

1 Time since first naturalization is key to explaining non-native plant
2 invasions on islands

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4 Running title: Naturalization time drives island invasions
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33 **Abstract**

34 Aim

35 Investigating the extent of insular invasions by non-native species (i.e. the number of islands
36 they occupy) is central to island conservation. However, interrelationships among plant life
37 history traits, naturalization histories, and island characteristics in determining island
38 occupancy by non-native plant species are poorly understood. We investigated whether island
39 occupancy by different non-native plant species varies according to their year of first
40 naturalization and whether periods of first naturalization differ across growth forms, dispersal
41 modes, and biogeographic origins. Then, we asked if non-native plants that naturalized more
42 recently occur more frequently on islands that are large, less isolated and close to urban areas.
43 We contrasted trends across growth forms, dispersal modes and biogeographic origins.

44

45 Location

46 264 offshore islands in northern Aotearoa New Zealand

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48 Taxa

49 Vascular plant species

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51 Methods

52 We combined field surveys and published data for 767 non-native plant species. We
53 categorized each species according to its growth form ($n = 3$), dispersal ($n = 4$) mode and
54 biogeographic origin ($n = 5$) and identified its year of first naturalization in Aotearoa New
55 Zealand. We tested our hypotheses using ANCOVA and generalized linear models (GLMs).

56

57 Results

58 There were similar relationships between island occupancy and year of first naturalization in
59 Aotearoa New Zealand across all trait and biogeographic origin categories but there were
60 distinctly different times of first naturalization within each category. First naturalization times
61 of herbaceous species, those with unspecialized dispersal modes, and those originating from
62 Eurasia and the Mediterranean basin were disproportionately earlier than other categories.
63 Non-native plants with more recent first naturalization occur more frequently on large islands
64 close to urban areas, but not on less isolated ones. Relationships with island characteristics
65 did not differ among trait and biogeographic origin categories.

66

67 Main Conclusions

68 Overall, time of first naturalization is more important than trait and biogeographic origin
69 categories in explaining non-native plant invasion patterns on islands. Since there were
70 similar relationships between island occupancy and year of first naturalization in Aotearoa
71 New Zealand for all categories, management bodies should focus on species of trait and
72 biogeographic origin categories that naturalized more frequently in recent times (e.g. woody
73 species from other regions within Oceania), and on large islands close to urban areas.
74 Introduction and naturalization histories provide essential context for interpreting the role of
75 plant traits and biogeographic origin in understanding plant invasions on islands.

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78 **Key Words:** *Aotearoa New Zealand, Biogeographic origin, Dispersal modes, Growth forms,*
79 *Island occupancy, Island biogeography, Naturalization time, Plant species*

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87 **Introduction**

88 The introduction of non-native species poses considerable threats to islands' biodiversity
89 (Dawson et al., 2017, Kueffer et al., 2010; Lonsdale, 1999; Pyšek & Richardson, 2006;
90 Walentowitz et al., 2023). Given the large numbers of species introduced to islands (Sax &
91 Gaines, 2008, Wilkinson 2004), understanding the extent of insular invasions (i.e. the number
92 of islands occupied) and which island characteristics influence the spread of non-native
93 species is central to island conservation (Lloret et al., 2005, Mologni et al., 2021; Moser et
94 al., 2018). However, the interrelationships between these elements and introduction and
95 naturalization histories and life history traits of non-native species remain poorly understood
96 (Kinlock et al., 2022; Lloret et al., 2005; Schrader et al. 2021a).

97 Rates of non-native species naturalization have changed over time (Sax & Gaines,
98 2008), primarily because of varying introduction rates. The overall number of new species
99 introduced worldwide remained relatively low until the 19th century (Seebens et al., 2017).

100 After that, numbers largely increased, except for the period 1914–1945, marked by two major
101 wars and rather constant introduction numbers (Pyšek et al., 2020; Seebens et al., 2017).

102 After 1945, rates of non-native species introduction increased and are not slowing, correlated
103 with increasing volumes of commerce worldwide (Pyšek et al., 2020; Seebens et al., 2017).

104 Identifying the time of first introduction and naturalization and how they relate to later spread
105 will help forecast future range expansions and guide management focused on early
106 containment or eradication.

107 Species introduced earlier are likely to occupy a larger number of islands than recent
108 introductions (Moyano et al., 2022). However, island colonization is also regulated by island
109 characteristics (MacArthur & Wilson, 1967). Large islands offer a bigger target for dispersing
110 propagules and therefore are more likely to be occupied first (i.e. target island effect,
111 Lomolino, 1990). Similarly, less isolated islands are more easily reached by dispersing
112 propagules and thus will be occupied before more isolated islands (MacArthur & Wilson,
113 1967 but see Moser et al., 2018). Since the introduction and naturalization of non-native plant
114 species often rely on humans, islands that are inhabited or close to human communities are
115 likely to be more readily invaded (Sullivan et al., 2005; Timmins & Williams, 1991,
116 Walentowitz et al., 2023). Interactions between these determinants of island invasions and
117 island occupancy by non-native plants have been rarely investigated.

118 Time of introduction and naturalization, island occupancy, and island characteristics
119 that favour initial colonization of islands are likely to be related to plant life history traits. For
120 instance, a disproportionate number of herbaceous European species were initially introduced
121 in many parts of the world for agricultural purposes and in an attempt to “recreate” a
122 landscape familiar to settlers (Crosby, 2004; Lenzner et al., 2022). Island occupancy might be
123 higher for non-native plant species that disperse more efficiently, either because of shorter
124 life cycles (e.g. herbs) or because of morphological adaptations allowing for long-distance
125 dispersal (Arjona et al., 2018; Mologni et al., 2022; Negoita et al., 2016). Likewise, species
126 originating from biogeographic areas with climates similar to the invaded range are likely to
127 have an advantage when establishing new populations (Richardson & Pyšek, 2012). These
128 traits might also affect how they relate to island characteristics. For instance, long-distance
129 dispersal species might show weaker relationships with isolation (Arjona et al., 2018; Negoita
130 et al., 2016), while high propagule pressure in species with short life cycles might mask a
131 target island effect.

132 Here, we aim to better understand the role of naturalization histories and plant traits in
133 explaining non-native plant invasions on islands. We calculated non-native plant species
134 richness on 264 islands off the northern coast of Aotearoa New Zealand, for a total of 831
135 non-native vascular plant species. We asked: (1) Does island occupancy by non-native plant
136 species vary according to year of first naturalization in Aotearoa New Zealand and, if so, is
137 the relationship consistent across different growth forms, dispersal modes, and biogeographic
138 origins? (2) Do periods of naturalization in Aotearoa New Zealand differ within and among
139 different growth forms, dispersal modes and biogeographic origins? (3) If non-native plant
140 species colonize larger, less isolated islands that are close to urban areas first, does the
141 average (a) area, (b) isolation and (c) distance from urban areas of the islands where a species
142 occurs respectively increase, decline and decline with its year of first naturalization? If so, do
143 the relationships differ across growth forms, dispersal modes and biogeographic origins?

144

145 **Methods**

146 *Study system*

147 The study system comprises 264 islands off the northern coast of Aotearoa New Zealand's
148 North Island (hereafter the 'mainland'). The study system spans six hundred kilometres (34–
149 38 °S, 172–179 °E, Figure 1). The islands vary in area from 0.00002 to 277.21 km², and
150 range from islands connected to the mainland at low tides to those >50 km from the coast
151 (Mologni et al. 2022). Most are land-bridge islands that were connected to the mainland
152 during the last glacial maximum (Poor Knight Islands excepted, Fleming, 1979). They
153 supported warm temperate rain forests (sensu Grubb et al., 2013) before human settlement
154 (circa 740 years ago, Wilmshurst et al., 2014). Humans cleared, burned and farmed many
155 islands (Bellingham et al., 2010), some of which were abandoned in only the last few
156 decades. Northern Aotearoa New Zealand is the most populous region of the country

157 (Statistics NZ Tatauranga Aotearoa, 2018), and 41 islands are still inhabited within the study
158 system.

159

160 *Data collection*

161 The total number of non-native vascular plant species occupying each of the 264 islands was
162 quantified by collating previously published material and field surveys (see Mologni et al.,
163 2021 for more details). We obtained permits for all unpublished surveys by either the New
164 Zealand's Department of Conservation or Māori and private island owners. For each species,
165 we obtained the year of their first naturalization from the New Zealand Plant Conservation
166 Network and Gatehouse (2008). Of the 831 non-native species, only 767 were included in our
167 analyses. Thirty-three were removed due to taxonomic uncertainty, while 31 were not
168 included due to a lack of first-naturalization data.

169 We classified non-native species according to their growth form, dispersal mode and
170 biogeographic origin (Table 1). Four categories of growth forms were identified: graminoids,
171 forbs, woody species, and lianas and climbers. Five categories of dispersal modes were
172 defined based on morphological adaptations to long-distance dispersal: wind-, water-, animal-
173 dispersed, unspecialized and short-distance (Arjona et al., 2018; Burns, 2019; Mologni et al.,
174 2022; Negoita et al., 2016). Nine categories of biogeographic origin were delineated:
175 Cosmopolitan, Eurasian (excl. Mediterranean basin), Mediterranean (i.e. Mediterranean
176 basin), Tropical, Southern African, Oceanian (excluding Aotearoa New Zealand, and mainly
177 Australian), North American (i.e. species restricted to North America), Other American (i.e.
178 species ranging across all America or temperate South America) and Macaronesian (see
179 Table 1 for more details). All species could be categorized based on their growth forms;
180 however, 18 and 33 species were respectively excluded from dispersal mode and
181 biogeographic origin analyses due to a lack of data. Of these categories, six were removed

182 due to the small sample size: lianas and climbers (number of species = 40), water-dispersed
183 (n = 16), cosmopolitan (n = 36), North American (n = 31), Other American (n = 15),
184 Macaronesian (n = 5). In total, we used 3 categories for growth forms, 4 for dispersal modes
185 and 6 for biogeographic origins (Table 1). Species were categorized by consulting Acta
186 Plantarum (2007), Brandt et al. (2021), and the New Zealand Plant Conservation Network
187 (2023), in addition to our professional expertise.

188 Island characteristics were quantified using ArcGIS 10.5.1 (ESRI, 2011). The size of
189 an island was calculated as the total surface area in km². Isolation was measured as the total
190 amount of water surrounding each island (Carter et al., 2020; Diver, 2008; Negoita et al.,
191 2016). The greater the amount of water, the more isolated an island is. For each island, we
192 created several concentric buffers with radii of 250, 500, 1000, 1500, 2000, 2500 and 3000
193 m. Then, we calculated the amount of water within each radius (%). To quantify the effect of
194 isolation from human communities, we calculated the shortest distance from the nearest
195 urban area (m). We defined as urban built-up areas having a population of 1000 people or
196 more (Statistics NZ Tatauranga Aotearoa, 2018).

197

198 *Statistical analyses*

199 To test whether island occupancy varies according to year of first naturalization and across
200 growth forms, dispersal modes, and biogeographic origins, we used a generalized linear
201 model (GLM). We set the number of islands occupied by each species as the dependent
202 variable and their year of first naturalization time in Aotearoa New Zealand as the
203 independent variable. Since the dependent variable was right-skewed even after
204 transformation, we modelled data using linear, quasipoisson and negative binomial models
205 for all species. Results were consistent and, for simplicity, we opted for a linear model. We
206 assessed differences in slopes among different trait and biogeographic origin categories with

207 ANCOVA tests. The number of occurrences was log-transformed in all linear models to
208 conform to assumptions of normality.

209 To test whether non-native species show distinct periods of their naturalization across
210 growth forms, dispersal modes, and biogeographic origins, modes were extracted for each
211 trait and biogeographic origin category after binning the year of naturalization by 20-year
212 periods. Additionally, we contrasted the distribution of each category of growth forms,
213 dispersal modes and biogeographic origin using ANOVAs.

214 To test whether non-native plants that naturalized more recently occur more
215 frequently on islands that are large, less isolated and close to urban areas, we set the average
216 area, isolation and distance from urban areas of occupied islands by each species as the
217 dependent variables in three separate linear models. The years of first naturalization in
218 Aotearoa New Zealand was used as the independent variable. To conform to assumptions of
219 normality, the average area, isolation and distance from urban areas of occupied islands were
220 respectively log-, arcsine- and square root-transformed. Tests were run for all species first,
221 then ANCOVA tests were used to contrast slopes among different trait and biogeographic
222 origin categories. All analyses were conducted in R (R Core Team, 2023). A post hoc Tukey
223 test and Bonferroni correction were applied for multiple testing at all stages (Hothorn et al.,
224 2008).

225

226 **Results**

227 The earlier a non-native plant species had naturalized in Aotearoa New Zealand, the more
228 islands it occupies (Figure 1 & Table S1). However, the slopes of relationships between the
229 number of islands occupied and year of first naturalization did not differ significantly among
230 different non-native plant growth forms, dispersal modes or biogeographic origins (Table
231 S2).

232 The non-native plant species on the islands had mainly two 20-year bin periods of
233 peak naturalization – 1850 to 1870 and 1970 to 1990 (Figure 2, Table 2 & Table S3). Non-
234 native graminoids and forbs on the islands had disproportionately first naturalized in
235 Aotearoa New Zealand between 1850–1870, whereas woody species had disproportionately
236 first naturalized between 1970–1990 (Figure 2, Table 2 & Table S3). Differences between
237 non-native woody species and other growth forms were significant, as were differences
238 between forbs and graminoids (ANOVA and Tukey tests: Table S4). Among dispersal
239 modes, unspecialized and animal-dispersed species on the islands had disproportionately first
240 naturalized in Aotearoa New Zealand between 1850–1870, whereas short-distance species
241 had disproportionately first naturalized between 1970–1990 (Figure 2, Tables 2 & Table S3).
242 Wind-dispersed species first naturalized disproportionately 1930–1950 (Figure 2, Tables 2 &
243 Table S3). Differences between unspecialized species and other dispersal modes were
244 significant (ANOVA and Tukey tests: Table S4). Non-native Eurasian and Mediterranean
245 plant species on the islands had disproportionately first naturalized during 1850–1870, those
246 from Southern Africa in both 1930–1950 and 1970–1990, and those from Oceania and
247 tropical regions during 1970–1990 (Figure 2, Table 2 & Table S3); there was a significant
248 difference in first naturalization dates between Eurasian and Mediterranean species and those
249 from other biogeographic regions (ANOVA, Tukey tests: Table S4).

250 The average area and distance from urban areas of islands where a species occurred
251 respectively increased and declined with year of first naturalization (Table 3). No significant
252 relationship was detected between the average isolation of islands where a species occurred
253 and year of first naturalization (Table 3). There were no significant differences in non-native
254 plants' growth forms, dispersal modes and biogeographic origins for the average area,
255 isolation and distance from urban areas of islands where a species occurred (Table 3 & Table
256 S5).

257

258 **Discussion**

259 We found similar negative relationships between island occupancy and year of first
260 naturalization across growth forms, dispersal modes, and biogeographic origins of non-native
261 plants on islands in northern Aotearoa New Zealand. We also identified differences in periods
262 of first naturalization in Aotearoa New Zealand among growth forms, dispersal modes and
263 biogeographic origins. Consistent with studies in the same system (Mologni et al., 2021),
264 large islands close to urban areas were more frequently colonized first by invading species.
265 However, once the year of first naturalization was accounted for, and in contrast to an earlier
266 study (Mologni et al., 2022), there were no differences across categories of growth forms,
267 dispersal modes and biogeographic origin.

268 Once the year of naturalization was accounted for, it was a better predictor than life
269 history traits in explaining non-native plant invasions on the islands. Date of first
270 naturalization has previously been identified as a crucial factor in explaining non-native plant
271 invasions (Fristoe et al., 2021, Moyano et al., 2022) but the interrelationship between time of
272 naturalization and plant traits as a predictor of invasion success on islands has been little
273 explored (but see Moyano et al., 2022). In our study, non-native woody plant species have
274 colonized proportionately fewer islands but most of them have naturalized more recently than
275 other growth forms. Herbaceous species used for agriculture were introduced and naturalized
276 disproportionately early, whereas woody, ornamental species have naturalized
277 disproportionately recently (Gatehouse, 2008). Among dispersal modes, more unspecialized
278 plant species naturalized early, perhaps reflecting a higher efficiency in exploiting human-
279 related pathways. Among biogeographic origins, more plant species from Eurasia and the
280 Mediterranean naturalized early, a reflection of European settlement (Gatehouse, 2008;
281 Lenzner et al., 2022). In particular, species originating from other regions within Oceania

282 (mostly Australian species) are disproportionately woody (60.3%, see Table S6), and these
283 species might be a particular focus for elimination for managers seeking to reduce invasions
284 on the islands in our study system.

285 If species occupy a similar number of islands irrespective of trait and biogeographic
286 origin categories, we might expect that, in time, more non-native woody species will invade
287 more islands, as the earlier-naturalized herbaceous species have already done (Howell, 2019).
288 Even so, woody species may show longer time lags since they will be generally slower to
289 reach reproductive maturity than herbaceous species (Grime, 2002). In support of this view,
290 there have been few introductions of new non-native conifer species to Aotearoa New
291 Zealand since its 1993 Biosecurity Act significantly reduced introduction rates of non-native
292 species (Seebens et al., 2017). Nonetheless, their naturalization has continued at a linear rate
293 to the present (Howell, 2019), potentially because some species introduced before 1993
294 reached reproductive maturity only recently, although other factors, such as recent co-
295 invasions by mutualists, can also determine more recent invasions by woody plants (Gardner
296 & Early, 1996, Dickie et al., 2010). Similar trends might be expected for other categories that
297 were more consistently introduced in recent times.

298 Previous work in the study system found that graminoids and long-distance dispersal
299 species have higher occupancy of islands than other trait categories (Mologni et al., 2022).
300 However, despite their higher occupancy of islands, graminoids and long-distance dispersal
301 species are no different in their relationship between year of first naturalization and island
302 occupancy than other life forms and dispersal modes. This implies they require a similar
303 amount of time to other categories for island colonization. Other processes might explain why
304 they occupy more islands in our study system. For instance, they might have been purposely
305 introduced more consistently than other categories. Alternatively, species introduced and
306 naturalized early might negatively impact the establishment of other non-native species (i.e.

307 priority effect, Fraser et al., 2015, Catford et al., 2022). For instance, herbaceous European
308 and Mediterranean non-native species that naturalized early and invaded the islands might
309 reduce the probability of successful establishment by herbaceous non-native species from
310 other biogeographic regions through niche pre-emption (Fukami, 2015). Notably, European
311 sward-forming grasses were introduced and naturalized in the mid-19th century (Allan, 1936)
312 and, in exposed perimeters of islands, these grasses can be effective barriers to colonization
313 by other native and non-native plants (Esler, 1988, Richardson et al., 2018).

314 Non-native plants that naturalized more recently occurred more frequently on large
315 islands, consistently with the target island effect (Lomolino, 1990). However, large islands in
316 the region are also inhabited (Mologni et al., 2021). Whether the more consistent colonization
317 of large islands reflects autonomous or human-mediated dispersal is unclear, both processes
318 though are likely at play. Since they both increase the likelihood of large over small island
319 colonization, island area might then simultaneously encapsulate a “geographic” target island
320 effect (i.e. species colonize larger islands first because they offer a bigger target) and a
321 “human-mediated” target islands effect (i.e. human inhabit larger islands, disproportionately
322 introducing non-native species). For example, the first invasions of the palm
323 *Archontophoenix cunninghamiana* (native to Australia, first naturalized in 1992) on large,
324 inhabited islands almost certainly derived from gardens in which it was widely planted in the
325 1980s–1990s (Cameron, 2000; Sullivan et al., 2005) but its wider bird dispersal into natural
326 forests (Christianini, 2006, Sheppard, 2013) is likely to have resulted in its wider invasion on
327 those islands and from them to other islands.

328 Non-native plants that naturalized more recently did not occur more frequently on less
329 isolated islands. However, a negative effect of isolation on non-native species richness was
330 previously identified (Mologni et al., 2021). Perhaps this pattern is simply more difficult to
331 capture, and, over time, less isolated islands still accumulate more species than more isolated

332 islands through multiple colonization events. In contrast, non-native plants that naturalized
333 more recently occurred more frequently on islands that are close to urban areas. Since non-
334 native species often use human-related pathways (Pyšek et al., 2020; Sinclair et al., 2020),
335 islands close to urban areas are more likely to be colonized just by proximity (Sullivan et al.,
336 2005; Timmins & Williams, 1991). For non-native species in this study system, isolation
337 might be more strongly associated with humans than with geography.

338 The relationships between island characteristics and time since first naturalization
339 were similar across trait and biogeographic origin categories. This suggests that the same
340 processes regulate the initial colonization of these islands, irrespective of plant life history
341 traits. This is surprising since a growing body of literature has identified differences in island
342 species distributions based on traits (Conti et al., 2022; Mologni et al., 2022; Schrader et al.,
343 2021b; Schrader et al., 2023; Walentowitz et al., 2022). Alternatively, other traits might be
344 more important. For instance, wind-pollinated species are independent of biological carriers
345 for dispersing their pollen, an advantage in newly occupied islands (Cox et al., 1997).
346 Propagule pressure is another trait that likely influences the ability to rapidly colonize more
347 islands (Lockwood et al., 2005). Future studies should assess a broad suite of traits when
348 assessing island invasion.

349 The year of naturalization was used to identify periods of non-native plant species'
350 naturalizations. However, in many cases, long periods of time can elapse between the
351 introduction, naturalization and identification of newly naturalized species. This is
352 particularly important for earlier introductions, more likely to go unnoticed due to a lack of
353 awareness of biological invasions. Conversely, we can expect current monitoring activities to
354 be more effective in identifying new non-native species (Pyšek et al., 2020). Unfortunately,
355 accounting for this form of error will be challenging.

356 Since similar relationships between island occupancy and year of first naturalization
357 were identified for all trait and biogeographic origin categories despite different periods of
358 naturalization, management bodies should focus on species of categories that naturalized
359 more frequently in recent times, such as woody species (Howell, 2019) originating from other
360 regions within Oceania and short-distance dispersal species, and on large islands close to
361 urban areas. Additionally, this work builds on and aligns with a growing body of literature
362 implementing traits in island biogeography (Ottaviani et al., 2020; Schrader et al., 2023;
363 Whittaker et al., 2014), and shows that introduction and naturalization histories provide
364 essential context for interpreting the role of plant traits and biogeographic origin in
365 understanding plant invasions on islands.

366

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561 **Data availability statement**

562 Data are available as supplementary material. Island locations (longitude and latitude) are not
563 publicly available due to private ownership and issues of data sovereignty of concern to
564 Māori.

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575 **Table 1** Description and the number of species for growth form, dispersal mode and
 576 biogeographic origin categories across 264 Aotearoa New Zealand offshore islands.

Category	Description	Number of species
Growth forms		767
Graminoids	Grasses, sedges and rushes	120
Forbs	Herbaceous, non-graminoid	466
Woody species	Trees and shrubs	141
Climbers and lianas	Herbaceous or woody climbers	40
Dispersal modes		749
Wind-dispersed	Plumes, wings, dust diaspores (anemochory)	139
Water-dispersed	Buoyant propagules, e.g. corky tissues, air pockets (hydrochory)	16
Animal-dispersed	Fleshy fruits or adhesive barbs (endo and epizoochory)	155
Unspecialized	No evident or prevalent morphological adaptations (unspecialized)	281
Short-distance	Morphological adaptations for short-distance dispersal only (ballochory, myrmecory)	158
Biogeographic origin		734
Eurasian	Europe (excl. Mediterranean basin), Asia, North Africa and the Arabian Peninsula	302
Cosmopolitan	Cosmopolitan, subcosmopolitan and temperate areas of the northern hemisphere	36
Other American	All America (North, Central, South) or Temperate South America	15
Mediterranean	Mediterranean basin	94
North American	Nearctic	31
Oceanian	Australia, Pacific	61
Southern African	South Africa, Eswatini, Lesotho, and Namibia	66

Tropical	Tropical Africa, Asia, Central and South America, Madagascar and Pantropical	124
Macaronesian	Macaronesia	5

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579 **Table 2** Peak of naturalization (mode) and the number of species introduced per 20-
580 year bin period for growth form, dispersal mode and biogeographic origin categories across
581 264 Aotearoa New Zealand offshore islands.

Category	Peak of naturalization (Mode)	Number of species introduced
Growth forms		
Graminoids	1850–1870	34
Forbs	1850–1870	129
Woody species	1970–1990	49
Dispersal modes		
Wind-dispersed	1930–1950	27
Animal-dispersed	1850–1870	37
Unspecialized	1850–1870	89
Short-distance	1970–1990	34
Biogeographic origin		
Eurasian	1850–1870	100
Mediterranean	1850–1870	27
Oceanian	1970–1990	20
Southern African	1930–1950, 1970–1990	15, 15
Tropical	1970–1990	27

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583 **Table 3** Top section: separate linear relationships between the average area, isolation
584 and distance from urban areas (average urban) of occupied islands and year of first

585 naturalization across all species. Entries are model, estimate, t- and p-value. Bottom section:
 586 Interaction terms extracted from ANCOVA models between the average area, isolation and
 587 distance from urban areas (average urban) of occupied islands and time since first
 588 naturalization by growth forms, dispersal modes and biogeographic origin. Entries are model,
 589 sum of squares, F- and p-value. P-values are in bold when significant (<0.05).

Overall models			
Model	Estimate	t-value	p-value
log (average area) ~ year of first naturalization	0.00 ± 0.00	3.68	<0.001
arcsine (average isolation) ~ year of first naturalization	-0.00 ± 0.00	0.44	0.662
square root (average urban) ~ year of first naturalization	-0.25 ± 0.03	-7.60	<0.001
Interaction terms			
Model	Sum of squares	F-value	p-value
log (average area) ~ year of first naturalization * growth forms	3.60	1.72	0.180
log (average area) ~ year of first naturalization * dispersal modes	8.30	2.55	0.054
log (average area) ~ year of first naturalization * biogeographic origin	3.40	0.78	0.536
arcsine (average isolation) ~ year of first naturalization * growth forms	0.04	0.67	0.513
arcsine (average isolation) ~ year of first naturalization * dispersal modes	0.15	1.60	0.188
arcsine (average isolation) ~ year of first naturalization * biogeographic origin	0.15	1.23	0.295
square root (average urban) ~ year of first naturalization * growth	3346.00	0.90	0.406
square root (average urban) ~ year of first naturalization * dispersal	9098.00	1.54	0.203
square root (average urban) ~ year of first naturalization * biogeographic origin	14467.00	1.93	0.105

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593 **Figure Captions**

594 **[single column] Figure 1** Relationships between the number of islands occupied by a
595 species and their year of first naturalization for growth form (top), dispersal mode (centre)
596 and biogeographic origin (bottom) categories across 264 Aotearoa New Zealand offshore
597 islands. No statistical differences were identified in contrasting categories.

598 **[double column] Figure 2** Frequency plots displaying the number of species naturalized
599 per 20-year bin period for growth form (top), dispersal mode (centre) and biogeographic
600 origin (bottom) categories across 264 Aotearoa New Zealand offshore islands. Numbers
601 illustrate the number of species introduced at the mode (naturalization peak) by category.
602 Two modes were identified for Southern African species.

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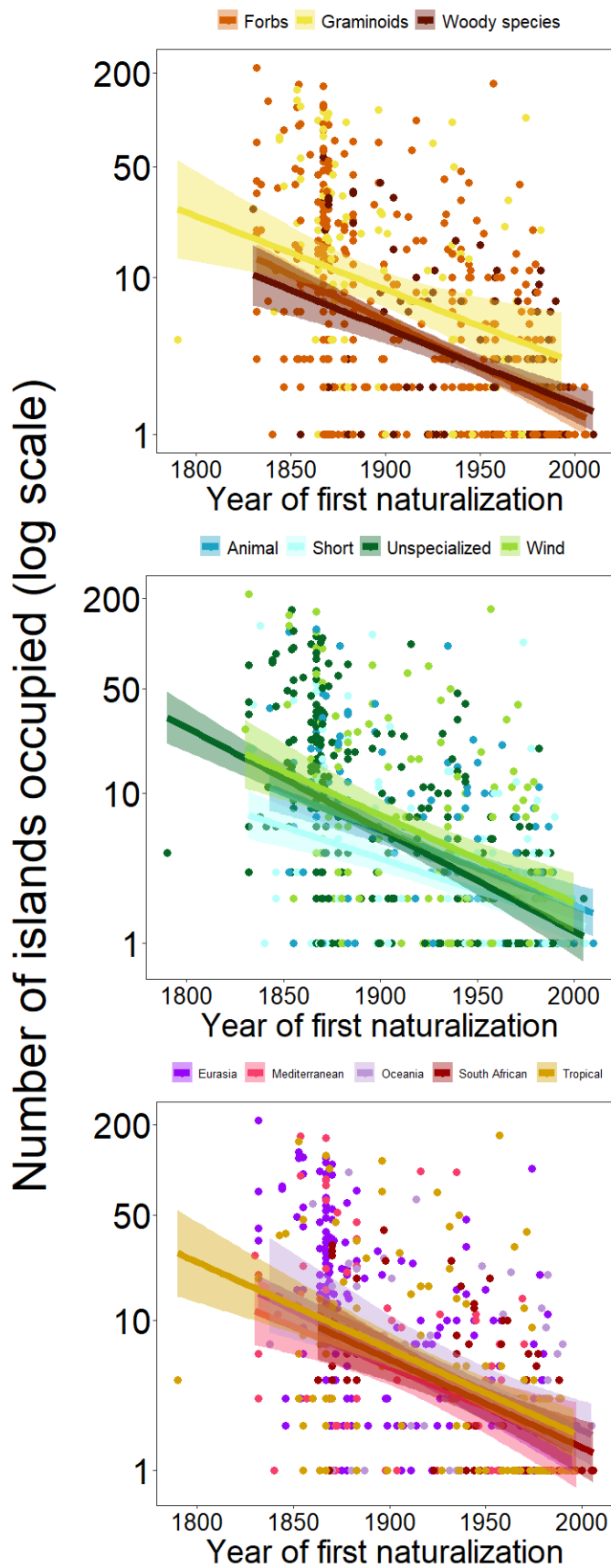
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613 **Figure 1**



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