

1 Time since first naturalization is key to explaining non-native plant
2 invasions on islands

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4 Running title: Naturalization time drives island invasions
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33 **Abstract**

34 Aim

35 Investigating the extent of insular invasions by non-native species (i.e. the number of islands
36 they occupy) is central to island conservation. However, interrelationships among plant life
37 history traits, naturalization histories, and island characteristics in determining island
38 occupancy by non-native plant species are poorly understood. We investigated whether island
39 occupancy by different non-native plant species declines in relation to their year of first
40 naturalization and whether periods of first naturalization differ among growth forms,
41 dispersal modes, and biogeographic origins. Then, we asked if non-native plants that
42 naturalized more recently occur more frequently on islands that are large, less isolated, and
43 close to urban areas. We contrasted trends across growth forms, dispersal modes and
44 biogeographic origins.

45

46 Location

47 264 offshore islands in northern Aotearoa New Zealand

48

49 Taxa

50 Vascular plant species

51

52 Methods

53 We combined field surveys and published data for 767 non-native plant species on the
54 islands. We categorized each species according to its growth form (n = 3), dispersal (n = 4)
55 mode and biogeographic origin (n = 5) and identified its year of first naturalization in
56 Aotearoa New Zealand. We tested our hypotheses using ANCOVA and generalized linear
57 models (GLMs).

58

59 Results

60 There were similar declines in island occupancy in relation to year of first naturalization in
61 Aotearoa New Zealand across all trait and biogeographic origin categories. First
62 naturalization times of herbaceous species, those with unspecialized dispersal modes, and
63 those originating from Eurasia and the Mediterranean basin were disproportionately earlier
64 than other categories. Non-native plants with more recent first naturalization occur more
65 frequently on large islands close to urban areas, but not on less isolated ones. Relationships
66 with island characteristics did not differ among trait and biogeographic origin categories.

67

68 Main Conclusions

69 Overall, time of first naturalization was more important than trait and biogeographic origin
70 categories in explaining non-native plant invasion patterns on islands. Since there were
71 similar relationships between island occupancy and year of first naturalization in Aotearoa
72 New Zealand for all categories, management bodies should focus on non-native plant species
73 of trait and biogeographic origin categories that have naturalized recently (e.g. woody species
74 from other regions within Oceania), and on large islands close to urban areas. Introduction
75 and naturalization histories provide essential context for interpreting the role of plant traits
76 and biogeographic origin in understanding plant invasions on islands.

77

78

79 **Key Words:** *Aotearoa New Zealand, Biological invasions, Biogeographic origin, Dispersal*
80 *modes, Growth forms, Island occupancy, Island biogeography, Naturalization time, Plant*
81 *species*

82

83 Introduction

84 The introduction of non-native species poses considerable threats to islands' biodiversity
85 (Dawson et al., 2017, Kueffer et al., 2010; Lonsdale, 1999; Pyšek & Richardson, 2006;
86 Walentowitz et al., 2023). Given the large numbers of species introduced to islands (Sax &
87 Gaines, 2008, Wilkinson 2004), understanding the extent of insular invasions (i.e. the number
88 of islands occupied) and which island characteristics influence the spread of non-native
89 species is central to island conservation (Lloret et al., 2005, Mologni et al., 2021; Moser et
90 al., 2018). However, the interrelationships between these elements and introduction and
91 naturalization histories and life history traits of non-native species remain poorly understood
92 (Kinlock et al., 2022; Lloret et al., 2005; Schrader et al. 2021a).

93 Rates of non-native species naturalization have changed over time (Sax & Gaines,
94 2008), primarily because of varying introduction rates. The overall number of new species
95 introduced worldwide remained relatively low until the 19th century (Seebens et al., 2017).
96 After that, numbers largely increased, except for the period 1914–1945, marked by two major
97 wars and rather constant introduction numbers (Pyšek et al., 2020; Seebens et al., 2017).
98 After 1945, rates of non-native species introduction increased and are not slowing, correlated
99 with increasing volumes of commerce worldwide (Pyšek et al., 2020; Seebens et al., 2017).
100 Identifying the time of first introduction and naturalization and how they relate to later spread
101 will help forecast future range expansions and guide management focused on early
102 containment or eradication.

103 Species introduced earlier are likely to occupy a larger number of islands than recent
104 introductions (Moyano et al., 2022). However, island colonization is also regulated by island
105 characteristics (MacArthur & Wilson, 1967). Large islands offer a bigger target for dispersing
106 propagules and therefore are more likely to be occupied first (i.e. target island effect,
107 Lomolino, 1990). Similarly, less isolated islands are more easily reached by dispersing
108 propagules and thus will be occupied before more isolated islands (MacArthur & Wilson,

109 1967 but see Moser et al., 2018). Since the introduction and naturalization of non-native plant
110 species often rely on humans, islands that are inhabited or close to human communities are
111 likely to be more readily invaded (Sullivan et al., 2005; Timmins & Williams, 1991,
112 Walentowitz et al., 2023). Interactions between these determinants of island invasions and
113 island occupancy by non-native plants have been rarely investigated.

114 Time of introduction and naturalization, island occupancy, and island characteristics
115 that favour initial colonization of islands are likely to be related to plant life history traits. For
116 instance, a disproportionate number of herbaceous European species were initially introduced
117 in many parts of the world for agricultural purposes and in an attempt to “recreate” a
118 landscape familiar to settlers (Crosby, 2004; Lenzner et al., 2022). Island occupancy might be
119 higher for non-native plant species that disperse more efficiently, either because of shorter
120 life cycles (e.g. herbs) or because of morphological adaptations allowing for long-distance
121 dispersal (Arjona et al., 2018; Mologni et al., 2022; Negoita et al., 2016). Likewise, species
122 originating from biogeographic areas with climates similar to the invaded range are likely to
123 have an advantage when establishing new populations (Richardson & Pyšek, 2012). These
124 traits might also affect how they relate to island characteristics. For instance, long-distance
125 dispersal species might show weaker relationships with isolation (Arjona et al., 2018; Negoita
126 et al., 2016), while high propagule pressure in species with short life cycles might mask a
127 target island effect.

128 Here, we aim to better understand the role of naturalization histories and plant traits in
129 explaining non-native plant invasions on islands. We calculated non-native plant species
130 richness on 264 islands off the northern coast of Aotearoa New Zealand, for a total of 831
131 non-native vascular plant species. We hypothesized that: (1) the earlier a non-native plant
132 species had naturalized in Aotearoa New Zealand the more islands it would occupy, and that
133 the relationships between occupancy and years since naturalization would differ among

134 growth forms, dispersal modes, and biogeographic origins; (2) periods of naturalization in
135 Aotearoa New Zealand would differ within and among distinct growth forms, dispersal
136 modes and biogeographic origins; e.g., we expected that European and Mediterranean
137 herbaceous non-native species would have naturalized consistently earlier; (3) non-native
138 plant species would have colonized larger, less isolated islands that are close to urban areas
139 first.

140

141 **Methods**

142 *Study system*

143 The study system comprises 264 islands off the northern coast of Aotearoa New Zealand's
144 North Island (hereafter the 'mainland'). The study system spans six hundred kilometres (34–
145 38 °S, 172–179 °E, Figure 1). The islands vary in area from 0.00002 to 277.21 km², and
146 range from islands connected to the mainland at low tides to those >50 km from the coast
147 (Mologni et al. 2022). Most are land-bridge islands that were connected to the mainland
148 during the last glacial maximum (Poor Knight Islands excepted, Fleming, 1979). They
149 supported warm temperate rain forests (sensu Grubb et al., 2013) before human settlement
150 (circa 740 years ago, Wilmschurst et al., 2014). Humans cleared, burned and farmed many
151 islands (Bellingham et al., 2010), some of which were abandoned in only the last few
152 decades. Northern Aotearoa New Zealand is the most populous region of the country
153 (Statistics NZ Tatauranga Aotearoa, 2018), and 41 islands are still inhabited within the study
154 system.

155

156 *Data collection*

157 The total number of non-native vascular plant species occupying each of the 264 islands was
158 quantified by collating previously published material and field surveys (see Mologni et al.,

159 2021 for more details). We obtained permits for all unpublished surveys by either the New
160 Zealand's Department of Conservation or Māori and private island owners. We updated all
161 plant species lists for each island to follow consistent vascular plant taxonomy (Allan
162 Herbarium, 2023a, b). For each non-native plant species, we obtained the year of its first
163 naturalization from the New Zealand Plant Conservation Network and Gatehouse (2008). Of
164 the 831 non-native species, only 767 were included in our analyses. Thirty-three were
165 removed due to taxonomic uncertainty, while 31 were not included due to a lack of first-
166 naturalization data.

167 We classified non-native species according to their growth form, dispersal mode and
168 biogeographic origin (Table 1). We evaluated three categories of growth forms – graminoids,
169 forbs, woody species – and excluded lianas and climbers because of their relatively low
170 number of species ($n = 40$). We evaluated four categories of dispersal modes that were based
171 on morphological adaptations to long-distance dispersal: wind-, animal-dispersed,
172 unspecialized and short-distance (Arjona et al., 2018; Burns, 2019; Mologni et al., 2022;
173 Negoita et al., 2016). We excluded water-dispersed species for which the number of species
174 was low ($n = 16$) and a further 18 species for which we lacked data on their dispersal modes.
175 We evaluated five categories of biogeographic origins: Eurasian (excl. Mediterranean basin),
176 Mediterranean (i.e. Mediterranean basin), Tropical, Southern African, Oceanian (excluding
177 Aotearoa New Zealand, and mainly Australian), omitting 87 species from other regions (see
178 Table 1 for more details). Species were categorized by consulting Acta Plantarum (2007),
179 Brandt et al. (2021), and the New Zealand Plant Conservation Network (2023), in addition to
180 our professional expertise.

181 Island characteristics were quantified using ArcGIS 10.5.1 (ESRI, 2011). The size of
182 an island was calculated as the total surface area in km^2 . Isolation was measured as the total
183 amount of water surrounding each island (Carter et al., 2020; Diver, 2008; Negoita et al.,

184 2016). The greater the amount of water, the more isolated an island is. For each island, we
185 created several concentric buffers with radii of 250, 500, 1000, 1500, 2000, 2500 and 3000
186 m. Then, we calculated the amount of water within each radius (%). To quantify the effect of
187 isolation from human communities, we calculated the shortest distance from the nearest
188 urban area (m). We defined as urban built-up areas having a population of 1000 people or
189 more (Statistics NZ Tatauranga Aotearoa, 2018).

190

191 *Statistical analyses*

192 To test whether island occupancy by non-native plants declines in relation to their year of
193 first naturalization and whether relationships differed among growth forms, dispersal modes,
194 and biogeographic origins, we used a generalized linear model (GLM). We set the number of
195 islands occupied by each species as the dependent variable and their year of first
196 naturalization time in Aotearoa New Zealand as the independent variable. Since the
197 dependent variable was right-skewed even after transformation, we modelled data using
198 linear, quasipoisson and negative binomial models for all species. Results were consistent
199 and, for simplicity, we opted for a linear model. We assessed differences in slopes among
200 different trait and biogeographic origin categories with ANCOVA tests. The number of
201 occurrences was log-transformed in all linear models to conform to assumptions of normality.

202 To test whether non-native species show distinct periods of their naturalization among
203 growth forms, dispersal modes, and biogeographic origins, modes were extracted for each
204 trait and biogeographic origin category after binning the year of naturalization by 20-year
205 periods. Additionally, we contrasted the distribution of each category of growth forms,
206 dispersal modes and biogeographic origin using ANOVAs.

207 To test whether non-native plants that naturalized more recently occur more
208 frequently on islands that are large, less isolated and close to urban areas, we calculated the

209 average area, isolation and distance from urban areas of islands that each species occupied
210 and set these as the dependent variables in three separate linear models. The years of first
211 naturalization in Aotearoa New Zealand was used as the independent variable. To conform to
212 assumptions of normality, the average area, isolation and distance from urban areas of
213 occupied islands were respectively log-, arcsine- and square root-transformed. Tests were run
214 for all species first, then ANCOVA tests were used to contrast slopes among different trait
215 and biogeographic origin categories. All analyses were conducted in R (R Core Team, 2023).
216 A post hoc Tukey test and Bonferroni correction were applied for multiple testing at all
217 stages (Hothorn et al., 2008).

218

219 **Results**

220 The earlier a non-native plant species had naturalized in Aotearoa New Zealand, the more
221 islands it occupied (Figure 1 & Table S1). However, the slopes of the negative relationships
222 between the number of islands occupied and year of first naturalization did not differ
223 significantly among different non-native plant growth forms, dispersal modes or
224 biogeographic origins (Table S2).

225 The non-native plant species on the islands had mainly two 20-year bin periods of
226 peak naturalization – 1850 to 1870 and 1970 to 1990 (Figure 2, Table 2 & Table S3). Non-
227 native graminoids and forbs on the islands had disproportionately first naturalized in
228 Aotearoa New Zealand between 1850–1870, whereas woody species had disproportionately
229 first naturalized between 1970–1990 (Figure 2, Table 2 & Table S3). Differences between
230 non-native woody species and other growth forms were significant, as were differences
231 between forbs and graminoids (ANOVA and Tukey tests: Table S4). Among dispersal
232 modes, unspecialized and animal-dispersed species on the islands had disproportionately first
233 naturalized in Aotearoa New Zealand between 1850–1870, whereas short-distance species

234 had disproportionately first naturalized between 1970–1990 (Figure 2, Tables 2 & Table S3).
235 Wind-dispersed species first naturalized disproportionately 1930–1950 (Figure 2, Tables 2 &
236 Table S3). Differences between unspecialized species and other dispersal modes were
237 significant (ANOVA and Tukey tests: Table S4). Non-native Eurasian and Mediterranean
238 plant species on the islands had disproportionately first naturalized during 1850–1870, those
239 from Southern Africa in both 1930–1950 and 1970–1990, and those from Oceania and
240 tropical regions during 1970–1990 (Figure 2, Table 2 & Table S3); there was a significant
241 difference in first naturalization dates between Eurasian and Mediterranean species and those
242 from other biogeographic regions (ANOVA, Tukey tests: Table S4).

243 The average area and distance from urban areas of islands where non-native species
244 occurred respectively increased and declined with year of their first naturalization (Table 3).
245 However, no significant relationship was detected between the average isolation of islands
246 where non-native species occurred and their year of first naturalization (Table 3). There were
247 no significant differences in non-native plants' growth forms, dispersal modes and
248 biogeographic origins for the average area, isolation and distance from urban areas of islands
249 where a species occurred (Table 3 & Table S5).

250

251 **Discussion**

252 The occupancy of non-native plants on islands in northern Aotearoa New Zealand declined in
253 relation to their years of first naturalization but, contrary to our hypothesis, the relationships
254 were similar among growth forms, dispersal modes, and biogeographic origins. There were
255 differences in periods of first naturalization in Aotearoa New Zealand among growth forms,
256 dispersal modes and biogeographic origins of the non-native plants across the islands, as we
257 hypothesized. Consistent with studies in the same system (Mologni et al., 2021), large islands
258 close to urban areas were more frequently colonized first by invading species. However, once

259 the year of first naturalization was accounted for, and in contrast to an earlier study (Mologni
260 et al., 2022), there were no differences among categories of growth forms, dispersal modes
261 and biogeographic origin.

262 Once the year of naturalization was accounted for, it was a better predictor than life
263 history traits in explaining non-native plant invasions on the islands. Date of first
264 naturalization has previously been identified as a crucial factor in explaining non-native plant
265 invasions (Fristoe et al., 2021, Moyano et al., 2022) but the interrelationship between time of
266 naturalization and plant traits as a predictor of invasion success on islands has been little
267 explored (but see Moyano et al., 2022). In our study, non-native woody plant species have
268 colonized proportionately fewer islands but most of them have naturalized more recently than
269 other growth forms. Herbaceous species used for agriculture were introduced and naturalized
270 disproportionately early, whereas woody, ornamental species have naturalized
271 disproportionately recently (Gatehouse, 2008). Among dispersal modes, more unspecialized
272 plant species naturalized early, perhaps reflecting a higher efficiency in exploiting human-
273 related pathways. Among biogeographic origins, more plant species from Eurasia and the
274 Mediterranean naturalized early, a reflection of European settlement (Gatehouse, 2008;
275 Lenzner et al., 2022). In particular, species originating from other regions within Oceania
276 (mostly Australian species) are disproportionately woody (60.3%, see Table S6), and these
277 species might be a particular focus for elimination for managers seeking to reduce invasions
278 on the islands in our study system.

279 If species occupy a similar number of islands irrespective of trait and biogeographic
280 origin categories, we might expect that, in time, more non-native woody species will invade
281 more islands, as the earlier-naturalized herbaceous species have already done (Howell, 2019).
282 Even so, woody species may show longer time lags since they will be generally slower to
283 reach reproductive maturity than herbaceous species (Grime, 2002). In support of this view,

284 there have been few introductions of new non-native conifer species to Aotearoa New
285 Zealand since its 1993 Biosecurity Act significantly reduced introduction rates of non-native
286 species (Seebens et al., 2017). Nonetheless, their naturalization has continued at a linear rate
287 to the present (Howell, 2019), potentially because some species introduced before 1993
288 reached reproductive maturity only recently, although other factors, such as recent co-
289 invasions by mutualists, can also determine more recent invasions by woody plants (Gardner
290 & Early, 1996, Dickie et al., 2010). Similar trends might be expected for other categories that
291 were more consistently introduced in recent times.

292 Previous work in the study system found that graminoids and long-distance dispersal
293 species have higher occupancy of islands than other trait categories (Mologni et al., 2022).
294 However, even though they had higher island occupancy, the relationships between year of
295 first naturalization and island occupancy for graminoids and long-distance dispersal species
296 did not differ, respectively, from those of other life forms and dispersal modes. This implies
297 they require a similar amount of time to other categories for island colonization. Other
298 processes might explain why they occupy more islands in our study system. For instance,
299 they might have been purposely introduced more consistently than other categories.
300 Alternatively, species introduced and naturalized early might negatively impact the
301 establishment of other non-native species (i.e. priority effect, Fraser et al., 2015, Catford et
302 al., 2022). For instance, herbaceous European and Mediterranean non-native species that
303 naturalized early and invaded the islands might reduce the probability of successful
304 establishment by herbaceous non-native species from other biogeographic regions through
305 niche pre-emption (Fukami, 2015). Notably, European sward-forming grasses were
306 introduced and naturalized in the mid-19th century (Allan, 1936) and, in exposed perimeters
307 of islands, these grasses can be effective barriers to colonization by other native and non-
308 native plants (Esler, 1988, Richardson et al., 2018).

309 Non-native plants that naturalized more recently occurred more frequently on large
310 islands, consistent with the target island effect (Lomolino, 1990). However, large islands in
311 the region are also inhabited (Mologni et al., 2021). Whether the more consistent colonization
312 of large islands reflects autonomous or human-mediated dispersal is unclear, both processes
313 though are likely at play. Since they both increase the likelihood of large over small island
314 colonization, island area might then simultaneously encapsulate a “geographic” target island
315 effect (i.e. species colonize larger islands first because they offer a bigger target) and a
316 “human-mediated” target islands effect (i.e. humans inhabit larger islands, disproportionately
317 introducing non-native species). For example, the first invasions of the palm
318 *Archontophoenix cunninghamiana* (native to Australia, first naturalized in 1992) on large,
319 inhabited islands almost certainly derived from gardens in which it was widely planted in the
320 1980s–1990s (Cameron, 2000; Sullivan et al., 2005) but its wider bird dispersal into natural
321 forests (Christianini, 2006, Sheppard, 2013) is likely to have resulted in its wider invasion on
322 those islands and from them to other islands.

323 Non-native plants that naturalized more recently did not occur more frequently on less
324 isolated islands. However, a negative effect of isolation on non-native species richness was
325 previously identified (Mologni et al., 2021). Perhaps this pattern is simply more difficult to
326 capture, and, over time, less isolated islands still accumulate more species than more isolated
327 islands through multiple colonization events. In contrast, non-native plants that naturalized
328 more recently occurred more frequently on islands that are close to urban areas. Since non-
329 native species often use human-related pathways (Pyšek et al., 2020; Sinclair et al., 2020),
330 islands close to urban areas are more likely to be colonized just by proximity (Sullivan et al.,
331 2005; Timmins & Williams, 1991). For non-native species in this study system, isolation
332 might be more strongly associated with humans than with geography.

333 The relationships between island characteristics and time since first naturalization
334 were similar across trait and biogeographic origin categories. This suggests that the same
335 processes regulate the initial colonization of these islands, irrespective of plant life history
336 traits. This is surprising since a growing body of literature has identified differences in island
337 species distributions based on traits (Conti et al., 2022; Mologni et al., 2022; Schrader et al.,
338 2021b; Schrader et al., 2023; Walentowitz et al., 2022). Alternatively, other traits might be
339 more important. For instance, wind-pollinated species are independent of biological carriers
340 for dispersing their pollen, an advantage in newly occupied islands (Cox et al., 1997).
341 Propagule pressure is another trait that likely influences the ability to rapidly colonize more
342 islands (Lockwood et al., 2005). Future studies should assess a broad suite of traits when
343 assessing island invasion.

344 The year of naturalization was used to identify periods of non-native plant species'
345 naturalizations. However, in many cases, long periods of time can elapse between the
346 introduction, naturalization and identification of newly naturalized species. This is
347 particularly important for earlier introductions, more likely to go unnoticed due to a lack of
348 awareness of biological invasions. Conversely, we can expect current monitoring activities to
349 be more effective in identifying new non-native species (Pyšek et al., 2020). Unfortunately,
350 accounting for this form of error will be challenging.

351 Since similar relationships between island occupancy and year of first naturalization
352 were identified for all trait and biogeographic origin categories despite different periods of
353 naturalization, management bodies should focus on species of categories that naturalized
354 more frequently in recent times, such as woody species (Howell, 2019) originating from other
355 regions within Oceania and short-distance dispersal species, and on large islands close to
356 urban areas. Additionally, our work builds on and aligns with a growing body of literature
357 integrating traits in island biogeography (Ottaviani et al., 2020; Schrader et al., 2023;

358 Whittaker et al., 2014), and shows that introduction and naturalization histories provide
359 essential context for interpreting the role of plant traits and biogeographic origin in
360 understanding plant invasions on islands.

361

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562 **Data availability statement**

563 Data are available as supplementary material and from the Manaaki Whenua data repository
564 at <https://doi.org/10.7931/ndkt-zw49>. Island locations (longitude and latitude) are not
565 publicly available due to private ownership and issues of data sovereignty of concern to
566 Māori.

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577 **Table 1** Description and the number of species for growth form, dispersal mode and
 578 biogeographic origin categories across 264 Aotearoa New Zealand offshore islands.

Category	Description	Number of species
Growth forms		767
Graminoids	Grasses, sedges and rushes	120
Forbs	Herbaceous, non-graminoid	466
Woody species	Trees and shrubs	141
Climbers and lianas	Herbaceous or woody climbers	40
Dispersal modes		749
Wind-dispersed	Plumes, wings, dust diaspores (anemochory)	139
Water-dispersed	Buoyant propagules, e.g. corky tissues, air pockets (hydrochory)	16
Animal-dispersed	Fleshy fruits or adhesive barbs (endo and epizoochory)	155
Unspecialized	No evident or prevalent morphological adaptations (unspecialized)	281
Short-distance	Morphological adaptations for short-distance dispersal only (ballochory, myrmecory)	158
Biogeographic origin		734
Eurasian	Europe (excl. Mediterranean basin), Asia, North Africa and the Arabian Peninsula	302
Cosmopolitan	Cosmopolitan, subcosmopolitan and temperate areas of the northern hemisphere	36
Other American	All America (North, Central, South) or Temperate South America	15
Mediterranean	Mediterranean basin	94
North American	Nearctic	31
Oceanian	Australia, Pacific	61
Southern African	South Africa, Eswatini, Lesotho, and Namibia	66

Tropical	Tropical Africa, Asia, Central and South America, Madagascar and Pantropical	124
Macaronesian	Macaronesia	5

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581 **Table 2** Peak of naturalization (mode) and the number of species introduced per 20-
582 year bin period for growth form, dispersal mode and biogeographic origin categories across
583 264 Aotearoa New Zealand offshore islands.

Category	Peak of naturalization (Mode)	Number of species introduced
Growth forms		
Graminoids	1850–1870	34
Forbs	1850–1870	129
Woody species	1970–1990	49
Dispersal modes		
Wind-dispersed	1930–1950	27
Animal-dispersed	1850–1870	37
Unspecialized	1850–1870	89
Short-distance	1970–1990	34
Biogeographic origin		
Eurasian	1850–1870	100
Mediterranean	1850–1870	27
Oceanian	1970–1990	20
Southern African	1930–1950, 1970–1990	15, 15
Tropical	1970–1990	27

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585 **Table 3** Top section: separate linear relationships between the average area, isolation
586 and distance from urban areas (average urban) of occupied islands and year of first

587 naturalization across all species. Entries are model, estimate, t- and p-value. Bottom section:
 588 Interaction terms extracted from ANCOVA models between the average area, isolation and
 589 distance from urban areas (average urban) of occupied islands and time since first
 590 naturalization by growth forms, dispersal modes and biogeographic origin. Entries are model,
 591 sum of squares, F- and p-value. P-values are in bold when significant (<0.05).

Overall models			
Model	Estimate	t-value	p-value
log (average area) ~ year of first naturalization	0.00 ± 0.00	3.68	<0.001
arcsine (average isolation) ~ year of first naturalization	-0.00 ± 0.00	0.44	0.662
square root (average urban) ~ year of first naturalization	-0.25 ± 0.03	-7.60	<0.001
Interaction terms			
Model	Sum of squares	F-value	p-value
log (average area) ~ year of first naturalization * growth forms	3.60	1.72	0.180
log (average area) ~ year of first naturalization * dispersal modes	8.30	2.55	0.054
log (average area) ~ year of first naturalization * biogeographic origin	3.40	0.78	0.536
arcsine (average isolation) ~ year of first naturalization * growth forms	0.04	0.67	0.513
arcsine (average isolation) ~ year of first naturalization * dispersal modes	0.15	1.60	0.188
arcsine (average isolation) ~ year of first naturalization * biogeographic origin	0.15	1.23	0.295
square root (average urban) ~ year of first naturalization * growth	3346.00	0.90	0.406
square root (average urban) ~ year of first naturalization * dispersal	9098.00	1.54	0.203
square root (average urban) ~ year of first naturalization * biogeographic origin	14467.00	1.93	0.105

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595 **Figure Captions**

596 **[single column] Figure 1** Relationships between the number of islands occupied by a
597 species and their year of first naturalization for growth form (top), dispersal mode (centre)
598 and biogeographic origin (bottom) categories across 264 Aotearoa New Zealand offshore
599 islands. No statistical differences were identified in contrasting categories.

600 **[double column] Figure 2** Frequency plots displaying the number of species naturalized
601 per 20-year bin period for growth form (top), dispersal mode (centre) and biogeographic
602 origin (bottom) categories across 264 Aotearoa New Zealand offshore islands. Numbers
603 illustrate the number of species introduced at the mode (naturalization peak) by category.
604 Two modes were identified for Southern African species.

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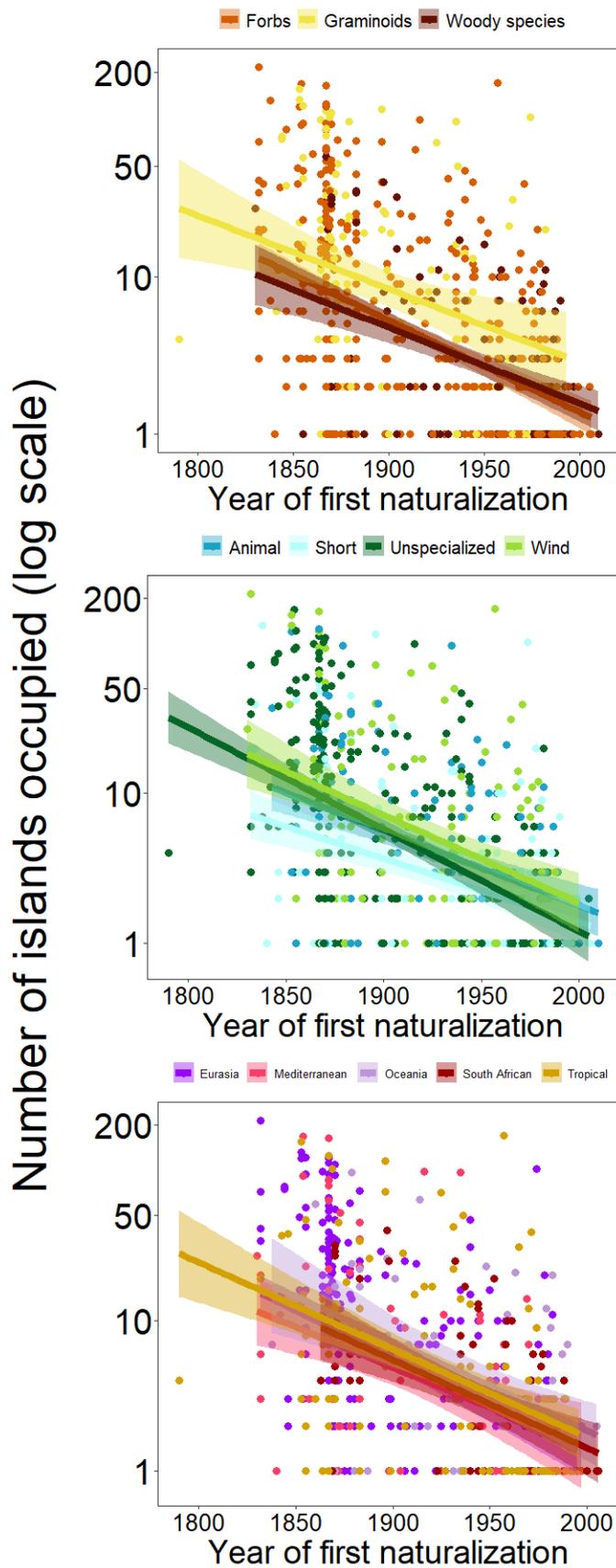
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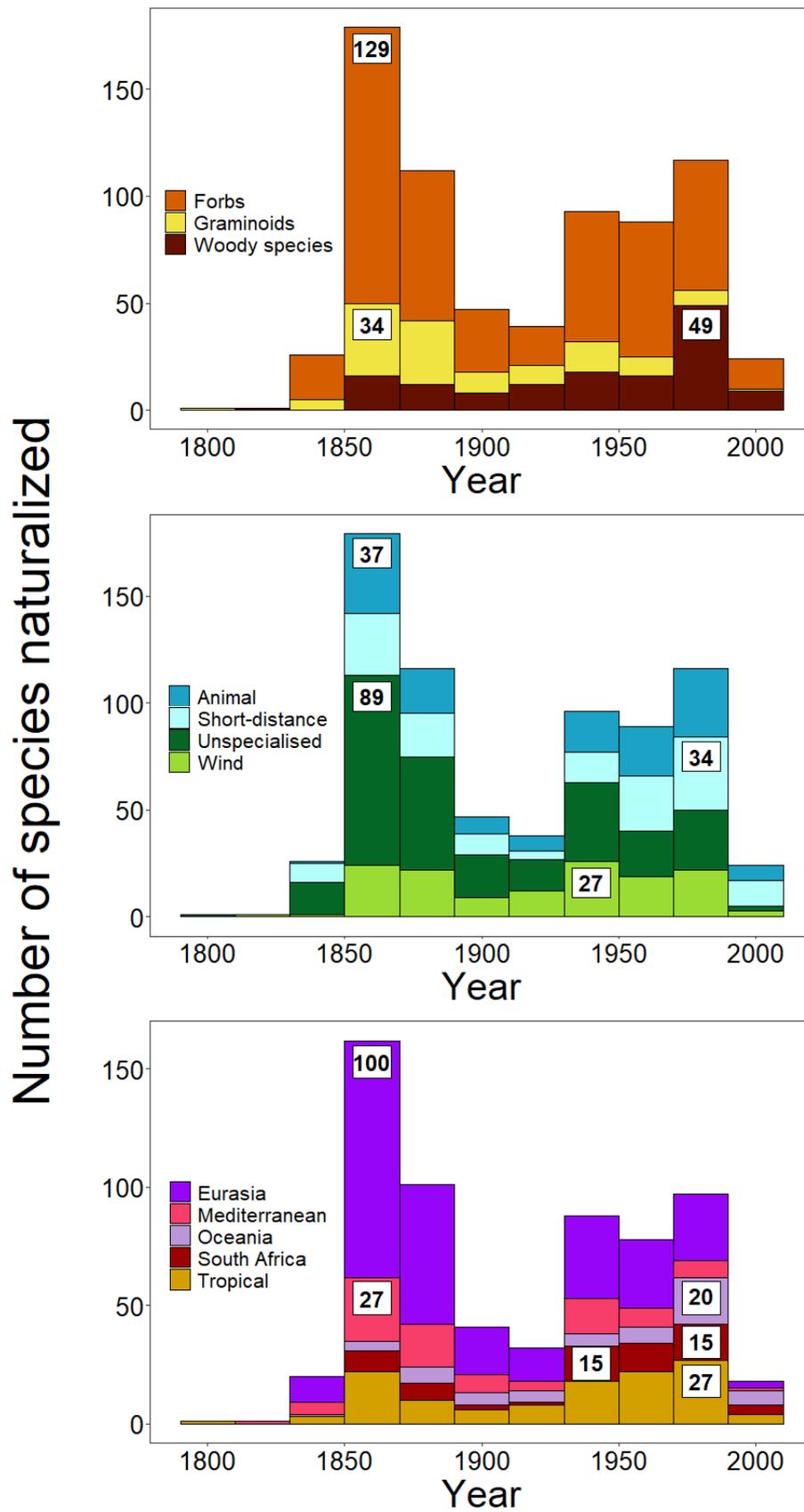
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615 **Figure 1**



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617 **Figure 2**



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