A probabilistic approach to estimating timber harvest location

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Abstract

Determining the harvest location of timber is crucial to enforcing international 23 regulations designed to protect natural resources and to tackle illegal logging and 24 associated trade in forest products. Stable Isotope Ratio Analysis (SIRA) can be used to 25 verify claims of timber harvest location by matching levels of naturally-occurring stable 26 isotopes within wood tissue to location-specific ratios predicted from reference data 27 ('isoscapes'). However, overly simple models for predicting isoscapes have so far limited 28 the confidence in derived estimates of timber provenance. In addition, most use cases 29 have limited themselves to differentiating between a small number of pre-determined 30 location options. Here, we present a new SIRA data analysis pipeline, designed to infer 31 the harvest location of a focal tree out of a continuous, arbitrarily large area. We use 32 Gaussian Processes to robustly estimate isoscapes from reference wood samples, and 33 overlay with species distribution data to compute, for every pixel in the study area, the 34 probability of it being the origin of the examined timber. This is the first time, to our 35 knowledge, that this approach is applied to determining timber provenance, providing 36 probabilistic results rather than a binary outcome. Additionally, we include an active 37 learning tool to identify locations from which additional samples would maximize the 38 improvement to model performance, allowing for optimisation of field efforts. We 39 demonstrate our approach on a set of SIRA data from seven oak species in the USA as a 40 proof of concept. Our method can determine the harvest location up to within 520 km 41 from the true origin of the sample and outperforms the state-of-the-art approach. 42 Incorporating species distribution data improves accuracy by up to 36%. The 43 future-sampling locations proposed by our tool decrease the variance of resultant 44 isoscapes by up to 86% more than sampling the same number of locations at random. 45 The method we present here greatly advances the toolset available for verification of 46

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47 timber harvest location, will empower authorities worldwide in enforcing anti-deforestation

⁴⁸ legislation and will help protect natural resources.

⁴⁹ Keywords: stable isotopes, origin traceability, timber provenance, illegal logging, isoscapes,

50 Gaussian Processes

⁵¹ **1** Introduction

Unsustainable exploitation of natural resources is the largest driver of terrestrial biodiversity loss 52 after land-use change (Díaz et al., 2019) and a major conservation challenge globally. To 53 prevent a sixth mass extinction (Barnosky et al., 2011), nearly 200 nations have recently agreed 54 on a new set of targets and goals under the Kunming-Montreal Global Biodiversity Framework. 55 In particular, Target 5 of the agreement includes the objective to "ensure that the use, 56 harvesting and trade of wild species is sustainable, safe and legal, preventing 57 overexploitation" (2022 UN Biodiversity Conference, 2022). Meeting this ambitious target will 58 require overcoming a key element of unsustainable use of natural resources: the illegal harvest 59 of threatened tree species. 60

Legal frameworks have been established to combat illegal logging and trade in illegally harvested 61 timber, such as the Convention on the International Trade in Endangered Species (CITES), the 62 US Lacey Act (amended 2008), the UK Timber Regulation (2021), the EU Deforestation 63 Regulation (EUDR; 2023) and the Australian Illegal Logging Prohibition Act (2012). The new 64 policies place additional traceability and reporting requirements on companies trading in wood 65 and agricultural products (Dormontt et al., 2015). For example, the EUDR requires operators 66 to record and report the coordinates of production location (forest or farm), and enforcement 67 officials will be expected to scrutinize those claims of harvest location. Despite the 68

comprehensive legislation already in place and the international commitments under current 69 adoption, enforcement of such regulations remains a challenge. Illegally harvested timber is 70 shipped under false declarations of origin or mixed into legal shipments, and methods for 71 verifying geographical location have so far only been able to determine the correct location out 72 of a few pre-determined options, mostly at country-level resolution (Watkinson et al., 2022a; 73 Horacek et al., 2009; Muñoz-Redondo et al., 2021). This challenge is greatly intensified by the 74 new EUDR legislation adopting precise geographical location (GPS point or polygon for plot of 75 land) as a determinant of the legal status of timber. 76

1.1 Stable Isotope Ratio Analysis to verify provenance

⁷⁸ Well-established scientific techniques enable measurement of the chemical, anatomical and
⁷⁹ genetic features of plants from a tissue sample (Deklerck, 2023), with ever increasing precision
⁸⁰ and availability. When compared against a robust physical reference collection, these attributes
⁸¹ of the tissue can be used to (in-)validate declared species and origin claims, and support
⁸² enforcement officials in their efforts to detect, for example, illegally harvested timber or fraud in
⁸³ supply chains.

Stable isotope ratio analysis (SIRA) is one of the most promising technologies in this context. 84 Several elements within biological tissues (mainly Hydrogen, Oxygen, Carbon, Sulfur, Nitrogen) 85 have multiple naturally-occurring stable isotopes, whose ratios vary predictably across space, in 86 correlation with environmental conditions (West et al., 2010; Siegwolf et al., 2022; Gay et al., 87 2022; Pederzani and Britton, 2019). The heavy isotopes of these elements do not undergo 88 radioactive decay, and their proportion can be readily detected by mass spectrometry (Boner 89 et al., 2007). The isotopic composition of elements incorporated into the tissues of a plant is 90 determined by soil properties, climate, metabolic fractionation and other biotic and abiotic 91

conditions characteristic of the species and the habitat in which the individual grows (Siegwolf 92 et al., 2022; Camin et al., 2017; Horacek et al., 2009; van der Sleen et al., 2017; Gay et al., 93 2022). Hence, differences in stable isotope ratios among individuals correspond to the 94 environment they grew in, and can be used to discriminate between plants from different 95 geographic areas. SIRA has proven useful in determining risk of illegally harvested material in a 96 wide variety of contexts, for example, forest products (Watkinson et al., 2020; Boner et al., 97 2007), wildlife trafficking (Bowen et al., 2005; Koehler et al., 2019; Wunder and Norris, 2008; 98 Vander Zanden et al., 2015), ivory trade (Van der Merwe et al., 1990; Ziegler et al., 2016), 99 agricultural products (Camin et al., 2016; Saadat et al., 2022), fish/seafood (Cusa et al., 2022; 100 Silva et al., 2021; Kroetz et al., 2020), precious metals (Kirk et al., 2003), and natural and 101 synthetic illegal drugs (Kurashima et al., 2004; Casale et al., 2005), but without the spatial 102 discrimination required by the new timber legislation. 103

104 1.2 Modelling approach

Current modelling practices for the use of SIRA to verify harvest location of both legally and 105 illegally harvested forest products require improvement. The use of SIRA is currently limited by 106 the simplistic models used, as well as by the limited number of reference samples used as input 107 data for such models. Reference sampling campaigns are costly and budgetary needs are often 108 underestimated, with sampling locations often taking into account relative ease of sampling 109 rather than areas that yield a gain in model prediction accuracy (Schmitz et al., 2019). There 110 has been considerable development of isoscapes ("isotope landscapes"), given that stable 111 isotopic variation is a continuous spatial variable in nature (West et al., 2010). These isoscapes 112 are geospatial maps that show the isotopic variation of the material of interest (West et al., 113 2010). While the potential of isoscapes for determining forest product origins has long been 114

recognized, few rigorous methods exist to achieve this task. The existing methods use simple prediction strategies such as linear regression (Watkinson et al., 2020, 2022b), which do not fully leverage the information contained in isotope ratio data.

Gaussian Process (GP) regression, which is closely related to kriging in geostatistics literature, 118 is a class of flexible regression models which use the values in sampled points to estimate the 119 values in surrounding points (Li and Heap, 2008; Deklerck, 2023; Williams and Rasmussen, 120 2006). A key advantage of GP regression is that it can quantify the uncertainty of its own 121 predictions based on the inferred spatial covariance structure of the samples. The importance of 122 quantifying the uncertainty of predictions is increasingly recognized in safety-critical (Jankowiak 123 et al., 2020) and forensic (Su and Srihari, 2010; Swofford and Champod, 2022) machine 124 learning applications. Additionally, GP regression facilitates inference of a sparsely sampled 125 variable of interest from variables that are highly correlated with it but more densely sampled 126 (Adhikary et al., 2017; Kanankege et al., 2018). In the context of plant harvest location 127 estimation, this translates to inferring stable isotope ratios from atmospheric drivers (such as 128 precipitation, temperature and water vapor pressure) known to influence the stable isotope 129 signal in wood (Horacek et al., 2009; Siegwolf et al., 2022). This then provides a powerful tool 130 for predicting the isotopic composition in areas that have not yet been sampled. However, 131 previous work on timber isoscapes used GP regression primarily as a spatial interpolation 132 technique without a probabilistic interpretation (Gori et al., 2018; Watkinson et al., 2022a). 133 Others used approximate GP models to derive variance estimates for origin determination in 134 animals (Ma et al., 2020; St. John Glew et al., 2019). 135

Here, we develop GP-based probabilistic machine learning models to infer timber harvest
 location by directly modelling timber isoscapes from SIRA data, with the aid of atmospheric
 predictors and species distribution data. We show that probabilistic modeling greatly enhances

the utility of SIRA in estimating the geographical origin of timber, and, assisted by a reference
 dataset (Gasson et al., 2021), can be used to guide future sample collection by identifying
 sampling locations that will minimize prediction uncertainty.

¹⁴² 2 Materials and Methods

143 2.1 Data sets

We use data from 87 trees of the genus *Quercus*, sampled across the contiguous United States, 144 as described in (Watkinson et al., 2020). Stable isotope ratio measurements were done 145 following the protocol described in (Boner et al., 2007). Each entry contained stable isotope 146 ratio measurements of oxygen δ^{18} O (ratio between ¹⁸O and ¹⁶O), hydrogen δ^{2} H (ratio between 147 2 H and 1 H), carbon δ^{13} C (ratio between 13 C and 12 C) and sulfur δ^{34} S (ratio between 34 S and 148 32 S) as well as the GPS coordinates of the sampled tree. As stable isotope ratios are largely 149 driven by environmental conditions such as precipitation, temperature, humidity and so on, 150 publicly available datasets for these factors are used to aid the inference of isoscapes. We used 151 the following atmospheric data: $\delta^2 H$ and $\delta^{18} O$ isotopic composition of precipitation (Bowen and 152 Revenaugh, 2003), water vapor (Borbas, 2015) (found to be associated with δ^{13} C by 153 (Watkinson et al., 2020)), reflected shortwave radiation (NEO, 2023) and precipitation 154 (multi-satellite) (Huffman et al., 2020), both of which were found to be associated with $\delta^{34}S$ 155 (Watkinson et al., 2020). For each of those data types, we used monthly means averaged over 156 a number of years to minimize the impact of weather patterns in specific years (see (Watkinson 157 et al., 2020) for precise year ranges). 158

¹⁵⁹ To inform the priors (probability distributions representing the prior belief on possible tree ¹⁶⁰ locations) of the models we develop, we used species inventory data across the natural range of

¹⁶¹ each species within the United States (Wilson et al., 2013), downloaded from:

https://www.fs.usda.gov/rds/archive/Catalog/RDS-2013-0013 on 09/12/2022. This
 data is available as species-specific raster layers of tree abundance at 250m resolution. We then
 used the function project() of the R package *terra* (Hijmans, 2022) to bilinearly aggregate
 abundance data so that it matched the spatial resolution of other spatial data in the pipeline.

¹⁶⁶ 2.2 Model architecture

Figure 1 presents an overview of the data sets and components comprising our model and output. We use a rectangular grid to represent the study area. Grid points are placed every 0.125 degree latitude (\approx 14 km) and every 0.06 degree longitude (\approx 4.3-6.0 km), which allows us to approximate spatial probability distributions with high accuracy. For every isotope ratio (IR), we fit a GP regression model to the training data to obtain the posterior mean and variance of the isotope ratio for every point of the grid (see Appendix for the full detail on implementation).

Gaussian Processes are a class of flexible regression models, which enable the modeler to 174 quantify the uncertainty about specific predictions. A GP regression model assumes that the 175 responses (in our case, isotope ratios) at different locations are jointly normally distributed 176 (Gaussian). The model is defined by three elements: (1) the mean, for which we use a 177 constant, (2) the covariance function for which we use the Matern function (Williams and 178 Rasmussen, 2006) with separate scaling parameters for latitude and longitude and (3) the noise 179 parameter. The choice of mean and covariance functions reflects prior knowledge and modelling 180 assumptions about the regression problem. The covariance function expresses the amount of 181 information about unobserved locations contained in nearby observed values. The function 182 parameters as well as the noise parameter are estimated by maximizing the likelihood of the 183

training data, in contrast to standard kriging approaches in geostatistics literature, which use
 approximate techniques based on summary statistics. We use GPyTorch (Gardner et al., 2018)
 to efficiently find the maximum likelihood parameter estimates.

The input to the GP consists of the coordinates and/or the climate variable values at the grid 187 point. For a combination of observed stable isotope ratios (y_O, y_H, y_C, y_S) (meaning δ^{18} O, δ^2 H, 188 δ^{13} C, δ^{34} S), we compute the likelihood of this observation at every point in the grid, using the 189 four GP regression models estimated in the previous step. This likelihood is the product of 190 likelihoods for each isotope ratio as we assume independence between isotopes. Given the prior 191 and the likelihood, we compute the posterior probability of each grid point being the harvest 192 location of the sample by multiplying the prior and the likelihood and normalizing so that the 193 probabilities sum up to 1. For ease of interpretation, the output is a map with highest-posterior 194 density (HPD) regions indicated for several probability levels (15%, 30%, 50%, 75%, 90%, 195 95%). 196

To incorporate atmospheric data into the isoscape we use monthly averages of the atmospheric variables listed in Section 2.1. We use a linear covariance term to model the covariance component corresponding to the variation in the respective atmospheric variables. The linear covariance function models a linear relationship between the atmospheric variable and the response and is mathematically equivalent to Bayesian linear regression (Williams and Rasmussen, 2006). The overall covariance function is then the sum of the spatial and linear terms and can be seen in Appendix A.

We use the spatial density maps developed by (Wilson et al., 2013) to design two prior distributions for sample origin that account for the spatial distribution of oak species. The first, which we call the *density prior*, holds that the probability of a sample originating at a grid cell is proportional to the basal area (average amount of area occupied by tree stems per unit of

space) recorded at the grid cell. The second, which we call the *range prior*, assigns equal probability to every grid cell where above-zero basal area has been recorded. In addition, both priors allow for a small probability that a sample might occur outside its observed range - we set that probability to 0.01 and diffuse it uniformly over all grid points within the contiguous United States where the species does not occur according to (Wilson et al., 2013).

213 **2.3** Performance evaluation

We perform 5-fold cross-validation on the data set and report the average values of all 214 performance metrics over all data points. Samples with incomplete or ambiguous species 215 information and samples collected in botanical gardens outside of their species' native range are 216 excluded from the test sets, but not from the training sets, resulting in a total of 74 test 217 samples across the 5 folds. We report performance of our models as well as our implementation 218 of the approach by (Watkinson et al., 2020) averaged across the five cross-validation folds. 219 Rigorously evaluating the performance of our models is a non-trivial task as each model 220 produces a distribution over possible locations, rather than a single location. For this reason, we 221 have defined several metrics to investigate different aspects of probabilistic harvest location 222 prediction: 223

Predictive log-likelihood and log-posterior: We report the log-likelihood and the
 log-posterior of observing the sample at its true origin. Both of those measure how well
 the model fits the test data.

Mode distance: We report the great circle distance between the true location and the
 mode of the posterior distribution, i.e. the highest scored location according to the
 model. This metric measures the accuracy of the highest-scored locations, but it does not

account for the amount of uncertainty in model predictions.

3. Mean absolute error (MAE): To investigate how distant our predicted locations are from
 the truth, we report the expected distance between the true location and a location
 sampled randomly from the posterior distribution returned by our model

$$MAE = \int_{\mathbf{x}\in A} d(\mathbf{x}_t, \mathbf{x}) p(\mathbf{x}|\bar{y}, S) d\mathbf{x}$$

where d() is the great circle distance between the two points and $p(\mathbf{x}|\bar{y},S)$ is the 234 posterior probability of \mathbf{x} being the harvest location. A perfect prediction would have the 235 distance of 0. This metric will favour predictions concentrated around the true location 236 over equally dispersed predictions concentrated elsewhere. It will also favour less dispersed 237 predictions generally. For the method of Watkinson *et al.*, which only outputs a region of 238 plausible locations, we assume a uniform distribution within the region highlighted by the 239 model. In practice, isoscapes often predict similar isotope ratio values at distant locations, 240 so even a statistically efficient method might yield a high MAE value. 241

4. Area scored higher than the true location (ASH): The behaviour of MAE is influenced by
the shape of the posterior distribution, which favours unimodal over multimodal shapes.
We report the total surface area corresponding to the points that the model considers
more plausible than the true origin of the sample.

$$ASH = \int_{\mathbf{x}\in A} I\left[score(\mathbf{x}|\bar{y},S) > score(\mathbf{x}_t|\bar{y},S)\right] d\mathbf{x}$$

where I(.) is the indicator function that yields 1 when the statement is true and 0 otherwise. For all GP models, the score is the posterior probability of harvest location,

whereas for the method of Watkinson *et al.* we take the score to be the negative of the minimum value of the threshold that results in the true location being included in the highlighted region. In contrast to MAE, this metric is likely to give a low value to a posterior distribution that is concentrated in several small areas as long as one of those areas contains the true location. For example, if the true location could be a county in New York or a county in West Virginia, this would give a low ASH but high MAE as the two counties are far apart.

255 2.4 Guiding future sampling efforts

Field sample collections are time-consuming and expensive. We can optimize future field collections using informed prediction of where additional sample points are most needed for increasing origin estimation accuracy. The isoscape variance estimates provided by GPs can be used to guide future sampling efforts, which in turn will maximize the performance of the model. This approach is known as *active learning* in the machine learning literature. Here, we propose a strategy to minimize the error of our isoscape estimates by carefully choosing future sampling locations.

Early attempts at efficient active learning in GPs involved collecting samples at points with 263 highest response variance or, equivalently, picking a set of points that maximizes the entropy of 264 responses (Cressie, 2015). Unfortunately, this approach tends to recommend collecting samples 265 on the boundaries of the study area, which is wasteful as the newly collected samples improve 266 isoscapes in a smaller fraction of the study area than if they were placed away from the 267 boundary. This motivated researchers to propose several criteria for optimizing 268 sampling (Guestrin et al., 2005; Ramakrishnan et al., 2005). Here, we adopt an approach similar 269 to that of (Guestrin et al., 2005) with a few modifications designed to address the large size of 270

our spatial grid, which renders their original method computationally intractable for our data set. We seek to maximize the *average* reduction in predictive variance across our study area that can be achieved by adding a sample to the training set. With S the set of sampled locations and Gthe set of grid points, we define the information gain (IG) associated with adding a new point (\mathbf{x}^*) to the training data set as

$$IG(\mathbf{x}^*) = \sum_{\mathbf{x}\in G} \left[\left(\log(\sigma^2(\mathbf{x}|S)) - \log(\sigma^2(\mathbf{x}|S \cup \{\mathbf{x}^*\})) \right) \right]$$
(1)

where the predictive variances are computed using Equation A.4. The algorithm then picks the 276 point in the grid that yields the highest IG. Importantly, the predictive variances depend only on 277 the sampling locations and not on the sampled values, so it is possible to sequentially propose 278 multiple sampling points before collecting the samples. Our method sequentially proposes 279 sample collection points until a user-specified number of samples is reached. We assume that 280 samples can only be collected in locations where at least one of the species is present. Thus, 281 grid points that lie outside every species range are excluded from the procedure. Our active 282 learning strategy requires repeatedly computing a large number of predictive variances for 283 varying training sets. To reduce computation time, we randomly downsample our grid to 15000 284 points before running the analysis. In addition, we assume that the reduction in variance is 285 negligible for grid points situated more than 15 degrees away from the newly sampled point (x^*) 286 in longitude or more than 7.5 degrees in latitude. 287

288 **3** Results

²⁸⁹ 3.1 Model accuracy and comparison

The plausible location areas identified by our models consisted of points within an average 290 distance of 520-870 kilometers from the true location of the oak tree samples, depending on 291 model settings. Even with a relatively small training data set of 69 - 70 training samples 292 (depending on the cross-validation fold), our model is able to exclude the vast majority of the 293 study area from consideration as a possible source of the focal sample. All our models 294 outperform the state-of-the-art method for determining timber harvest location (Watkinson 295 et al., 2020) in most or all metrics. Table 1 shows performance metrics for all the models on 296 the test data set. 297

Incorporating species distribution information improves prediction performance for every model and every metric examined except the log-likelihood, which is computed independently of the prior. Informative priors improve MAE by 16% to 35% and ASH by 15% to 57% with most improvement for the pure spatial model and least for the spatial+atmospheric model. The more informative density prior gives better accuracy than the range prior according to all the metrics. Predicted probability maps for a few test points are shown in Fig. 2 and 3.

The spatial-only GP model gives the closest location predictions to the true location of the tree samples, except when a flat prior is used. In general, the spatial-only model and the combined spatial+atmospheric model give similar results on all metrics and outperform the atmospheric-only model in almost all settings. Somewhat surprisingly, the combined model does not outperform the spatial-only model. This might be due to the relatively small dataset used here or the choice of atmospheric predictors, and remains to be tested as we continue to expand our reference databases. The predictions of atmospheric GP models appear qualitatively

different from those from the purely spatial GP, perhaps because atmospheric model predictions emphasize geographical areas with distinct climate patterns, such as Appalachia or the Gulf Coast. Unsurprisingly, the purely spatial GP identifies areas that are more spatially cohesive but do not share any obvious physical features.

315 3.2 Guiding future sampling efforts

We investigated the performance of our active learning strategy on the US oak data set. For the spatial-only model, we let our method propose 10 new sampling locations to add to the training data set in the first cross-validation fold and computed the predictive variances before and after including the proposed locations.

The resulting isoscape standard deviation maps are shown in Figure 4. Our active learning 320 strategy proposes sampling locations in currently undersampled regions with high predictive 321 variance and sampling in those areas results in a visible improvement. The highest decrease in 322 predictive variance was observed for $\delta^2 H$ while the lowest decrease was observed for $\delta^{18} C$. Most 323 of the chosen locations are close to, but not at the boundary of, the allowed sampling area. 324 To investigate the efficiency of our active learning procedure, we compared isoscape variances 325 resulting from active learning with those resulting from adding the same number of points 326 sampled randomly from the allowed sampling area. We generated 100 such variance maps and 327 compared the average variance (across the allowed sampling area) of those maps with the maps 328 in Fig. 4. Appendix B shows the average predictive variances as a function of the number of 329 points added for both random and active learning sampling strategies. We see that our active 330 learning strategy results in a substantially faster decrease in predictive variances. After adding 331 10 samples, the reduction in variance achieved by our active learning method is between 64%332 $(\delta^{13}C)$ and 86% $(\delta^{18}O)$ greater than the average reduction achieved by the same number of 333

335 4 Discussion

4.1 Harvest location estimation

To halt illegal logging, to enforce timber regulations and to protect biodiversity in forested 337 landscapes, we need to be able to accurately estimate timber harvest location. Although several 338 examples exist of applying SIRA for timber origin questions (Gori et al., 2018; Watkinson et al., 339 2020; Kagawa and Leavitt, 2010), these approaches do not take full advantage of (1)340 atmospheric and species distribution datasets available or (2) state-of-the-art probabilistic 341 machine learning models. In addition, many SIRA use-cases limit themselves to a classification 342 problem (country X versus country Y) compared to a continuous assignment problem (true 343 harvest location). In response to growing evidence of fraud in supply chains, legislation 344 increasingly requires operators to trace back to plot (for example the EU Deforestation 345 Regulation). Consequently, determining the true harvest location will likely become increasingly 346 important. In this work we present a new computational pipeline which aims at taking 347 advantage of both (1) and (2) while predicting the harvest location as a continuous variable. 348 The accuracy of our models depends on the specific modelling approach and the data sets used. 349 Using prior information about species distribution results in a considerable increase in accuracy 350 regardless of which model is used by all metrics considered. The impact of adding species 351 distribution data appears to be greater for the spatial-only model than models that use 352 atmospheric information. This could be due to climate patterns influencing both species 353 distributions (habitat suitability) and the values of the atmospheric variables that we 354 incorporated in our models, which renders species distribution information more redundant once 355

³⁵⁶ atmospheric variables have been included in the model.

Within timber tracing literature, our method bears the most resemblance to the work 357 of (Watkinson et al., 2020), which uses linear regression to predict isoscapes based on 358 atmospheric data. Their approach assumes a constant variance across the study area. In 350 contrast, our method estimates the predictive variances based on the spatial covariance 360 structure learned from the reference data, which enables us to translate differences in sampling 361 density across regions into varying levels of confidence in isoscapes across space. Our method 362 also assumes a linear relationship between atmospheric predictors and isoscapes, but our GP 363 formulation implicitly integrates over plausible values of regression parameters, which should 364 lead to more robust predictions compared to standard linear regression. In addition, our 365 approach makes use of species distribution data, which yields substantially improved predictions 366 compared to uninformative priors. Finally, our approach enables us to propose locations for 367 further sample collection that maximize the utility of the samples. 368

The estimation of spatial covariance structure has recently attracted attention in animal stable 360 isotope studies. (Ma et al., 2020) recently proposed a method that uses probabilistic 370 precipitation isoscapes derived from a GP (Courtiol et al., 2019), which are then calibrated to 371 produce isoscapes for the species of interest. (St. John Glew et al., 2019) introduced a model 372 combining spatial and environmental effects using a novel likelihood approximation for isoscape 373 estimation, though the main focus of their work is isoscape modelling, not origin estimation. 374 These approaches differ from ours in that 1) they rely on Laplace approximations for isoscape 375 estimation rather than exact likelihood maximization; 2) they use ordinary least-squares 376 regression to account for atmospheric predictors, whereas our method uses a Bayesian approach 377 via a linear covariance term; and 3) they do not aim to actively improve isoscapes through 378 additional sampling. A common feature between these models and ours is using a grid to 379

compute the posterior distribution of origins, which to the best of our knowledge was first
 considered by (Wunder, 2010).

³⁸² Our current best performing model can estimate the harvest location for *Quercus* species to 520 ³⁸³ km across the east of the United States. Future field expeditions will lead to an improvement, ³⁸⁴ especially if the identified priority locations are targeted (see 4.2). The presented model will be ³⁸⁵ adapted to other use cases, with a focus largely on endangered tropical species which are under ³⁸⁶ high logging pressure.

We expect that our models will be more accurate once more timber samples become available. 387 The size of the current data set of wood samples available to this study (87 samples) is quite 388 small relative to the area of the contiguous United States, which inevitably results in large 389 predictive variance in many areas. In addition to reducing uncertainty about undersampled 390 areas, larger data sets (in the range of hundreds to thousands of samples collected from across 391 the US) should also enable researchers to use more complex GP models, including models with 392 heterogeneous noise (Binois et al., 2018), or deep GP models where the covariance function is 393 modelled by a neural network (Wilson et al., 2016). 394

4.2 Guiding future collection efforts

³⁹⁶ Under the World Forest ID Programme (Gasson et al., 2021), tens of thousands of tree samples ³⁹⁷ are being collected globally, and are being analysed by different techniques, including SIRA, to ³⁹⁸ build georeferenced databases which can be used to identify timber harvest origin. Our active ³⁹⁹ learning approach can be used to inform future sample collection efforts and increase model ⁴⁰⁰ accuracy that can be achieved within a fixed sampling budget. This will be especially important ⁴⁰¹ in tropical regions, where reaching sampling sites can be difficult, time intensive and expensive. ⁴⁰² A good sampling design can substantially improve model performance (Contina et al., 2022),

and our method can be used to adapt sampling efforts as more data is analysed. Our current approach focuses on minimizing predictive variances without considering the impact of newly sampled points on model parameters. Extending our approach to *non-myopic* sampling (Krause and Guestrin, 2007), which considers the impact on model parameters, would constitute an interesting future research direction. Another avenue for improving our approach would be to augment our IG criterion to reflect the varying investment in collecting samples as a function of the time, logistics, and financial cost of reaching the desired sampling location.

410 5 Conclusion

The accurate estimation of geographic origin of globally traded wood products is a critical step 411 in combating illegal logging and associated trade, by supporting authorities' ability to verify 412 claims made by traders at any supply chain node. In this work we presented a novel analytical 413 pipeline that brings together and incorporates multiple data types and algorithms. This 414 methodology is able to accurately predict timber product origin and can be used to optimize 415 future field sampling to further increase accuracy and precision. We hope that this work will 416 inspire more efforts to expand reference collections of wood samples, and that governments and 417 companies will more routinely use the technological tools at their disposal to have more 418 oversight over their supply chains and promote a more sustainable use of natural resources. 419

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7 Conflict of interest statement

⁴³⁵ The authors declare that they have no conflicts of interest.

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Table 1: Mean test set performance for all the models used in the study. Best values across all models are shown in bold and underlined whereas values that are not significantly different from the best values (Wilcoxon signed-rank test, p=0.05) are shown in bold. The Spatial-only GP combined with the density prior gives the highest predictive log-likelihood and log-posterior and the lowest MAE and ASH values for all priors used. The Spatial-only model outperforms the other models when range or density priors are used, while the Atmospheric+Spatial model performs best in terms of MAE and ASH when flat priors are used. The inclusion of species distribution information decreases MAE and ASH values for all models used. All of our models outperform the earlier method of Watkinson *et al.* (Watkinson et al., 2020) on most or all metrics.

model	prior	log L	mode distance (km)	MAE (km)	log-posterior	ASH (km^2)
Spatial-only	flat	<u>-6.964</u>	433	809	-9.582	470000
Spatial-only	range	<u>-6.964</u>	435	600	-9.537	327000
Spatial-only	density	-6.964	400	<u>520</u>	-9.059	203000
Atmospheric-only	flat	-7.362	531	870	-9.972	576000
Atmospheric-only	range	-7.362	505	606	-9.797	450000
Atmospheric-only	density	-7.362	534	567	-9.428	311000
Atmospheric+Spatial	flat	-7.149	408	794	-9.518	382000
Atmospheric+Spatial	range	-7.149	<u>399</u>	627	-9.431	315000
Atmospheric+Spatial	density	-7.149	463	536	<u>-8.978</u>	213000
Watkinson <i>et al.</i>	NA	NA	886	859	NA	691000

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Figure 1: Model workflow. We use a training set of isotope ratios from trees collected at known locations and atmospheric data layers ("Reference data"). We fit a Gaussian process regression model to infer isoscapes and associated variance estimates, and compute the likelihood of observing the IR value for each element across the study area. To estimate the source of material with uncertain provenance ("Samples from unknown origin"), the isoscapes are then combined with prior information on the geographical distribution of the species, to yield a probability distribution of origin for the sample. We visualize predicted probability maps by plotting highest-posterior density regions for several probability levels (15%, 30%, 50%,75%,90% and 95%, dark blue to light green).



Figure 2: Harvest location predictions from the three models for 5 points from the test set using the range prior. Darker shades denote areas with higher posterior probability with thresholds set so that the total probability of the colored area is equal to a specified value (see color chart). The red cross indicates the actual location of the tree.



Figure 3: Harvest location predictions from the three models for 5 points from the test set using the density prior. Darker shades denote areas with higher posterior probability with thresholds set so that the total probability of the colored area is equal to a specified value (see color chart). The red cross indicates the actual location of the tree.



Figure 4: Maps showing predictive standard deviations for the four isotopes before and after adding 10 sample locations proposed by our active learning method for the spatial-only model. Standard deviations are only shown within the allowed sampling area, which is the union of ranges for the species in our data set. The red dots show the proposed locations. Our method proposes locations in areas with high predictive variance, particularly for δ^2 H and δ^{34} S. Adding the proposed locations leads to a marked reduction of variance in the neighboring areas.

A More detail on Gaussian Processes

In the following, we give a brief overview of GPs. For an in-depth discussion, see Williams and Rasmussen (Williams and Rasmussen, 2006).

GPs provide a flexible framework for regression, which enables the modeler to quantify the uncertainty of specific inferences. A GP is a random process such that all of its marginals are jointly normally distributed (Gaussian). Let $\mathbf{x} = [x_{lon}, x_{lat}]$ be the GPS coordinates of a sample. For any set of positions $\mathbf{x_1}, \mathbf{x_2}, \ldots, \mathbf{x_n}$, the responses $y_1, y_2 \ldots, y_n$ at those positions are assumed to be jointly normally distributed

$$\begin{bmatrix} y_1 \\ y_2 \\ \vdots \\ y_n \end{bmatrix} \sim \mathcal{N} \left(\begin{bmatrix} m(\mathbf{x_1}) \\ m(\mathbf{x_2}) \\ \vdots \\ m(\mathbf{x_n}) \end{bmatrix}, \begin{bmatrix} k(\mathbf{x_1}, \mathbf{x_1}) & k(\mathbf{x_1}, \mathbf{x_2}) & \dots & k(\mathbf{x_n}, \mathbf{x_n}) \\ k(\mathbf{x_2}, \mathbf{x_1}) & k(\mathbf{x_2}, \mathbf{x_2}) & \dots & k(\mathbf{x_2}, \mathbf{x_n}) \\ \vdots & \vdots & \ddots & \vdots \\ k(\mathbf{x_n}, \mathbf{x_1}) & k(\mathbf{x_n}, \mathbf{x_2}) & \dots & k(\mathbf{x_n}, \mathbf{x_n}) \end{bmatrix} + \sigma^2 \mathbf{I} \right)$$
(A.2)

626 where:

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1. The mean function $m(\mathbf{x})$ describes the *a priori* expected value of the response y at location \mathbf{x} . We use the constant mean $m_c(\mathbf{x}) = \theta_m$ for all \mathbf{x} , where θ_m is a parameter to be estimated from the data.

⁶³⁰ 2. The covariance function $k(\mathbf{x_1}, \mathbf{x_2})$ describes the *a priori* covariance between responses at ⁶³¹ locations $\mathbf{x_1}$ and $\mathbf{x_2}$. This is also a parameterized function. Popular choices of k are the ⁶³² squared exponential $k_{se}(\mathbf{x_1}, \mathbf{x_2}) = A \exp(-|\mathbf{x_1} - \mathbf{x_2}|^2/\rho^2)$, or the Matern

function (Williams and Rasmussen, 2006), which both reflect the common assumption
 that similar predictor values will lead to similar response values. In this work, we use the
 Matern function with separate scaling parameters for latitude and longitude to model

spatial covariance.

 $_{637}$ 3. The noise parameter σ^2 models measurement error.

4. I is the $n \times n$ identity matrix.

⁶³⁹ We write \mathbf{y} , \mathbf{m} and K to denote the responses, means and the covariance matrix of the training ⁶⁴⁰ data, respectively, so that we can write Eq. A.2 as $\mathbf{y} \sim \mathcal{N}(\mathbf{m}, K + \sigma^2 \mathbf{I})$. The choice of mean ⁶⁴¹ and covariance functions reflects prior knowledge and modelling assumptions about the ⁶⁴² regression problem. The function parameters as well as the noise parameter σ are estimated by ⁶⁴³ maximizing the likelihood of the training data. We use GPyTorch (Gardner et al., 2018) to ⁶⁴⁴ efficiently find the maximum likelihood parameter estimates.

After parameter estimation, the GP regression model can be used to predict responses at previously unseen data points. Let S be the locations and responses comprising the training data set. Since the responses at training and test points are assumed to be jointly Gaussian, the conditional distribution of the response at a test point \mathbf{x}^* given the training data is also Gaussian with mean

$$\mu(\mathbf{x}^*|S) = m(\mathbf{x}^*) + \mathbf{k}^*(K + \sigma^2 \mathbf{I})^{-1}(\mathbf{y} - \mathbf{m})$$
(A.3)

where $\mathbf{k}^* = [k(\mathbf{x}^*, \mathbf{x}_1), k(\mathbf{x}^*, \mathbf{x}_2), \dots, k(\mathbf{x}^*, \mathbf{x}_n)]$ is the vector of *a priori* covariances between responses at \mathbf{x}^* and training data points. The posterior variance of y^* is given by

$$\sigma^{2}(\mathbf{x}^{*}|S) = k(\mathbf{x}^{*}, \mathbf{x}^{*}) + \sigma^{2} - \mathbf{k}^{*}(K + \sigma^{2}\mathbf{I})^{-1}\mathbf{k}^{*\top}$$
(A.4)

- see (Williams and Rasmussen, 2006) for a derivation.

⁶⁵³ For a specific response value y^+ , its likelihood of being observed at \mathbf{x}^* is just the Gaussian

₆₅₄ probability density with mean μ and variance σ^2 found by applying Equations A.3 and A.4

$$p(y^* = y^+ | \mathbf{x}^*, S) = \frac{1}{\sqrt{2\pi\sigma^2(\mathbf{x}^*|S)}} \exp\left(\frac{-(y^+ - \mu(\mathbf{x}^*|S))^2}{2\sigma^2(\mathbf{x}^*|S)}\right)$$
(A.5)

For a sample $\bar{y} = (y_O, y_H, y_C, y_S)$ of observed isotope ratio values (meaning δ^{18} O, δ^{2} H, δ^{13} C, δ^{34} S), the Bayes' theorem gives the posterior distribution of possible harvest locations:

$$p(\mathbf{x}|\bar{y}, S) = \frac{p(\mathbf{x}) \prod_{i \in \{O, H, C, S\}} p_i(y_i | \mathbf{x}, S)}{\int_{\mathbf{x} \in A} p(\mathbf{x}) \prod_{i \in \{O, H, C, S\}} p_i(y_i | \mathbf{x}, S) d\mathbf{x}}$$
(A.6)

where the probabilities p_i are computed from the GP models for the respective isotopes using Equation A.5 and A is the study area. The integral in the denominator is computed by averaging the probabilities over the spatial grid.

B Active learning performance



Figure 5: Average predictive variances for $\delta^{18}O$, δ^2H , $\delta^{13}C$ and $\delta^{34}S$ as a function of the number of samples added to the base training data set; blue - active learning strategy; red - random sampling (shaded area denotes values within two standard deviations of the mean across $n_r = 100$ simulations).