

1 **Measuring historical pollution: natural history collections as tools for public health and**
2 **environmental justice research**

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18

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20 **Key words:** temporal sampling, spatial sampling, environmental justice, health disparities,
21 exposure risk, natural history collections, One Health

22 **ABSTRACT**

23 ***Background***

24 Through the industrial era, environmental pollution has been unevenly distributed in the
25 environment, disproportionately impacting disenfranchised communities. The distribution of
26 pollution is thus a question of environmental justice and public health that requires policy
27 solutions. However, we lack robust quantitative data on pollutants for many locations and time
28 periods because environmental monitoring is largely reactive—i.e., pollutants are monitored
29 only after they are recognized as harmful and circulating in the environment at elevated levels.
30 Without comprehensive historical pollution data, it is difficult to understand the full,
31 intergenerational determinants and consequences of pollutants on environmental and human
32 health. Here, we promote the use of biological specimens in natural history collections as an
33 underutilized source of quantitative pollution data for tracking environmental change over nearly
34 two centuries and for informing justice-centered policy solutions.

35 ***Objectives***

36 First, we discuss the need for quantitative pollution data in environmental research and its
37 implications for public health and policy. We then examine the capacity of biological specimens
38 to serve as tools for tracking environmental pollutants through space and time. We then present
39 a framework for how pollution datasets from specimens can be paired with spatially and
40 temporally matched human health datasets to inform and evaluate public policy. Finally, we
41 identify challenges and research directions associated with the use of quantitative pollution
42 datasets.

43 ***Discussion***

44 Biological specimens present a unique opportunity to fill critical gaps in the environmental
45 record—and to generate historical data for emerging pollutants—to engage public health and
46 policy questions that we have previously lacked data to address. This work demands diverse
47 expertise and partnerships across the sciences, and between researchers and communities
48 affected by pollution, to connect data generated from specimens with urgent questions about
49 environmental health and justice, and to find solutions to some of the most pressing
50 environmental challenges of the twenty-first century.

51 **MAIN TEXT**

52 A key feature of environmental pollution is that it is unevenly distributed across space and time,
53 often with marginalized communities experiencing higher rates of harmful exposure (United
54 Church of Christ Commission for Racial Justice 1987, Bullard 2008, Mohai et al. 2009, Hill 2020,
55 Goldsmith and Bell 2021, Tessum et al. 2021). How environmental pollution is (and has been)
56 distributed—and its unequal impacts on different human populations—is thus a question tied to
57 issues of historical inequality, environmental racism, environmental justice, and public health
58 and policy more broadly (Mohai et al. 2009). The ability to link pollution exposures to public
59 health and policy decisions, however, is constrained by a lack of quantitative spatial and
60 temporal data on pollution concentrations in the environment (Institute of Medicine 1999, Payne-
61 Sturges and Gee 2006, Mohai et al. 2009, Wang et al. 2019).

62 Combining spatial and temporal data is necessary for understanding the uneven impacts of past
63 and present pollution. For example, Colmer et al. (2020) show that significant disparities in air
64 pollution exposure have persisted across communities in the continental U.S. over the last four
65 decades, even as overall concentrations of air pollution have dropped across this period. While
66 this study used one of the most comprehensive pollution datasets to date, including 36 years of
67 air pollution data, this time frame does not reflect the full scope of many pollutants' impacts,
68 circulation, and persistence in the environment. Many health outcomes associated with pollution
69 manifest later in life as a product of long-term exposure. Thus, to fully understand the links
70 between human health and pollution requires long-term data on when and where pollutants
71 occur. Without robust spatial and temporal data of historical pollution levels it remains difficult to
72 understand how pollution loads have changed (and are changing) in the environment, which is
73 consequential for developing policies that effectively redress past harm and prevent widening
74 health disparities (Arcaya and Figueroa 2017). Here, we propose that the missing environmental
75 data necessary for understanding the impacts and extent of pollution, as well as generating
76 historical data for emerging pollutants, may come from an underutilized data source: biological
77 specimens in natural history collections.

78 Over the last 200 years, there have been significant public and private efforts to collect plants
79 and animals from around the world for biological research. As a result, hundreds of millions of
80 specimens are now stored in natural history collections, where they are primarily used to
81 catalogue the world's biodiversity and understand its origins (National Academies of Sciences,
82 Engineering, and Medicine 2021). Incidentally, these specimens also function as *time capsules*,
83 preserving pieces of the environments they inhabited (e.g., Figure 1). These specimens offer
84 researchers material records of past environments that can be used to quantify spatiotemporal
85 changes in environmental pollution. For example, in a textbook example that led to key
86 environmental legislation in the U.S. (and a focus of Rachel Carson's book *Silent Spring*),
87 researchers analyzed a time series of bird eggs from natural history collections, uncovering how
88 the insecticide DDT was responsible for eggshell thinning and declines in bird populations in the
89 U.S. and U.K. (Ratcliffe 1967, Hickey and Andersen 1968). Bird and fish specimens have also
90 been used to track the rise of heavy metals and microplastics in terrestrial and aquatic
91 environments (e.g. Vo et al. 2011, Bond et al. 2015, Hou et al. 2021, Ilechukwu et al. 2023), and

92 bird specimens have been used to reconstruct over a century of atmospheric pollution across
93 the U.S. Industrial Belt (Figure 1, DuBay and Fuldner 2017). Given the sampling abundance,
94 historical breadth, and spatially explicit nature of specimen collections, they can be used to
95 understand pollution concentrations in the environment from decades before systematic efforts
96 to monitor the environment were established. Furthermore, the groundwork for using specimens
97 to link pollution to human health has already been done, with decades of research
98 demonstrating that the levels of over a hundred pollutants can be reliably estimated from
99 specimens (Table S1), including some of the most concerning pollutants for human health. This
100 capacity of specimens to be used as indicators of environmental health is implicit in GeoHealth
101 and One Health frameworks that acknowledge how geography, plant and animal health, and
102 public health are closely intertwined (American Geophysical Union, n.d.; Center for Disease
103 Control and Prevention, n.d.).

104 Here, we suggest that natural history collections have unique potential as time capsules to
105 inform human health, public policy, and environmental justice, providing opportunities to
106 reconstruct past environments and fill critical gaps in the environmental sampling record. With
107 this perspective, we first discuss the need for quantitative spatial and temporal pollution data.
108 We situate this discussion within a justice-oriented framework of environmental health, but the
109 need for quantitative historical pollution data is broadly applicable to public health and policy.
110 Second, we examine the robust history of using specimens to track and quantify environmental
111 pollution, which falls within the areas of ecotoxicology and bioindicator research. In this section,
112 we argue that bioindicator research has always been rooted in human health concerns, but few
113 studies to date have realized their full potential in linking pollution loads and exposure risk to
114 community-level health impacts and policy. Lastly, we present a framework for how natural
115 history specimens can contribute to a spatial and temporal understanding of past pollution—and
116 for emerging pollutants—that can be linked to spatially and temporally matched human datasets
117 to inform public health and policy decision-making. This work builds toward a multidisciplinary
118 approach that spans the natural, physical, and social sciences, and requires collaboration
119 between research communities and the public, to find creative and collective solutions to some
120 of our most pressing environmental and health challenges.



Figure 1: An example of how specimens function as *time capsules*, preserving pieces of the environments they inhabited (from DuBay and Fuldner [2017]). The figure compares two Field Sparrows (*Spizella pusilla*) collected within the U.S. Industrial Belt, one from 1906 and one from 1996. The specimen from 1906 is covered in black carbon particulate from coal burning in the region at that time. The lower panels show SEM micrographs of belly feathers plucked from the specimens in the upper panel. The SEM micrographs of the Field Sparrow from 1906 show aggregates of black carbon sphericals, while the feather from the 1996 specimen lacks black carbon deposition.

121 **The consequences of incomplete data on historical pollution**

122 Understanding pollution in the environment is critical for understanding human health because
123 pollution is one of the most acute forces by which environmental disparities manifest in peoples'
124 lives. In 2003, the term “One Health” was coined at a meeting hosted by the Wildlife
125 Conservation Society to capture the notion that, “the health of people is closely connected to the
126 health of animals and our shared environment” (Center for Disease Control and Prevention,
127 n.d.). To fully understand the connections between human health and the environment, we must
128 also understand the sociopolitical and economic systems that shape our environments and
129 maintain pollution in the first place (Murphy 2008, Liboiron 2021). In acknowledging these
130 connections (referred to by sociologists as the *political economy of the environment* [Rudel et al.
131 2011]), we can begin to understand human health as influenced by social structures (i.e., the
132 social determinants of health) that shape where, when, and to what extent (or in what
133 concentrations) pollution occurs and the resulting differences in exposure between populations.
134 One of the clearest examples of these connections is the impact of U.S. infrastructure and
135 housing policy that has enforced decades of residential placement, segregation, and redlining
136 across the U.S., concentrating marginalized populations in environmentally degraded and
137 polluted areas (Singer 2011, Taylor 2014, Collins et al. 2015, Castellón 2021, Lane et al. 2022).
138 Another example lies in the polluting industries themselves, which often follow a ‘path of least
139 resistance’, resulting in the placement of polluting facilities in, or adjacent to, communities that
140 have been politically and economically disenfranchised (Taylor 2014, Mohai and Saha 2015,
141 Begum et al. 2022). In the U.S., policies and industrial siting decisions disproportionately impact
142 Black, Indigenous and Latinx communities, poorer communities, LGBTQ+ communities,
143 populations experiencing greater incidence of mental health illness, and people with disabilities
144 (Payne-Sturges and Gee 2006, Taylor 2014, Liévanos 2018, Tessum et al. 2019, Goldsmith and
145 Bell 2021, Tessum et al. 2021, Chakraborty 2022). These effects also compound at the
146 intersections of different populations and identities (Homan et al. 2021). Within these
147 subpopulations there are higher incidences of cardiovascular disease, cancer, respiratory illness
148 (such as asthma), cognitive disorders, and preterm birth, among others adverse health impacts
149 (Payne-Sturges and Gee 2006, Culhane and Goldenberg 2011, Collins et al. 2017). Data on the
150 environmental conditions that communities have experienced in the context of social structures
151 are thus necessary for informing efforts to redress past and ongoing health inequities and for
152 mitigating future inequity.

153 A key shortcoming of efforts to measure pollutants is that environmental monitoring generally
154 begins only after pollutants have reached critical levels in the environment and have resulted in
155 observable effects on human or other animal populations. This time lag in monitoring means
156 that we often lack sufficient retrospective pollution data, which in turn skews our historical
157 understanding of the buildup and life cycle of pollutants in the environment. For example, in
158 1955 the U.S. passed its first federal air pollution legislation, the Air Pollution Control Act, in
159 response to decades of research and activism to reduce atmospheric black carbon pollution
160 (also known as soot) from coal burning (Stradling 1999). The Air Pollution Control Act
161 established the first national network to monitor air quality, but black carbon pollution in the U.S.
162 had already peaked decades earlier, and by 1955, levels of black carbon pollution were

163 declining (Stradling 1999, Bond et al. 2007, DuBay and Fuldner 2017). Environmental
 164 monitoring—as it has been deployed—is thus characteristically reactive, resulting in an
 165 incomplete record of pollutants in the environment.

166 While some spatially- and temporally-explicit datasets currently exist that have been used to
 167 understand the origins and impacts of pollutants (e.g. Colmer et al. 2020), current pollution
 168 datasets often insufficiently capture: (1) the time scales at which pollutants impact, circulate in,
 169 and persist in the environment, (2) the joint spatial and temporal heterogeneity of pollution, and
 170 (3) the co-production and co-distribution of multiple pollutants. A lack of spatially and temporally
 171 paired data limits our ability to answer longstanding and emerging questions that have been
 172 identified by environmental health researchers, some of which we highlight in Table 1.

173 **Table 1.** Critical questions that have been identified by environmental health and justice researchers. The
 174 references provide examples of where each question has been identified.

Critical questions about environment health and justice	References
<i>Who is (and has been) exposed to which pollutants?</i>	Brulle and Pellow 2006 Mohai et al. 2009
<i>What are the transgenerational impacts of exposure on health outcomes and communities?</i>	Fitzgerald et al. 1998, Anway et al. 2005 Schmidt 2013, Hoke and McDade 2014 Lombó et al. 2015, Bové et al. 2019 Shukla et al. 2019
<i>How do pollutants co-vary across the environment, and what are the impacts of multiple stressors on human health?</i>	Morello-Frosch and Shenassa 2006, US EPA 2007, Dominici et al. 2010 Billionnet et al. 2012, Levy et al. 2014 Rodgers et al. 2020
<i>What are the potential pathways of human exposure (e.g., exposure from the outdoor vs. built environment)?</i>	Wallace 1991, Rosenbaum et al. 2011 Li et al. 2021
<i>How will future land use and environmental change influence the distribution and accumulation of pollutants?</i>	Tang et al. 2005, Schilling et al. 2008
<i>How do changes in biodiversity and land-use impact the flow of pollutants through food webs and ecosystems?</i>	Maiti and Chowdhury 2013 Lepak et al. 2019, Nilsen et al. 2019
<i>What are the cumulative effects of multiple exposures and chronic low-level exposure?</i>	Di et al. 2017, Nilsen et al. 2019
<i>What are the consequences and efficacy of past policy to reduce exposure to pollutants?</i>	Crabb and Leroy 2012

175 Filling spatial and temporal gaps in the environmental record will aid in efforts to tackle the
 176 questions in Table 1, and many more. Natural history specimens—in their capacity to track
 177 environmental change—offer data and historical perspective that can work in service of health
 178 equity and environmental justice.

179 **Natural history specimens and their use as bioindicators**

180 Over the last two centuries, millions of zoological and botanical specimens have been collected
 181 throughout much of the world (Figure 2 and 3) (National Academies of Sciences, Engineering,
 182 and Medicine 2021). These specimens are housed across hundreds of natural history
 183 collections globally where they are available as a source of data. The origins of many of these

184 collections, like those in the U.S. and Europe, are inextricably tied to extractive and imperialistic
185 practices of Western science, which sever specimens from the environments they once
186 inhabited (MacKenzie 1990; Quintero Toro 2012; DuBay, Palmer Droguett, Piland. 2020;
187 Kilbourne 2020). We propose, however, that using specimens to understand past environments
188 can help reconnect specimens to the places from which they came to address place-based
189 environmental challenges.

190 With each specimen, information is recorded about when and where it was collected—dates are
191 often documented to the level of calendar day and year, and locations are usually documented
192 to the level of county (Figure 4), township, etc., but can be as specific as cross-streets or
193 georeferenced latitudes and longitudes (Figures 2 and 3B). In addition, numerous types of
194 tissues have been preserved for posterity—dried study specimens of birds, mammals, insects,
195 and plants; frozen tissues; and wet specimens (preserved in fluid) of fishes, birds, mollusks,
196 reptiles, and amphibians (Webster 2017). Although spatial and temporal gaps in sampling exist,
197 the breadth of sampling spans urban-rural gradients, land cover types, land-use gradients, and
198 geopolitical boundaries. Time-series of specimens can thus be analyzed to address questions
199 across various spatial and temporal scales of interest, for example, at local, regional, or
200 continental scales, and across seasons, decades, or centuries of environmental change.
201 Different spatial scales of potential interest can be seen in Figures 2-4.

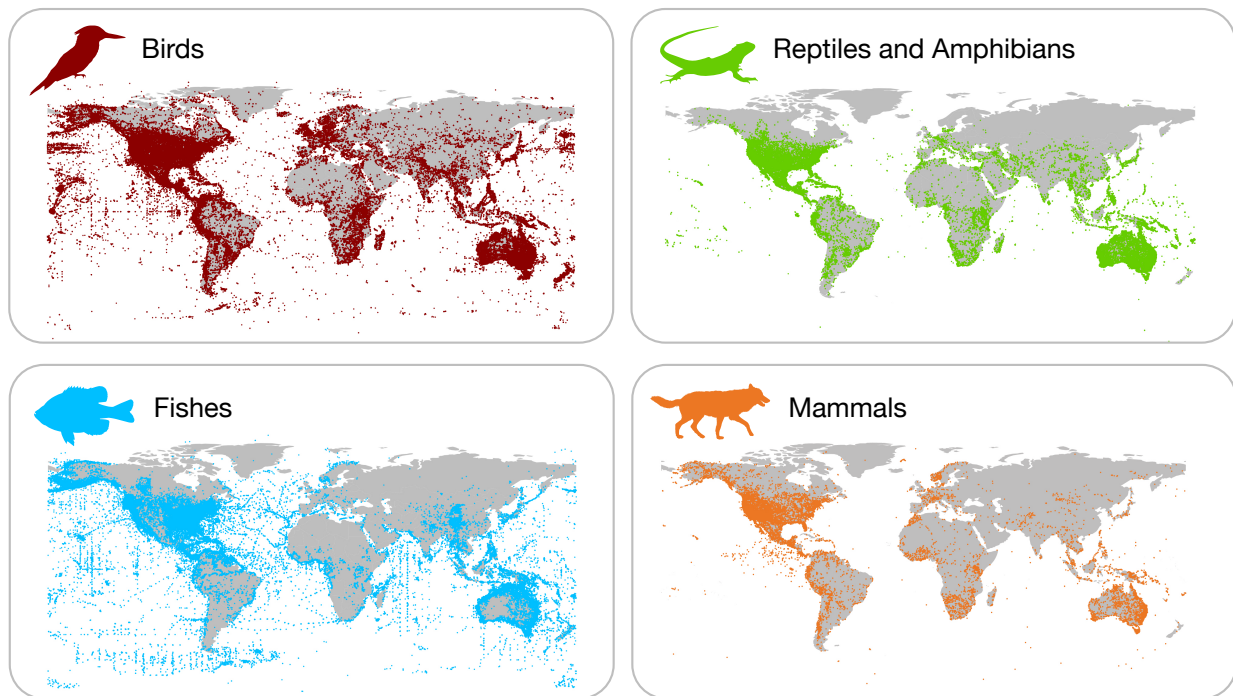


Figure 2: Global maps showing the extent of vertebrate specimens from over 125 natural history collections. Each point represents an individual georeferenced specimen, from 1850 to present. Sample sizes: Birds ($n = 2,074,502$), Reptiles and Amphibians ($n = 296,720$), Fishes ($n = 502,430$), and Mammals ($n = 68,227$). Maps were compiled from specimen records on Vertnet.org, an aggregate specimen database of global natural history collections.

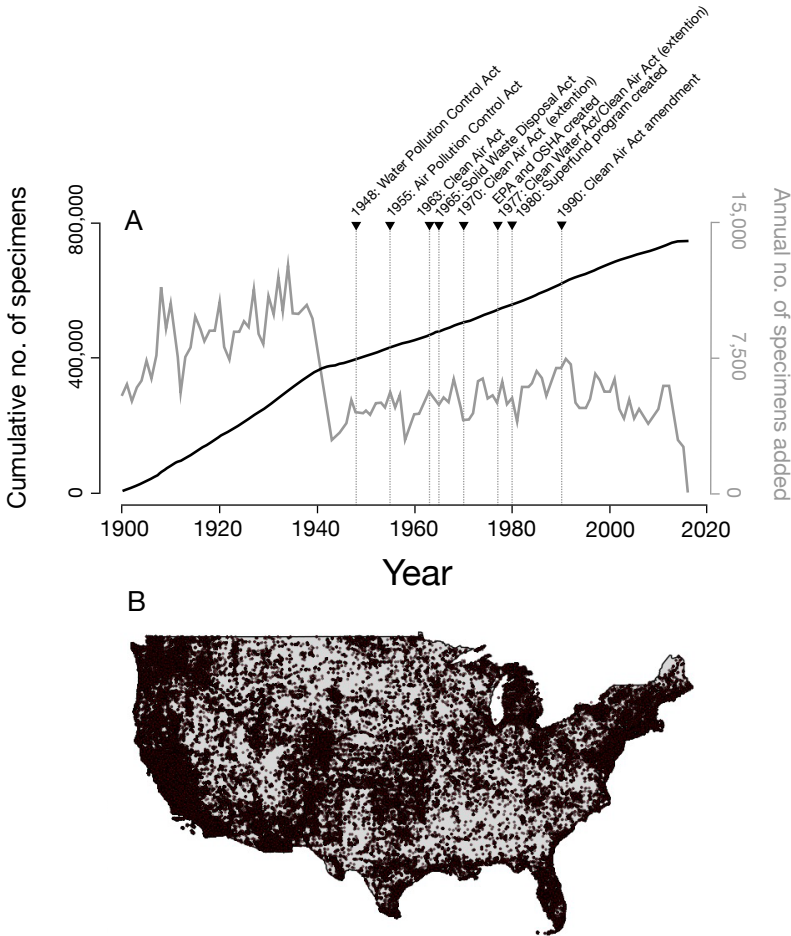


Figure 3: Vertebrate specimens from the continental U.S. from 1900 to 2016, compiled from Vertnet.org. (A) Specimen acquisition through time with a timeline of key environmental legislation in the U.S. Black line shows cumulative specimens in collections through time. Gray line shows the number of specimens entering collections each year. (B) Specimen map. Each point represents a single specimen. The sharp drop in specimens in the last five years of the time series is likely artificial, reflecting a time lag in new acquisitions being entered into museums databases.

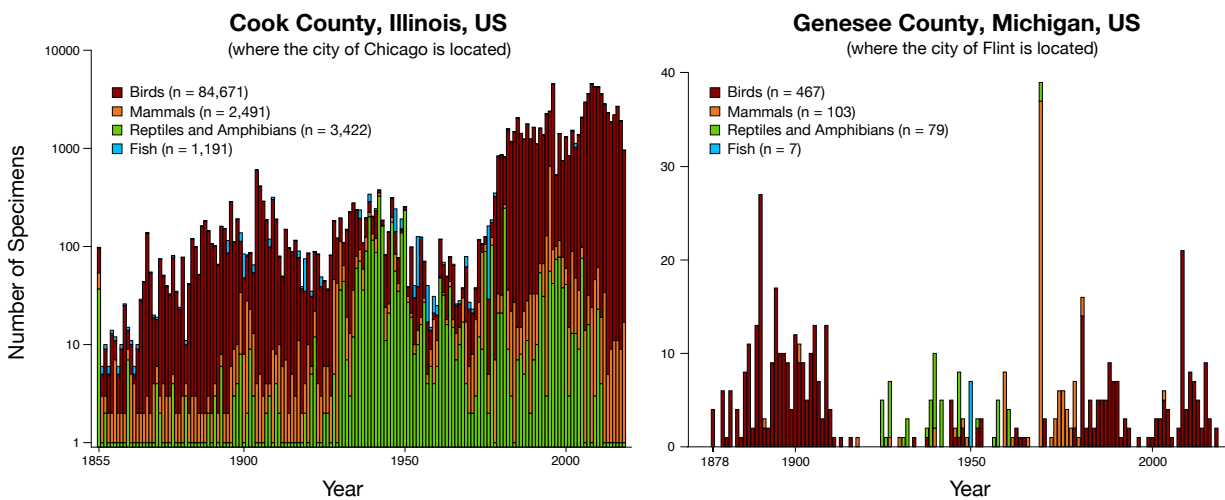


Figure 4: Examples of vertebrate specimen sample sizes through time from two U.S. counties with long histories of environmental contamination, Cook Co. and Genesee Co., which include the cities of Chicago and Flint, respectively, and their surrounding areas. These plots show the range of variation in historical sampling that exists in natural history collections at finer spatial scales. Plots were compiled from specimen records on Vertnet.org.

202 Since the 1960s, the potential for using plants and animals to understand environmental
203 pollutants has been an active area of scientific research (e.g., Ratcliffe 1967, Hickey and
204 Andersen 1968). For example, studies have examined how animals sequester and
205 bioaccumulate pollutants from the environment in their tissues (feathers, hairs, bones, muscles,
206 organs, exoskeletons, etc.) (relevant studies can be found in Table S1). In trees, tar spots and
207 leaf/needle damage reflect exposure to air pollutants like sulfur dioxide and ozone (e.g. Oshima
208 1974, Bevan and Greenhalgh 1976). Bivalves in aquatic systems, like mussels, filter pollutants
209 from the water column and sequester these pollutants in soft tissues, which have been
210 measured to assess heavy metal contamination and pollutants from plastics like phthalates and
211 polychlorinated biphenyls (PCBs) (e.g. Marigómez et al. 2013), and fish have been used to
212 assess changes in microplastics in the environment (Hou et al. 2021). These examples highlight
213 a small subset of studies that have documented how plants and animals capture pollutants from
214 the environments they inhabit. These examples—and others mentioned throughout the text —
215 also highlight the various ways in which organisms bioaccumulate pollutants. For example,
216 black carbon particles accumulate superficially on an organism, such as on bird feathers, while
217 heavy metals bioaccumulate in animal tissues, like bone. A non-exhaustive list of more than a
218 hundred pollutants that have been assessed in most major plant and animal groups can be
219 found in Table S1.

220 The motivation for much of this research on pollution levels in plant and animal tissues is largely
221 predicated on the assumption that organisms are reliable bioindicators that can be used to
222 better understand exposure and the broader impacts of pollutants on ecosystem health and
223 human health. However, there is a gap between measuring pollutants in specimens and their
224 use in public health research, in part because of the complex analyses required to relate
225 specimen-based pollution data to human health outcomes. Two other reasons may also help
226 explain why natural history specimens have been underutilized in environmental health
227 research. First, most bioindicator studies assess either temporal dynamics alone (from a single
228 location or a handful of locations through time), or spatial variation for a given point in time. That
229 is, most research does not jointly assess spatial and temporal dynamics, which limits how this
230 methodology can be used for questions of public health and policy given the spatial and
231 temporal complexity of environmental pollution. Second, there are disciplinary silos that prevent
232 useful exchanges and data integration. On one hand, the natural history museum community
233 has access to specimens but has been predominantly invested in questions related to
234 biodiversity (Webster 2017). On the other hand, public health, policy, and environmental justice
235 communities are driving questions about human health and equity, but they may have less
236 familiarity with, or access to, biological collections. A necessary first step in realizing the full
237 potential of biological collections for public health is to find ways to connect disciplinary
238 communities, which has significant potential for mutual enrichment, as exemplified in ongoing
239 efforts to use natural history specimens to understand emerging pathogens and zoonotic
240 diseases, and to track and monitor animal vectors of disease (for detailed accounts of this work
241 see Cook et al. 2020, Colella et al. 2021, and Thompson et al. 2021).

242 **Linking historical pollution data with public health and policy**

243 The *time capsule* feature of natural history specimens makes them uniquely valuable for
244 reconstructing the environmental record for pollutants of public health concern, as well as a key
245 data source for understanding emerging pollutants in the future. By analyzing pollutants from a
246 time series of specimens, there is significant potential to build spatially dynamic datasets of
247 pollutants over long periods of time, which can be paired with spatially and temporally matched
248 datasets of human health and demography, environmental policy, and more. In Box 1, we
249 outline a potential workflow to illustrate these connections using lead pollution as an example.

250 Spatially- and temporally-explicit datasets would provide a foundation for quantitative
251 assessment of salient questions, for example, about how pollution levels and exposure risk are
252 moderated or exacerbated by policy choices and social contexts, or how exposure risk is
253 associated with disparities in public health outcomes over time. Past and current pollution data
254 could also be paired with demographic data to explore links between exposure risk and
255 population density, race, income, and other factors to identify pollutants to which communities
256 are at risk of being exposed. Pollution data from biological specimens can be used to inform
257 environmental and land-use policy; they can also be used to estimate exposure risks in parental
258 and grandparental generations to better understand the transgenerational impacts of exposure
259 on human health. Furthermore, pollution data could be compared with mortality or morbidity
260 data of various diseases to assess associations between exposure risk, health outcomes, and
261 disease pathology. This is particularly relevant for diseases such as neurological disorders that
262 manifest later in life and potentially result from long-term exposure to pollutants at low levels
263 (Ritz and Yu 2021). For these diseases, we have lacked datasets of appropriate time scale to
264 understand their pathologies, and temporal pollution maps could provide the key to better
265 understanding the relationships between exposure and disease. Methodological models already
266 exist for joining spatial and temporally matched dataset, for example, in linking air quality to
267 health outcomes, like for asthma (Friedman et al. 2001) and COVID-19 mortality (Villeneuve
268 and Goldberg 2020, Wu et al. 2020, Kerr et al. 2021). What we propose here is to extend these
269 types of methods to robust, quantitative pollution datasets built using time series of natural
270 history specimens.

271 **Challenges and big questions moving forward**

272 With quantitative pollution data, there are several public health and environmental challenges
273 that we can begin to address with robust, spatially dynamic temporal data. Addressing these
274 challenges will demand diverse expertise and creativity, opening up a transdisciplinary space for
275 community-based research and engagement while bridging the natural, physical, and social
276 sciences. Some of the challenges that robust pollution datasets will help clarify and contribute to
277 include:

278 (1) *Understanding the transgenerational impacts of exposure and the impacts of long-*
279 *term exposure, chronic low-level exposure, and time lags in exposure on health*
280 *outcomes:* Understanding these impacts has been challenging not only because we

281 have lacked the necessary breadth of temporal data on exposure, but because of the
282 time lags between exposure and the onset of health conditions. The temporal breadth of
283 data from specimens (that spans multiple generations) is situated to tackle this exact
284 challenge.

285 *(2) Disentangling cocktails of exposure and the associations between individual*
286 *pollutants and public health outcomes:* Pollutants often co-vary through time and space,
287 and individuals are exposed to multiple pollutants across their lifetime. These dynamics
288 present major challenges for disentangling the impacts of distinct pollutants on health
289 outcomes and establishing the origins of pollutants to hold polluters accountable. It will
290 be important to develop temporal and spatial sampling schemes that allow researchers
291 to delineate the respective and combined effects of multiple pollutants on public health
292 outcomes, while being able to trace pollutants to their industrial origins. For example,
293 chemical analyses were critical in linking General Motors to PCB waste that entered the
294 Raquette and St. Lawrence Rivers to hold them legally accountable for remediation and
295 damages (DeCaprio et al. 2005, Begum et al. 2022). The spatial and temporal breadth of
296 natural history specimens makes them well suited to tackle these challenges by
297 capturing the subtleties of how pollutants co-vary across the environment. Achieving this
298 will bring us closer to moving from correlation to causality.

299 *(3) Calibrating historical levels of exposure and deposition with modern sample data:* We
300 can assess trends of pollutant levels through time, but how does the pollutant level in an
301 individual animal/specimen relate to the level being emitted and deposited in the
302 environment? By pairing modern environmental pollution data with pollutant levels in
303 specimens sampled at present day, we can calibrate pollution models to better estimate
304 historical pollution levels in the environment. There is a lot that we can learn by
305 understanding relative trends in historical pollution but having estimates of the actual
306 concentrations/levels in the environment can help us tackle some of the other challenges
307 mentioned here.

308 *(4) Understanding how pollutants move through the environment and biological systems:*
309 Just because a plant or animal was exposed to a pollutant does not necessarily translate
310 to human exposure. It will be important to understand how pollutants move through the
311 environment and across trophic levels (i.e., food webs), while defining the pathways
312 through which humans are exposed from the environment and understanding when and
313 how human exposure is coupled or decoupled from non-human animal exposure—e.g.,
314 what is the extent of lead exposure coming from the external environment vs. the built
315 environment, like from water pipes? This work will allow us to better understand how
316 exposure relates to the outdoor environment vs. the built environment, while
317 disentangling the two, and will help uncover possible routes of human exposure.

318 *(5) Understanding the role of organisms in filtering or concentrating pollutants in the*
319 *environment:* To what extent are plants and animals providing an ecosystem service that
320 benefits human health by removing pollutants from the environment? In contrast, how

321 are organisms concentrating and increasing pollutants locally by accumulating pollutants
322 from one location and depositing them in another location (e.g., via movement coupled
323 with death, excretion, molting, etc.)? Furthermore, how do these competing dynamics
324 change through space and time and with biodiversity loss, land conversion, and habitat
325 degradation? While a key feature of the framework presented here is that specimens
326 can help us understand pollution dynamics through space and time, these data can also
327 be used to understand the interactions between pollution, biodiversity, and exposure
328 risks.

329 *(6) Understanding the mechanisms of accumulation, storage, and degradation of*
330 *pollutants in organisms and biological samples:* While much research has shown that
331 organisms bioaccumulate pollutants from their environments, it will be critical to better
332 understand how this accumulation happens, and if/how pollutants degrade in natural
333 history specimens over time. We expect, for example, that pollutants such as heavy
334 metals and persistent organic pollutants (like PCBs) remain at stable levels in specimens
335 through time in collections, but information about the “life cycle” of a pollutant in
336 specimens will help researchers refine the methodology proposed here and understand
337 its limitations.

338 These challenges and research questions highlight the complexity and interdisciplinarity needed
339 in studying environmental health. Our intention here is not to downplay the practical and
340 methodological challenges in addressing these questions, which will demand creative,
341 interdisciplinary teams of ecologists, medical professionals, social scientists, public health
342 experts, communities most impacted by environmental pollution, activists, engineers, historians
343 and others. We believe, however, that these challenges are not insurmountable, and they create
344 space to build diverse teams and partnerships to tackle some of our most significant public
345 health challenges of the twenty-first century.

346 **Conclusions**

347 Natural history collections present a unique opportunity to understand pollution concentrations
348 in the environment over nearly two centuries. These data are especially informative for time
349 periods and locales that lack environmental monitoring data, which is applicable for most
350 pollutants because environmental monitoring often only starts after a pollutant is discovered to
351 be a public health issue. The spatial and temporal scope of specimen collections offer an
352 opportunity to fill these gaps in the environmental record to better understand pollution, its
353 history and origins, and its impacts on the environment and human health.

354 With a growing recognition of the links between human health, the environments in which we
355 live and work, and the social structures that create and perpetuate environmental inequity, there
356 is an opportunity to expand the One Health framework beyond its emphasis on infectious
357 disease to advance research that tackles the history and impacts of environmental pollution on
358 public health outcomes and environmental inequity. Natural history collections can support this
359 work through their capacity to generate quantitative spatiotemporal data on pollutants. This work

360 will not only require diverse teams and partnerships, but it also calls for increased funding and
361 investment into specimen-based work to further develop collections as essential tools and to
362 build the necessary collaborations across disciplines and communities. With creative and
363 transdisciplinary efforts to leverage collections, we have an unparalleled opportunity to make
364 significant progress in understanding and mitigating the impacts of environmental change.

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368 **BOX 1.** A case study to reconstructing lead pollution from natural history specimens: the need
369 for comprehensive data, a proposed methodology, and its implications for public health and
370 policy.

Background	In 2014, the Flint Water Crisis brought lead pollution back to the forefront of national discourse in the U.S.; this public health crisis has real and lasting consequences for those in Flint, and also serves as an important point of public awareness (Olson and Fedinick 2018). However, it is also important to recognize that the discourse around lead exposure in Flint has focused (and continues to focus) on a single pathway of exposure—leaded pipes and drinking water—whereas in the past, public attention focused on other pathways of exposure, like lead-based paint (Mushak and Crocetti 1989, Marino et al. 2011) and leaded gasoline (Annest et al. 1983). Each one of these pathways of lead exposure is of significant public health concern, but treated separately they limit a comprehensive understanding of the full impacts of lead in the environment and on public health.
Problem and its Consequences	<p>We lack comprehensive data on where lead is in the environment. That is, we lack data on the 'life cycle' of lead once it enters the environment: where it goes, how it moves through the environment, the various pathways of human exposure, and how these things have changed across geographic space and through time (i.e., its 'history'), particularly in relation to land-use change, biodiversity loss, and policy (Tovar-Sánchez et al. 2018, Kalani et al. 2021, Levin et al. 2021). As in Flint, lead exposure is often attributed to drinking water, but the movement and cycle of lead through our environments is much more complex and can happen in tandem with other pollutants (e.g., mercury) that are also of significant public health concern.</p> <p>The public health consequences of environmental pollutants like lead are global, and disproportionately impact children, lower income countries, and marginalized communities (IHME 2019, Ritchie and Roser 2022). Without understanding lead pollution levels in the environment—and how these levels have changed through geographic space and time—it remains difficult to: (1) understand the independent and cumulative effects of different pathways of lead exposure, (2) tease apart the independent and cumulative effects of exposure to multiple pollutants, and (3) understand the impacts of chronic low-level lead exposure and its cascading social and health consequences across generations. Understanding these complexities are necessary to inform policy and legal measures to most effectively mitigate harm, protect vulnerable populations, and repair our relationships with the environment and build healthier relationships moving forward.</p>

<p>Solution</p>	<p>Quantifying lead levels in natural history specimens can uniquely help us reconstruct the life cycle of lead in the environment. These datasets would capture the biological component of environmental lead (e.g., identifying organismal burdens of lead, how lead moves through food webs and between aquatic and terrestrial environments). These data could then be paired with complementary datasets that capture abiotic components of lead, like lead in water samples and in sediment cores. By integrating across biotic and abiotic datasets, we could build a comprehensive and foundational understanding of where lead is (and has been) in the environment.</p>
<p>Measuring lead from specimens</p>	<p>Lead concentrations can be quantified from organisms with spectrometry methods. For example, total elemental concentrations of lead can be measured in plant tissues (roots, leaves, bark pockets, growth rings, etc.), lichens, mollusk soft tissues, fungal fruiting bodies, arthropod tissues (the whole organism or products like spider webs), and vertebrate tissues (feathers, hair, scales, bone, kidney, liver, spleen, eggs, etc.). See Supplementary Table 1 for a list of studies that provide detailed methodology for quantifying lead from diverse tissue types using spectrophotometry.</p> <p>While spectrophotometry methods are frequently used to quantify lead burdens in organisms and have been informative for understanding the impacts of lead on the environment, they have two big limitations that hinder their ability for scaling up: (1) the methodology is time consuming and (2) often requires destroying the sample. However, recent methodological advancements with X-ray fluorescence (XRF) allow for non-destructive high-throughput sampling, significantly increasing the rate of data collection without destroying the sample (Specht et al. 2014, Specht et al. 2016). In vertebrates, skeletal lead is the predominant source of body burden (Barry 1975), and Specht and colleagues have been successfully developing XRF methods for quantifying lead concentrations in bones, as well as from blood spots (e.g., Specht et al. 2018, Specht and Weisskopf 2019, Specht et al. 2019, Hampton et al. 2021, Specht et al. 2021). With these types of methodological advances (and future methods development) it will be possible to design feasible sampling schemes to generate a robust understanding of the biological component of environmental lead.</p>
<p>Health implications</p>	<p>Detailed data on the spatial and temporal extent of lead in the environment would provide a quantitative framework for better understanding its consequences on human health and disease pathology, particularly its long-term and transgenerational impacts. For example, these data could be paired with spatially and temporally matched public health datasets, like US Medicare insurance claims for neurological/neurodegenerative conditions, which manifest later in life. Given that Medicare was established in 1965 in the US for individuals over the age of 65, Medicare insurance claim data captures potential health outcomes from over 120 years of environmental exposure. When analyzed in conjunction, these paired datasets could help us better understand correlative links between environmental lead, lifetime exposure risk, and disease pathology. This hypothetical example is one of the many ways in which detailed environmental pollution data could inform our understanding of the health consequences of pollution across time scales (e.g., over someone's lifetime or across generations) that have been previously difficult to address.</p>
<p>Policy implications</p>	<p>How can we understand the impacts of past environmental policy regarding lead without understanding how pollution loads have changed through geographic space and time? Detailed environmental data on lead has both upstream and downstream implications for policy.</p> <p><i>Upstream (historical) policy implications:</i> What role has policy played in producing lead pollution distributions, and what has been the efficacy of past policy reforms designed to mitigate lead pollution? Quantitative temporal lead data would allow us to answer these questions and better understand the impacts and efficacy of policy in meeting its goals. To illustrate this point, in 2017 we reconstructed 135 years of atmospheric black carbon pollution in the U.S. Manufacturing Belt from bird specimens (DuBay and Fuldner 2017). With these data we were able to assess the various roles of federal and city-level</p>

legislation in mitigating black carbon pollution. With future studies, the spatial and temporal specificity of datasets derived from biological collections would allow for policy assessments across various spatial scales of interest (local, regional, federal) as well as various time scales.

Downstream (future) policy implications: Informed by new, quantitative data on environmental lead (and other pollutants), what are effective and equitable policy responses? What information can these data provide for Superfund sites and other federal clean-up efforts or for future land use decisions? Can these data empower vulnerable populations and the communities most impacted by pollutants in their efforts to push for policy measures that best serve their needs and protect their health? How does a better understanding of past policy inform future policy decisions? These are all consequential questions that impact public health and well-being that hinge on a comprehensive understanding of the 'life' and 'history' of pollutants in the environment.

References:

American Geophysical Union. n.d. *Fact Sheet on Geohealth*. <https://www.agu.org/Share-and-Advocate/Share/Policy-makers/Position-Statements/Fact-sheet-Geohealth>

Annest, J.L., Pirkle, J.L., Makuc, D., Neese, J.W., Bayse, D.D. and Kovar, M.G., 1983. Chronological trend in blood lead levels between 1976 and 1980. *New England Journal of Medicine*, 308(23), pp.1373-1377.

Anway, M.D., Cupp, A.S., Uzumcu, M. and Skinner, M.K., 2005. Epigenetic transgenerational actions of endocrine disruptors and male fertility. *science*, 308(5727), pp.1466-1469.

Arcaya, M.C. and Figueroa, J.F., 2017. Emerging trends could exacerbate health inequities in the United States. *Health Affairs*, 36(6), pp.992-998.

Barry, P.S., 1975. A comparison of concentrations of lead in human tissues. *Occupational and Environmental Medicine*, 32(2), pp.119-139.

Begum, T.F. and Carpenter, D.O., 2022. Review of Environmental Contamination at Akwesasne and Associated Health Outcomes. *Journal of Indigenous Research*, 10(2022), p.1.

Bevan, R.J. and Greenhalgh, G.N., 1976. *Rhytisma acerinum* as a biological indicator of pollution. *Environmental Pollution* (1970), 10(4), pp.271-285.

Billionnet, C., Sherrill, D. and Annesi-Maesano, I., 2012. Estimating the health effects of exposure to multi-pollutant mixture. *Annals of epidemiology*, 22(2), pp.126-141.

Bond, T.C., Bhardwaj, E., Dong, R., Jogani, R., Jung, S., Roden, C., Streets, D.G. and Trautmann, N.M., 2007. Historical emissions of black and organic carbon aerosol from energy-related combustion, 1850–2000. *Global biogeochemical cycles*, 21(2)

Bond, A.L., Hobson, K.A. and Branfireun, B.A., 2015. Rapidly increasing methyl mercury in endangered ivory gull (*Pagophila eburnea*) feathers over a 130 year record. *Proceedings of the Royal Society B: Biological Sciences*, 282(1805), p.20150032.

Bové, H., Bongaerts, E., Slenders, E., Bijmens, E.M., Saenen, N.D., Gyselaers, W., Van Eyken, P., Plusquin, M., Roeffaers, M.B., Ameloot, M. and Nawrot, T.S., 2019. Ambient black carbon particles reach the fetal side of human placenta. *Nature communications*, 10(1), pp.1-7.

Brulle, R.J. and Pellow, D.N., 2006. Environmental justice: human health and environmental inequalities. *Annu. Rev. Public Health*, 27, pp.103-124.

Bullard, R.D., 2008. *Dumping in Dixie: Race, class, and environmental quality*. Avalon Publishing-(Westview Press).

Castellón, I.G., 2021. Cancer Alley and the fight against environmental racism. *Vill. Envtl. LJ*, 32, p.15.

Center for Disease Control and Prevention. n.d. *One Health Basics*.
<https://www.cdc.gov/onehealth/basics/index.html>

Chakraborty, J., 2022. Disparities in exposure to fine particulate air pollution for people with disabilities in the US. *Science of The Total Environment*, 842, p.156791.

Colella, J.P., Bates, J., Burneo, S.F., Camacho, M.A., Carrion Bonilla, C., Constable, I., D'Elía, G., Dunnum, J.L., Greiman, S., Hoberg, E.P. and Lessa, E., 2021. Leveraging natural history biorepositories as a global, decentralized, pathogen surveillance network. *PLoS Pathogens*, 17(6), p.e1009583.

Collins, T.W., Grineski, S.E. and Chakraborty, J., 2015. Household-level disparities in cancer risks from vehicular air pollution in Miami. *Environmental Research Letters*, 10(9), p.095008.

Collins, T.W., Grineski, S.E. and Morales, D.X., 2017. Sexual orientation, gender, and environmental injustice: Unequal carcinogenic air pollution risks in greater Houston. *Annals of the American Association of Geographers*, 107(1), pp.72-92.

Colmer, J., Hardman, I., Shimshack, J. and Voorheis, J., 2020. Disparities in PM_{2.5} air pollution in the United States. *Science*, 369(6503), pp.575-578.

Cook, J.A., Arai, S., Armién, B., Bates, J., Bonilla, C.A.C., Cortez, M.B.D.S., Dunnum, J.L., Ferguson, A.W., Johnson, K.M., Khan, F.A.A. and Paul, D.L., 2020. Integrating biodiversity infrastructure into pathogen discovery and mitigation of emerging infectious diseases. *Bioscience*, 70(7), pp.531-534.

Crabb, A. and Leroy, P., 2012. The handbook of environmental policy evaluation. Routledge.

Culhane, J.F. and Goldenberg, R.L., 2011, August. Racial disparities in preterm birth. In *Seminars in perinatology* (Vol. 35, No. 4, pp. 234-239). WB Saunders.

DeCaprio, A.P., Johnson, G.W., Tarbell, A.M., Carpenter, D.O., Chiarenzelli, J.R., Morse, G.S., Santiago-Rivera, A.L., Schymura, M.J. and Akwesasne Task Force on the Environment, 2005. Polychlorinated biphenyl (PCB) exposure assessment by multivariate statistical analysis of serum congener profiles in an adult Native American population. *Environmental Research*, 98(3), pp.284-302.

Di, Q., Wang, Y., Zanobetti, A., Wang, Y., Koutrakis, P., Choirat, C., Dominici, F. and Schwartz, J.D., 2017. Air pollution and mortality in the Medicare population. *New England Journal of Medicine*, 376(26), pp.2513-2522

Dominici, F., Peng, R.D., Barr, C.D. and Bell, M.L., 2010. Protecting human health from air pollution: shifting from a single-pollutant to a multi-pollutant approach. *Epidemiology* (Cambridge, Mass.), 21(2), p.187

DuBay, S.G. and Fuldner, C.C., 2017. Bird specimens track 135 years of atmospheric black carbon and environmental policy. *Proceedings of the National Academy of Sciences*, 114(43), pp.11321-11326.

DuBay, S., Palmer Droguett, D.H. and Piland, N., 2020. Global inequity in scientific names and who they honor. *BioRxiv*, pp.2020-08.

Fitzgerald, E.F., Hwang, S.A., Bush, B., Cook, K. and Worswick, P., 1998. Fish consumption and breast milk PCB concentrations among Mohawk women at Akwesasne. *American journal of epidemiology*, 148(2), pp.164-172.

Friedman, M.S., Powell, K.E., Hutwagner, L., Graham, L.M. and Teague, W.G., 2001. Impact of changes in transportation and commuting behaviors during the 1996 Summer Olympic Games in Atlanta on air quality and childhood asthma. *Jama*, 285(7), pp.897-905.

Goldsmith, L. and Bell, M.L., 2022. Queering Environmental Justice: Unequal Environmental Health Burden on the LGBTQ+ Community. *American Journal of Public Health*, 112(1), pp.79-87.

Hampton, J.O., Specht, A.J., Pay, J.M., Pokras, M.A. and Bengsen, A.J., 2021. Portable X-ray fluorescence for bone lead measurements of Australian eagles. *Science of The Total Environment*, 789, p.147998.

Hickey, J.J. and Anderson, D.W., 1968. Chlorinated hydrocarbons and eggshell changes in raptorial and fish-eating birds. *Science*, 162(3850), pp.271-273

Hill, M.K., 2020. Understanding environmental pollution. Cambridge University Press.

Hoke, M.K. and McDade, T., 2014. Biosocial inheritance: A framework for the study of the intergenerational transmission of health disparities. *Annals of Anthropological Practice*, 38(2), pp.187-213.

Homan, P., Brown, T.H. and King, B., 2021. Structural intersectionality as a new direction for health disparities research. *Journal of Health and Social Behavior*, 62(3), pp.350-370.

Hou, L., McMahan, C.D., McNeish, R.E., Munno, K., Rochman, C.M. and Hoellein, T.J., 2021. A fish tale: a century of museum specimens reveal increasing microplastic concentrations in freshwater fish. *Ecological Applications*, p.e02320

Ilechukwu, I., Das, R.R. and Reimer, J.D., 2023. Review of microplastics in museum specimens: An under-utilized tool to better understand the Plasticene. *Marine Pollution Bulletin*, 191, p.114922.

Institute for Health Metrics and Evaluation (IHME). GBD Compare. Seattle, WA: IHME, University of Washington; 2019

Institute of Medicine (US) Committee on Environmental Justice, 1999. Toward environmental justice: research, education, and health policy needs.

Kalani, T.J., South, A., Talmadge, C., Leibler, J., Whittier, C. and Rosenbaum, M., 2021. One map: Using geospatial analysis to understand lead exposure across humans, animals, and the environment in an urban US city. *One Health*, 13, p.100341.

Kerr, G.H., Goldberg, D.L. and Anenberg, S.C., 2021. COVID-19 pandemic reveals persistent disparities in nitrogen dioxide pollution. *Proceedings of the National Academy of Sciences*, 118(30).

Kilbourne, B., 2020. Natural History, the Curious Institution. *Ecotone*, 16(2), pp.96-100.

Lane, H.M., Morello-Frosch, R., Marshall, J.D. and Apte, J.S., 2022. Historical redlining is associated with present-day air pollution disparities in US cities. *Environmental science & technology letters*, 9(4), pp.345-350.

Lepak, R.F., Hoffman, J.C., Janssen, S.E., Krabbenhoft, D.P., Ogorek, J.M., DeWild, J.F., Tate, M.T., Babiarz, C.L., Yin, R., Murphy, E.W. and Engstrom, D.R., 2019. Mercury source changes and food web shifts alter contamination signatures of predatory fish from Lake Michigan. *Proceedings of the National Academy of Sciences*, 116(47), pp.23600-23608.

Levin, R., Vieira, C.L.Z., Rosenbaum, M.H., Bischoff, K., Mordarski, D.C. and Brown, M.J., 2021. The urban lead (Pb) burden in humans, animals and the natural environment. *Environmental research*, 193, p.110377.

Levy, I., Mihele, C., Lu, G., Narayan, J. and Brook, J.R., 2014. Evaluating multipollutant exposure and urban air quality: pollutant interrelationships, neighborhood variability, and nitrogen dioxide as a proxy pollutant. *Environmental health perspectives*, 122(1), pp.65-72.

Li, L., Sangion, A., Wania, F., Armitage, J.M., Toose, L., Hughes, L. and Arnot, J.A., 2021. Development and evaluation of a holistic and mechanistic modeling framework for chemical emissions, fate, exposure, and risk. *Environmental health perspectives*, 129(12), p.127006.

Liboiron, M., 2021. *Pollution is colonialism*. Duke University Press.

Liévanos, R.S., 2018. Retooling CalEnviroScreen: Cumulative pollution burden and race-based environmental health vulnerabilities in California. *International Journal of Environmental Research and Public Health*, 15(4), p.762.

Lombó, M., Fernández-Díez, C., González-Rojo, S., Navarro, C., Robles, V. and Herráez, M.P., 2015. Transgenerational inheritance of heart disorders caused by paternal bisphenol A exposure. *Environmental pollution*, 206, pp.667-678.

MacKenzie, J.M. ed., 1990. *Imperialism and the natural world*. Manchester: Manchester University Press.

Maiti, S.K. and Chowdhury, A., 2013. Effects of anthropogenic pollution on mangrove biodiversity: a review. *Journal of Environmental Protection*, 2013.

Marigómez, I., Zorita, I., Izagirre, U., Ortiz-Zarragoitia, M., Navarro, P., Etxebarria, N., Orbea, A., Soto, M. and Cajaraville, M.P., 2013. Combined use of native and caged mussels to assess biological effects of pollution through the integrative biomarker approach. *Aquatic toxicology*, 136, pp.32-48.

Marino, P.E., Landrigan, P.J., Graef, J., Nussbaum, A.B.R.A.H.A.M., Bayan, G.R.E.G.O.R.Y., Boch, K. and Boch, S.T.E.V.E.N., 1990. A case report of lead paint poisoning during renovation of a Victorian farmhouse. *American Journal of Public Health*, 80(10), pp.1183-1185.

Mohai, P., Pellow, D. and Roberts, J.T., 2009. Environmental justice. *Annual review of environment and resources*, 34, pp.405-430.

Mohai, P. and Saha, R., 2015. Which came first, people or pollution? Assessing the disparate siting and post-siting demographic change hypotheses of environmental injustice. *Environmental research letters*, 10(11), p.115008.

Morello-Frosch, R. and Shenassa, E.D., 2006. The environmental “riskscape” and social inequality: implications for explaining maternal and child health disparities. *Environmental health perspectives*, 114(8), pp.1150-1153.

Murphy, M., 2008. Chemical regimes of living. *Environmental History*, 13(4), pp.695-703.

Mushak, P. and Crocetti, A.F., 1989. Determination of numbers of lead-exposed American children as a function of lead source: Integrated summary of a report to the US Congress on childhood lead poisoning. *Environmental research*, 50(2), pp.210-229.

National Academies of Sciences, Engineering, and Medicine, 2021. *Biological collections: ensuring critical research and education for the 21st century*. National Academies Press

Nilsen, E., Smalling, K.L., Ahrens, L., Gros, M., Miglioranza, K.S., Picó, Y. and Schoenfuss, H.L., 2019. Critical review: grand challenges in assessing the adverse effects of contaminants of emerging concern on aquatic food webs. *Environmental Toxicology and Chemistry*, 38(1), pp.46-60.

Olson, E. D, Fedinick, K. P., 2018. What's in your water? Flint and beyond. Natural Resources Defense Council. Available at: <https://www.nrdc.org/sites/default/files/whats-in-your-water-flint-beyond-report.pdf>, Accessed 13 April 2022

Oshima, R.J., 1974. A viable system of biological indicators for monitoring air pollutants. *Journal of the Air Pollution Control Association*, 24(6), pp.576-578.

Payne-Sturges, D. and Gee, G.C., 2006. National environmental health measures for minority and low-income populations: tracking social disparities in environmental health. *Environmental Research*, 102(2), pp.154-171.

Quintero Toro, C., 2012. *Birds of empire, birds of nation: A history of science, economy, and conservation in United States-Colombia relations*. Ediciones Uniandes-Universidad de los Andes.

Ratcliffe, D.A., 1967. Decrease in eggshell weight in certain birds of prey. *Nature*, 215(5097), pp.208-210.

Ritchie, H. and Rose, M. 2022. "Lead Pollution". Published online at OurWorldInData.org. Retrieved from: '<https://ourworldindata.org/lead-pollution>'

Ritz, B. and Yu, Y., 2021. Invited perspective: Air pollution and dementia: Challenges and opportunities. *Environmental Health Perspectives*, 129(8), p.081301.

Rodgers, K., McLellan, I., Peshkur, T., Williams, R., Tonner, R., Knapp, C.W., Henriquez, F.L. and Hursthouse, A.S., 2020. The legacy of industrial pollution in estuarine sediments: spatial

and temporal variability implications for ecosystem stress. *Environmental geochemistry and health*, 42, pp.1057-1068.

Rosenbaum, R.K., Huijbregts, M.A., Henderson, A.D., Margni, M., McKone, T.E., Van De Meent, D., Hauschild, M.Z., Shaked, S., Li, D.S., Gold, L.S. and Jolliet, O., 2011. USEtox human exposure and toxicity factors for comparative assessment of toxic emissions in life cycle analysis: sensitivity to key chemical properties. *The International Journal of Life Cycle Assessment*, 16(8), pp.710-727.

Rudel, T.K., Roberts, J.T. and Carmin, J., 2011. Political economy of the environment. *Annual Review of Sociology*, 37, pp.221-238.

Schilling, K.E., Jha, M.K., Zhang, Y.K., Gassman, P.W. and Wolter, C.F., 2008. Impact of land use and land cover change on the water balance of a large agricultural watershed: Historical effects and future directions. *Water Resources Research*, 44(7).

Schmidt, C.W., 2013. Uncertain inheritance: transgenerational effects of environmental exposures.

Shukla, A., Bunkar, N., Kumar, R., Bhargava, A., Tiwari, R., Chaudhury, K., Goryacheva, I.Y. and Mishra, P.K., 2019. Air pollution associated epigenetic modifications: transgenerational inheritance and underlying molecular mechanisms. *Science of The Total Environment*, 656, pp.760-777.

Singer, M., 2011. Down cancer alley: the lived experience of health and environmental suffering in Louisiana's chemical corridor. *Medical Anthropology Quarterly*, 25(2), pp.141-163.

Specht, A.J., Kirchner, K.E., Weisskopf, M.G. and Pokras, M.A., 2019. Lead exposure biomarkers in the Common Loon. *Science of the Total Environment*, 647, pp.639-644.

Specht, A.J., Lin, Y., Weisskopf, M., Yan, C., Hu, H., Xu, J. and Nie, L.H., 2016. XRF-measured bone lead (Pb) as a biomarker for Pb exposure and toxicity among children diagnosed with Pb poisoning. *Biomarkers*, 21(4), pp.347-352.

Specht, A.J., Obrycki, J.F., Mazumdar, M. and Weisskopf, M.G., 2021. Feasibility of lead exposure assessment in blood spots using energy-dispersive X-ray fluorescence. *Environmental Science & Technology*, 55(8), pp.5050-5055.

Specht, A.J., Parish, C.N., Wallens, E.K., Watson, R.T., Nie, L.H. and Weisskopf, M.G., 2018. Feasibility of a portable X-ray fluorescence device for bone lead measurements of condor bones. *Science of The Total Environment*, 615, pp.398-403.

Specht, A. and Weisskopf, M., 2019. Feasibility of lead exposure assessment in blood spots using x-ray fluorescence. *Environmental Epidemiology*, 3, p.379.

Specht, A.J., Weisskopf, M. and Nie, L.H., 2014. Portable XRF technology to quantify Pb in bone in vivo. *Journal of biomarkers*, 2014.

Stradling, D., 1999. *Smokestacks and progressives: Environmentalists, engineers and air quality in America, 1881-1951* (p. 131). Baltimore: Johns Hopkins University Press.

Tang, Z., Engel, B.A., Pijanowski, B.C. and Lim, K.J., 2005. Forecasting land use change and its environmental impact at a watershed scale. *Journal of environmental management*, 76(1), pp.35-45.

Taylor, D., 2014. Toxic communities. In *Toxic Communities*. New York University Press.

Tessum, C.W., Apte, J.S., Goodkind, A.L., Muller, N.Z., Mullins, K.A., Paoella, D.A., Polasky, S., Springer, N.P., Thakrar, S.K., Marshall, J.D. and Hill, J.D., 2019. Inequity in consumption of goods and services adds to racial–ethnic disparities in air pollution exposure. *Proceedings of the National Academy of Sciences*, 116(13), pp.6001-6006.

Tessum, C.W., Paoella, D.A., Chambliss, S.E., Apte, J.S., Hill, J.D. and Marshall, J.D., 2021. PM2.5 polluters disproportionately and systemically affect people of color in the United States. *Science Advances*, 7(18), p.eabf4491.

Thompson, C.W., Phelps, K.L., Allard, M.W., Cook, J.A., Dunnum, J.L., Ferguson, A.W., Gelang, M., Khan, F.A.A., Paul, D.L., Reeder, D.M. and Simmons, N.B., 2021. Preserve a voucher specimen! The critical need for integrating natural history collections in infectious disease studies. *Mbio*, 12(1), pp.e02698-20.

Tovar-Sánchez, E., Hernández-Plata, I., Martínez, M.S., Valencia-Cuevas, L. and Galante, P.M., 2018. Heavy metal pollution as a biodiversity threat. *Heavy Metals*, 383.

United Church of Christ. Commission for Racial Justice, 1987. *Toxic wastes and race in the United States: A national report on the racial and socio-economic characteristics of communities with hazardous waste sites*. Public Data Access.

US EPA, 2007. *The Multi-Pollutant Report: Technical Concepts and Examples*.

Villeneuve, P.J. and Goldberg, M.S., 2020. Methodological considerations for epidemiological studies of air pollution and the SARS and COVID-19 coronavirus outbreaks. *Environmental Health Perspectives*, 128(9), p.095001.

Vo, A.T.E., Bank, M.S., Shine, J.P. and Edwards, S.V., 2011. Temporal increase in organic mercury in an endangered pelagic seabird assessed by century-old museum specimens. *Proceedings of the National Academy of Sciences*, 108(18), pp.7466-7471.

Wang, M., Aaron, C.P., Madrigano, J., Hoffman, E.A., Angelini, E., Yang, J., Laine, A., Vetterli, T.M., Kinney, P.L., Sampson, P.D. and Sheppard, L.E., 2019. Association between long-term exposure to ambient air pollution and change in quantitatively assessed emphysema and lung function. *Jama*, 322(6), pp.546-556.

Wallace, L.A., 1991. Comparison of risks from outdoor and indoor exposure to toxic chemicals. *Environmental Health Perspectives*, 95, pp.7-13.

Webster, M.S. ed., 2017. *The extended specimen: emerging frontiers in collections-based ornithological research*. CRC Press.

Wu, X., Nethery, R.C., Sabath, M.B., Braun, D. and Dominici, F., 2020. Air pollution and COVID-19 mortality in the United States: Strengths and limitations of an ecological regression analysis. *Science advances*, 6(45), p.eabd4049