment (number of cubs surviving to their first birthday per female, conditional on reproduction)	
Number of adult females in the pride ²	Reproduction probability	<p>Reproduction is mainly driven by takeover dynamics and interpride competition (Packer 2023), with small prides being unable to defend their cubs against outside males or defend their territories against larger neighbouring prides, and large prides attracting more frequent male takeovers and suffering greater within-pride feeding competition. We thus expect a u-shaped response of reproduction to the number of adult females in the pride (Packer 2023), which can be detected by including a quadratic term.</p>
Coalition size	Nomadic male survival	<p>Successful takeovers are affected by the size of both resident and nomadic coalitions (Bygott et al. 1979; Packer & Pusey 1983a; Borrego et al. 2018).</p>
	Resident male survival	
	Nomadic male takeover	
	Resident male eviction	
	Young subadult survival	

Number of nomadic coalitions in the home range	Old subadult survival	Nomadic coalitions taking over prides can increase the mortality of subadults and older young through infanticide and forced dispersal (Packer & Pusey 1987; Elliot et al. 2014; Packer 2023). Protective encounters by mothers with nomadic coalitions can lead to injuries and lower survival of adult females (Pusey & Packer 1994; Packer & Pusey 1997).
	Young male survival	
	Adult female survival	
	Resident male survival	More nomadic coalitions increase takeover rates (Borrego et al. 2018). Although this has not been explicitly tested, higher numbers of nomadic males could also lead to more encounters with resident males, potentially affecting their survival.
	Resident male eviction	
	Reproduction probability	Higher numbers of nomadic coalitions in the population can lead to more takeovers, increasing cub mortality due to infanticide (Bertram 1975; Pusey & Packer 1994; Whitman et al. 2004; Borrego et al. 2018).
	Recruitment (number of cubs surviving to their first birthday per female, conditional on reproduction)	
Number of adult females in the pride : Number of nomadic coalitions in the home range	Reproduction probability	While it has not yet been explicitly tested, this interaction would enable us to understand whether the effect of nomads on reproduction can be counterbalanced by females in the pride.
	Recruitment (number of cubs surviving to their first birthday per female, conditional on reproduction)	
Age	Adult female survival	Testing for senescence and age-dependent reproduction.
	Reproduction probability	
Age ²	Reproduction probability	Females in our population have been observed to reproduce between 2.5 and 15 years old, but most reproduce between 3 and 10 years old. We should thus observe lower reproduction probabilities for young and old females.

333 *Habitat*

334

335 Lions in our study population inhabit two different habitats (plains and woodland)
336 where vital rates can display different patterns. Food availability in the plains strongly
337 varies between seasons and is particularly scarce in the dry season (Schaller 1972;
338 Sinclair & Norton-Griffiths 1995; Packer et al. 2005; Sinclair et al. 2013). On the
339 other hand, lions in the woodland benefit from a somewhat continuous food
340 availability throughout the whole year (Hanby & Bygott 1979; VanderWaal et al.
341 2009; Packer et al. 2005). We thus tested for the season-specific effect of habitat on
342 all lion vital rates except for the probability of young males becoming nomadic (ϕ_{YM}),
343 due to the lack of data on this transition. As for density, we did not test for sex-
344 specific habitat effects on the survival of old subadults. We accounted for differences
345 in detection probabilities between habitats by including a habitat effect on all stage-
346 specific detection probabilities.

347

348 *Correlation among covariates and year random effect*

349

350 We checked for correlations between covariates using the Pearson correlation
351 coefficient for two density-dependent (continuous) variables (using the *cor* function
352 from the *stats* R package; R Core Team 2022), and the biserial correlation coefficient
353 for a density-dependent (continuous) variable and the categorical habitat variable
354 (using the *binomial.cor* function of the *lrm* R package version 1.2-0; Rizopoulos
355 2007). We considered two variables to be uncorrelated when the absolute value of
356 the correlation coefficient was under 0.5. In addition to density, season, and habitat,

357 we included a yearly season-specific random effect in all models to account for
358 among-year variation unexplained by density or habitat.

359

360 *Standardization of continuous covariates*

361

362 We standardized all non-binary covariates using the approach described by Gelman
363 (2008):

364

$$365 \text{covariate}_{\text{scaled}} = \frac{(\text{covariate}_{\text{unscaled}} - \mu_{\text{covariate}_{\text{unscaled}}})}{2 \cdot \sigma_{\text{covariate}_{\text{unscaled}}}} \text{ (Equation 1)}$$

366

367 where μ and σ are respectively the mean and standard deviation of a given unscaled
368 covariate. In comparison with the common standardization by one standard
369 deviation, this standardization approach enables the comparison of the effect sizes
370 of both categorical (i.e. habitat) and continuous covariates (i.e. density-dependent
371 variables).

372

373 *Implementation using NIMBLE*

374

375 We used NIMBLE (version 1.0.1 of the *nimble* package; de Valpine et al. 2017;
376 2022) to implement both the multistate capture-recapture model and the generalized
377 linear mixed models in a Bayesian framework. For the multistate capture-recapture
378 model, to decrease the runtime and memory requirements of the Markov chain
379 Monte Carlo algorithm (MCMC), we created a custom distribution integrating over
380 latent states, based on Nater et al. (2020; see Appendix S2 for details). We used

381 non-informative priors for all parameters and ran the MCMC for four chains of 60,000
382 iterations with no thinning and a burn-in phase of 15,000 iterations the multistate
383 model and 10,000 for the GLMM. We tested for parameter extrinsic identifiability
384 using prior-posterior overlap (Gimenez et al. 2009) and assessed model fit using
385 posterior predictive checks (Conn et al. 2018). The detailed methods are available in
386 Appendix S2. All analyses were performed in R 4.2.2 (R Core Team 2022) using
387 RStudio (Posit team 2023). R code for running analyses and plotting results is
388 available on Zenodo [citation placeholder] and on GitHub at [Github link placeholder].
389

390 **Results**

391 Socially- and spatially-explicit density dependence of vital rates 392

393
394 Most vital rates were influenced by at least one measure of density at the
395 group or home-range level, the only exception being adult-female survival. Moreover,
396 some density effects varied between seasons (Fig. 2, Fig. 3, Fig. 4, and Appendix
397 S5: Fig. S1). Many vital rates also differed between the plains and woodland
398 habitats, but the degree of vital-rate variation due to density dependence was
399 generally higher than that due to habitat (Fig. 2 and Appendix S5: Fig. S1). In Figure
400 2 and Figure 3, we highlight the lack of response of adult-female survival to the
401 density measures we considered (Fig. 2a). In addition, we show the most compelling
402 examples of how lion vital rates respond to various density measures at the group
403 (reproduction probability, and old-subadult, resident-male, and nomadic-male
404 survival; Figs. 2b-d and Figs. 3a-d) and home-range levels (recruitment; Fig. 3f). We
405 also show notable examples of seasonal differences in density effects on lion vital

406 rates (old-subadult and resident-male survival; Figs. 2c, d). In the following, all
407 results are presented using the median of the posterior distribution for each
408 parameter and the 90% credible interval (more stable than the 95% CRI, following
409 Kruschke 2014) on the probability (for survival and transition rates and reproduction
410 probability) or natural scale (for recruitment).

411

412 Among all vital rates for which we tested the effect of density, the survival of adult
413 females was the only one not markedly affected by at least one density measure at
414 the group or home-range level (Fig. 2a and Appendix S5: Fig. S3). Otherwise, many
415 vital rates were largely affected by density variables at the group level (Fig. 2 and
416 Appendix S5: Fig. S1). The number of adult females in the pride negatively affected
417 young subadult survival in the dry season (with a median survival probability of 0.98
418 [0.95, 0.99] with 2 females in the pride and 0.95 [0.90, 0.99] with 8 females). The
419 number of females in a pride also affected reproduction probability in the wet
420 season, with a quadratic effect indicating a higher reproduction probability in small
421 and large prides compared to prides of average size (0.21 [0.17, 0.28] with 2 females
422 in the pride, 0.15 [0.12, 0.20] with 8 females, and 0.17 [0.13, 0.23] with 12 females;
423 Fig. 2b and Fig. 3a). The effect of the number of females in the pride on old subadult
424 survival strongly differed between seasons (Fig. 2c and Fig. 3b). In larger prides with
425 more adult females, old-subadult survival decreased in the wet season (0.94 [0.89,
426 0.97] with 2 females in the pride to 0.89 [0.82, 0.94] with 8 females) but increased in
427 the dry season (from 0.92 [0.82, 0.98] to 0.98 [0.94, 1.0]). In contrast, the number of
428 adult females in the pride did not affect young-male survival or recruitment (Appendix
429 S5: Fig. S1).

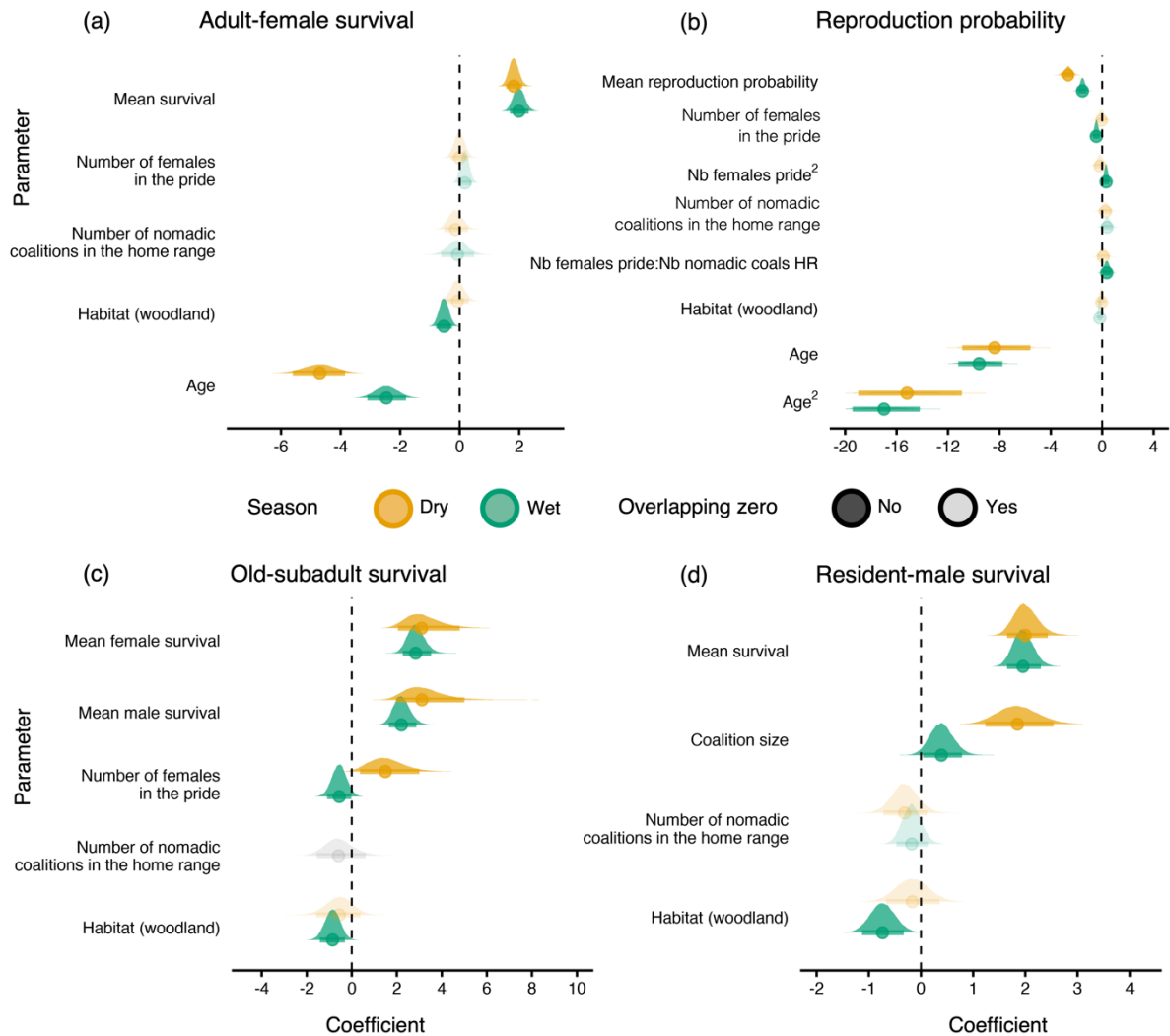
430

431 Adult males were affected by density measures at the group level as well, with
432 resident-male eviction probability decreasing with the size of the resident coalition in
433 the wet (from 0.0034 [0.00019, 0.017] for a coalition of 2 males to 0.00065
434 [0.000029, 0.0046] with 3 males) and dry season (from 0.035 [0.016, 0.062] to 0.015
435 [0.0052, 0.032]) (Appendix S5: Fig. S1). Resident-male survival increased with
436 coalition size in both seasons but showed large differences in the seasonal response
437 to coalition size (see Fig. 2d and Fig. 3c), with survival increasing more strongly with
438 larger coalitions in the dry season (from 0.89 [0.85, 0.92] for a coalition of 2 males to
439 0.95 [0.91, 0.97] with 3 males) than in the wet season (from 0.88 [0.84, 0.91] to 0.89
440 [0.86, 0.92]). Moreover, while the size of a nomadic coalition did not affect takeover
441 probability in the wet season, larger nomadic coalitions had higher chances to take
442 over a pride in the dry season (from 0.28 [0.20, 0.37] for a coalition of 2 males to
443 0.40 [0.28, 0.54] with 3 males; Fig. 3d). Nomadic coalition size also increased
444 nomadic-male survival both in the wet (from 0.88 [0.77, 0.95] for a coalition of 2
445 males to 0.96 [0.85, 0.99] with 3 males) and dry season (from 0.98 [0.93, 1.0] to 1.00
446 [0.99, 1.0]) (Appendix S5: Fig. S1).

447

448 In addition, at the home-range level, the number of nomadic coalitions negatively
449 affected recruitment in the wet season (from 0.54 [0.43, 0.67] cubs surviving their
450 first year per reproducing female with 2 nomadic coalitions in the home range to 0.33
451 [0.17, 0.61] cubs with 5 coalitions; Fig. 3e). By contrast, nomadic coalitions in the
452 home range positively affected dry-season survival of young subadults (with survival
453 probabilities ranging from 0.93 [0.87, 0.97] with no nomadic coalition in the home
454 range to 0.99 [0.96, 1.0] with 2 coalitions) and young males (from 0.82 [0.71, 0.93] to
455 1.0 [0.91, 1.0]), with both vital rates showing a particularly strong seasonal response

456 to nomadic coalitions (Appendix S5: Fig. S1). This unexpected positive effect of
457 nomadic males might be attributable to favourable environmental conditions. As
458 described by Borrego et al. (2018), increasing numbers of nomadic coalitions
459 coincide with years where wet-season rainfall is abundant; such conditions could
460 have positive effects on young-subadult and young-male survival rates. More
461 nomadic coalitions in the home range of a pride in the wet season also increased the
462 probability of eviction of resident males (from 0.0022 [0.000081, 0.014] with 1
463 nomadic coalition in the home range to 0.019 [0.0026, 0.062] with 4 coalitions).
464 While we found no effect of nomadic males on reproduction probability, reproduction
465 was affected by the interaction between the number of nomadic coalitions in the
466 home range and the number of females in the pride in the wet season. That is,
467 increasing numbers of nomadic coalitions had larger effects on reproduction
468 probability in prides with higher numbers of females (with 4 nomadic coalitions in the
469 home range of a pride, reproduction probability was 0.25 [0.16, 0.37] in prides of 4
470 females and 0.31 [0.19, 0.45] with 10 females; Fig. 2b and Fig. 3f and Appendix S5:
471 Fig. S1). However, we found no effect of nomadic coalitions on the survival of old
472 subadults and resident males (Fig. 2c and Fig. 2d).



473 **Figure 2 – Seasonal effects of habitat and density variables at the group**

474 **and home-range level on lion vital rates.** Using a Bayesian multistate capture-

475 recapture and Bayesian GLMMs, we investigated the presence of seasonal patterns

476 in the response of lion survival, transition, and reproductive rates to the habitat type

477 (woodland or plains), within-group density (number of adult females and coalition

478 size), and the number of nomadic coalitions in the home range. The figure

479 represents the effect sizes of these covariates on adult-female (a) survival and (b)

480 reproduction probability; and on the survival of (c) old subadults; and (d) resident

481 males. Each plot represents, on the logit scale, the median (dots) and 90% Credible

482 Interval (CRI; lines) of each coefficient obtained from the multistate capture-

483 recapture model and the GLMMs. The density plots above each estimate show the

484 posterior distribution of each parameter. Shaded dots and CRIs indicate coefficients
485 with 90% CRI overlapping zero.

486

487 Habitat effects on vital rates

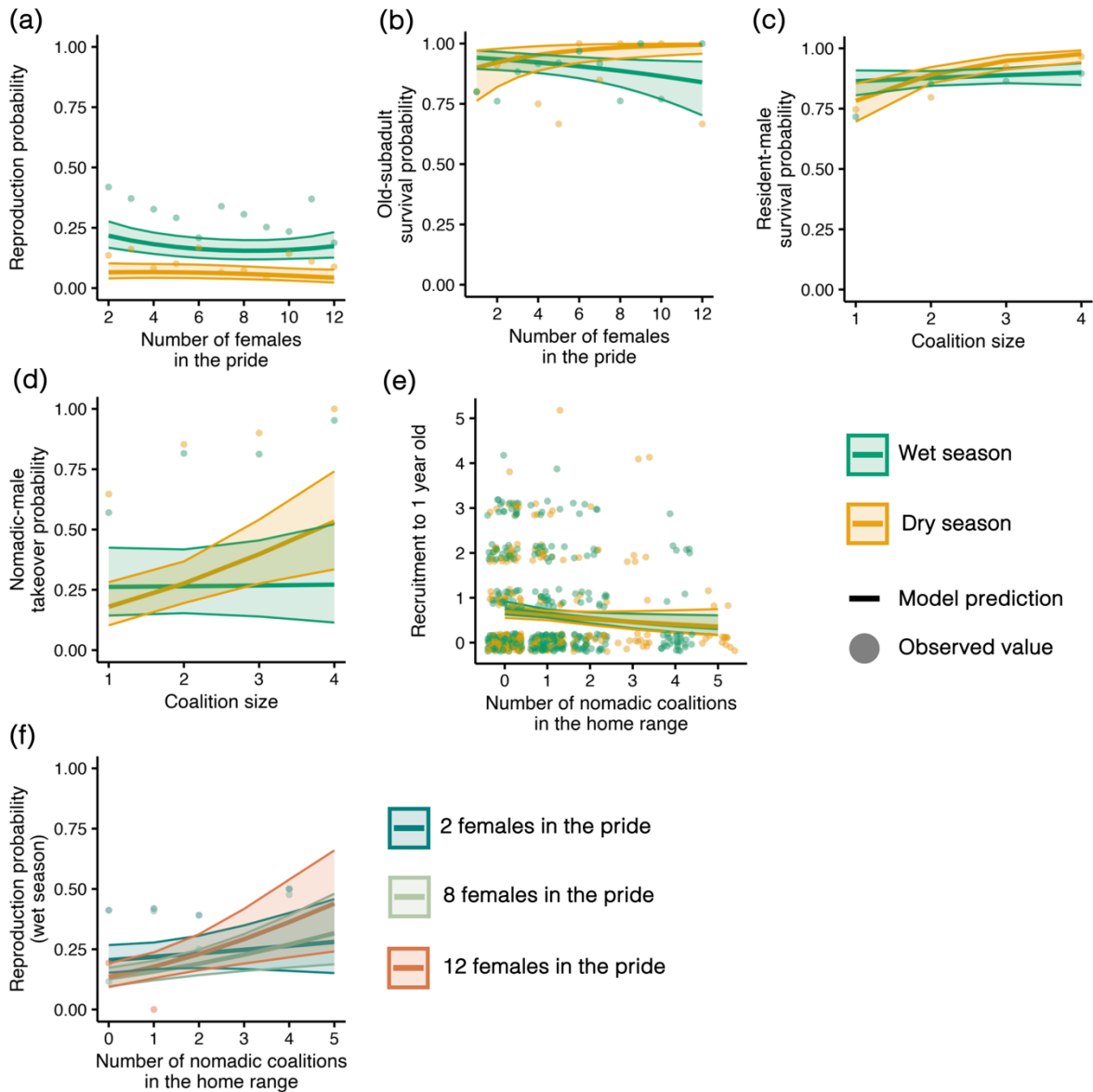
488

489 In addition to density, we found effects of habitat (plains or woodland) on most
490 vital rates, but these effects largely varied depending on the season and life-history
491 stage (Fig. 2). Overall, while we found no differences in survival between the plains
492 and the woodland in the dry season, survival was lower in the woodland in the wet
493 season compared to the plains (e.g., the survival probability of old subadults was
494 0.83 [0.74, 0.90] in the woodland and 0.92 [0.88, 0.95] in the plains, and resident
495 males had a survival probability of 0.77 [0.70, 0.83] in the woodland and 0.88 [0.84,
496 0.91] in the plains; see Fig. 2c, and Fig. 2d). Unlike density, the habitat did affect
497 adult-female survival, which decreased from 0.88 [0.85, 0.91] in the plains to 0.82
498 [0.76, 0.86] in the woodland in the wet season (Fig. 2a). The survival of nomadic
499 males also decreased in the woodland in the dry (0.97 [0.92, 0.99] in the plains and
500 0.85 [0.68, 0.97] in the woodland) and wet season (0.85 [0.76, 0.93] and 0.74 [0.56,
501 0.88]), while recruitment increased from 0.60 [0.51, 0.71] in the plains to 0.96 [0.79,
502 1.2] cubs per female in the woodland in the dry season (Appendix S5: Fig. S1).

503 Additionally, habitat-specific takeover probabilities for nomadic males strongly varied
504 between seasons, with takeover probability increasing from 0.26 [0.18, 0.35] in the
505 plains to 0.47 [0.30, 0.68] in the woodland in the dry season but decreasing from
506 0.30 [0.21, 0.40] in the plains to 0.15 [0.066, 0.28] in the woodland in the wet
507 season. However, we found no differences in young-male emigration probability and
508 female reproduction probability between habitats.

509

510 Finally, older females had a lower probability of survival, especially in the dry season
511 (0.99 [0.98, 0.99] at 3 years old and 0.87 [0.84, 0.90] at 13 years old) compared to
512 the wet season (0.97 [0.96, 0.98] and 0.89 [0.86, 0.91]; see Fig. 2a), with seasonal
513 differences in survival increasing with age. Similarly, age had a quadratic effect on
514 female reproduction probability in both seasons, indicating a lower reproduction
515 probability for young (at 4 years old, 0.26 [0.23, 0.30] in the wet season and 0.091
516 [0.072, 0.11] in the dry season) and old females (at 12 years old, 0.28 [0.23, 0.32] in
517 the wet season and 0.10 [0.076, 0.14] in the dry season) compared to 8 year-old
518 females (0.46 [0.41, 0.50] in the wet and 0.18 [0.15, 0.22] in the dry season; Fig. 2b).



519 **Figure 3 – Seasonal effects of socially- and spatially-explicit density**
 520 **measures on lion vital rates.** Using a Bayesian multistate capture-recapture, we
 521 investigated the presence of seasonal patterns in the response of lion survival and
 522 transition rates to the habitat type (woodland or plains), group density (number of
 523 adult females and coalition size), and the number of nomadic coalitions in the home
 524 range of a pride. The figure represents the model predictions of the response (a)
 525 reproduction probability and (b) old-subadult survival to the effect of the number of
 526 females in the pride; (c) resident-male survival and (d) nomadic-male takeover
 527 probability to male coalition size; (e) recruitment to 1 year old to the number of

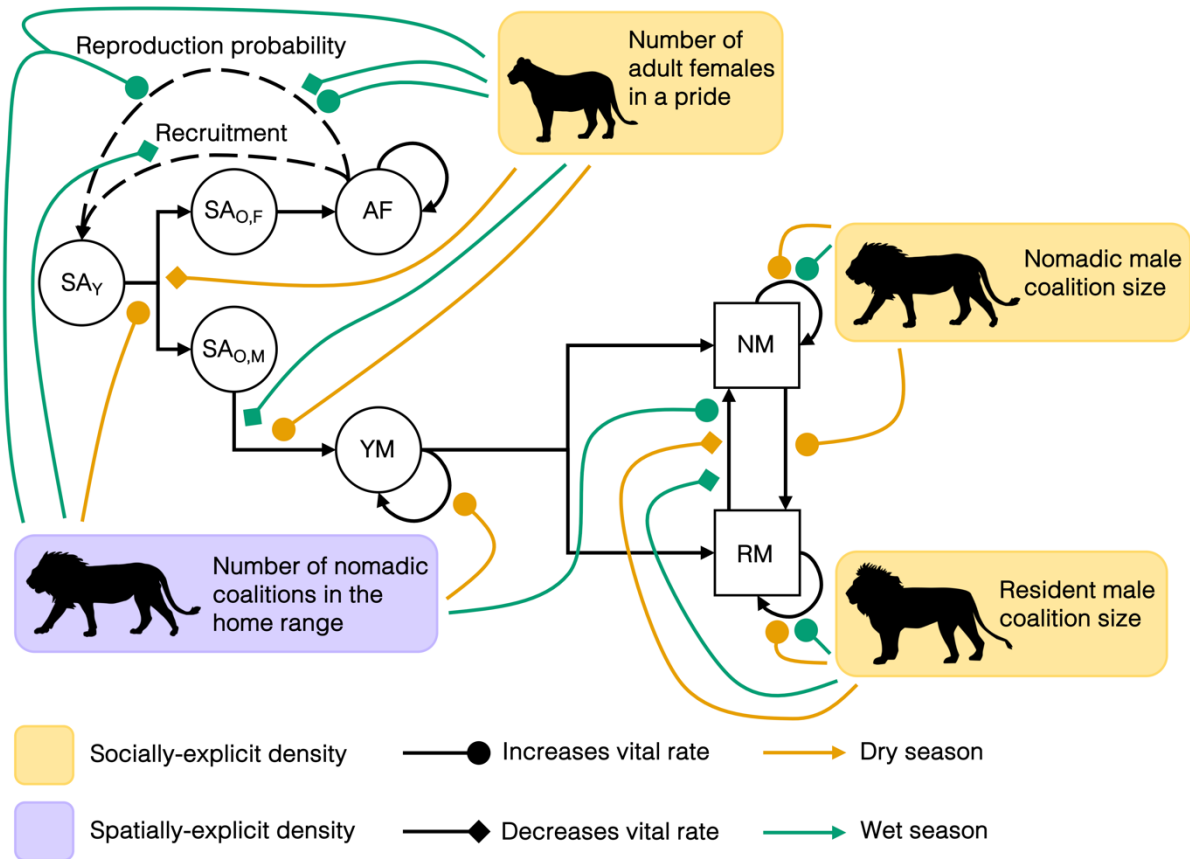
528 nomadic coalitions in the home range of a pride; and (f) wet-season reproduction
529 probability to the number of nomadic coalitions in the home range of a pride
530 depending on the number of females in the pride. Each plot represents the median
531 estimate (line) and 90% Credible Interval (CRI; lines) of each vital-rate prediction
532 derived from the output of the multistate capture-recapture model and the GLMMs.

533

534 Parameter identifiability and model fit

535

536 We found no strong evidence of non-identifiability for either the multistate
537 capture-recapture model or the GLMMs (Appendix S5: Fig. S4). Additionally, the
538 posterior predictive checks showed that the GLMMs fitted the data appropriately
539 (Appendix S5). This was also largely the case for the multistate capture recapture
540 model, with the exception of a few metrics (e.g., number of nomadic males becoming
541 residents or number of resident males becoming nomadic). For these, posterior
542 predictive checks suggested some estimation bias, and the results for the
543 corresponding vital rates (e.g., takeover or eviction probabilities) should be
544 interpreted with caution.



Silhouettes available on phylopic.org.
 The female and nomadic-male silhouettes were designed by Gabriela Palomo-Munoz and available under the CC BY 4.0 licence (<https://creativecommons.org/licenses/by/4.0/>).
 The resident-male silhouette was designed by Lisa Nicvert.

545 **Figure 4 – Seasonal effects of socially- and spatially-explicit density**
 546 **measures on lion vital rates.** Socially- and spatially-explicit density measures
 547 (yellow and purple boxes) have positive and negative effects on the different vital
 548 rates of Serengeti lions (round and diamond arrowheads; only the effects of
 549 covariates for which the coefficient 90% CRIs do not overlap 0 are represented), with
 550 differences in these effects between the dry and wet seasons (orange and green
 551 arrows). The complexity of vital-rate density dependence emphasizes the need to
 552 account for socially- and spatially-explicit considerations of density to assess the role
 553 of density feedbacks in shaping vital-rate variation in social species.

554 **Discussion**

555

556 Our study unveiled strong effects of local measures of density on the vital
557 rates of the Serengeti lion population, with seasonal differences in these effects for
558 some vital rates. Our results show variation both in the magnitude and direction of
559 vital-rate responses to a combination of season-specific socially-, and spatially-
560 explicit density measures at the group and home-range levels (Fig. 4). Importantly,
561 our results show strong effects of nomadic coalitions on key processes such as
562 reproduction and takeover dynamics. In addition, while the effects of season-specific
563 density were overall stronger than that of the habitat, we found lower survival
564 probabilities in the woodland in the wet season, and seasonal differences in the
565 effect of habitat type for various vital rates. Interestingly, our results indicate that
566 habitat and age were the only variables affecting the survival of adult females—a key
567 vital rate in many long-lived species (e.g., Eberhardt and Siniff 1977; Gaillard et al.
568 1998; Hunter et al. 2010). Unlike the other vital rates, adult-female survival thus
569 appeared buffered against changes in density measures considered in our study.
570 Overall, our findings emphasize the need for studies accounting for socially- and
571 spatially-explicit considerations of density when investigating vital-rate density
572 dependence in social and potentially other species. Moreover, our results highlight
573 the necessity to assess the effects of environment-density interactions, which can
574 play a key role in shaping vital-rate variability in a context of strong environmental
575 seasonality (Gamelon et al. 2017; Conquet et al. 2023).

576

577 Socially-explicit density dependence

578

579 With vital rates being affected by density measures at multiple scales (e.g., group or
580 population level), density feedbacks can affect social species through complex
581 pathways. For instance, Ausband et al. (2021) showed that both population density
582 and group size affected reproduction of grey wolves (*Canis lupus*), with interacting
583 effects of the two density measures. Additionally, in African wild dogs (*Lycaon*
584 *pictus*), different measures of density at the intra-group level (e.g., number of pups or
585 number of adults) had very variable effects on dispersal probabilities, with sex-
586 specific responses (Behr et al. 2020). Our results highlight this complexity, revealing
587 that also lion demography is affected by a combination of density measures at the
588 group (pride and male coalition size) and home-range level (number of nomadic
589 coalitions). As expected, the size of a social group (i.e. pride or male coalition) had
590 important, complex effects on many vital rates, corroborating previous findings on
591 density effects on lion vital rates. More specifically, larger male coalitions gave an
592 advantage to males both in survival and in gaining (for nomads) or maintaining the
593 tenure of a pride (for residents) (Bygott et al. 1979; Packer & Pusey 1983a; Borrego
594 et al. 2018). Additionally subadult survival decreased in prides with more females
595 and female reproduction probability was higher in small and large prides than in
596 prides of average size. This may be explained by females struggling to defend
597 smaller prides—where resident coalitions are often absent (Packer et al. 1988;
598 Pusey & Packer 1994)—and a greater competition between coalitions for larger
599 prides resulting in more frequent coalition takeovers (Packer & Pusey 1987), leading
600 to higher young mortality due to infanticide and forced dispersal (Packer 2023), and
601 consequently to more frequent reproduction events (Bertram 1975; Packer et al.

602 1988). In addition, within-pride competition for food is stronger in large prides, where
603 individuals are consequently thinner than in smaller prides, leading to reduced
604 survival rates (Packer 2023). While we found effects of the number of females on
605 reproduction probabilities, our results showed no such effects on recruitment (i.e.,
606 the number of cubs surviving to one year old). This is contrary to previous studies,
607 which found notable effects of pride size on female reproductive output (Packer et al.
608 1990; Packer 2023). This might be due to our analysis underestimating the number
609 of reproducing solitary females (see Appendix S5 for more details), who often must
610 settle in low-quality habitats, causing high rates of litter loss (Packer 2023). Overall,
611 however, our results might indicate that belonging to a pride of at least two lionesses
612 may be key to raising cubs until their first birthday, but two or ten females does not
613 make any discernible difference. Low recruitment in small prides could also possibly
614 be concealed by a strong effect of other density measures, such as the number of
615 nomadic coalitions in the home range of a pride.

616

617 While males are often overlooked in demographic studies, they are an important part
618 of the life history of many species (Rankin & Kokko 2007) and often play a key role in
619 shaping their demography (e.g., Borrego et al. 2018; Penteriani et al. 2011). In
620 species where male infanticide due to nomadic individuals replacing residents is
621 prominent, males may have particularly strong effects on vital rates, with potentially
622 drastic consequences for population dynamics and strong population declines when
623 males are especially targeted by regulation or trophy hunting activities (e.g. Swenson
624 2003; Whitman et al. 2004). Nonetheless, while the effect of nomadic individuals on
625 population demography has been extensively assessed in birds (Penteriani et al.
626 2006; 2011), the role of nomadic males in shaping demography is rarely accounted

627 for in mammals. Despite data and modelling limitations (see Appendix S5 for
628 details), we found important effects of nomad abundance on several vital rates,
629 which confirm previous findings. For example, the probability of a female reproducing
630 in the wet season increased with the number of nomadic coalitions in the home
631 range of a pride, especially in prides with more females, which are more attractive to
632 nomads (Packer & Pusey 1987). Additionally, as suggested by Borrego et al. (2018),
633 takeover dynamics leading to infanticide—as indicated by the increased eviction
634 probability—had negative effects on recruitment (i.e., the number of cubs surviving
635 the first year per female, conditional on reproduction). Opposite responses of
636 reproduction probability and recruitment to an increased presence of nomadic
637 males—and higher takeover rates—are expected because females who lose their
638 cubs following a takeover can mate soon after (Bertram 1975; Packer et al. 1988).
639 Overall, our results show that nomads can play a key role in shaping vital rates in
640 mammal populations, emphasizing the need to invest efforts in monitoring nomadic
641 or transient individuals to better understand the demography of populations.

642

643 Despite most lion vital rates showing important responses to at least one measure of
644 density, our results suggest that adult-female survival is affected only by the habitat
645 and age, and not by the density measures we considered. Population dynamics of
646 long-lived species are typically sensitive to variation in the survival of adult females
647 (e.g., Eberhardt & Siniff 1977; Gaillard et al. 1998; Hunter et al. 2010; but see Gerber
648 & Heppell 2004); the response of such key vital rates to density could therefore have
649 important consequences on population dynamics. For example, under environmental
650 conditions causing population declines, the absence of compensating density
651 feedbacks acting as a buffer against adverse environmental effects (e.g., Reed et al.

2013; Paniw et al. 2019) could prevent populations from recovering. The absence of buffering density dependence could have dire consequences for many populations facing increasing climate-change and anthropogenic pressures with negative effects on vital rates (e.g., Vinks et al. 2021, Conquet et al. 2023). Conversely, the lack of negative density effects on key vital rates such as adult female survival, could favour populations experiencing strong negative density feedbacks in other vital rates. This could contribute to limiting overcompensatory density dependence in populations experiencing strong negative feedbacks coupled with adverse environmental conditions (Coulson et al. 2001; Fauteux et al. 2021). Overall, however, our results reveal important density effects on the vital rates of Serengeti lions at the group and home-range levels, as well as indications of vital-rate responses to population size (Appendix S5: Fig. S3). These findings thus emphasize the need for a systematic assessment of the effects of a socially- and spatially-explicit consideration of density.

Vital-rate responses to season-density interactions

Context dependence in density effects have been widely described in various species, with age-specific and sex-specific density effects (e.g., Fay et al. 2017), and vital-rate responses to density varying among climatic conditions (e.g., Dierickx et al. 2019). Such environment-density interactions can have critical effects on population persistence (Coulson et al. 2001; Gamelon et al. 2017) and are therefore paramount to account for. Lions in the Serengeti experience strong seasonal rainfall patterns driving prey availability (Norton-Griffiths et al. 1975; Packer et al. 2005; Sinclair et al. 2013) and these environmental patterns lead to seasonality in lion vital rates, similarly to several other systems (Letcher et al. 2015; Payo-Payo et al. 2022;

677 Conquet et al. 2023). The key role of seasonal environmental patterns in driving
678 variations in vital-rate responses to density (e.g., Barbraud & Weimerskirch 2003,
679 Sandvig et al. 2017) is supported by our results. For example, positive or negative
680 density effects can be intensified in a given season, as exemplified by the stronger
681 increase in resident-male survival with higher coalition size in the dry compared to
682 the wet season. Larger male coalitions might be more successful at hunting more
683 and larger prey, ensuring their survival during times of prey scarcity. Additionally,
684 environmental seasonality can lead to opposite density effects between seasons. For
685 example, in the wet season, old subadults fared worse in large prides compared to
686 prides with less females, but the opposite was true in the dry season. This pattern
687 likely arose because our analysis estimates apparent survival and does not
688 discriminate between survival and permanent emigration. Under favourable
689 environmental conditions such as that occurring in the wet season, subadults
690 approaching adulthood may be more likely to emigrate in response to higher lion
691 densities in large prides, causing the observed season-specific effect of density on
692 apparent survival.

693

694 While density feedbacks could be key in allowing populations to persist under the
695 predicted changes in seasonality (Conquet et al. 2023), changes in seasonal
696 patterns could also increase negative density effects, potentially leading to
697 population declines (Gamelon et al. 2017; Paniw et al. 2019). For example, in lions,
698 a shift towards drier seasons could strengthen the negative effect of nomads on
699 recruitment, and of the number of females on young subadult survival. If not
700 counterbalanced, for example by wet-season dynamics, such effects could be
701 detrimental to the recruitment of young in the population, thereby critically hampering

702 population persistence. Understanding how such changes in seasonal patterns will
703 affect populations experiencing strong seasonality and density feedbacks (e.g.,
704 Hansen et al. 1999; Lima et al. 2002; Marra et al. 2015) requires investigating the
705 presence of season-density interactions, as such interactions are likely to play a
706 crucial role in populations where key demographic processes (e.g., reproduction or
707 dispersal) are restricted to a specific period of the year (e.g., Lima et al. 2002; Lok et
708 al. 2013; Marra et al. 2015).

709

710 Habitat effects in lion vital rates

711

712 Similar to seasonality, different habitats can expose populations to very different
713 environmental conditions (e.g., resources availability or temperatures), with
714 consequential effects on vital rates (e.g., Ozgul et al. 2006; Swift et al. 2020). While
715 density had stronger effects on lion vital rates than the habitat, we nonetheless found
716 differences in vital rates between the plain and woodland lion prides, as well as
717 seasonal patterns in habitat effects. The two habitats differ mostly in terms of prey
718 availability, with plain lions experiencing an important decline in food availability in
719 the dry season, when the migrating herds of herbivores continue their migration
720 toward the north of the Serengeti to find food, while lions in the woodland have
721 access to similar amounts of prey most of the whole year (Packer et al. 2005). In the
722 dry season, conditions are thus more favourable in the woodland, leading to higher
723 recruitment rates compared to the plains. However, the survival of most stages was
724 lower in the wet season in the woodland compared to the plains, because of the
725 stronger increase in prey availability in the plains between the dry and wet season

726 compared to the stable abundance of prey in the woodland between seasons
727 (Packer 2023).
728
729 Variations in environmental conditions among habitat types can lead to differences in
730 density feedbacks among these habitats (e.g., Pärn et al. 2012; Marra et al. 2015),
731 potentially leading to tradeoffs in inhabiting better-quality habitats with stronger
732 negative density effects. While our models did not assess habitat-density interactions
733 and seasonal variation in such interactions, previous studies on the Serengeti lion
734 indicate that density feedbacks might be stronger in the woodland, where living
735 conditions are supposedly more favourable (Hanby & Bygott 1979). Further
736 investigations on seasonal patterns of habitat-density interactions could thus help
737 better understand how habitat differences shape the demography of species beyond
738 African lions through density feedbacks, and assess the potential consequences of
739 changes in habitat structure under anthropogenic land use or climate change.

740

741 Conclusion

742

743 Vital-rate density dependence is common across taxa and can be an important driver
744 of vital-rate variations, possibly more so than environmental variables. Density can
745 therefore be a key factor shaping demography, especially in species where sociality
746 is at the heart of life history. In such cases, therefore, assessing the effect of density
747 on vital rates requires investigating the relative effects of different measures of
748 socially- and spatially-explicit density that are relevant to each study system.
749 Moreover, vital rates can show complex responses to environment-density
750 interactions, and accounting for such interactions is therefore paramount to

751 understanding how density affects vital rates, more importantly for populations
752 experiencing environmental periodic patterns (e.g., seasonality). Our work not only
753 contributes to the body of literature emphasizing the importance of density in shaping
754 demography but additionally shows that density feedbacks can affect the
755 demography of social species through complex pathways involving density
756 measures at different scales. Consequently, assessing vital-rate responses to
757 density measures beyond group or population size, and accounting for socially- and
758 spatially-explicit considerations of density and their interactions with the environment
759 when estimating vital rates could provide a valuable insight on how density
760 dependence shapes demography in species where such complex feedbacks are
761 likely to be at play. Although methodological and data limitations did not allow for
762 such complexity in our study, assessing vital-rate responses to interacting density
763 measures would undoubtedly provide further invaluable insights on the role of
764 intraspecific density in shaping population demography (see e.g., Behr et al. 2020).
765 Additionally, accounting for the effects of interspecific density would allow for a more
766 exhaustive understanding of density feedbacks, as interspecific interactions can play
767 a key role in shaping population dynamics (Morrissette et al. 2010; Qu erou e et al.
768 2021). Studies accounting for these factors would enable capturing the full picture of
769 the role of density feedbacks in vital-rate variations, consequently leading to a better
770 assessment of the persistence of species beyond the Serengeti lion.

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772

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Appendix S1 – Study area and habitat types

The lion study population was monitored in a 2000-km² area in the Serengeti National Park, Tanzania (-2°27' N, 34°48' E) (Packer & Pusey 1987; Fig. S1). The population inhabits two main habitat types: In the plains, food availability is strongly seasonal, with migratory herbivores passing through in the wet season but a scarcer prey availability in the dry seasons (Packer et al. 2005). Conversely, in the woodland, lions have access to resident herbivores the whole year.

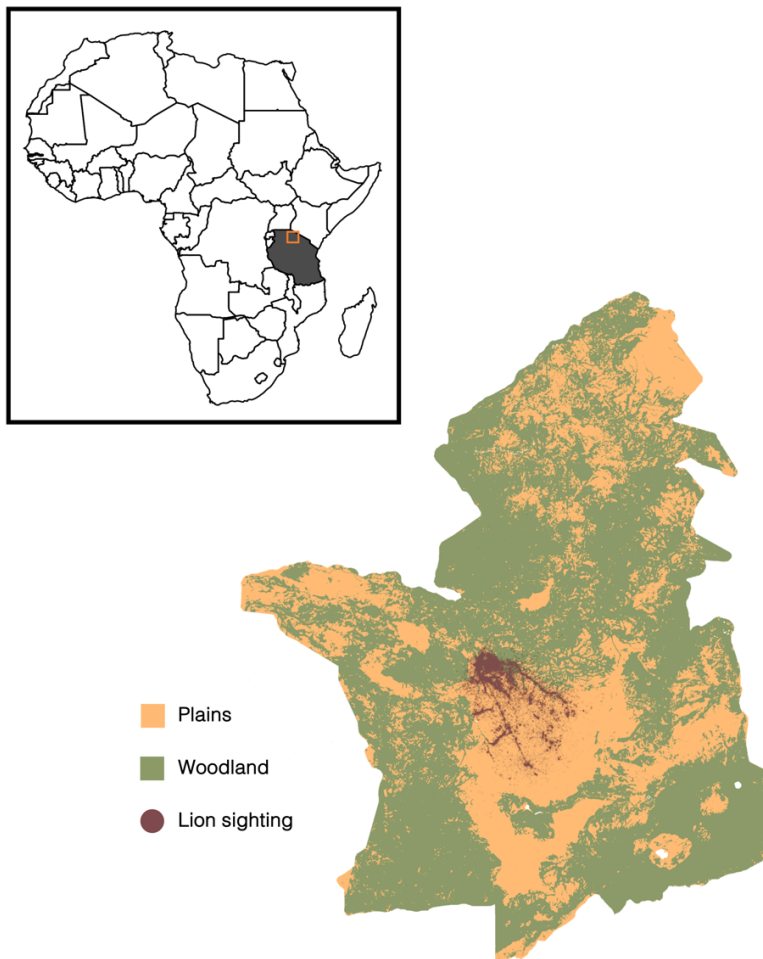


Figure S1 – Study area, habitat types, and lion sightings between 1984 and 2014. The studied population lives in the Serengeti National Park, Tanzania (inset map), and inhabits a region characterized by two main habitats: the plains (light orange areas), where food availability is strongly seasonal, and the woodland (light green areas), where lions have access to prey the whole year. The data on vegetation categories has been obtained by Grant Hopcraft from Reed et al. (2009) and is available at <https://serengetidata.weebly.com/>. Each transparent maroon dot represents the sighting of a single individual between 1984 and 2014.

37 **References – Appendix S1**

38

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Appendix S2 – Details on the model structure and custom likelihood distribution

Model structure

We used a Bayesian multistate capture-recapture model (Lebreton & Pradel 2002; Schaub et al. 2004) to estimate survival and transition rates as well as detection probabilities of pride individuals and nomadic males for the Serengeti lion population. The true, “latent” state of each individual in a given year, \mathbf{z}_t , is among 12 possible states. The first 10 states correspond to the 10 life-history stages we considered: (1) Young subadult (SA_Y ; 1–1.5 years) and old subadult (1.5–2 years), separated into (2) females ($SA_{O,F}$) and (3) males ($SA_{O,M}$), (4) adult females (AF ; > 2 years), young males—(5) YM_1 (2–2.5 years), (6) YM_2 (2.5–3 years), (7) YM_3 (3–3.5 years), and (8) YM_4 (3.5–4 years)—, (9) nomadic male (NM ; > 2 years and nomadic), and (10) resident male (RM ; > 2 years and resident in a different pride). In addition, to take advantage of the dead-recovery data available for 105 individuals (i.e., lions found dead, as opposed to lions who died or left the study area unwitnessed), we followed Gauthier and Lebreton (2008) and used an additional, observable (11) newly dead state. This approach allows lions in any state to transition to newly dead with a probability of $1 - \text{survival}$. Newly dead lions then transition to an absorbing, unobserved (12) permanently dead state with a probability of 1, and remain permanently dead afterwards.

The state process matrix (Fig. S1a) contains the transition probabilities among all 12 latent states. More specifically, these probabilities are conditional on the sex ratio (φ ratio, fixed at 0.55; representing the proportion of lionesses and thus the probability of an individual being female), state-specific survival (σ_s), young-male emigration and transition to nomadic male (φ_{Em}^{YM} and φ_{YM}), resident-male eviction (φ_{Ev}), and nomadic-male takeover (φ_T). The observation process matrix (Fig. S1b) contains the probabilities of observing a lion in its true state (i.e., detection probabilities). Due to the data collection method relying on finding a collared female in each pride, we assumed all lions belonging to a pride to have the same detection probability and therefore only estimated pride and nomad detection probabilities (p_{pride} and p_{NM}). In addition, we estimated the probability to observe a dead lion (p_{dead}).

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State_{t+1}

State _t	SA _y	SA _{0,F}	SA _{0,M}	SA _{0,F} · ϕ ratio	SA _{0,M} · (1 - ϕ ratio)	AF	YM ₁	YM ₂	YM ₃	YM ₄	NM	RM	New. Perm. dead
SA _y	0	σ _{SA_y}	0	0	0	0	0	0	0	0	0	0	1 - σ _{SA_y}
SA _{0,F}	0	0	0	0	0	σ _{SA_{0,F}}	0	0	0	0	0	0	1 - σ _{SA_{0,F}}
SA _{0,M}	0	0	0	0	0	0	σ _{SA_{0,M}}	0	0	0	0	0	1 - σ _{SA_{0,M}}
AF	0	0	0	0	0	0	0	σ _{AF}	0	0	0	0	1 - σ _{AF}
YM ₁	0	0	0	0	0	0	0	σ _{YM₁} · (1 - φ ^{YM₁)}	0	0	σ _{YM₁} · φ ^{YM₁}	σ _{YM₁} · φ ^{YM₁ · (1 - φ^{YM₁)}}	1 - σ _{YM₁}
YM ₂	0	0	0	0	0	0	0	0	σ _{YM₂} · (1 - φ ^{YM₂)}	0	0	σ _{YM₂} · φ ^{YM₂}	1 - σ _{YM₂}
YM ₃	0	0	0	0	0	0	0	0	0	σ _{YM₃} · (1 - φ ^{YM₃)}	0	σ _{YM₃} · φ ^{YM₃}	1 - σ _{YM₃}
YM ₄	0	0	0	0	0	0	0	0	0	0	σ _{YM₄} · φ ^{YM₄}	σ _{YM₄} · φ ^{YM₄ · (1 - φ^{YM₄)}}	1 - σ _{YM₄}
NM	0	0	0	0	0	0	0	0	0	0	0	σ _{NM} · φ ^T	1 - σ _{NM}
RM	0	0	0	0	0	0	0	0	0	0	0	σ _{RM} · φ ^{E_y}	1 - σ _{RM}
New. dead	0	0	0	0	0	0	0	0	0	0	0	0	0
Perm. dead	0	0	0	0	0	0	0	0	0	0	0	0	1

(a)

Observed_t

State _t	SA _y	SA _{0,F}	SA _{0,M}	AF	YM ₁	YM ₂	YM ₃	YM ₄	NM	RM	Dead	Unobserved
SA _y	1	0	0	0	0	0	0	0	0	0	0	0
SA _{0,F}	0	p _{pride}	0	0	0	0	0	0	0	0	0	1 - p _{pride}
SA _{0,M}	0	0	p _{pride}	0	0	0	0	0	0	0	0	1 - p _{pride}
AF	0	0	0	p _{pride}	0	0	0	0	0	0	0	1 - p _{pride}
YM ₁	0	0	0	0	p _{pride}	0	0	0	0	0	0	1 - p _{pride}
YM ₂	0	0	0	0	0	p _{pride}	0	0	0	0	0	1 - p _{pride}
YM ₃	0	0	0	0	0	0	p _{pride}	0	0	0	0	1 - p _{pride}
YM ₄	0	0	0	0	0	0	0	p _{pride}	0	0	0	1 - p _{pride}
NM	0	0	0	0	0	0	0	0	p _{pride}	0	0	1 - p _{pride}
RM	0	0	0	0	0	0	0	0	0	p _{pride}	0	1 - p _{pride}
Dead	0	0	0	0	0	0	0	0	0	0	p _{dead}	1 - p _{dead}

(b)

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Figure S1 – State and observation process matrices. (a) The state process matrix represents the transitions among all twelve true states between time t (rows) and $t+1$ (columns), conditional on the sex ratio (ϕ ratio) and the survival (σ) and transition parameters (ϕ). (b) The observation process matrix represents detection probabilities (p), that is, probabilities of observing an individual in a given state (columns) depending on its true state (rows).

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Custom likelihood distribution

Given the high number of parameters estimated in our model, we used the opportunity offered by NIMBLE (de Valpine et al. 2017) to create custom distributions and built a custom likelihood distribution allowing us to integrate over latent states (Turek et al. 2016). This avoids the estimation of the true state of each individual at each timestep, consequently greatly reducing the dimension of the MCMC posterior distribution. Additionally, instead of the arrays commonly used in Bayesian multistate models, we rely on vectors (\mathbf{pi} and \mathbf{Zpi}), allowing us to use one-dimensional linear algebra instead of matrix algebra to estimate the probabilities and transitions between states. This reduces the memory requirements and running time of the model (by removing latent states corresponding to the true state of an individual at a given time; see Nater et al. 2020 for details). To create this distribution (*dDHMMlionKF*, referring to discrete Hidden Markov Model for lions, including known fate), we used the *nimbleFunction* function of the *nimble* package (version 1.0.1; de Valpine et al. 2017) and provide a description of the various parameters used in the function below. At each time step t , the vector of observed state probabilities \mathbf{Zpi} is updated depending on the possible true, latent states and the detection probabilities (dp). Similarly, the vector of latent state probabilities \mathbf{pi} is updated depending on the preceding observations and the survival and transition rates (*surv*, *emigYM*, *transYMNM*, *takeover*, and *eviction*). The log-likelihood $logL$ is updated at each timestep t by the sum of the vector of observed state probabilities \mathbf{Zpi} .

Below, we print the code for the custom distribution. The code can also be found on GitHub at [GitHub link placeholder] and on Zenodo [citation placeholder].

```

# States (S):

# 1 Subadult 1
# 2 Subadult 2 Female
# 3 Subadult 2 Male
# 4 Adult Female
# 5 Young Male 1
# 6 Young Male 2
# 7 Young Male 3
# 8 Young Male 4
# 9 Nomadic Male
# 10 Resident Male
# 11 Newly dead
# 12 Permanently dead

# Observations (O):

# 1 seen as Subadult 1
# 2 seen as Subadult 2 Female
# 3 seen as Subadult 2 Male
# 4 seen as Adult Female
# 5 seen as Young Male 1
# 6 seen as Young Male 2
# 7 seen as Young Male 3
# 8 seen as Young Male 4
# 9 seen as Nomadic Male
# 10 seen as Resident Male
# 11 seen dead
# 13 not seen

dDHMM_lionKF <- nimbleFunction(
  run = function(

    ## Argument type declarations

    x = double(1),          # Vector containing the observed capture history
data length = double(),    # Length of the capture history
    init = double(1),      # Initial state probabilities
    survSA1 = double(1),   # State-specific survival
    survSA2F = double(1),
    survSA2M = double(1),
    survAF = double(1),
    survYM = double(1),
    survNM = double(1),
    survRM = double(1),
    transYMM = double(1), # Between-state transitions
    emigYM = double(1),
    takeover = double(1),
    eviction = double(1),
    dpPride = double(1),   # Detection probabilities
    dpNM = double(1),
    dpDead = double(1),
    log = double()){      # Logical argument specifying whether the log of
the likelihood should be returned

    logL <- 0              # Initialise log-likelihood
    pi <- init              # Initialise state probabilities

```

```

for(t in 1:length){      # Iterate over observations

  # x = "recorded as"
  # pi = probability of each latent state, conditioned on preceding
observations
  # Zpi = probability of current observed capture, conditioned on each
possible latent state

  Zpi <- pi # Initialise Zpi with the values of pi to avoid assigning
            # values to Zpi when the observation probability of a given
            # latent state in a given observed state is 1 (e.g. Zpi[12]
            # when x[t] == 13, as permanently dead individuals will
            # always be unobserved).

  # Detection probabilities

  if(x[t] == 1){

    # We do not assign any value to Zpi[1] here because the latent state 1
    # "young subadults" is the first state defined in our model. Therefore,
    # in the capture histories, observations are either (1) an NA if
    # the first capture of an individual took place when it was older than
    # 1.5 years, or (2) a 1 if the first capture happened when
    # it was between 1 and 1.5 years old.

    Zpi[2] <- 0
    Zpi[3] <- 0
    Zpi[4] <- 0
    Zpi[5] <- 0
    Zpi[6] <- 0
    Zpi[7] <- 0
    Zpi[8] <- 0
    Zpi[9] <- 0
    Zpi[10] <- 0
    Zpi[11] <- 0
    Zpi[12] <- 0

  }

  if(x[t] == 2){

    Zpi[1] <- 0
    Zpi[2] <- pi[2] * dpPride[t]
    Zpi[3] <- 0
    Zpi[4] <- 0
    Zpi[5] <- 0
    Zpi[6] <- 0
    Zpi[7] <- 0
    Zpi[8] <- 0
    Zpi[9] <- 0
    Zpi[10] <- 0
    Zpi[11] <- 0
    Zpi[12] <- 0

  }

  if(x[t] == 3){

    Zpi[1] <- 0

```

```

Zpi[2] <- 0
Zpi[3] <- pi[3] * dpPride[t]
Zpi[4] <- 0
Zpi[5] <- 0
Zpi[6] <- 0
Zpi[7] <- 0
Zpi[8] <- 0
Zpi[9] <- 0
Zpi[10] <- 0
Zpi[11] <- 0
Zpi[12] <- 0

}

if(x[t] == 4){

Zpi[1] <- 0
Zpi[2] <- 0
Zpi[3] <- 0
Zpi[4] <- pi[4] * dpPride[t]
Zpi[5] <- 0
Zpi[6] <- 0
Zpi[7] <- 0
Zpi[8] <- 0
Zpi[9] <- 0
Zpi[10] <- 0
Zpi[11] <- 0
Zpi[12] <- 0

}

if(x[t] == 5){

Zpi[1] <- 0
Zpi[2] <- 0
Zpi[3] <- 0
Zpi[4] <- 0
Zpi[5] <- pi[5] * dpPride[t]
Zpi[6] <- 0
Zpi[7] <- 0
Zpi[8] <- 0
Zpi[9] <- 0
Zpi[10] <- 0
Zpi[11] <- 0
Zpi[12] <- 0

}

if(x[t] == 6){

Zpi[1] <- 0
Zpi[2] <- 0
Zpi[3] <- 0
Zpi[4] <- 0
Zpi[5] <- 0
Zpi[6] <- pi[6] * dpPride[t]
Zpi[7] <- 0
Zpi[8] <- 0
Zpi[9] <- 0
Zpi[10] <- 0

```

```

Zpi[11] <- 0
Zpi[12] <- 0

}

if(x[t] == 7){

Zpi[1] <- 0
Zpi[2] <- 0
Zpi[3] <- 0
Zpi[4] <- 0
Zpi[5] <- 0
Zpi[6] <- 0
Zpi[7] <- pi[7] * dpPride[t]
Zpi[8] <- 0
Zpi[9] <- 0
Zpi[10] <- 0
Zpi[11] <- 0
Zpi[12] <- 0

}

if(x[t] == 8){

Zpi[1] <- 0
Zpi[2] <- 0
Zpi[3] <- 0
Zpi[4] <- 0
Zpi[5] <- 0
Zpi[6] <- 0
Zpi[7] <- 0
Zpi[8] <- pi[8] * dpPride[t]
Zpi[9] <- 0
Zpi[10] <- 0
Zpi[11] <- 0
Zpi[12] <- 0

}

if(x[t] == 9){

Zpi[1] <- 0
Zpi[2] <- 0
Zpi[3] <- 0
Zpi[4] <- 0
Zpi[5] <- 0
Zpi[6] <- 0
Zpi[7] <- 0
Zpi[8] <- 0
Zpi[9] <- pi[9] * dpNM[t]
Zpi[10] <- 0
Zpi[11] <- 0
Zpi[12] <- 0

}

```



```

if(x[t] == 10){

Zpi[1] <- 0
Zpi[2] <- 0
Zpi[3] <- 0
Zpi[4] <- 0
Zpi[5] <- 0
Zpi[6] <- 0
Zpi[7] <- 0
Zpi[8] <- 0
Zpi[9] <- 0
Zpi[10] <- pi[10] * dpPride[t]
Zpi[11] <- 0
Zpi[12] <- 0

}

if(x[t] == 11){

Zpi[1] <- 0
Zpi[2] <- 0
Zpi[3] <- 0
Zpi[4] <- 0
Zpi[5] <- 0
Zpi[6] <- 0
Zpi[7] <- 0
Zpi[8] <- 0
Zpi[9] <- 0
Zpi[10] <- 0
Zpi[11] <- pi[11] * dpDead[t]
Zpi[12] <- 0

}

if(x[t] == 13){

Zpi[1] <- 0
Zpi[2] <- pi[2] * (1 - dpPride[t])
Zpi[3] <- pi[3] * (1 - dpPride[t])
Zpi[4] <- pi[4] * (1 - dpPride[t])
Zpi[5] <- pi[5] * (1 - dpPride[t])
Zpi[6] <- pi[6] * (1 - dpPride[t])
Zpi[7] <- pi[7] * (1 - dpPride[t])
Zpi[8] <- pi[8] * (1 - dpPride[t])
Zpi[9] <- pi[9] * (1 - dpNM[t])
Zpi[10] <- pi[10] * (1 - dpPride[t])
Zpi[11] <- pi[11] * (1 - dpDead[t])

# We do not assign any value to Zpi[12] here because individuals in
# the latent state 12 "permanently dead" are always unobserved
# (observed state 13). The value of Zpi[12] is therefore the one it has
# been initialised with (pi[12])

}

sumZpi <- sum(Zpi) # Log-likelihood contribution of given
# observed state x
logL <- logL + log(sumZpi) # Overall log likelihood

```

```

# Transition probabilities

if(t != length){

pi[1] <- 0
pi[2] <- Zpi[1] * survSA1[t] * 0.55
pi[3] <- Zpi[1] * survSA1[t] * (1 - 0.55)
pi[4] <- Zpi[2] * survSA2F[t] + Zpi[4] * survAF[t]
pi[5] <- Zpi[3] * survSA2M[t]
pi[6] <- Zpi[5] * survYM[t] * (1 - emigYM[t])
pi[7] <- Zpi[6] * survYM[t] * (1 - emigYM[t])
pi[8] <- Zpi[7] * survYM[t] * (1 - emigYM[t])
pi[9] <- Zpi[5] * survYM[t] * emigYM[t] * transYMNM[t] +
        Zpi[6] * survYM[t] * emigYM[t] * transYMNM[t] +
        Zpi[7] * survYM[t] * emigYM[t] * transYMNM[t] +
        Zpi[8] * survYM[t] * transYMNM[t] +
        Zpi[9] * survNM[t] * (1 - takeover[t]) +
        Zpi[10] * survRM[t] * eviction[t]
pi[10] <- Zpi[5] * survYM[t] * emigYM[t] * (1 - transYMNM[t]) +
        Zpi[6] * survYM[t] * emigYM[t] * (1 - transYMNM[t]) +
        Zpi[7] * survYM[t] * emigYM[t] * (1 - transYMNM[t]) +
        Zpi[8] * survYM[t] * (1 - transYMNM[t]) +
        Zpi[9] * survNM[t] * takeover[t] +
        Zpi[10] * survRM[t] * (1 - eviction[t])
pi[11] <- Zpi[1] * (1 - survSA1[t]) +
        Zpi[2] * (1 - survSA2F[t]) +
        Zpi[3] * (1 - survSA2M[t]) +
        Zpi[4] * (1 - survAF[t]) +
        Zpi[5] * (1 - survYM[t]) +
        Zpi[6] * (1 - survYM[t]) +
        Zpi[7] * (1 - survYM[t]) +
        Zpi[8] * (1 - survYM[t]) +
        Zpi[9] * (1 - survNM[t]) +
        Zpi[10] * (1 - survRM[t])
pi[12] <- Zpi[11] + Zpi[12]

pi <- pi / sumZpi # Normalise
}
}

returnType(double())

if(log) return(logL) else return(exp(logL)) # Return log-likelihood
}
)

```

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99 **x** Vector of the observed capture history data

100 **length** Length of the capture history

101 **init** Initial state probabilities

102 **survSA1** Young-subadult survival

103 **survSA2F** Female old-subadult survival

104 **survSA2M** Male old-subadult survival

105	survAF	Adult-female survival
106	survYM	Young-male survival
107	survNM	Nomadic-male survival
108	survRM	Resident-male survival
109	transYMNM	Probability of transition between young male and nomadic male
110	emigYM	Young-male emigration probability
111	takeover	Nomadic-male takeover probability
112	eviction	Resident-male eviction probability
113	dpPride	Pride member detection probability
114	dpNM	Nomadic-male detection probability
115	dpDead	Dead detection probability
116	log	Logical parameter defining whether the log likelihood is returned
117	logL	Log likelihood of the observed capture history
118	pi	Latent state probability conditional on observations in previous steps
119	Zpi	Current observed capture probability conditional on each latent state
120	sumZpi	Likelihood of a given observation, or marginal probability of current observed
121	capture	

122 **References – Appendix S2**

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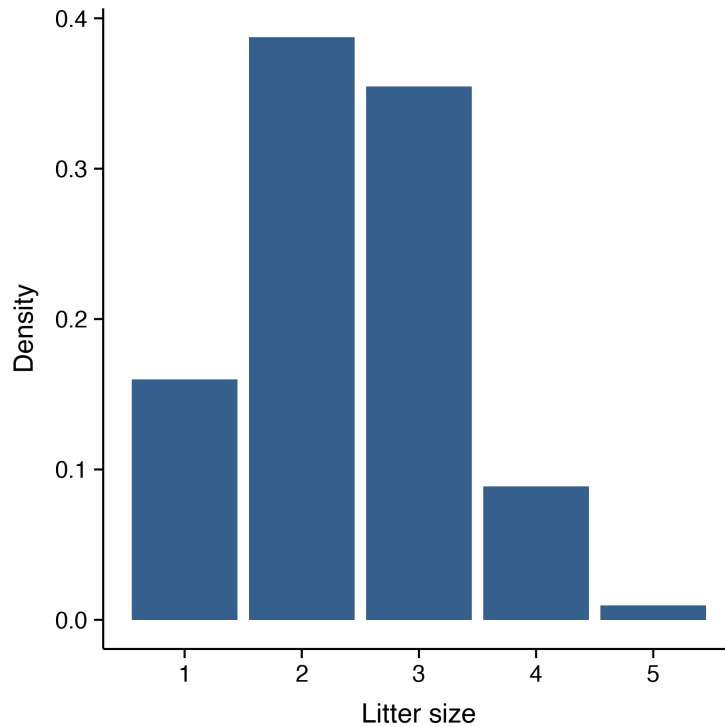
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1 **Appendix S3 – Female recruitment**

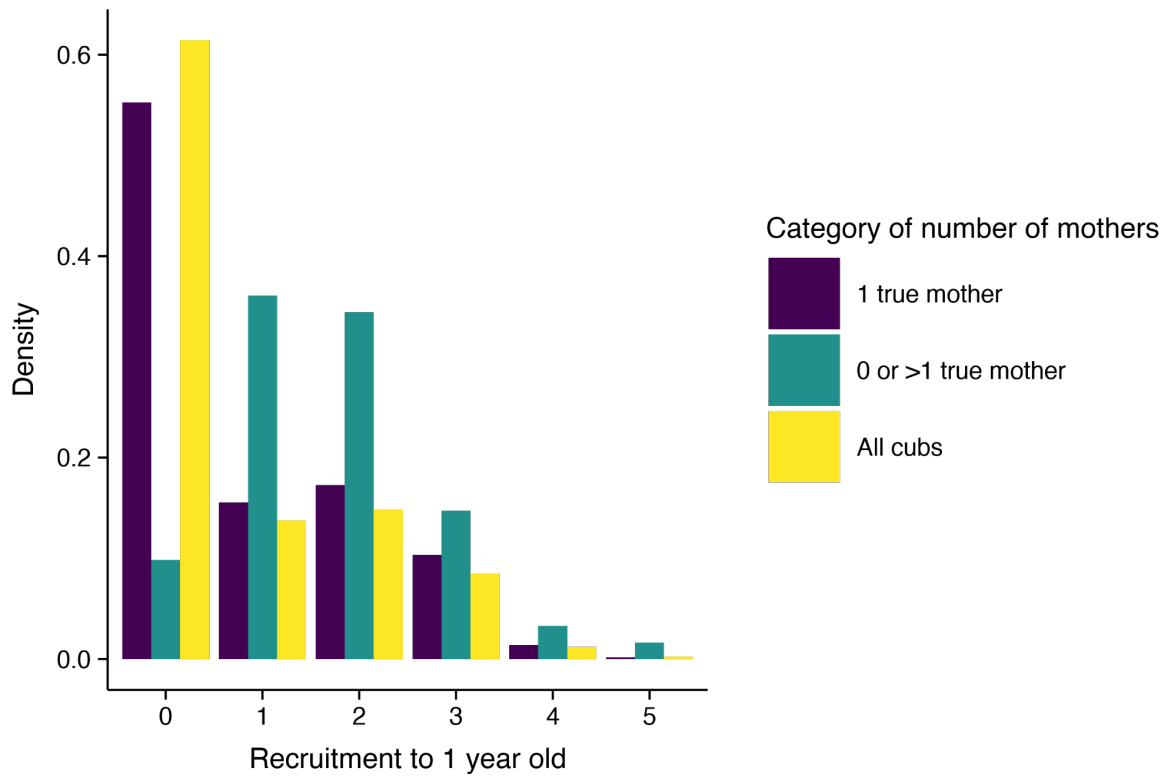
2

3 In our study, following previous research on the Serengeti lion, we defined
4 recruitment as the number of cubs reaching their first birthday (Packer et al. 2001). Because
5 females raise their cubs in crèches, we could not unequivocally assign a true mother to 42%
6 of the cubs (i.e., at least two females could be the mother or the cub had no potential mother
7 assigned). While in previous studies females could be assigned 0.5 cubs (Packer et al.
8 2001), we relied on observed data on litter size (i.e., integers only) for females identified as
9 the only known mother of cubs to assign the remaining cubs to females. That is, we first
10 used data on cubs with known mothers to assign the total number of cubs with a unique ID—
11 i.e., regardless of whether they survived their first year—to the right females in each
12 seasonal timestep t . From the obtained number of identified cubs per female per timestep t ,
13 we created an observed litter-size distribution. We used this distribution to assign a litter ID
14 to the cubs left with several potential mothers and born on the same day. For example, for a
15 group of five cubs born on the same day in the same pride and two possible mothers, two
16 different litters of two and three cubs are more likely to be created than a litter of five cubs
17 from a single female (Fig. S1). For each litter, we then chose the potential mothers in order
18 of priority: (1) among the potential mothers assigned to the cub by the observer, or, if all
19 potential mothers already had alive, independent offspring (i.e. young < 2 years old), (2)
20 randomly among the adult females (i.e. > 2 years old) belonging to the natal pride of the
21 cubs in the birth season of the cubs.



22 **Figure S1 – Distribution of female litter size in the dataset.** We used data on
23 cubs with a single assigned mother to create an observed distribution of litter size (i.e.,
24 number of cubs per female including cubs lost before their first birthday) and assign mothers
25 to cubs with no or several potential mothers.

26
27 To obtain the recruitment per reproducing female (i.e., the number of yearlings), we followed
28 the life history of each cub and removed it from the litter if it died before its first birthday. In
29 addition, we assigned zero cubs to females who lost their litter (recognized by lactation
30 stains with no cubs observed; Packer 2023). In some cases, the number of cubs observed in
31 a given pride was too high for the litter size per female to be kept at the maximum observed
32 litter size (i.e., five cubs). This is likely because some females in the focal pride were not
33 observed in the birth season of the cubs, and we therefore did not assign those cubs to any
34 female (<1% of the total number of cubs). The resulting distribution of female recruitment
35 (i.e., number of cubs ≥ 1 year old per reproducing female) on the whole dataset resembles
36 that of the observed recruitment (Fig. S2).



37 **Figure S2 – Distribution of female recruitment in the dataset.** We used data on
 38 cubs with a single assigned mother (1 true mother) to create an observed distribution of litter
 39 size (i.e., number of cubs per female including cubs lost before their first birthday) and
 40 assign mothers to cubs with no or several potential mothers (0 or >1 true mother). Although
 41 the distribution for cubs with no or more than one potential mother does not match that for
 42 the cubs with a single potential mother, the final distribution of recruitment (i.e., number of
 43 yearlings per female) in the full dataset (all cubs) matches it quite well.

44 **References – Appendix S3**

45

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1 **Appendix S4 – Number of nomadic coalitions in the home range of** 2 **a pride**

3

4 The effect of nomadic males on lion demography has previously been assessed by
5 looking at the number of nomadic coalitions entering the study area, that is, at the population
6 level (Borrego et al. 2018). Throughout the study period (1984–2014), nomadic coalitions
7 (i.e., coalitions of males above 2 years old that do not belong to a pride) in the study area
8 have been recorded through opportunistic sightings during monitoring of prides (Borrego et
9 al. 2018).

10 In our study, we assessed the response of survival, stage transitions, and reproductive rates
11 to the presence of nomadic males by testing for the effect of the number of nomadic
12 coalitions within any given pride home range. Because resident males spend only about
13 15% of their time with females of the pride (Packer 2023), we calculated separate home
14 ranges for resident males and for other pride individuals (i.e., subadults, young males, and
15 adult females of the same pride). That is, we used the GPS locations of individuals in a given
16 male coalition or pride to compute the 95% kernel utilization distribution using the *kernelUD*
17 and *getverticesHR* functions of the *adehabitatHR* R package (version 0.4.20; Calenge
18 2006)—with the *ad hoc* method “href” for the smoothing parameter of the bivariate normal
19 kernel. Using the utilization distribution of each group (i.e., resident-male coalition or pride),
20 we assessed the presence of nomadic coalitions by computing the overlap between the
21 home range of a group and the GPS locations of nomads in a given coalition, using the *over*
22 function of the *sp* R package (version 1.4-7; Pebesma & Bivand 2005; Bivand et al. 2013).
23 We added a nomadic coalition to the list of coalitions in a home range if the overlap was >0 ,
24 that is, if at least one individual in the focal nomadic coalition was observed in the home
25 range of a coalition or pride. We could not calculate a home range for resident coalitions or
26 prides for which we only had five or less locations and thus assigned NA to the number of
27 nomadic coalitions in the home range of these groups.

28 **References – Appendix S4**

29

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1 **Appendix S5 – Additional results, parameter identifiability, and**
 2 **posterior predictive checks**

3
 4

Effects of density-dependent factors and habitat on lion vital rates

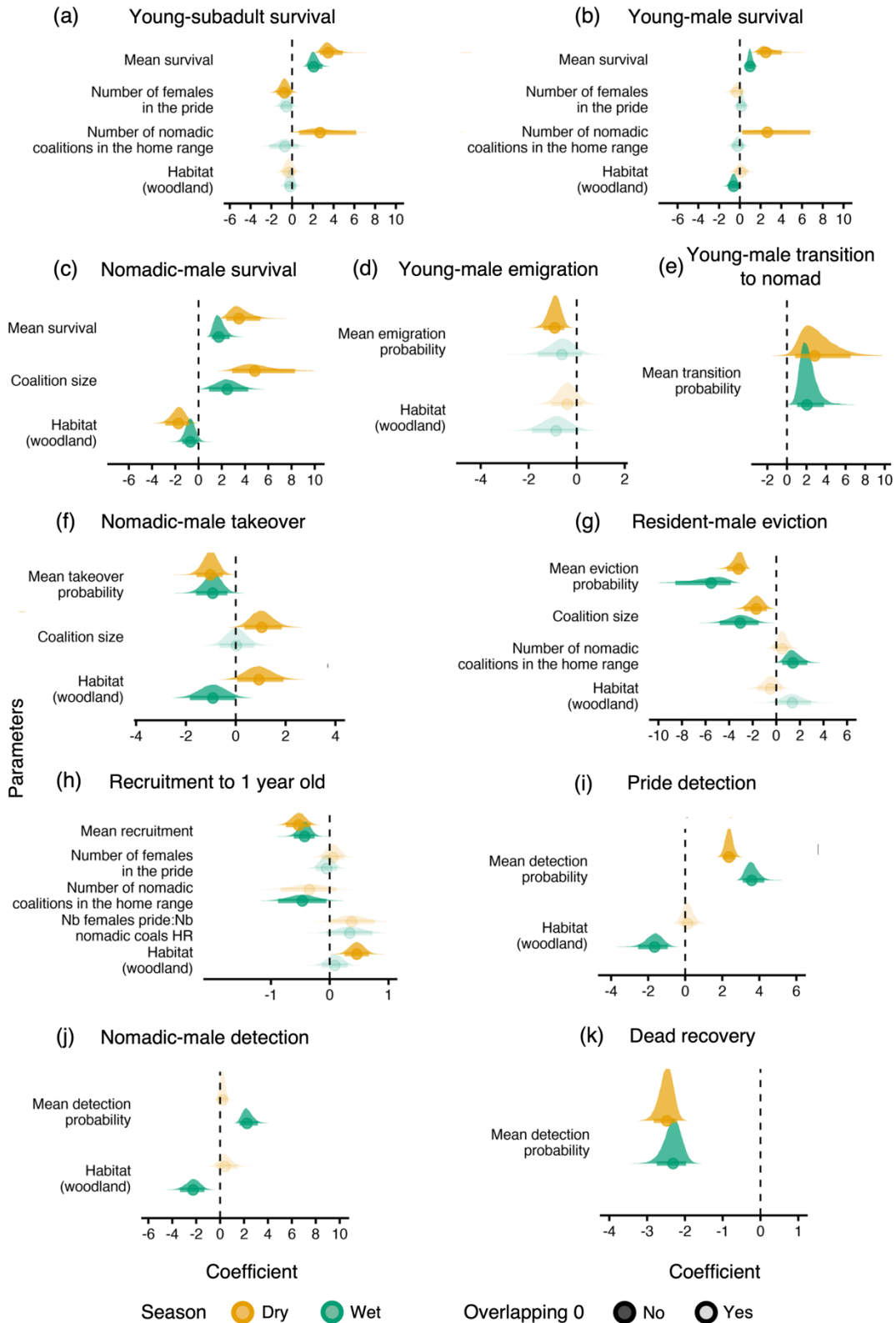


Figure S1 - Seasonal effects of habitat and within- and among-group density variables on lion vital rates. We investigated season-specific effects of within-group density (number of adult females and coalition size), the number of nomadic males in the home range, and habitat (plains or woodland) on the survival of (a) young subadults, (b) young males, and (c) nomadic males; (d) young-male emigration probability and (e) their probability to become nomadic after emigrating; probabilities of (f) nomadic-male takeover and (g) resident-male eviction; (h) recruitment to 1 year old; and detection probabilities of (i) pride individuals, (j) nomads, and (k) dead individuals. On each plot, the median (dots) and 90% Credible Interval (CRI; lines) of each coefficient (on the logit scale) were obtained from the posterior samples of the multistate capture-recapture model and the GLMMs. Density plots show the posterior distribution of each parameter. Shaded dots and CRIs are used for coefficients of effects for which there was little evidence in the data (i.e., 90% CRI overlapping with zero).

In contrast with the other vital rates, which were affected by at least one measure of density (Fig. S1), our results suggest that adult-female survival is affected only by the habitat and age, and not by the density measures we considered. However, previous findings have indicated negative effects of neighbours on female survival due to higher wounding rates (Mosser & Packer 2009). In their study, Mosser and Packer investigated the response of adult-female survival to the number of individual neighbours (males or females only, or both), while we focused exclusively on the effect of nomadic males by calculating the number of nomadic coalitions in the home range of a pride without regard to the number of neighbours; this could explain the discrepancies between our results and that of previous studies. Importantly, however, our results indicate a potential negative effect of population size—which might be correlated with the number of neighbours—on adult-female survival (Fig. S3). Therefore, in this specific case, the lower-level density measures we included in our models might not be able to provide additional insights on the response of adult-female survival to density. Additionally, effects of neighbours could, be grasped by the habitat. Our results indicate a lower wet-season survival rate of adult females in the woodland than in the plains. While lions in both habitats can profit from a high prey availability in the wet season, lion—and thereby neighbour—densities can strongly increase in good quality habitats such as the woodland, where prey availability is more consistent between seasons (Hanby & Bygott 1979; VanderWaal et al. 2009). Therefore, neighbour lions in general might have stronger effects on female survival than nomadic coalitions specifically.

40 Previous studies showed notable effects of pride size on female reproductive output, with a
41 higher number of cubs per female in average-sized prides (Packer et al. 1990; Packer 2023).
42 However, while we found an effect of the number of females in a pride on the probability of a
43 female reproducing in the wet season, our results showed no effect of females on
44 recruitment (i.e., the number cubs surviving to one year old). Previous studies focused on
45 female overall reproductive output, whereas we partitioned this output into two components:
46 reproduction probability (i.e., the probability to become a reproducing female) and
47 recruitment to 1 year old (i.e., the number of cubs reaching their first birthday per
48 reproducing female). Although this approach enables us to assess the seasonal effects of
49 density and habitat on each of these components, this partitioning potentially introduces a
50 bias in the estimation of reproduction probability, which might be underestimated in our
51 analyses. This is due to the lack of data on pregnancy resulting in lost litters for some
52 females, especially solitary lionesses, who often fail to recruit cubs due to their limited
53 access to high-quality territories (Packer 2023), and whose reproduction is seldom recorded.
54 While females who were not seen reproducing (i.e., pregnant, with lactation stains, or with
55 small cubs) had a recruitment of 0 in previous studies, we assigned them a reproduction of 0
56 and *NA* cubs. Many solitary females were thus considered as non-reproducing and excluded
57 from the recruitment analysis despite some of them possibly having had unobserved cubs
58 that did not survive until their first birthday. Our attribution of reproduction to adult females
59 associated with the lack of an effect of pride size on recruitment indicates that solitary
60 females struggle to raise cubs until their first birthday because they have to settle in poor-
61 quality habitats and suffer more from wounding (Packer 2023). Consequently, belonging to a
62 pride of at least two lionesses may be key to raising cubs until their first birthday, but two or
63 ten females does not make any discernible difference.

64

65 Overall, our definition of reproduction and recruitment leads to a lower number of females
66 with 0 recruited cubs in our data, and any underestimation of reproduction probability
67 subsequently leads to a corresponding overestimation in the recruitment per reproducing
68 female. As a result, the combined reproductive output remains consistent with the measure
69 used in previous analyses (e.g. Packer et al. 1990; Packer 2023), and investigating the
70 season-specific effects of density and habitat on each component of reproduction is still
71 possible—granted that the source of the bias is not correlated with these variables. Our
72 results thus indicate that the effect of the number of females on overall reproduction might
73 be more strongly influenced through probability of reproduction rather than recruitment.
74 Alternatively, the discrepancies between previous results and ours might arise because,
75 while our model does not account for differences in density effects between habitats or
76 across time, effects of pride size are largely driven by habitat quality, which has varied over

77 time (Packer 2023). Additionally—although we could not test for it—recruitment is driven to a
78 considerable extent by the ability of the resident coalition to fend off rivals (Bygott et al.
79 1979; Pusey & Packer 1994).

80

81 While results on male survival and takeover dynamics confirm previous findings, the
82 estimates on nomadic- and resident-male vital rates should be interpreted cautiously. In our
83 study population, lions are followed via the GPS localisation of prides and opportunistic
84 sightings of isolated and nomadic individuals (Borrego et al. 2018). Although capture-
85 recapture models enable to account for differences in detection probability (Lebreton et al.
86 1992; Lebreton & Pradel 2002), the lack of observed data—here more specifically on
87 nomad-resident transitions—can pose limitations on vital-rate estimations (Bailey et al. 2010;
88 Griffith et al. 2016). Similarly, lack of data on specific life-history stages and transitions can
89 limit the interpretation of density effects on demographic processes. For example, contrary to
90 previous findings (Elliot et al. 2014; Packer 2023), our analysis unexpectedly indicates a
91 positive effect of nomadic coalitions on the survival of young subadults and young males in
92 the dry season. Because our model only estimates apparent survival (i.e., does not
93 distinguish mortality from permanent emigration), an increase in young-male apparent
94 survival might be a consequence of a decrease in permanent emigration due to the pressure
95 exerted by high numbers of nomadic coalitions. However, changes in the detection of
96 nomads across the study period might bias the observed numbers of nomadic coalitions, as
97 nomadic males are only found opportunistically in the study area (Borrego et al. 2018). Such
98 limitations could be overcome by the use of combined capture-recapture and telemetry data
99 (e.g. Johnson et al. 2010; Bird et al. 2014), or of auxiliary data sources such as previous
100 publications or expert knowledge (e.g. Bauduin et al. 2020). Nevertheless, the interpretation
101 of current vital-rate predictions and population projections relying on them needs to take into
102 account the uncertainty in estimates (Fieberg & Ellner 2001; Ellner et al. 2002).

103

104 Posterior distributions for random year effects

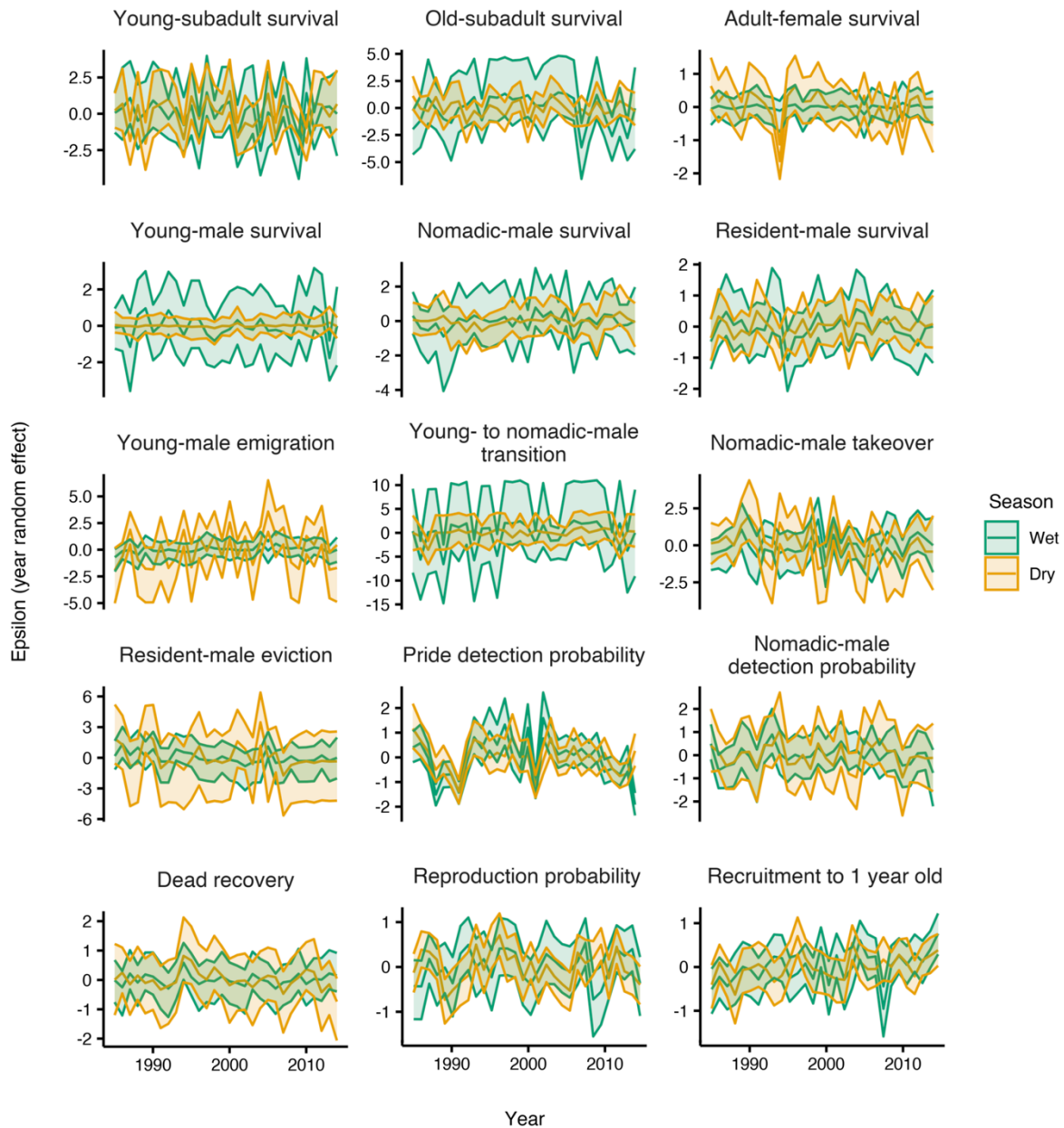
105

106 The season-specific yearly random-effect parameters showed seasonal differences in most
107 years for most vital rates and detection probabilities (Fig. S2), indicating a potential effect of
108 a seasonal variable our models did not explicitly account for. While we did not find any
109 noticeable temporal trend in the random effects, their variation was higher in the wet season
110 for most survival rates, and in the dry season for most transition rates. This stronger yearly
111 variation in specific seasons for specific groups of vital rates could be an indicator of
112 important seasonal factors that were not included in our model. Young- to nomadic-male
113 transition was an exception to this pattern, as the variability in random effects was much

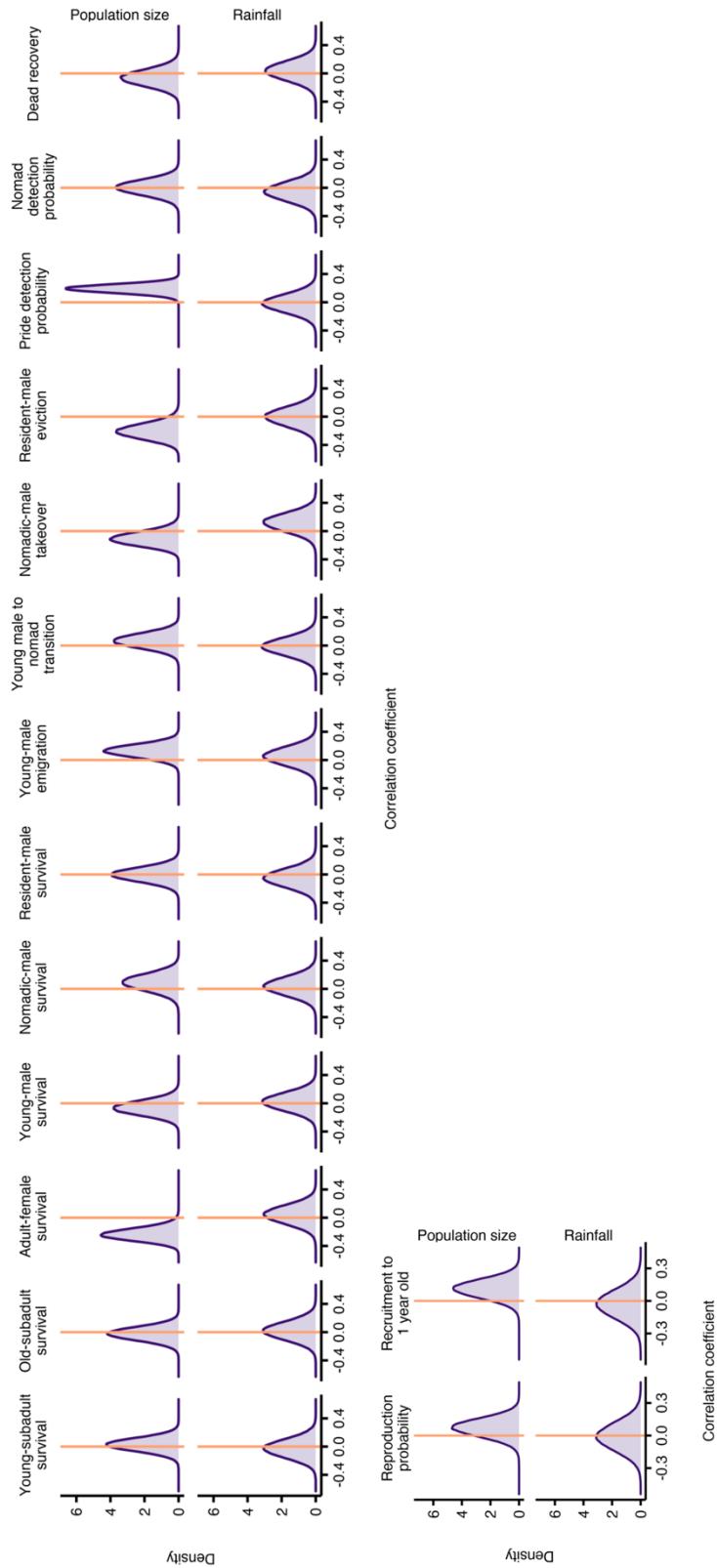
114 greater in the wet season. While the lack of data on this transition rate prevented us from
115 testing for the effect of density, this variability is likely due to a covariate linked especially to
116 the wet season that we did not explicitly account for.

117

118 Because of the complexity of our models and our decision to focus on the effect of socially-
119 explicit density measures, we could not include effects of overall population size and rainfall
120 in our multistate capture-recapture model and GLMMs. However, to assess a potential effect
121 of these two variables, we calculated the Pearson correlation coefficients between both
122 variables and every posterior sample of every vital rate and detection parameter. The
123 resulting posterior distributions of correlation coefficients indicate possible additional effects
124 of seasonal rainfall for all vital rates (Fig. S3). We also find evidence for potential effects of
125 overall population size on most vital rates, excluding adult-female survival, resident-male
126 eviction, and pride detection probability (Fig. S3). This gives additional indications of the
127 density independence of the survival of adult females and of the presence of strong density
128 effects on the other vital rates.



129 **Figure S2 - Season-specific yearly random effects.** In each model, we included a
 130 season-specific yearly random effect. For old-subadult survival, while the intercept depends
 131 on the sex, the random effect is shared for both males and females. The figure shows the
 132 season-specific mean random effect value (line) and the 90% credible interval (shaded
 133 ribbon) as a function of the year.



167 **Figure S3 - Distribution of the Pearson correlation coefficients between the**
 168 **season-specific random effects of each vital rate model and two covariates: rainfall**
 169 **and population size.** Posterior distributions (purple) of Pearson correlation coefficients
 170 between each MCMC sample of season-specific yearly random effects and potential

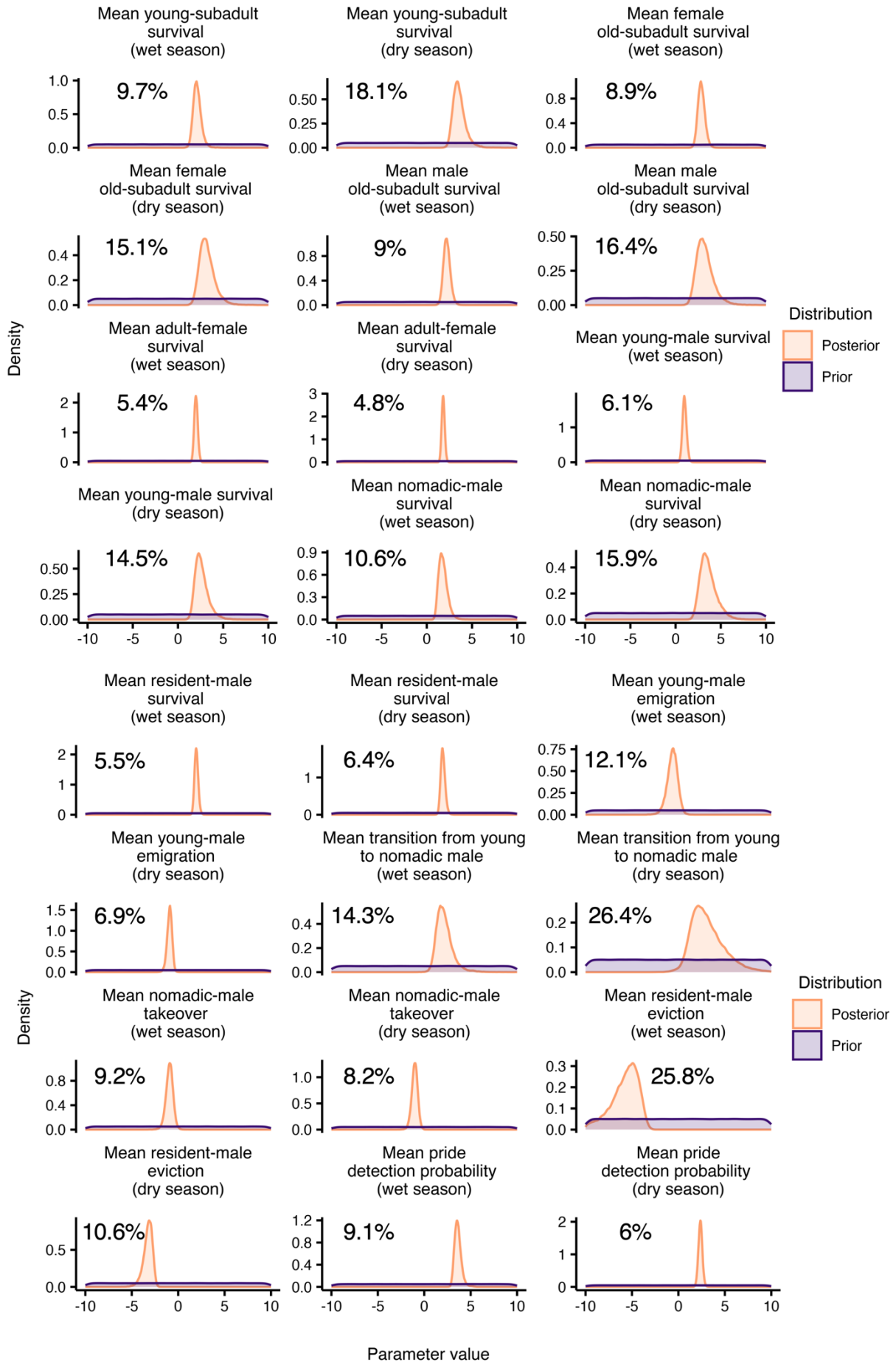
171 additional covariates overall population size and seasonal rainfall. The orange vertical line
172 marks “no correlation” (i.e., correlation coefficient = 0). The distribution of correlation
173 coefficients enables us to identify vital rates for which variation might be associated with
174 changes in population size or rainfall (no or a small overlap with the orange line, which
175 corresponds to a correlation coefficient of 0), or not (large overlap with 0).

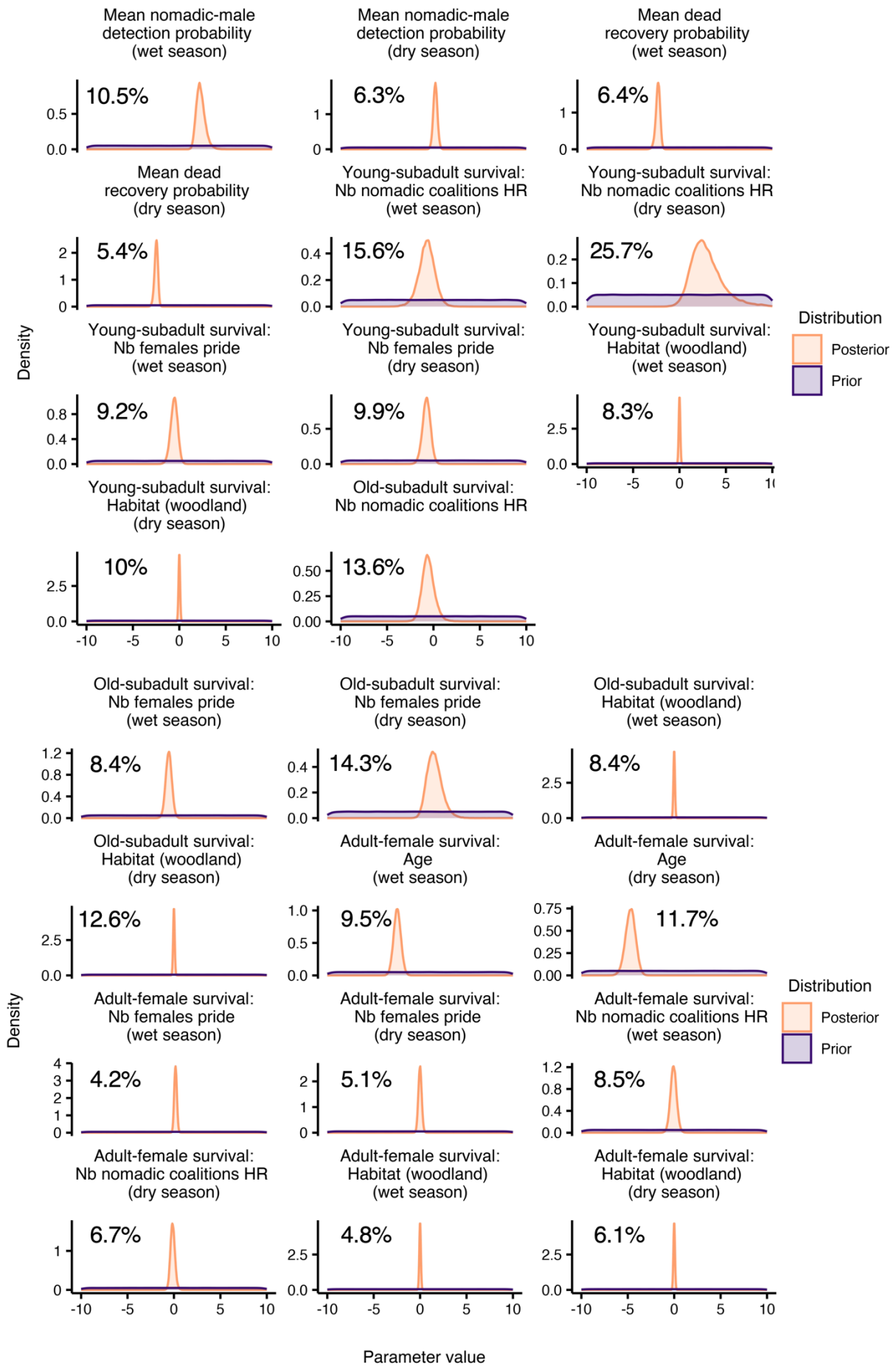
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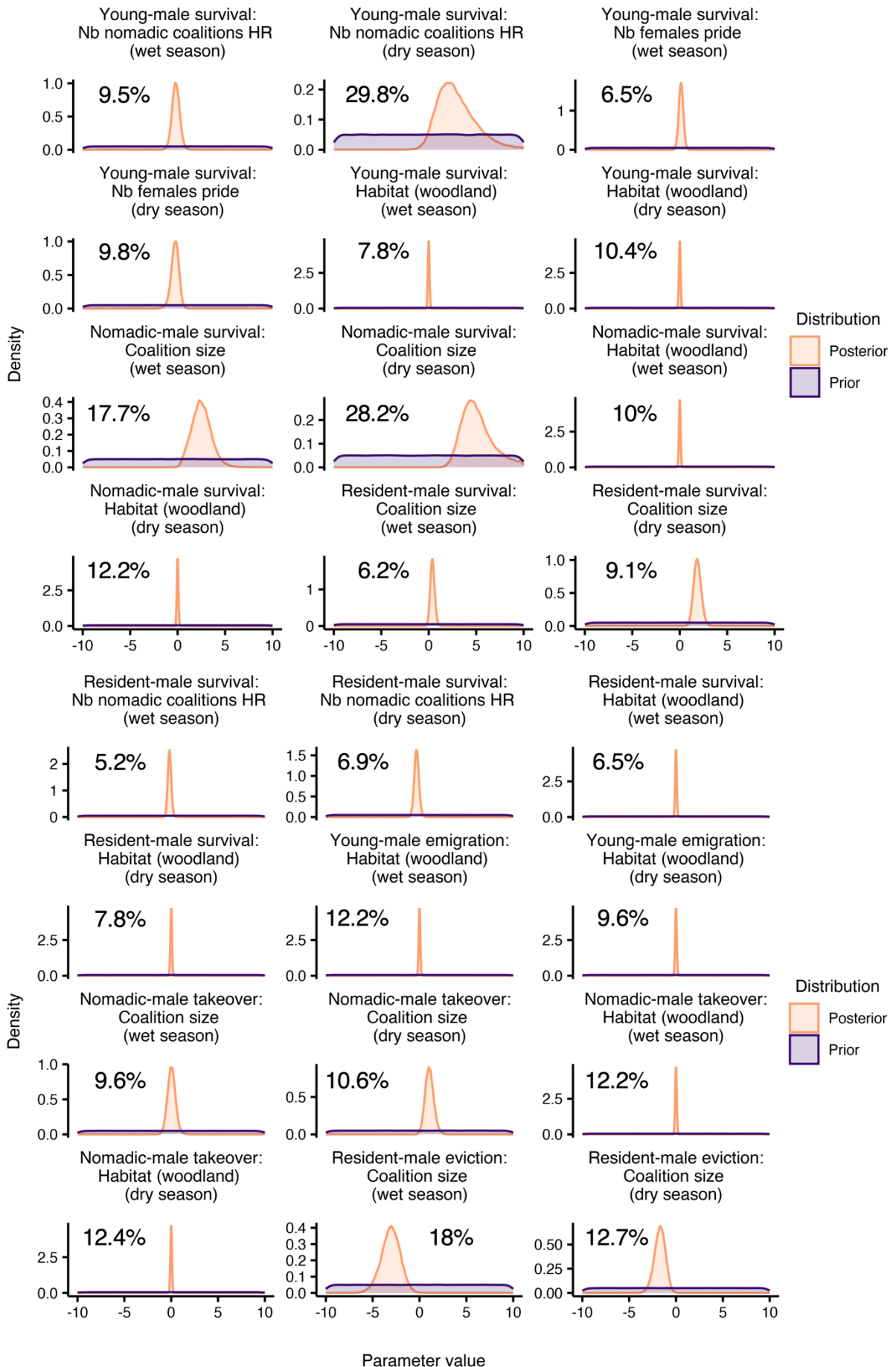
177 Extrinsic identifiability

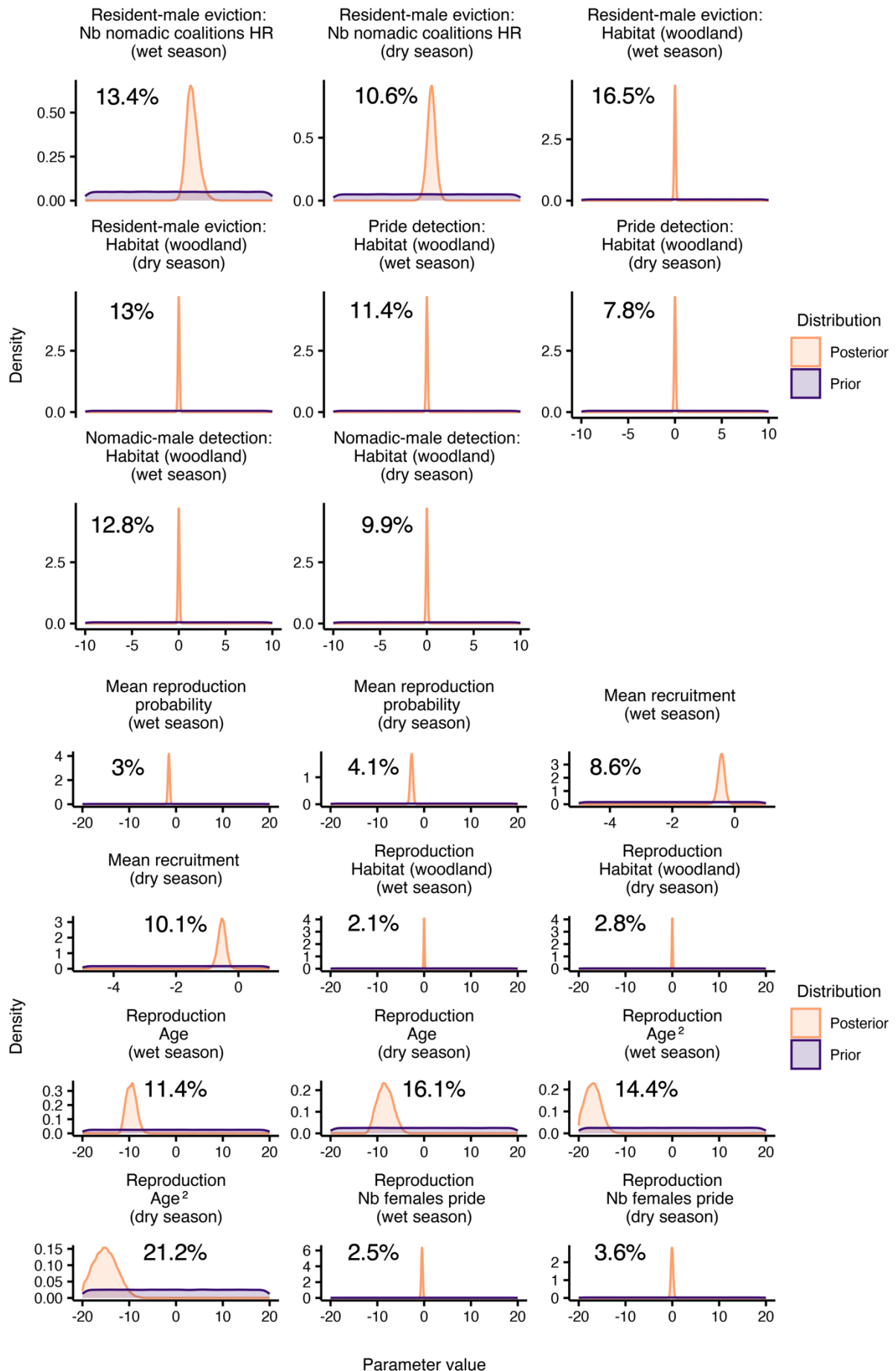
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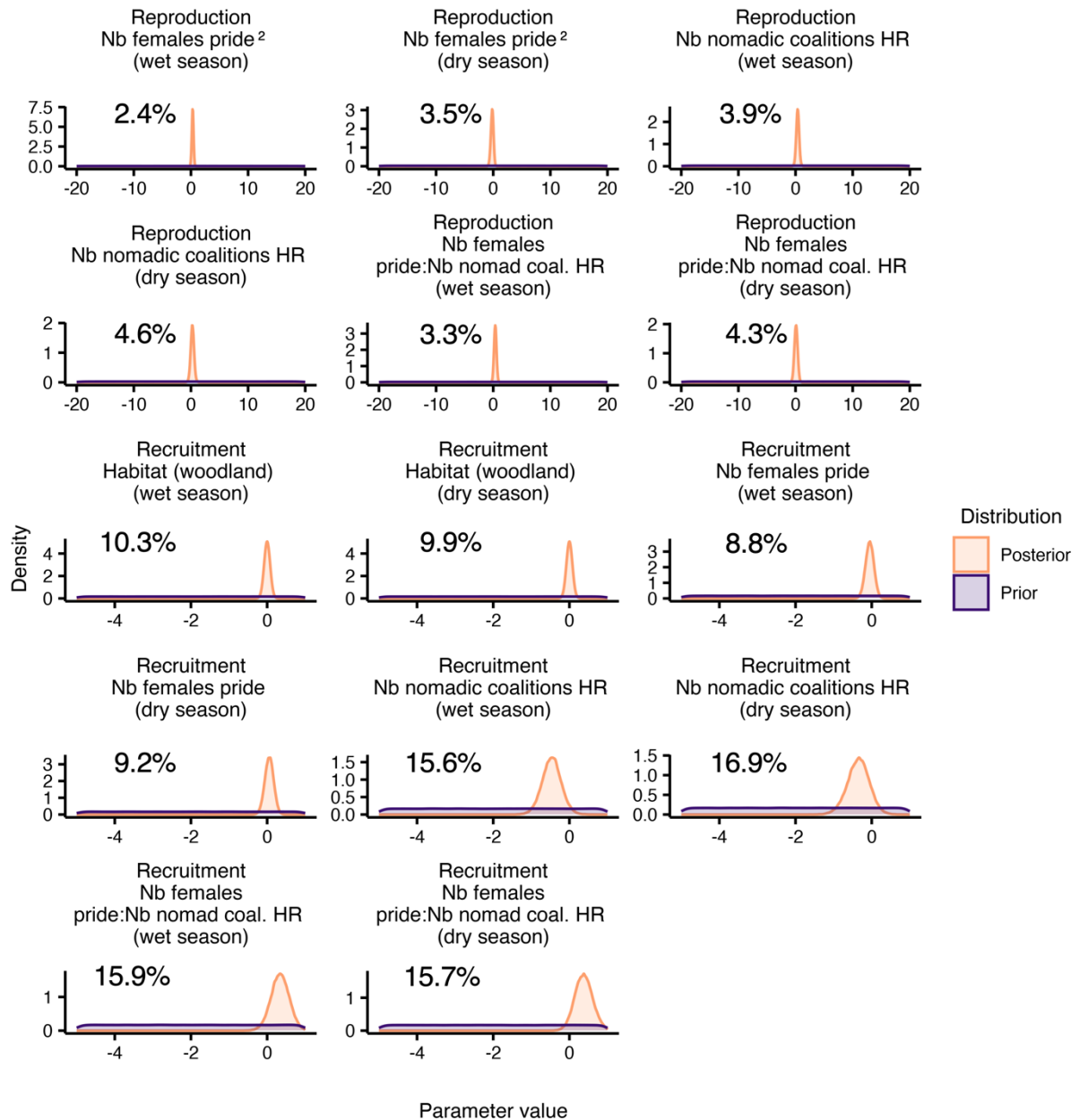
179 We assessed parameter extrinsic identifiability to detect near-redundancy in our model
180 parameters by calculating the overlap between the prior and posterior distributions (following
181 Garrett & Zeger 2000). For various classes of models, a parameter is commonly considered
182 as weakly identifiable when its prior and posterior distributions overlap by more than 35%
183 (Garrett & Zeger 2000; Gimenez et al. 2009). In our case, this threshold was reached for
184 none of the estimated parameters (Fig. S4), suggesting no major issues with extrinsic
185 identifiability for any of the parameters.











190 **Figure S4 - Overlap between the prior and posterior distributions of each**
 191 **estimated parameter.** For each estimated parameter, we assessed extrinsic identifiability
 192 by calculating the overlap between the prior (purple density plots) and the posterior
 193 distribution (orange density plots). A percentage of overlap above 35% indicates weak
 194 identifiability.

195

196 Posterior predictive checks

197

198 We assessed model fit for both the multistate capture-recapture model and the GLMMs by
 199 performing posterior predictive checks (Conn et al. 2018). We first defined a set of metrics to
 200 be calculated from the lion capture histories (e.g. total number of recaptures or number of

201 recaptures in a given state S , see below) and from the reproduction and recruitment data
202 (e.g. mean recruitment per female, see below). For each metric, we compared the observed
203 value to the distribution of values obtained from simulated datasets. To produce these
204 simulated datasets, we first sampled 500 sets of posterior values for each parameter of the
205 corresponding model—including random effects, which we did not re-sample from the
206 estimated standard deviations of the vital rate-specific random effects. For each sampled set
207 of parameters, we used observed covariate values to simulate 10 new reproduction and
208 capture-history datasets, for the latter starting from the true state of each individual on its
209 first capture. We therefore obtained 5000 simulated datasets for each model and calculated,
210 as for the observed data, the following metrics:

211

212 For the reproduction data:

213

- 214 ● Proportion of females reproducing
- 215 ● Mean age of reproducing females
- 216 ● Mean number of cubs (recruited to 1 year old) per reproducing female

217

218 For the capture histories:

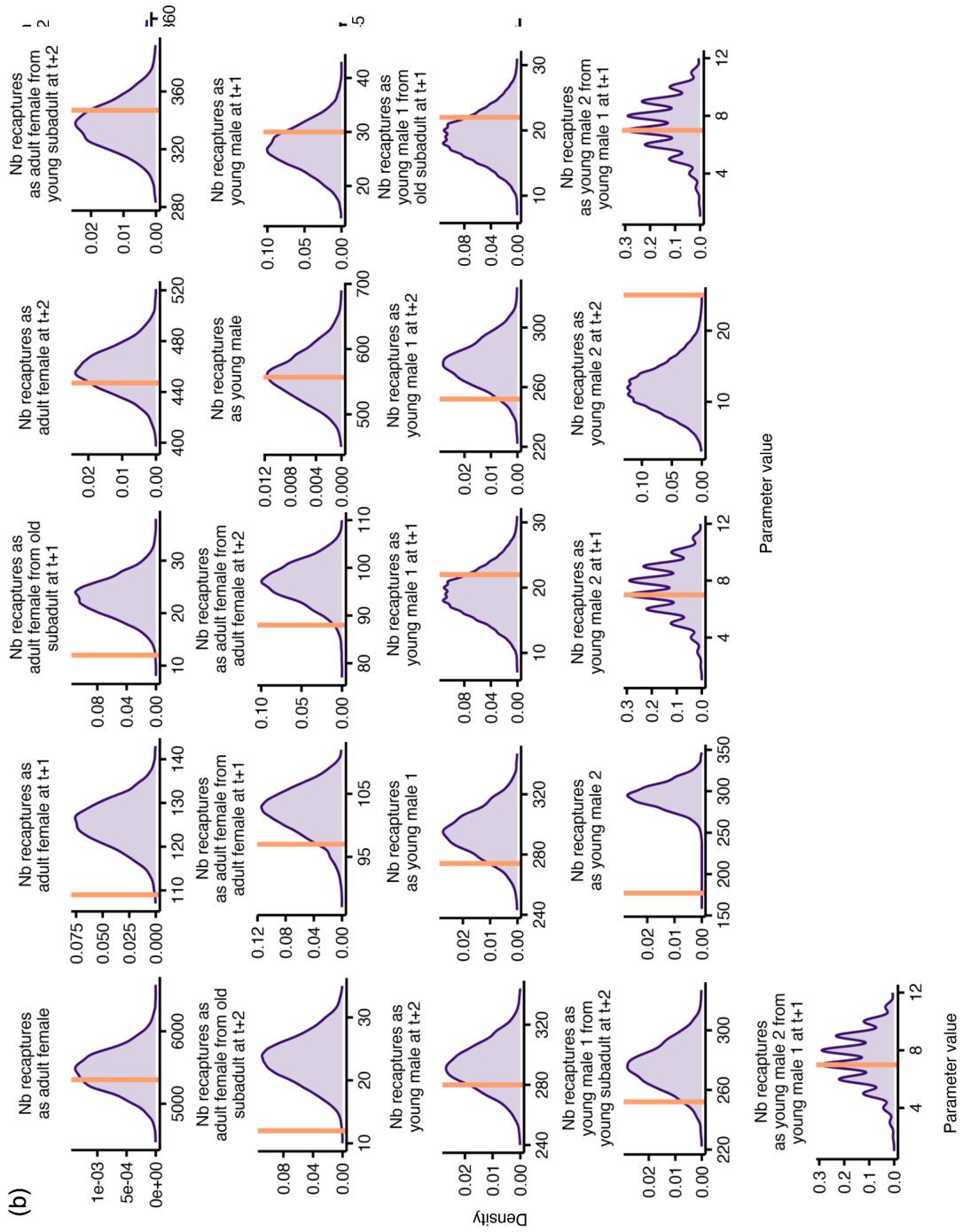
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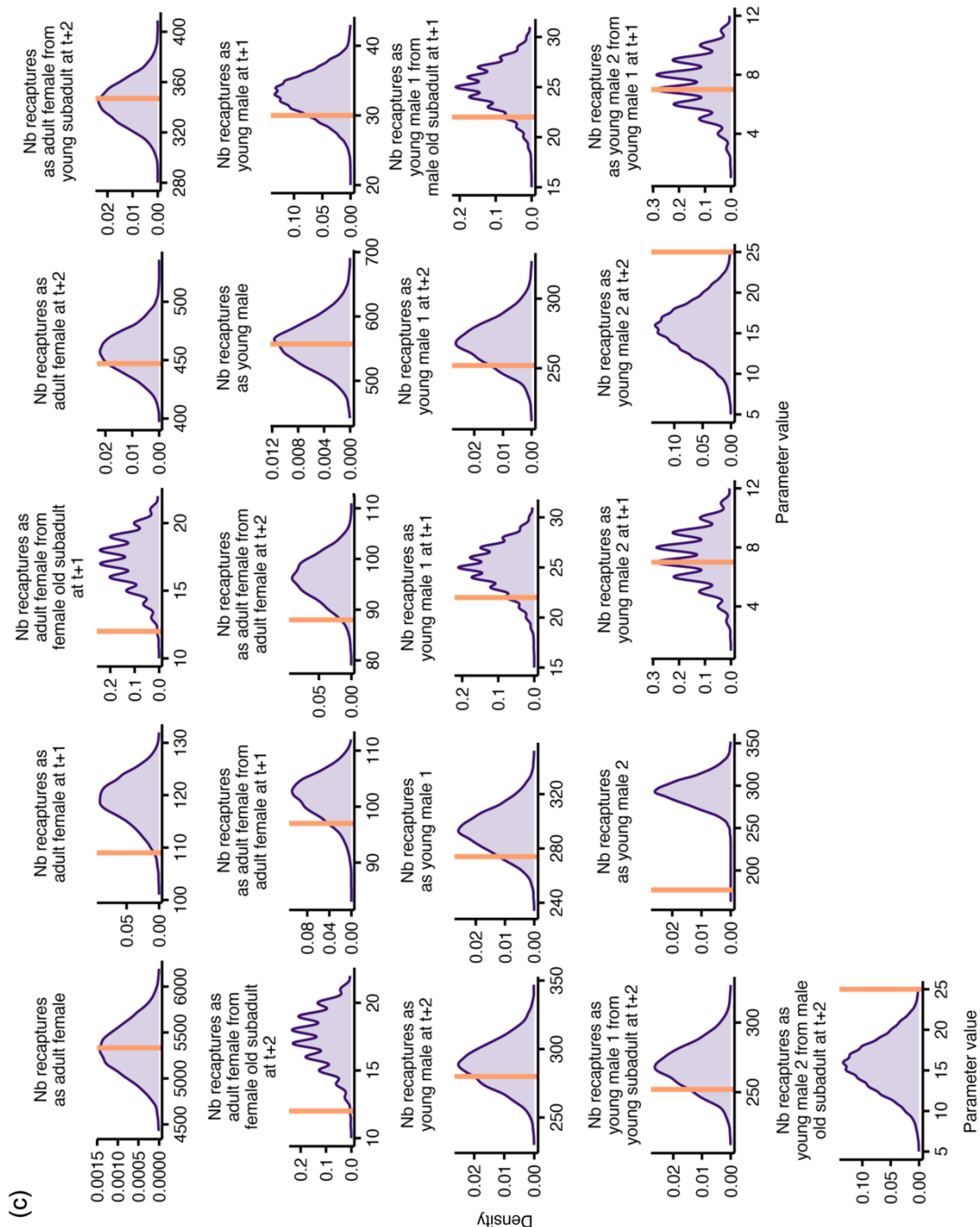
- 220 ● Total number of recaptures (overall, at $t+1$, and at $t+2$)
- 221 ● Number of recaptures as female old subadult (overall and at $t+1$)
- 222 ● Number of recaptures as male old subadult (overall and at $t+1$)
- 223 ● Number of recaptures as young male (overall, at $t+1$, and at $t+2$)
- 224 ● Number of recaptures in each of the four young-male stages (overall, at $t+1$, and at
225 $t+2$)
- 226 ● Number of male old subadults becoming young male 1 (at $t+1$)
- 227 ● Number of young subadults becoming young male 1 (at $t+2$)
- 228 ● Number of young male 1 becoming young male 2 (at $t+1$)
- 229 ● Number of male old subadults becoming young male 2 (at $t+2$)
- 230 ● Number of young male 2 becoming young male 3 (at $t+1$)
- 231 ● Number of young male 1 becoming young male 3 (at $t+2$)
- 232 ● Number of young male 3 becoming young male 4 (at $t+1$)
- 233 ● Number of young male 2 becoming young male 4 (at $t+2$)
- 234 ● Number of recaptures as nomadic male (overall, at $t+1$, and at $t+2$)
- 235 ● Number of male old subadults becoming nomadic males (at $t+2$)
- 236 ● Number of young male 1 becoming nomadic males (at $t+1$ and $t+2$)
- 237 ● Number of young male 2 becoming nomadic males (at $t+1$ and $t+2$)

- 238 ● Number of young male 3 becoming nomadic males (at $t+1$ and $t+2$)
- 239 ● Number of young male 4 becoming nomadic males (at $t+1$ and $t+2$)
- 240 ● Number of nomadic males becoming nomadic males (at $t+1$ and $t+2$)
- 241 ● Number of resident males becoming nomadic males (at $t+1$ and $t+2$)
- 242 ● Number of recaptures as resident male (overall, at $t+1$, and at $t+2$)
- 243 ● Number of male old subadults becoming resident males at $t+2$
- 244 ● Number of young male 1 becoming resident males (at $t+1$ and $t+2$)
- 245 ● Number of young male 2 becoming resident males (at $t+1$ and $t+2$)
- 246 ● Number of young male 3 becoming resident males (at $t+1$ and $t+2$)
- 247 ● Number of young male 4 becoming resident males (at $t+1$ and $t+2$)
- 248 ● Number of nomadic males becoming resident males (at $t+1$ and $t+2$)
- 249 ● Number of resident males becoming resident males (at $t+1$ and $t+2$)
- 250 ● Number of recaptures as adult female (overall, at $t+1$, and at $t+2$)
- 251 ● Number of female old subadults becoming adult females (at $t+1$ and $t+2$)
- 252 ● Number of young subadults becoming adult females (at $t+2$)
- 253 ● Number of adult females becoming adult females (at $t+1$ and $t+2$)
- 254 ● Number of dead recoveries

255

256 Comparing the observed and simulated values for each metric allowed us to determine
 257 which vital rate in the lion life cycle was poorly estimated by the two models and to improve
 258 the model accordingly. For example, an earlier model assuming an even (i.e. 0.5) female-to-
 259 male sex ratio led to an underestimated number of females in the simulated datasets
 260 compared to the observed capture histories (Fig. S5a). Adjusting the sex ratio to 0.55
 261 improved estimates of the number of females (Fig. S5b). In addition, a previous version of
 262 the model did not discriminate between male and female old subadults and assumed the
 263 same mean survival for both sexes. In that model, posterior predictive checks pointed to
 264 issues in transitions between subadults and adult females or young males. While estimating
 265 sex-specific mean survival rates for old subadults improved the precision and accuracy of
 266 predictions on the number of recaptured adult females, it did not improve predictions related
 267 to young males (Fig. S5c).



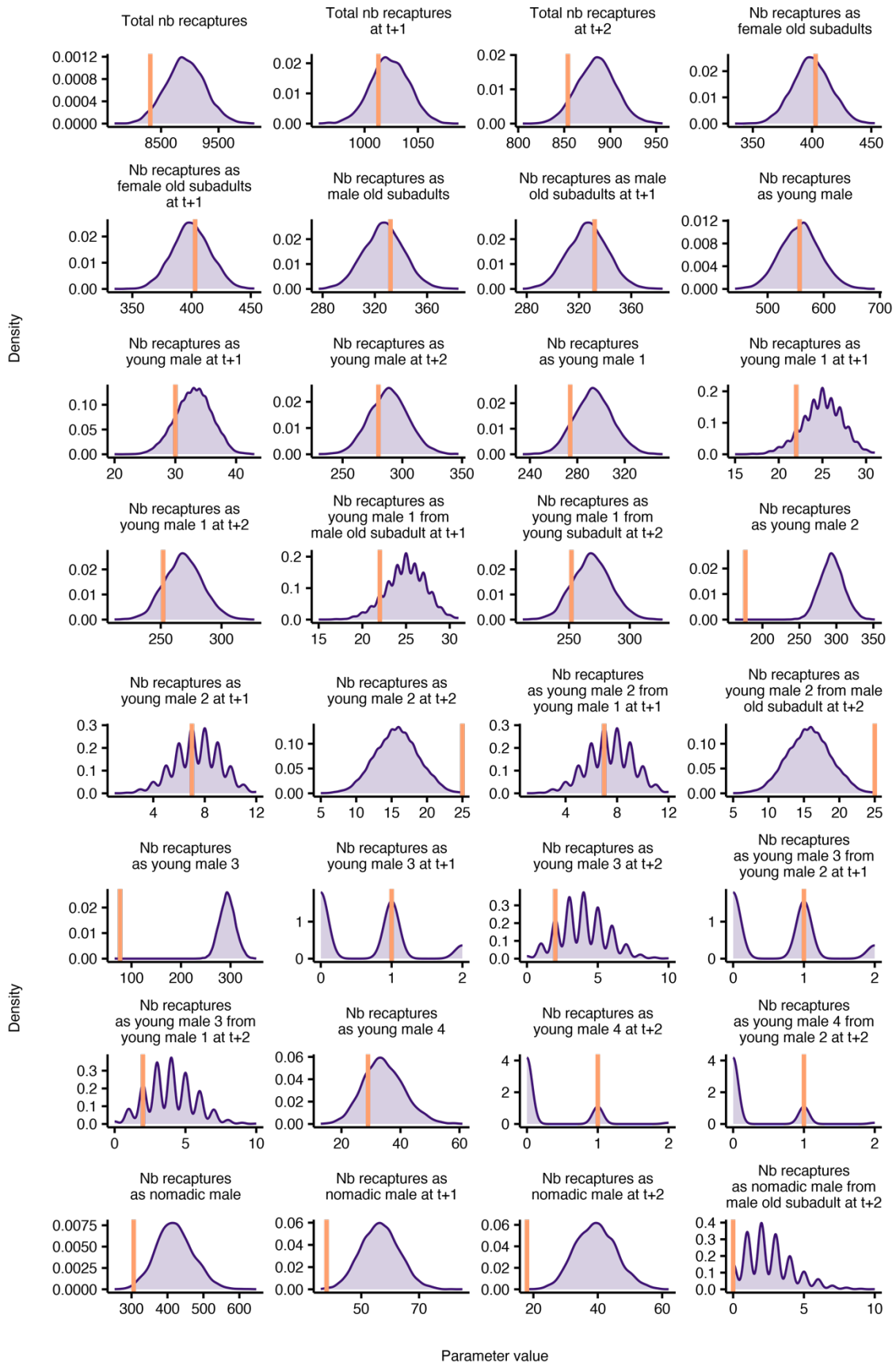


270 **Figure S5 - Simulated and observed values of metrics calculated on capture**
 271 **histories for the posterior predictive checks in three models with different structures.**
 272 We calculated a set of metrics on the observed data (orange vertical line) and the associated
 273 5000 simulated datasets (corresponding to 10 datasets simulated for each of 500 sets of
 274 sampled parameters; purple density plots). This figure compares the posterior predictive
 275 checks of three model assumptions: (a) Sex ratio of 0.5; (b) sex ratio of 0.55; and (c) sex
 276 ratio of 0.55 and sex-specific intercepts for the survival of old subadults.
 277 In the final model, for most metrics, the simulated distributions included the observed value
 278 (Fig. S6), and the Bayesian p-values (i.e., the proportion of simulated values higher than the
 279 observed value) were close to 0.5, indicating satisfactory fit (Fig. S7). However, some

280 discrepancies remain and should be discussed; mainly, the number of individuals recaptured
281 as young male 2–4 is greatly underestimated. This is likely a consequence of the limited
282 amount of data on transitions to and from young-male stages leading to issues estimating
283 the related parameters and thereby to discrepancies between the observed and simulated
284 values. In addition, the number of resident males becoming nomadic is overestimated, while
285 the number of nomadic males becoming resident is underestimated. This points to issues
286 estimating the parameters linked to takeover dynamics, indicating that more data is needed
287 to estimate such parameters properly. This could be achieved, for example, by integrating
288 additional data sources, such as telemetry data, or expert knowledge to increase information
289 about when males leave or join a pride (Johnson et al. 2010; Bird et al. 2014; Bauduin et al.
290 2020). Overall, parameters linked to young, resident, and nomadic males, as well as future
291 population projections relying on the predictions of these vital rates should be interpreted
292 with caution.

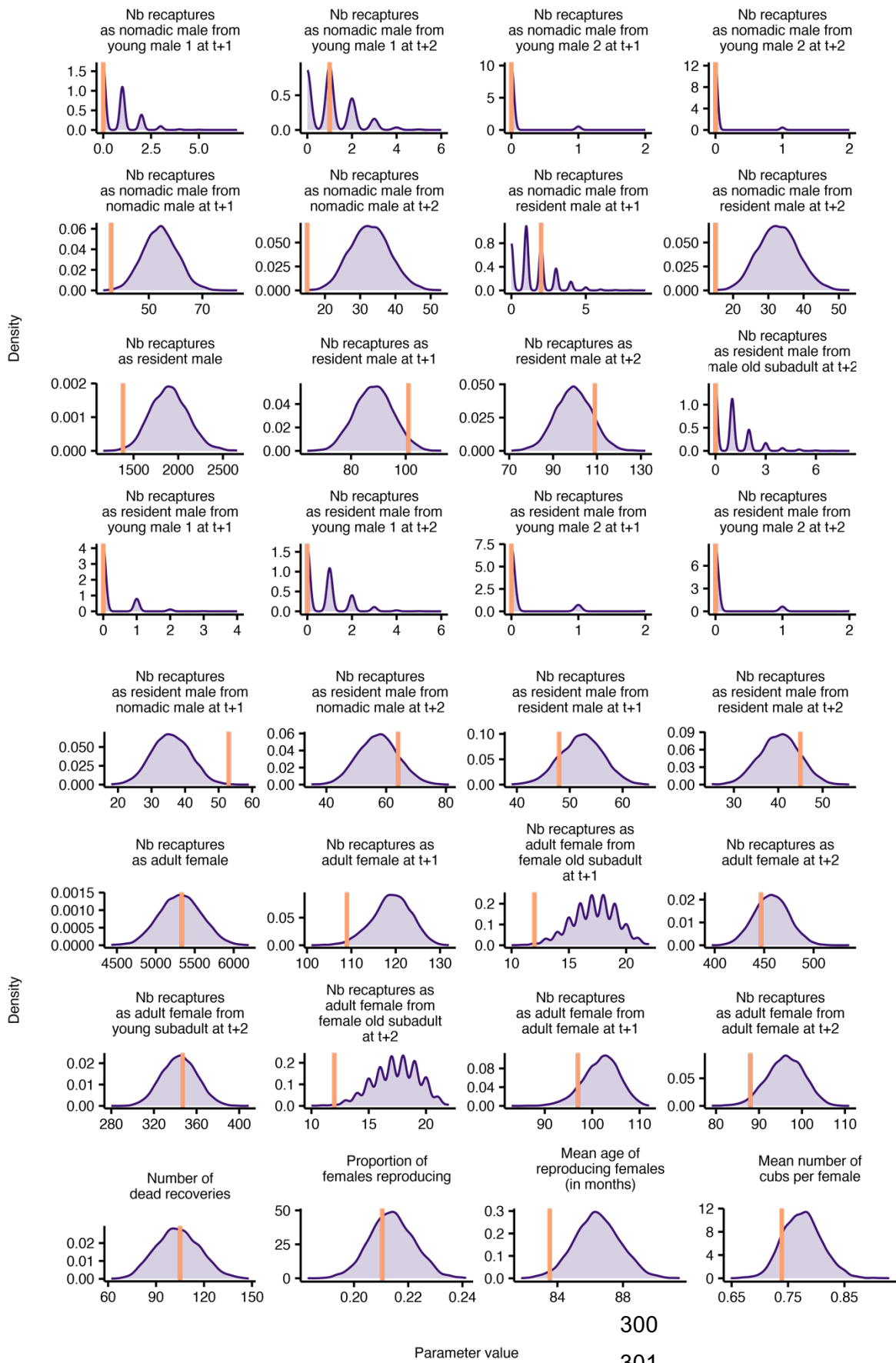
293

294 In addition, we used the posterior distributions of the parameters defining reproduction rates
295 to predict the season-specific reproduction probability and recruitment in each year. The
296 predicted values and 95% credible intervals do correspond to the observed values (Fig. S8),
297 giving further indication of a good model fit.



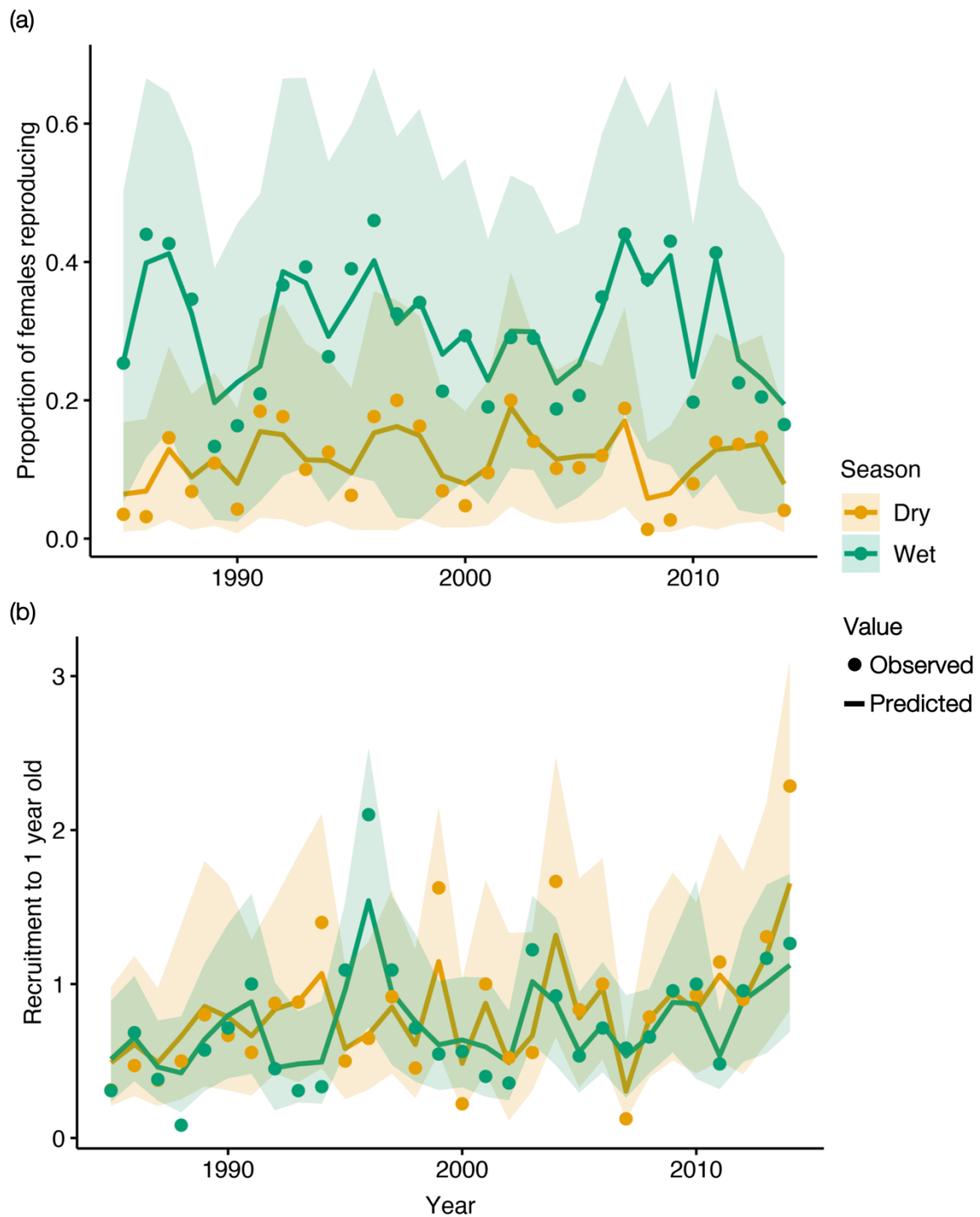
298

299



303 **Figure S6 - Simulated and observed values of metrics calculated on capture**
304 **histories and reproduction data for the posterior predictive checks.** For the capture
305 histories and the reproduction dataset, we calculated a set of metrics on the observed data
306 (orange vertical line) and the associated 5000 simulated datasets (corresponding to 10
307 datasets simulated for each of 500 sets of sampled parameters; purple density plots).
308

310 **Figure S7 - Bayesian p-values of each metric used for the posterior predictive**
311 **checks.** For each metric calculated on 5000 simulated capture histories and reproduction
312 datasets, we computed the Bayesian p-value (i.e., the proportion of simulated values higher
313 than the observed value). (a) For metrics associated with a given timestep t (i.e. $t+1$ and
314 $t+2$), we calculated one p-value for each t of the capture history, obtaining 59 p-values for
315 metrics calculated at $t+1$ and 58 for those calculated at $t+2$. (b) For metrics associated with
316 the whole dataset, we only calculated one p-value. The orange horizontal line corresponds
317 to a p-value of 0.5, indicating a perfect correspondence between the observed and simulated
318 metric.



319 **Figure S8 - Observed and predicted reproduction probability and recruitment.**
 320 For each year, we predicted the season-specific proportion of (a) females reproducing in the
 321 population and (b) the season-specific recruitment (i.e., number of cubs reaching one year
 322 old per female) using the posterior distributions of the parameters defining these
 323 reproductive rates to compare our mean model predictions (lines) and their 95% credible
 324 intervals to the observed data (dots).

325 **References – Appendix S5**

326

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