1 Time of naturalization is key to explaining non-native plant invasio	ons on
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3 4 5	Running title: Naturalization time drives island invasions
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7	*Fabio Mologni <sup>1</sup>
8	Peter J. Bellingham <sup>2</sup>
9	Ewen K. Cameron <sup>3</sup>
10	Anthony E. Wright <sup>4</sup>
11	
12	
13	
14	
15	
16	1. Department of Biology
17	University of British Columbia Okanagan
18	1177 Research Road, Kelowna, BC V1V 1V7, Canada
19	
20	2. Manaaki Whenua—Landcare Research, Lincoln, Aotearoa New Zealand
21	
22	
23	3. Auckland War Memorial Museum, Auckland, Aotearoa New Zealand
24	
25	4. Canterbury Museum, Christchurch, Aotearoa New Zealand
26	
27	
28	* Corresponding author
29	Postal address: Department of Biology, University of British Columbia Okanagan, 1177
30	Research Road, Kelowna, BC V1V 1V7, Canada
31	<i>Phone:</i> +12369709033
32	Email: <u>fabio.mologni@ubc.ca</u>
33	ORCiD: https://orcid.org/0000-0003-4750-9974

34	Abstra	ct

35 Aim

36	Relatively long periods of time can elapse between the naturalization and spread of non-
37	native plant species. However, time lags on islands are poorly understood, especially if
38	integrating plant life histories. We asked whether (1) there is a time lag in the invasion
39	process, (2) there were distinct periods of naturalization in the non-native plants that invaded
40	islands and (3) non-native plants that naturalized more recently occur more frequently on
41	islands that are large, less isolated and close to urban areas. Then, we contrasted trends across
42	growth forms, dispersal modes and biogeographic origins.
43	
44	Location
45	264 offshore islands in northern Aotearoa New Zealand
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47	Taxa
48	Vascular plant species
49	
50	Methods
51	We combined field surveys and published data for 848 non-native plant species. We
52	categorized each species according to its growth form, dispersal mode and biogeographic
53	origin and identified its year of naturalization in Aotearoa New Zealand. We contrasted
54	period of naturalization, time lags and relationships with island area, isolation and distance
55	from the nearest urban area by growth form, dispersal mode and biogeographic origin using
56	ANCOVA and generalized linear models (GLMs).
57	
58	Results

59	We identified time lags, similar across all trait and biogeographic origin categories. Species
60	with different trait and biogeographic origin categories first naturalized at different periods in
61	time. Herbaceous species, those with unspecialized dispersal modes, and those originating
62	from Eurasia and the Mediterranean basin were disproportionately introduced earlier than
63	other categories. Non-native plants that naturalized more recently occur more frequently on
64	large islands close to urban areas, but not less isolated ones. Relationships with island
65	characteristics did not differ among trait and biogeographic origin categories.
66	
67	Main Conclusions
68	Overall, we found that the time of naturalization is more important than trait and
69	biogeographic origin categories in explaining non-native plant invasion patterns on islands.
70	Since similar time lags were identified for all categories, management bodies should focus on
71	species of trait and biogeographic origin categories that naturalized more consistently in
72	recent times (e.g. woody species from other regions within Oceania), and on large islands
73	close to urban areas. Plant life histories do not always play a role in explaining plant
74	distributions on islands.
75	
76	
77	Key Words: Aotearoa New Zealand, Biogeographic origin, Dispersal modes, Growth forms,
78	Island biogeography, Naturalization time, Plant species, Time lags
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#### 84 Introduction

The introduction of non-native species poses considerable threats to islands' biodiversity 85 (Kueffer et al., 2010; Lonsdale, 1999; Pyšek & Richardson, 2006). Large numbers of species 86 87 have been introduced to islands (Sax & Gaines, 2008, Wilkinson 2004) but, for some of 88 these, a long time can elapse before they spread (i.e. there is a time lag, or invasion debt; 89 González-Moreno et al., 2017; Kowarik, 1995; Seabloom et al., 2006). Time lags in invasion 90 might be regulated by species' life history traits (Duncan, 2021, Fristoe et al., 2023; Moyano, 91 2023). However, our understanding of time lags within insular environments remains limited, 92 and how the introduction and life histories of non-native species intersect is still largely 93 unexplored (Kinlock et al., 2022; Schrader et al. 2021a). 94 Rates of non-native species introduction have varied over time. The number of new species introduced worldwide remained relatively low until the 19th century (Seebens et al., 95 96 2017). After that, numbers increased consistently until now, with the exception of the period 97 1914–1945, marked by two major wars and rather constant introduction numbers (Pyšek et 98 al., 2020; Seebens et al., 2017). After 1945, rates of non-native species introduction increased 99 and are not slowing, correlated with increasing volumes of commerce worldwide (Pyšek et 100 al., 2020; Seebens et al., 2017). Clearly quantifying time lags between first introduction and 101 naturalization and later spread will help forecast future range expansions and guide 102 management focused on early containment or eradication. 103 Species introduced earlier are likely to occupy a larger number of islands than recent 104 introductions. However, island colonization is also regulated by island characteristics 105 (MacArthur & Wilson, 1967). Large islands offer a bigger target for dispersing propagules

and therefore are more likely to be occupied first (i.e. target island effect, Lomolino, 1990).

107 Similarly, less isolated islands are more easily reached by dispersing propagules and thus will

108 be occupied before more isolated islands (MacArthur & Wilson, 1967). Since the

introduction and naturalization of non-native species often rely on humans, islands that are
inhabited or close to human communities are likely to be more readily invaded (Sullivan et
al., 2005; Timmins & Williams, 1991). Interactions between these determinants of island
invasions and time lags have not been adequately investigated.

113 Time of introduction, time lags, and island characteristics that favour initial colonization of islands are likely to be related to plant life-histories. For instance, a 114 115 disproportionate number of herbaceous European species were initially introduced in many 116 parts of the world for agricultural purposes and in an attempt to "recreate" a landscape 117 familiar to settlers (Crosby, 2004; Lenzner et al., 2022). Time lags might be shorter for 118 species that disperse more efficiently, either because of shorter life cycles (e.g. herbs) or 119 because of morphological adaptations allowing for long-distance dispersal (Arjona et al., 120 2018; Mologni et al., 2022; Negoita et al., 2016 but see Duncan, 2021). Similarly, species 121 that originate from biogeographic areas with similar climates (i.e. climate matching) are 122 likely to have an advantage when establishing new populations (Richardson & Pyšek, 2012). 123 These traits might also affect how they relate to island characteristics. For instance, long-124 distance dispersal species might show weaker relationships with isolation (Arjona et al., 125 2018; Negoita et al., 2016), while high propagule pressure in species with short life cycles 126 might mask a target island effect.

Here, we aim to better understand the role of introduction histories and plant traits in explaining non-native plant invasions on islands. We calculated non-native species richness on 264 islands off the northern coast of Aotearoa New Zealand, for a total of 848 non-native vascular plant species. We asked: (1) Is there a time lag in the invasion process on the islands? If so, is it consistent across growth forms, dispersal modes, and biogeographic origins? (2) Do the non-native plants on the islands show distinct periods of naturalization in Aotearoa New Zealand? If so, were periods of naturalization similar across growth forms,

dispersal modes and biogeographic origins? (3) If non-native plant species colonize larger,
less isolated islands that are close to urban areas first, does the average (a) area, (b) isolation
and (c) distance from urban areas of the islands where a species occurs respectively decline,
increase and increase with time since first naturalization? If so, is it consistent across growth
forms, dispersal modes and biogeographic origins?

139

### 140 Methods

141 Study system

142 The study system comprises 264 islands off the northern coast of Aotearoa New Zealand's 143 North Island (hereafter the 'mainland'). The study system spans six hundred kilometres (34-144 38 °S, 172–179 °E, Figure 1). The islands vary in area from 0.00002 to 277.21 km<sup>2</sup>, and 145 range from islands connected to the mainland at low tides to those >50 km from the coast. 146 Most are land-bridge islands that were connected to the mainland during the last glacial 147 maximum (Poor Knight Islands excepted, Fleming, 1979). They supported warm temperate 148 rain forests before human settlement (circa 740 years ago, Wilmshurst et al., 2014). Humans 149 cleared, burned and farmed many islands (Bellingham et al., 2010), some of which were 150 abandoned in only the last few decades. Northern Aotearoa New Zealand is the most populous region of the country (Statistics NZ Tatauranga Aotearoa, 2018), and 41 islands are 151 152 still inhabited within the study system. A more detailed description of the study system is 153 provided by Mologni et al., 2022.

154

155 Data collection

156 The total number of non-native vascular plant species occupying each island was quantified 157 by collating previously published material and field surveys (see Mologni et al., 2021 for 158 more details). For each species, we obtained the year of their naturalization from the New 159 Zealand Plant Conservation Network and Gatehouse (2008). Then, we used as a proxy for 160 their time since first naturalization (i.e. the number of years they resided in Aotearoa New 161 Zealand) the current year (2023) minus the year of the first record. Of the 848 non-native 162 species, only 784 were included in our analyses. Thirty-three were removed due to taxonomic 163 uncertainty, while 31 were not included due to a lack of first-naturalization data.

164 We classified non-native species according to their growth form, dispersal mode and 165 biogeographic region (Table 1). Five categories of growth forms were identified: graminoids, 166 forbs, woody species, lycophytes and lianas and climbers. Five categories of dispersal modes 167 were defined based on morphological adaptations to long-distance dispersal: wind-, water-, 168 animal-dispersed, unspecialized and short-distance (Arjona et al., 2018; Burns, 2019; 169 Mologni et al., 2022; Negoita et al., 2016). Nine categories of biogeographic origin were delineated: Cosmopolitan, Eurasia (excl. Mediterranean basin), Mediterranean (i.e. 170 171 Mediterranean basin), Tropical, Southern Africa, Oceania (excl. Aotearoa New Zealand), 172 North America (i.e. species restricted to North America), Other American (i.e. species 173 ranging across all America or temperate South America) and Macaronesia (see Table 1 for 174 more details). All species could be categorized based on their growth forms; however, 18 and 175 40 species were respectively excluded from dispersal mode and biogeographic origin analyses due to a lack of data. Of these categories, seven were removed due to the small 176 177 sample size: lycophytes (number of species = 1), lianas and climbers (n = 41), water-178 dispersed (n = 16), cosmopolitan (n = 36), North America (n = 31), Other American (n = 16), 179 Macaronesia (n = 6). In total, we used 3 categories for growth forms, 4 for dispersal modes 180 and 6 for biogeographic origins (Table 1). 181 Island characteristics were quantified using ArcGIS 10.5.1 (ESRI, 2011). The size of

an island was calculated as the total surface area in km<sup>2</sup>. Isolation was measured as the total
amount of water surrounding each island (Carter et al., 2020; Diver, 2008; Negoita et al.,

184 2016). The greater the amount of water, the more isolated an island is. For each island, we 185 created several concentric buffers with radii of 250, 500, 1000, 1500, 2000, 2500 and 3000 186 m. Then, we calculated the amount of water within each radius (%). To quantify the effect of 187 isolation from human communities, we calculated the shortest distance from the nearest 188 urban area (m). We defined as urban built-up areas having a population of 1000 people or 189 more (Statistics NZ Tatauranga Aotearoa, 2018).

190

191 Statistical analyses

192 To test for time lags in the invasion process, we used a generalized linear model (GLM). We 193 set the number of islands occupied by each species as the dependent variable and years since 194 first naturalization time as the independent variable. Since the dependent variable was right-195 skewed even after transformation, we modelled data using linear, quasipoisson and negative 196 binomial models for all species. Results were consistent and, for simplicity, we opted for a 197 linear model. We assessed differences in slopes among different trait and biogeographic 198 origin categories with ANCOVA tests. The number of occurrences was log-transformed in all 199 linear models to conform to assumptions of normality.

To test whether non-native species show distinct periods of their naturalization to Aotearoa New Zealand, modes were extracted for each trait and biogeographic origin category after binning the year of naturalization by 20-year periods. Additionally, we contrasted the distribution of each category of growth forms, dispersal modes and biogeographic origin using ANOVAs.

To test whether non-native plants that naturalized more recently occur more frequently on islands that are large, less isolated and close to urban areas, we set the average area, isolation and distance from urban areas of occupied islands by each species as the dependent variables in three separate linear models. Years since first naturalization was used

as the independent variable. To conform to assumptions of normality, the average area,
isolation and distance from urban areas of occupied islands were respectively log-, arcsineand square root-transformed. Tests were run for all species first, then ANCOVA tests were
used to contrast slopes among different trait and biogeographic origin categories. All analyses
were conducted in R (R Core Team, 2023). A post hoc Tukey test and Bonferroni correction
were applied for multiple testing at all stages (Hothorn et al. 2008).

215

#### 216 **Results**

There was a time lag in the invasion process: the earlier a non-native plant species had naturalized in Aotearoa New Zealand, the more islands it occupied (Figure 1 & Table S1). The slopes of relationships between time since first naturalization and number of islands occupied did not differ significantly among different non-native plant growth forms, dispersal modes or biogeographic origins (Table S2).

222 The non-native plant species on the islands had mainly two 20-year bin periods of 223 peak naturalization – 1850 to 1870 and 1970 to 1990 (Figure 2, Table 2 & Table S3). Non-224 native graminoids and forbs on the islands had disproportionately first naturalized in 225 Aotearoa New Zealand between 1850–1870, whereas woody species had disproportionately first naturalized between 1970–1990 (Figure 2, Table 2 & Table S3). Differences between 226 227 non-native woody species and other growth forms were significant (ANOVA and Tukey 228 tests: Table S4). Among dispersal modes, unspecialized and animal-dispersed species on the 229 islands had disproportionately first naturalized in Aotearoa New Zealand between 1850– 230 1870, whereas short-distance species had disproportionately first naturalized between 1970-231 1990 (Figure 2, Tables 2 & Table S3). Wind-dispersed species first naturalized 232 disproportionately 1930–1950 (Figure 2, Tables 2 & Table S3). Differences between 233 unspecialized species and other dispersal modes were significant (ANOVA and Tukey tests:

Table S4). Non-native Eurasian and Mediterranean plant species on the islands had
disproportionately first naturalized during1850–1870, those from Southern Africa in both
1930–1950 and 1970–1990, and those from Oceania and tropical regions during 1970–1990
(Figure 2, Table 2 & Table S3); a significant difference was identified between Eurasian and
Mediterranean species and those from other biogeographic regions (ANOVA, Tukey tests:
Table S4).

The average area and distance from urban areas of islands where a species occurred respectively declined and increased with time since first naturalization (Table 3). No significant relationship was detected between the average isolation of islands where a species occurred and time since first naturalization (Table 3). There were no significant differences in non-native plants' growth forms, dispersal modes and biogeographic origins for the average area, isolation and distance from urban areas of islands where a species occurred (Table 3 & Table S5).

247

#### 248 **Discussion**

249 Time lags in the invasion process are commonly reported (Duncan, 2021; Kowarik, 1995; 250 Pyšek et al., 2020; Seabloom et al., 2006). Here, we found evidence of time lags in the invasion of non-native plants on islands in northern Aotearoa New Zealand. Moreover, there 251 252 were differences in periods of first naturalization among growth forms, dispersal modes and 253 biogeographic origins. Consistent with studies in the same system (Mologni et al., 2021), 254 large islands close to humans were more frequently colonized first by invading species. 255 However, once time of first naturalization was accounted for, and in contrast to an earlier 256 study (Mologni et al., 2022), there were no differences across categories of growth forms, 257 dispersal modes and biogeographic origin.

258 Once time of naturalization was accounted for, it was a better predictor than life 259 history traits in explaining non-native plant invasions on the islands. For example, non-native 260 woody plant species have proportionately colonized fewer islands but most of them have 261 naturalized more recently than other growth forms. Herbaceous species used for agriculture 262 were introduced and naturalized disproportionately early, whereas woody, ornamental species 263 naturalized disproportionately recently (Gatehouse, 2008). Among dispersal modes, more 264 unspecialized plant species naturalized early, perhaps reflecting a higher efficiency in 265 exploiting human-related pathways. Among biogeographic origins, more plant species from 266 Eurasia and the Mediterranean naturalized early, a reflection of European settlement 267 (Gatehouse, 2008; Lenzner et al., 2022). In particular, species originating from other regions 268 within Oceania (mostly Australian species) are disproportionately woody (60.3%, see Table 269 S6), indicating this trait combination is of utmost importance for future management 270 endeavours.

271 If there are similar invasion time lags irrespective of trait and biogeographic origin 272 categories, we might expect that, in time, more non-native woody species will invade more 273 islands, as the earlier-naturalized herbaceous species have already done (Howell, 2019). Even 274 so, the time lag for woody species may still be longer since woody species will generally be 275 slower to reach reproductive maturity than herbaceous species (Grime 2002). In support of 276 this view, the naturalization of non-native conifers in Aotearoa New Zealand continued at a 277 linear rate to this day (Howell, 2019), even though there had been no or few introductions of 278 new non-native conifer species to the country since its 1993 Biosecurity Act significantly 279 reduced introduction rates of non-native species (Seebens et al., 2017). Similar trends might 280 be expected for other categories that were more consistently introduced in recent times. 281 Previous work in the study system found that graminoids and long-distance dispersal

282 species have higher occupancy rates than other trait categories and distributional patterns

283 more similar to native species (Mologni et al., 2022). Here, we found that their invasion 284 success is not associated with shorter time lags. Other processes might be at play. For 285 instance, these species might have been purposely introduced more consistently than other 286 categories. Alternatively, species introduced and naturalized early might negatively impact 287 the establishment of other non-native species (i.e. priority effect, Fraser et al., 2015, Catford 288 et al. 2022). For instance, herbaceous European and Mediterranean non-native species that 289 naturalized early and invaded the islands might reduce the probability of successful 290 establishment by herbaceous non-native species from other biogeographic regions through 291 niche pre-emption (Fukami 2015).

292 Non-native plants that naturalized more recently occurred more frequently on large 293 islands, consistently with the target island effect (Lomolino, 1990). However, large islands in 294 the region are also inhabited (Mologni et al., 2021). Whether the more consistent colonization 295 of large islands reflects autonomous or human-mediated dispersal is unclear, both processes 296 though are likely at play. Since they both increase the likelihood of large over small island 297 colonization, island area might then simultaneously encapsulate a "geographic" target island 298 effect (i.e. species colonize larger islands first because they offer a bigger target) and a 299 "human-mediated" target islands effect (i.e. human inhabit larger islands, disproportionately 300 introducing non-native species).

Non-native plants that naturalized more recently did not occur more frequently on less isolated islands. However, a negative effect of isolation on non-native species richness was previously identified (Mologni et al., 2021). Perhaps this pattern is simply more difficult to capture, and, over time, less isolated islands still accumulate more species than more isolated islands through multiple colonization events. In contrast, non-native plants that naturalized more recently occurred more frequently on islands that are close to urban areas. Since nonnative species often use human-related pathways (Pyšek et al., 2020; Sinclair et al., 2020),

islands close to urban areas are more likely to be colonized just by proximity (Sullivan et al.,
2005; Timmins & Williams, 1991). For non-native species in this study system, isolation
might be more strongly associated with humans than with geography.

311 The relationships between island characteristics and time since first naturalization 312 were similar across trait and biogeographic origin categories. This suggests that the same 313 processes regulate the initial colonization of these islands, irrespective of plant life histories. 314 This is surprising since a growing body of literature has identified differences in island 315 species distributions based on traits (Conti et al., 2022; Mologni et al., 2022; Schrader et al., 316 2021b; Schrader et al., 2023; Walentowitz et al., 2022). Alternatively, other traits might be 317 more important. For instance, wind-pollinated species are independent of biological carriers 318 for dispersing their pollen, an advantage in newly occupied islands (Cox et al., 1997). 319 Propagule pressure is another trait that likely influences the ability to rapidly colonize more 320 islands (Lockwood et al., 2005). Future studies should assess a broad suite of traits when 321 assessing the importance of time lags in the invasion process on islands. 322 The year of naturalization was used to identify periods of non-native plant species' 323 naturalizations. However, in many cases, long periods of time can elapse between the 324 introduction, naturalization and identification of newly naturalized species. This is

325 particularly important for earlier introductions, more likely to go unnoticed due to a lack of

326 awareness of biological invasions. Conversely, we can expect current monitoring activities to

be more effective in identifying new non-native species (Pyšek et al., 2020). Unfortunately,

accounting for this form of error will be challenging.

329 Since similar time lags were identified for all trait and biogeographic origin categories 330 despite different periods of naturalization, management bodies should focus on species of 331 categories that naturalized more consistently in recent times, such as woody species (Howell, 332 2019) originating from other regions within Oceania and short-distance dispersal species, and

334	growing body of literature implementing traits in island biogeography (Ottaviani et al., 2020;
335	Schrader et al., 2023; Whittaker et al., 2014), suggesting, however, that plant life histories do
336	not always play a role in explaining plant distributions on islands.
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338	References
339	Arjona, Y., Nogales, M., Heleno, R., & Vargas, P. (2018). Long-distance dispersal
340	syndromes matter: Diaspore-trait effect on shaping plant distribution across the
341	Canary Islands. <i>Ecography</i> , 41(5), 805-814. https://doi.org/10.1111/ecog.02624
342	Bellingham, P. J., Towns, D. R., Cameron, E. K., Davis, J. J., Wardle, D. A., Wilmshurst, J.
343	M., & Mulder, C. P. H. (2010). New Zealand island restoration: Seabirds, predators,
344	and the importance of history. New Zealand Journal of Ecology, 34(1), 115–136.
345	Burns, K. C. (2019). Evolution in Isolation: The Search for an Island Syndrome in Plants.
346	Cambridge University Press.
347	Carter, Z. T., Perry, G. L. W., & Russell, J. C. (2020). Determining the underlying structure
348	of insular isolation measures. Journal of Biogeography, 47(4), 955–967.
349	https://doi.org/10.1111/jbi.13778
350	Catford, J. A., Wilson J. R., Pyšek, P., Hulme, P. E., Duncan, R. P. (2022). Addressing
351	context dependence in ecology. Trends in Ecology & Evolution, 37(2), 158-170.
352	Conti, L., Méndez-Castro, F. E., Chytrý, M., Götzenberger, L., Hájek, M., Horsák, M.,
353	Jiménez-Alfaro, B., Klimešová, J., Zelený, D., & Ottaviani, G. (2022). Insularity
354	promotes plant persistence strategies in edaphic island systems. Global Ecology and
355	Biogeography, 31(4), 753-764. https://doi.org/10.1111/geb.13465
356	Cox, P. A., Grubb, P. J., Chaloner, W. G., Harper, J. L., & Lawton, J. H. (1997). Abiotic
357	pollination: An evolutionary escape for animal-pollinated angiosperms. Philosophical

on large islands close to urban areas. Additionally, this work builds on and aligns with a

- 358 Transactions of the Royal Society of London. Series B: Biological Sciences,
- 359 *333*(1267), 217–224. https://doi.org/10.1098/rstb.1991.0070
- 360 Crosby, A. W. (2004). *Ecological Imperialism: The Biological Expansion of Europe*, 900361 1900. Cambridge University Press.
- 362 Diver, K. C. (2008). Not as the crow flies: Assessing effective isolation for island
- 363 biogeographical analysis. *Journal of Biogeography*, *35*(6), 1040–1048.
- 364 https://doi.org/10.1111/j.1365-2699.2007.01835.x
- 365 Duncan, R. P. (2021). Time lags and the invasion debt in plant naturalizations. *Ecology*
- 366 *Letters*, 24(7), 1363–1374. https://doi.org/10.1111/ele.13751
- 367 ESRI. (2011). ArcGIS desktop: Release 10.
- Fleming, C. A. (1979). *The geological history of New Zealand and its life*. Auckland
  University Press.
- 370 Fraser, C. I., Banks, S. C., & Waters, J. M. (2015). Priority effects can lead to
- underestimation of dispersal and invasion potential. *Biological Invasions*, *17*(1), 1–8.
  https://doi.org/10.1007/s10530-014-0714-1
- 373 Fristoe, T. S., Bleilevens, J., Kinlock, N. L., Yang, Q., Zhang, Z., Dawson, W., Essl, F.,
- 374 Kreft, H., Pergl, J., Pyšek, P., Weigelt, P., Dufour-Dror, J.-M., Sennikov, A. N.,
- 375 Wasowicz, P., Westergaard, K. B., & Van Kleunen, M. (2023). Evolutionary
- imbalance, climate and human history jointly shape the global biogeography of alien
- 377 plants. Nature Ecology & Evolution. https://doi.org/10.1038/s41559-023-02172-z
- 378 Fukami, T. (2015). Historical contingency in community assembly: integrating niches,
- 379 species pools, and priority effects. *Annual Review of Ecology, Evolution, and*380 *Systematics*, 46, 1-23.
- 381 Gatehouse, H. A. W. (2008). Ecology of the naturalization and geographic distribution of the
   382 non-indigenous seed plant species of New Zealand. Lincoln University.

383	González-Moreno, P., Pino, J., Cózar, A., García-de-Lomas, J., & Vilà, M. (2017). The
384	effects of landscape history and time-lags on plant invasion in Mediterranean coastal
385	habitats. Biological Invasions, 19(2), 549-561. https://doi.org/10.1007/s10530-016-
386	1314-z
387	Grime, J. P. (2002). Plant strategies, vegetation processes, and ecosystem properties. Second
388	edition. Wiley, Chichester.
389	Howell, C. J. (2019). Naturalized status of exotic conifers in New Zealand. New Zealand
390	Journal of Botany, 57(4), 227-237. https://doi.org/10.1080/0028825X.2019.1626744
391	Kinlock, N. L., Dehnen-Schmutz, K., Essl, F., Pergl, J., Pyšek, P., Kreft, H., Weigelt, P.,
392	Yang, Q., & van Kleunen, M. (2022). Introduction history mediates naturalization and
393	invasiveness of cultivated plants. Global Ecology and Biogeography, 31(6), 1104-
394	1119. https://doi.org/10.1111/geb.13486
395	Kowarik, I. (1995). Time lags in biological invasions with regard to the success and failure of
396	alien species. In P. Pyšek, K. Prach, M. Rejmánek, & M. Wade (Eds.), Plant
397	Invasions—General aspects and special problems (pp. 15–38).
398	Kueffer, C., Daehler, C. C., Torres-Santana, C. W., Lavergne, C., Meyer, JY., Otto, R., &
399	Silva, L. (2010). A global comparison of plant invasions on oceanic islands.
400	Perspectives in Plant Ecology, Evolution and Systematics, 12(2), 145–161.
401	https://doi.org/10.1016/j.ppees.2009.06.002
402	Lenzner, B., Latombe, G., Schertler, A., Seebens, H., Yang, Q., Winter, M., Weigelt, P., van
403	Kleunen, M., Pyšek, P., Pergl, J., Kreft, H., Dawson, W., Dullinger, S., & Essl, F.
404	(2022). Naturalized alien floras still carry the legacy of European colonialism. Nature
405	Ecology & Evolution, 6(11), Article 11. https://doi.org/10.1038/s41559-022-01865-1
406	Lockwood, J. L., Cassey, P., & Blackburn, T. (2005). The role of propagule pressure in
407	explaining species invasions. Trends in Ecology & Evolution, 20(5), 223–228.

- 408 https://doi.org/10.1016/j.tree.2005.02.004
- 409 Lomolino, M. V. (1990). The Target Area Hypothesis: The Influence of Island Area on
- 410 Immigration Rates of Non-Volant Mammals. *Oikos*, 57(3), 297–300.
- 411 https://doi.org/10.2307/3565957
- 412 Lonsdale, W. M. (1999). Global patterns of plant invasions and the concept of invasibility.
- 413 *Ecology*, 80(5), 1522–1536. https://doi.org/10.1890/0012-
- 414 9658(1999)080[1522:GPOPIA]2.0.CO;2
- 415 Mac Arthur, R., & Wilson, E. (1967). *The theory of island biogeography*. Princeton
  416 University Press.
- 417 Mologni, F., Bellingham, P. J., Cameron, E. K., Dinh, K., Wright, A. E., & Burns, K. C.
- 418 (2022). Functional traits explain non-native plant species richness and occupancy on
- 419 northern New Zealand islands. *Biological Invasions*, 24(7), 2135–2154.
- 420 https://doi.org/10.1007/s10530-022-02762-1
- 421 Mologni, F., Bellingham, P. J., Tjørve, E., Cameron, E. K., Wright, A. E., Burns, K. C., &
- 422 Munoz, F. (2021). Similar yet distinct distributional patterns characterize native and
- 423 exotic plant species richness across northern New Zealand islands. *Journal of*

424 *Biogeography*, 48(7), 1731–1745. https://doi.org/10.1111/jbi.14110

- 425 Moyano, J. (2023). Origins of successful invasions. Nature Ecology & Evolution.
- 426 https://doi.org/10.1038/s41559-023-02190-x
- 427 Negoita, L., Fridley, J. D., Lomolino, M. V., Mittelhauser, G., Craine, J. M., & Weiher, E.
- 428 (2016). Isolation-driven functional assembly of plant communities on islands.
- 429 *Ecography*, *39*(11), 1066–1077. https://doi.org/10.1111/ecog.01551
- 430 New Zealand Plant Conservation Network. (2023). https://www.nzpcn.org.nz/
- 431 Ottaviani, G., Keppel, G., Götzenberger, L., Harrison, S., Opedal, Ø. H., Conti, L., Liancourt,
- 432 P., Klimešová, J., Silveira, F. A. O., Jiménez-Alfaro, B., Negoita, L., Doležal, J.,

433	Hájek, M., Ibanez, T., Méndez-Castro, F. E., & Chytrý, M. (2020). Linking Plant
434	Functional Ecology to Island Biogeography. Trends in Plant Science, 25(4), 329–339.
435	https://doi.org/10.1016/j.tplants.2019.12.022
436	Pyšek, P., Hulme, P. E., Simberloff, D., Bacher, S., Blackburn, T. M., Carlton, J. T., Dawson,
437	W., Essl, F., Foxcroft, L. C., Genovesi, P., Jeschke, J. M., Kühn, I., Liebhold, A. M.,
438	Mandrak, N. E., Meyerson, L. A., Pauchard, A., Pergl, J., Roy, H. E., Seebens, H.,
439	Richardson, D. M. (2020). Scientists' warning on invasive alien species. Biological
440	Reviews, 95(6), 1511–1534. https://doi.org/10.1111/brv.12627
441	Pyšek, P., & Richardson, D. M. (2006). The biogeography of naturalization in alien plants.
442	Journal of Biogeography, 33(12), 2040–2050. https://doi.org/10.1111/j.1365-
443	2699.2006.01578.x
444	R Core Team. (2023). R: A language and environment for statistical computing. R
445	Foundation for Statistical Computing. https://www.r-project.org/
446	Richardson, D. M., & Pyšek, P. (2012). Naturalization of introduced plants: Ecological
447	drivers of biogeographical patterns. New Phytologist, 196(2), 383-396.
448	https://doi.org/10.1111/j.1469-8137.2012.04292.x
449	Schrader, J., Wright, I. J., Kreft, H., Weigelt, P., Andrew, S. C., Abbott, I., & Westoby, M.
450	(2023). ETIB-T: An Equilibrium Theory of Island Biogeography for plant traits.
451	Journal of Biogeography, 50(1), 223-234. https://doi.org/10.1111/jbi.14526
452	Seabloom, E. W., Williams, J. W., Slayback, D., Stoms, D. M., Viers, J. H., & Dobson, A. P.
453	(2006). Human impacts, plant invasion, and imperiled plant species in California.
454	Ecological Applications, 16(4), 1338–1350. https://doi.org/10.1890/1051-
455	0761(2006)016[1338:HIPIAI]2.0.CO;2
456	Seebens, H., Blackburn, T. M., Dyer, E. E., Genovesi, P., Hulme, P. E., Jeschke, J. M.,
457	Pagad, S., Pyšek, P., Winter, M., Arianoutsou, M., Bacher, S., Blasius, B., Brundu,

- 458 G., Capinha, C., Celesti-Grapow, L., Dawson, W., Dullinger, S., Fuentes, N., Jäger,
- 459 H., ... Essl, F. (2017). No saturation in the accumulation of alien species worldwide.
  460 *Nature Communications*, 8(1), 14435. https://doi.org/10.1038/ncomms14435
- 461 Sinclair, J. S., Lockwood, J. L., Hasnain, S., Cassey, P., & Arnott, S. E. (2020). A framework
- 462 for predicting which non-native individuals and species will enter, survive, and exit
- 463 human-mediated transport. *Biological Invasions*, 22(2), 217–231.
- 464 https://doi.org/10.1007/s10530-019-02086-7
- Schrader, J., Westoby, M., Wright, I. J., Kreft, H., & Sfenthourakis, S. (2021). Disentangling
  direct and indirect effects of island area on plant functional trait distributions. Journal
  of Biogeography, 48(8), 2098–2110. https://doi.org/10.1111/jbi.14138
- Schrader, J., Wright, I. J., Kreft, H., & Westoby, M. (2021). A roadmap to plant functional
  island biogeography. Biological Reviews, 96(6), 2851–2870.
- 470 https://doi.org/10.1111/brv.12782
- 471 Statistics NZ Tatauranga Aotearoa. (2018). Urban Area 2017 (generalised version).
- 472 Sullivan, J. J., Timmins, S. M., & Williams, P. A. (2005). Movement of exotic plants into
- 473 coastal native forests from gardens in northern New Zealand. *New Zealand Journal of*474 *Ecology*, 29(1), 1-10.
- Timmins, S. M., & Williams, P. A. (1991). Weed numbers in New Zealand's forest and scrub
  reserves. *New Zealand Journal of Ecology*, *15*(2), 153–162.
- 477 Walentowitz, A., Troiano, C., Christiansen, J. B., Steinbauer, M. J., & Barfod, A. S. (2022).
- 478 Plant dispersal characteristics shape the relationship of diversity with area and
- 479 isolation. *Journal of Biogeography*, 49(9), 1599–1608.
- 480 https://doi.org/10.1111/jbi.14454
- 481 Whittaker, R. J., Rigal, F., Borges, P. A. V., Cardoso, P., Terzopoulou, S., Casanoves, F., Pla,
- 482 L., Guilhaumon, F., Ladle, R. J., & Triantis, K. A. (2014). Functional biogeography of

483	oceanic islands and the scaling of functional diversity in the Azores. Proceedings of
484	the National Academy of Sciences, 111(38), 13709–13714.
485	https://doi.org/10.1073/pnas.1218036111
486	Wilmshurst, J. M., Moar, N. T., Wood, J. R., Bellingham, P. J., Findlater, A. M., Robinson, J.
487	J., & Stone, C. (2014). Use of pollen and ancient DNA as conservation baselines for
488	offshore islands in New Zealand. Conservation Biology, 28(1), 202–212.
489	https://doi.org/10.1111/cobi.12150
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491	Data availability statement
492	Data are available as supplementary material. Island locations (longitude and latitude) are not
493	publicly available due to private ownership and issues of data sovereignty of concern to
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## 505 Tables

**Table 1**Description and the number of species for growth form, dispersal mode and507biogeographic origin categories across 264 Aotearoa New Zealand offshore islands.

Category	Description	Number of species
Growth forms		
Graminoids	Grasses, sedges and rushes	125
Forbs	Herbaceous, non-graminoid	472
Woody species	Trees and shrubs	145
Lycophytes	Lycophytes	1
Climbers and lianas	Herbaceous or woody climbers	41
	Dispersal modes	766
Wind-dispersed	Plumes, wings, dust diaspores (anemochory)	141
Water-dispersed	Buoyant propagules, e.g. corky tissues, air pockets (hydrochory)	16
Animal-dispersed	Fleshy fruits or adhesive barbs (endo and epizoochory)	157
Unspecialized	No evident or prevalent morphological adaptations (unspecialized)	286
Short-distance	Morphological adaptations for short-distance dispersal only (ballochory, myrmecory)	166
Biogeographic origin		
Eurasia	Europe (excl. Mediterranean basin), Asia, North Africa and the Arabian Peninsula	310
Cosmopolitan	Cosmopolitan, subcosmopolitan and temperate areas of the northern hemisphere	36
Other American	All America (North, Central, South) or Temperate South America	16
Mediterranean	Mediterranean basin	94
North America	Nearctic	31
Oceania	Australia, Pacific	63

Southern Africa	South Africa, Eswatini, Lesotho, and Namibia	66
Tropical	Tropical Africa, Asia, Central and South America, Madagascar and Pantropical	127
Macaronesia	Macaronesia	6

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530 Table 2 Peak of naturalization (mode) and the number of species introduced per 20531 year bin period for growth form, dispersal mode and biogeographic origin categories across
532 264 Aotearoa New Zealand offshore islands.

Category	Peak of introduction (Mode)	Number of species introduced		
	Growth forms			
Graminoids	1850–1870	34		
Forbs	1850–1870	129		
Woody species	1970–1990	51		
	<b>Dispersal modes</b>			
Wind-dispersed	1930–1950	27		
Animal-dispersed	1850–1870	37		
Unspecialized	1850–1870	89		
Short-distance	1970–1990	37		
	biogeographic origin			
Eurasia	1850–1870	100		
Mediterranean	1850–1870	27		
Oceania	1970–1990	20		
South Africa	1930–1950, 1970–1990	15, 15		
Tropical	1970–1990	27		

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541 **Table 3** Top section: separate linear relationships between the average area, isolation 542 and distance from urban areas (average urban) of occupied islands and time since first 543 naturalization across all species. Entries are model, estimate, t- and p-value. Bottom section: 544 Interaction terms extracted from ANCOVA models between the average area, isolation and 545 distance from urban areas (average urban) of occupied islands and time since first 546 naturalization by growth forms, dispersal modes and biogeographic origin. Entries are model, 547 sum of squares, t- and p-value. P-values are in bold when significant (<0.05).</p>

Overall models						
Model	Estimate	t-value	p-value			
log (average area) ~ time since first naturalization	$-0.00\pm0.00$	-3.68	<0.001			
arcsine (average isolation) ~ time since first naturalization	$0.00\pm0.000$	0.44	0.662			
square root (average urban) ~ time since first naturalization	$0.25\pm0.03$	7.60	<0.001			
Interaction	terms					
Model	Sum of squares	t-value	p-value			
log (average area) ~ time since first naturalization *growth forms	3.60	1.71	0.181			
log (average area) ~ time since first naturalization *dispersal modes	8.30	2.55	0.054			
log (average area) ~ time since first naturalization *biogeographic origin	3.40	0.78	0.536			
arcsine (average isolation) ~ time since first naturalization *growth forms	0.04	0.67	0.513			
arcsine (average isolation) ~ time since first naturalization *dispersal modes	0.15	1.60	0.188			
arcsine (average isolation) ~ time since first naturalization *biogeographic origin	0.15	1.23	0.295			
square root (average urban) ~ time since first naturalization *growth	3345.00	0.90	0.405			
square root (average urban) ~ time since first naturalization *dispersal	9098.00	1.54	0.203			
square root (average urban) ~ time since first naturalization *biogeographic origin	14467.00	1.93	0.105			

# 552 Figure Captions

Figure 1 Relationships between the number of islands occupied by a species and years
since first naturalization for growth form (top), dispersal mode (centre) and biogeographic
origin (bottom) categories across 264 Aotearoa New Zealand offshore islands. No statistical
differences were identified in contrasting categories.

557	Figure 2	Frequency plots displaying the number of species naturalized per 20-year bin
558	period for gr	owth form (top), dispersal mode (centre) and biogeographic origin (bottom)
559	categories ac	ross 264 Aotearoa New Zealand offshore islands. Numbers illustrate the number
560	of species in	roduced at the mode (introduction peak) by category. Two modes were
561	identified for	Southern African species. Asterisks indicate significant differences in the
562	number of sp	becies naturalized by 20-year intervals (p-value <0.05).
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**577 Figure 2** 



## 580 Supplementary Material

581 **Table S1** Relationships between the number of islands occupied and time since first 582 naturalization of all species across 264 Aotearoa New Zealand offshore islands. Entries are 583 model utilized estimate t and p value. P values are in hold when significant (20.05)

583 model utilized, estimate, t- and p-value. P-values are in bold when significant (<0.05).

Model	Estimate	t-value	p-value
Linear (log-transformed)	$0.01 \pm 0.00$	14.85	<0.001
Quasipoisson	$0.02\pm0.00$	10.49	<0.001
Negative binomial	$0.02\pm\ 0.00$	17.82	<0.001

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599 **Table S2** ANOVA and Tukey tests for multiple comparisons of relationships between 600 the number of islands occupied by a species and its time since first naturalization across 601 categories of growth forms, dispersal modes and biogeographic origin. In brackets is the level 602 of significance of ANOVA tests (NS > 0.05, \* < 0.05, \*\* < 0.01, \*\*\* < 0.001). Columns are 603 categories and estimates, t- and p-values of Tukey tests.  $\alpha$  was set at 0.003 after applying a

604 Bonferroni correction for multiple comparisons.

Category	Estimate	t-value	p-value			
Growth forms (NS)						
Graminoids - Forbs	$0.28\pm0.31$	0.90	0.633			
Woody - Forbs	$0.13\pm0.22$	0.62	0.805			
Woody - Graminoids	$-0.14 \pm 0.33$	-0.43	0.902			
	Dispersal modes (NS)					
Unspecialized - Short-distance	$-0.27 \pm 0.24$	-1.09	0.698			
Unspecialized - Animal	$-0.43 \pm 0.26$	-1.68	0.332			
Unspecialized - Wind	$0.41 \pm 0.28$	1.47	0.458			
Wind - Animal	$-0.02 \pm 0.29$	-0.06	1.000			
Wind - Short-distance	$0.15\pm0.28$	0.53	0.951			
Short-distance - Animal	$-0.17 \pm 0.25$	-0.66	0.911			
	Biogeographic origin (NS	)				
Eurasia - Mediterranean	$0.19\pm0.37$	0.53	0.984			
Eurasia - Oceania	$0.33 \pm 0.33$	1.01	0.850			
Eurasia - Southern Africa	$0.20\pm0.34$	0.60	0.975			
Eurasia - Tropical	$0.30\pm0.27$	1.10	0.805			
Mediterranean - Oceania	$0.14\pm0.42$	0.32	0.998			
Mediterranean - Southern Africa	$0.01 \pm 0.43$	0.02	1.000			
Mediterranean - Tropical	$0.11\pm0.38$	0.28	0.999			
Oceania - Southern Africa	$-0.13 \pm 0.40$	-0.32	0.998			
Oceania - Tropical	$-0.03 \pm 0.34$	-0.08	1.000			
Tropical - Southern Africa	$0.10\pm0.36$	0.28	0.999			

**Table S3**Peak of naturalization (Mode) and the number of species introduced by 20-<br/>year bin period for growth form, dispersal mode and biogeographic origin categories across<br/>264 Aotearoa New Zealand offshore islands. 

Category	1790- 1810	1810–30	1830–50	1850–70	1870–90	1890- 1910	1910–30	1930–50	1950–70	1970–90	1990- 2010
	Growth forms										
Graminoids	1	0	5	34	30	10	9	15	10	9	2
Forbs	0	0	21	129	72	29	18	62	64	63	14
Woody species	0	1	0	16	13	8	12	18	16	51	10
	Dispersal modes										
Wind	0	1	1	24	22	9	12	27	20	22	3
Animal	0	0	1	37	21	8	7	19	23	33	8
Unspecialized	1	0	15	89	54	20	15	38	21	30	3
Short- distance	0	0	9	29	22	10	5	14	27	37	13
				Bio	ogeographic	c origin					
Eurasia	0	0	11	100	59	20	14	35	29	28	3
Mediterranea n	0	1	5	27	18	8	4	15	8	7	1
Oceania	0	0	1	4	7	5	5	5	7	20	6
South Africa	0	0	0	9	7	2	1	15	12	15	4
Tropical	1	0	3	22	10	6	8	18	22	27	4

- 615 **Table S4** ANOVA and Tukey tests for multiple comparisons of naturalization time
- 616 frequency distributions across categories of growth forms, dispersal modes and
- biogeographic origins. In brackets is the level of significance of ANOVA tests (NS > 0.05, \*

618 < 0.05, \*\* < 0.01, \*\*\* < 0.001). Columns are categories and estimates, t- and p-values of

619 Tukey tests. P-values are in bold when significant,  $\alpha$  was set at 0.003 after applying a

620 Bonferroni correction for multiple comparisons.

Category	Estimate	t-value	p-value			
Growth forms (***)						
Graminoids - Forbs	$-12.95 \pm 4.71$	-2.75	0.016			
Woody - Forbs	30.62 ± 4.44	6.89	<0.001			
Woody - Graminoids	43.57 ± 5.71	7.63	<0.001			
	Dispersal modes (***)					
Unspecialized - Short-distance	$-26.15 \pm 4.63$	-5.65	<0.001			
Unspecialized - Animal	-24.11 ± 4.71	-5.12	<0.001			
Unspecialized - Wind	$20.92 \pm 4.88$	4.29	<0.001			
Wind - Animal	$-3.198 \pm 5.50$	-0.58	0.9373			
Wind - Short-distance	$-5.24 \pm 5.43$	-0.96	0.7682			
Short-distance - Animal	$2.04 \pm 5.28$	0.39	0.9802			
	Biogeographic origin (***	)				
Eurasia - Mediterranean	$-0.21 \pm 5.32$	-0.04	1.000			
Eurasia - Oceania	42.10 ± 6.37	6.60	<0.001			
Eurasia - Southern Africa	37.12 ± 6.17	6.02	<0.001			
Eurasia - Tropical	$26.76 \pm 4.85$	5.51	<0.001			
Mediterranean - Oceania	42.31 ± 7.44	5.68	<0.001			
Mediterranean - Southern Africa	37.33 ± 7.27	5.13	<0.001			
Mediterranean - Tropical	$26.97\pm6.19$	4.35	<0.001			
Oceania - Southern Africa	$-4.99 \pm 8.07$	-0.62	0.971			
Oceania - Tropical	-15.35 ± 7.11	-2.16	0.190			
Tropical - Southern Africa	$-10.36 \pm 6.93$	-1.49	0.555			

Table S5 Tukey tests for multiple comparisons of relationships between the average
 area, isolation and distance from urban areas (urban) of occupied islands and time since first
 naturalization across categories of growth forms, dispersal modes and biogeographic origins.
 Columns are island characteristics and t-values of Tukey tests. α was set at 0.003 after

626 applying a Bonferroni correction for multiple comparisons.

Category	Area	Isolation	Urban				
Growth forms							
Graminoids - Forbs	-0.50	0.26	1.90				
Woody - Forbs	-0.80	2.16	-0.208				
Woody - Graminoids	-0.01	1.08	-1.880				
	Dispersal mo	des					
Unspecialized - Short-distance	-0.93	1.08	1.182				
Unspecialized - Animal	1.77	-2.00	2.348				
Unspecialized - Wind	0.08	0.98	-0.867				
Wind - Animal	1.68	-0.74	1.262				
Wind - Short-distance	-0.75	2.00	0.175				
Short-distance - Animal	2.72	-3.00	1.237				
	Biogeographic o	rigins					
Eurasia - Mediterranean	-0.23	-1.72	1.86				
Eurasia - Oceania	1.42	1.75	-0.07				
Eurasia - Southern Africa	-0.10	-1.80	1.19				
Eurasia - Tropical	0.74	0.91	0.78				
Mediterranean - Oceania	1.31	2.86	-1.68				
Mediterranean - Southern Africa	0.12	0.04	-0.64				
Mediterranean - Tropical	0.76	2.31	-1.23				
Oceania - Southern Africa	-1.25	-2.98	1.08				
Oceania - Tropical	-0.76	-0.94	0.69				
Tropical - Southern Africa	0.66	2.42	-0.54				

**Table S6**Contingency table of biogeographical origin and growth forms (top) and

629 dispersal modes (bottom).

	Graminoids	Forbs	Woody species	Total
Eurasia	53	200	45	298
Mediterranean	8	63	19	90
Oceania	13	12	38	63
South Africa	6	49	9	64
Tropical	20	72	19	111
Total	100	396	130	626

	Wind-dispersed	Animal-dispersed	Unspecialized	Short-distance	Total
Eurasia	47	66	132	58	303
Mediterranean	11	15	35	33	94
Oceania	15	23	12	11	61
South Africa	20	3	27	14	64
Tropical	23	35	38	23	119
Total	116	142	244	139	641