# 1 Habitat suitability maps for Australian flora and fauna under CMIP6 climate

- 2 scenarios
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- 17 Abstract

18 **Background**: Spatial information about the location and suitability of areas for native plant and animal

- 19 species under different climate futures is an important input to land use and conservation planning and
- 20 management. Australia, renowned for its abundant species diversity and endemism, often relies on

21 modelled data to assess species distributions due to the country's vast size and the challenges associated

22 with conducting on-ground surveys on such a large scale. Modelled habitat suitability maps use

23 information about known occurrences of species and predict suitable areas for species using climate, soil

- 24 and landscape information. **Results**: Using MaxEnt, we produced Australia-wide habitat suitability maps
- under RCP2.6-SSP1, RCP4.5-SSP2, RCP7.0-SSP3 and RCP8.5-SSP5 climate futures for 1,382 terrestrial
- 26 vertebrates and 9,251 vascular plants at 5km<sup>2</sup> for open access. This represents 60% of all Australian

27 mammal species, 77% of amphibian species, 50% of reptile species, 71% of bird species and 44% of

28 vascular plant species. We also include tabular data which includes summaries of total quality-weighted

- 29 habitat area of species under different climate scenarios and time periods. Conclusions: These habitat
- 30 suitability maps can be used as input data for landscape and conservation planning or species
- 31 management, particularly under different climate change scenarios in Australia.

#### 32 Keywords

Atlas of Living Australia, CliMAS, bioclimatic variables, species distribution, species range, amphibian,
 reptile, bird, mammal, vascular plant, natural capital accounting, biodiversity, biodiversity data, climate
 suitability, suitability mapping, Maxent, spatial conservation planning, climate change, WorldClim,
 habitat suitability, niche

#### 37 Data Description

#### 38 Introduction

39 Rich spatial and temporal information about the effect of climatic and environmental change on species 40 distributions is necessary to ensure robust species management and conservation policy more broadly 41 (Bryan et al., 2014; Hanson et al., 2019; Leclère et al., 2020; Summers et al., 2012). Identifying areas 42 where species occur now, as well as areas which may be suitable in the future, is a crucial aspect of 43 decision making under uncertainty (Summers et al., 2012). The availability of resources for conservation, 44 including financial, staffing and land availability, is limited and exacerbates the challenge of conservation 45 planning during climate change (Hanson et al., 2019). These constraints have sparked the need for more 46 strategic landscape and conservation planning methods, such as spatial prioritisation, to identify the most 47 effective conservation solutions (Tulloch et al., 2015). Spatial information on where species are now and 48 where suitable areas may be in the future is the foundation of efficient planning for conservation action, 49 particularly in areas where local conditions are more sensitive to climate change (Summers et al., 2012). 50

Australia is a hyper-diverse country with high levels of species endemism (Chapman, 2009; Coleman,
2016). Unfortunately, Australia also has some of the highest recorded numbers of contemporary
extinctions worldwide and more than 1900 species and ecological communities are even now under threat
(Woinarski et al., 2019; Australian Government Department of Agriculture and the Environment, 2021).
Given the extensive and severe range and population declines of many threatened species (Bergstrom et

al., 2021; Kearney et al., 2018; Woinarski et al., 2019), many more species are also predicted to have a
high risk of extinction in the future (Garnett et al., 2022). To ensure the conservation of Australia's
unique biodiversity, identifying and protecting important areas for species such as climate refugia is key
to planning for resilience and adaptive capacity (Reside et al., 2014). To fulfill this task, underlying data
on species location and the habitat suitability of areas for species under different climate futures is
required.

62

63 There are many ways to assess suitable areas for species, and one popular approach is to use the 64 maximum entropy method (henceforth, MaxEnt). MaxEnt is a niche-based general-purpose machine 65 learning method with a simple and precise mathematical formulation which is particularly well-suited for 66 species distribution modelling with presence-only data (Phillips et al., 2006). Generating MaxEnt models 67 for individual species at continental scales presents challenges around the processing and storage of large 68 volumes of data. Graham et al. (2019) developed a comprehensive spatial dataset of 1,872 terrestrial and 69 freshwater vertebrate species distributions using the Intergovernmental Panel on Climate Change's 70 (IPCC) AR4 Coupled Model Intercomparison Project 3 (CMIP3) future climate projections (Meehl et al., 71 2007) and made them freely available through a web-based portal known as 'CliMAS'. Although the 72 CliMAS models led to many applied outcomes (Maxwell et al., 2019; Ward et al., 2022), the website was 73 retired in 2020, in recognition of the fact that there have been two major updates by the IPCC and the 74 current projections are based on CMIP6. For conservation planning to progress, an improved and enlarged 75 suite of freely available spatial data, based on up-to-date climate projections and extended for a much 76 broader range of species including vascular plants, is needed.

77

We developed habitat suitability maps for Australian flora and fauna under different climate futures using
a MaxEnt approach. We produced freely accessible Australia-wide habitat suitability maps for 1,441
terrestrial vertebrates and 9,251 vascular plants. This represents 60% of all Australian mammal species,
77% of amphibian species, 50% of reptile species, 71% of bird species and 44% of vascular plant species.

We fit these models using 7 bioclimatic variables and 11 soil and landscape variables under 4 climate scenarios, 8 GCMs and 1 ensemble average, and 5 time periods. These habitat suitability maps are best used as input data to represent species or biodiversity values for conservation planning, particularly under different climate change scenarios in Australia.

## 86 Methods

The workflow for this study was adapted from the CliMAS project (Graham et al., 2019) (Figure 1). The first step involved compiling and collecting the input data which consisted of occurrence point data as well as climate, soil and landscape variables. We then used MaxEnt to fit models of habitat suitability using climate, soil and landscape variables. We conducted a variable selection procedure which considered the statistical and ecological importance of variables to refine the predictor variables as well as validating the models. We then used the lambda files produced in the model fitting step to project species habitat suitability under future climate scenarios and time periods.



94

- 95 Figure 1 Workflow of the MaxEnt modelling procedure. Input data is represented as green, variable selection procedure is
- 96 represented as purple, MaxEnt modelling procedure is represented as grey and the output files are represented as orange.

#### 97 Input data

#### 98 Species occurrence points

99 Species occurrence records which were used to fit the historical climate models were sourced from the 100 Australian Atlas of Living Australia (ALA) (Atlas of Living Australia, 2012), the Queensland Museum, 101 and CSIRO. Vascular plant occurrence point data were acquired through from the Queensland Museum. 102 Vertebrate species occurrence were records acquired through ALA went through an additional data 103 cleaning process prior to modelling (see Graham et al., 2019). We used the points originally applied in the 104 CliMAS project as of 2012 for vertebrates, and the vascular plant point compiled but never modelled with 105 for the CliMAS project. Throughout these sources we obtained occurrence point data for 197 mammals 106 (60% coverage), 523 birds (71% coverage), 530 reptiles (50%), 191 amphibians (77%) and 9,251 vascular 107 plants (44% coverage). MaxEnt uses background sample points as pseudoabsences and recommends the 108 use of target groups in sample selection (Philips et al. 2009). Each background file contained between 109 60,000 to 250,000 points depending on the taxonomic group, in which MaxEnt takes a subsample of 110 10,000 points.

#### 111 Environmental variables

112 We used a combination of bioclimatic, soil and landscape variables as predictors to fit the MaxEnt models. For the climate variables, we downloaded spatial data at a 5km<sup>2</sup> resolution on historical and 113 114 future CMIP6 modelled bioclimatic variables through the WorldClim database (www.worldclim.org, 115 accessed on September 2020). Bioclimatic variables summarise monthly temperature and rainfall values 116 into 19 more biologically meaningful variables. Bioclimatic variables were downloaded for eight global 117 climate models (GCMs): BCC-CSM2-MR, CNRM-CM6-1, CNRM-ESM2-1, CanESM5, GFDL-ESM4, 118 IPSL-CM6A-LR, MIROC-ES2L, MIROC6, MRI-ESM2-0, for four shared socioeconomic (SSP) and 119 representative concentration pathway (RCP) combinations: RCP2.6-SSP1, RCP4.5-SSP2, RCP7.0-SSP3 120 and RCP8.5-SSP5 and 5 time-periods (1990, 2030, 2050, 2070 and 2090). As we did not have access to

the following two files: IPSL-CM6A-LR SSP2-4.5 2030 and MRI-ESM2-0 SSP5-8.5 2030, we linearly
 interpolated values. All climate scenarios, bioclimatic variables were clipped to the extent of Australia

123 prior to modelling.

- 124
- 125 We downloaded 15 environmental variables from the Soil and Landscape Grid of Australia database
- 126 (https://www.csiro.au/en/research/natural-environment/land/soil-and-landscape-grid-of-australia, accessed

127 on Sep 2021) to use as environmental predictors of habitat suitability. Additionally, we downloaded the

128 Interim Biogeographic Regionalisation for Australia (IBRA) as an indication of the inherent spatial

129 differences in biome across Australia. Soil and landscape variables were clipped and masked to the extent

130 of Australia and scaled to the same resolution as the bioclimatic data.

#### 131 MaxEnt modelling procedure

#### 132 Model fitting

All habitat suitability models were fit in MaxEnt Version 3.4.1. Maxent models were first run with 10 replicates validated using a cross validation method to train the model and to compute model validation statistics. At this stage, habitat suitability values are calculated as values between 0 and 1 with no threshold applied and were later converted to values between 0 and 100. Important outputs of the MaxEnt modelling procedure include a .csv file containing statistical information to inform variable selection and model validation as well as the 'lambdas file', which is a text file containing the regression coefficients, or lambdas, fit by MaxEnt during modelling.

#### 140 Variable selection

141 The variables included in the final MaxEnt model runs were informed by analysing the variable

142 contributions and importance percentages calculated using a full MaxEnt model run, information about

143 variable complexity (Low et al., 2021), as well as ecological knowledge based on several published

144 models of terrestrial vertebrate and vascular plant climate and habitat suitability. The goal of variable

selection was to reduce the number of predictor variables from the initial 35 variables chosen as potential environmental predictors to avoid overfitting. Although Maxent is considered to be robust to multicollinearity among variables (Feng et al., 2019), including excessive numbers of predictors can affect the model's ability to make inferences outside of the training data.

149

150 We reviewed variables included within several Australian biodiversity modelling efforts of terrestrial 151 vertebrates (Graham et al., 2019), and vascular plants (Butt et al., 2013; Gallagher et al., 2019). We then 152 performed a full MaxEnt model run which included the 35 variables described in the above section, for 153 each species. We reviewed the importance of variables based on the average percent contribution and 154 percent importance values across all species. The percent contribution is a measure of the contribution of 155 each variable towards model fit after each iteration of the MaxEnt model, while the percent importance is 156 a measure of the importance of each variable towards model fit for the final MaxEnt model. We also 157 categorised bioclimatic variables based on complexity and favoured simple variables as they tended to be 158 less correlated with one another (Low et al., 2021).

159

160 This combined approach to variable selection resulted in 18 variables which moved through to the model 161 fitting stage (Table 1): 7 bioclimatic variables and 11 soil and landscape variables. All bioclimatic 162 variables selected for this study were included in CliMAS models (Graham et al., 2019) and similar 163 modelling efforts for Australian plants (Gallagher et al., 2019), and all bioclimatic variables with the 164 exception of BIO15 were considered to be simple climate variables (Low et al., 2021) (Table 1). All 165 bioclimatic variables except for BIO05 had high or moderate importance values in the full model. 166 Similarly, we included additional soil and landscape variables (Hageer et al., 2017) based on their use in 167 recent biodiversity models (Gallagher et al., 2019), and we favored soil and landscape variables that were 168 simpler.

#### Code Variable Name Contribution<sup>1</sup> **Ecological Rationale** Importance<sup>2</sup> Bioclimatic variables 8.72 18.21 Influences thermal tolerances of species. BIO1 Annual Mean Temperature 6.33 9.92 Max Temperature of BIO5 Influences upper thermal tolerances of species through extreme temperatures. Warmest Month Min Temperature of Coldest 4.30 8.66 BIO6 Influences lower thermal tolerances of species through extreme temperatures. Month 8.60 10.81 Average annual rainfall which influences water availability. BIO12 Annual Precipitation Precipitation of Wettest 17.67 7.77 Maximum rainfall in the wettest month which influences maximum water BIO13 Month availability. 14.93 8.45 BIO14 Precipitation of Driest Month Minimum rainfall in the driest month which influences minimum water availability. 12.13 13.20 Standard deviation of rainfall in the annually which influences the variation in BIO15 Precipitation Seasonality water availability. Soil and landscape variables AWC Available Water Capacity 0.94 The amount of water held by the soil for future use. 0.68 BDW Bulk Density (Whole Earth) 0.89 1.17 Soil's ability to function for structural support, water and nutrient and microbial life movement, and soil aeration. Clay 1.04 Promotes water retention and reduces air circulation in soil. CLY 0.95 Depth of Soil DES 2.00 Defines the root space and volume of soils available. 1.29 ECE Electroconductivity 3.39 5.21 Movement of nutrients within the soil which influences the availability of soil nutrients. 2.37 Elevation influences soil properties and air pressure. Elevation 1.57 elev pHc pH 5.43 Affects the amount of nutrients that are water soluble in soil. 4.30 slope Slope Relief 1.81 Influences soil properties and creates varving microclimates. 1.00 SLT Silt 2.63 2.10 Promotes water retention and creates relatively porous soil conditions. SND Sand 1.60 1.60 Promotes water drainage and air circulation in soil. SOC Organic Carbon 5.17 3.05 Promotes soil structure by providing a food source for micro-organisms.

#### 169 Table 1 Summary of the bioclimatic, soil and landscape variable selected in the final MaxEnt model.

170 <sup>1</sup>Average (mean) percent contribution in the final models for each environmental variable across all species. A measure of the contribution of each variable towards model fit 171 after each iteration of the MaxEnt model.

172 <sup>2</sup> Average (mean) percent importance in the final models for each environmental variable across all species. A measure of the importance of each variable measure depends the 173 resulting decrease in training AUC on the final MaxEnt model.

#### 174 Model validation

Once variables were selected, models were re-run, and 175 176 model performance was assessed based on the area under 177 the curve (AUC) value, with AUC values of 0.7 or below 178 indicating poor performance. This process resulted in 33 179 birds, 4 vascular plants, 1 mammal, 0 reptiles and 0 180 amphibians with AUC valus less than 0.7. The median 181 AUC across all models was 0.9714. Prior to using species 182 data, please ensure you check the AUC value which is

contained withing the maxentResults.csv file.



*Figure 2 Distribution of AUC values for species models.* 

### 184 Model projections

183

185 Using the best model selected in the model fitting procedure we projected species-level MaxEnt models

- 186 under the future climate scenarios RCP2.6-SSP1, RCP4.5-SSP2, RCP7.0-SSP3 and RCP8.5-SSP5, 8
- 187 GCMs, for 1 historical time-period (1990) and 4 future time-periods (2030, 2050, 2070, 2090) using the
- 188 lambda files produced in the model fitting step. Using the predicted habitat suitability data, we then
- 189 calculated an ensemble average (mean), minimum and maximum habitat suitability (to capture model

190 variance) across 8 GCMs for each species, climate scenario and time-period.

### 191 Geospatial calculations

192 To describe the patterns of habitat suitability across time in an accessible tabular format we calculated the

- 193 total quality-weighted sum of habitat suitability for each species under different climate scenarios at each
- 194 time period (Equation 1). We first adjusted the resolution of the rasters to  $1 \text{km}^2$ , therefore the quality-
- 195 weighted habitat area (qwHA) sum corresponds to the 'habitat area' in km<sup>2</sup>. For example, if the probable
- 196 habitat suitability in a cell is equal to 1, the cell is equivalent to 1km<sup>2</sup>, whereas if the probable habitat

suitability in a cell is equal to 0.3, the cell is equivalent to  $0.3 \text{km}^2$ . Noting that the quality-weighted habitat area is not equivalent to the realised area available for a species given ecological or land use constraints which can both influence habitat availability and suitability for species. The probability of habitat suitability (*p*) was summed across raster cells (*xy*), for each species (*j*), year (*y*) and climate scenario (*c*):

$$202 \qquad \qquad qwHA_{jyc} = \sum_{i=1}^{n} p_{jyc,xy}$$

203

(Equation 1)

To describe how the patterns of habitat suitability may have changed across space under different climate scenarios or years, we summarised raster data for each species in multiple ways. For each species (t) we calculated changes in habitat suitability (s) by subtracting future time periods and climate scenarios (yc)by historical climate niche  $(p^h)$ . Where positive values indicate areas that increase in suitability in the future and negative values indicate areas that decrease in climate suitability in the future. We provide visual representation of this information in Figure 6, and included the absolute and proportional change in habitat area in the tabular summaries provided for species:

$$s_t^{yc} = p_t^h - p_t^{yc}$$

212

219

(Equation 2)

To spatially identify important areas of climate refugia which was done for Figure 5, we multiplied the historical habitat suitability matrix by the habitat suitability in each future climate scenario and year combination. For each the cell, the probability of habitat suitability values per cell (p), for each species (t), year (y) and climate scenario (c) were multiplied by the future habitat suitability. Cell values were then divided by 100, and the resulting cell value represents climate refugia (r) between 0 to 100.

218 
$$r_t^{yc} = (p_t^h * p_t^{yc})/100$$

(Equation 3)

#### 220 **Re-use potential**

#### 221 Code availability

222 For each species, MaxEnt models were run directly from the terminal using java and bash syntax and

- 223 were ultimately executed using SLURM on a high-performance Linux-based computer cluster. Additional
- 224 modelling and geospatial analyses were processed using a shell file executed using SLURM on the
- 225 computer cluster. Data and geospatial analyses were conducted in R version 4.0.1 (R Core Team., 2020),
- key libraries include the 'tidyverse' (Wickham et al., 2019), 'sf' (Pebesma, 2018) and 'raster' (Hijmans,
- 227 2021). We used Python version 3.8.3 as well as the Geospatial Data Abstraction Library (GDAL). The
- scripts used in to generate this data is available in the GitHub repository, (see,
- 229 <u>https://github.com/CarlaBirdy/MaxEnt-habitat-models</u>).

#### 230 Dataset

231 Individual species' maps for historical and future minimum, mean and maximum ensembled habitat 232 suitability, as well as the MaxEnt lambda file and summary reports produced in this study are publicly 233 accessible for download on the open-access companion GigaDB database (which upholds the FAIR 234 principles, Wilkinson et al., 2016). This dataset includes species-level historical (1970-2000 centered on 235 1990) and the future minimum, mean and maximum habitat suitability projections for 1,382 terrestrial 236 vertebrates (182 amphibians, 487 birds, 178 mammals and 535 reptiles) and 9,251 vascular plants under 4 237 climate scenarios and 5 time periods, this data equates to 521,017 rasters that are compressed using 238 Lempel–Ziv–Welch (lzw) compression. Additionally, for each species we have included a .csv file which 239 contains the total quality-weighted habitat areas (in km<sup>2</sup>) for each species under each different climate 240 scenario and time period. We have also consolidated these tables across all species and included this 241 tabular data. A complete list of the species for which habitat suitability maps were produced can be found 242 in the companion GigaDB database.

#### 243 Spatial resolution of data

This data is presented at 5km<sup>2</sup> resolution which is aligned with the climate data used as key inputs to the 244 245 MaxEnt model. The data can be subsequently downscaled to finer resolutions, however assumptions will 246 have to be made about how habitat suitability is distributed across cells. The current resolution of this data 247 is best utilised to understand general trends across space and time. To demonstrate the resolution, we 248 present the southern cassowary (Casuarius casuarius) which is known to occur in the Wet Tropics region 249 of Queensland, Australia. Current suitable areas for the southern cassowary are predicted to occur 250 between Townsville to Cooktown, with an isolated area around the Iron Range (Figure 3). Taking the 251 most severe climate change scenario (RCP8.5 - SSP5), the environmental space for the southern 252 cassowary is predicted to reduce over time around its central habitat in the Atherton Tablelands. The 253 maps for the southern cassowary can be compared with (Graham et al., 2019) for reference.





255 Figure 3 This habitat suitability distribution is for the Southern cassowary (Casuarius casuarius) and presents its historical

- suitability projection. This zoomed in location map highlights the resolution of the data, and how the habitat suitability
- 257 distribution for the Southern cassowary is modeled over time under the RCP8.5 SSP5 scenario.

#### 258 Species-level data summary

259 The dataset includes suitability maps for species under different climate scenarios and time periods using 260 an ensemble average approach. Through the process of ensemble averaging, the minimum and maximum 261 suitability maps were also produced. These maps can be compared to understand the bounds of how 262 climate change may generally impact habitat suitability in the future. The importance of incorporating 263 multiple GCM projections can be seen by the variation among the minimum, mean, and maximum 264 suitability maps (Figure 4). For the common wallaroo (Macropus robustus), the differences between the 265 minimum, mean, and maximum suitability maps are most apparent under worsening climate scenarios. 266 Areas across the southern parts of Australia remain suitable across all three suitability maps, compared to 267 areas in the central and northern parts of their range becoming progressively less suitable. These trends are consistent with other macropod modelling studies that also suggest suitability for the common 268 269 wallaroo will track south as climate scenarios worsen (Ritchie & Bolitho, 2008). The maps for the 270 common wallaroo can also be compared with (Graham et al., 2019) for reference.



271

- 272 Figure 4 Minimum, mean, and maximum suitability value across GCMs. These habitat suitability distributions are for the
- 273 common wallaroo (Macropus robustus) for four future emission scenarios in the year 2090.

#### 274 Spatial changes over time

275 Taking this a step further, geospatial calculations can also be applied to determine the differences between years or climate scenarios. This can be conducted to identify areas of refugia (Equation 3), or the location 276 277 and magnitude of change between different time periods (Equation 2). To calculate refugia, historical and 278 future suitability maps can be multiplied together to accentuate areas in space that are suitable in both 279 time periods. To calculate changes in habitat suitability, historical and future suitability maps can be 280 subtracted from one another to accentuate areas in space that have changed in suitability across time periods. Using the snow gum (Eucalyptus pauciflora) as an example, we find refugia in the alpine region 281 282 of Australia is predicted to decline for the snow gum under worsening climate scenarios, with declines 283 being most severe in the year 2090 (Figure 5, top). Across all climate scenarios habitat suitability is 284 declining from all areas of the snow gum's range, and we did not identify areas of increases (Figure 5, 285 bottom).



Figure 5 These refugia and habitat suitability change maps are for the snow gum (Eucalyptus pauciflora). The Top Panel present climate change refugia for four future emission scenarios in the years 2030 and 2050. Darker blue on the refugia maps represent areas that have high predictive suitability historically as well in future time periods. The Bottom Panel present changes in habitat suitability for four future emission scenarios in the years 2030 and 2050. Darker orange areas indicate places that decrease in suitability compared to the previous time period, and white areas indicate no change in suitability.

#### 292 Changes in quality-weighted habitat area

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293 The dataset also includes a tabular summary of quality-weighted habitat area in km<sup>2</sup> for each species

under different climate scenarios and time periods (Equation 1 and Equation 2). The quality-weighted

habitat area values can be analysed and plotted to understand the how climate change may impact habitat area for single species or groups of species in the future (Figure 6). When this data is summarised across all species, we can show that in 2030 the distribution of change in habitat area are similar across the four climate scenarios. However, in 2090 the distribution of change in habitat area follows a different pattern across climate scenarios with progressively more species loosing progressively more habitat area as climate scenario worsens (Figure 6).



301

Figure 6 Histogram of the number of species and their relative change in quality weighted habitat area between 1990 and each
 future time period (2030, 2050, 2070, 2090).

# 304 **Discussion**

305 Spatial data on the suitability of areas for species is an important input to guide conservation planning,

306 policy and management. The objective of this paper was to develop habitat suitability maps for Australian

flora and fauna under different climate futures using a MaxEnt approach. This data has been developed in a way that is consistent across species and enables users to analyse how different climate futures may impact the habitat suitability for biodiversity more generally across Australia. This data can also be used for species-level analysis and can be a starting point for additional analyses which utilise either geospatial information or tabular information that could take into consideration additional information like land use, conservation actions or species ecology.

313

314 This spatial and tabular dataset is ideal for users that would like to understand how the habitat suitability 315 of areas for species is predicted to change over time or under different climate scenarios. Due to its 5km<sup>2</sup> 316 spatial resolution, the data is best for understanding broader spatial trends that can be integrated into 317 spatial planning (Maxwell et al., 2019), rather than more local management such as identifying specific 318 sites for translocation (Eyre et al., 2022). For example, as presented above, these maps can be combined 319 to evaluate how habitat suitability changes over time (Figure 6) or over space and time (Figure 5), which 320 can then be considered into conservation or monitoring plans in areas which are predicted to lose or gain 321 suitable areas for the species. These analyses can be conducted at a species or a taxonomic group level to 322 support conservation actions for species of interest (e.g., threatened species) or for biodiversity values 323 more generally. Spatial information about species could be directly utilised to develop management or 324 monitoring plans that consider how climate change may impact the species habitat area.

325

When using and interpreting the data contained in this data set it is important to consider the following limitations and considerations. This dataset presents the habitat suitability of areas for species under different climate scenarios and time periods. These maps are not distribution maps, rather they present habitat suitability based on climate, soil and landscape characteristics. These maps have not been thresholded nor do they consider dispersal (Graham et al., 2019), land use (Kapitza et al., 2021), biophysical capacity (Briscoe et al., 2023), or attributes that may be important for species of interest (e.g., fire or vegetation structure e.g., Eyre et al., 2022). The occurrence points used for this analysis were those 333 originally used for the CliMAS work, and the ALA data were passed through an additional rigorous 334 cleaning process. This process helped reduce the spatial bias and noise in the occurrence points (Phillips 335 et al., 2009); however, more broadly there are sampling biases that influence the distribution of 336 occurrence points, such as land tenure. To improve on the models, an integrated pathway to ALA into the 337 modelling procedure would be ideal as this would ensure up-to-date input data. However, this can also 338 come with challenges as occurrence data is required to have the same temporal resolution to the historical 339 or current climate data (i.e., 1990 in this study). MaxEnt models are prone to overfit but are also less 340 influenced by collinearity than statistical models, we tried mitigating the impacts of overfitting the 341 MaxEnt models by conducting variable selection. There are a multitude of other methods to model 342 suitability and species distributions that have their own use cases and limitations (Elith & Graham, 2009). 343 344 To spatially target conservation actions, spatial information about the location and suitability of areas for 345 species is needed. This study provides a comprehensive data set of predicted habitat suitability under 4 346 climate futures, while also incorporating the uncertainty across GCMs. We are providing a spatial and tabular data product at the Australian scale and at 5km<sup>2</sup> resolution that can be used to inform research and 347 348 decision making at local, regional and national scales. This data can be applied within strategic 349 conservation planning approaches and can be used to identify important areas for species consecration 350 (Tulloch et al., 2015). Spatial information about current and future suitable areas for species is a key 351 component of conservation planning, particularly as the impact of climate change on species and 352 biodiversity is uncertain. 353 354 Data availability 355 All spatial and tabular data are freely accessible in the companion GigaDB repository. 356

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363	
364	Conflict of interest
365	The authors declare no conflicts of interest.
366	References
367	Australian Government Department of Agriculture and the Environment. (2021). Species Profile and
368	Threats Database (SPRAT). http://www.environment.gov.au/cgi-bin/sprat/public/sprat.pl
369	Bergstrom, D. M., Wienecke, B. C., van den Hoff, J., Hughes, L., Lindenmayer, D. B., Ainsworth, T. D.,
370	Baker, C. M., Bland, L., Bowman, D. M. J. S., Brooks, S. T., Canadell, J. G., Constable, A. J.,
371	Dafforn, K. A., Depledge, M. H., Dickson, C. R., Duke, N. C., Helmstedt, K. J., Holz, A., Johnson,
372	C. R., Shaw, J. D. (2021). Combating ecosystem collapse from the tropics to the Antarctic.
373	Global Change Biology, 27(9), 1692–1703. https://doi.org/10.1111/gcb.15539
374	Briscoe, N. J., Morris, S. D., Mathewson, P. D., Buckley, L. B., Jusup, M., Levy, O., Maclean, I. M. D.,
375	Pincebourde, S., Riddell, E. A., Roberts, J. A., Schouten, R., Sears, M. W., & Kearney, M. R.
376	(2023). Mechanistic forecasts of species responses to climate change: The promise of biophysical
377	ecology. In Global Change Biology (Vol. 29, Issue 6, pp. 1451–1470). John Wiley and Sons Inc.
378	https://doi.org/10.1111/gcb.16557
379	Bryan, B. A., Nolan, M., Harwood, T. D., Connor, J. D., Navarro-Garcia, J., King, D., Summers, D. M.,
380	Newth, D., Cai, Y., Grigg, N., Harman, I., Crossman, N. D., Grundy, M. J., Finnigan, J. J., Ferrier,
381	S., Williams, K. J., Wilson, K. A., Law, E. A., & Hatfield-Dodds, S. (2014). Supply of carbon

- 382 sequestration and biodiversity services from Australia's agricultural land under global change.
- 383 Global Environmental Change, 28, 166–181. https://doi.org/10.1016/j.gloenvcha.2014.06.013
- Butt, N., Pollock, L. J., & Mcalpine, C. A. (2013). Eucalypts face increasing climate stress. *Ecology and*

385 *Evolution*, *3*(15), 5011–5022. https://doi.org/10.1002/ece3.873

- Chapman, A. D. (2009). Numbers of Living Species in Australia and the World. Biodiversity Information
   Services.
- Coleman, S. (2016). Australia state of the environment 2016: built environment. In *independent report to the Australian Government Minister for the Environment and Energy*.
- 390 https://doi.org/10.30547/mediascope.1.2018.6
- 391 Elith, J., & Graham, C. H. (2009). Do they? How do they? WHY do they differ? on finding reasons for
- differing performances of species distribution models. *Ecography*, *32*(1), 66–77.
- 393 https://doi.org/10.1111/j.1600-0587.2008.05505.x
- 394 Eyre, A. C., Briscoe, N. J., Harley, D. K. P., Lumsden, L. F., McComb, L. B., & Lentini, P. E. (2022).
- 395 Using species distribution models and decision tools to direct surveys and identify potential
- translocation sites for a critically endangered species. *Diversity and Distributions*, 28(4), 700–711.
- 397 https://doi.org/10.1111/ddi.13469
- 398 Feng, X., Park, D. S., Liang, Y., Pandey, R., & Papeş, M. (2019). Collinearity in ecological niche
- 399 modeling: Confusions and challenges. *Ecology and Evolution*, *9*(18), 10365–10376.
- 400 https://doi.org/10.1002/ece3.5555
- Gallagher, R. V., Allen, S., & Wright, I. J. (2019). Safety margins and adaptive capacity of vegetation to
  climate change. *Scientific Reports*, 9(1), 1–11. https://doi.org/10.1038/s41598-019-44483-x
- 403 Garnett, S., Hayward-Brown, B. K., Kopf, R. K., Woinarski, J. C. Z., Cameron, K. A., Chapple, D. G.,
- 404 Copley, P., Fisher, A., Gillespie, G., Latch, P., Legge, S., Lintermans, M., Moorrees, A., Page, M.,
- 405 Renwick, J., Birrell, J., Kelly, D., & Geyle, H. M. (2022). Australia's most imperilled vertebrates.
- 406 *Biological Conservation*, 270, 109561. https://doi.org/10.1016/j.biocon.2022.109561

- 407 Graham, E. M., Reside, A. E., Atkinson, I., Baird, D., Hodgson, L., James, C. S., & VanDerWal, J. J.
- 408 (2019). Climate change and biodiversity in Australia: a systematic modelling approach to
- 409 nationwide species distributions. Australasian Journal of Environmental Management, 26(2), 112–
- 410 123. https://doi.org/10.1080/14486563.2019.1599742
- 411 Hageer, Y., Esperón-Rodríguez, M., Baumgartner, J. B., & Beaumont, L. J. (2017). Climate, soil or both?
- 412 Which variables are better predictors of the distributions of Australian shrub species? *PeerJ*,
- 413 2017(6). https://doi.org/10.7717/peerj.3446
- 414 Hanson, J. O., Schuster, R., Strimas-Mackey, M., & Bennett, J. R. (2019). Optimality in prioritizing
- 415 conservation projects. *Methods in Ecology and Evolution*, *10*(10), 1655–1663.
- 416 https://doi.org/10.1111/2041-210X.13264
- 417 Hijmans, R. J. (2021). raster: Geographic Data Analysis and Modeling (R package version 3.5-9).
- 418 https://CRAN.R-project.org/package=raster.
- 419 Kapitza, S., Van Ha, P., Kompas, T., Golding, N., Cadenhead, N. C. R., Bal, P., & Wintle, B. A. (2021).
- 420 Assessing biophysical and socio-economic impacts of climate change on regional avian biodiversity.

421 Scientific Reports, 11(1). https://doi.org/10.1038/s41598-021-82474-z

- 422 Kearney, S. G., Cawardine, J., Reside, A. E., Fisher, D. O., Maron, M., Doherty, T. S., Legge, S., Silcock,
- 423 J., Woinarski, J. C. Z., Garnett, S. T., Wintle, B. A., & Watson, J. E. M. (2018). The threats to
- 424 Australia's imperilled species and implications for a national conservation response. *Pacific*
- 425 Conservation Biology, Chapman 2009. https://doi.org/10.1071/PC18024
- 426 Leclère, D., Obersteiner, M., Barrett, M., Butchart, S. H. M., Chaudhary, A., De Palma, A., DeClerck, F.
- 427 A. J., Di Marco, M., Doelman, J. C., Dürauer, M., Freeman, R., Harfoot, M., Hasegawa, T.,
- 428 Hellweg, S., Hilbers, J. P., Hill, S. L. L., Humpenöder, F., Jennings, N., Krisztin, T., ... Young, L.
- 429 (2020). Bending the curve of terrestrial biodiversity needs an integrated strategy. *Nature*,
- 430 585(October 2018). https://doi.org/10.1038/s41586-020-2705-y

- 431 Low, B. W., Zeng, Y., Tan, H. H., & Yeo, D. C. J. (2021). Predictor complexity and feature selection
- 432 affect Maxent model transferability: Evidence from global freshwater invasive species. *Diversity*

433 *and Distributions*, 27(3), 497–511. https://doi.org/10.1111/ddi.13211

- 434 Maxwell, S. L., Reside, A., Trezise, J., McAlpine, C. A., & Watson, J. E. (2019). Retention and
- 435 restoration priorities for climate adaptation in a multi-use landscape. *Global Ecology and*
- 436 *Conservation*, *18*, e00649. https://doi.org/10.1016/j.gecco.2019.e00649
- 437 Meehl, G. A., Covey, C., Delworth, T., Latif, M., Mcavaney, B., Mitchell, J. F. B., Stouffer, R. J., &
- Taylor, K. E. (2007). *The WCRP CMIP3 Multi-model Dataset: A New Era in Climate Change Research.* http://cera-www.dkrz.de/IPCC DDC/
- Pebesma, E. (2018). Simple Features for R: Standardized Support for Spatial Vector Data. *The R Journal*, *10*(1), 439–446.
- Phillips, S. B., Aneja, V. P., Kang, D., & Arya, S. P. (2006). Maximum entropy modeling of species
  geographic distributions. *Ecological Modelling*, 6(2–3), 231–252.
- 444 https://doi.org/10.1016/j.ecolmodel.2005.03.026
- 445 Phillips, S. J., Dudík, M., Elith, J., Graham, C. H., Lehmann, A., Leathwick, J., & Ferrier, S. (2009).
- 446 Sample selection bias and presence-only distribution models: Implications for background and
- 447 pseudo-absence data. *Ecological Applications*, 19(1), 181–197. https://doi.org/10.1890/07-2153.1
- R Core Team. (2020). *R: A language and environment for statistical computing* (4.0.1). R Foundation for
  Statistical Computing.
- 450 Reside, A. E., Welbergen, J. A., Phillips, B. L., Wardell-Johnson, G. W., Keppel, G., Ferrier, S.,
- Williams, S. E., & Vanderwal, J. (2014). Characteristics of climate change refugia for Australian
  biodiversity. *Austral Ecology*, *39*(8), 887–897. https://doi.org/10.1111/aec.12146
- 453 Ritchie, E. G., & Bolitho, E. E. (2008). Australia's savanna herbivores: Bioclimatic distributions and an
- 454 assessment of the potential impact of regional climate change. *Physiological and Biochemical*
- 455 Zoology, 81(6), 880–890. https://doi.org/10.1086/588171

- 456 Summers, D. M., Bryan, B. A., Crossman, N. D., & Meyer, W. S. (2012). Species vulnerability to climate
- 457 change: Impacts on spatial conservation priorities and species representation. *Global Change*

458 Biology, 18(7), 2335–2348. https://doi.org/10.1111/j.1365-2486.2012.02700.x

- 459 Tulloch, V. J. D., Tulloch, A. I. T., Visconti, P., Halpern, B. S., Watson, J. E. M., Evans, M. C.,
- 460 Auerbach, N. A., Barnes, M., Beger, M., Chadès, I., Giakoumi, S., McDonald-Madden, E., Murray,
- 461 N. J., Ringma, J., & Possingham, H. P. (2015). Why do We map threats? Linking threat mapping
- 462 with actions to make better conservation decisions. *Frontiers in Ecology and the Environment*,
- 463 *13*(2), 91–99. https://doi.org/10.1890/140022
- 464 Ward, M., Tulloch, A., Stewart, R., Possingham, H. P., Legge, S., Gallagher, R. V., Graham, E. M.,
- 465 Southwell, D., Keith, D., Dixon, K., Yong, C., Carwardine, J., Cronin, T., Reside, A. E., & Watson,
- J. E. M. (2022). Restoring habitat for fire-impacted species' across degraded Australian landscapes.
   *Environmental Research Letters*, *17*(8). https://doi.org/10.1088/1748-9326/ac83da
- 468 Wickham, H., Averick, M., Bryan, J., Chang, W., McGowan, L. D., François, R., Grolemund, G., Hayes,
- 469 A., Henry, L., Hester, J., Kuhn, M., Lin Pedersen, T., Miller, E., Milton Bache, S., Müller, K.,
- 470 Ooms, J., Robinson, D., Paige Seidel, D., Spinu, V., ... Yutani, H. (2019). Welcome to the
- 471 tidyverse. Journal of Open Source Software, 4(43), 1686.
- 472 Wilkinson, M. D., Dumontier, M., Aalbersberg, Ij. J., Appleton, G., Axton, M., Baak, A., Blomberg, N.,
- 473 Boiten, J. W., da Silva Santos, L. B., Bourne, P. E., Bouwman, J., Brookes, A. J., Clark, T., Crosas,
- 474 M., Dillo, I., Dumon, O., Edmunds, S., Evelo, C. T., Finkers, R., ... Mons, B. (2016). Comment:
- 475 The FAIR Guiding Principles for scientific data management and stewardship. *Scientific Data*, *3*.
- 476 https://doi.org/10.1038/sdata.2016.18
- 477 Woinarski, J., Braby, M. F., Burbidge, A. A., Coates, D., Garnett, S. T., Fensham, R. J., Legge, S. M.,
- 478 McKenzie, N. L., Silcock, J. L., & Murphy, B. P. (2019). Reading the black book: The number,
- 479 timing, distribution and causes of listed extinctions in Australia. *Biological Conservation*,
- 480 *239*(November), 108261. https://doi.org/10.1016/j.biocon.2019.108261

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