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Neglected biodiversity of fish assemblages associated with Antipatharia (black corals) on tropical shallow reef ecosystems

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Abstract

Addressing anthropogenic threats compromising the persistence of tropical marine ecosystems requires an understanding of the fundamental ecological functions these organisms fulfil. Habitat provision is a major function of corals in tropical marine ecosystems, although most research in this area has concentrated on scleractinians (hard corals). Here, we provide one of the first empirical studies of fish communities on shallow tropical reefs associated with another, lesser-known hexacoral group – the antipatharians (black corals). We quantify i) the abundance, and taxonomic and functional diversity of fish communities associated with antipatharians, and ii) the type of associations between the fish and the antipatharian colonies. Surveys were conducted on an artificial reef (*SS Yongala* shipwreck) and on a coral reef (Orpheus Island) in the central Great Barrier Reef, Australia. We documented 28 different species of fish within seven trophic groups and 23 functional entities associated with antipatharians, predominantly using the colonies as shelter. Antipatharians support both taxonomically distinct fish assemblages (>40% of species) and unique types of associations with the fishes. At the functional level, we observed large overlap in the fish community between antipatharians and scleractinians, reflecting

35 their shared ecological roles, although antipatharians support significantly higher functional diversity.
36 Given the similarity in functional composition of fish assemblages utilising both antipatharians and
37 scleractinians, the presence of antipatharians may help buffer the effects of ongoing hard corals decline
38 in tropical marine ecosystems. Overall, our study provides empirical evidence of the important role of
39 antipatharians in supporting fish functional and taxonomic diversity on shallow tropical reefs.

40

41 **Introduction**

42 Identifying, understanding and maintaining ecological functions is essential to sustaining ecosystems
43 in the face of current anthropogenic stressors (Bellwood et al., 2004a; Brandl et al., 2019; Hughes et
44 al., 2017a). Traditionally, ecological studies have focussed on the taxonomic composition of
45 assemblages. However, there has been an increasing focus on understanding the functional roles of
46 species in recognition of the fact that relatively few taxa perform key ecological functions (Bellwood
47 et al., 2004a; Harborne et al., 2017; Naeem et al., 2012). While species richness plays an important role
48 in buffering reef ecosystem functions (Lefcheck et al., 2021), the presence and abundance of
49 functionally important species are particularly important for ecosystem resilience (McGill et al., 2006).
50 For example, on coral reef ecosystems, a global analysis of unique trait combinations of fish showed
51 that, in the Central Indo-Pacific, about one-third of ecological functions are provided by only one
52 species (Mouillot et al., 2014).

53 Structural complexity is one of the most important ecological traits on reefs and has been
54 associated with key coral taxa (Darling et al., 2017; González-Barrios & Álvarez-Filip, 2018; Kerry &
55 Bellwood, 2015a). Three-dimensional habitat strongly influences the composition and diversity of a
56 range of reef-associated taxa, particularly fishes (Jones & Syms, 1998; Wilson et al., 2009). Reef
57 complexity has been shown to influence species richness, density, biomass and trophic structure of fish
58 assemblages (Behrens, 1987; Beukers et al., 1997; Darling et al., 2017; Gratwicke & Speight, 2005).
59 The abundance and size of holes or cavities mediates predation dynamics and juvenile fish survivorship,
60 thereby influencing the composition of fish communities in different ways (Almany, 2004; Darling et
61 al., 2017; Lingo & Szedlmayer, 2006). Habitat complexity on coral reefs can be provided by both the
62 underlying reef substrate and by habitat-forming sessile benthos, such as corals, algae and sponges. The
63 loss of habitat-forming benthos and resulting loss of habitat complexity therefore compromises the
64 ecological functioning of coral reefs (Graham & Nash, 2013) and makes them less likely to recover
65 from disturbances (Graham et al., 2015).

66 Most studies examining the importance of habitat-forming benthos on coral reefs have focussed
67 on scleractinians (hard corals). For example, the abundance of scleractinians with complex growth
68 forms (e.g. *Acropora* and *Pocillopora*) is often correlated with the composition of fish communities on
69 shallow reefs (<30 m depth) (Beukers et al., 1997; Darling et al., 2012; Kerry & Bellwood, 2015a).

70 Similarly, a study in the Great Barrier Reef (GBR) showed that tabular *Acropora* spp. had
71 disproportionate effects on the distribution of large reef fish communities, even when that morphology
72 constituted a small fraction (4%) of the total benthic cover (Kerry & Bellwood, 2015a). Moreover, a
73 branching morphology can provide fine-scale structural complexity for small-bodied and/or juvenile
74 fishes to refuge from predators (Beukers et al., 1997). Despite the importance of coral morphology and
75 size (Fisher, 2023; Zawada et al., 2019), most coral reef monitoring programs only document live coral
76 cover, without considering structural complexity.

77 Structural complexity at reefscape scales can be estimated by visual scores (Gratwicke &
78 Speight, 2005), although these approaches can be easily influenced by surveyor perspectives. More
79 recently, photogrammetry has enabled quantitative analysis of reef structural complexity (Ferrari et al.,
80 2016; Friedman et al., 2012; Kornder et al., 2021). Moreover, photogrammetry has been used to quantify
81 the total volume of shelter (habitat) provided by different scleractinian growth forms or ‘shelter
82 volumes’; and predictive models of shelter volume (a 3D metric) can be estimated based on 2D metrics
83 (area or diameter) for each major growth form (Aston et al., 2022; Urbina-Barreto et al., 2021).
84 Therefore, it is currently possible to quantify the shelter volume of different coral morphologies and
85 investigate the link between coral complexity and reef fish abundance at finer spatial scales (Urbina-
86 Barreto et al., 2022). To date, this research has focussed almost entirely on scleractinians, with no
87 attempts to quantify the importance of other habitat-forming benthic groups in providing habitat
88 complexity.

89 Antipatharians - commonly known as black corals – are a sister group to the scleractinians,
90 within the class Hexacorallia. Antipatharians occur in all worlds’ oceans except for the Arctic, at depths
91 ranging from 1 to 8,900 m (Molodtsova et al., 2008; Pasternak, 1977; Wagner et al., 2012). Unlike
92 scleractinians, antipatharians do not produce a calcium carbonate skeleton but a thorny axial skeleton
93 (brown or black in colouration) composed of different scleroproteins (Goldberg, 1978; Goldberg et al.,
94 1994). Antipatharians have a range of morphologies including flabellate (fan-like), whip-like, bottle-
95 brush-like, and branching (which can be either small bush-like or large arborescent colonies) (Wagner
96 et al., 2012). Despite limited studies on the topic, it is known that antipatharians provide important
97 habitat complexity supporting an array of fish. For example, (Boland & Parrish, 2005) examined the
98 diversity and movement patterns of fish associated with branching antipatharians between 52 m and 73
99 m depth in Hawaii. While their study was conducted in a mesophotic coral ecosystem (MCEs; 30-150
100 m depth reefs), 95% of the fish recorded also occur on shallow reefs (Boland & Parrish, 2005). A study
101 on the subtropical eastern Atlantic on antipatharian forests (i.e. dense aggregations of branching
102 antipatharian colonies) found that 90% of fish functional richness inhabiting antipatharian forests at
103 mesophotic depths, were also found on shallow reefs, although the dominant species varied between
104 shallow and mesophotic depths (Bosch et al., 2023). In temperate mesophotic ecosystems (TMEs) in

105 the Mediterranean Sea, an array of fishes – including species of both conservation interest and high
106 commercial value – were associated with antipatharian forests (Chimienti et al., 2022)

107 Despite the clear importance of antipatharians as habitat for a wide range of fish species across
108 a range of ecosystems, there is currently little information on their role on shallow tropical reefs.
109 Moreover, no studies have examined whether antipatharians host a different fish community to the one
110 in association with neighbouring scleractinian corals. Here, we provide the first assessment of the fish
111 community structure associated with antipatharians in shallow reef ecosystems, and explore how this
112 previously overlooked benthic taxon influences fish communities on these reefs. We quantified the fish
113 species richness and density, and recorded the behaviour of fishes in close association with both
114 antipatharian and scleractinian colonies at two sites in the central Great Barrier Reef (GBR) to
115 investigate: i) the fish community structure associated with antipatharians; and ii) the effects of coral
116 taxon, corals complexity properties (area and shelter volume), and reef sites, on the fish communities.
117 This information aims to improve our understanding of the role of antipatharians in supporting
118 functional and taxonomic diversity on reefs.

119

120 **Methods**

121 *Field sites:* Multiple *in situ* surveys were conducted at two locations on the central Great Barrier Reef,
122 Queensland Australia, between May and October 2021: the SS *Yongala* wreck and Orpheus Island. The
123 *Yongala* is 107 m long wreck located ~22 km from mainland (-19.291, 147.627) and sits between 14 m
124 and 29 m depth. The wreck is a world-renowned dive-site for its high fish abundance; however, with
125 the exception of one study of the fish species richness conducted in the late 1990's (Malcolm et al.,
126 1999), scientific studies of the abundance and diversity of fish and benthic fauna are lacking. The wreck
127 supports both antipatharian and scleractinian corals and therefore represents a great opportunity to
128 investigate the influence of both coral taxa on fish communities. Because the *Yongala* is in essence an
129 'artificial' reef, and to explore the generality of our results across shallow coral reefs, we also collected
130 data on a well-studied shallow reef that, like *Yongala*, supports interspersed populations of
131 antipatharians and scleractinians between 13 m and 16 m depth - Orpheus Island (18.616, 146.519, at
132 Little Pioneer and Iris Point). *Yongala* and Orpheus Island are ~142 km apart (Map in Supplementary
133 1a). Both sites are within No-Take Marine Protected Areas (Marine Park zones), and *Yongala* is a
134 Commonwealth Cultural Heritage Site.

135

136 *Corals area and shelter volume:* Photos with a scale were taken to estimate the planar area of each of
137 the coral colonies using the software ImageJ (Bourne, 2010). For scleractinians, we recorded the planar
138 area as viewed from above, which is the traditional approach (Rogers C. G.; Garrison R.; Grober Z. M.
139 H; Franke M. A., 1994) and also the method used by (Urbina-Barreto et al., 2021) to develop predictive

140 models of shelter volume. For antipatharians, planar area was calculated based on width (diameter) and
141 height of the colonies as viewed from the side, which are considered the best estimators of surface area
142 for non-scleractinian branching bushy-like coral colonies (Santavy et al., 2013). The shelter volume
143 (dm^3) of all coral colonies was calculated using the predictive models (based on the colonies diameter)
144 of (Urbina-Barreto et al., 2021) for branching, massive and tabular colonies. No predictive models are
145 available for encrusting and foliose growth forms; thus these two morphologies were treated as massive.
146 Predictions of shelter volume were made using log-scale colony diameters, which were nearly identical
147 to the predicted shelter volumes when using area (Shelter volume figures and images of coral
148 morphologies are presented in Supplementary 1b).

149

150 *Fish surveys:* Four-minute long stationary videos of both antipatharian and scleractinian colonies were
151 filmed at 30 fps on SCUBA during daylight hours (1100 - 1400). Coral colonies were stochastically
152 (i.e. as encountered on the dives) filmed in pairs (one for each coral taxon) – where the antipatharian
153 and scleractinian were at the same depth, ≤ 10 m apart, and filmed at the same time or one immediately
154 after the other. At *Yongala*, 17 coral pair videos were filmed (34 colonies) at two depth ranges: 14-20
155 m & 21-27 m (eight and nine colony pairs respectively). At Orpheus Island, scleractinians are not
156 abundant beyond 14 m depth; therefore, six coral pair videos (12 colonies) were filmed at 14 m depth,
157 where both coral taxa coexist, and to maintain a similar depth range to the other site.

158

159 *Video analysis:* We identified observed fish communities to species and recorded the maximum number
160 of fish individuals of a particular species that occurred in a single frame of a video (MaxN) using
161 EventMeasure (SeaGIS, Melbourne Australia), which allowed us to estimate the abundance of each fish
162 species observed. This represents a conservative estimate of the minimum number of individuals known
163 to have been present in the sampling area (each colony) over the filming period and provides the
164 maximum number of individuals of a particular species occurring in a single frame of a video (Lyle et
165 al., 2007). We also recorded the ‘behaviour’ of each fish associating with coral colonies as follows:

- 166 • *HovA* – hovering around (< 50 cm around the colonies)
- 167 • *HovH* – hovering around and hiding (hovering < 50 cm around the colonies and seeking refuge
168 among the coral structure)
- 169 • *Stat_next* – static (resting next to the coral colonies)
- 170 • *Stat_in* – static (static within or on the coral colonies)
- 171 • *Feed* – feeding on polyps (in the case of corallivores), or feeding on algae on top of coral (e.g.
172 *Scarus*)
- 173 • *Clean* – being cleaned (by *Labroides dimidiatus*)
- 174 • *Pass* – passing by

175 Filming of footage initialised immediately after encountering the colonies, and for video analysis
176 fish were not counted for the first ten seconds. This approach maximises the data being collected –
177 considering we had time constraints of working at deeper depths on *Yongala* – while optimising
178 the likelihood of detecting both cryptic and mobile species that may not be present if longer wait
179 times were used (Bohnsack & Bannerot, 1986; Willis, 2001). This approach was standardised
180 across all videos (filming and analyses). Fishes passing by (*Pass*) were recorded, however these
181 were not considered for further analysis because of the uncertainty of their association with the
182 corals. For instance, fish passing could have been foraging but were not observed consuming their
183 prey.

184

185 *Statistical analysis:* All analyses were conducted in R 4.4.0 (R Core Team, 2024). The shelter volume
186 (dm^3) of all coral colonies was calculated using the predictive models proposed by Urbina-Barreto et
187 al. (2021) as described above. To standardise fish abundance and species richness, we used the area of
188 the surveyed coral colonies (i.e. we used fish density m^{-2} and fish richness m^{-2}).

189 To explore the effect of depth, we fit generalised linear mixed-effects models (GLMMs)
190 examining the effect of the two depth bands at *Yongala* (14–20 m and 22–27 m) and coral order
191 (Antipatharia and Scleractinia) on fish density m^{-2} and species richness m^{-2} . GLMMs were fit using
192 restricted maximum likelihood with a gamma distribution (log-link) and a random effect of colony pair
193 ID. For each model, we included the effect of coral order and fit models using a full interaction term.
194 Where interactions were significant, we examined pairwise contrasts and split the analysis by coral
195 order for subanalyses using generalised linear models (no pair ID effect). The effect of depth on fish
196 species richness m^{-2} for scleractinians did not meet the assumption of homogeneity of variance;
197 therefore, we fit a generalised least squares (GLS) model with the variance function ‘varIdent’ to
198 account for heteroscedasticity across groups (i.e. unequal variances of species richness m^{-2} at each depth
199 band). To investigate the effect of site and coral taxa on fish density m^{-2} and species richness m^{-2} , we
200 again fit gamma log-link GLMMs using colony pair as random effect. Additionally for both depth and
201 site models, we created AICc model selection tables that confirmed full-factorial model results (see
202 Supplementary 2).

203 To quantify reef fish functional diversity supported by antipatharian and scleractinian colonies,
204 we constructed a multidimensional trait space using the R package mFD (v1.1.2; Magneville et al.,
205 2022). We selected six ecologically relevant traits reflecting key aspects of fish functional roles on coral
206 reefs: fish body size (0–7 cm, 7.1–15 cm, 15.1–30 cm, 30.1–50 cm, 50.1–80 cm, and >80 cm), diet
207 category (herbivore-detritivores, macroalgae-herbivores, sessile invertivores, mobile invertivores,
208 planktivorous, piscivores and omnivores), mobility (sedentary, mobile within a reef or mobile between
209 reefs), activity period (diurnal, diurnal and nocturnal or nocturnal), group size (solitary, pairing, small
210 groups of 3–20 individuals, medium groups of 20–50 individuals, or large groups of >50 individuals)
211 and vertical position in the water column (benthic, benthopelagic or pelagic). Fish traits were chosen

212 following the framework developed by Mouillot et al. (2014) and compiled from FishBase (Froese &
213 Pauly, 2015) and supplemented with values from primary literature where database entries were missing
214 or ambiguous. We calculated pairwise Gower distances between functional entities which
215 accommodates mixed trait types and allocates equal weighting (Legendre & Legendre, 2012). A
216 Principal Coordinates Analysis (PCoA) was then applied to the distance matrix using the first four
217 principal axes. The number of axes to construct functional trait space was chosen based on mean
218 absolute deviation (msD) based the deviation between initial trait-based distances and distances in the
219 functional space (Maire et al., 2015). To visualise this analysis, we plotted convex hulls of functional
220 trait space from fish assemblages on antipatharians, scleractinians and the global trait space (all colonies
221 combined) using ggplot (Wickham, 2016). Data, analyses and summary table of fish functional analyses
222 are available in Supplementary 3. Additionally, fish functional diversity indices were estimated as
223 follows: a) a GLMM was fit using a gaussian distribution with site as random effect, to assess mean
224 functional entities (i.e., unique combinations of functional traits); b) Functional redundancy (i.e.,
225 multiple species performing same or similar ecological roles) was estimated using a Generalized
226 Additive Models for Location, Scale, and Shape (GAMLSS) model with a gamma distribution to
227 account for under-dispersed and skewed data.

228 All GLMMs were fit using restricted maximum likelihood (REML) via the package *glmmTMB*
229 (Brooks, 2022) and model diagnostics (i.e., assumptions of normality, homogeneity or variances, no
230 overdispersion) were assessed using the package *DHARMA* (Harting, 2022). Post-hoc analysis,
231 estimated marginal means, and pair-wise contrast were done using the package *emmeans* (Lenth et al.,
232 2022); and predicted values were calculated using the *predict* function with 95% confidence intervals.
233 The GAMLSS model was fit using the *gamlss* package (Stasinopoulos & Rigby, 2007) and residuals
234 were examined using the function *residuals* in the same package. All models' formulas, results and
235 summary statistics – marginal means, and 95% confidence intervals are reported in Supplementary 2.

236 To visualise the overall significance of the different variables (site, coral taxa, shelter volume)
237 on explaining the fish community structure, a constrained ordination using distance-based redundancy
238 analysis (dbRDA; Legendre & Andersson, 1999) was conducted using the package *vegan* (Oksanen,
239 2022), with the variables overlaid as a vector. This was followed by a permutation-based multivariate
240 ANOVA (PERMANOVA) of the dbRDA to identify significant ($p < 0.05$) variables driving the fish
241 community structure. Then, a one-way permutation-based multivariate ANOVA (PERMANOVA) was
242 performed using the function “*adonis2*” in *vegan* to further assess findings. To validate our
243 interpretation on the PERMANOVA, we performed a PERMDIST test *betadisp* (a multivariate
244 equivalent to levene's test for homogeneity of variance), test results are available in Supplementary 2.

245 **Results**

246 *i) Effect of depth on fish density m^{-2} and species richness m^{-2} at Yongala*

247 We found no significant ($p > 0.05$) differences between depths for either fish density m^{-2} or species
248 richness m^{-2} (Table 1; Supplementary 2). Mean fish density m^{-2} for antipatharians in the shallower
249 depth range (14–20 m) was 2.23 ± 0.93 , and 0.65 ± 0.32 for scleractinians. In the deeper depth range
250 (22–27 m), antipatharians hosted an average fish density m^{-2} of 1.99 ± 0.80 and scleractinians hosted
251 fish densities of 0.58 ± 0.24 . Similar to fish density, species richness did not strongly vary across depth
252 bands. Mean species richness m^{-2} for antipatharians in the shallower depth range (14–20 m) was 0.089
253 ± 0.024 , and 0.052 ± 0.014 for scleractinians. In the deeper depth range (22–27 m), antipatharians hosted
254 an average fish density m^{-2} of 0.054 ± 0.014 and scleractinians hosted fish densities of 0.124 ± 0.032 .
255 Despite having slightly higher richness, scleractinians at the deeper Yongala reef band (22–27 m) were
256 not significantly different upon examining pairwise differences in a post-hoc analysis ($p > 0.05$; see
257 Supplementary 2). Since depth range had no overall effect at *Yongala*, we pooled the fish density m^{-2}
258 and species richness m^{-2} across the two depth bands for subsequent analyses, examining corals in terms
259 of site (*Yongala* vs. *Orpheus*) and coral order (*Antipatharia* vs. *Scleractinia*).

260 *ii) Effect of coral taxon and site on fish density m^{-2} and species richness m^{-2}*

261 Fish density m^{-2} varied significantly between the coral taxa (Table 1 – Model 3), with Antipatharians
262 hosting on average 3.55 times the fish density per area as Scleractinians (Supplementary 2). Coral taxon
263 did not have a significant effect on fish richness m^{-2} (Table 1 – Model 4; Supplementary 2). For
264 antipatharians, fish density m^{-2} decreased from 2.13 ± 0.56 fish m^{-2} at *Yongala*, to 0.35 ± 0.15 fish m^{-2}
265 at *Orpheus*. Likewise, for scleractinians fish density decreased from 0.63 ± 0.192 m^{-2} at *Yongala*, to
266 0.09 ± 0.04 m^{-2} at *Orpheus* (Figure 2; Supplementary 2). At *Yongala*, fish richness m^{-2} for antipatharians
267 was 0.071 ± 0.013 m^{-2} , and 0.090 ± 0.016 m^{-2} for scleractinians. At *Orpheus*, richness m^{-2} was $0.026 \pm$
268 0.008 m^{-2} antipatharians, and 0.016 ± 0.005 m^{-2} for scleractinians. Site had a significant effect on fish
269 density and richness per m^2 (Table 1 – Models 3 & 4, respectively), with *Yongala* having 6.44 and 3.93
270 times the fish density and species richness m^{-2} as *Orpheus*, respectively (Supplementary 2).

271 *iii) Fish community structure associated with antipatharians*

272 A total of 28 fish species were recorded in close association with antipatharians (20 different species at
273 *Yongala* and 13 at *Orpheus*), from 11 families (Figure 1a; data in Appendix 1), and 23 functional entities
274 (Table 1 in Supplementary 3). The most common and abundant species were *Neopomacentrus azysron*,
275 *Rhabdamia gracilis*, *Chromis nitida*, *N. bankieri*, *Verulux cypselurus*, *Ostorhinchus cladophilos*,
276 *Cheilodipterus quinquelineatus*, and *N. cyanomos*. The most frequently observed fish behaviours were
277 HovH and HovA, with around 89% and 10% of individuals displaying these behaviours, respectively
278 (Figure 1c). Marginal fish density at *Yongala* was 2.12 ± 0.5 (mean fish density $m^{-2} \pm SE$), and $0.35 \pm$
279 0.5 m^{-2} at *Orpheus* (Figure 2). The fish communities associated with scleractinians in comparison to the
280 antipatharians show the following patterns: they shared >40% of the species recorded, the most common
281 fish behaviour was HoA (~81% of fish), in contrast to HoH, which was dominant for antipatharians but

282 represented only 18% of the fish associated with scleractinians (Figure 1c). At functional level, we
283 found that there is an overlap of approximately 37% of functional entities (Table 1 in Supplementary
284 3).

285 *iv) Functional diversity indices of coral-associated fish assemblages*

286 The overall total number of fish functional entities associated with the corals was 30, of which 12 were
287 exclusively associated with antipatharians and seven uniquely associated with scleractinians (Figure
288 4b; Supplementary 2). Marginal mean estimate of functional entities for antipatharians was 3.9 ± 0.3 ,
289 and 2.5 ± 0.3 for scleractinians, and it was significantly different between coral taxa ($p < 0.05$; Table 1;
290 Supplementary 2). Marginal estimated mean of functional redundancy for antipatharians was $1.16 \pm$
291 0.04 , and 1.11 ± 0.04 for scleractinians, and it was not significantly different between coral taxa (p
292 > 0.05 ; Table 1; Supplementary 2). Functional richness (FRic) was quantified as the volume of the
293 convex hull occupied by species in trait space, representing the total functional space filled by a
294 community. When plotted, we found that the convex hulls generated by antipatharians and
295 scleractinians respectively were similar in size and shape and largely overlapped to generate the global
296 functional trait space. (Figure 5c).

297 *v) Overall significance of the different variables driving the fish community structure*

298 The dbRDA analysis (PERMANOVA; pseudo- $F = 1.86$, 999 permutations, p (perm) = 0.001) showed
299 that coral taxon was the only variable that had a significant influence ($p < 0.05$) on the fish community
300 (Supplementary 2). For visualisation, the different variables added in the model (coral order, shelter
301 volume, and sites) were plotted as vectors according to the magnitude and direction of the relationship
302 and overlaid on the fish community observed. Significant differences in the fish community between
303 the coral taxa was found and confirmed by the one-way permutation test (PERMANOVA; pseudo- $F =$
304 3.13 , 999 permutations, p (perm) = 0.001). The PERMDISP test confirmed equal dispersion within the
305 two coral taxa ($F = 0.97$, $p = 0.302$).

306

307 **Discussion**

308 The ongoing decline of reefs globally has prompted greater interest in the functional roles of different
309 reef-associated taxa for preserving functional coral reef ecosystems (Bellwood et al., 2004a; Darling et
310 al., 2012; McLean et al., 2021). Nonetheless, studies of key ecological functions in corals (e.g. reef
311 accretion and habitat provision) have focused mostly on scleractinians. While reef accretion is mainly
312 attributable to scleractinians and calcifying algae, other benthic taxa provide important habitat
313 complexity that supports coral reef biodiversity and ecosystem functioning. Our study represents one
314 of the first to examine the role of antipatharians in supporting fish communities on shallow tropical reef
315 ecosystems. We provide empirical evidence for the contribution of antipatharians for habitat provision

316 in shallow tropical reefs. We found that antipatharians support a diverse range of fish species that utilise
317 them for a range of different purposes. They support unique fish species and functional entities, and
318 unique types of associations with fishes compared to scleractinians, but there are also overlaps between
319 the coral taxa. At the functional level, fish assemblages display considerable similarity between
320 antipatharians and scleractinians, suggesting that antipatharians could provide some redundancy of fish
321 functional roles if scleractinians decline. Consequently, our findings underscore that, owing to their
322 unique fish assemblages and the overlap with those linked to scleractinians, antipatharians serve as
323 essential components of shallow tropical marine ecosystems.

324

325 *Fish communities associated with antipatharians*

326 Of the 28 different species and 23 functional entities associated with antipatharians, suggesting a high
327 diversity of functional roles with the fish community supported by antipatharians. The most common
328 and abundant species were primarily using the colonies as shelter (HovH, HovA behaviours; Figure
329 1c). This is not surprising considering the ample shelter capacity that branching corals provide for small-
330 bodied fish and/or juvenile fish (Beukers et al., 1997; Kerry & Bellwood, 2015b), which is the most
331 common morphology of antipatharians on both surveyed sites (Figure 3). This type of association
332 (hovering around the colonies for protection) was not restricted to small-bodied fish – at both sites, we
333 also recorded larger fish (e.g. *Lutjanus russellii*, *L. carponotatus*, *Platax pinnatus*) hovering behind
334 antipatharian colonies (Figure 3a). These larger fish may be using the antipatharian colonies to shelter
335 from strong currents or to ambush prey. Additionally, we documented corallivorous fish (e.g.
336 *Chaetodon rainfordi*, *Heniochus acuminatus*) feeding on antipatharian polyps (Figure 2b). We also
337 observed several *Gobiodon* species using antipatharians as habitat (Figure 3c; see also Allen et al.,
338 2004), but we were not able to quantify the abundance of these cryptic fish using MaxN through video
339 analysis. Nonetheless, further studies of cryptic reef fishes and their symbiosis with antipatharians
340 deserves attention due to their important role in coral reef energy transfer (Brandl et al., 2019).

341 The use of antipatharians as nocturnal shelter by predator fishes has been reported from
342 mesophotic reefs in Hawaii (Boland & Parrish, 2005). In this current study conducted during daylight
343 hours, we documented predator fishes (e.g. *Plectropomus leopardus*, *Cephalopholis boenak*) laying
344 static next to or under antipatharian colonies (Stat_in behaviour), which were potentially sheltering or
345 waiting to ambush smaller fish. While not recorded during our stationary videos, we also observed other
346 species of conservation interest and commercial value, such as the marble-grouper (*Epinephelus*
347 *fuscoguttatus*), laying static among antipatharian colonies (Figure 3d). In addition, cleaner wrasse
348 (*Labroides dimidiatus*) – which uses the antipatharians as habitat and refuge – attracted larger fish (e.g.
349 *Diagramma pictum labiosum*, *Platax teira*) which hovered next to the coral colonies to get cleaned
350 (Figure 3e).

351 Predator-prey interactions are considered one of the most fundamental ecological dynamics on
352 coral reefs (Hixon & Beets, 1993). In the present study, we observed several large predator fish (e.g.
353 *Carangoides fulvoguttaus*, *Plectropomus maculatus*, *Lutjanus monostigma*, etc.) passing by; and
354 despite these fishes being likely foraging, we did not observe any actual predation events on the
355 stationary cameras due to their limited field of view. Nonetheless, schools of the greater amberjack
356 (*Seriola dumerili*), the bluefin tuna (*Tunnus thynnus*), and the yellowmouth barracuda (*Sphyræna*
357 *viridensis*) have been documented searching for fish prey among antipatharian forests on TMEs in the
358 Mediterranean (Chimienti et al., 2020). Consequently, antipatharians are important for a range of fish
359 species, providing both protection for prey and foraging opportunities for predators.

360

361 *Influence of the different variables on the fish community observed*

362 *Site.* We found site to have a significant effect on the fish density m^{-2} and richness m^{-2} (Table 1), which
363 was not unexpected considering that fish communities on shipwrecks are known to differ to those found
364 on natural reefs (Nieves-Ortiz et al., 2021; Sánchez-Caballero et al., 2021). Nonetheless, differences in
365 fish richness and density are also evident from studies comparing both similar and distinct coral reef
366 morphologies, and are often driven by site-level factors (Galbraith et al., 2021; Gilby et al., 2016).
367 Despite fish density m^{-2} and richness m^{-2} being higher at *Yongala* (Figure 1 & 2), our dbRDA analysis
368 of fish community composition – testing the influence of sites, coral taxa, and shelter volume – revealed
369 no significant effect of sites on community structure (Figure 4). This suggests that site-specific factors
370 did not significantly shape overall fish community composition.

371 At *Yongala*, depth did not have a significant effect on the fish community, and most of the fish
372 recorded occur across the wreck depth gradient (14-29 m depth). The one exception was *Ostorhinchus*
373 *cladophilos*, which is not typically found above 20 m depth (Froese & Pauly 2023). Sixty-two per cent
374 of the fish recorded in this study inhabit mesophotic ecosystems, slightly lower than a study from
375 Hawaii which reported that 95% of the fish documented in association with antipatharians on
376 mesophotic reefs are also occur on shallow reefs (Boland & Parrish, 2005). Nonetheless, in the eastern
377 Atlantic – where antipatharian forest are found at mesophotic depths – the most common functional
378 entities and species shifted between shallow and mesophotic reefs, even when 90% of the fish functional
379 entities were shared between shallow and mesophotic reefs (Bosch et al., 2023). Thus, antipatharians
380 might promote specialisation of reef fishes along the reef depth gradient (Bosch et al., 2023), which is
381 yet to be investigated on tropical reefs.

382

383 *Area and Shelter volume.* In addition to colony area (m^2), we used shelter volume (dm^3) to
384 quantify one of the most important ecological functions of corals – shelter provision (Urbina-Barreto et
385 al., 2021, 2022). Importantly, shelter volume encompasses both the area of the coral and its morphology,

386 both of which influence specific ecological functions (Kerry & Bellwood, 2015a; Lingo & Szedlmayer,
387 2006). In light of the lack of proxies to estimate shelter volume specifically for antipatharians, we use
388 the ones developed for scleractinians (Urbina-Barreto et al., 2021) considering that ‘branching’
389 morphology of scleractinians and antipatharians varies in terms of branch thickness and arrangement,
390 and antipatharians tend to exhibit more intricate and complex structures. As such, applying the same
391 morphological proxies to both coral groups is a conservative approach. Indeed, our results suggest that
392 these proxies adequately capture shelter volume in antipatharians. Notably, our models results were
393 similar regardless of whether area or shelter volume was used. Nonetheless, dedicated proxies for
394 antipatharians would be preferable for future studies.

395 Structural complexity is a key predictor of both fish abundance and species richness on coral
396 reefs (Darling et al., 2017; Graham & Nash, 2013; Urbina-Barreto et al., 2022). However, in our study
397 shelter volume influenced fish richness, but not abundance. Despite greater shelter volume comprising
398 more habitat, niche space within a colony is more homogenous than at the colony perimeter (Boström-
399 Einarsson et al., 2014; Holbrook & Schmitt, 2002; Robertson, 1996). Therefore, our results could be
400 related to large colonies - with homogenous internal shelter volume - regulating fish abundance through
401 competitive interactions. Additionally, nuanced relationships between patch habitat area and edge
402 interactions with surrounding habitats are often associated to species richness, but not to abundance
403 (Fonseca, 2008; Hattori & Shibuno, 2015). For instance, fish species richness may be enhanced around
404 the colony perimeter where the habitat is more complex and where opportunities for interactions with
405 surrounding habitat are optimised (Hattori & Shibuno, 2015).

406

407 *Coral taxon.* The density m^{-2} , and number of functional entities of fish communities varied significantly
408 between antipatharians and scleractinians, nonetheless, there was not a significant difference on species
409 richness m^{-2} and both coral taxa supported seven functional groups (Figure 1c & 5). There is high
410 overlap of fish species associated with both coral taxa, and despite some species found in unique
411 association with either scleractinians or antipatharians (Figure 5a), none of these species are considered
412 as either antipatharian or scleractinian specialist (Froese & Pauly, 2015). One potential explanation for
413 the difference in fish richness among coral taxa is the type of association with the corals. For instance,
414 while the most abundant families (Apogonidae, Pomacentridae, Labridae) were shared between both
415 coral taxa, Lutjanidae was 80% more abundant for antipatharians (Figure 2a). Within the family
416 Lutjanidae, *L. russellii* and *L. carponotatus* were only recorded in association with antipatharians, and
417 both fish species appeared to be using the colonies as shelter from currents (HovH behaviour; Figure
418 3a). Similar specific interactions have been observed for scleractinians; for example, some fish use
419 tabular *Acropora* colonies to protect themselves from solar irradiance (Kerry & Bellwood, 2015b).

420 Therefore, species-specific associations may contribute to dissimilarities in the fish species associating
421 with antipatharians and scleractinians.

422 At the functional level, fish communities associated with antipatharians exhibited significantly
423 higher functional diversity, with a greater number of unique functional entities compared to those
424 associated with scleractinians (Figure 5b), suggesting that antipatharians can support fish faunas with
425 more varied ecological roles. In contrast, analysis of functional redundancy shows that antipatharian-
426 associated communities have slightly higher redundancy than the ones in association with
427 scleractinians, meaning more species share similar ecological roles, although the difference was not
428 significant. The functional space analysis illustrates these dynamics (Figure 5c), with antipatharian and
429 scleractinian communities showing large overlap in their functional spaces, indicating shared ecological
430 roles, while also revealing distinct differences that underscore the broader functional range supported
431 by each coral taxon. The hulls in the plot further emphasize this overlap and variation, highlighting the
432 unique contributions of each coral taxa to the global functional space (Figure 5c). While higher
433 functional diversity implies more unique roles and higher redundancy suggests more overlapping roles,
434 these findings are not necessarily contradictory but reflect complex ecological dynamics. Antipatharians
435 may provide diverse niches that foster unique functional roles while also supporting multiple species
436 within those roles, potentially enhancing resilience against species loss. Scleractinian-associated
437 communities, with fewer unique functional entities and slightly lower redundancy, may be more
438 vulnerable to losing critical ecological functions. These findings, in the context of functional diversity
439 and vulnerability in tropical reef fish faunas (Mouillot et al., 2014), highlight the critical influence of
440 coral type on both the variety and stability of ecological roles in reef ecosystems.

441 Variation in fish communities may be in part attributable to intrinsic differences in
442 morphological complexity between the two coral taxa. The differences in complexity and branching
443 arrangement between antipatharians and scleractinians are shown in Figure 3 (f,g,h). Antipatharians do
444 not grow as massive or encrusting colonies, and all growth forms extend vertically off the substrate,
445 thereby increasing the exposed area available for habitat. Moreover, the canopy-like effect created by
446 most antipatharian growth forms can enhance fine-scale hydrodynamic conditions (e.g. upwelling) that
447 promote the retention of plankton and juvenile fish, which benefits planktivorous, invertivorous and
448 piscivorous fish species (Guizien & Ghisalberti, 2017). Additionally, habitat spaces provided by
449 densely branched colonies might also influence fish density due to the schooling behaviour of most
450 planktivore fishes, and the refuge availability and survivorship for juvenile and small-bodied fish. While
451 shelter volume provides a quantitative measure of the space available for shelter, it is based on colony
452 area or diameter (Urbina-Barreto et al., 2021); therefore, it does not capture the elevation from the
453 substrate (colony height). This could explain why neither shelter volume nor area had a significant
454 effect on fish abundance.

455 Numerous studies have identified colony height as a more influential factor driving fish
456 assemblages than surface area or coral shape (Fisher, 2023; Harborne et al., 2012). Therefore, future
457 studies should quantify both shelter volume and colony height of corals when examining their
458 correlation with fish assemblages. Additionally, the development of proxies specifically for
459 antipatharians could enable finer-scale morphological differences to be captured. This information will
460 enable trait-based approaches to understanding coral reef function to be extended to a wider range of
461 benthos, rather than just scleractinians – an important approach considering scleractinians are not
462 necessarily the dominant habitat-forming benthos in many shallow tropical ecosystems.

463

464 *Implications for conservation*

465 The importance of trait-based approaches to support and guide local and regional conservation
466 strategies in light of the current coral reefs crisis is now well recognised (Bellwood et al., 2004b; Hughes
467 et al., 2017a; McLean et al., 2021). However, most studies utilising trait-based approaches in coral reef
468 ecology and the influence of benthic communities on fish assemblages focus on scleractinians (Darling
469 et al., 2017; Fisher, 2023; Harborne et al., 2012). Our study highlights that other coral taxa can
470 significantly influence reef fish communities, playing an important role in providing three-dimensional
471 habitat complexity on shallow tropical reefs. Other habitat-forming benthic groups have been
472 previously considered (e.g. octocorals and sponges; González-Murcia et al., 2023; Moynihan et al.,
473 2022); however, antipatharians are commonly neglected from coral reef monitoring programs and
474 studies. A greater effort to quantify the abundance and ecological roles of the different benthic groups
475 would lead to a more holistic understanding of how the different benthic taxa interact to support coral
476 reef biodiversity.

477 While antipatharians are not abundant in the shallowest depths (<10 m), they are common in
478 most other reef depths, in both shallow and mesophotic reefs (Molodtsova et al., 2023; Wagner et al.,
479 2012). Importantly, antipatharians are less susceptible to the phenomenon known as bleaching (Gress
480 et al., 2021), and other climate related stressors (Godefroid et al., 2023) than scleractinians. Given the
481 impact of bleaching events on scleractinians (Hughes et al., 2017b, 2018), the importance of other coral
482 taxa in supporting and maintaining reef ecological functions requires a greater understanding to account
483 for in conservation strategies.

484 Human activities such as fisheries have led to some antipatharian species being listed as “near
485 threatened” by the International Union for Conservation of Nature (IUCN) Red List of the
486 Mediterranean (Bo et al., 2008, 2017). Nonetheless, the status of antipatharian species outside the
487 Mediterranean remains unknown despite evidence of declines on some tropical reefs (Boland & Parrish,
488 2005; Gress & Kaimuddin, 2021; Grigg, 2004). Considering the relevance of antipatharians in

489 supporting reef biodiversity, we argue that a greater effort should be afforded to understanding the role
490 of antipatharians and their status worldwide.

491

492

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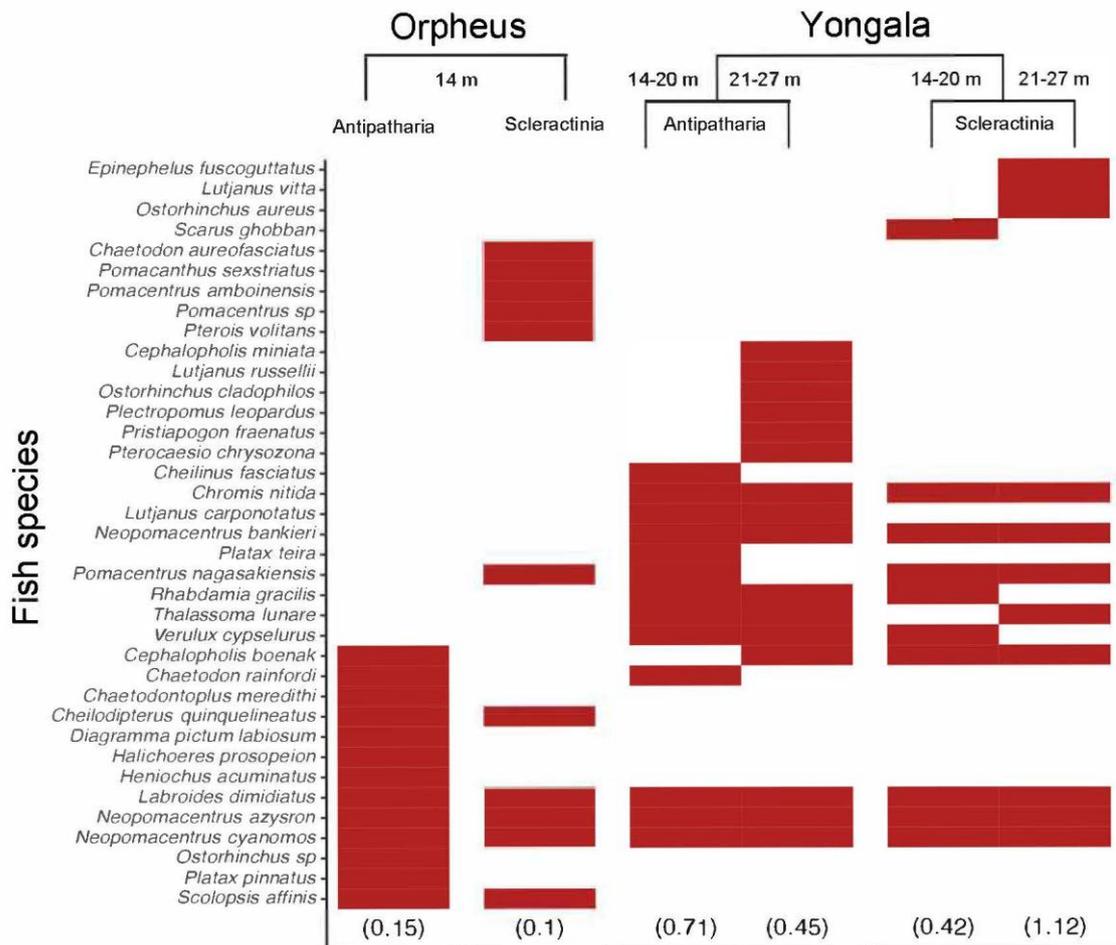
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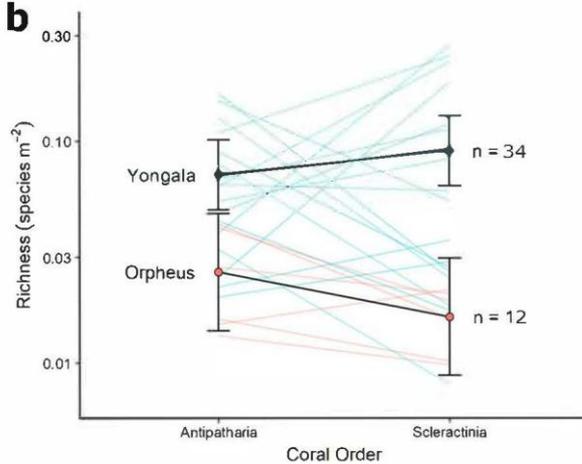
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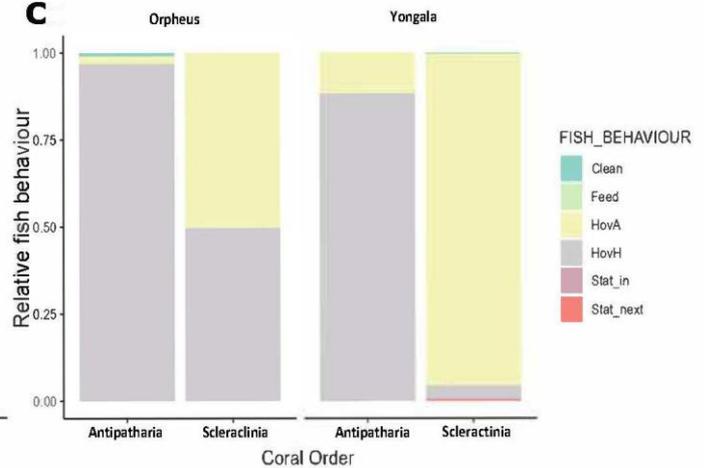
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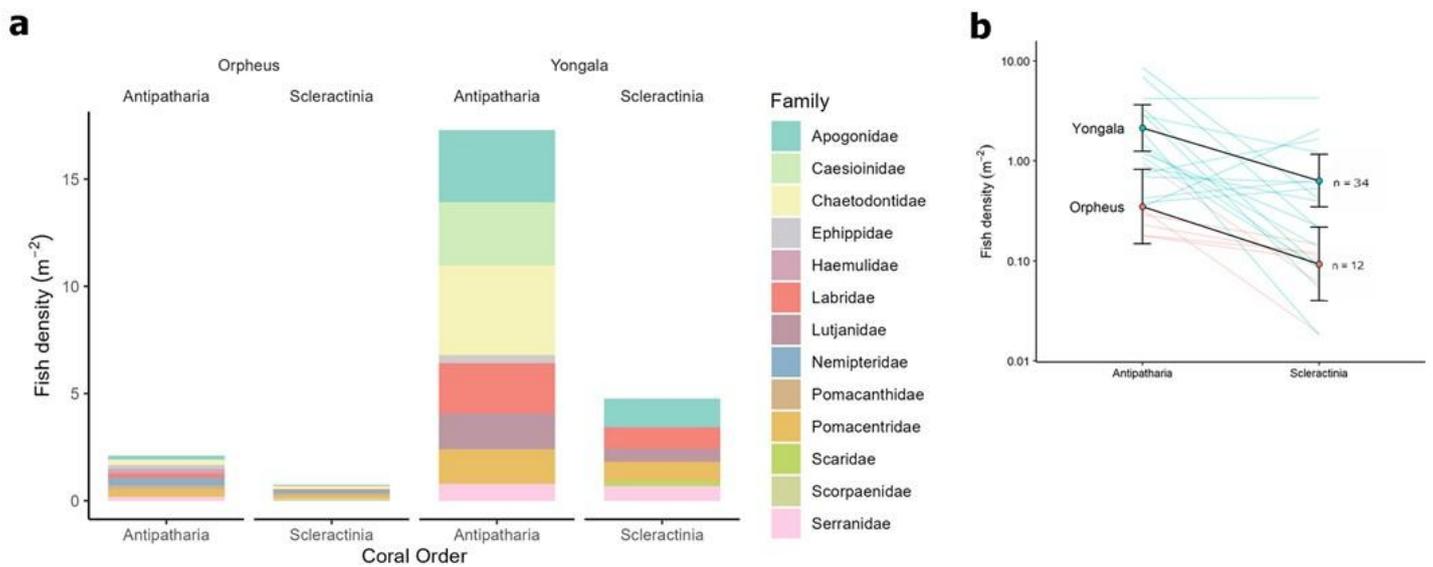


771 **Figure 1** | Fish species richness associating with antipatharians and scleractinians at *Yongala* and
 772 *Orpheus*: **a**) Fish species present (red bars) or absent (white bars) for each coral taxa, study site and

773 depth; numbers in parenthesis are the total relative fish species richness m^{-2} . **b)** Marginal effects plot of
 774 mean fish richness m^{-2} predicted for each coral taxa (Model 4 in Table 1). Coloured lines connect colony
 775 pairs (one antipatharian and one scleractinian) for each of the two sites: 34 colonies at *Yongala* ($n = 17$
 776 colony pairs), and 12 colonies at Orpheus ($n = 6$ colony pairs) surveyed. **c)** Relative fish behaviour
 777 showing the proportional contribution of each type of behaviour ().

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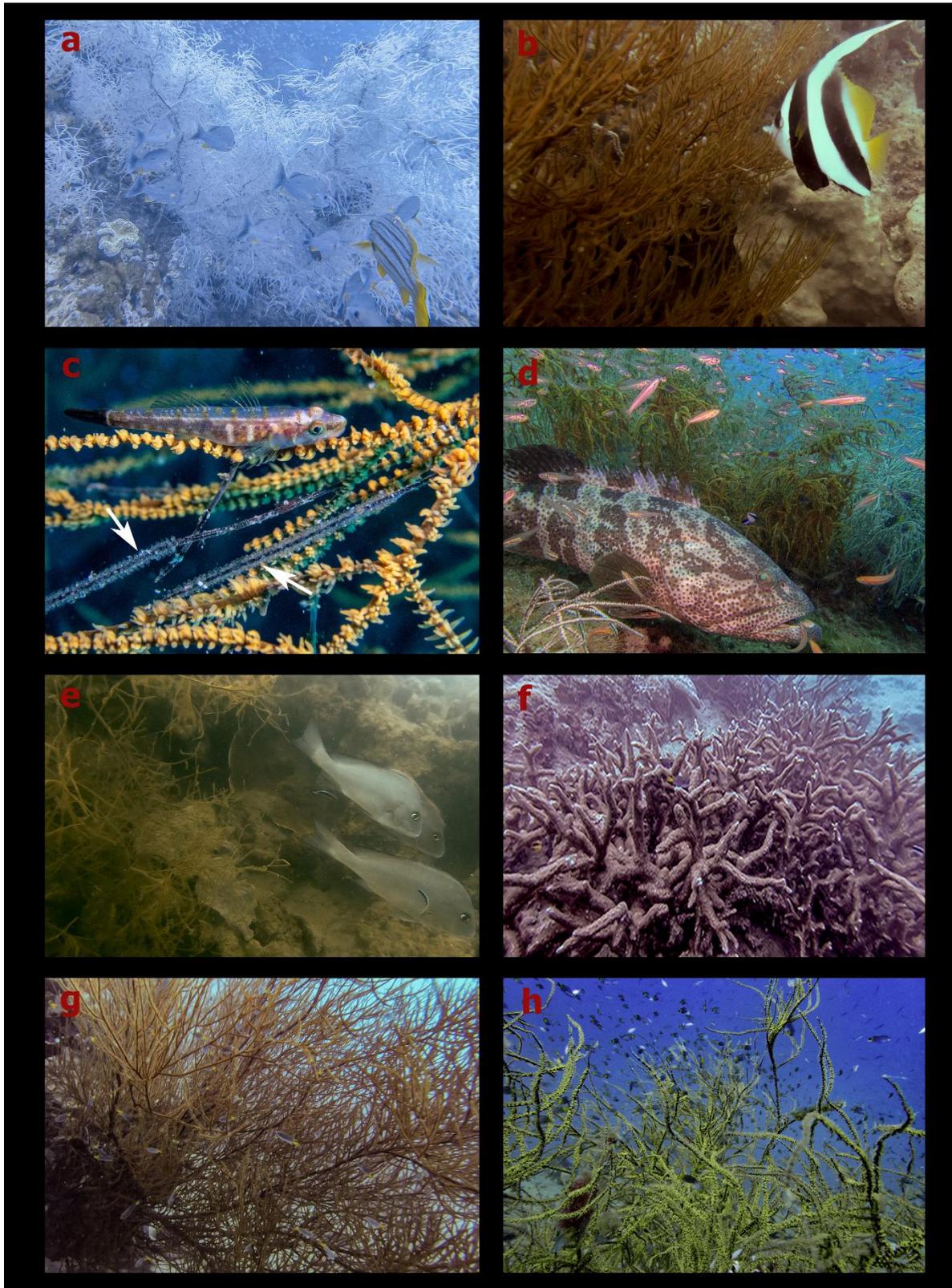
781 **Figure 2** | Fish density associating with antipatharians and scleractinians at *Yongala* and Orpheus: **a)**
 782 Fish density m^{-2} showing the contribution of each family proportional to the average density of fish
 783 within the family. **b)** Marginal effects plot of mean fish density m^{-2} predicted for each coral taxa (Model
 784 3 in Table 1). Coloured lines connect colony pairs (one antipatharian and one scleractinian) for each of
 785 the two sites: 34 colonies at *Yongala* ($n = 17$ colony pairs), and 12 colonies at Orpheus ($n = 6$ colony
 786 pairs) surveyed.

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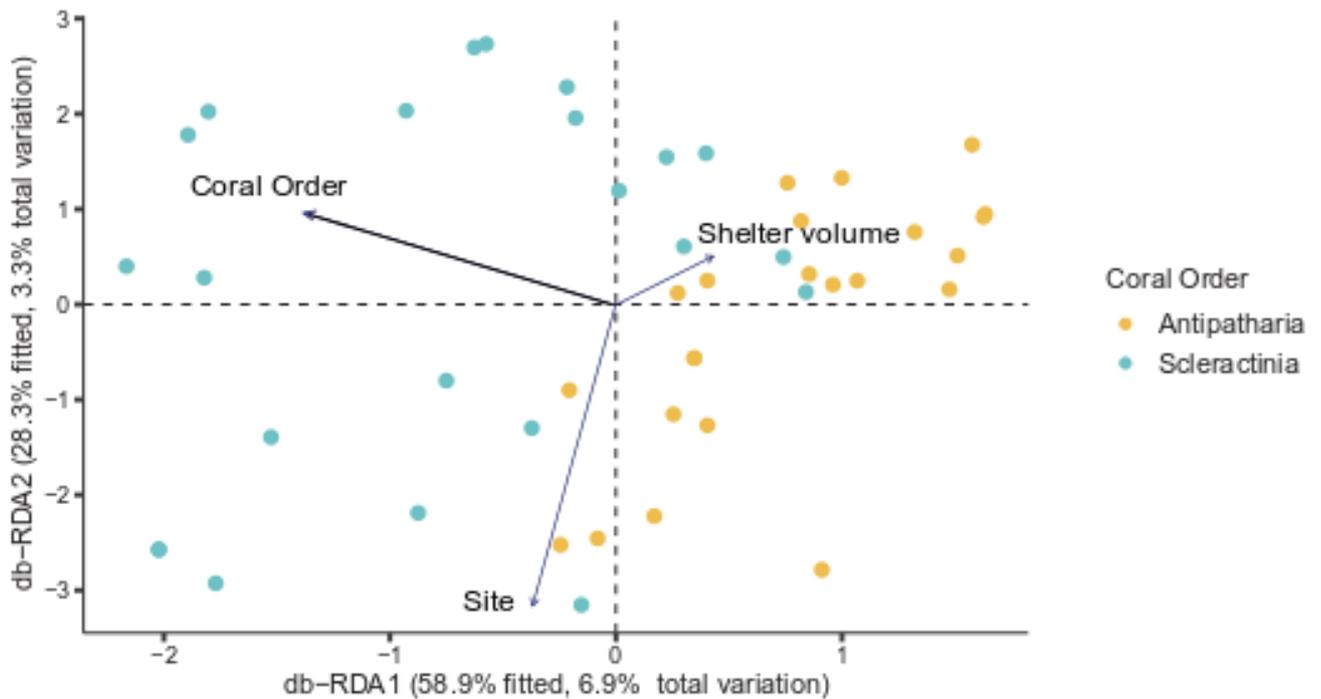


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792 **Figure 3** | Examples of interactions between fish and antipatharians and scleractinians documented in
 793 this study: **a)** *Lutjanus russellii* and *L. carponotatus* behind a white antipatharian colony sheltering from
 794 the current at *Yongala*. **b)** *Heniochus acuminatus* feeding on the polyps of an antipatharian colony at
 795 Orpheus. **c)** *Bryaninops tigris* residing on an antipatharian colony at *Yongala*; white arrows show its

796 eggs deposited on the colony branches. **d)** *Epinephelus fuscoguttatus* laying among antipatharian
 797 colonies at *Yongala*. **e)** *Diagramma pictum* being cleaned by *Labroides dimidiatus* while hovering next
 798 to an antipatharian colony at Orpheus. **f)** A range of fish species hiding among a branching scleractinian
 799 colony at Orpheus. **g)** A range of fish species hiding among a branching antipatharian colony at
 800 Orpheus. **h)** A range of fish species sheltering among a branching antipatharian at *Yongala*. (Photos:
 801 Erika Gress).

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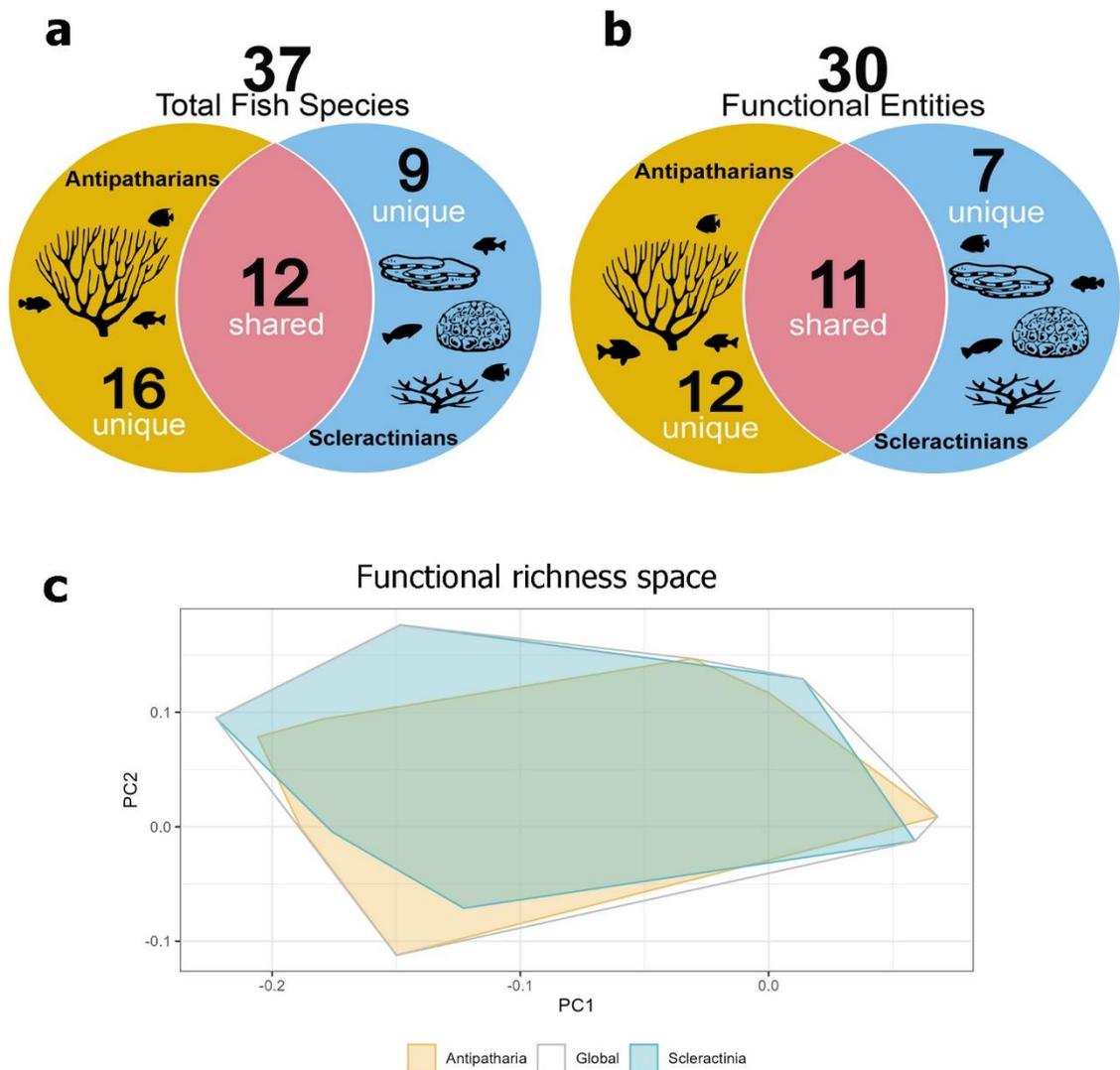


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804 **Figure 4** | Distance-based redundancy analysis (dbRDA) of fish communities associated to
 805 antipatharian (yellow dots) and scleractinian (blue dots) colonies. Vectors (arrows) represent the
 806 different variables tested on their significance as drivers of the fish community. The length and direction
 807 of the arrow represents the magnitude and direction of the relationship. Coral taxon (thicker arrow) was
 808 identified as the only significant variable ($p < 0.05$) influencing the fish community (Supplementary 2).

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812 **Figure 5** | Fish communities associated with antipatharians and scleractinians. **a)** Unique and shared
 813 fish species associated to each coral taxon. **b)** Unique and shared functional entities associated to each
 814 coral taxon. **c)** Coral-associated fish functional richness space illustrating the distribution of functional
 815 richness for antipatharians (yellow hull), scleractinians (blue hull), and the global (i.e., overall, white
 816 hull).

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821 **Table 1** | Summary table of the generalised linear mixed-effect models (GLMMs), generalised linear
 822 models (GLMs), generalised least square model (GLS), and Generalized Additive Models for
 823 Location, Scale, and Shape (GAMLSS) model. Model numbers (first column) correspond to the name
 824 given on Supplementary 2, which contains full model summaries, estimated marginal means and
 825 contrast analysis results. Significant factors in each model are in bold.

Model	Question	Formula [model type]	Parameters	Test statistic	p-value
Model 1	Depth on fish density m^{-2} , for all colony pairs at <i>Yongala</i> site	Density $m^{-2} \sim$ Depth range \times Coral taxon + (1 Pair) [GLMM]	Intercept (14–20m, Antipatharia): Depth range effect (22–27m): Coral taxon effect (Scleractinia): Depth \times coral taxon interaction:	1.91 -0.19 -2.15 0.00	0.057 0.85 0.032 0.99
Model 2	Depth on fish richness m^{-2} , for all colony pairs at <i>Yongala</i> site (two depth bands)	Richness $m^{-2} \sim$ Depth range \times Coral taxon + (1 Pair) [GLMM]	Intercept (14–20m, Antipatharia): Depth range effect (22–27m): Coral taxon effect (Scleractinia): Depth \times coral taxon interaction: (note: all confidence intervals overlap and no significant pairwise differences found during post-hoc analysis)	-8.99 -1.34 -1.39 2.59	< 0.001 0.18 0.17 0.01
Model 2a	Depth on fish richness m^{-2} , for antipatharians at <i>Yongala</i>	Richness $m^{-2} \sim$ Depth range [GLM]	Intercept (14–20m): Depth range effect (22–27m):	-11.6 -1.73	< 0.001 0.083
Model 2b	Depth on fish richness m^{-2} , for scleractinians at <i>Yongala</i>	Richness $m^{-2} \sim$ Depth range [GLS]	Intercept (14–20m): Depth range effect (22–27m):	-13.5 0.90	< 0.001 0.38

Model 3	Coral taxon and site on fish density m ⁻²	Density m ⁻² ~ Coral taxon * Site + (1 Pair) [GLMM]	Intercept (Antipatharia, Orpheus): Coral taxon effect (Scleractinia): Site effect (Yongala): Coral taxon × Site interaction:	-2.48 -2.33 3.66 0.16	0.013 0.020 < 0.001 0.87
Model 4	Coral taxon and site on fish richness m ⁻²	Richness m ⁻² ~ Coral taxon * Site + (1 Pair) [GLMM]	Intercept (Antipatharia, Orpheus): Coral taxa effect (Scleractinia): Site effect (Yongala): Coral taxon × site interaction:	-12.1 -1.10 2.87 1.44	< 0.001 0.27 0.0041 0.15
Model 4	Coral taxon on mean fish entities	Functional entities ~ Coral taxon + (1 SITE) [GLMM]	Intercept (Antipatharia): Coral taxa effect (Scleractinia)	4.30 2.87	< 0.001 0.0064
Model 6	Coral taxon on mean fish functional redundancy	Functional redundancy ~ Coral taxon [GAMLSS]	Intercept (Antipatharia): Coral taxa effect (Scleractinia)	4.30 -1.10	< 0.001 0.276

826 *Density* m⁻² refers to the standardised fish abundance per m²; and *Richness* m⁻² refers to the number of fish species
827 observed per m².

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836 Galbraith; Formal analysis: Erika Gress, Gemma Galbraith, Kevin Bairos-Novak; Writing - original
837 draft: Erika Gress; Writing – review and editing: Erika Gress, Gemma Galbraith, Kevin Bairos-Novak,
838 Tom Bridge; Funding acquisition: Erika Gress, Tom Bridge.

839 **Data availability:** Appendix 1 contains the datasets generated during the current study are available
840 via the following link: <https://figshare.com/s/f0f6d8fc855a81628b66>. Appendix 2 contains the dataset
841 generated for fish functional analyses available via the following link:
842 <https://figshare.com/s/da1b6baa16db0c0c47f8>. Both currently private-for-peer review.