1	Phenological sensitivity to climate across taxa and local habitats in
2	a high-Arctic arthropod community
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17	H.S.G.: Conducted data analysis, statistics and drafted the manuscript. T.T.H.:
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20	N.M.S.: Oversaw collection of the field data.
21	All authors participated with intellectual contributions and revised the manuscript.
22	
23	Data accessibility
24	Arthropod monitoring data is available through the open-source GEM-database:
25	https://data.g-e-m.dk/. The R-scripts necessary to replicate the findings of this
26	manuscript will be made available at the time of publication in an open-access
27	repository.
28	
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35 <u>Abstract</u>

Arthropods respond to climate change by shifting their phenology in the spring and summer seasons. These phenological shifts are rarely uniform, and taxa show distinct variation in the direction and magnitude of phenological responses to climate drivers. To gain insights into the most climate-sensitive taxa and forecast the implications of climate change on community-wide activity and biotic interactions, it is important to understand how the climate affects the timing of activity of different taxa in local sites within a community. Here, we examined the temporal trends of arthropod phenology, and associations between arthropod phenological responses and climate predictors using arthropod monitoring data from five different habitats in high-Arctic Greenland covering a 25-year period. We found that, for most taxa, advanced arthropod phenology was associated with earlier snowmelt, and, to a lesser extent, warmer temperatures. However, the magnitude of advanced activity varied considerably between arthropod taxa and local habitats. Our study also revealed that pollinators were the most climate-sensitive group, with advanced and, in some habitats, shortened seasonal activities. Late active taxa and late snow melting habitats advanced phenology at greater magnitudes than early active taxa and early snow melting habitats. The magnitude of phenological shifts of arthropod taxa was dependent on habitat, highlighting the substantial spatial variation in phenological responses. Overall, our results demonstrate that high-Arctic arthropods are capable of tracking local climate drivers of phenology well, but the phenological responses of arthropod taxa to global climate change are complex, and community-wide trends may mask the variation in direction and magnitude of phenological shifts in different taxa and locally adapted populations.

Keywords: arthropods, climate change, phenology, high-Arctic, snowmelt,

temperature, long-term monitoring, phenological mismatch

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75 Introduction

76 Phenology in a global change context

77 The globe is undergoing unprecedented climatic changes, which have 78 implications for ecosystems across the world. Numerous ecological changes have 79 been associated with global warming (Halsch et al., 2021; Parmesan, 2006; Walther et al., 2002), but the most reported effects are shifts in phenological events 80 81 (Thackeray et al., 2016; Visser & Both, 2005). Long-term studies report advancements of arthropod phenological events in response to warming (Menzel et 82 83 al., 2006; Parmesan & Yohe, 2003; Post et al., 2018; Root et al., 2003; Vitasse et al., 84 2021). For example, terrestrial insects in European alpine (Vitasse et al., 2021) and north-temperate areas (Cohen et al., 2018) have shown advanced emergence of 6.0 85 86 and 4.15 days per decade, respectively. While most studies on phenological change 87 are from temperate regions of Europe and North America (Cohen et al., 2018), arthropods in the Arctic are experiencing even greater warming (AMAP, 2017; IPCC, 88 89 2021: Rantanen et al., 2022). Due to the short arctic growing seasons, arthropods begin their active season as soon as temperatures become favourable in the spring 90 (Danks, 2004). Consequently, small changes in temperatures during the growing 91 92 season can have a significant impact on the phenology of arctic arthropods, making 93 them particularly sensitive to climate change (Bolduc et al., 2013; Høye, Post, 94 Meltofte, Schmidt, et al., 2007; Post et al., 2018; Tulp & Schekkerman, 2008). This is especially interesting because arthropods dominate in animal species richness and 95 abundance in the Arctic and play important roles in ecosystem functioning (Barrio et 96 97 al., 2017; Hodkinson, 2013; Hodkinson & Coulson, 2004).

98

99 Variation in the direction and magnitude of phenological change

100 Arthropod taxa can exhibit greatly varying phenological sensitivities to 101 environmental changes, which may depend on specific life history traits such as voltinism and body size (Diamond et al., 2011; Gillespie et al., 2017; Pacifici et al., 102 103 2017). Univoltine species tend to advance and shorten their phenology compared to 104 multivoltine species that often delay their late season phenology (Glazaczow et al., 105 2016). Temperature sensitivity among species can also explain differential 106 phenological responses (Buckley, 2022; Thackeray et al., 2016), where early active 107 species advance their spring emergence and late active species delay their fall 108 activity (Bartomeus et al., 2011; Brooks et al., 2017; Gallinat et al., 2015; Kharouba 109 et al., 2014). Furthermore, differences in phenological responses can be associated 110 with greater sensitivity to environmental cues that are not affected by climate 111 change, such as photoperiod (Bale et al., 2002; Danks, 2007). The extent of 112 phenological shifts is also highly site-specific (Chmura et al., 2019; Kharouba et al., 2014; Primack et al., 2009), especially in Arctic snow-dominated environments 113 114 characterized by topographic heterogeneity creating pronounced spatial variation in 115 timing of arthropod emergence (Kankaanpää et al., 2018). Temperature and particularly timing of snowmelt are important environmental drivers of interannual 116 variation in Arctic arthropod spring emergence (Høye & Forchhammer, 2008). 117 118 However, predicting arthropod population- and community-level responses to climate 119 change remains a challenge, especially when multiple environmental drivers act in 120 concert to determine arthropod phenology.

121

122 Heterogeneous phenological shifts can reshape a community

Arthropod populations in habitats with early snowmelt emerge earlier
than those in late snow melting habitats. This local scale emergence affects the

125 duration of activity for the entire arthropod community (Phillimore et al., 2010; Roy et al., 2015) and defines the community phenological niche, which is the overall 126 127 temporal occupancy of individual taxa and population phenology in a season (Post, 128 2019). If species and populations differ in their sensitivity to climate drivers (Brooks et al., 2017; Prevey et al., 2017; Stalhandske et al., 2014; Tyukmaeva et al., 2020; 129 130 Valtonen et al., 2014), warmer seasons could lead to rapid advanced emergence in 131 sites with early snowmelt compared to sites with late snowmelt or vice versa. As a 132 result, overlap in duration of activity could increase or shorten, which ultimately 133 affects the temporal synchronization of arthropod community activity and thus the 134 community phenological niche breadth (Post, 2019). For instance, dominant plant species at Zackenberg have shown reduced community flowering because late 135 136 flowering populations advanced their flowering time faster than early flowering 137 populations (Høye et al., 2013). It remains unclear how the phenological niche 138 breath of arthropod taxa is affected by a changing environment (But see Altermatt, 139 2012; Gutierrez & Wilson, 2021; Phillimore et al., 2010; Roy et al., 2015), and if the 140 total duration of arthropod community activity season has changed. Changes in arthropod community activity can lead to a temporal mismatch between trophic 141 142 levels, such as plant-pollinator or predator-prey interactions (Pyke et al., 2016; 143 Reneerkens et al., 2016; Visser & Both, 2005), which could have serious 144 consequences for ecosystem functioning (Samplonius et al., 2021). Therefore, 145 understanding how rates of phenological change vary among taxa in the face of increasing climatic variability is crucial. 146

147 Comprehensive long-term capture data is essential to gain insights into 148 the impact of climate change on the phenology of arthropods at the community level 149 (Coulson et al., 2014; Gillespie et al., 2020; Hodkinson, 2013; Inouye, 2022). The

150 arthropod monitoring program at Zackenberg, Northeast Greenland has provided 151 long-term standardized collection data since 1996 (Schmidt et al., 2019), offering a 152 unique opportunity to investigate the ecological impacts of climate change. Here, we 153 used this data to investigate the variation in phenological responses of terrestrial and semiaquatic arthropods from different local habitats to the timing of snowmelt and 154 temperature using abundance data from weekly trappings. Our objective was 155 156 threefold. First, we aimed to provide an assessment of the temporal dynamics on arthropod phenology for all available family-habitat combinations. In accordance with 157 158 previous short-term studies on phenological responses of arthropods (Høye, Post, 159 Meltofte, & Schmidt, 2007), we expected temporal advancements in phenological responses and prolonged activity periods of most arthropod taxa. Second, we aimed 160 161 to evaluate the relative contribution of snowmelt timing and temperature on the 162 phenological responses (peak activity and duration of activity). We expected the variation in phenology of arthropods to be driven by timing of snowmelt rather than 163 164 temperature, but early active taxa and taxa in early snow melting habitats were expected to respond more strongly to temperature than late active taxa. Third, we 165 aimed to assess whether phenological responses to climate variation differed among 166 taxa and habitats to derive knowledge on the variation in phenological sensitivity to 167 168 climate variables within the community. We expected the climate sensitivity of early 169 season taxa and populations to be greater than the sensitivity of late season taxa 170 and populations, resulting in longer overall activity seasons at the community level.

171

172 Materials and methods

173 Study site and arthropod sampling

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Arthropods were collected at the Zackenberg Research Station, located

175 in high-Arctic Northeast Greenland (74°28' N; 20°34' W). The collection of 176 arthropods was conducted from 1996 to 2020 through the Greenland Ecosystem 177 Monitoring Program and included seven different plots. Plot 2 - 7 consisted of eight 178 yellow pitfall traps (1997 to 2006) later reduced to four pitfall traps (1996 and 2007 to 2020) and plot 1 included four window traps where specifically limnitic insect 179 180 emergence and aerial activity was monitored. The traps opened at snowmelt in late 181 May to early June, operated during the growing season and ended by 1st September, 182 which often coincided with freeze up. The traps were emptied weekly at fixed dates, 183 unless the weather prohibited handling the samples in which case the traps were 184 emptied the following day. After collection, specimens were stored in 75% ethanol, later changed to 96% ethanol, and transported to Denmark where the arthropods 185 186 were sorted by the technicians at the Department of Ecoscience at Aarhus University 187 to the family level for spiders and most insects, superfamily level for Aphidoidea, 188 Chalcidoidea and Coccoidea, and subclass for other arthropods, and subsequently 189 counted. The data is publicly available at data.g-e-m.dk. To enable comparison of 190 arthropod capture numbers between years, we focused on arthropod data from only 191 June, July and August. Further, weekly abundance counts for each arthropod group were standardized by calculating individuals per trap per day for each plot. 192

The plots represented pond (Plot 1), wet fen (Plot 2), mesic heath (Plots 3 and 4), snow bed (Plot 6) and arid heath (Plots 5 and 7) habitats where each habitat differed in plant community composition, soil moisture and the timing of snowmelt. The wet fen habitat was primarily dominated by mosses and grasses and has high soil moisture and early snowmelt. The mesic heath habitat was dominated by lichens, *Cassiope tetragona*, *Dryas* sp. and *Salix arctica* and typically had snowmelt two weeks later than the fen and arid heath area. The arid heath habitat

was composed primarily of lichens, *Dryas* sp. and grasses, had relatively low soil
 moisture and experienced early snowmelt. A small islet in a shallow pond
 represented the pond habitat (Schmidt et al., 2019).

203 Our analysis focused on the most abundant arthropod taxa within the 204 community, which we defined as taxa for which at least 50 individuals were caught 205 per plot in a season. The arthropod taxa that were sufficiently represented based on 206 this criterion were soil mites (Acari), collembolans, spiders and the insect orders; 207 Diptera, Hymenoptera, Hemiptera and Lepidoptera (see Supporting Information S1). 208 Some traps showed major spikes in the number of mites caught, which was linked to 209 the capture of mite-infested bumblebees in a trap thus resulting in the capture of 210 several hundred mites. These spike numbers were changed to the average of the 211 other traps in a plot from identical capture periods (see Supporting Information S1). 212 An identical problem with large spikes in abundance estimates was found for the 213 spider family Lycosidae caused by the capture of many juveniles along with female 214 individuals. This was managed by changing spike numbers to the average of the 215 other traps in a plot from the same capture period. Another spider family; Linyphiidae exhibited deviations in annual adult activity patterns causing problems when 216 generating the phenological curves as bimodal distributions were found for some 217 218 years. The different species of spiders within Linyphildae may overwinter in different 219 life stages causing bimodal distributions in abundance estimates across a season. A 220 distinction between juveniles and adults was not made in the data for all years and so, abundance estimates for Linyphiidae could not be corrected. However, it was 221 222 carefully considered if satisfying phenological curves could be generated across a 223 season for each year and thus, it was regarded as appropriate to keep this family in 224 the study (see Supporting Information S1). This approach was identical to the

approach taken in Høye et al. (2007).

226

227 Climate variables

228 We selected temperature and snowmelt timing as climate variables to determine the effect of climate on emergence phenology as these variables 229 previously was found to influence the timing of arthropod emergence (Høye & 230 231 Forchhammer, 2008; Kankaanpää et al., 2018). A meteorological station located 232 centrally and within 600 m from all plots operated through the entire study period and 233 measured soil (0-, 5-, and 10-cm depth) and air (2 m above the ground) temperature 234 on an hourly basis. Air temperature, rather than soil temperature, was chosen to 235 explain temperature variations in the activity levels of arthropods in this study (please 236 see Supporting Information S2 for a detailed justification of the choice of air 237 temperature). Timing of snowmelt was estimated as the date by which less than 10 238 cm of snow was measured by an automatic ultrasonic snow depth sensor installed at 239 the meteorological station (Skov, 2020). Years 2009, 2013 and 2019 had limited snow accumulation resulting in the estimation of very early snow melting dates. As 240 these years became obvious outliers, we used soil temperature data (averaged from 241 0, 5 and 10 cm depth) to estimate more reliable snow melting dates for the years 242 243 2009, 2013 and 2019 as well as 1996 where no snow depth data was available, 244 following the method in Rixen et al. (2022). We identified the time period towards the 245 end of the winter when ground temperatures were stable near 0 °C and subsequently started fluctuating when the snow cover disappeared (defined as the 246 247 zero-curtain window). From this, we defined the date of snowmelt as when the mean daily soil temperature rose above +1 °C after a period with diurnal fluctuations of less 248 than 2 K and mean daily temperatures between -1 °C and 1 °C (Rixen et al., 2022). 249

This also enabled us to estimate a snow melt day for the year 1996 where no snow depth data was available. The correlation between day of snowmelt estimated from the snow depth sensor and soil temperature data can be seen in Figure S2.1.

253 We used hourly data on air temperature measured at 2 m height at the Zackenberg climate station (Downloaded: 13th January 2022). We then compiled 254 temperature predictors for each phenological event separately. Temperature in the 255 256 period before emergence is closely related to arthropod development (Gillooly et al., 257 2002) and may thus be a good predictor of adult arthropod emergence. Therefore, 258 we first calculated the temperature predictor as temperatures through the 30 days 259 prior to a phenological event, so as not to use temperature values largely occurring after the phenological event to predict the response. This was done by calculating 260 261 the mean date of the phenological event across years for each family and habitat 262 combination and then extracting the mean temperature during the 30 days prior to 263 this date for each year.

264

265 Quantifying phenology

Annual onset, peak and end of emergence of arthropod taxa across 266 each habitat and each year were calculated using generalized additive modelling 267 268 (GAM). We predicted a non-linear phenological development across each season, 269 however, as the shape was unknown, it was ideal to take advantage of the partial 270 smoothing method in GAMs to model the seasonal development in capture rates (Guisan et al., 2002). In addition, GAMs provide fairly accurate estimations of 271 272 species phenology despite gaps in the distribution due to varying sample size 273 (Moussus et al., 2010). Curves of arthropod abundance were fit across a season 274 using GAM, assuming a poisson distribution with k = 4 (basis dimensions) to ensure

275 appropriate smoothing and a log link function. We used the package 'mqcv' version 276 1.8-40 to fit the GAMs (Wood, 2017). Some taxa had low capture numbers in some 277 plots in some years, and consequently, we restricted our analyses to years and plots 278 where at least 50 individuals of a given taxa were caught. Also, the three phenological events should be possible to calculate (the taxa must be present in at 279 280 least two weeks) and a sensible seasonal curve must be generated. Annual onset, 281 peak and end of the activity season were then calculated as the day at which 10%, 282 50% and 90% of the accumulated abundance (area under the curve) was reached, 283 respectively (Figure 1). The duration of the activity period was estimated as the number of days between onset and end of the activity season. 284



Figure 1– Example of fitting phenological curves using GAM (Poisson family) for abundance data of
arthropod taxa to estimate onset, peak and end of emergence. Data from Culicidae in plot 1 in year
2019.

291 Statistical analyses

- 292 Temporal trends in climate and arthropod phenology
- 293 Temporal trends in air temperature and timing of snowmelt were
- 294 calculated as the slope of the regression against year as a continuous predictor
- using simple linear regression. Additionally, linear regression was used to calculate
- the temporal trends in the onset, peak and end of arthropod activity as the slope of

297 the regression against year. This was done for each taxon for which it was possible 298 to estimate phenology of the activity period for at least five years across the time 299 series data. Data for analysis of temporal trends was sufficient for 15 taxa (full list 300 available in Supplementary Information S1). To test if temporal trends of arthropod 301 phenology differed between taxa, functional group and habitat, we used multiple 302 regression with each individual phenological event as the response variable and 303 year, plot and taxa or functional group as predictor variables. When functional 304 groups were compared, we added random intercepts for taxa to account for 305 nonindependence of observations among taxa within functional groups. In addition, a 306 year-plot interaction was included to allow for the year effect to vary among sites. The functional groups were defined as; pollinators (Diptera families and 307 308 Nymphalidae), predators (Aranae families), herbivores (Hemiptera families), 309 parasitoids (Hymenoptera families) and decomposers (Acari and Collembola) (Table S1.1). 310

311

312 Correlation between climate variables

313 The climate variables included in this study (timing of snowmelt and air temperature) may be highly correlated. An appropriate correlation analysis between 314 315 climate variables was not feasible because specific air temperature predictor values 316 were calculated for the individual average emergence date for each arthropod taxa in 317 each plot. To appropriately examine a potential issue of multicollinearity, variance inflation factors (VIFs) were derived using the R package 'car' version 3.1-1 for 318 319 timing of snowmelt and temperature in all family and plot combinations for each 320 phenological event (Zuur et al., 2010). VIF values were also derived for timing of 321 snowmelt and soil temperature measured at the same climate station at a depth of 0

- 10 cm to compare the correlation of the two temperature variables with timing of
 snowmelt. We used a threshold criterion of 10 such that predictors with values above
 a VIF > 10 were considered contributing greatly to multicollinearity.

325

326 Effect of environmental predictors on phenological events

327 To determine the effect of snowmelt and temperature on arthropod 328 phenology, we used timing of snowmelt and temperature as explanatory variables of 329 each phenological response variable while accounting for the random slope and 330 intercept parameters of arthropod taxa and plot in linear mixed models (LMM) using 331 the 'Ime4' R package version 1.1-31 (Bates et al., 2015). As we use taxa and plot specific estimates of the temperature predictor, we separated the within-subject 332 333 effects from between-subject effects that we cannot account for in the mixed models 334 by implementing within-subject centering in the model (van de Pol & Wright, 2009). 335 This is done by subtracting an average temperature value for each arthropod taxa 336 and plot combination from the specific temperature value for each arthropod taxa and plot combination. This new temperature predictor derived is then included as a 337 fixed effect in the LMM that expresses the within and between arthropod taxa and 338 plot temperature variation component. We also detrended the climate variables by 339 340 adding year as a covariate in the model. Interactions between climate variables and 341 plot, taxon and functional group was included in the model. A significant interaction 342 term indicated that the slope of the linear relationship between phenological response and climate differed between taxa, functional group or plot. 343

To determine the differences in phenology of individual taxa among plots, multiple linear regression analysis with timing of snowmelt and temperature as predictors and plot as a fixed effect was performed. Plot was included as a

347 categorical variable and to make the desired comparisons of slopes between plots, customized contrasts were used. We also tested all possible interactions between 348 plots and climate variables. We did not include a random intercept for year, as there 349 350 was no within-year replication of the site-specific environmental variables. A significant interaction term indicated that the slope of the linear relationship between 351 352 a phenological response and climate variable differed among habitats. The best 353 models were selected based on ANOVA comparison between nested models. The chosen model is clearly indicated for each analysis in tables with results in the 354 355 supplementary information. 356 357 Results 358 Warmer summers but high interannual variation in climate variables 359 Snowmelt timing and spring air temperature has not changed significantly during the 25-year study period at Zackenberg (Snowmelt: R^2 = 360

361 0.01, $F_{1,23} = 0.25$, P = 0.62, -1.6 days earlier per decade; Average spring air

362 temperature: $R^2 = 0.02$, $F_{1,23} = 0.52$, P = 0.48, 0.04 °C per decade), but show

increasing variation since the mid-2000s (Figure 2). However, summer air

temperature has significantly increased by 0.6 ± 0.3 °C per decade during the study

365 period (average summer air temperature: $R^2 = 0.16$, $F_{1,23} = 4.49$, P = 0.045). A low

366 level of multicollinearity between timing of snowmelt and air temperature was found

367 (see Supplementary Information S4).



Figure 2 – Interannual variation and trends of (a) timing of snowmelt and (b) average seasonal air
temperature for spring (April, May) and summer (June, July, August) at Zackenberg, Greenland
during the study period of 1996 - 2020. Only summer air temperature changed significantly during the
study period.

373

374 The phenological niche of arthropods differ among taxa and habitats

375 The timing of activity varied widely among arthropod taxa and habitats 376 (Figure 3). Families of pollinators are active early and late in the season. As an

- example, Chironomidae is present early in the season compared to the late
- 378 emerging Phoridae. The parasitoids and herbivores are active late in the season with
- 379 considerable variation between habitats. Within Ichneumonidae, the peak activity
- occurs on average on day 201 in the wet fen with early snowmelt as opposed to day
- 381 220 in the mesic heath with late snowmelt. A longer duration of the activity season is
- 382 predominant for decomposers and predators.



Figure 3 – The average peak phenological event (when 50% of the season capture was reached) of all taxa across multiple years represented by at least 50 individuals per season for each plot and habitat (day of year 140 – 240). Associated error bars are given. The arthropods are listed in order of activity such that the top families are active earlier than the below families. The associated error bars indicate the standard error of the mean. Estimations of all phenological events are available in Supporting Information.

390

391 Temporal advancements in community phenology but weak overall directional shift

392 Consistent with the lack of temporal trends in climate variables, there 393 were limited phenological shifts across time (only 18% of all cases showed a significant shift across time) and significant phenological shifts were almost entirely 394 395 advancements (Figure S3.1). Of all the family and plot combinations exhibiting a significant phenological shift, 51% (20/39) showed a significant change in the onset 396 397 of emergence. Only 23% (9/39) and 26% (10/39) of the family and plot combinations 398 showed a significant shift in the peak and end of activity, respectively. The average community peak phenology advanced by 2.6 ± 0.4 days per decade. There was no 399

400 significant change in the community duration of activity. Multiple regression analysis 401 revealed a significant difference in the rate of temporal advancements among 402 arthropod taxa (peak and duration of activity) and habitats (only for duration of 403 activity) (Table S3.2). When merging taxa in functional groups, there was a 404 significant difference among groups in the temporal trends of the peak activity and 405 duration of activity (see Supporting Information S3). Predators had significantly 406 extended the duration of their activity by 3.9 ± 1.9 days per decade. The pollinators 407 shifted their phenology considerably in the wet fen habitat with peak phenology 408 advancing 6.8 ± 1.3 days per decade on average and Sciaridae advancing the most 409 (advancing 10.2 ± 2.7 days per decade).

410

411 Advanced arthropod phenology with earlier snowmelt and increasing temperatures

We only present results of the effect of climate on the peak and duration
(difference between onset and end of activity) of activity to facilitate a better
understanding of the results.

415 The timing of snowmelt was a better predictor of arthropod phenology than air temperature, as temperature explained less variation in phenological events 416 (see Supporting Information S4). After detrending the time series by year, the 417 418 significance of the linear temperature-phenology relationship diminished, while most 419 linear phenology-snowmelt relationships persisted. Most taxa exhibited a significant 420 response of peak phenology to timing of snowmelt, temperature or both, except Aphidoidea and Thomisidae (Table S5.1). The most common trend was advanced 421 422 phenology in response to earlier snowmelt and increasing temperature (Figure 4, 5 & 423 Table S5.1). The average community phenology advanced by 0.35 ± 0.06 days per 424 earlier snowmelt day. In response to temperature, average community phenology

425 advanced by 1.11 ± 0.51 days per 1 °C increase (Table S5.3). Most taxa responded 426 to snowmelt and temperature by either extending or shortening the duration of activity (Figure 4, 5 & Table S5.2). Shifts in duration of activity was only significant in 427 428 response to snowmelt with a slight increase of -0.12 ± 0.05 days per advanced 429 snowmelt day, but substantial variation in duration of activity in response to climate 430 among taxa and habitat was found (Table S5.3). 12 out of 15 taxa showed a 431 significant positive effect of snowmelt timing on peak phenology (Table S5.1), 432 meaning that these taxa advanced their phenology in response to timing of 433 snowmelt. Seven taxa showed a significant negative effect of air temperature (results 434 not shown), thereby advancing their peak phenology in warmer years. However, 435 after detrending the time series data, the significance of the linear relationships 436 between arthropod phenology and temperature decreased such that five taxa 437 showed a significant effect of air temperature (Table S5.1).





439 Figure 4 – The phenological sensitivity (days per shifted snowmelt day) in (A) peak and (B) duration 440 of the activity season for taxa in each plot within habitats with each panel representing the different 441 habitats where arthropods were collected. The model is controlled for the effect of temperature. The 442 arthropods are listed in order of activity such that the top families are active earlier than the below 443 families. The size of each point represents the number of observations in each family and plot 444 combination. Significant shifts are represented in red ($P \le 0.05$), marginally significant shifts in orange 445 $(0.05 < P \le 0.10)$ and nonsignificant shifts in blue (P > 0.05). We refer to trends as significant when the 95% confidence interval (CI) for a parameter of the fitted models did not overlap zero. The 446 447 associated error bars indicate the 95% CI's. To ease comparison, all panels are equally scaled 448 causing a cut-off of Cl's. In the mesic heath and arid heath panels, two points are shown indicating 449 the two plots in these habitats. If points have different significance levels, the points are slightly 450 separated to ease visibility.



452

453 Figure 5 – The phenological sensitivity (days per °C) in (A) peak and (B) duration of the activity 454 season for taxa in each plot with each panel representing the different habitats where arthropods 455 were collected. The model is controlled for the effect of snowmelt. The arthropods are listed in order 456 of activity such that the top families are active earlier than the below families. The size of each point 457 represents the number of observations in each family and plot combination. Significant shifts are 458 represented in red ($P \le 0.05$), marginally significant shifts in orange ($0.05 < P \le 0.10$) and 459 nonsignificant shifts in blue (P > 0.05). We refer to trends as significant when the 95% confidence 460 interval (CI) for a parameter of the fitted models did not overlap zero. The associated error bars 461 indicate the 95% Cl's. To ease comparison, all panels are equally scaled causing a cut-off of Cl's. In 462 the mesic heath and arid heath panels, two points are shown indicating the two plots in these 463 habitats. If points have different significance levels, the points are slightly separated to ease visibility. 464

465 Differential phenological responses among functional groups and early vs late active
 466 taxa

467

Taxa within functional groups responded with the same directional

468 change in phenology to climate variables, even though the rate of change varied for some taxa (see Supporting Information S7 for details). When comparing slopes 469 470 between functional groups, pollinators and parasitoids was found to advance their 471 phenology (Pollinators, slope: 0.52 ± 0.07 ; Parasitoids, slope: 0.55 ± 0.13) to a greater extent in response to earlier snowmelt compared to predators and 472 decomposers (Predators, slope: 0.11 ± 0.08 ; Decomposers, slope: 0.25 ± 0.08) 473 474 (Table S6.1). Predators and decomposers extended the duration of their activity strongly in response to earlier snowmelt (Predators, slope: -0.48 ± 0.07 ; 475 476 Decomposers, slope: -0.30 ± 0.06) compared to all other functional groups. 477 Pollinators did not show significant changes in duration of activity in response to snowmelt (Table S6.2), but considerable variation among families and habitats was 478 479 found in this group (Figure 4, Table S5.2 & S6.2). As temperatures increased, the 480 peak activity of pollinators became substantially earlier (Pollinators, slope: -1.57 ± 481 0.52). There was no significant change in the duration of activity of functional groups 482 in response to temperature (Table S6.2).

483 In response to earlier snowmelt, late active taxa advanced their peak phenology by 0.47 ± 0.07 days per day earlier snowmelt whereas early active taxa 484 advanced their peak phenology by only 0.26 ± 0.07 days per day earlier snowmelt 485 486 (Table S6.3, Figure S6.1). In response to temperature, early active taxa advanced 487 their peak phenology by -1.71 ± 0.62 days per 1 °C increase, while late active taxa 488 did not advance their phenology, however, the slopes were not significantly different (Table S6.3). Early active taxa extended their activity period with 0.21 ± 0.03 days 489 490 per day earlier snowmelt, while late active taxa did not change the period of activity 491 $(Slope: 0.01 \pm 0.04).$

492

The random effect of the interaction between plot and arthropod taxa

493 explained up to 73% of the total variation in phenological responses to climate
494 variables (not shown but see Supporting Information S7). Thus, a detailed
495 investigation of the spatial variation in phenological responses to climate within each
496 taxon separately was highly relevant.

497

498 Spatial variation in phenological responses to climate

499 We found strong support for spatial variation in arthropod phenology indicated by a significant effect of habitat as explanatory variable for taxa found at 500 501 different sites (Figure 4 & 5, Table S7.6 & S7.7). Populations from late snow melting 502 mesic heath and snow bed habitats showed more rapid peak phenological shifts in response to earlier snowmelt than populations from early snow melting arid heath 503 504 and wet fen habitats (Figure 4). Whether taxa shortened or extended their activity 505 period in response to snowmelt and temperature was dependent on habitat for 506 families of pollinators (Figure 4 & 5, Table S5.2).

507 Spatial variation in phenological responses to snowmelt in particular, but 508 also temperature, were pronounced for families of pollinators. For Chironomidae, Muscidae, Phoridae and Sciaridae, there were stronger phenological responses in 509 510 late compared to early snow melting habitats (Table S5.1). Muscidae, Chironomidae 511 and Phoridae advanced peak phenology at a greater rate in response to earlier 512 snowmelt in late snow melting mesic heath habitats than in the early snow melting 513 arid heath and wet fen habitats. In addition, Muscidae shortened the duration of 514 activity in the wet fen habitat (Table S5.2). Families in the pond habitat advanced 515 their phenology at fast rates in response to temperature, e.g., Muscidae peak 516 phenology was 4.63 ± 1.47 days earlier per 1 °C, which was significantly greater than 517 in other habitats (Table S5.1). In addition, Sciaridae experienced shortened activity

periods in the pond habitat in response to warmer summers (Table S5.2). Culicidae
was also active for shorter periods in the pond habitat in response to earlier
snowmelt (Table S5.2, Figure 4). Duration shortened in warmer years and years with
earlier snowmelt because the end of activity advanced at faster rates than onset to
increasing temperatures and advancing snowmelt (Figure S7.1, S7.2).

523 For Herbivores, Parasitoids and Decomposers, most families extended 524 their activity periods in response to earlier snowmelt in the late snow melting mesic heath habitat, which was not found in the early snow melting arid heath habitat 525 526 (Table S5.2). Additionally, the significance of slopes in response to climate variables 527 was greater in the mesic heath habitat (Figure 4, 5), but the statistical analysis did not find any significant differences in rates of advanced peak phenology between 528 529 habitats (except for Coccoidea advancing peak phenology more in mesic heath 530 habitats) (Table S5.1). The two dominant predator families; Lycosidae and 531 Linyphildae, extended their activity periods in all habitats in response to warmer 532 summers and earlier snowmelt, but Linyphiidae delayed peak phenology in response 533 to earlier snowmelt whereas Lycosidae advanced peak phenology (Figure 4).

534

535 Discussion

The Zackenberg area of Northeast Greenland has experienced warmer summers but also large interannual variation in timing of snowmelt and temperature in the past 25 years. Consistent with a lack of directional change in climate conditions, the arthropod community phenology showed only weak advancements through the study period, and variation amongst taxa and habitats in the rate of temporal phenological advancements were large. Timing of snowmelt was a key driver of arthropod phenology, which was particularly evident following detrending of

543 the phenological time series where the temperature-phenology relationship was 544 markedly reduced but the snowmelt-phenology relationship persisted. We found that 545 while most arthropod taxa in this high-Arctic community were able to track climate 546 change well, the climate-associated phenological shifts varied among taxa and populations in different habitat types. Specifically, pollinators demonstrated rapid 547 phenological advancements and, in some habitats, shortened seasonal activity in 548 549 response to earlier snowmelt and warmer summers, while predators responded by extending their seasonal activity. We were able to show that late active taxa and 550 551 populations in late snow melting habitats advanced their phenology more in 552 response to earlier snowmelt than early active taxa and populations in early snow melting habitats. Our results indicate that the community-wide trends may mask the 553 554 variation in direction and magnitude of phenological shifts in different taxa and locally 555 adapted populations, thus emphasizing the idiosyncratic nature of individual taxa and population responses to climate change. These findings further highlight the 556 557 substantial heterogeneity in phenological responses to climate change in arthropods and could facilitate a mismatch in the seasonal timing of interacting organisms. 558

Our study highlights the importance of timing of snowmelt as an 559 560 environmental predictor of arthropod phenology, which could be a response to the 561 large interannual variation in snowmelt patterns (Høye, Post, Meltofte, & Schmidt, 562 2007; Pedersen et al., 2016). Importantly, the reduced temperature-phenology 563 relationship and the persisting snowmelt-phenology relationship found after detrending by year shows that multiple drivers of phenology may exist and 564 565 accentuates that the importance of climate variables can vary depending on region. 566 Thus, for predicting Arctic arthropod phenological responses to climate change, 567 incorporating information on timing of snowmelt, other than temperature, is

important, which is also suggested from other cold region arthropod phenology
studies (Kankaanpää et al., 2018; Leingartner et al., 2014; Mortensen et al., 2016;
Ovaskainen et al., 2013).

571 We found a greater phenological sensitivity of arthropod taxa to timing of snowmelt in late compared to early snow melting habitats demonstrating local spatial 572 573 variation in the strength of arthropod phenological responses to timing of snowmelt. 574 Tundra environments can create substantial spatial variation in arthropod emergence and abundance across small spatial scales because of local differences in timing of 575 576 snowmelt (Høye & Forchhammer, 2008; Kankaanpää et al., 2018; Koltz, Schmidt, et 577 al., 2018) and temperature at the soil surface due to variation in vegetation and snow cover (For example, Bjorkman et al., 2020; Elmendorf et al., 2012). In an 578 579 environment with inherently high spatial variability (Kankaanpää et al., 2018) 580 combined with large interannual variation in climate variables, the environmental 581 controls of phenology as well as the rate of phenological responses of arthropod taxa 582 to climate may vary substantially. Thus, populations of taxa from late snow melting 583 habitats may be more plastic allowing them to track timing of snowmelt. Conversely, it can be a disadvantage for taxa in early snow melting habitats to track timing of 584 585 snowmelt as the interannual variation in snowmelt increases the possibility of taxa 586 being exposed to unfavourable environmental conditions (Regan & Sheldon, 2023; 587 Rixen et al., 2022), such as exposure to freezing. This could explain the strong 588 responses of populations of taxa in late snow melting habitats to timing of snowmelt 589 compared to populations in early snow melting habitats. These results highlight the 590 implications of failing to capture the spatial variation in phenological responses to 591 climate change as it may encompass under- or overestimating the vulnerability of 592 arthropod communities to climate warming because the sensitivity of taxa to climate

593 change is highly habitat dependent.

594 We find extended activity periods for most arthropod taxa in the late 595 snow melting mesic heath habitat, which could indicate that arthropod taxa attain a 596 wider phenological niche in early snow melting years. Conversely, some taxa from early snow melting habitats experienced shorter activity periods and thus a reduction 597 598 in their phenological niche. We cannot reject that the coarse taxonomic resolution in 599 this study masks information on species composition between plots. Different arthropod species and populations vary in their phenology within a community. 600 601 Hence, the site-specific phenologies could be attributed different species 602 compositions in early and late snow melting sites (Bowden & Buddle, 2010; Hansen et al., 2016). Previous studies from Zackenberg have demonstrated different species 603 604 assemblages of muscid flies and spiders between habitats (Bowden et al., 2018; 605 Loboda et al., 2018). Thus, the spatial variation in phenological shifts could be due to interspecific variation in climatic sensitivity such that late active species respond 606 607 more rapidly to warmer seasons. The likelihood of an underlying taxonomic 608 variability in phenological sensitivity of taxa to climate change in our study may entail a conservative estimate of the magnitude of phenological shifts in response to a 609 610 changing climate and we could be underestimating the true ecological effects of a 611 changing climate.

Our findings document varying rates of phenological responses among functional groups to environmental cues. Pollinator and parasitoid peak phenology were earlier in years with earlier snowmelt than peak phenology of predators. Some families of pollinators had shorter activity periods in years with early snowmelt, though this trend was highly habitat dependent. The phenology of herbivores and decomposers were not as strongly related to climate variables compared to other

618 functional groups suggesting that other abiotic or biotic cues may drive the phenology of these taxa. These results confirm that most arthropod taxa in this high-619 620 Arctic community are able to track environmental changes, but also highlights the 621 variation in climate-associated phenological responses among functional groups. This strongly increases the complexity of phenological impacts in a community 622 623 context. Ultimately, this may entail major reorganization in the arthropod community, 624 which could have considerable impacts on the way organisms interact within ecological networks (Walther, 2010). For example, earlier and shorter pollinator 625 626 activity periods could temporally decouple them from flowering host plants leading to 627 a reduction in pollination services (Pyke et al., 2016, Schmidt et al., 2016), and extended activity of predators could influence top-down control in a food web (Koltz, 628 629 Classen, et al., 2018). Further, as the community-level activity of arthropods at 630 Zackenberg has advanced in the past 25 years, the risk of trophic mismatch between 631 breeding Arctic shorebirds' insectivorous chicks and the availability of arthropod prey 632 increases (Both et al., 2009; Reneerkens et al., 2016). Meanwhile, the host-633 parasitoid interaction could be kept intact as parasitoids seem to be following their hosts (families of predators and pollinators). We do suspect that the different 634 635 responses of arthropod taxa to climate in this study could reflect their life history 636 strategies (Buckley, 2022; Gallinat et al., 2015). As temperatures rise, the growth 637 rate of insects increase (Chaves et al., 2015), and univoltine insects, that dominate 638 in the Arctic (Høye et al., 2020), are therefore expected to advance their fall senescence. On the other hand, invertebrate species may extend their activity 639 640 seasons by producing additional generations in response to longer growing seasons 641 (Altermatt, 2010; Braune et al., 2008; Kerr et al., 2020; Roy & Sparks, 2000), hence 642 the shorter activity seasons we find for some taxa in warmer years. We also found

that the predators are rapidly extending the duration of their activity in response to a
warmer climate. This trend could be attributed climate warming facilitating the
production of additional generations of spider species, which has been documented
for the wolf spider *Pardosa glacialis* at Zackenberg (Høye et al., 2020). Thus, the
extended duration of activity of the spider families found in our study across the past
25 years could reflect an adaptation in reproductive strategies.

649 Our results point towards pollinators being particularly climate sensitive. At one of our local sites, pollinators became active 6.8 days earlier per decade, 650 651 which is greater than what has been observed for temperate pollinators. For 652 example, wild bees in North America have advanced their phenology in the range of 0.8 and 1.62 days per decade in the past 40 - 50 years (Bartomeus et al., 2011; 653 654 Dorian et al., 2022). In addition, species of temperate butterflies and bees show 655 patterns of increased flight duration (Altermatt, 2010; Dorian et al., 2022; Michielini et al., 2021), and while pollinators have not changed the duration of activity across 656 657 years in our study, the important pollinator families Chironomidae and Muscidae shortened the length of their activity seasons in early snow melting years in the wet 658 fen. From the perspective of pollination services in the community, shifts in the timing 659 of activity and changes in flight duration of pollinators is concerning because it may 660 influence the phenological overlap with plant species that depend on pollination for 661 662 reproductive success, such as Dryas sp. At Zackenberg (Tiusanen et al., 2016). 663 Relevant to this, the dominant plant species at Zackenberg have been found to rapidly advance their phenology (Høye, Post, Meltofte, Schmidt, et al., 2007; Høye et 664 al., 2013; Iler et al., 2013) with average peak flowering occurring 8.8 days earlier per 665 decade (Iler et al., 2013). In comparison, the average peak activity of pollinators in 666 this study was 3.1 days earlier per decade. While the pollination services at 667

Zackenberg seem to be stable till now (Cirtwill et al., 2022), there is a concern that
the temporal overlap between flowering plants and pollinators may be further
reduced (Schmidt et al., 2016).

671 The heterogenous phenological responses among arthropod taxa and habitats entail that community- and family-level changes in phenology does not 672 673 necessarily correspond. A few dominant taxonomic groups may drive the community 674 response or compensate with dynamics that moderate the community shift, for example if climate change is causing changes in species composition with varying 675 676 seasonal activities (Walther, 2010). Our study would certainly benefit by linking 677 species-level changes in phenology to our findings on family-, population- and community-level phenology (Walters et al., 2013). However, our results have 678 679 important implications in terms of incorporating both spatial and temporal 680 components of phenological variation among arthropod taxa to climate variables, which is generally lacking in studies investigating arthropod population and 681 682 community phenology (Ellwood et al., 2012; Samplonius et al., 2021). At the same 683 time, we demonstrate that arthropod taxa respond strongly to a changing climate, and on a community-level, arthropods seem to be tracking the changing climate 684 quite well, however, arthropod taxa track climate variables at different rates and this 685 686 diversity in phenological shifts can lead to substantial reshaping of communities with 687 potential implications for species interactions.

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