

1 INTEGRATING ECOLOGICAL FEEDBACKS ACROSS SCALES

2 AND LEVELS OF ORGANIZATION

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ABSTRACT

25

26 In ecosystems, species interact in various ways with other species, and with their local
27 environment. In addition, ecosystems are coupled in space by diverse types of flows.
28 From these links connecting different ecological entities can emerge circular pathways
29 of indirect effects: feedback loops. This contributes to creating a nested set of ecological
30 feedbacks operating at different organizational levels as well as spatial and temporal
31 scales in ecological systems: organisms modifying and being affected by their local
32 abiotic environment, demographic and behavioral feedbacks within populations and
33 communities, and spatial feedbacks occurring at the landscape scale. Here, we review
34 how ecological feedbacks vary in space and time, and discuss the emergent properties
35 they generate such as species coexistence or the spatial heterogeneity and stability of
36 ecological systems. With the aim of identifying similarities across scales, we identify
37 the abiotic and biotic modulators that can change the sign and strength of feedback
38 loops and show that these feedbacks can interact in space or time. Our review shows
39 that despite acting at different scales and emerging from different processes, feedbacks
40 generate similar macroscopic properties of ecological systems across levels of organization.
41 Ultimately, our contribution emphasizes the need to integrate such feedbacks to improve
42 our understanding of their joint effects on the dynamics, patterns, and stability of ecological
43 systems.

44 **Keywords:** Feedback loops, temporal and spatial scales, level of organization, emergent
45 properties, self-organization, ecosystem patterns, species coexistence, stability, ecosystem

46 functioning, niche construction

INTRODUCTION

47

48 The discovery of art caves depicting species interactions (*e.g.*, Viereck and Rudner, 1957), or
49 more recently, narratives and paintings about self-organized patterns in ecosystems (Walsh
50 et al., 2023) illustrate that human curiosity toward patterns and interdependencies in nature
51 has been a long-standing interest. In the last century, the development of cybernetics (*i.e.*,
52 the study of regulation and dynamics of interconnected systems; Wiener, 1948; Maruyama,
53 1963) has led feedbacks to become increasingly studied. In a set of connected entities, a
54 feedback loop corresponds to a circular path of dependencies. This means that a focal
55 entity modulates its own dynamics by changing other entities composing the system,
56 which, in turn, affect its variations. Examples involve social norms that regulate people's
57 behaviors in complex societies (Fehr and Fischbacher, 2004), or nitrogen-fixing plants
58 that increase the local availability of nitrogen on which they grow (Menge and Hedin,
59 2009). The feedback concept now infuses all complex systems including climatology (Rial
60 et al., 2004; Kemp et al., 2022), history (Downey et al., 2016; Centeno et al., 2023), political
61 and economic sciences (Casillas and Kammen, 2010; Leonard et al., 2021), and ecology
62 (DeAngelis et al., 1986).

63 In ecology, the adoption of a system's approach to quantifying indirect effects began with
64 the early development of ecosystem ecology (Margalef, 1963; Hannon, 1973; Patten and
65 Odum, 1981), which was built using a complex system approach, with compartments
66 (biotic and abiotic) mapped into a network of flows of resources and energy (Lindeman,
67 1942; Odum, 1968; O'Neill et al., 1987). Since then, feedbacks have been used across all

68 levels of organization from individuals interacting with their environment (Cuddington
69 et al., 2009), communities (Zelnik et al., 2022), ecosystems (Veldhuis et al., 2018), including
70 at larger spatial scale (Phillips, 2016) to explain how emergent macroscopic properties
71 of complex adaptive systems could derive from the interdependency of smaller entities
72 (Levin, 1998). Incorporating such feedbacks often leads to interdisciplinary outcomes. For
73 instance, in the context of ongoing global changes, knowledge of feedbacks is central to
74 understanding how ecosystems participate in the global carbon balance, and ultimately
75 whether ecosystems will amplify or dampen climate change in a warmer world (Field
76 et al., 2007; Heimann and Reichstein, 2008). More practically, the mechanistic understand-
77 ing of feedbacks has been crucial to design effective restoration programs for degraded
78 ecosystems such as coastal areas or shallow lakes (Suding et al., 2004; Byers et al., 2006;
79 Silliman et al., 2015).

80 Hitherto, despite the development of feedback-based studies at different levels of organiza-
81 tion in ecology, we lack an integrative understanding of their global implications. Indeed,
82 because feedbacks are observed at different scales, they may generate similar emergent
83 properties and principles across levels of organization (see Table 1 for examples). A unified
84 framework of feedbacks going from organisms to the landscape extent could help to
85 better understand the pervasiveness of feedbacks, how they contribute to the organization
86 of ecological systems by linking scales and levels of organization, and how they determine
87 the response of ecological systems to different type of disturbances. Indeed, feedbacks are
88 key mechanisms of the regulation of ecological systems and contribute to their stability
89 properties. Feedbacks are thus involved in the response of ecosystems to disturbances and
90 more broadly to human-induced global changes.

91 With this contribution, we emphasize that ecological feedback loops connecting entities
92 (from organisms to ecosystems) give rise to emergent properties across scales. For this
93 purpose, we first give a general definition that includes a set of properties characterizing
94 each ecological feedback. Then, we present how feedbacks vary across space and time,
95 and show that, while acting at different scales, they interact in both space and time. We
96 also show that these feedback loops generate emergent properties that impact the patterns
97 of coexistence, functioning, stability, and heterogeneity of ecological systems. We give
98 examples where a better knowledge of feedbacks has helped to build conservation and
99 restoration measures (Table 2). Finally, after acknowledging mechanisms that modulate
100 the strength and sign of these loops, we propose open avenues for studying ecological
101 feedbacks.

102 **PROPERTIES OF FEEDBACK LOOPS**

103 *System and ecological entities*

104 Let us consider a system composed of a set of entities interacting with each other. In ecology,
105 these entities can represent individuals, or a species interacting with its local environment
106 (abiotic or biotic), but also higher levels of organization such as complete ecosystems
107 connected by flows of resources, organisms, and information across landscapes. Each
108 entity acts on others through both direct and indirect pathways, creating a network of
109 dependencies (Hutchinson 1948). Feedbacks emerge when an entity modulates, through
110 these different paths, the direction and speed of its own dynamic. A feedback loop is

111 therefore defined relative to a response variable measured on the focal entity such as body
112 temperature, population abundance, or carbon or nutrients stocks in ecosystems.

113 *Feedbacks emerge from links between entities*

114 Feedbacks emerge when there is a circular path of directed links (Fig. 1). In the simplest
115 context, a feedback loop appears when two entities are bidirectionally connected. Here,
116 because the feedback is made up of two steps, goes to another entity and then back to the
117 focal entity, the feedback is said to be of length two. In consumer-resource systems, the
118 consumer generates negative feedback on its own demography because it is continuously
119 lowering its resource availability (Fig. 1). At higher spatial scales, bidirectional flows
120 of resources connecting forests and streams (*e.g.*, through aquatic insect emergence and
121 plant litter subsidies) generate spatial feedbacks across the landscape (Leroux and Loreau,
122 2012; Baruch et al., 2021). Yet, in more complex settings, feedbacks can involve far
123 more than two links. For instance, in intransitive competitive communities, long loops
124 can emerge because there is no best competitor (rock-paper-scissors game; see Fig. 1):
125 all species beat some competitors and are weaker than some others (Kerr et al., 2002;
126 Allesina and Levine, 2011). Together, this entails that for a set of connected entities, the
127 interdependency of entities can result not only from direct links but also from the many
128 emerging feedback loops of varying length. Importantly, the definition of the feedback
129 loop, and the quantification of its length, depends on the defined system, and in particular
130 on the nature of its entities. Demographic regulation, for instance, corresponds to a
131 negative feedback between individuals of a given population, directly (*e.g.*, interference

132 competition) or indirectly (*e.g.*, through a resource in exploitative competition). If one
133 wants to consider only the population level, or to ignore the resource dynamics, then the
134 underlying feedback could be represented in a phenomenological way (as in the logistic
135 growth) and the feedback would be of length one (of the population on itself). Hence, the
136 way the system is defined and its scale impose a “perceptual bias” (*sensu* Levin, 1992) on
137 the feedback identified and their lengths (see also O’Neill et al., 1987). Here we base our
138 definition on a mechanistic representation of feedbacks, with a minimal length of two.

139 *Strength, sign, and dynamical properties*

140 Feedbacks are characterized by their length, but also their strength and sign. On one side,
141 positive feedbacks emerge when the circular path benefits the focal species (*e.g.*, plants
142 attracts pollinators, while pollinators feed on nectar in a mutually beneficial interaction;
143 Fig. 1). Such positive loops are destabilizing because they can self-reinforce over time
144 (Maruyama, 1963; DeAngelis et al., 1986), potentially leading to a runaway growth (“orgy
145 of mutual benefaction”, May, 1981) and/or alternative stable states and tipping point dy-
146 namics in ecosystems (Scheffer and Carpenter, 2003). On the contrary, negative feedbacks,
147 for instance generated by consumptive interactions, tend to be regulation forces that limit
148 exploding dynamics (see Emergent properties from feedback loops section).

149 The quantification of the feedback strength and sign relies on a multiplicative network
150 approach derived from the input–output theory in economy (Hannon, 1973). In a network
151 of connected entities, the strength and sign of a loop is given by the product of all signed
152 weights of the links forming the loop. Then, positive feedbacks can emerge from negative

153 interactions when the loop contains an even number of negative interactions (see Fig. 1).
154 For instance, in a two-species competitive community, each species competes with the
155 other, therefore limiting the negative effect that it experiences from its competitor (positive
156 feedback on each species). Accessing the feedback sign alone provides information about
157 whether the system will be stable or not. Notably, qualitative loop analysis can be used to
158 infer the stability of a system from the sign of links between entities (May, 1973a; Levins,
159 1974). Moreover, the feedback strength depends on the cumulation of the link weights
160 along the feedback loop. The weight of a link can correspond to interaction strengths for
161 food-webs or plant-pollinator networks (*i.e.*, the effect that a species has on the growth
162 of another for instance obtained by the partial derivative of the per capita growth rate
163 with respect to the density of another species, typically summarized in the Jacobian
164 matrix; see Ramesh and Hall, 2023 or Zelnik et al., 2022 for the generic framework), or to
165 material flows between populations, communities or ecosystems in a spatial context. Many
166 methods have been proposed to quantify the weight of links (as reviewed in Wootton
167 and Emmerson, 2005). In practice, direct interactions are often difficult to measure due
168 to possible influence of indirect effects (Lawlor, 1979; Paine, 1992), timescale, functional
169 response, and context dependencies (Berlow et al., 1999; Wootton and Emmerson, 2005).
170 Yet, assuming all link weights are known, the strength of a loop corresponds to the product
171 of the weights of links involved in the loop. Interestingly, when using feedbacks within an
172 ecological network (*i.e.*, the different links of the feedback loop correspond to ecological
173 interactions), a general pattern seems to emerge. Ecological interactions are known to be
174 heterogeneous in strength, with ecological communities typically having few strong and
175 many weak interactions (McCann et al., 1998; Wootton and Emmerson, 2005). Therefore,

176 as a loop becomes longer it will likely contain weak interactions so that the product
177 of links will become weaker. This leads to a predicted pattern, where longer loops are
178 expected to be weaker, that is often observed in nature (Neutel et al., 2002, 2007; Lever
179 et al., 2023). Moreover, because each link connecting entities has a unit (*e.g.*, consumption
180 rate in food-webs in $\text{mass}\cdot\text{m}^{-2}\cdot\text{year}^{-1}$, or interaction strength in interaction networks in
181 year^{-1} ; De Ruiter et al., 1995). This implies that two feedbacks of different lengths have
182 different dimensions, making them impossible to be compared (Ulanowicz, 2004). To avoid
183 this dimensionality issue, studies have proposed to define direct interactions relatively to
184 self-links (*e.g.*, corresponding to regulation mechanisms) by dividing interaction strengths
185 by self-links, thereby making each link non-dimensional (Neutel and Thorne, 2014; Zelnik
186 et al., 2022). Since its first application on ecosystem flows (Hannon, 1973), this network
187 approach has been applied in a wide variety of contexts, including the integration and
188 interdependence of species within communities (Zelnik et al., 2022) or the evolution of
189 species in mutualistic networks (Guimarães et al., 2017). To summarize, quantifying
190 feedback strength and sign allows comparing different loops and understanding whether
191 a given loop will self-amplify or counteract the effects of a perturbation. However, this
192 is not sufficient if one wants for instance to quantify the contribution of a given loop on
193 the transient dynamics following a perturbation (*e.g.*, how much a given feedback loop
194 increases or decreases the effect of a disturbance and the recovery time). Therefore, a
195 complementary approach has been proposed, which we detail in Box 1.

Timescales are the essence

196

197 Feedbacks are characterized by the focal entity on which it acts (*e.g.*, an organism , an
198 ecosystem), as well as a timescale. Each process and entity has a characteristic timescale:
199 decomposition is faster in freshwater compared to terrestrial ecosystems (Gounand et al.,
200 2020), and a given feedback duration will be perceived differently by species that have very
201 different generation times (*e.g.*, a phytoplankton vs a tree species if considering primary
202 producers; Steele and Henderson, 1994). For a feedback loop, the timescale corresponds to
203 the sum of all durations of processes involved in the loop. It therefore always generates a
204 time lag between the cause and its effect (Higashi and Patten, 1989). In the case of nutrient
205 recycling feedback on plants for instance, the delay is due to decomposition and nutrient
206 release by decomposers (McClaugherty et al., 1984). Comparing the generation timescale
207 of the focal entity and the one of the feedback is critical to determine how the loop may
208 impact the focal entity. If the timescale of the feedback loop is high compared to the one of
209 the entity, its effect may generate slow-fast dynamics and induce cycles (Daufresne and
210 Hedin, 2005; Pastor and Durkee Walker, 2006; Barraquand et al., 2017) or long transients
211 (Hastings, 2010; Hastings et al., 2018; Miller and Allesina, 2023). When the feedback has
212 a very long timescale compared to the entity of interest, the feedback loop may impact
213 the system but via long-term legacy effects (Albertson et al., 2022). Let us illustrate this
214 point by considering phytoplankton. Phytoplankton participates in capturing carbon and
215 regulating the climate, which ultimately conditions its own environment, but the timescale
216 at which this feedback operates makes it irrelevant to investigate phytoplankton life-cycle.
217 However, in the long term, the feedback between plankton and the climate has been shown

218 to strongly shape the dynamics of planktonic assemblages (Gibbs et al., 2006; Slater et al.,
219 2022). Hence, depending on the timescale defined by the entity of interest (here cell versus
220 assemblage), the same feedback may matter or not, depending on the temporal scale one
221 looks at. On the contrary, the timescale of the feedback can be smaller than the one of
222 the entity, in which case it is important for the focal entity: in forest-savanna systems for
223 instance, the fire-mediated feedback loop, which determines the canopy cover and the
224 strategies of plants, occurs at a smaller timescale compared to the life-span of trees, and
225 determines the bistable dynamics in the system (Pausas and Bond, 2020).

226 Framing feedbacks in terms of timescales has received a lot of attention in the context of
227 climatic mitigation measures such as carbon sequestration to understand when will carbon
228 sequestered in biomass or the ocean be remineralized (*i.e.*, return to the atmosphere, Table
229 2; Boyd et al., 2019; Friggens et al., 2020). Also, this strongly matters when comparing the
230 timescales of ecological and evolutionary processes (see Govaert et al., 2019; Fronhofer
231 et al., 2023 for dedicated reviews on this topic).

232 **FEEDBACKS ACROSS LEVELS OF ORGANIZATION**

233 Organisms interact with its abiotic environment but is itself embedded in a community,
234 while populations, communities and ecosystems are part of a landscape and interconnected
235 through flows of individuals, information and resources. The feedback concept can be
236 applied at any of these levels of organization. Here, we identify three types of feedbacks
237 depending on the level they act on: feedbacks from organisms interacting with their abiotic
238 environment, demographic and behavioral feedbacks in populations and communities

239 and spatial feedbacks at landscape scale (Fig. 2, Table 1).

240 *Feedbacks between organisms and their abiotic environment*

241 From the observation that organisms interact with their abiotic environment, Tansley
242 defined the concept of "ecosystem" encapsulating both species and their environment
243 (Tansley, 1935). Later, Lewontin suggested thinking of the environment as a third helix of
244 the DNA because the environment constrains the evolution of organisms and is in turn
245 modified by organisms (Lewontin, 2000, see also Patten, 1982). Though we do not focus on
246 eco-evolutionary feedbacks here, this metaphor illustrates the importance of interactions
247 between organisms and their environment.

248 Organism-environment interactions can involve modifications of the physical habitat
249 and landforms that feed back on the focal species' growth, survival, or abundance (*i.e.*,
250 biogeomorphic feedback; Temmink et al., 2022). Examples include intertidal sea-grass
251 meadows that stabilize the sediments (soil accretion) and change the hydrodynamic
252 system via their rooting and shooting systems (Bouma et al., 2013; Maxwell et al., 2017),
253 or coastal plants that capture sediments transported by wind and help maintain dunes
254 that limit the erosion by waves and sand resuspension (Zarnetske et al., 2012). All these
255 mechanisms foster plant survival by buffering against the physical stress induced by
256 anoxic conditions, or waves and wind erosion (Silliman et al., 2015; Maxwell et al., 2017),
257 and therefore have been used as knowledge to guide restoration of coastal areas (Table
258 2). These engineering species "modify their own niche and/or each other's niches" (*sensu*
259 Odling-Smee et al., 1996), a wide phenomenon called niche construction. Such habitat

260 modifications in turn affect the ecology of other species, thereby creating new feedback
261 loops. For instance, seagrass recruitment provides shelter from predation to meso-grazers
262 that graze on seagrass epiphytes and invasive algae, which benefits seagrasses (Maxwell
263 et al., 2017). Hence, niche construction encapsulates both biotic and abiotic changes (see
264 next section for biotic niche construction; Odling-Smee et al., 1996, 2013).

265 Organism-environment interactions often involve the modulation of local resource avail-
266 ability. On one side, consumers deplete their resources, which generates a negative
267 feedback loop: increased consumption leads to lower resource availability (negative niche
268 construction). Such a loop can switch from negative to positive for engineering or facilitat-
269 ing species (positive niche construction; Kylafis and Loreau, 2008). In intertidal systems,
270 sea-grass meadows or mussel-beds increase organic deposition, availability of oxygen and
271 nutrients, either by trapping them in their rooting system for sea-grass (Maxwell et al.,
272 2017) or by ammonium excretion in mussel beds Pfister, 2007. Ultimately, it positively
273 affects their growth. Similarly, facilitation by nurse species in stressed ecosystems (arid,
274 salted marshes, alpine) enhances local availability of resources (through the increase in
275 water infiltration, organic matter and by limiting evaporation; Filazzola and Lortie, 2014)),
276 which enhances seedling recruitment. Feedbacks bounding species and the availability of
277 resources are also observed at the ecosystem-level. Each species is involved in a recycling-
278 mediated feedback loop, where organic matter is recycled and in return positively or
279 negatively affects each trophic level through a bottom-up cascade (Loreau, 2010). These
280 loops between the ecology of species and their ever-changing local environment set the
281 theater for evolution to play (Lewontin, 1978; Odling-Smee et al., 1996; Phillips, 2016).

Population and community-level feedbacks

282

283 Individuals do not simply interact with their abiotic environment, they also eat, compete
284 or cooperate with each other. These interactions generate demographic and behavioral
285 feedbacks within and between populations of the same or different species (Fig. 2, central
286 panel).

287 Within populations, demography is modulated by the balance between growth and reg-
288 ulation. Contrary to intrinsic population growth, that acts by definition positively
289 on population demography, regulation from density-dependent mechanisms such as in-
290 traspecific competition for resources slows down population growth (*e.g.*, due to stronger
291 intraspecific competition or to the accumulation of specialized enemies), thereby acting as
292 a negative loop (see also Table 2 human-induced demographic regulation of harvested
293 systems). Negative density-dependent mechanisms regulate population growth, however,
294 conspecific density-dependence mechanisms such as cooperation can also positively affect
295 population growth (Gil et al., 2019). For example, group defense in meerkats reduces
296 the individual risk of being eaten by predators. Another example comes from the collec-
297 tive foraging of coral-reef fishes that increases their foraging rate when they have more
298 conspecific in their neighborhood (Gil et al., 2020). At low density, these social or sexual
299 interactions between individuals generate positive feedback between population demog-
300 raphy and growth rate (positive density-dependence). By contrast, as the population
301 grows, density-dependent negative feedbacks overcome positive ones. Such an interplay
302 of positive and negative feedback loops at the population level leads to Allee effects (Allee,
303 1931, Box 2), where the population can collapse under a defined threshold (minimum

304 viable population) as it becomes insufficient to sustain the cooperation level required for
305 population persistence (see also Kylafis and Loreau, 2008; Koffel et al., 2021).

306 Species are also embedded in a complex web of diverse interactions with other species.
307 The intertwinement of interaction in networks such as in plants-pollinators, and food-webs
308 illustrate this ‘entangled bank of species’ (Darwin, 1876), describing the diversity and
309 complexity of interspecific interactions. In a network, direct interactions between pairs
310 of species are well pictured. Direct interactions also generate a multitude of less visible
311 indirect loops that create strong interdependence of species (Ulanowicz, 2004). Because
312 positive and negative interactions such as competition, mutualism, exploitation coexist in
313 communities (Fontaine et al., 2011; Poccock et al., 2012), these loops vary in sign, strength
314 and length. Consequently, when a species enters a community, by interacting with other
315 species it modulates the emergent feedbacks that any species in the community experiences
316 (Arnoldi et al., 2022). Recently, Zelnik et al (2022) proposed the term “interaction horizon”
317 to describe the maximal length of indirect pathways connecting species that significantly
318 contribute to the community patterns and dynamics. In particular, using the network
319 approach to decompose the direct and indirect effects of perturbations (Box 1), they found
320 that indirect paths increase in importance when species interact more strongly together,
321 that is, when species assemblages are more integrated (*i.e.*, interdependent).

322 *Spatial feedbacks coupling ecosystems*

323 Feedbacks can also emerge in space when species, individuals or material flows are
324 exchanged at the landscape extent (Fig. 2). Such feedbacks have been addressed in

325 all theoretical frameworks of spatial ecology (metapopulations, metacommunities and
326 metaecosystems) and will be discussed hereafter. Considering metapopulations (*i.e.*, sets
327 of local populations linked by the dispersal of individuals), spatial feedbacks can emerge
328 from the dispersal of individuals through local patches, and thus affect the general state
329 of the metapopulation (*e.g.*, total abundance or occupancy). For instance, the “spatial
330 rescue” effect relies on the idea of a positive feedback between population abundance and
331 spatial flows of organisms. Specifically, migration from locally abundant populations
332 increases the size of small populations and therefore decreases their risk of extinction
333 (Brown and Kodric-Brown, 1977). In turn, the number of individuals dispersing from
334 these small populations increases (positive density-dependent dispersal, *e.g.*, Fonseca and
335 Hart, 1996, for an empirical example) and can rescue other populations, thereby increasing
336 the metapopulation occupancy (Hanski and Gyllenberg, 1993). Further, spatial flows
337 of individuals link local and spatial feedbacks in various ways. If local feedbacks are
338 dominated by intraspecific competition, a common expectation is that the dispersal of
339 individuals will balance competitive constraints among patches (Fretwell and Lucas, 1969).
340 If, on the other hand, the establishment of local populations is constrained by local positive
341 feedbacks (*e.g.* local Allee effects), the effectiveness of dispersal at the landscape scale is
342 reduced. For instance, local Allee effects constrain the range expansion of the gypsy moth,
343 an invasive species in North America (Tobin et al., 2007). Finally, from an evolutionary
344 point of view, spatial heterogeneity in the local states of patches is expected to select for
345 lower dispersal levels and weaker spatial feedbacks (Hastings, 1983; Parvinen et al., 2020).
346 Similarly, in metacommunities (*i.e.*, sets of local species assemblages linked by dispersal),
347 spatial feedbacks emerge from the dispersal of species through the different local assem-

348 blages. For instance, when competition among species dominates in metacommunities,
349 low dispersal leads to high local competition and competitive exclusion, so that species
350 are sorted differently in different patches depending on local environmental conditions
351 (“species sorting”; Leibold et al., 2004). When dispersal is high, spatial feedbacks homoge-
352 nize local communities making competition happen at the scale of the metacommunity
353 and may lead to competitive exclusion at this level (Mouquet and Loreau, 2003). While
354 most metacommunity works assume undirected dispersal (diffusion), dispersal can also be
355 context-dependent (*i.e.*, when it depends on the presence of resources, competitors, preda-
356 tors and mutualists, Cote et al., 2013; Fronhofer et al., 2015; Trekels and Vanschoenwinkel,
357 2019). Spatial flows of species are then directly coupled to locally dominant feedbacks.
358 In addition to flows of species, resources or information connect ecosystems in space and
359 can also generate spatial feedbacks (meta-ecosystems; Polis et al., 1997; Loreau et al., 2003;
360 Massol et al., 2011; Gounand et al., 2018a). An interesting example comes from oceanic
361 salmon that migrate upstream for reproduction. By doing so, salmon produce carcasses
362 that foster biofilm growth (Rüegg et al., 2011), and serve as resources for freshwater insects.
363 In turn, salmon juveniles feed on these insects, which positively increase their growth
364 and survival (Giannico and Hinch, 2007). A recent synthesis on spatial flows of resources
365 shows that their magnitude can be as high as the local ecosystem fluxes in some ecotones
366 (*e.g.*, freshwater-terrestrial; Gounand et al., 2018b), suggesting that the spatial feedbacks
367 generated by resource flows can be as important as local feedbacks within ecosystems.
368 These flows of species and resources generate negative and positive spatial feedbacks
369 through other ecosystems (Montagano et al., 2018). Resource flows between ecosystems
370 can relax local limitations in nutrients or carbon in each local ecosystem, therefore gener-

371 ating a mutually beneficial spatial loop (Pichon et al., 2023). However, spatial feedbacks
372 can also be negative, when dispersing individuals carry parasites or diseases that de-
373 crease the demography of local populations (anti-rescue effect; Harding and McNamara,
374 2002) or when subsidy flows correspond to low-quality resources ((Kelly et al., 2014); see
375 Modulators of feedback properties).

376 **EMERGENT PROPERTIES FROM FEEDBACK LOOPS**

377 We next review the effects of positive and negative feedback loops on species coexistence,
378 the functioning, stability and emergent spatial patterns of ecological systems.

379 *Feedbacks have contrasting effects on species coexistence*

380 Positive feedbacks can have positive effects on species coexistence. In drylands, the
381 facilitation-driven positive feedbacks between nutrients and water in soil and some facili-
382 tating species create a favorable environment for their growth but also sustain some species
383 that would not be able to persist without facilitation (Filazzola and Lortie, 2014; Bulleri
384 et al., 2016). This heterospecific positive density-dependence acts on mortality: higher
385 density of the facilitating species decreases the mortality of other species. In theoretical
386 models, this mechanism fosters species coexistence (Gross, 2008; Gil et al., 2019, see also
387 Aubier, 2020 for similar results between conspecifics).

388 Such positive effects on coexistence are however not systematic. Because of the cost
389 associated with facilitating another species that also competes for resources and space
390 (Schöb et al., 2014), the species benefiting from facilitation may eventually out-compete the

391 facilitating species when environmental conditions allow its persistence alone and decrease
392 coexistence (Koffel et al., 2021). When the positive density-dependence acts on conspecific
393 growth rates, it can also reinforce the competitive hierarchy thereby decreasing coexistence
394 (*i.e.*, reducing further the growth rate of least competitive species that already have a lower
395 abundance compared to best-competing species; Wang et al., 1999). Similar effects can be
396 observed with nutrient recycling that can promote the dominance of a single plant when it
397 increases the loss of the nutrient that limits their competitor growth (Daufresne and Hedin,
398 2005), or the dominance of a mobile consumer connecting ecosystems by foraging: if the
399 consumer forage in the ecosystem that is fertilized by its competitors, it benefits from the
400 positive recycling feedback and can exclude other mobile consumers (Peller et al., 2021).
401 Positive feedbacks may also generate priority effects when species pre-empt or modify
402 available niches, ultimately limiting further invasions (Drake, 1991; Fukami, 2015). Such
403 priority effects may be reinforced over evolutionary timescales when early-arriving species
404 adapt to local conditions and radiate available niches (“monopolization hypothesis”,
405 De Meester et al., 2002; Leibold et al., 2022). These mechanisms may explain the patterns of
406 dominance and low species diversity in some ecosystems such as peatlands, boreal forests,
407 or coastal mangroves (Zobel et al., 2023). For instance, in bog areas where the pH is low, a
408 moss species (*Sphagnum*) can colonize and modulate the local conditions by increasing
409 acidity in its neighborhood and limiting decomposition (Clymo, 1984), which promotes its
410 ecosystem dominance. By contrast, in areas where pH is higher, the community is more
411 diverse and composed of grasses, forbes, and sedges (Laine et al., 2021). Similarly, in the
412 context of biotic invasion, an invader experiencing a positive feedback while invading a
413 resident species will enter a community and replace the resident species (*i.e.*, invasional

414 meltdown; 'Arnoldi et al., 2022). By contrast, when positive feedbacks are more frequent
415 between native plant species than between invasive ones, they offer a mechanism of
416 resistance against invasions (Yin et al., 2022).

417 Links between negative feedbacks and species coexistence have a long history in ecol-
418 ogy. Classical theory suggests that coexistence is enhanced when intraspecific negative
419 feedbacks are stronger than interspecific competition (*e.g.*, when species have large niche
420 differences; Levins, 1966). The later work of Janzen and Connell (Janzen 1970, Connell,
421 1971) suggested that the accumulation of pathogens near adult trees inhibiting the sur-
422 vival and recruitment of their juveniles would favor coexistence (a mechanism known as
423 negative-conspecific density-dependence or self-regulation). Since then, there has been
424 accumulated evidence of the positive effect of negative density-dependence on coexistence
425 using both experiments (Klironomos, 2002; Mangan et al., 2010; Teste et al., 2017), regional
426 observations (Johnson et al., 2012; LaManna et al., 2017), and theory on plant-soil feedbacks
427 (Bever et al., 1997; Loeuille and Leibold, 2014; Eppinga et al., 2018). This is corroborated by
428 the negative relationship between the species abundances in plant communities and the
429 strength of the negative feedback loop with the soil: least abundant species are the ones
430 involved in the stronger negative feedbacks with the soil (Mangan et al., 2010; Johnson
431 et al., 2012).

432 While we simplified the explanation above by splitting positive and negative feedback
433 loops, in communities, feedbacks of different signs are generally intertwined. In addition,
434 given that their signs can change along environmental gradients and timescales (see
435 Modulators of feedback properties section; Box 2), understanding how each loop relates
436 to species coexistence can be tricky. In this regard, recent developments of niche theory

437 explicitly link the feedbacks emerging from organism-environment interactions (both
438 biotic or abiotic environment), the niche of each species and species coexistence (see Koffel
439 et al., 2021 for details on the mathematical framework). Another possible way of unveiling
440 the links between different feedback loops and the patterns of species coexistence is to
441 acknowledge the scale at which each feedback loop acts (see Zobel et al., 2023). At the
442 scale of an individual, or a patch of vegetation, species-level positive feedbacks driven by
443 facilitation may promote species coexistence by enhancing environmental conditions and
444 expanding the niche of other species (Bulleri et al., 2016; Koffel et al., 2021). Yet, when this
445 positive niche construction scales-up to be dominant at the ecosystem level, it promotes
446 the dominance of a single species (*e.g.*, due to priority effects or monopolization). Over
447 timescales, niche construction may shift to ecosystem-level negative feedbacks that limit
448 such cases of dominance, by accumulating resources that promote competition-driven
449 coexistence mechanisms (Box 2). Therefore, in many ecosystems the maintenance of a high
450 number of species may tie in the balance of feedbacks having various signs and acting at
451 different scales. We argue for the development of a theory investigating how feedbacks
452 acting at different scales modulate species coexistence.

453 *The central role of organism-abiotic resource feedbacks on ecosystem functioning*
454 *and development*

455 Feedbacks are key to understanding how ecosystems and landscapes function. At the root
456 of many ecosystems lies a positive feedback loop between plants, performing photosyn-
457 thesis, and decomposers that close the recycling loop of matter, a loop that can also be

458 accelerated by herbivores (Mazancourt et al., 1998). Depending on whether decomposers
459 are limited by either carbon or nutrients, plants and decomposers compete for nutrients or
460 become mutualists, respectively (Daufresne and Loreau, 2001). In the latter case, plants
461 and decomposers benefit from the by-product of the other (decomposers decompose litter
462 produced by plants and make nutrients available again for plants) making an autocatalytic
463 loop (*sensu* Veldhuis et al., 2018; Fig. 3). As long as this self-reinforcing loop is fed with
464 energy (*i.e.*, light for photosynthesis), resource competition leads to the selection of species
465 with the highest resource-use intensity (lowest R^* according to resource competition the-
466 ory, Tilman, 1982) within plants and decomposers, and consequently increases biomass
467 and energy production while minimizing resource losses (Loreau, 1998; see also Box 2).
468 Taken together, this simple system shows that feedbacks can generate auto-catalytic pro-
469 cesses that determine the development and functioning of the ecosystem (Odum, 1988;
470 Ulanowicz, 2009; Lenton et al., 2021). Interestingly, such ecosystem principles can be ex-
471 tended to the landscape extent, where subsidy flows can connect ecosystems with different
472 functioning (Harvey et al., 2023). Freshwater or benthic ecosystems tend to have a net
473 heterotrophic functioning (respiration > primary production; Gounand et al., 2018b, 2020),
474 while terrestrial and pelagic ecosystems have on average a net autotrophic functioning
475 but transfer less efficiently energy to the higher-trophic levels (Shurin et al., 2006), making
476 them carbon sources at the landscape scale (Fig. 3). In addition, these ecosystems cor-
477 respond to communities with different carbon, nitrogen and phosphorous needs (Elser
478 et al., 2000). Hence, when looking at terrestrial-freshwater or benthic-pelagic ecotones,
479 it appears a spatial analogy of the plant-decomposer relationship: when nutrients and
480 detritus are spatially exchanged and meet the local needs of communities within each

481 ecosystem (*i.e.*, the exported by-products of an ecosystem bring the resource limiting the
482 other). In that case, a positive feedback loop can emerge at the landscape scale, fostering
483 the landscape-scale production and maximizing the use of nutrients and carbon between
484 ecosystems (Pichon et al., 2023; see also Modulators section). Taken together, we notice
485 similar autogenic development of ecological systems driven by autocatalytic loops at
486 different spatial scales.

487 *Feedbacks properties determine ecosystem stability and fragility*

488 Feedbacks are also intrinsically linked to the stability of ecological systems. While positive
489 feedbacks can maintain an ecosystem in a high biomass state, they can also promote its
490 fragility. Indeed, the strong positive interdependencies of entities can serve as a medium
491 to amplify and propagate perturbations across the whole system. In multiple ecosystems
492 such as lakes, coral-reefs, and drylands, where strong positive feedbacks are observed,
493 under some conditions, small perturbations can lead to a sudden change in ecosystem
494 state (so-called “catastrophic shifts “; Scheffer, 2009; Kéfi et al., 2016). Similar alternative
495 ecosystem states have also been described at the population and community levels (Table
496 1; Fig. 3). Priority effects from positive feedbacks discussed in the coexistence section can
497 generate alternative community states depending on the order of assembly in the com-
498 munity (“historical contingency”; Case, 1990; Fukami and Nakajima, 2011). Importantly,
499 the stability of communities is impacted differently when feedbacks are species-specific
500 (*i.e.*, depends on the species identity, such as a pollinator specialized on one plant) or
501 aggregated (*i.e.*, which does not depend on the species identity; *sensu* Karatayev et al.,

2023). With aggregated feedbacks all species experience the same Allee effect, and therefore a single perturbation can drive the collapse of the community simultaneously (Lever et al., 2014; Karatayev et al., 2023). With positive species-specific feedbacks, each species experiences a specific Allee threshold, which thereby does not necessarily cause cascades of extinctions in the community (Aguadé-Gorgorió et al., 2023). For example in plant-pollinator communities, because specialist species experience stronger positive feedbacks, they are more vulnerable to perturbations than generalist species (Saavedra et al., 2013). Because some traits of interacting partners can be lost over evolutionary timescales (*e.g.* loss of genes to produce arginine in leaf-cutter ants), positive interdependencies between species can be strengthened (*e.g.*, ants rely on fungus for arginine; Ellers et al., 2012). In some cases, such high specialization to a partner can lead to population collapse due to an evolutionary trap (Singer and Parmesan, 2018) or evolutionary ‘cul-de-sac’ at longer timescales (Perez-Lamarque et al., 2022). Related impacts of positive dependencies can be drawn for spatial systems. In this case, the spatial dependencies emerge from species dispersal or spatial flows of resources. When such flows are important, a small local perturbation can spread in space through a traveling wave changing the state of the connected populations (Keitt et al., 2001; Villa Martín et al., 2015; Saade et al., 2023), or leading to alternative ecosystem states (Gounand et al., 2014; Fig. 3). To summarize, mutual benefits create dependencies and integration of entities, thereby fostering their joint persistence or possible collapse. In other words, interdependency can beget fragility (Vespignani, 2010; Centeno et al., 2015; Brummitt et al., 2015).

Conversely, negative feedbacks have been related to homeostasis, or regulation (Odum, 1969; Wiener, 1948). Negative feedbacks stabilize ecological systems, avoiding infinite

525 growth, either by mediating a return of the system to its equilibrium after a perturbation
526 (*i.e.*, resilience), or by constraining dynamics in cycles (*i.e.*, avoiding runaway). Cycles are
527 observed when long-delayed negative loops overcome shorter ones (Barraquand et al.,
528 2017; Lever et al., 2023), as exemplified by the slow-fast dynamics of shallow lakes where
529 the interaction between a slow loop with phosphorus and a fast loop with turbidity
530 induces primary producer cycles between macrophytes and microalgae (van Nes and
531 Scheffer, 2007). Theory also suggests that cycles emerge when the time delay of negative
532 feedbacks is long compared to focal entity timescale (*e.g.*, long delayed demographic
533 regulation compared to slow population' growth; (May, 1973b)). In such cases, the system is
534 destabilized because growth and negative regulation feedbacks are out-of-phase (Ramesh
535 and Hall, 2023; Yang et al., 2023). These delayed feedbacks can be promoted by legacy
536 effects that can persist for years (Hastings et al., 2007; Albertson et al., 2022). For example,
537 drought legacy effects were shown to switch plant-soil feedbacks of two grassland species
538 from positive to more negative (and potentially destabilizing) feedbacks (Kaisermann et al.,
539 2017). In this experiment, drought led to turnovers in soil fungal communities, which may
540 explain the change in feedback sign, although the precise mechanism and the impact of
541 the change in feedback sign on the stability of the system were not elucidated (see De Vries
542 et al., 2023 and references therein for further discussion on the topic).

543 Because ecological systems host both positive and negative interaction types, loops of
544 opposite signs are intertwined. Theoretical studies investigating the impact of this diversity
545 of interactions in communities showed that inclusion of negative interactions in mutualistic
546 communities can stabilize them (*i.e.*, generates negative loops; Mougi and Kondoh, 2014;
547 Coyte et al., 2015). This is also corroborated in a simple plant-pollination-herbivore module,

548 where coexistence is favored by the mutualistic part (positive feedback) while stability
549 is favored by the herbivory (negative feedback) so that overall maintenance requires a
550 specific balance between the two interactions (Yacine and Loeuille, 2022).

551 *Feedback loops induce spatial heterogeneity*

552 Lastly, feedbacks can generate emergent spatial patterns (self-organization; Fig. 3; Rietkerk
553 and van de Koppel, 2008). Different patterns can be distinguished depending on the type
554 of feedback that has generated them (scale-dependent or density-dependent). On the one
555 hand, local feedbacks between species and their abiotic resources generate self-organized
556 patterns and have been observed across a wide range of biological systems (Rietkerk and
557 van de Koppel, 2008; Kondo and Miura, 2010), including drylands (Rietkerk, 2004; Kéfi
558 et al., 2007), planktonic communities (Okubo, 1980), salt marsh communities (Zhao et al.,
559 2021), and mussel beds (Liu et al., 2014). When seeing these ecosystems from the sky,
560 one may see a two-phase mosaic with species aggregated in space separated by open
561 areas. This spatial heterogeneity results from two antagonistic forces acting at different
562 spatial scales: a short-range facilitation (*e.g.*, by plants or mussels) and a long-range
563 competition due to the redistribution of nutrients or water in space. The balance between
564 the two determines the sign and the strength of the feedback, as well as the type of pattern
565 observed. In particular, when competition is high, these patterns show a regular shape
566 (Turing-like; Klausmeier, 1999; Rietkerk and van de Koppel, 2008), while they have more
567 irregular structures, characterized by a scale-free patch size distribution (*i.e.*, power-law)
568 when facilitation dominates the system (Kéfi et al., 2007; Scanlon et al., 2007). Importantly,

569 the emergent feedback from the balance of interactions in space is somehow similar to the
570 temporal balance of feedback loops found to trigger oscillations in shallow lakes (van Nes
571 and Scheffer, 2007): both spatial patterns and oscillatory dynamics emerge when there are
572 differences in scale (space or time) between loops of contrasting signs.

573 On the other hand, biogeomorphic feedbacks in intertidal ecosystems described previously
574 generate another type of self-organized pattern, referred to as phase separation (see Liu
575 et al., 2016 for review). Contrary to Turing patterns, they involve a density-dependent
576 aggregation and are found across a wide range of systems (Ge and Liu, 2021; Siteur et al.,
577 2023). In seagrass meadows for example, because seagrass patches favor the accumulation
578 of sediments, it limits grazing by waterfowl during low tide as they cannot feed on dense
579 seagrass patches elevated above the water and thereby graze on the water-logged hollows
580 (Van Der Heide et al., 2012). Grazing pressure is therefore determined by the spatial
581 distribution of the seagrass, which in turn is controlled by grazing (via consumption). Ulti-
582 mately, this feedback between grazing/foraging behavior and seagrass density generates
583 emergent spatial patterns (Ge and Liu, 2021). These types of patterns differ from the Turing
584 ones as the mean patch-size coarse-grain over time and their patch-size distribution is best
585 described by a log-normal distribution (Siteur et al., 2023).

586 Spatial heterogeneity can also emerge from single negative or positive feedbacks. Because
587 mobile consumers such as predators actively track hotspots of resources, they can produce
588 heterogeneity in the spatial distribution of their prey by feeding preferentially on some
589 areas rather than on others (Barraquand and Murrell, 2013). In addition, by spending
590 more time in these areas, they excrete more nutrients locally, which can positively feed
591 back on their growth (Anderson et al., 2010; McLoughlin et al., 2016). For example,

592 nutrient-rich guano deposited by birds on islands has been shown to leach to neighboring
593 marine ecosystems, boost fish growth, and positively affect sea-bird demography in return
594 (Graham et al., 2018; McInturf et al., 2019). Hence, localized animal wastes generate
595 nutrient hotspots and spatial heterogeneity (Monk and Schmitz, 2022; Ferraro et al., 2022;
596 Johnson-Bice et al., 2023). Interestingly, the interaction between this positive recycling loop
597 and a negative one acting at lower scale can erase spatial heterogeneity. In the example of
598 bird guano, islands where birds deposit large amounts of nutrients can also be invaded by
599 coconut trees that benefit from the bird nutrient enrichment. By consuming the nutrients
600 locally, and inhibiting the deposition of guano because seabirds prefer to roost on non-
601 invasive trees, these invasive trees limit the emergence of heterogeneity (Young et al., 2010;
602 Table 2). By being antagonists and acting at different scales, feedback loops therefore may
603 cancel each other.

604 **MODULATORS OF FEEDBACK PROPERTIES**

605 *Physical constraints can switch the sign of feedbacks*

606 Physical constraints are a strong modulator of feedbacks across spatial scales: in arid
607 ecosystems, local slopes determine how water is redistributed in space and acts in fine
608 on the scale of interspecific competition. Higher slopes lead to change from spot to stripe
609 patterns due to water flowing downhill between each stripe; this structures competition in
610 space: competition is stronger in the direction of the slope compared to orthogonally to the
611 slope (Klausmeier, 1999; Deblauwe et al., 2011). Similarly, because of gravity, the structure

612 of ecosystems (*e.g.*, convex versus concave profiles; Lindeman, 1942) changes the strength
613 of spatial feedbacks through ecosystems. Watersheds or streams receive more resource
614 flows than grasslands or forests (Gounand et al., 2018b), which modulates trophic cascades
615 (Leroux and Loreau, 2008) and how many subsidies are exported back. In addition, in
616 alpine, arid, or salt marshes ecosystems, the sign of the feedback among plants can change
617 with the level of stress (“stress-gradient hypothesis”; (Callaway et al., 2002; Maestre et al.,
618 2009)). In particular, what is generally being observed across ecosystems, is that stress
619 promotes positive facilitation-driven feedbacks (McCluney et al., 2012; He et al., 2013;
620 Piccardi et al., 2019).

621 *Species traits can change the sign and strength of feedback loops*

622 Stoichiometry of organisms (*i.e.*, elemental composition) modulates the recycling-mediated
623 feedback loops at different spatial scales. At the ecosystem-level, when plants get richer
624 in carbon due to nitrogen depletion for instance or predation risk (Hawlena and Schmitz,
625 2010), their detritus have a higher C:N ratio due to more recalcitrant carbon structures (*e.g.*,
626 lignin), which slows down the decomposition process (Cherif and Loreau, 2013; Hawlena
627 et al., 2012) and reduces the strength of nutrient cycling. In a meta-ecosystem context,
628 we previously stressed that spatial positive feedbacks can emerge through ecosystems
629 (see Emergent properties from feedback loops section), but the stoichiometry of spatial
630 flows may also exacerbate local stoichiometric mismatches between consumers and their
631 resources (*e.g.*, decomposers and detritus). This happens when a large magnitude of
632 nutrient-poor terrestrial litter falls into lakes or streams (Kelly et al., 2014), and leads

633 to reduced secondary production, eutrophication, or hypoxia (Subalusky et al., 2015;
634 Dutton et al., 2018). In this case, terrestrial and aquatic ecosystems are involved in a
635 negative feedback loop: terrestrial ecosystems benefit from nutrient-rich spatial subsidies,
636 while stoichiometric mismatch impairs freshwater functioning (Pichon et al., 2023). Trait
637 variation can largely constrain the sign and intensity of these feedbacks. In arid ecosystems,
638 whether plants are involved in positive (facilitation) or negative (exploitation) niche
639 construction with the soil depends on their position on the slow-fast functional gradient:
640 contrary to slow facilitating species, fast exploitative ones exhibit high density tissues
641 and leaf-dry mass adapted to the conditions beneath canopies (Liancourt et al., 2005;
642 Butterfield and Briggs, 2011). As a consequence, variation in species traits can change
643 the emergent patterns in ecosystems such as intertidal areas where the stiffness and the
644 density of plant shoots have been shown to modulate the flow velocity, the sedimentation
645 rates, and ultimately, the scale-dependent feedback (Zarnetske et al., 2012; Bouma et al.,
646 2013; Schwarz et al., 2015 in coastal areas and Box 2).

647 Last, we want to emphasize species with uncooperative strategies (cheaters), and their
648 cascading effect on the stability of communities. Cheating is a phenomenon in cooperative
649 systems where some species have evolved an uncooperative strategy by benefiting from
650 an interaction without paying the associated cost (*e.g.*, Klironomos, 2003; Genini et al.,
651 2010). While there are constraints upon cheating (Perez-Lamarque et al., 2020), when a
652 cheater enters the community, it exercises a negative effect on its mutualistic partners,
653 consequently switching some mutually beneficial feedbacks to negative ones. Such changes
654 of the emergent biotic feedbacks ultimately affect the stability of communities (*e.g.*, Mougé
655 and Kondoh, 2012; Coyte et al., 2015). Duchenne et al. (2023) recently suggested that when

656 cheaters emerge among pollinators, they reduce the community-level positive feedback
657 between mutualistic species, which negatively impacts community persistence.

658 **TOP-DOWN AND BOTTOM-UP VIEWS OF FEEDBACK'**

659 **REGULATION**

660 We reviewed existing knowledge on the tight bounds between entities and macroscopic
661 systems' behavior in complex adaptive systems. A question remains, however: how are
662 feedbacks modulating the system? Feedbacks between entities contribute to the dynamics,
663 patterns and stability of the whole system (see "Emergent properties" section). In this
664 regard, the system can be constrained by the feedbacks between lower-level entities (*i.e.*,
665 feedbacks as bottom-up regulating forces). At the same time, the system's organization
666 may reciprocally impose structural or stability constraints on the interacting entities (*i.e.*,
667 the system is a top-down regulating force). To illustrate this top-down view of systems'
668 functioning, let us consider the non-random organization of mutualistic networks and
669 food-webs. Nestedness is often observed in mutualistic networks, with specialist species
670 interacting more with a subset of the more generalist species (Bascompte et al., 2003).
671 This non-random structure has been shown to reduce interspecific competition among
672 mutualistic partners (Bastolla et al., 2009), widen the conditions of coexistence (Rohr et al.,
673 2014), and ultimately promote the stability of mutualistic communities (Thebault and
674 Fontaine, 2010). In the same vein, in food webs, studies have sought to explain food-
675 chain lengths or the skewed distribution of interaction strengths towards a dominance

676 of weak links in food webs (De Ruiter et al., 1995; Wootton and Emmerson, 2005): longer
677 food-chains are less stable (Pimm and Lawton, 1977) and weak links promote food-webs
678 stability (Yodzis, 1981; McCann et al., 1998; Neutel et al., 2002; Rooney et al., 2006). Notably,
679 such system's non-random organization may explain why empirical networks include
680 remarkably few feedback loops (Albergante et al., 2014; Domínguez-García et al., 2014;
681 Johnson and Jones, 2017). Hence, under such "systemic selection" (*sensu* Borrelli et al.,
682 2015), the macroscopic properties of a system (*e.g.*, stability, robustness or coherence)
683 constrain the links connecting species, the strength of interactions, and ultimately the
684 emergent feedbacks in the community.

685 CONCLUSION

686 Since the early cybernetic interest in positive and negative feedbacks, ecological studies
687 have scaled-up from organism level to the whole landscape to understand how feedback
688 loops could generate diverse, emergent properties. Throughout our contribution, we
689 emphasize, from locally interacting species to spatially connected ecosystems, the ubiquity
690 of feedbacks interacting across space and time and levels of organization, and their im-
691 portance in generating macroscopic patterns, such as species diversity, emergent spatial
692 patterns or the functioning and stability of ecological systems. Furthermore, properties
693 such as species traits, or physical constraints modulate the strength and sign of feedback
694 loops, and ultimately the pattern they generate across scales.

695 All the research reviewed suggests that knowledge about ecological feedbacks is essential
696 to improve our fundamental understanding of the interdependence of ecological systems

697 across scales and levels of organization. Given their contribution to the functioning and
698 stability of ecological systems, acknowledging the links connecting species and ecosystems
699 (*i.e.*, ecological interdependencies) could help improve the design of effective conservation
700 and restoration measures that integrate local and spatial dependencies.

BOX 1: QUANTIFYING THE DYNAMICAL EFFECT OF A FEEDBACK LOOP

Let us consider the recycling loop that links herbivores and decomposers as an illustration (Fig. below). Herbivores feed soil decomposers through their carcasses and excretions. Microbes and fungi decompose organic matter, making nutrients available again for plants to grow and indirectly benefiting herbivores that feed on plants. Quantifying the strength of this recycling loop is not sufficient in itself to understand how the feedback loop modulates the transient response of herbivores (or any other compartment) to a perturbation such as nutrient enrichment. To do so, one can compare how herbivores deviate from its equilibrium following nutrient enrichment, as well as their recovery dynamics with and without the feedback loop. For the system in the Figure of Box 1, it can be done by replacing the link between decomposers and the nutrient compartments with a controlled inflow of nitrogen, corresponding to the flow at equilibrium when the loop is closed (*i.e.*, breaking the recycling loop). Then, comparing the system with and without the feedback loop allows us to understand how the feedback loop contributes to transiently increasing or decreasing the perturbation even if the long-term equilibrium in the two situations might be the same. This method has been for instance applied to quantify the impact of the feedback emerging from nutrient recycling on the asymptotic stability of food web (Quévreux et al., 2021).

713

BOX 2: FEEDBACKS ACROSS TEMPORAL SCALES (1/2)

Ecological feedbacks do not only vary in space, but also in time, along the development of populations, communities or ecosystems. At the population level, positive feedback loops can dominate at small populations, while negative feedbacks (*e.g.*, resource competition) seem inevitable at high population levels. Starting at low levels, Allee effects will constrain the development of the population, as too few individuals are present for cooperation or group defense to be effective. This creates a minimum viable population. If, however, the population manages to pass this threshold, the positive feedback will act to favor population growth, eventually leading to a state where individuals are abundant and resources are scarce. Competition for resources then creates a dominant negative feedback. When reaching a demographic threshold, there is therefore a switch in the sign of the dominant feedback loop that individuals experience: from positive feedbacks to negative ones. Such changes in the sign of dominant feedbacks can also happen along ontogenies. This is for instance the case of some shrubs in drylands: adult shrubs facilitate the establishment of juveniles, while juveniles compete with adults for the availability of water and nutrients (Miriti, 2006). When juveniles grow, adult plants no longer facilitate their recruitment but compete for the resource availability, therefore changing the sign of the dominant feedbacks from positive to negative. From a community point of view Yin et al. (2022) highlight how local feedbacks can vary in time within communities, along succession. The authors show that facilitative interactions are globally as frequent as competitive interactions in New Jersey grasslands over fifty years. They also reveal that facilitation among species dominates at the early stages of their settlement (colonization probability and general occurrence probability), while it is less common at later stages (survival of species in the patch and the growth of their population). This leads to a succession of dominant feedbacks (from positive to negative) that is akin to the one proposed above within populations (Allee effects).

BOX 2: FEEDBACKS ACROSS TEMPORAL SCALES (2/2)

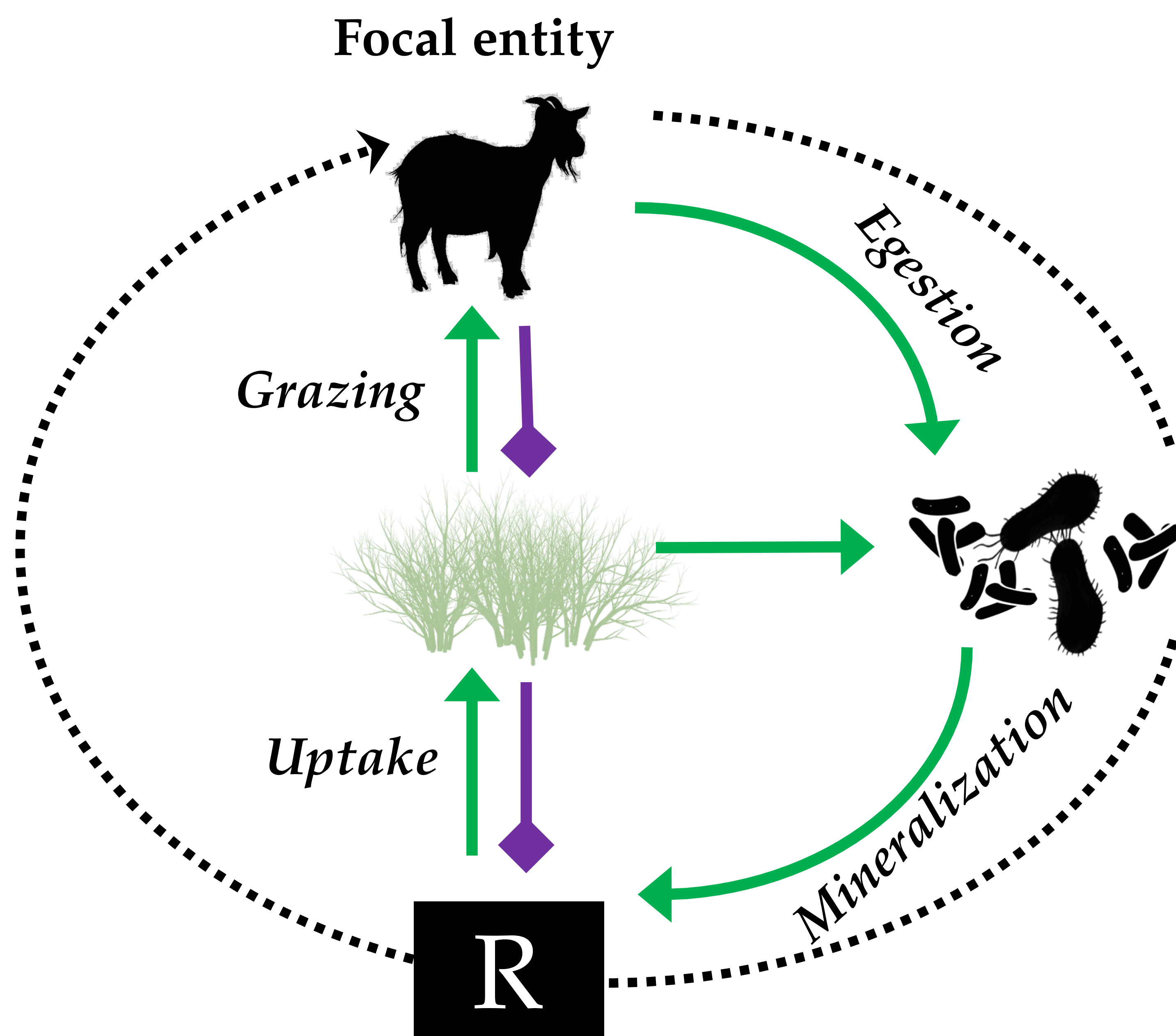
At the ecosystem level, temporal succession of feedbacks along successional stages have also been discussed. In their seminal work, Odum and Margalef suggested that ecosystem development would lead to increasing dominance of negative feedbacks regulating ecosystem functions and increasing its stability to external perturbations (Margalef, 1963; Odum, 1969). While the argument was mostly verbal at the time, there is now evidence of such changes along successional times. In nitrogen-poor ecosystems, facilitation dominates at early-successional stages because some nitrogen-fixing plants increase the availability of nitrogen, which positively affects their growth capacities (positive feedbacks with soil). As succession goes, higher availability of nitrogen drives phosphorus limitation and competitive exclusion of early nitrogen-fixing species by late-colonizing ones eventually leading to dominant negative feedbacks during the late stages (Menge and Hedin, 2009; Koffel et al., 2021). Finally, this succession from positive to negative feedbacks can also be observed in the context of species invasion. For instance, the “enemy release hypothesis” suggests that invasive species may escape pathogens at initial stages, therefore leading to positive (or weakly negative) feedback with the soil at initial stages. However, the accumulation of pathogens in later stages contribute to stronger negative feedbacks (Klironomos, 2002; Diez et al., 2010). Importantly, because feedbacks change over successional stages, they can scale up to produce self-organized patterns at different times. This is the case in intertidal systems, where fast colonizer plants do not self-organize in patches because they produce many seedlings that rapidly occupy most of the landscape and stabilize existing wetland channels (Schwarz et al., 2018). Later colonizing plants, on the other hand, are characterized by a higher lateral expansion which leads to stronger biogeomorphic feedbacks and the emergence of new vegetation-induced channels.

716 **Figure Box 1: Understanding a feedback loop using two complementary approaches.** The
717 feedback strength is obtained by multiplying the weight of the links along the circular path
718 (here, egestion, mineralization, uptake, and grazing processes; left panel). The feedback
719 strength estimates whether a loop will self-amplify (when positive) or limit (when negative)
720 the effects of a perturbation. This also allows to compare the strength of different loops
721 in a system. As a complementary approach, to understand how this feedback modulates
722 the transient dynamics of the focal entity (here the herbivore) following a perturbation
723 (here nutrient enrichment), one needs to control for the loop, and compare the dynamics
724 of recovery and the distance to equilibrium with the feedback and when controlling for it
725 (right panel). “R” stands for resources.

Complementary approaches to quantify feedbacks

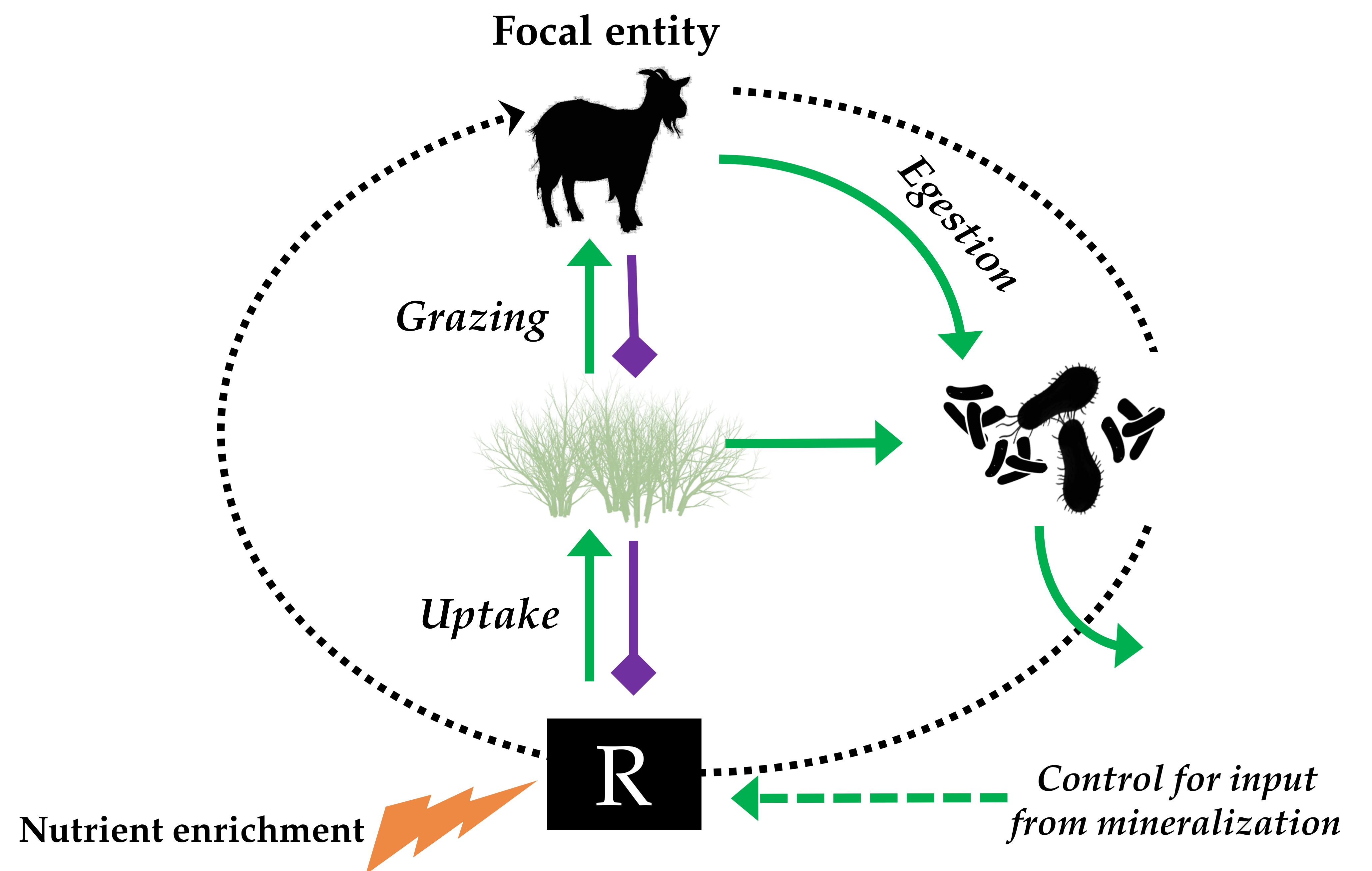
Network approach for feedback strength

Hannon 1973, Neutel et al., 2002



Loop-control approach for feedback dynamical property

Quévreur et al., 2021, Theis et al. 2021



Feedback strength = Egestion * Mineralization *
Uptake * Grazing

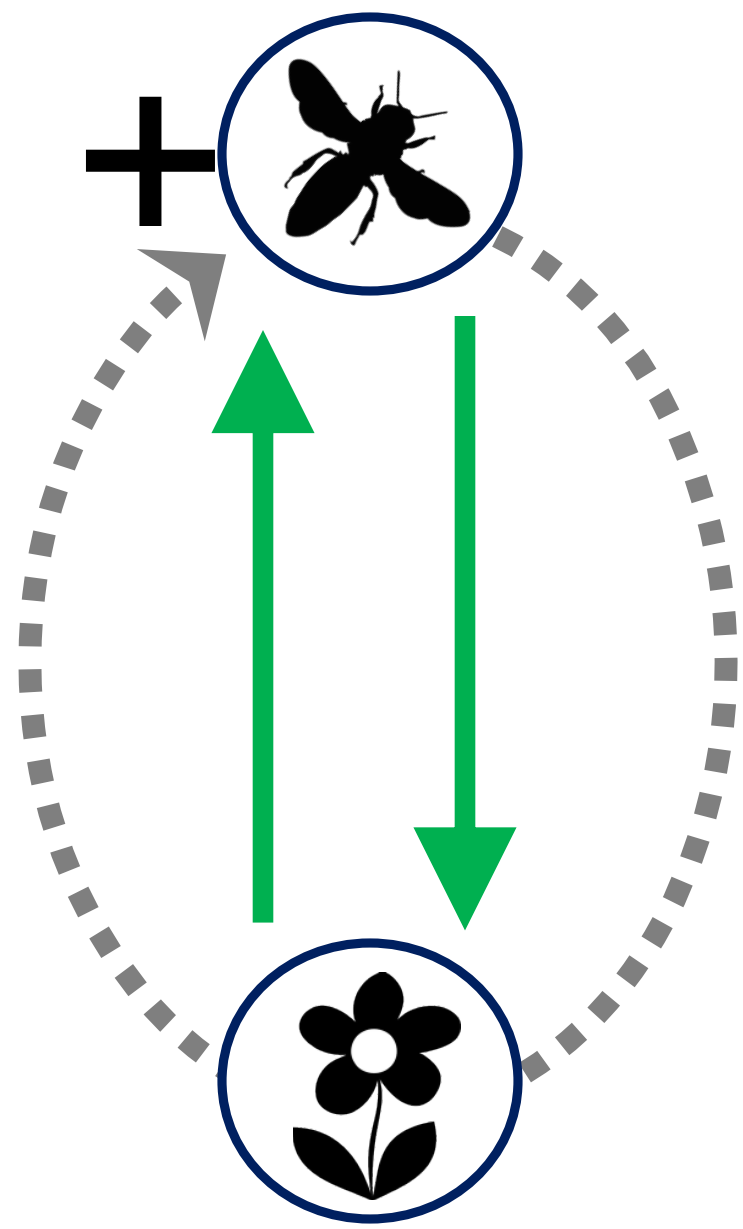
Feedback dynamical effect = difference in transient dynamics between control and full loop

726 **Figure 1: The length and sign of the feedback loop.** Feedbacks can emerge when
727 two species interact through mutualism (positive feedback) or between a species and
728 its resource (negative feedback). Some feedback loops have longer lengths such as in
729 intransitive competitive networks, where the feedback sign is determined by the number
730 of negative links involved in the loop. “R” stands for resources.

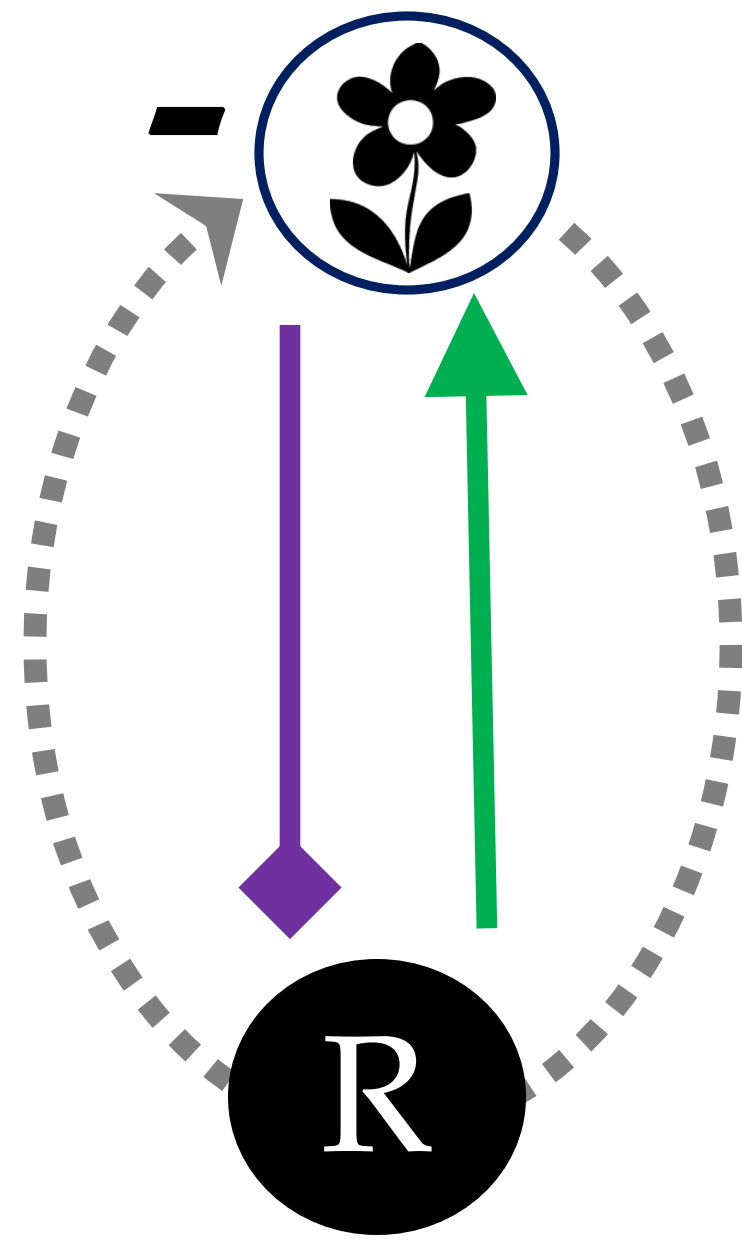
Feedback length & sign

Feedbacks of length 2

Facilitation
positive loop

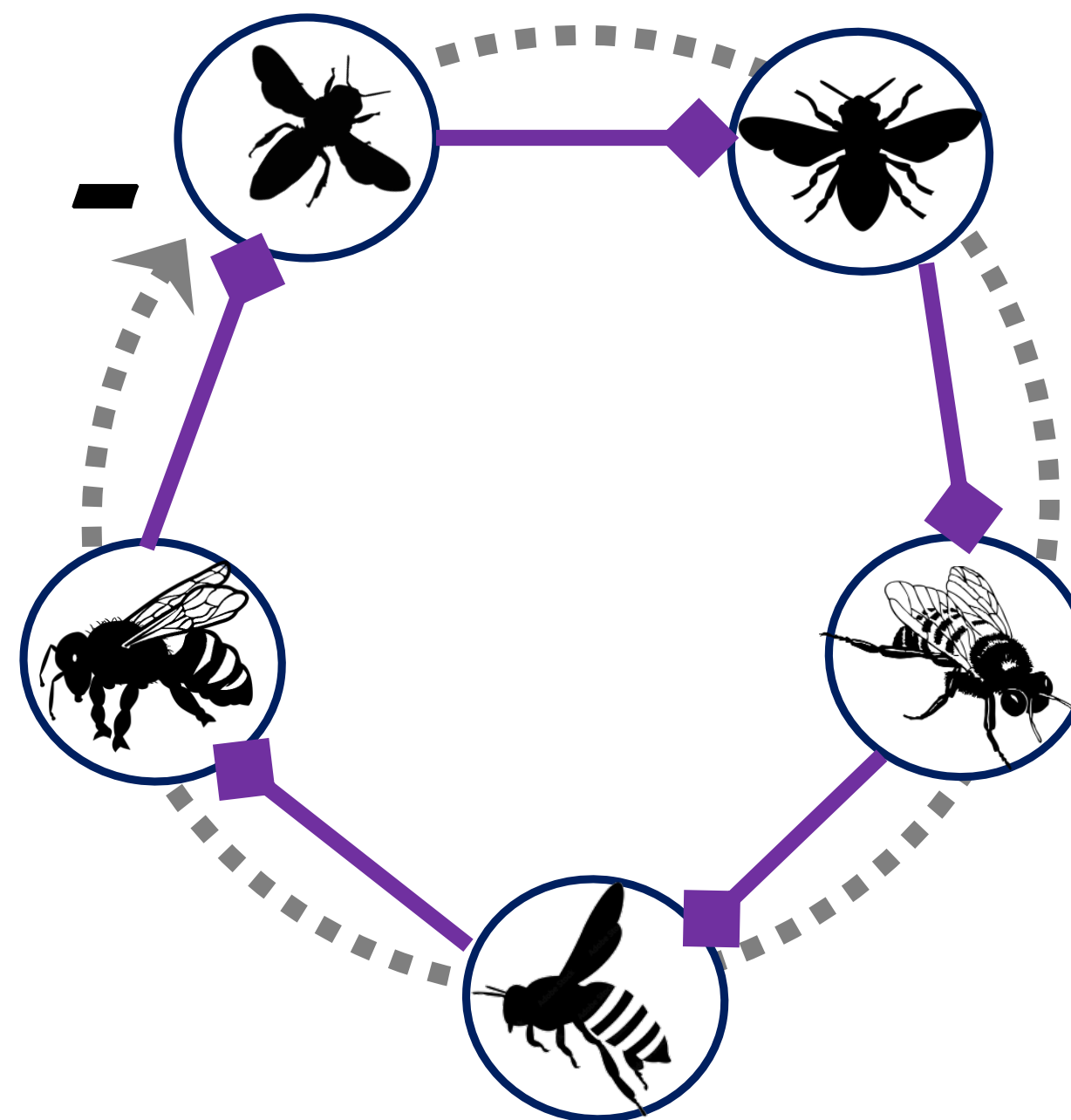


Consumption
negative loop

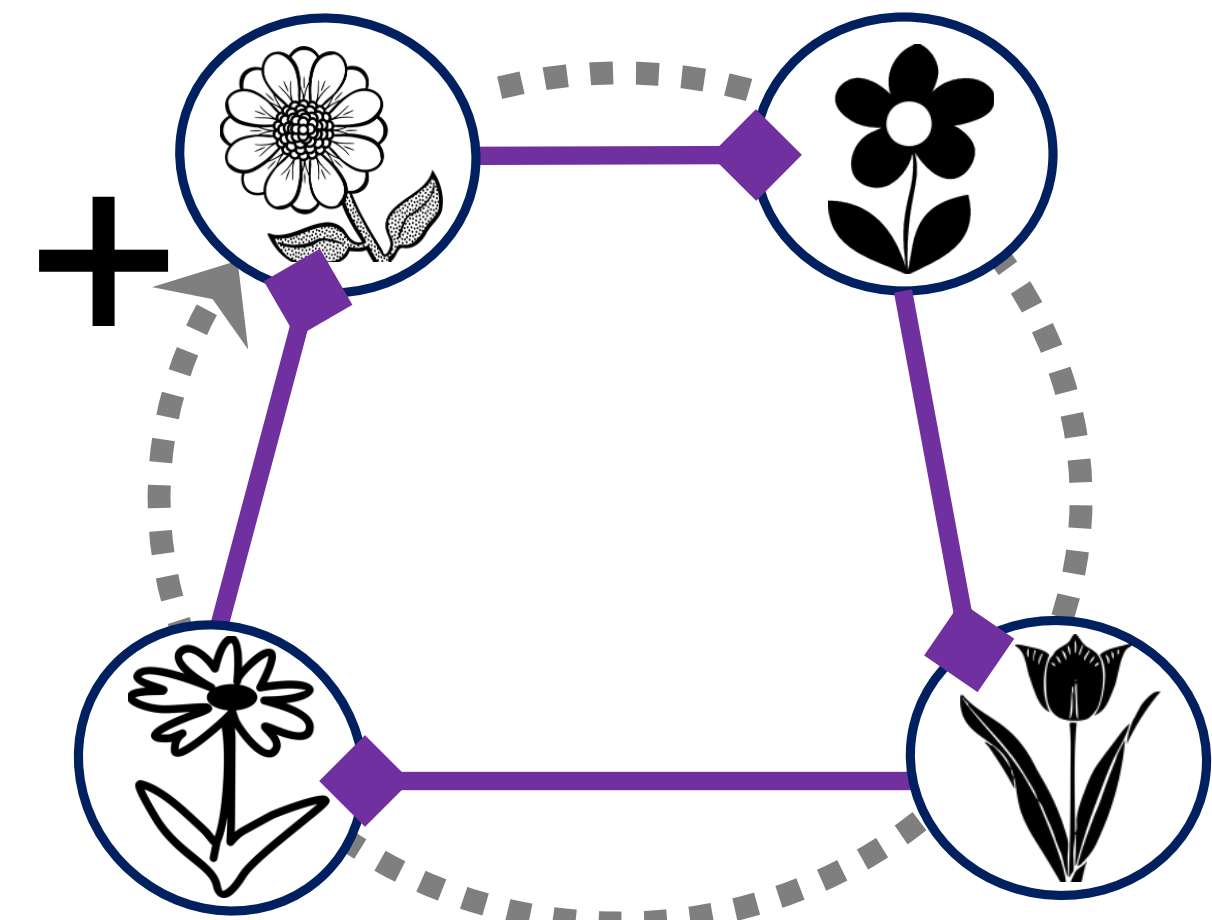


Longer feedbacks and feedback signs

Negative
intransitive loop



Positive
intransitive loop



→ Positive link

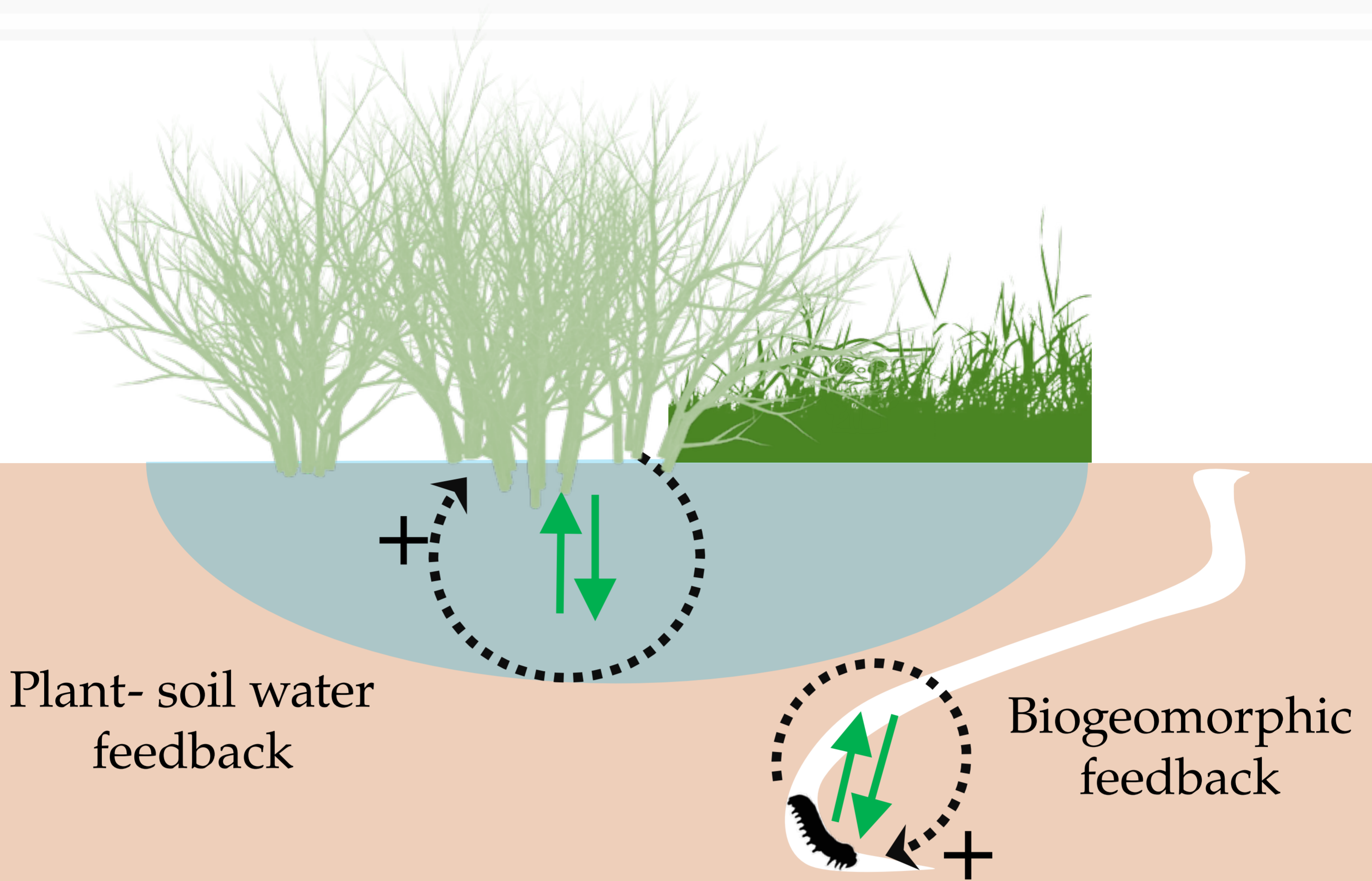
→ Negative link

→ Feedback loop

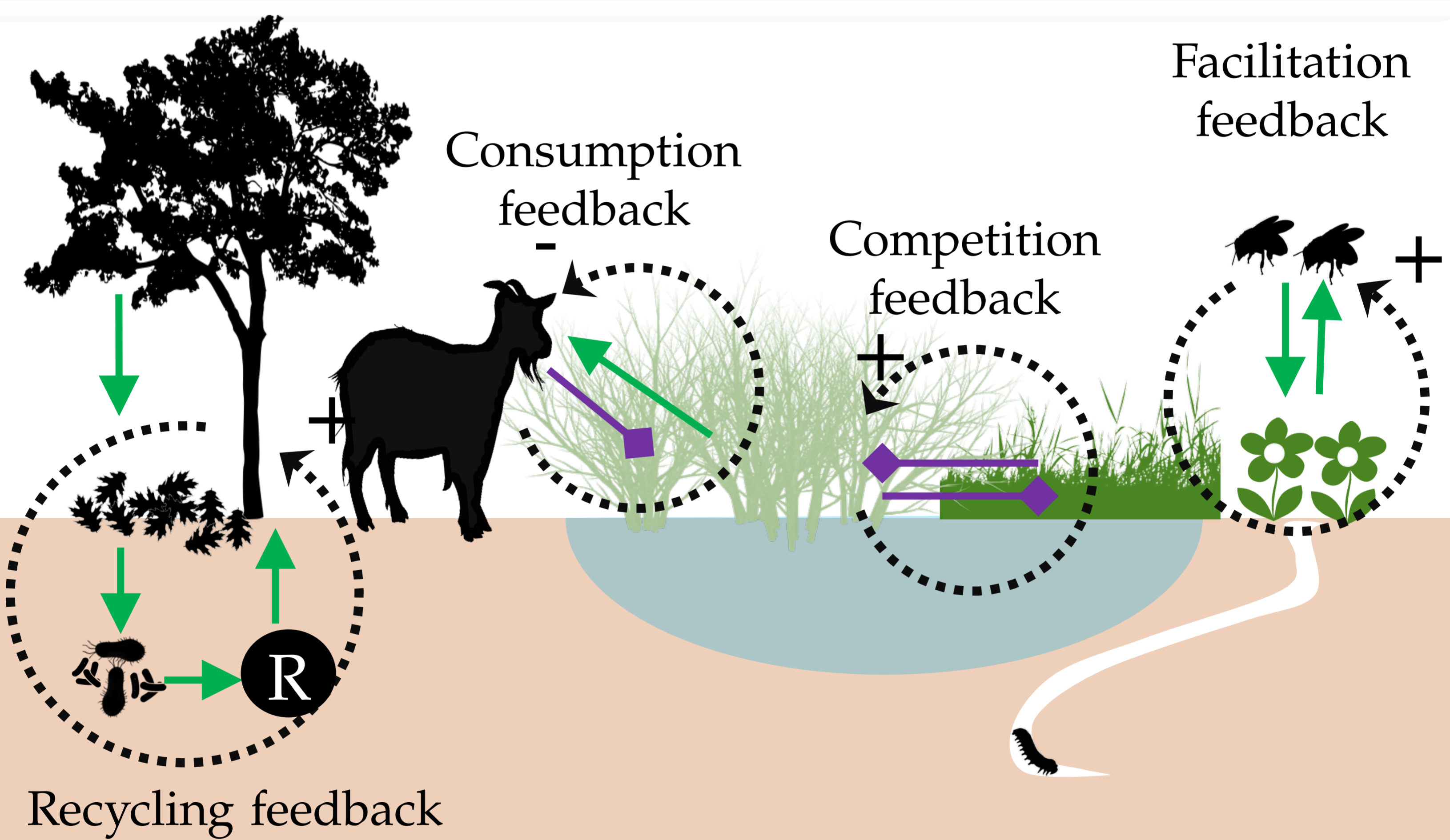
731 **Figure 2: Ecological feedbacks across spatial scales.** At the scale of an individual or a
732 species, interactions with the local abiotic environment generate feedback with resource
733 availability or landforms (biogeomorphic feedbacks), but also more generally through
734 niche construction processes (*e.g.*, habitat creation or modification). At the population
735 or community level, the individual or species may also be involved in demographic or
736 behavioral feedbacks emerging from competition, facilitation, and density-dependent
737 behavior. Last, at the landscape scale, populations, communities and ecosystems exchange
738 individuals, resources and information, which generates spatial feedbacks linking local
739 and landscape scale dynamics. For detailed examples not represented in this figure (*e.g.*,
740 behavioral feedbacks or dispersal spatial feedback) see Table 1. “R” stands for resources.

Feedbacks across spatial scales

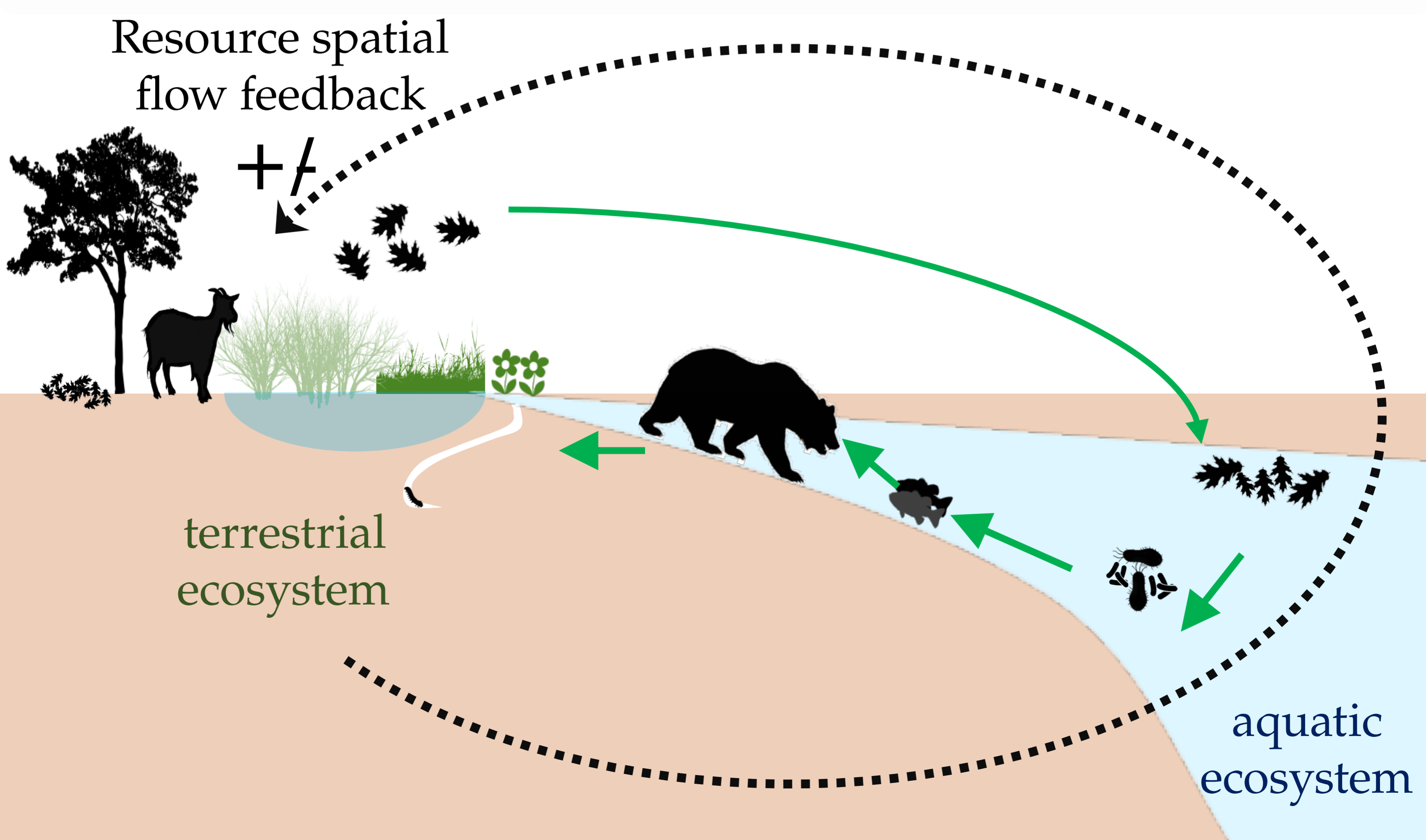
Local organisms-abiotic feedbacks



Community-scale biotic feedbacks



Spatial feedbacks



741 **Figure 3: Mapping the emergent properties from feedback loops.** The position of each
742 emergent property corresponds to a level of organization (from individual to ecosystem)
743 and a spatial scale (local or spatial). See Table 1 for examples for each emergent property
744 and the associated references.

Emergent properties from feedback loops

Spatial scale	Spatial	Condition-dependent dispersal	Patchy-resource distribution
	Local	Source-sink dynamics	Nutrient/carbon source-sink
		Rescue/anti-rescue	Mass effects, species sorting
		Self-organized patterns	Priority effects
			Auto-catalysis
		Abiotic niche construction	Alternative community states
		Allee effect	Biotic niche construction
			Alternative ecosystem states
		Individual	Population
		Community	Ecosystem

Levels of organization

745 **Table 1: Examples of the different types of feedbacks and their associated emergent**
746 **properties**

Table 1: Examples of the different types of feedbacks and their associated emergent properties

<u>Type of feedback</u>	<u>Example</u>	<u>Emergent property</u>
Species-environment feedbacks	Biogeomorphic	Sea-grass meadows stabilize sediments and change hydrodynamic regime [1] Alternative stable states [2] Landscape formation [3-4]
	Involving resources	Plant facilitation increases the availability of nutrients, water, and limits water evaporation, which positively affect their growth and maintenance in drylands [5] Alternative stable states [5-7] Self-organized patterns [6,7]
Population and community-level feedbacks	Behavioural feedbacks	Fish populations in coral reefs display density-dependent foraging rates [8] Alternative stable states [8, 9]
		High predation risk areas generate nutrient hotspot through decomposition of carcasses and nutrient excretion, which fosters plant growth and the patch quality for herbivore consumption [10-13] Patchy distribution of resources [10-13]
	Demographic feedbacks	Positive feedback between plants and pollinators [14] Negative feedbacks between predators and preys [15] Alternative stable states [14] Long transients, cycles [15,16]
Spatial feedbacks	Feedbacks driven by dispersal	Dispersing individuals can experience density-dependent dispersal emerging from interspecific competition [17] or patch-dependent colonization rate [18] (<i>e.g.</i> which depends on the patch quality) creating a feedback between local and spatial dynamics Alternative stable states [19] Spatial heterogeneity and source-sink dynamics [18,20] Positive spatial feedbacks (rescue) or negative spatial feedbacks (anti-rescue) [19]
	Feedbacks driven by resource flows	Seabirds excrete guano that positive affect their growth through a cascading effect on coral reefs and fish [21, 22] Source-sink dynamics of carbon and nutrients [23] Spatial auto-catalytic loop on functioning [24] Emergent colimitation of resources [25]
		Bidirectional exchanges of resources between ecosystems (<i>e.g.</i> , at terrestrial-aquatic ecotone) [25-27] Alternative stable states [28] Diffusion-induced instabilities [29,30]

[1] Maxwell et al., 2017, [2] Carr et al., 2010, [3] Schwarz et al., 2018, [4] Temmink et al., 2022, [5] Rietkerk and van de Koppel 1997, [6] Klausmeir 1999, [7] Kéti et al., 2007, [8] De Roos and Persson 2002, [9] Gil et al., 2020, [10] Bump et al., 2009, [11] Schmitz et al., 2010, [12] Monk & Schmitz 2022, [13] Johnson-Bice et al. 2022, [14] Lever et al., 2014, [15] Barraquand et al., 2017, [16] Hasting et al., 2018, [17] Fronhofer et al., 2015, [18] Pulliam 1988, [19] Harding and McNamara 2002, [20] Hui et al., 2004, [21] Graham et al., 2018, [22] Benkwitt et al., 2021, [23] Gravel et al., 2010, [24] Pichon et al., 2003, [25] Marleau et al., 2015 [26] Bartels et al., 2012, [27] Klemmer et al., 2020, [28] Gounand et al., 2014 [29] Marleau et al., 2010, [30] McCann et al., 2021

747 **Table 2: Examples of how feedback knowledge can help for conservation, restoration**
748 **and mitigation measures across scales**

Table 2: Examples of how feedback knowledge can help for conservation, restoration and mitigation measures across scales

	<u>Description</u>	<u>Consequence</u>
Ecosystem restoration	Increasing the success of dryland restorations through higher seeding (fostering positive feedback from facilitation) and weed control (limiting competition on seedlings)	Higher establishment success of plants [1] More carbon sequestration and services provided by plants [2]
	Maximizing the restoration success of coastal ecosystems by planting propagules in clustered patches and adding clams (interspecific facilitation) to promote positive feedbacks during establishment	Higher survivorship, biomass and expansion of propagules [3-5] Landscape formation, carbon capture and storage [6]
	Managing the demography of harvested species to regulate population demography and maintain high catch or yield	Fisheries assessment and management control can, in theory, maintain fisheries at a maximal sustainable yield either by increasing the catch when population demography is high (<i>i.e.</i> , more negative feedbacks) or by adopting rebuilding plans for overfished species (promoting positive feedbacks) [7]
Conservation	Limiting the spread of invasive species such as island rats or foxes preying on seabirds, or coconut trees replacing birds' nesting habitats to maintain the positive recycling feedback loop from bird guano	Nitrogen depletion in soils that changes the stoichiometry of plants and the community composition [8,9] Disrupts the positive feedback mediated by guano that increased sea-bird demography, fostered island vegetation and coral-reefs' functioning [10,11]
	Protecting the spatial feedbacks between adjacent ecosystems such as at the forest/stream ecotone	Forest subsidies can cascade from organic matter to top consumers in streams (fish), where it fosters fish growth and food-web functioning [12]. Disruption of such coupling may lead to nutrient loading and stream eutrophication [13]
	Protecting specific marine areas by excluding fishing pressure to promote restoration of adjacent areas through spatial rescue (positive spillover via dispersal from the marine protected area)	Designing effective marine protected areas contribute to preserve larval sources, areas connectivity, and fish population abundances [14]
Climatic mitigation	Understanding the feedback loop involved in decomposition process (from organic carbon to carbon release in the atmosphere) to design efficient carbon sequestration measures	Mechanisms such as predation risk [15,16] and priming effect [17-19] can accelerate the loop of carbon decomposition and lead to more carbon loss and in shorter timescales
	Understanding animal-driven recycling loops involving nutrient and carbon cycles to design climatic mitigation measures and species conservation	Trophic downgrading of large animals breaks nutrient and carbon recycling loops and lead to reduced nitrogen and carbon cycles, higher soil-respiration rates, lower ecosystem functioning, and can ultimately switch ecosystems from carbon sink to source [20-24]

[1] Gómez-Aparicio 2009, [2] Shackelford et al., 2021, [3] Zhang et al., 2021 [4] Renzi et al., 2019, [5] Silliman et al., 2015, [6] Temmink et al., 2022, [7] Frank et Oremus, 2023, [8] Croll et al., 2005, [9] Young et al., 2010, [10] Graham et al., 2018, [11] Klemmer et al., 2020, [12] Tanentzap et al., 2014, [13] Harvey et al., 2016, [14] Gaines et al., 2010 , [15] Fontaine et al., 2004, [16] Fontaine et al., 2007, [17] Hawlena et al., 2010, [18] Hawlena et al., 2012, [19] Friggens et al., 2020, [20] Wilmers et al., 2012, [21] Dirzo et al., 2014, [22] Doughty et al., 2016, [23] Leroux et al., 2020, [24] Malhi et al., 2022

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