

1 **Title: Phenological responses to climate change across taxa and local habitats**
2 **in a high-Arctic arthropod community**

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15 **Open Research statement:** Arthropod phenological data and climate data are already published
16 and publicly available at <https://data.g-e-m.dk/>, with those publications properly cited in this
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23 snowmelt, temperature
24

31 Abstract

32 Climate change has led to pronounced phenological responses that typically vary across taxa. The
33 rate of warming is especially high in the Arctic, but comprehensive long-term data on phenological
34 changes is rare in this region, especially for arthropods – a diverse taxonomic group that form
35 important links to other trophic levels. Understanding the environmental drivers of arthropod
36 phenological variation is necessary to predict future trends across taxa and habitats. Here, we
37 analyze temporal trends and climate associations in arthropod phenology using 25 years of
38 standardized monitoring data from five different habitats in high-Arctic Greenland. We observed
39 earlier peak activity and extended activity periods in the arthropod community, but this directional
40 trend was weak, and the magnitude of responses varied considerably among taxa and local habitats.
41 Snowmelt timing was generally a strong driver of arthropod phenology, and a key cue for earlier
42 activity of late-active taxa whereas temperature was an important driver of early-active taxa.
43 Families of mixed feeders and parasitoids exhibited especially rapid phenological responses to
44 snowmelt and temperature, but with pronounced heterogeneity in responses among habitats. Our
45 findings highlight the complexity in arthropod community phenological responses to global
46 climate change. However, by estimating phenological metrics across entire activity seasons in a
47 functional and life-history trait perspective, general trends and consistent patterns can be identified
48 amidst this complexity.

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62 Introduction

63 The Earth is subject to rapid climatic changes with implications for ecosystems worldwide
64 (Parmesan & Yohe, 2003; Walther et al., 2002). Numerous ecological changes have been
65 associated with global warming (Halsch et al., 2021; Parmesan, 2006), with one of the most
66 widely reported effects being shifts in phenological events (Visser & Both, 2005; Walther,
67 2010). Long-term studies report earlier phenological firsts and peak phenology in response to
68 warming (CaraDonna et al., 2014; Cleland et al., 2007; Inouye, 2008; Menzel et al., 2020;
69 Parmesan & Yohe, 2003), but studies predominantly focus on plants (Clark & May, 2002;
70 Ramula et al., 2015). Being ectothermic, the developmental rate of arthropods is closely coupled
71 to temperature, which facilitates strong responses of arthropods to climate change (Cohen et al.,
72 2018; Halsch et al., 2021; Roslin et al., 2021; Thackeray et al., 2016). Despite their ecological
73 importance, huge diversity, and sensitivity to climate variations (Boggs, 2016; Gillespie et al.,
74 2020; Hodkinson & Jackson, 2005), there is a knowledge gap concerning arthropod community
75 phenological responses to changing growing conditions.

76

77 High-Arctic ecosystems are experiencing rapid climate warming, with average annual
78 temperatures rising at twice the global average rate (AMAP, 2017; Rantanen et al., 2022). While
79 temperature trends are relatively clear, the timing of snowmelt varies substantially from year to
80 year as snowmelt depends on both spring temperature and winter snowfall as well as its
81 distribution, which is influenced by factors such as wind and topography (Bjorkman et al., 2015;
82 Callaghan et al., 2011). Hence, earlier snowmelt in spring varies temporally and spatially
83 (Dauginis & Brown, 2021; Hernández-Henríquez et al., 2015; Pedersen et al., 2016). The
84 environmental condition in Arctic regions presents organisms with a very short window of

85 activity, making species especially vulnerable to the rapid and major changes occurring (Bale &
86 Hayward, 2010; Danks, 2004).

87

88 Warming has been linked to changes in arthropod community composition (Koltz, Schmidt, et
89 al., 2018), increased growth and reproduction (Culler et al., 2015; Høye et al., 2020), range shifts
90 (Chen et al., 2011; Ittonen et al., 2022) and modified phenology (Bolduc et al., 2013; Roslin et
91 al., 2021). At the same time, snowmelt timing, another key driver of Arctic arthropod phenology
92 (Høye & Forchhammer, 2008b; Rixen et al., 2022), determines the start of activity seasons.

93 Earlier snowmelt advances spring or summer phenology, while delayed snowmelt is likely to
94 delay phenology and shorten the period of optimal growth conditions. The interactions between
95 temperature and snowmelt timing makes the analysis and prediction of future phenological
96 patterns of Arctic arthropods complex and requires long time series data.

97

98 Arthropod functional groups and species can exhibit different phenological responses to climate
99 variation depending on specific life-history traits such as voltinism and body size (Diamond et
100 al., 2011; Gillespie et al., 2017; Pacifici et al., 2017). Univoltine species typically respond to
101 warming by advancing and shortening their phenology whereas multivoltine species tend to
102 delay their late-season phenology (Glazaczow et al., 2016). Species may also exhibit different
103 temperature sensitivities (Buckley, 2022; Thackeray et al., 2016). For example, early-active
104 species strongly advance their spring emergence (Roy & Sparks, 2000) while late-active species
105 often delay their fall activity in response to warming (Bartomeus et al., 2011; Gallinat et al.,
106 2015). Furthermore, different phenological responses can be associated with various sensitivities
107 to environmental cues unaffected by climate change, such as photoperiod (Bale et al., 2002;

108 Danks, 2007). Arctic arthropod phenological shifts may also vary between populations of the
109 same species, due to genetic differences between populations (Diez et al., 2012; Primack et al.,
110 2009) or to site-specific differences in the magnitude of warming (Nufio & Buckley, 2019).
111 Taken together, variation in phenological responses across populations, taxa and phenological
112 events, contribute to changes in community-level phenology, highlighting the complexity in
113 understanding and predicting responses to climate change.

114

115 The diverse life-history strategies among species and functional groups could give rise to various
116 phenological responses to warming and changing snowmelt regimes. Changes in the relative
117 timing of activity of individual arthropod taxa compared to other taxa in the community, termed
118 the phenological niche breadth (Post, 2019; Prevey et al., 2019), may increase the temporal
119 overlap and coexistence of species. However, this could also lead to a shorter duration of the
120 entire arthropod community and thereby strongly influence competitive and trophic interactions
121 (Rudolf, 2019; Sanz-Aguilar et al., 2015). On the other hand, differential shifts in the
122 phenological niche might result in temporal mismatches between trophic levels, such as plant-
123 pollinator or predator-prey interactions (Pyke et al., 2016; Reneerkens et al., 2016; Visser &
124 Both, 2005). Assessing the effect and relative importance of abiotic cues on arthropod phenology
125 and duration of activity can improve predictions of how climate change may shape the activity of
126 individual taxa and the arthropod community as well as ecological interactions in a food web.

127

128 Reliable predictions about the impact of climate change on phenology require comprehensive,
129 long-term monitoring schemes (Coulson et al., 2014; Gillespie et al., 2020; Hodkinson &
130 Jackson, 2005; Inouye, 2022). Unfortunately, Arctic regions lack long-term data with sufficient

131 spatial and temporal resolution, resulting in inadequate documentation of seasonal changes in
132 arthropod activity across entire communities. Furthermore, studies investigating how multiple
133 climate predictors influence the variation in arthropod phenology across diverse taxa and
134 phenology metrics, representing the entire active season, are scarce (But see Prather et al., 2023;
135 Roslin et al., 2021). The BioBasis program at Zackenberg in Northeast Greenland has been
136 conducting arthropod monitoring since 1996 (Schmidt et al., 2019), representing the longest-
137 running terrestrial arthropod monitoring program in the Arctic (Summary of monitoring
138 programmes in Gillespie et al., 2020). The program offers a unique opportunity to address
139 knowledge gaps regarding seasonal activity patterns of arthropod taxa in a remote and
140 challenging research area, which can help predict the ecological impacts of climate change.
141 Using this 25-year time series dataset, we explored variations in phenological responses
142 throughout the entire active season of terrestrial and semiaquatic arthropods in different local
143 habitats in relation to snowmelt timing and temperature.

144

145 We use this standardized dataset to test (i) whether arthropod phenology across all available
146 family-habitat combinations have temporally changed over the study period; (ii) whether
147 snowmelt timing or temperature are most closely associated to changes in arthropod phenology
148 (specifically, peak activity and duration of activity); (iii) how snowmelt timing and temperature
149 affect arthropod phenology and whether these relationships vary by habitat type or taxa as well
150 as functional groups and general life-history traits.

151

152 In accordance with findings from prior studies on phenological responses of Arctic arthropods
153 (Høye et al., 2007; Pearce-Higgins et al., 2005; Tulp & Schekkerman, 2008), we hypothesize that

154 earlier snowmelt and warming lead to advances in peak phenology across taxa and habitats.
155 Responses in the duration of activity are expected to be more variable among arthropods, as
156 some taxa will shorten (Culler et al., 2015), not change (Høye et al., 2014) or extend (Høye et al.,
157 2020; Karlsson, 2014) their activity seasons. We anticipate that snowmelt timing will be the
158 primary driver of arthropod phenological variation, rather than temperature (Bowden et al., 2018;
159 Høye et al., 2014; Høye & Forchhammer, 2008b). Advanced phenological shifts are expected
160 with earlier snow melt and warmer temperatures (Høye et al., 2007). The magnitude of
161 phenological responses will likely vary among taxa, making precise predictions challenging. As
162 observed in many previous studies (Bartomeus et al., 2011; Brooks et al., 2014; Pau et al., 2011;
163 Valtonen et al., 2011), we predict that early-active taxa will be more responsive to temperature
164 than late-active. Conversely, late-active taxa will be more responsive to the timing of snowmelt
165 (Høye & Forchhammer, 2008b).

166

167 Materials and methods

168 *Study site and arthropod sampling*

169 Arthropods were collected at the Zackenberg Research Station, located in high-Arctic Northeast
170 Greenland (74°28' N; 20°34' W). The collection of arthropods was conducted from 1996 to 2020
171 by BioBasis Zackenberg within the Greenland Ecosystem Monitoring Program and included
172 seven different plots (Schmidt et al., 2019). Plot 2 - 7 consisted of eight yellow pitfall traps (1997
173 to 2006) later reduced to four pitfall traps (also four traps in 1996 but otherwise from 2007 to
174 2020). Each pitfall sampling plot contained two rows of sections (first row: A – D and second
175 row: E – H) with one trap in each section. After 2006, pitfalls in row E – H were closed. Plot 1
176 included two window traps (flight-interception traps), constructed as a window screen held in

177 place by two angular aluminium bars between two chambers filled with water, detergent and salt,
178 where specifically limnic insect emergence and flight activity was monitored. Individual traps
179 were opened as they became snow-free (usually late May to early June), but for early snow
180 melting habitats (plot 2 and plots 5 & 7), snowmelt often occurred prior to sampling initiation.
181 Traps operated during the growing season and ended by 1st September, which often coincided
182 with freeze up. The traps were emptied weekly at fixed dates, unless the weather prohibited
183 handling the samples in which case the traps were emptied the following day. Further details
184 regarding sampling procedures are given in Schmidt et al. (2019). After collection, specimens
185 were stored in 75% ethanol, and transported to Denmark where the arthropods were sorted by
186 technicians at the Department of Ecoscience at Aarhus University. Spiders and most insects were
187 sorted to family level, Aphidoidea, Chalcidoidea and Coccoidea were sorted to super family
188 level, and other arthropods were sorted to subclass level, and all specimens were subsequently
189 counted. The data is publicly available at <http://data.g-e-m.dk>. As the field season slightly varied
190 from year-to-year depending on several factors such as timing of spring snowmelt as well as
191 logistical challenges involved with initiating and ending the field season, we focused on
192 arthropod data from only June, July and August which enabled comparison of arthropod capture
193 numbers among years. Further, weekly abundance counts for each arthropod group were
194 standardized by calculating individuals per trap per day for each plot.

195

196 The plots represented pond (Plot 1), wet fen (Plot 2), mesic heath (Plots 3 and 4) and arid heath
197 (Plots 5 and 7) habitats where each habitat differed in plant community composition, soil
198 moisture and the timing of snowmelt. The wet fen habitat was primarily dominated by mosses
199 and grasses and has high soil moisture and early snowmelt. The mesic heath habitat was

200 dominated by lichens, *Cassiope tetragona*, *Dryas* sp. and *Salix arctica* and typically had
201 snowmelt two weeks later than the fen and arid heath area. The arid heath habitat was composed
202 primarily of lichens, *Dryas* sp. and grasses, had relatively low soil moisture and experienced
203 early snowmelt. A small islet in a shallow pond represented the pond habitat (Schmidt et al.,
204 2019).

205
206 Our analysis focused on the most abundant arthropod taxa within the community, which we
207 defined as taxa for which at least 50 individuals were caught per plot in a season. The arthropod
208 taxa that were sufficiently represented based on this criterion were soil mites (Acari),
209 collembolans, three families of spiders and the insect orders; Diptera, Hymenoptera, Hemiptera
210 and Lepidoptera (see Supporting Information S1). Some traps showed major spikes in the
211 number of mites caught, which was linked to the capture of mite-infested bumblebees in a trap
212 thus resulting in the capture of several hundred mites. These spike numbers were changed to the
213 average of the other traps in a plot from identical capture periods (see Supporting Information
214 S1). An identical problem with large spikes in abundance estimates was found for the spider
215 family Lycosidae caused by the capture of many juveniles along with female individuals. This
216 was managed by changing spike numbers to the average of the other traps in a plot from the
217 same capture period. Another spider family; Linyphiidae exhibited deviations in annual adult
218 activity patterns causing problems when generating the phenological curves as bimodal
219 distributions were found for some years. The different species of spiders within Linyphiidae may
220 overwinter in different life stages causing bimodal distributions in abundance estimates across a
221 season. A distinction between juveniles and adults was not made in the data for all years and so,
222 abundance estimates for Linyphiidae could not be corrected. However, it was carefully

223 considered if normal distribution curves could be fit on abundance across a season for each year
224 and thus, it was regarded as appropriate to keep this family in the study (see Supporting
225 Information S1). This approach was identical to the approach taken in Høye & Forchhammer
226 (2008b).

227

228 We divided the arthropod taxa in the following broad functional groups; decomposers (Acari and
229 Collembola), predators (Linyphiidae, Lycosidae and Thomisidae), herbivores (Aphidoidea and
230 Coccoidea), parasitoids (Hymenoptera families; Ichneumonidae and Chalcidoidea) and mixed
231 feeders (Diptera families and Lepidoptera). Diptera (flies and mosquitoes) and Lepidoptera
232 (butterflies and moths) were recognized as mixed feeders because of their role as predators and
233 herbivores as larvae and as pollinators and predators in the adult life stage. These groupings were
234 in accordance with previous studies of this high-Arctic arthropod community (Høye et al., 2021,
235 Koltz et al., 2018). Further, we characterized the arthropods as being either surface-dwelling
236 (decomposers, predators and herbivores) or flying insects (parasitoids and mixed feeders). See
237 Supplementary Information S1 for more information.

238

239 *Climate variables*

240 We selected temperature and timing of snowmelt as climate variables to determine the effect of
241 climate on emergence phenology as these variables were previously found to influence the
242 timing of arthropod emergence (Høye & Forchhammer, 2008b; Kankaanpää et al., 2018). A
243 meteorological station located in a mesic heath habitat and within 600 m from all plots operated
244 through the entire study period and measured soil (0-, 5-, and 10-cm depth) and air (2 m above
245 the ground) temperature on an hourly basis (Downloaded: 13th January 2022). Air temperature,

246 rather than soil temperature, was chosen to explain temperature variations in the activity levels of
247 arthropods in this study (see Supporting Information S2 for a detailed justification of the choice
248 of air temperature) by compiling temperature predictors for each phenological event separately.
249 While a growing degree day model might offer a more precise representation of arthropod
250 responses (Cayton et al., 2015), this approach was not feasible due to limited knowledge
251 regarding the lower developmental threshold temperatures for many arthropod taxa in our study
252 and the use of coarse taxonomic resolution. Instead, we focused on the temperature in the period
253 before activity, which has been considered closely related to arthropod development (Gillooly et
254 al., 2002) and may therefore be a good predictor of adult arthropod phenological events. To this
255 end, we determined the temperature predictor as an average over the 30 days leading up to a
256 mean phenological date of an event minus the minimum standard deviation for each taxon-by-
257 plot combination, so as not to use temperature values largely occurring after the phenological
258 event to predict the response (Figure S2.4).

259

260 Timing of snowmelt was estimated as the date by which less than 10 cm of snow was measured
261 by an automatic ultrasonic snow depth sensor installed at the meteorological station (Skov et al.,
262 2020). Years 2009, 2013 and 2019 had limited snow accumulation resulting in the estimation of
263 very early snow melting dates. We used soil temperature data (averaged from 0, 5 and 10 cm
264 depth) to estimate more reliable snow melting dates for the years 2009, 2013 and 2019 as well as
265 1996 where no snow depth data was available, following the method in Rixen et al. (2022). For
266 those four years, we identified the time period towards the end of the winter when ground
267 temperatures were stable near 0 °C and subsequently started fluctuating when the snow cover
268 disappeared (defined as the zero-curtain window). From this, we defined the date of snowmelt as

269 when the mean daily soil temperature rose above +1 °C after a period with diurnal fluctuations of
270 less than 2 K and mean daily temperatures between -1 °C and 1 °C (Rixen et al., 2022). This also
271 enabled us to estimate a snow melt day for the year 1996 where no snow depth data was
272 available. The correlation between day of snowmelt estimated from the snow depth sensor and
273 soil temperature data can be seen in Figure S2.1.

274

275 *Quantifying phenology*

276 Annual onset, peak and end of activity of each arthropod taxa were calculated using generalized
277 additive modelling (GAM) on the abundance per trap per day for each plot in each season. We
278 predicted a non-linear phenological development across each season and used the partial
279 smoothing method in GAMs to model the seasonal development in capture rates (Guisan et al.,
280 2002). GAMs provide fairly accurate estimations of species phenology despite gaps in the
281 distribution due to varying sample size (Moussus et al., 2010). Curves of arthropod abundance
282 were fit across a season using GAM, assuming a poisson distribution with $k = 4$ (basis
283 dimensions) to ensure appropriate smoothing and constrain a symmetric activity curve as well as
284 a log link function. We used the package ‘*mgcv*’ version 1.8-40 to fit the GAMs (Wood, 2017).
285 Some taxa had low capture numbers in some plots in some years, and consequently, we restricted
286 our analyses to years and plots where at least 50 individuals of a given taxa were caught
287 calculated as the sum of individuals across all traps in a plot within a year. Also, the three
288 phenological events should be possible to calculate (the taxa must be present in at least two
289 weeks). Based on these criteria, the GAM curves provided a fit with a normal distribution for
290 most arthropod taxa and an approximately symmetric distribution in activity due to the
291 univoltine nature of most arthropods in the Arctic. Any indication of asymmetric phenological

292 curves or multimodal distributions led us to exclude the taxa in a specific year from the analysis.
293 When filtering for all these criteria, we excluded 768 years of taxa-by-plot combinations of 1,875
294 years in total. Annual onset, peak and end of the activity season were then calculated as the day
295 at which 10%, 50% and 90% of the accumulated abundance (area under the curve) was reached,
296 respectively (Figure S1.1). The duration of the activity period was estimated as the number of
297 days between onset and end of the activity season.

298

299 *Statistical analyses*

300 *Temporal trends in climate and arthropod phenology*

301 Temporal trends in air temperature and timing of snowmelt were calculated as the slope of the
302 regression against year as a continuous predictor using simple linear regression. Additionally,
303 linear regression was used to calculate the temporal trends in the onset, peak and end of
304 arthropod activity as the slope of the regression against year. This was done for each taxon for
305 which it was possible to estimate phenology of the activity period for at least five years across
306 the time series data. Data for analysis of temporal trends was sufficient for 15 taxa (full list
307 available in Supplementary Information S1). To test if temporal trends of arthropod phenology
308 differed among taxa, functional group and habitat, we used linear mixed models with each
309 individual phenological event as the response variable and year, taxa or functional group as
310 predictor. When functional groups were compared, we added random intercepts for taxa to
311 account for nonindependence of observations among taxa within functional groups (Table S5.1).
312 In addition, a year-plot interaction was included to allow for the year effect to vary among sites.

313

314 *Correlation between climate variables*

315 The climate variables included in this study (timing of snowmelt and air temperature) may be
316 highly correlated. Yet, a proper correlation analysis between climate variables was not feasible
317 because specific air temperature predictor values were calculated for the individual average
318 emergence date for each arthropod taxa in each plot. To appropriately examine a potential issue
319 of multicollinearity, variance inflation factors (VIFs) were derived using the R package ‘*car*’
320 version 3.1-1 for timing of snowmelt and temperature in all family and plot combinations for
321 each phenological event (Zuur et al., 2010). VIF values were also derived for timing of snowmelt
322 and soil temperature measured at the same climate station at a depth of 0 – 10 cm to compare the
323 correlation of the two temperature variables with timing of snowmelt. We used a threshold
324 criterion of 5 such that predictors with values above a $VIF > 5$ were considered contributing
325 greatly to multicollinearity (Chatterjee & Hadi, 2013).

326

327 *Effect of environmental predictors on phenological events*

328 To determine the effect of snowmelt and temperature on arthropod phenology, we used timing of
329 snowmelt and temperature as explanatory variables of each phenological response variable while
330 accounting for the random slope and intercept parameters of arthropod taxa and plot in linear
331 mixed models (LMM) using the ‘*lme4*’ R package version 1.1-31 (Bates et al., 2015). As we
332 used taxa and plot specific estimates of the temperature predictor, we separated the within-
333 subject effects from between-subject effects that we could not account for in the mixed models
334 by implementing within-subject centering in the model (van de Pol & Wright, 2009). This was
335 done by subtracting an average temperature value for each arthropod taxa and plot combination
336 from the specific temperature value for each arthropod taxa and plot combination. This new
337 temperature predictor derived expressed the within and between arthropod taxa and plot

338 temperature variation component and was included as a fixed effect in the LMM. We also
339 detrended the climate variables by adding year as a covariate in the model. Interactions between
340 climate variables and plot, taxon and functional group was included in the model (Table S5.1). A
341 significant interaction term indicated that the slope of the linear relationship between
342 phenological response and climate differed between taxa, functional group, or plot.

343
344 To determine the differences in phenology of individual taxa among plots, multiple linear
345 regression analysis with timing of snowmelt and temperature as predictors and plot as a fixed
346 effect was performed (Table 1). Plot was included as a categorical variable and to make the
347 desired comparisons of slopes between plots, customized contrasts were used. For all flying
348 insects, the random nested effect of trap type within plot and the random effect of trap type was
349 included to account for the different methods of trapping in plot 1 (window traps) and plots 2 – 7
350 (pitfall traps). If trap type accounted for much of the variance in the model, trap type was
351 included as a random effect in the final model (Supplementary Information S5.2). We also tested
352 all possible interactions between plots and climate variables. We did not include a random
353 intercept for year, as there was no within-year replication of the site-specific environmental
354 variables. A significant interaction term indicated that the slope of the linear relationship
355 between a phenological response and climate variable differed among habitats. The best model
356 was selected based on lowest AIC scores. The chosen model is clearly indicated for each analysis
357 in tables with results in the supplementary information.

358 **Table 1** – Multiple regression models to determine the effect of climate predictors; timing of snowmelt (Snowmelt,
359 the day of year where snow depth < 10 cm) and temperature (Temp, average temperature 30 days prior to a mean
360 phenological event minus one standard deviation) on the response variable; peak activity or duration of activity
361 (Phenological event). Plot is included as a covariate to determine differences in phenological responses of arthropod
362 taxa among plots. Best model was selected based on lowest AIC score.

Model	Description
Model 1, simple linear regression snowmelt	<i>Snowmelt + Plot</i>
Model 2, simple linear regression temperature	<i>Temp + Plot</i>
Model 3, all covariates linear regression	<i>Snowmelt + Temp + Plot</i>
Model 4, interaction snowmelt and plot	<i>Snowmelt x Plot</i>
Model 5, interaction temperature and plot	<i>Temp x Plot</i>
Model 6, all interactions	<i>Snowmelt x Plot + Temp x Plot</i>
Model 7, three-way interaction	<i>Snowmelt*Temp*Plot</i>

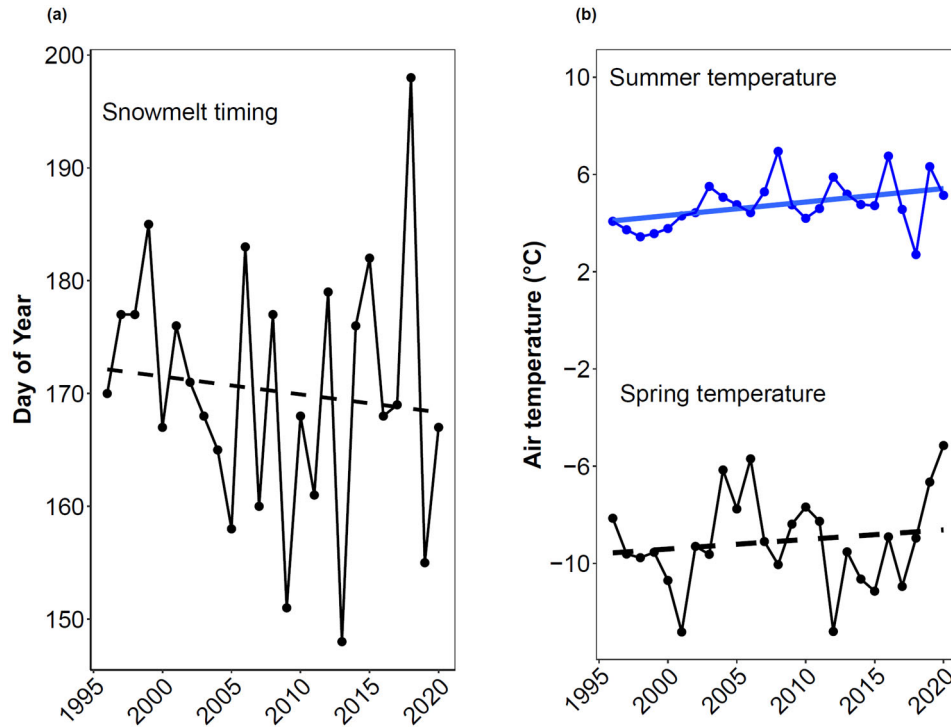
363

364

365 Results

366 *Environmental change*

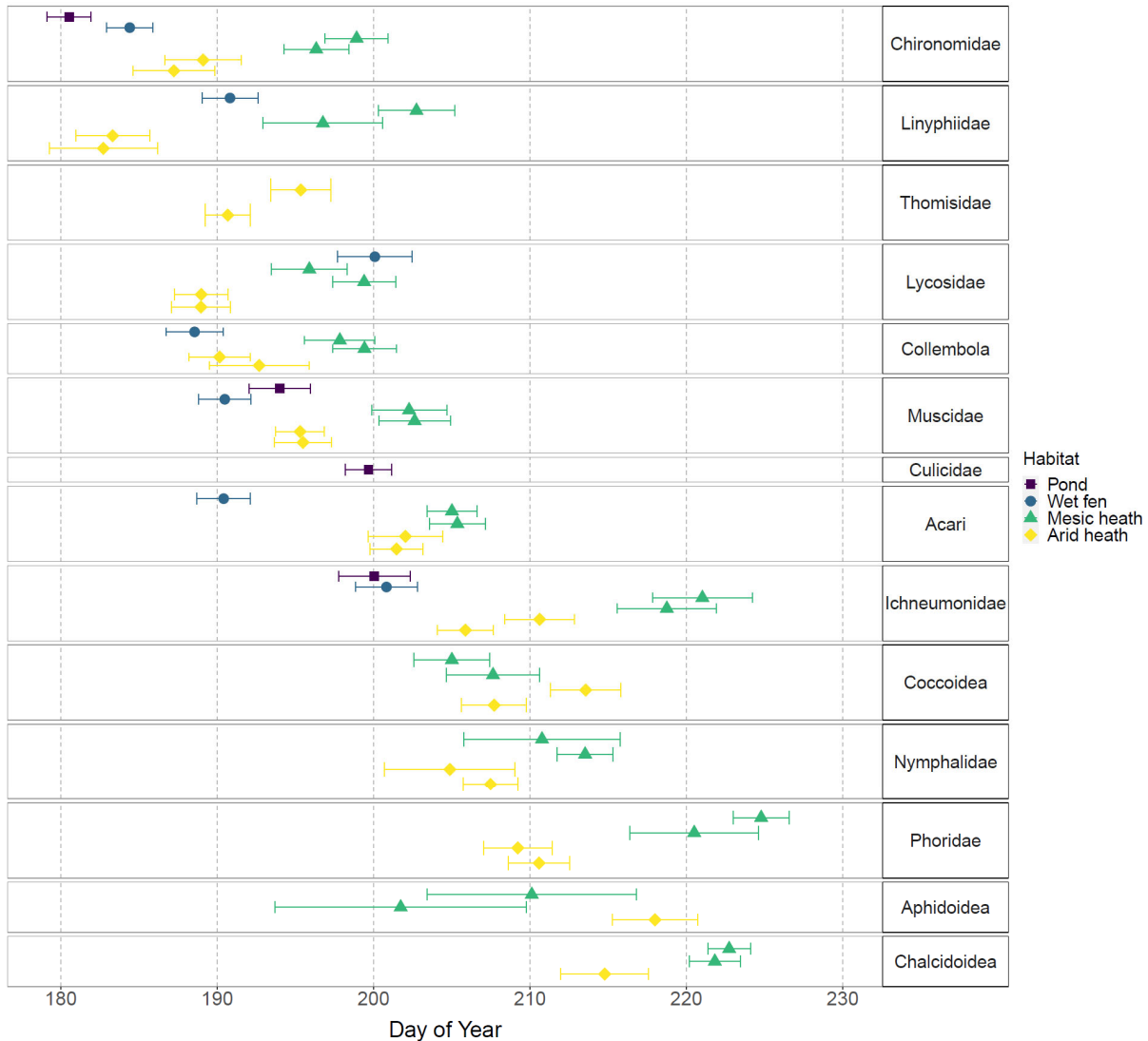
367 Snowmelt timing occurred earlier over the 25-year study period at Zackenberg (Snowmelt: $-1.6 \pm$
368 3.2 days earlier per decade), but this trend was not significant and a substantial amount of
369 among-year variation was found (Figure 1a, $R^2 = 0.01$, $P = 0.62$). Summer air temperature
370 significantly increased by 0.6 ± 0.3 °C per decade during the study period (Figure 1b, $R^2 = 0.16$,
371 $P = 0.045$), whereas spring air temperature did not change significantly (Average spring air
372 temperature: 0.04 ± 0.5 °C per decade). A low level of multicollinearity between timing of
373 snowmelt and air temperature was found (see Supplementary Information S4).



374
 375 **Figure 1** – Interannual variation and trends of (a) timing of snowmelt and (b) average seasonal air temperature for
 376 spring (April, May) and summer (June, July, August) at Zackenberg, Greenland during the study period of 1996 -
 377 2020. Only summer air temperature changed significantly during the study period (indicated by solid regression
 378 line).
 379

380 *The phenological niche of arthropods*

381 The timing of activity varied widely among arthropod taxa and habitats (Figure 2). Families of
 382 mixed feeders are active throughout most of the summer season. As an example, Chironomidae
 383 are present early in the season compared to Phoridae. The parasitoids and herbivores are active
 384 late in the season with considerable variation among habitats. Ichneumonidae peak activity
 385 occurs on average on day 201 in the wet fen with early snowmelt as opposed to day 220 in the
 386 mesic heath with late snowmelt. A longer duration of the activity season is predominant for
 387 decomposers and predators.



388
 389 **Figure 2** – The average peak phenological event (when 50% of the season capture was reached) of all taxa across
 390 multiple years represented by at least 50 individuals per season for each plot (day of year 140 – 240). The arthropods
 391 are listed in order of activity such that the top families are active earlier than the below families. The associated error
 392 bars indicate the standard error of the mean.
 393

394 *Temporal phenological responses*

395 We observed earlier peak activity in the arthropod community over the study period (Figure
 396 S3.1), but only 21% of family-by-plot combinations across onset, peak and end of activity
 397 showed a significant shift over time. The community peak phenology occurred 3.07 ± 0.4 days
 398 earlier per decade ($R^2_{conditional} = 0.56, P = <0.001$) and the community duration of activity
 399 extended by 0.79 ± 0.4 days per decade, although marginally nonsignificant ($R^2_c = 0.63, P =$

400 0.06). While trends indicate that arthropod activity becomes earlier, we observed strong variation
401 in responses among arthropod taxa and habitats in all phenological events (Figure S3.1).
402 Sciaridae (5.1 ± 1.5 days earlier per decade), Nymphalidae (8.3 ± 2.2 days earlier per decade),
403 Lycosidae (4.2 ± 1.3 days earlier per decade) and Collembola (5.9 ± 1.3 days earlier per decade)
404 showed the strongest responses across the study period. While most arthropods showed trends of
405 prolonged or no change in activity periods across the study period, herbivores exhibited trends of
406 shortened activity (7.8 ± 3.8 days per decade).

407

408 *Community phenological responses to snowmelt timing and temperature*

409 We only present results of the effect of climate on the peak and duration (difference between
410 onset and end of activity) of activity to facilitate a better understanding of the results.

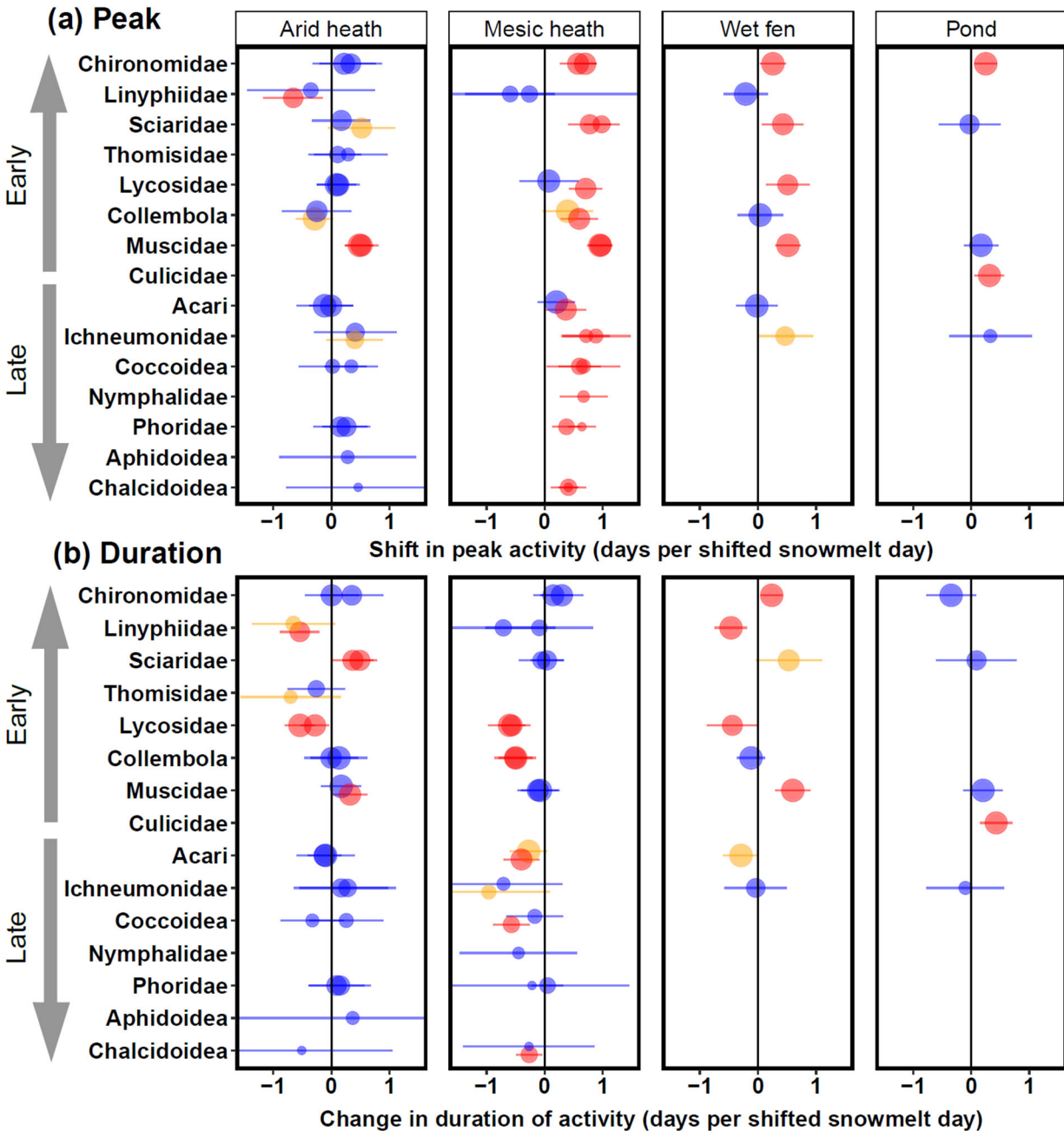
411

412 The average arthropod community peak activity was 0.26 ± 0.03 days earlier per day earlier
413 snowmelt ($R_c^2 = 0.68$, $P = <0.001$). Ten of 15 taxa showed a significant positive effect of
414 snowmelt timing on peak phenology (Table S5.3), indicating that these taxa started their peak
415 activity earlier in response to earlier timing of snowmelt. In response to temperature, the average
416 community peak activity was 1.11 ± 0.46 days earlier per 1°C increase ($R_c^2 = 0.68$, $P = 0.02$). Six
417 of 15 taxa showed a negative effect of temperature on peak phenology, indicating that these taxa
418 started their peak activity earlier in response to warming (Table S5.3).

419

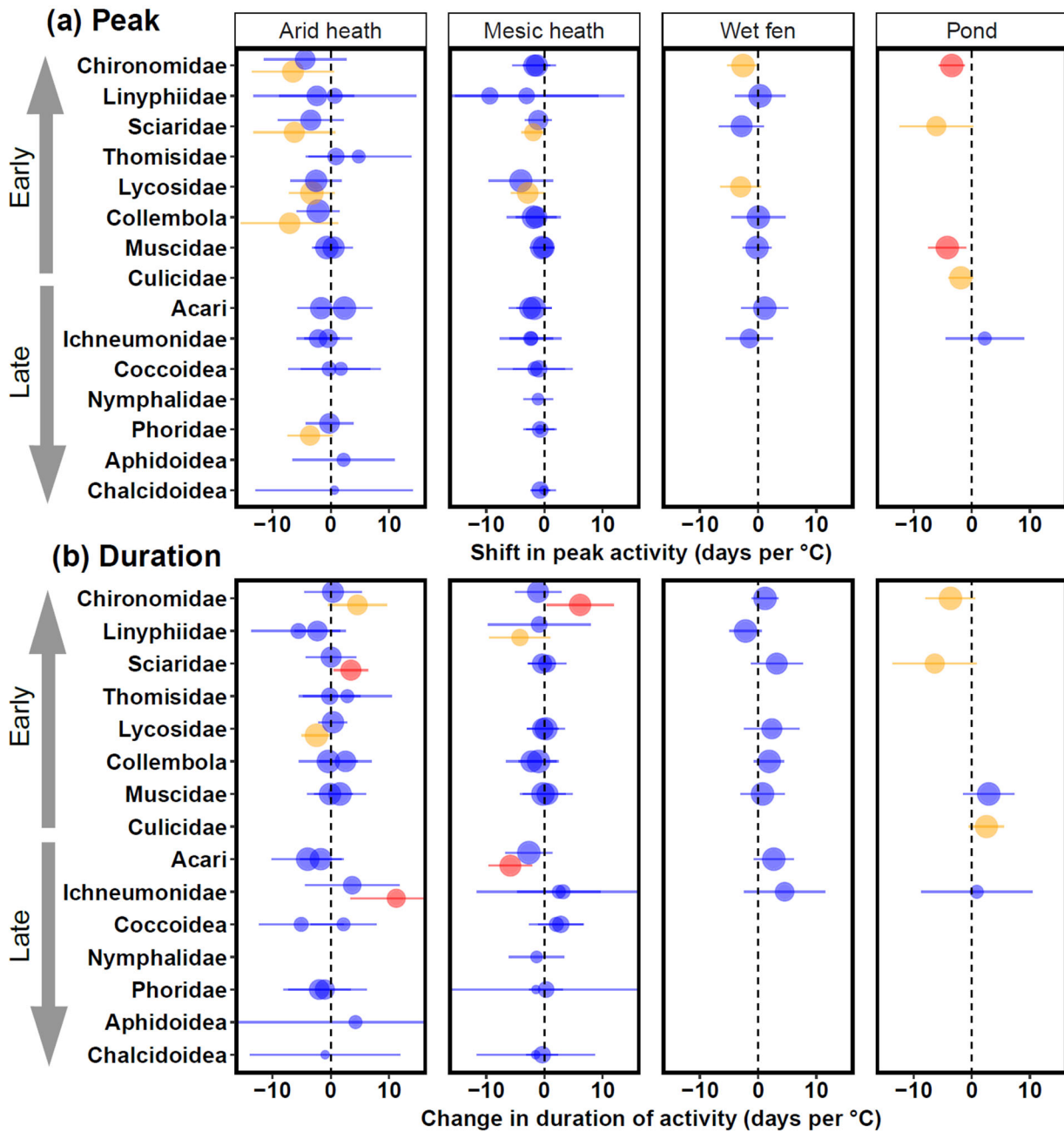
420 Most taxa responded to earlier snowmelt and warming by extending the duration of activity,
421 although the directional shift in responses to timing of snowmelt and temperature varied among
422 taxa and habitats (Figure 3 & 4). Nevertheless, we found that the average duration of activity in

423 the arthropod community in response to snowmelt timing extended at a rate of 0.10 ± 0.03 days
 424 per earlier snowmelt day ($R_c^2 = 0.63$, $P = <0.001$). In response to temperature, the community
 425 activity season did not change.



426
 427 **Figure 3** – The average phenological shift in days per change in snowmelt day in (a) peak and (b) duration of the
 428 activity season for taxa in each habitat. Taxa with peak phenological shifts that fall above zero advance their peak
 429 activity date with earlier snowmelt. Taxa with shifts in duration of activity that fall below zero extend the duration
 430 of activity with earlier snowmelt. Significant shifts are represented in red ($P \leq 0.05$), marginally significant shifts in
 431 orange ($0.05 < P \leq 0.10$) and nonsignificant shifts in blue ($P > 0.10$). The model is controlled for the effect of

432 temperature and year. The arthropods are listed in order of activity such that the top families are active earlier than
 433 the below families. The size of each point represents the number of observations in each family and plot
 434 combination. The associated error bars indicate the 95% confidence interval (CI). To ease comparison, all panels are
 435 equally scaled causing a cut-off of CI's. In the mesic heath and arid heath panels, two points are shown indicating
 436 the two plots in these habitats. If points have different significance levels, the points are slightly separated to ease
 437 visibility.
 438
 439



440
 441 **Figure 4** – The average phenological shift in days per 1 °C in (a) peak and (b) duration of the activity season for
 442 taxa in each habitat. Taxa with peak phenological shifts that fall below zero advance their peak activity with warmer
 443 temperatures. Taxa with shifts in duration of activity that fall above zero extend their activity with warmer

444 temperatures. Significant shifts are represented in red ($P \leq 0.05$), marginally significant shifts in orange ($0.05 < P \leq$
445 0.10) and nonsignificant shifts in blue ($P > 0.05$). The model is controlled for the effect of snowmelt and year. The
446 arthropods are listed in order of activity such that the top families are active earlier than the below families. The size
447 of each point represents the number of observations in each family and plot combination. The associated error bars
448 indicate the 95% confidence interval (CI). To ease comparison, all panels are equally scaled causing a cut-off of
449 CI's. In the mesic heath and arid heath panels, two points are shown indicating the two plots in these habitats. If
450 points have different significance levels, the points are slightly separated to ease visibility.

451
452

453 *Environmental drivers of phenology*

454 Model comparisons with timing of snowmelt and temperature as predictors showed that
455 snowmelt explained more variation in peak phenology than temperature (Table S5.6). In most
456 cases, however, the model including both predictors best explained arthropod phenology. The
457 exception was Aphidoidea, Thomisidae, and Acari (except in the mesic heath) that did not
458 respond strongly to snowmelt timing or temperature in their peak phenology. After detrending
459 the time series by year, the significance of the linear phenology-temperature relationship
460 diminished, while most linear phenology-snowmelt relationships persisted (Figure S6.3, S6.4).
461 Lycosidae was the only taxa for which temperature explained more variation in peak phenology
462 than snowmelt timing ($\Delta AIC > 2$).

463

464 For duration of activity, temperature was found to explain more variation than snowmelt timing
465 (Table S5.6), particularly for flying insects. We found that the model including both predictors
466 best explained arthropod duration of activity. Aphidoidea was the only taxon that did not respond
467 to either snowmelt timing or temperature in duration of activity.

468

469 We also found strong taxon-by-plot differences in the importance of environmental drivers of
470 phenology. Taxa in the pond habitat responded strongly to temperature compared to the other
471 habitats (Table S5.3). Timing of snowmelt was a better predictor of peak phenology in the mesic

472 heath habitat compared to the other habitats.

473

474 *Phenological responses among taxa and habitats to snowmelt timing and temperature*

475 While we found trends of earlier peak activity in the arthropod community in response to earlier
476 snowmelt and increasing temperature, the response varied substantially among taxa and habitats
477 (Figure 3 & 4). Earlier snowmelt caused earlier peak activity of all taxa responding to snowmelt
478 timing, except for Linyphiidae showing delayed activity, but the rate at which arthropods
479 responded with earlier peak activity to earlier snowmelt varied substantially among taxa. Taxa
480 from the mesic heath showed stronger responses of earlier peak activity to earlier snowmelt dates
481 than taxa from the arid heath and wet fen habitats (Figure 3 & Table S5.3). The peak
482 phenological responses to temperature among arthropod taxa was less clear compared to
483 snowmelt timing (Figure 4) and most taxon-by-plot phenological responses were not significant.

484

485 The effect of snowmelt timing and temperature on duration of activity also varied among taxa
486 and habitats, both in the directional effects (whether taxon-by-plot combinations extended or
487 shortened their activity periods) and in the rate of shortened or extended activity periods. For
488 example, Chironomidae and Sciaridae experienced shortened activity periods in the pond habitat
489 in response to warmer temperature, which was not found in other habitats (Figure 4 & Table
490 S5.4). Muscidae and Sciaridae responded with shorter activity periods to earlier snowmelt in the
491 wet fen and arid heath (Table S5.4). Some taxa experienced substantial changes in activity
492 periods to warming. For example, Ichneumonidae extended their activity period with 5.29 ± 1.62
493 days per °C on average across habitats, which was of a greater magnitude than other taxa (Table
494 S5.4). A few taxa exhibited contrasting responses in duration of activity to climate variables.

495 Acari and Linyphiidae extended their activity with earlier snowmelt but shortened their activity
496 in response to warmer temperatures.

497

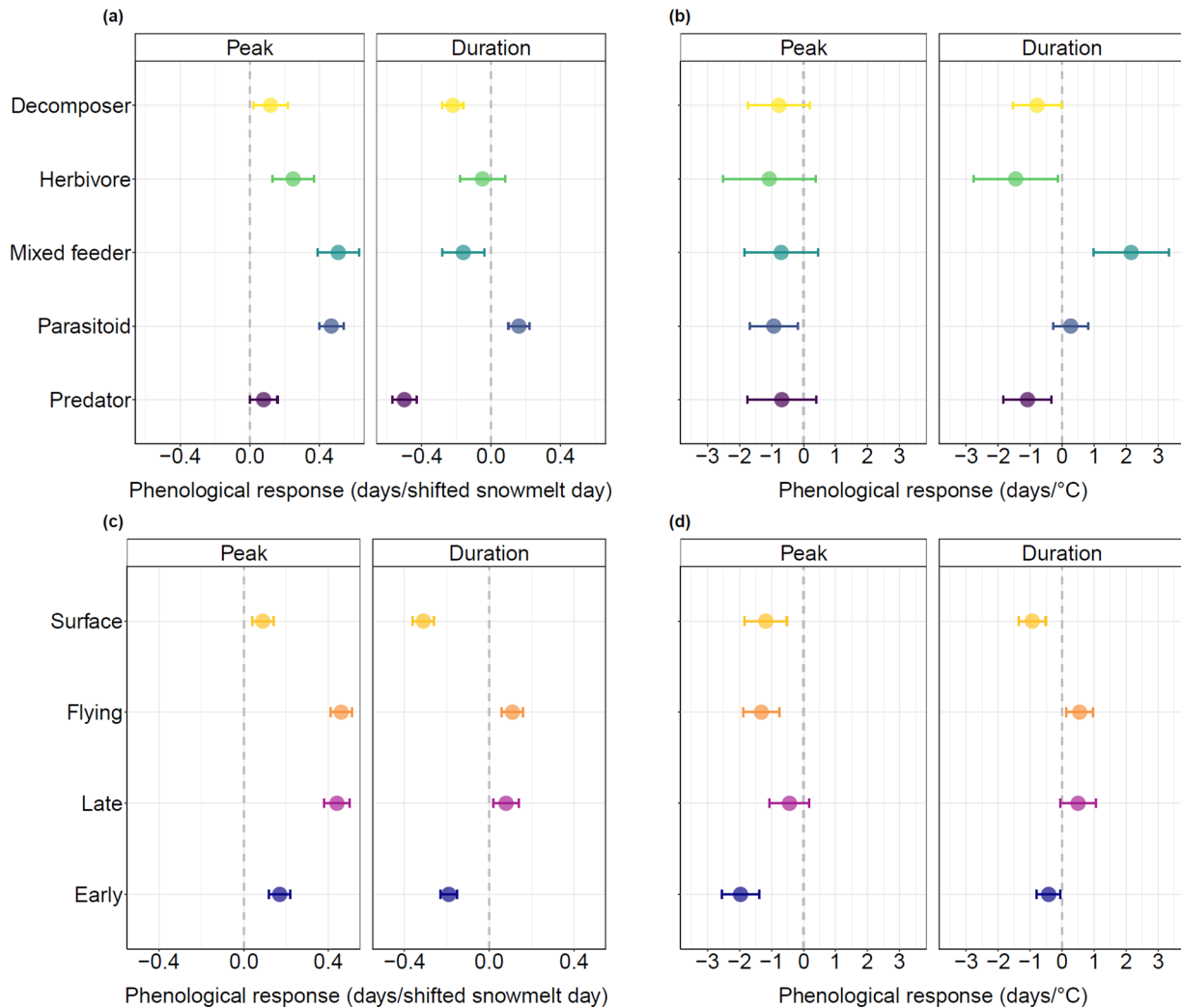
498 *Phenological responses among functional groups and groups with different traits*

499 Mixed feeders and parasitoids shifted their activity to earlier in the season (Mixed feeders, slope:
500 0.45 ± 0.11 ; Parasitoids, slope: 0.44 ± 0.04) at a significantly greater rate in response to earlier
501 snowmelt compared to the other functional groups (Figure 5). Herbivores also responded with
502 earlier peak activity to earlier snowmelt, however, at lower rates (slope: 0.25 ± 0.12).

503

504 Parasitoids was the only group to significantly shorten their duration of activity in response to
505 earlier snowmelt (slope: 0.16 ± 0.06). Predators extended the duration of their activity with
506 earlier snowmelt (slope: -0.50 ± 0.07), but exhibited trends towards shorter activity periods with
507 warming (Figure 5). No functional groups responded significantly in peak activity or duration of
508 activity to temperature (Figure 5), except peak phenology of parasitoids, which became earlier
509 with warming. Our results showed positive trends in the duration of activity of mixed feeders to
510 increasing temperature, suggesting overall extended activity periods with warming. Herbivores
511 showed trends of shorter duration of activity in response to warming, although this trend was not
512 significant.

513



514

515 **Figure 5** – The average phenological shift in (a, c) days per change in snowmelt day and (b, d) days per 1 °C in
 516 duration and peak of the activity season for each functional group (a, b) and arthropods with different traits (c, d).
 517 Average peak phenological shifts that fall above zero in response to snowmelt advance their peak activity with
 518 earlier snow melting days. Average peak phenological shifts that fall below zero in response to temperature advance
 519 their peak activity with warmer temperatures. An average change in duration of activity that fall below zero in
 520 response to snowmelt increase the length of the activity period with earlier snow melting days. An average change in
 521 duration of activity that fall above zero in response to temperature increase the length of the activity period with
 522 warmer temperatures. The associated error bars indicate the standard error of the mean.

523

524 Comparisons of responses between early and late active taxa as well as flying and surface-

525 dwelling taxa were modelled together to consider all interactions and effects. In response to

526 temperature, the peak activity of early active taxa was 1.97 ± 0.58 days earlier per 1 °C increase,

527 while late active taxa were less responsive (Figure 5). Instead, late active taxa were more

528 responsive to timing of snowmelt with peak phenology shifting 0.44 ± 0.06 days earlier per day
529 earlier snowmelt. Similarly, flying insects responded with earlier peak activity by 0.46 ± 0.04
530 days per day earlier snowmelt while surface-dwelling arthropods' peak activity barely changed.

531
532 Taxa with different traits responded very differently to timing of snowmelt and temperature in
533 the duration of phenology (Figure 5). Surface-dwelling and early active taxa extended the
534 duration of their activity period in response to earlier snowmelt, while flying and late active taxa
535 showed trends towards shortened activity in response to earlier snowmelt. At the same time,
536 surface-dwelling and early active taxa exhibited trends towards shortened activity periods to
537 warming, while flying and late active taxa showed trends of extended duration of activity periods
538 in response to warming.

539

540 Discussion

541 Our study highlights three main findings in congruence with our hypotheses:

542 (i) Arthropods displayed weak directional trends towards earlier activity over the study period,
543 with considerable variation in the rate of change among different arthropod taxa and habitats.

544 (ii) Timing of snowmelt was confirmed to be a key driver of arthropod phenology, but
545 temperature explained much of the variation in the duration of activity among taxa compared to
546 timing of snowmelt. Models including both climate variables received the strongest support
547 when testing climate effects on arthropod phenology.

548 (iii) This high-Arctic arthropod community showed responses towards earlier activity with
549 earlier timing of snowmelt and warmer temperatures. However, notable variation in climate-
550 associated phenological shifts were observed among taxa and habitats, both in the direction and

551 magnitude of phenological response. Some of this variation could be explained by the functional
552 roles and general life history traits of arthropods. Late-active taxa responded strongly to
553 snowmelt timing, whereas early-active taxa showed stronger responses to warming. Mixed
554 feeders and parasitoids were especially sensitive to snowmelt timing, showing significantly
555 earlier activity with earlier snowmelt. Predators showed contrasting phenological responses to
556 snowmelt and temperature, indicating that altered snowmelt patterns could counter the effects of
557 warming. This suggests that life-history traits of arthropod taxa may aid in determining
558 sensitivity to climate change, as proposed in other studies (Diamond et al., 2011; Dorian et al.,
559 2022; Pacifici et al., 2017; Zografou et al., 2021).

560

561 *Snowmelt as a key driver of Arctic arthropod phenology*

562 Our study highlights the importance of snowmelt timing as an environmental driver of arthropod
563 phenology, indicating that local snowmelt conditions, even when controlling for the effect of
564 temperature, play a pivotal role in shaping Arctic arthropod responses to global change (Hein et
565 al., 2014; Kankaanpää et al., 2018; Rixen et al., 2022; Saalfeld et al., 2019). Temperature
566 interacts with winter and spring precipitation in determining the duration of the snow-free season
567 (Callaghan et al., 2011), leading to only partial correlation between snowmelt timing and
568 temperature (Bjorkman et al., 2015; Wheeler et al., 2015). As the climate warms, temperature
569 and snowmelt is expected to increasingly decouple in some regions (Callaghan et al., 2011;
570 Raisanen, 2008), resulting in greater variability in snowmelt timing despite generally warmer
571 seasonal temperatures. The strong relationship we identified between phenology and snowmelt is
572 likely due to plastic responses to large temporal variations in snowmelt patterns (Høye et al.,
573 2007; Pedersen et al., 2016). Consequently, the anticipated changes in snowmelt dynamics could

574 enhance landscape-scale heterogeneity, leading to more localised variations in arthropod activity
575 patterns (Gillespie et al., 2016; Leingartner et al., 2014). Taken together, snowmelt currently
576 serves as an important determining factor of phenology and will likely remain essential in
577 shaping Arctic arthropod phenological responses to global change, potentially having
578 implications for arthropod population and community dynamics.

579

580 *Heterogeneity in phenological responses to climate change*

581 Our findings indicate heterogeneity in phenological sensitivity to environmental factors in this
582 high-Arctic arthropod community. A key factor shaping changes in arthropod phenology to
583 changing environmental conditions was their general life-history traits. The timing of snowmelt
584 demonstrated greater explanatory power for late-active arthropod taxa and taxa in habitats (mesic
585 heath) with later snowmelt compared to early-active taxa and taxa in habitats (arid heath and wet
586 fen) with earlier snowmelt (Figure 3 & 5). Late-active taxa thus exhibited greater plasticity to
587 snowmelt and snowmelt timing acts as an important environmental control on phenology.
588 Conversely, we found early-active taxa and taxa from early snow melting habitats to be more
589 responsive to temperature, indicating that temperature serves as a more reliable cue for activity
590 in the beginning of the growing season (Brooks et al., 2014; Roy & Sparks, 2000). This may
591 imply that early-active arthropods are approaching the limits of phenological plasticity to
592 snowmelt (Iler et al., 2013; Jochner et al., 2016; Stemkovski et al., 2023), hindering the ability of
593 arthropods to track the changing climate. Perhaps early-active taxa rely more on a stable abiotic
594 driver such as photoperiod, when snowmelt is early (Bowden et al., 2018; Høye &
595 Forchhammer, 2008a). By adjusting to photoperiod, arthropods can delay their activity until later
596 in the season when temperature conditions are more stable, thereby avoiding exposure to

597 freezing (Gehrmann et al., 2017; Regan & Sheldon, 2023; Rixen et al., 2022).

598

599 We also observed more pronounced advancements in peak phenology among flying and late-

600 active taxa in response to earlier snowmelt. Surface-dwelling and early active taxa prolonged

601 their activity period with earlier snowmelt, but shortened their activity period with warming.

602 Most Arctic arthropods are univoltine (Høye et al., 2020), and because they grow faster with

603 warming (Chaves et al., 2015; Culler et al., 2015), shortened activity periods could be common

604 in the Arctic. However, warmer environments could facilitate the production of additional

605 generations, a consequence of climate change that has been previously demonstrated (Altermatt,

606 2010; Braune et al., 2008; Kerr et al., 2020; Tobin et al., 2008), and has been observed in the

607 wolf spider *Pardosa glacialis* at Zackenberg (Høye et al., 2020). The extended activity periods

608 found in this study for a few functional groups and individual taxa may indicate that patterns of

609 voltinism in Arctic arthropods are shifting with a changing climate. Future studies should

610 therefore consider other types of distributions, such as bimodal distributions, when modelling

611 seasonal activity patterns of species (Hodgson et al., 2011).

612

613 These observations suggest that predicting phenological responses of arthropod taxa to future

614 climate change may depend on general trait characteristics such as early or late activity and

615 surface-dwelling or flying arthropods.

616

617 *Community and ecosystem consequences of climate and phenological changes*

618 We found mixed feeders, parasitoids and herbivores to show strong trends of earlier peak

619 phenology in years with earlier snowmelt. In years with early snowmelt and warmer

620 temperatures, Diptera families (categorized as mixed feeders) exhibited shorter activity periods
621 in certain habitats. Predators showed contradicting responses to climate variables as they
622 prolonged their activity period in early snow melting years but shortened their activity period in
623 response to warming. Herbivores and decomposers showed weaker correlations with climate
624 variables compared to other functional groups. The temporal trends of shorter activity periods
625 exhibited by herbivores could be linked to warming (Figure 5), but we found no significant
626 responses to snowmelt timing or temperature in our study. The length of activity seasons of
627 herbivores could instead be driven by strong interactions with their plant food source (Gillespie
628 et al., 2007; Koltz et al., 2022; Strathdee et al., 1993).

629
630 The diversity in phenological responses associated with climate change in this high-Arctic
631 arthropod community may entail major reorganization within the arthropod community under
632 climate change, potentially impacting interaction networks (Schmidt et al., 2017; Walther, 2010).
633 For instance, earlier pollinator activity could temporally decouple them from flowering host
634 plants, resulting in reduced pollination services (Pyke et al., 2016; Schmidt et al., 2016). Plants at
635 Zackenberg have been found to rapidly advance and shorten their flowering season (Høye et al.,
636 2013; Iler et al., 2017), but it remains uncertain if pollinators shift their phenology sufficiently to
637 track timing of flowering, particularly because plants seem to respond more rapidly to warming
638 than pollinators (mixed feeders) found in this study (Plants: -3.19 ± 0.06 d/°C (Iler et al., 2017);
639 Mixed feeders: -0.79 ± 1.15 d/°C). Extended or shortened predator activity could affect top-down
640 control in the food web (Koltz et al., 2018). Given that arthropod peak activity has shifted to
641 earlier in the season at Zackenberg in the past 25 years, there could be an increased risk of
642 trophic mismatch between Arctic shorebirds' insectivorous chicks and the availability of

643 arthropod prey (Both et al., 2009; Reneerkens et al., 2016). However, the host-parasitoid
644 interaction may remain intact, as parasitoids appear to track their hosts (families of predators and
645 mixed feeders) (Abrego et al., 2021). These findings indicate that climate change in the Arctic
646 may affect trophic interactions and food web dynamics.

647

648 An important caveat when interpreting our results is the rather coarse taxonomic resolution used
649 in our study, which may obscure important species-level variations. At the family-level
650 taxonomic resolution, we cannot differentiate between the effects of changing species
651 composition between years and the interannual variation in the phenology of individual species.
652 Consequently, we may derive erroneous conclusions, such as estimating an advanced emergence
653 date in early snow melting years when, in reality, early emerging species become more common.
654 Conversely, among the more diverse arthropod taxa, there may be underlying variations in
655 interspecific phenological responses to abiotic drivers that, when examined at a broader
656 taxonomic resolution, might appear as if there is no response to climate change. However, our
657 aim with the present study was to elucidate broader patterns in the phenological responses of
658 various taxonomical groups using robust abundance estimates allowing us to analyse phenology
659 and climate relationships with greater statistical power. In addition, it is essential to attain a
660 better understanding of community-level responses to climate change on various taxonomic
661 levels (Koltz, Schmidt, et al., 2018; Legagneux et al., 2014), because changes in the structure and
662 seasonal dynamics of broader taxonomic groups affect how ecosystems function.

663

664 *Conclusions*

665 Our study highlights that phenological responses across taxa and populations within an arthropod

666 community exposed to changing climatic conditions are complex and does not follow broad
667 overall patterns. Therefore, to gain a comprehensive understanding and make credible
668 predictions regarding future changes in arthropod seasonal activity, detailed investigations into
669 the environmental drivers of change at high spatial and taxonomic resolution is critical. We were
670 able to identify some general trends and consistent patterns in phenological responses to a
671 changing climate in this high-Arctic arthropod community. We found that snowmelt timing is an
672 important determinant of arthropod phenology, but the relative importance of snowmelt timing
673 and temperature as predictors of arthropod phenology varies among taxa and local habitats. The
674 magnitude of phenological responses to climate predictors also varied by taxa and habitat,
675 indicating that general predictions of phenological change for all taxa and locations is not
676 feasible. However, including information on functional groups and life-history traits will likely
677 strengthen our ability to predict future phenological trends of arthropods in cold regions.

678

679

680 Supplementary material

681 All supplementary information has been provided in separate files.

682

683 **Author contributions**

684 H.S.G.: Conceived the study, conducted data analyses, statistics, and drafted the manuscript.

685 T.T.H.: Acquired funding, participated in the design of the study, and in drafting the manuscript.

686 M.H.: Acquired funding and participated in drafting the manuscript. N.M.S.: Oversaw collection
687 of the field data.

688 All authors participated with intellectual contributions and revised the manuscript.

689

690 **Data accessibility**

691 Arthropod monitoring data and climate data is available through the open-source GEM-database:
692 <https://data.g-e-m.dk/>. The R-scripts necessary to replicate the findings of this manuscript are
693 available:

694 https://github.com/soerinegerlich/high_arctic_arthropod_phenology_manuscript

695

696 **Competing Interest Statement**

697 We have no conflict of interest to declare.

698

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709 References

- 710 Abrego, N., Roslin, T., Huotari, T., Ji, Y., Schmidt, N. M., Wang, J., Yu, D. W. & Ovaskainen,
711 O. (2021). Accounting for species interactions is necessary for predicting how arctic
712 arthropod communities respond to climate change. *Ecography*, 44, 1-12. doi:
713 10.1111/ecog.05547
- 714 Altermatt, F. (2010). Climatic warming increases voltinism in European butterflies and moths.
715 *Proc. R. Soc. B*, 277(1685), 1281-1287. <https://doi.org/10.1098/rspb.2009.1910>
- 716 AMAP. (2017). *Snow, Water, Ice and Permafrost in the Arctic (SWIPA) 2017*. A. M. a. A. P.
717 (AMAP).
- 718 Bale, J. S., & Hayward, S. A. L. (2010). Insect overwintering in a changing climate. *Journal of*
719 *Experimental Biology*, 213(6), 980-994. <https://doi.org/10.1242/jeb.037911>
- 720 Bale, J. S., Masters, G. J., Hodkinson, I. D., Awmack, C., Bezemer, T. M., Brown, V. K.,
721 Butterfield, J., Buse, A., Coulson, J. C., Farrar, J., Good, J. E. G., Harrington, R., Hartley,
722 S., Jones, T. H., Lindroth, R. L., Press, M. C., Symrnioudis, I., Watt, A. D., & Whittaker,
723 J. B. (2002). Herbivory in global climate change research: direct effects of rising
724 temperature on insect herbivores. *Glob Change Biol*, 8(1), 1-16. <https://doi.org/DOI>
725 10.1046/j.1365-2486.2002.00451.x
- 726 Bartomeus, I., Ascher, J. S., Wagner, D., Danforth, B. N., Colla, S., Kornbluth, S., & Winfree, R.
727 (2011). Climate-associated phenological advances in bee pollinators and bee-pollinated

728 plants. *Proc Natl Acad Sci U S A*, 108(51), 20645-20649.
729 <https://doi.org/10.1073/pnas.1115559108>

730 Bates, D., Machler, M., Bolker, B. M., & Walker, S. C. (2015). Fitting Linear Mixed-Effects
731 Models Using lme4. *J. Stat. Softw.*, 67(1), 1-48. <https://doi.org/DOI>
732 10.18637/jss.v067.i01

733 Bjorkman, A. D., Elmendorf, S. C., Beamish, A. L., Vellend, M., & Henry, G. H. R. (2015).
734 Contrasting effects of warming and increased snowfall on Arctic tundra plant phenology
735 over the past two decades. *Glob Change Biol*, 21(12), 4651-4661.
736 <https://doi.org/10.1111/gcb.13051>

737 Boggs, C. L. (2016). The fingerprints of global climate change on insect populations. *Curr Opin*
738 *Insect Sci*, 17, 69-73. <https://doi.org/10.1016/j.cois.2016.07.004>

739 Bolduc, E., Casajus, N., Legagneux, P., McKinnon, L., Gilchrist, H. G., Leung, M., Morrison, R.
740 I. G., Reid, D., Smith, P. A., Buddle, C. M., & Bety, J. (2013). Terrestrial arthropod
741 abundance and phenology in the Canadian Arctic: modelling resource availability for
742 Arctic-nesting insectivorous birds. *Can Entomol*, 145(2), 155-170.
743 <https://doi.org/10.4039/tce.2013.4>

744 Both, C., van Asch, M., Bijlsma, R. G., van den Burg, A. B., & Visser, M. E. (2009). Climate
745 change and unequal phenological changes across four trophic levels: constraints or
746 adaptations? *J Anim Ecol*, 78(1), 73-83. [https://doi.org/10.1111/j.1365-](https://doi.org/10.1111/j.1365-2656.2008.01458.x)
747 2656.2008.01458.x

748 Bowden, J. J., Hansen, O. L. P., Olsen, K., Schmidt, N. M., & Høye, T. T. (2018). Drivers of
749 inter-annual variation and long-term change in High-Arctic spider species abundances.
750 *Polar Biol.*, 41(8), 1635-1649. <https://doi.org/10.1007/s00300-018-2351-0>

751 Braune, E., Richter, O., Sondgerath, D., & Suhling, F. (2008). Voltinism flexibility of a riverine
752 dragonfly along thermal gradients. *Glob Change Biol*, 14(3), 470-482.
753 <https://doi.org/10.1111/j.1365-2486.2007.01525.x>

754 Brooks, S. J., Self, A., Toloni, F., & Sparks, T. (2014). Natural history museum collections
755 provide information on phenological change in British butterflies since the late-
756 nineteenth century. *Int. J. Biometeorol*, 58(8), 1749-1758.
757 <https://doi.org/10.1007/s00484-013-0780-6>

758 Buckley, L. B. (2022). Temperature-sensitive development shapes insect phenological responses
759 to climate change. *Curr Opin Insect Sci*, 52, 100897.
760 <https://doi.org/10.1016/j.cois.2022.100897>

761 Callaghan, T. V., Johansson, M., Brown, R. D., Groisman, P. Y., Labba, N., Radionov, V.,
762 Barry, R. G., Bulygina, O. N., Essery, R. L. H., Frolov, D. M., Golubev, V. N., Grenfell,
763 T. C., Petrushina, M. N., Razuvaev, V. N., Robinson, D. A., Romanov, P., Shindell, D.,
764 Shmakin, A. B., Sokratov, S. A., . . . Yang, D. Q. (2011). The Changing Face of Arctic
765 Snow Cover: A Synthesis of Observed and Projected Changes. *Ambio*, 40, 17-31.
766 <https://doi.org/10.1007/s13280-011-0212-y>

767 CaraDonna, P. J., Iler, A. M., & Inouye, D. W. (2014). Shifts in flowering phenology reshape a
768 subalpine plant community. *Proceedings of the National Academy of Sciences of the*
769 *United States of America*, 111(13), 4916-4921. <https://doi.org/10.1073/pnas.1323073111>

770 Cayton, H. L., Haddad, N. M., Gross, K., Diamond, S. E., & Ries, L. (2015). Do growing degree
771 days predict phenology across butterfly species? *Ecology*, 96(6), 1473-1479.
772 <https://doi.org/10.1890/15-0131.1>

773 Chatterjee, S., & Hadi, A. S. (2013). Regression Analysis by Example, 5th edition. *Int Stat Rev*,

774 81(2), 308-308. https://doi.org/10.1111/insr.12020_2

775 Chaves, L. F., Imanishi, N., & Hoshi, T. (2015). Population dynamics of *Armigeres subalbat*
776 (Diptera: Culicidae) across a temperate altitudinal gradient. *Bull. Entomol. Res*, 105(5),
777 589-597. <https://doi.org/10.1017/S0007485315000474>

778 Chen, I. C., Hill, J. K., Ohlemuller, R., Roy, D. B., & Thomas, C. D. (2011). Rapid Range Shifts
779 of Species Associated with High Levels of Climate Warming. *Science*, 333(6045), 1024-
780 1026. <https://doi.org/10.1126/science.1206432>

781 Clark, J. A., & May, R. M. (2002). Taxonomic bias in conservation research. *Science*,
782 297(5579), 191-192. DOI: 10.1126/science.297.5579.191b

783 Cleland, E. E., Chuine, I., Menzel, A., Mooney, H. A., & Schwartz, M. D. (2007). Shifting plant
784 phenology in response to global change. *Trends Ecol Evol*, 22(7), 357-365.
785 <https://doi.org/10.1016/j.tree.2007.04.003>

786 Cohen, J. M., Lajeunesse, M. J., & Rohr, J. R. (2018). A global synthesis of animal phenological
787 responses to climate change. *Nat Clim Change*, 8(3), 224-+.
788 <https://doi.org/10.1038/s41558-018-0067-3>

789 Coulson, S. J., Convey, P., Aakra, K., Aarvik, L., Avila-Jimenez, M. L., Babenko, A., Biersma,
790 E. M., Bostrom, S., Brittain, J. E., Carlsson, A. M., Christoffersen, K., De Smet, W. H.,
791 Ekrem, T., Fjellberg, A., Fureder, L., Gustafsson, D., Gwiazdowicz, D. J., Hansen, L. O.,
792 Holmstrup, M., . . . Zmudzynska-Skarbek, K. (2014). The terrestrial and freshwater
793 invertebrate biodiversity of the archipelagoes of the Barents Sea, Svalbard, Franz Josef
794 Land and Novaya Zemlya. *Soil Bio Biochem* 68, 440-470.
795 <https://doi.org/10.1016/j.soilbio.2013.10.006>

796 Culler, L. E., Ayres, M. P., & Virginia, R. A. (2015). In a warmer Arctic, mosquitoes avoid
797 increased mortality from predators by growing faster. *Proc. R. Soc. B*, 282(1815).
798 <https://doi.org/ARTN 20151549 10.1098/rspb.2015.1549>

799 Danks, H. V. (2004). Seasonal adaptations in arctic insects. *Integr Comp Biol*, 44(2), 85-94.
800 <https://doi.org/10.1093/icb/44.2.85>

801 Danks, H. V. (2007). The elements of seasonal adaptations in insects. *Can Entomol*, 139(1), 1-
802 44. <https://doi.org/DOI 10.4039/n06-048>

803 Dauginis, A. L. A., & Brown, L. C. (2021). Sea ice and snow phenology in the Canadian Arctic
804 Archipelago from 1997 to 2018. *Arct Sci*, 7(1), 182-207. <https://doi.org/10.1139/as-2020-0024>

805

806 Diamond, S. E., Frame, A. M., Martin, R. A., & Buckley, L. B. (2011). Species' traits predict
807 phenological responses to climate change in butterflies. *Ecology*, 92(5), 1005-1012.
808 <https://doi.org/doi.org/10.1890/10-1594.1>

809 Diez, J. M., Ibanez, I., Miller-Rushing, A. J., Mazer, S. J., Crimmins, T. M., Crimmins, M. A.,
810 Bertelsen, C. D., & Inouye, D. W. (2012). Forecasting phenology: from species
811 variability to community patterns. *Ecol Lett*, 15(6), 545-553.
812 <https://doi.org/10.1111/j.1461-0248.2012.01765.x>

813 Dorian, N. N., McCarthy, M. W., & Crone, E. E. (2022). Ecological traits explain long-term
814 phenological trends in solitary bees. *J Anim Ecol*. <https://doi.org/10.1111/1365-2656.13778>

815

816 Gallinat, A. S., Primack, R. B., & Wagner, D. L. (2015). Autumn, the neglected season in
817 climate change research. *Trends Ecol Evol*, 30(3), 169-176.
818 <https://doi.org/10.1016/j.tree.2015.01.004>

819 Gehrman, F., Hanninen, H., Liu, C., & Saarinen, T. (2017). Phenological responses to small-

820 scale spatial variation in snowmelt timing reveal compensatory and conservative
821 strategies in subarctic-alpine plants. *Plant Ecol Divers*, 10(5-6), 453-468.
822 <https://doi.org/10.1080/17550874.2018.1428693>

823 Gillespie, M., Hodkinson, I. D., Cooper, E. J., Bird, J. M., & Jonsdottir, I. S. (2007). Life history
824 and host-plant relationships of the rare endemic Arctic aphid *Acyrtosiphon calvulus* in a
825 changing environment. *Entomol Exp Appl*, 123(3), 229-237.
826 <https://doi.org/10.1111/j.1570-7458.2007.00547.x>

827 Gillespie, M. A. K., Alfredsson, M., Barrio, I. C., Bowden, J., Convey, P., Coulson, S. J., Culler,
828 L. E., Dahl, M. T., Daly, K. M., Koponen, S., Loboda, S., Marusik, Y., Sandstrom, J. P.,
829 Sikes, D. S., Slowik, J., & Høye, T. T. (2020). Circumpolar terrestrial arthropod
830 monitoring: A review of ongoing activities, opportunities and challenges, with a focus on
831 spiders. *Ambio*, 49(3), 704-717. <https://doi.org/10.1007/s13280-019-01185-y>

832 Gillespie, M. A. K., Baggesen, N., & Cooper, E. J. (2016). High Arctic flowering phenology and
833 plant-pollinator interactions in response to delayed snow melt and simulated warming.
834 *Environ. Res. Lett*, 11(11). <https://doi.org/10.1088/1748-9326/11/11/115006>

835 Gillespie, M. A. K., Birkemoe, T., & Sverdrup-Thygeson, A. (2017). Interactions between body
836 size, abundance, seasonality, and phenology in forest beetles. *Ecol Evol*, 7(4), 1091-1100.
837 <https://doi.org/10.1002/ece3.2732>

838 Gillooly, J. F., Charnov, E. L., West, G. B., Savage, V. M., & Brown, J. H. (2002). Effects of
839 size and temperature on developmental time. *Nature*, 417(6884), 70-73.
840 <https://doi.org/10.1038/417070a>

841 Glazaczow, A., Orwin, D., & Bogdziewicz, M. (2016). Increased temperature delays the late-
842 season phenology of multivoltine insect. *Sci Rep*, 6. <https://doi.org/10.1038/srep38022>

843 Guisan, A., Edwards, T. C., & Hastie, T. (2002). Generalized linear and generalized additive
844 models in studies of species distributions: setting the scene. *Ecol Modell*, 157(2-3), 89-
845 100. [https://doi.org/10.1016/S0304-3800\(02\)00204-1](https://doi.org/10.1016/S0304-3800(02)00204-1)

846 Halsch, C. A., Shapiro, A. M., Fordyce, J. A., Nice, C. C., Thorne, J. H., Waetjen, D. P., &
847 Forister, M. L. (2021). Insects and recent climate change. *Proc Natl Acad Sci U S A*,
848 118(2). <https://doi.org/10.1073/pnas.2002543117>

849 Hein, N., Feilhauer, H., Finch, O.-D., Schmidlein, S., & Löffler, J. (2014). Snow cover
850 determines the ecology and biogeography of spiders (Araneae) in alpine tundra
851 ecosystems. *Erdkunde*, 157-172. <https://doi.org/10.3112/erdkunde.2014.03.01>

852 Hernández-Henríquez, M. A., Déry, S. J., & Derksen, C. (2015). Polar amplification and
853 elevation-dependence in trends of Northern Hemisphere snow cover extent, 1971-2014.
854 *Environ. Res. Lett*, 10(4). <https://doi.org/10.1088/1748-9326/10/4/044010>

855 Hodgson, J. A., Thomas, C. D., Oliver, T. H., Anderson, B. J., Brereton, T. M., & Crone, E. E.
856 (2011). Predicting insect phenology across space and time. *Glob Change Biol*, 17(3),
857 1289-1300. <https://doi.org/10.1111/j.1365-2486.2010.02308.x>

858 Hodkinson, I. D. & Jackson, J. K. (2005). Terrestrial and freshwater invertebrates as
859 Bioindicators for Environmental Monitoring, with Particular Reference to Mountain
860 Ecosystems. *Environ. Manage*, 35(5), 649-666.

861 Høye, T. T., Eskildsen, A., Hansen, R. R., Bowden, J. J., Schmidt, N. M., & Kissling, W. D.
862 (2014). Phenology of high-arctic butterflies and their floral resources: Species-specific
863 responses to climate change. *Curr Zool*, 60(2), 243 - 251.

864 Høye, T. T., & Forchhammer, M. C. (2008a). The influence of weather conditions on the activity
865 of high-arctic arthropods inferred from long-term observations. *BMC Ecol*, 8, 8.

866 <https://doi.org/10.1186/1472-6785-8-8>

867 Høye, T. T., & Forchhammer, M. C. (2008b). Phenology of High-Arctic Arthropods: Effects of
868 Climate on Spatial, Seasonal, and Inter-Annual Variation. *Adv. Ecol. Res*, 299-324.
869 [https://doi.org/10.1016/s0065-2504\(07\)00013-x](https://doi.org/10.1016/s0065-2504(07)00013-x) (Advances in Ecological Research)

870 Høye, T. T., Kresse, J. C., Koltz, A. M., & Bowden, J. J. (2020). Earlier springs enable high-
871 Arctic wolf spiders to produce a second clutch. *Proc Biol Sci*, 287(1929), 20200982.
872 <https://doi.org/10.1098/rspb.2020.0982>

873 Høye, T. T., Post, E., Meltøfte, H., Schmidt, N. M., & Forchhammer, M. C. (2007). Rapid
874 advancement of spring in the High Arctic. *Curr Biol*, 17(12), R449-R451.
875 <https://doi.org/https://doi.org/10.1016/j.cub.2007.04.047>

876 Iler, A. M., Høye, T. T., Inouye, D. W., & Schmidt, N. M. (2013). Nonlinear flowering responses
877 to climate: are species approaching their limits of phenological change? *Philos Trans R*
878 *Soc Lond B Biol Sci*, 368(1624), 20120489. <https://doi.org/10.1098/rstb.2012.0489>

879 Iler, A.M., Inouye, D. W., Schmidt, N. M., & Høye, T. T. (2017). Detrending phenological time
880 series improves climate-phenology analyses and reveals evidence of plasticity. *Ecology*,
881 98(3), 647-655. <https://doi.org/10.1002/ecy.1690>

882 Inouye, D. W. (2008). Effects of climate change on phenology, frost damage, and floral
883 abundance of montane wildflowers. *Ecology*, 89(2), 353-362. <https://doi.org/Doi>
884 10.1890/06-2128.1

885 Inouye, D. W. (2022). Climate change and phenology. *Wiley Interdisciplinary Reviews-Climate*
886 *Change*. <https://doi.org/10.1002/wcc.764>

887 Ittonen, M., Hagelin, A., Wiklund, C., & Gotthard, K. (2022). Local adaptation to seasonal cues
888 at the fronts of two parallel, climate-induced butterfly range expansions. *Ecol Lett*, 25(9),
889 2022-2033. <https://doi.org/10.1111/ele.14085>

890 Jochner, S., Sparks, T. H., Laube, J., & Menzel, A. (2016). Can we detect a nonlinear response to
891 temperature in European plant phenology? *Int. J. Biometeorol*, 60(10), 1551-1561.
892 <https://doi.org/10.1007/s00484-016-1146-7>

893 Kankaanpää, T., Skov, K., Abrego, N., Lund, M., Schmidt, N. M., & Roslin, T. (2018).
894 Spatiotemporal snowmelt patterns within a high Arctic landscape, with implications for
895 flora and fauna. *Arct. Antarct. Alp. Res*, 50(1).
896 <https://doi.org/10.1080/15230430.2017.1415624>

897 Karlsson, B. (2014). Extended season for northern butterflies. *Int. J. Biometeorol*, 58(5), 691-
898 701. <https://doi.org/10.1007/s00484-013-0649-8>

899 Kerr, N. Z., Wepprich, T., Grevstad, F. S., Dopman, E. B., Chew, F. S., & Crone, E. E. (2020).
900 Developmental trap or demographic bonanza? Opposing consequences of earlier
901 phenology in a changing climate for a multivoltine butterfly. *Glob Change Biol*, 26(4),
902 2014-2027. <https://doi.org/10.1111/gcb.14959>

903 Koltz, A. M., Classen, A. T., & Wright, J. P. (2018). Warming reverses top-down effects of
904 predators on belowground ecosystem function in Arctic tundra. *Proc Natl Acad Sci U S*
905 *A*, 115(32), E7541-E7549. <https://doi.org/10.1073/pnas.1808754115>

906 Koltz, A. M., Gough, L., & McLaren, J. R. (2022). Herbivores in Arctic ecosystems: Effects of
907 climate change and implications for carbon and nutrient cycling. *Ann N Y Acad Sci*,
908 1516(1), 28-47. <https://doi.org/10.1111/nyas.14863>

909 Koltz, A. M., Schmidt, N. M., & Høye, T. T. (2018). Differential arthropod responses to
910 warming are altering the structure of Arctic communities. *R Soc Open Sci*, 5(4), 171503.
911 <https://doi.org/10.1098/rsos.171503>

- 912 Legagneux, P., Gauthier, G., Lecomte, N., Schmidt, N. M., Reid, D., Cadieux, M. C., Berteaux,
913 D., Bety, J., Krebs, C. J., Ims, R. A., Yoccoz, N. G., Morrison, R. I. G., Leroux, S. J.,
914 Loreau, M., & Gravel, D. (2014). Arctic ecosystem structure and functioning shaped by
915 climate and herbivore body size. *Nat Clim Change*, 4(5), 379-383.
916 <https://doi.org/10.1038/Nclimate2168>
- 917 Leingartner, A., Krauss, J., & Steffan-Dewenter, I. (2014). Elevation and experimental snowmelt
918 manipulation affect emergence phenology and abundance of soil-hibernating arthropods.
919 *Ecol Entomol*, 39(4), 412-418. <https://doi.org/10.1111/een.12112>
- 920 Menzel, A., Yuan, Y., Matiu, M., Sparks, T., Scheifinger, H., Gehrig, R., & Estrella, N. (2020).
921 Climate change fingerprints in recent European plant phenology. *Glob Chang Biol*.
922 <https://doi.org/10.1111/gcb.15000>
- 923 Moussus, J. P., Julliard, R., & Jiguet, F. (2010). Featuring 10 phenological estimators using
924 simulated data. *Methods Ecol. Evol*, 1(2), 140-150. <https://doi.org/10.1111/j.2041-210X.2010.00020.x>
- 926 Nufio, C. R., & Buckley, L. B. (2019). Grasshopper phenological responses to climate gradients,
927 variability, and change. *Ecosphere*, 10(9). <https://doi.org/ARTN e02866>
928 10.1002/ecs2.2866
- 929 Pacifici, M., Visconti, P., Butchart, S. H. M., Watson, J. E. M., Cassola, F. M., & Rondinini, C.
930 (2017). Species' traits influenced their response to recent climate change. *Nat Clim*
931 *Change*, 7(3), 205-+. <https://doi.org/10.1038/Nclimate3223>
- 932 Parmesan, C. (2006). Ecological and evolutionary responses to recent climate change. *Annu. Rev.*
933 *Ecol. Evol. Syst*, 37, 637-669. <https://doi.org/10.1146/annurev.ecolsys.37.091305.110100>
- 934 Parmesan, C., & Yohe, G. (2003). A globally coherent fingerprint of climate change impacts
935 across natural systems. *Nature*, 421(6918), 37-42. <https://doi.org/10.1038/nature01286>
- 936 Pau, S., Wolkovich, E. M., Cook, B. I., Davies, T. J., Kraft, N. J. B., Bolmgren, K., Betancourt,
937 J. L., & Cleland, E. E. (2011). Predicting phenology by integrating ecology, evolution
938 and climate science. *Glob Change Biol*, 17(12), 3633-3643.
939 <https://doi.org/10.1111/j.1365-2486.2011.02515.x>
- 940 Pearce-Higgins, J. W., Yalden, D. W., & Whittingham, M. J. (2005). Warmer springs advance
941 the breeding phenology of golden plovers *Pluvialis apricaria* and their prey (Tipulidae).
942 *Oecologia*, 143(3), 470-476. <https://doi.org/10.1007/s00442-004-1820-z>
- 943 Pedersen, S. H., Tamstorf, M. P., Abermann, J., Westergaard-Nielsen, A., Lund, M., Skov, K.,
944 Sigsgaard, C., Mylius, M. R., Hansen, B. U., Liston, G. E., & Schmidt, N. M. (2016).
945 Spatiotemporal characteristics of seasonal snow cover in Northeast Greenland from in
946 situ observations. *Arct. Antarct. Alp. Res*, 48(4), 653-671.
947 <https://doi.org/10.1657/Aaar0016-028>
- 948 Post, E. (2019). The Phenological Niche. In *Time in Ecology: A Theoretical Framework [MPB*
949 *61]* (4 ed., pp. 67-93). Princeton University Press.
950 <https://doi.org/https://doi.org/10.1515/9780691185491-006>
- 951 Prather, R. M., Dalton, R. M., Barr, B., Blumstein, D. T., Boggs, C. L., Brody, A. K., Inouye, D.
952 W., Irwin, R. E., Martin, J. G. A., Smith, R. J., Van Vuren, D. H., Wells, C. P.,
953 Whiteman, H. H., Inouye, B. D., & Underwood, N. (2023). Current and lagged climate
954 affects phenology across diverse taxonomic groups. *Proc Biol Sci*, 290(1990), 20222181.
955 <https://doi.org/10.1098/rspb.2022.2181>
- 956 Prevey, J. S., Rixen, C., Ruger, N., Hoye, T. T., Bjorkman, A. D., Myers-Smith, I. H.,
957 Elmendorf, S. C., Ashton, I. W., Cannone, N., Chisholm, C. L., Clark, K., Cooper, E. J.,

958 Elberling, B., Fosaa, A. M., Henry, G. H. R., Hollister, R. D., Jonsdottir, I. S., Klanderud,
959 K., Kopp, C. W., . . . Wipf, S. (2019). Warming shortens flowering seasons of tundra
960 plant communities. *Nat Ecol Evol*, 3(1), 45-52. [https://doi.org/10.1038/s41559-018-0745-](https://doi.org/10.1038/s41559-018-0745-6)
961 6

962 Primack, R. B., Ibanez, I., Higuchi, H., Lee, S. D., Miller-Rushing, A. J., Wilson, A. M., &
963 Silander, J. A. (2009). Spatial and interspecific variability in phenological responses to
964 warming temperatures. *Biol. Conserv.*, 142(11), 2569-2577.
965 <https://doi.org/10.1016/j.biocon.2009.06.003>

966 Pyke, G. H., Thomson, J. D., Inouye, D. W., & Miller, T. J. (2016). Effects of climate change on
967 phenologies and distributions of bumble bees and the plants they visit. *Ecosphere*, 7(3).
968 <https://doi.org/10.1002/ecs2.1267>

969 Raisanen, J. (2008). Warmer climate: less or more snow? *Clim Dyn*, 30(2-3), 307-319.
970 <https://doi.org/10.1007/s00382-007-0289-y>

971 Ramula, S., Johansson, J., Lindén, A., & Jonzén, N. (2015). Linking phenological shifts to
972 demographic change. *Clim Res*, 63(2), 135-144. <https://doi.org/10.3354/cr01289>

973 Rantanen, M., Karpechko, A. Y., Lipponen, A., Nordling, K., Hyvarinen, O., Ruostenoja, K.,
974 Vihma, T., & Laaksonen, A. (2022). The Arctic has warmed nearly four times faster than
975 the globe since 1979. *Commun. Earth Environ*, 3(1). [https://doi.org/10.1038/s43247-022-](https://doi.org/10.1038/s43247-022-00498-3)
976 00498-3

977 Regan, C. E., & Sheldon, B. C. (2023). Phenotypic plasticity increases exposure to extreme
978 climatic events that reduce individual fitness. *Glob Chang Biol*.
979 <https://doi.org/10.1111/gcb.16663>

980 Reneerkens, J., Schmidt, N. M., Gilg, O., Hansen, J., Hansen, L. H., Moreau, J., & Piersma, T.
981 (2016). Effects of food abundance and early clutch predation on reproductive timing in a
982 high Arctic shorebird exposed to advancements in arthropod abundance. *Ecol Evol*,
983 6(20), 7375-7386. <https://doi.org/10.1002/ece3.2361>

984 Rixen, C., Hoye, T. T., Macek, P., Aerts, R., Alatalo, J. M., Anderson, J. T., Arnold, P. A.,
985 Barrio, I. C., Bjerke, J. W., Bjorkman, M. P., Blok, D., Blume-Werry, G., Boike, J.,
986 Bokhorst, S., Carbognani, M., Christiansen, C. T., Convey, P., Cooper, E. J., Cornelissen,
987 J. H. C., . . . Zong, S. W. (2022). Winters are changing: snow effects on Arctic and alpine
988 tundra ecosystems. *Arct Sci*. <https://doi.org/10.1139/as-2020-0058>

989 Roslin, T., Antao, L., Hallfors, M., Meyke, E., Lo, C., Tikhonov, G., Delgado, M. D., Gurarie,
990 E., Abadonova, M., Abduraimov, O., Adrianova, O., Akimova, T., Akkiev, M., Ananin,
991 A., Andreeva, E., Andriychuk, N., Antipin, M., Arzamascev, K., Babina, S., . . .
992 Ovaskainen, O. (2021). Phenological shifts of abiotic events, producers and consumers
993 across a continent. *Nat Clim Change*, 11(3). <https://doi.org/10.1038/s41558-020-00967-7>

994 Roy, D. B., & Sparks, T. H. (2000). Phenology of British butterflies and climate change. *Glob*
995 *Change Biol*, 6(4), 407-416. [https://doi.org/DOI 10.1046/j.1365-2486.2000.00322.x](https://doi.org/DOI%2010.1046/j.1365-2486.2000.00322.x)

996 Rudolf, V. H. W. (2019). The role of seasonal timing and phenological shifts for species
997 coexistence. *Ecol Lett*, 22(8), 1324-1338. <https://doi.org/10.1111/ele.13277>

998 Sanz-Aguilar, A., Carrete, M., Edelaar, P., Potti, J., & Tella, J. L. (2015). The empty temporal
999 niche: breeding phenology differs between coexisting native and invasive birds. *Biol*
1000 *Invasions*, 17(11), 3275-3288. <https://doi.org/10.1007/s10530-015-0952-x>

1001 Schmidt, N. M., Hansen, L. H., Hansen, J., Berg, T. B., & Meltofte, H. (2019). BioBasis -
1002 Conceptual design and sampling procedures of the biological monitoring programme
1003 within Zackenberg Basic. <https://g-e-m.dk/fileadmin/g-e->

1004 m/Zackenberg/BioBasis_manual_2019.pdf. [Accessed 21 March 2022].

1005 Schmidt, N. M., Hardwick, B., Gilg, O., Høye, T. T., Krogh, P. H., Meltofte, H., Michelsen, A.,
1006 Mosbacher, J. B., Raundrup, K., Reneerkens, J., Stewart, L., Wirta, H., & Roslin, T.
1007 (2017). Interaction webs in arctic ecosystems: Determinants of arctic change? *Ambio*,
1008 46(Suppl 1), 12-25. <https://doi.org/10.1007/s13280-016-0862-x>

1009 Schmidt, N. M., Kankaanpää, T., Tiisanen, M., Reneerkens, J., Versluijs, T. S. L., Hansen, L.
1010 H., Hansen, J., Gerlich, H. S., Høye, T. T., Cirtwill, A. R., Zhemchuzhnikov, M. K.,
1011 Pena-Aguilera, P., & Roslin, T. (2023). Little directional change in the timing of Arctic
1012 spring phenology over the past 25 years. *Curr Biol*.
1013 <https://doi.org/10.1016/j.cub.2023.06.038>

1014 Schmidt, N. M., Mosbacher, J. B., Nielsen, P. S., Rasmussen, C., Høye, T. T., & Roslin, T.
1015 (2016). An ecological function in crisis? The temporal overlap between plant flowering
1016 and pollinator function shrinks as the Arctic warms. *Ecography*, 39(12), 1250-1252.
1017 <https://doi.org/10.1111/ecog.02261>

1018 Skov, K., Sigsgaard, C., Mylius, M. R., Lund, M. (2020). *Zackenberg Ecological Research*
1019 *Operations, GeoBasis, Guidelines and sampling procedures for the geographical*
1020 *monitoring programme of Zackenberg Basic*. [https://g-e-m.dk/fileadmin/g-e-](https://g-e-m.dk/fileadmin/g-e-m/GEM/GeoBasis_Manual_Comp2020.pdf)
1021 [m/GEM/GeoBasis_Manual_Comp2020.pdf](https://g-e-m.dk/fileadmin/g-e-m/GEM/GeoBasis_Manual_Comp2020.pdf) [Accessed 16 September 2022].

1022 Stemkovski, M., Bell, J. R., Ellwood, E. R., Inouye, B. D., Kobori, H., Lee, S. D., Lloyd-Evans,
1023 T., Primack, R. B., Templ, B., & Pearse, W. D. (2023). Disorder or a new order: How
1024 climate change affects phenological variability. *Ecology*, 104(1).
1025 <https://doi.org/10.1002/ecy.3846>

1026 Strathdee, A. T., Bale, J. S., Block, W. C., Coulson, S. J., Hodkinson, I. D., & Webb, N. R.
1027 (1993). Effects of Temperature Elevation on a Field Population of *Acyrtosiphon*
1028 *Svalbardicum* (Hemiptera, Aphididae) on Spitsbergen. *Oecologia*, 96(4), 457-465.
1029 <https://doi.org/10.1007/Bf00320502>

1030 Saalfeld, S. T., McEwen, D. C., Kesler, D. C., Butler, M. G., Cunningham, J. A., Doll, A. C.,
1031 English, W. B., Gerik, D. E., Grond, K., Herzog, P., Hill, B. L., Lagasse, B. J., & Lanctot,
1032 R. B. (2019). Phenological mismatch in Arctic-breeding shorebirds: Impact of snowmelt
1033 and unpredictable weather conditions on food availability and chick growth. *Ecol Evol*,
1034 9(11), 6693-6707. <https://doi.org/10.1002/ece3.5248>

1035 Thackeray, S. J., Henrys, P. A., Hemming, D., Bell, J. R., Botham, M. S., Burthe, S., Helaouet,
1036 P., Johns, D. G., Jones, I. D., Leech, D. I., Mackay, E. B., Massimino, D., Atkinson, S.,
1037 Bacon, P. J., Brereton, T. M., Carvalho, L., Clutton-Brock, T. H., Duck, C., Edwards,
1038 M., . . . Wanless, S. (2016). Phenological sensitivity to climate across taxa and trophic
1039 levels. *Nature*, 535(7611), 241-245. <https://doi.org/10.1038/nature18608>

1040 Tobin, P. C., Nagarkatti, S., Loeb, G., & Saunders, M. C. (2008). Historical and projected
1041 interactions between climate change and insect voltinism in a multivoltine species. *Glob*
1042 *Change Biol*, 14(5), 951-957. <https://doi.org/10.1111/j.1365-2486.2008.01561.x>

1043 Tulp, I., & Schekkerman, H. (2008). Has prey availability for arctic birds advanced with climate
1044 change? Hindcasting the abundance of tundra arthropods using weather and seasonal
1045 variation. *Arctic*, 61(1), 48-60. <Go to ISI>://WOS:000254658200005

1046 Valtonen, A., Ayres, M. P., Roininen, H., Poyry, J., & Leinonen, R. (2011). Environmental
1047 controls on the phenology of moths: predicting plasticity and constraint under climate
1048 change. *Oecologia*, 165(1), 237-248. <https://doi.org/10.1007/s00442-010-1789-8>

1049 van de Pol, M. V., & Wright, J. (2009). A simple method for distinguishing within- versus

1050 between-subject effects using mixed models. *Anim Behav*, 77(3), 753-758.
1051 <https://doi.org/10.1016/j.anbehav.2008.11.006>
1052 Visser, M. E., & Both, C. (2005). Shifts in phenology due to global climate change: the need for
1053 a yardstick. *Proc R Soc B: Biol Sci*, 272(1581), 2561-2569.
1054 <https://doi.org/10.1098/rspb.2005.3356>
1055 Walther, G. R. (2010). Community and ecosystem responses to recent climate change. *Philos*
1056 *Trans R Soc Lond B Biol Sci*, 365(1549), 2019-2024.
1057 <https://doi.org/10.1098/rstb.2010.0021>
1058 Walther, G. R., Post, E., Convey, P., Menzel, A., Parmesan, C., Beebee, T. J. C., Fromentin, J.
1059 M., Hoegh-Guldberg, O., & Bairlein, F. (2002). Ecological responses to recent climate
1060 change. *Nature*, 416(6879), 389-395. [https://doi.org/DOI 10.1038/416389a](https://doi.org/DOI%2010.1038/416389a)
1061 Wheeler, H. C., Høye, T. T., Schmidt, N. M., & Svenning, J.-C., Forchhammer, M.C. (2015).
1062 Phenological mismatch with abiotic conditions - implications for flowering in Arctic
1063 plants. *Ecology*, 96(3), 775 - 787.
1064 Wood, S. N. (2017). *Generalized Additive Models - An Introduction with R* (2nd ed.). Chapman
1065 and Hall/CRC. <https://doi.org/https://doi.org/10.1201/9781315370279>
1066 Zografou, K., Swartz, M. T., Adamidis, G. C., Tilden, V. P., McKinney, E. N., & Sewall, B. J.
1067 (2021). Species traits affect phenological responses to climate change in a butterfly
1068 community. *Sci Rep*, 11(1). [https://doi.org/ARTN 3283](https://doi.org/ARTN%203283)
1069 [10.1038/s41598-021-82723-1](https://doi.org/10.1038/s41598-021-82723-1)
1070 Zuur, A. F., Ieno, E. N., & Elphick, C. S. (2010). A protocol for data exploration to avoid
1071 common statistical problems. *Methods Ecol Evol*, 1(1), 3-14.
1072 <https://doi.org/10.1111/j.2041-210X.2009.00001.x>
1073