# 1 Co-crediting system for carbon and biodiversity

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### 24 Science for Society

Humankind is facing both climate and biodiversity crises. To ameliorate carbon emissions, 25 carbon offsetting and crediting systems have become well established, whereas equivalent 26 programs are nonexistent for biodiversity. Here we propose a scheme that offers tradable 27 credits for combined aboveground and soil carbon and biodiversity, where species richness of 28 various soil organisms constitutes a biodiversity proxy. We argue that multidiversity - based 29 on high-throughput molecular identification of soil animals, fungi, bacteria, protists and 30 plants - offers a cost-effective method that captures much of the terrestrial biodiversity. We 31 anticipate that such a voluntary crediting system increases the quality of carbon projects and 32 33 may contribute to much of global biodiversity funding in a 10-year perspective.

### 34 Summary

Carbon crediting and land offsets for biodiversity protection are implemented to tackle the 35 challenges of increasing greenhouse gas emissions and loss of global biodiversity, but these 36 two mechanisms are not optimal when considered separately. Focusing solely on carbon 37 capture - the primary goal of most carbon-focused offsetting commitments - often results in 38 the establishment of non-native, fast-growing monocultures that negatively affect biodiversity 39 40 and soil-related ecosystem services. Soil contributes a vast proportion of global biodiversity and contains traces of aboveground organisms. Here we introduce a carbon and biodiversity 41 42 co-crediting scheme based on the multi-kingdom molecular analysis and carbon analysis of soil samples and remote sensing for above-ground carbon analysis. Combined, such a co-43 crediting scheme could help halt biodiversity loss by incentivising industry and governments 44 to fully account for biodiversity in carbon sequestration projects, prioritising protection before 45 restoration and promoting socially and environmentally sustainable land stewardship in 46 society's journey towards a 'Net Positive' future. 47

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Keywords: biodiversity crediting, carbon crediting, soil biodiversity, DNA metabarcoding,
molecular identification, ecological sustainability, net positive effect, biodiversity banking,
offsetting, conservation

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### 53 Introduction

The release of greenhouse gases and the continued expansion of agriculture and forestry have 54 collectively resulted in massive losses of native biodiversity worldwide<sup>1-3</sup>. Today, land use 55 changes constitute the primary threat to species worldwide, with climate change driving key 56 additional risks such as increased physiological stress and loss of suitable habitat or mutualistic 57 partners<sup>4,5</sup>. In an effort to mitigate climate change, there has been a tremendous interest by 58 industry, governments and other parts of society to rapidly develop schemes to sequester 59 carbon, either through technological inventions for locking carbon into the substrate or through 60 nature-based solutions, such as the mass planting of trees<sup>6</sup>. The problem is that these carbon 61 62 capture solutions are often deleterious to biodiversity, for example by promoting rapidly growing tree monocultures instead of natural vegetation<sup>7,8</sup>. Furthermore, short-rotation 63 bioenergy plantations only poorly mitigate climate change relative to fossil resources and fail 64 to support biodiversity<sup>9</sup>. 65

Biodiversity is crucial for ecosystem functioning and for increasing resistance to perturbations, 66 particularly in stressful and increasingly unpredictable environmental conditions<sup>10</sup>. Positive 67 biodiversity-ecosystem functioning effects are inherent to all domains of life - from 68 microorganisms such as bacteria and fungi to macroorganisms<sup>11</sup>. Much of the biodiversity is 69 built up of rare species<sup>12</sup>, which can have a disproportionate effect on ecosystems by 70 71 performing unique ecosystem services, such as generating micro-climates, controlling diseases, and promoting tight nutrient cycling<sup>13</sup>. Recent estimates indicate that only around one quarter 72 of the required funding sources are invested into biodiversity globally<sup>14</sup>. 73

74 Programs for offsetting carbon and environmental protection have a short history. In 1989, the US-based AES company invested 2 million USD to offset carbon emissions by planting and 75 conserving rainforest in Guatemala. In the early 1990s, the first carbon crediting initiatives 76 77 were developed to support land owners practising sustainable management of agroecosystems, grasslands, and forests that promoted carbon sequestration in aboveground biomass and 78 topsoil<sup>15</sup>. Similarly, conservation offsetting programs have been pursued to counterbalance 79 agriculture-related or industrial land degradation<sup>16</sup>. For example, a payment-for-ecosystem-80 services program has been implemented in Costa Rica since 1997<sup>17</sup>, and the Chinese Green for 81 Grain program has been developed to prevent erosion since 1999<sup>18</sup>. In the 2000s, offsetting 82 schemes for habitats of endangered species were developed in California<sup>19</sup>. While offsetting is 83 related to compensating harm elsewhere and regarded as the last resort for conservation<sup>16</sup>, 84

biodiversity credits are designed solely to promote conservation<sup>20</sup>. Although carbon crediting 85 schemes increasingly account for biodiversity effects, no large-scale, operating biodiversity 86 crediting schemes exist (Table 1). In 2022, principles of tradable biodiversity crediting schemes 87 were developed for terrestrial and aquatic biota<sup>21</sup>. In spite of the current biodiversity crisis, the 88 relatively slow evolution of biodiversity credits is likely due to the multitude of alternative 89 biodiversity metrics<sup>22,23</sup>, the lack of consensus around biodiversity baselines<sup>24</sup>, and difficulties 90 in accurately surveying and quantifying biodiversity<sup>25</sup> compared to estimating carbon 91 sequestration potential. Here we explore what we call the biodiversity and carbon co-crediting 92 93 concept, and discuss how its implementation could transform the conservation, restoration, and off-setting landscape to help societies achieve a greener road to Net Zero – a state of balance 94 between anthropogenic emissions and anthropogenic removals. 95

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# 97 The biodiversity-carbon co-crediting concept

We propose that biodiversity should be explicitly incorporated into carbon marketing schemes to increase environmental and social sustainability in land use. All carbon marketing schemes should adopt the no net biodiversity loss criterion. For this integration to become possible, biodiversity needs to be robustly and efficiently measured. This is required to allow for assessments of how various forms of intervention (such as reforestation, habitat restoration, agroforestry, etc.) may affect biodiversity as compared to previous land use or management practice.

However, biodiversity is currently difficult and costly to comprehensively estimate across 105 multiple ecosystem components such as soil, wood, air, and tree canopy. Recent advances in 106 high-throughput DNA sequencing analysis of environmental samples (eDNA analysis) offer a 107 promising tool for rapid, cost-effective evaluation of biodiversity, a technique now thoroughly 108 validated for water, soil and bulk animal samples<sup>26</sup>. If we accept that eDNA-based biodiversity 109 assessments are today feasible and well validated<sup>26,27</sup>, encapsulating a much larger proportion 110 of the planet's biodiversity than previously possible through manual assessments, the next 111 question of implementation is: how can metrics of biodiversity be aligned with carbon markets? 112

Past research has shown that market-based incentives can be important mechanisms for driving conservation policy<sup>17</sup>. Integrating biodiversity credits into existing or novel carbon crediting mechanisms could encourage landowners to proactively manage land. If biodiversity and

carbon are rewarded in the same scheme, land managers are more likely to optimise different 116 types of benefits for the particular land cover in the region<sup>28,29</sup>. For instance, today only net 117 gains in carbon can be considered under Net Zero schemes, which directly incentivises cutting 118 and reforestation, instead of protecting a forest from felling in the first place (Fig. 1). Restoring 119 rather than protecting is always a significantly worse and more expensive solution in both 120 biodiversity and carbon storage terms, given the manifold advantages of an old-growth forest, 121 as compared to any form of tree planting or natural regeneration. Today many carbon crediting 122 schemes are blindly focused on 'cheap carbon' - supporting interventions that in fact lead to 123 124 the lowest levels of both carbon storage and biodiversity in the medium term (Fig. 1). Therefore, all carbon offsets and credits affecting land use and land cover should have baseline 125 estimates for biodiversity effects. 126

Directly coupling carbon and biodiversity credits would prevent other perverse incentives, such as excess fertilisation and planting monocultures, that strongly favour carbon over biodiversity<sup>7</sup>. A framework that directly integrates carbon and biodiversity credits needs to be practical and well-tested, setting clear rules that are easy to follow<sup>30</sup>. From a global perspective, these rules would benefit from general biodiversity policies<sup>31</sup>.

As most carbon crediting schemes account only for aboveground biomass production, carbon storage in soil remains usually overlooked. In some regions, topsoil carbon stocks alone are comparable in size to aboveground carbon, but vary greatly across biomes and land cover types<sup>32</sup>, such as lower biomass accumulation in nutrient-poor rainforests. Soil carbon stores also tend to increase with sustainable land management including organic farming, moderate grazing pressure, selective timber harvesting and establishment of mixed plantations<sup>33-35</sup>.

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### **139 Biodiversity proxies**

Because soil contributes to a vast proportion of biodiversity (most terrestrial species have at least part of their life cycle underground), productivity and functioning of terrestrial ecosystems<sup>11,36</sup>, soil biodiversity has a potential to constitute a proxy for biodiversity and ecosystem health in most forest, grassland, and agricultural habitats<sup>37</sup>. Soil biodiversity analyses follow the HAND(Y) principle - they are High-tech, Accurate, Novel, Detailed and Yielding. Comprehensive assessment of soil biodiversity, including both macro- and microorganisms, can be carried out using an internationally standardised soil sampling scheme (e.g., SoilBON) coupled with cross-kingdom global analyses of soil biota<sup>38-40</sup>. Such analyses
can also help us develop a better global picture of cryptic biodiversity, such as where hotspots
of micro-organismal diversity are located<sup>37,38</sup>. Using soil biodiversity as a metric to evaluate
the impact of reforestation and habitat restoration increases the ease of measuring, comparing
and monitoring biodiversity across diverse landscapes and over time<sup>41</sup>.

In biology, genes, individuals and species are the main ecological units. Since genes and individuals are more difficult to measure and species are easy to grasp, the species-level metrics are of greatest public and conservational interest, with species richness, effective number of species and multidiversity as the best biodiversity proxies<sup>22,42</sup>. Functional and phylogenetic diversity offer additional insights into ecosystem functioning<sup>43</sup>, although ascertaining impacts on ecosystem functions requires further information about the ecology of individual species, which is lacking for most soil organisms.

Species also differ in abundance, which in turn affects their contributions to ecosystem 159 functioning. The redundancy and additionality of rare species can be difficult to assess due to 160 the low statistical power in observational studies. Rare species are often habitat specialists or 161 sensitive to anthropogenic impact<sup>13</sup>. Therefore, rare, especially threatened species, can be 162 considered more important from the conservation and crediting perspectives<sup>21,44</sup>. However, the 163 conservation status for the vast majority of species has never been assessed<sup>3</sup>. One pragmatic 164 approach is to weigh the importance of all species equally, until we can rank all species based 165 on their conservation value or distinguish them by function and habitat, for example identifying 166 keystone forest species in re/afforestation projects<sup>45</sup>. The rapidly growing traits and occurrence 167 databases (e.g., GBIF, www.gbif.org; TRY Plant Trait Database, www.try-db.org; 168 FungalTraits<sup>46</sup>) may facilitate the identification of target species for ecosystem restoration in 169 the near future. 170

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# 172 The crediting process

To maximise biodiversity and carbon benefits, a robust crediting system should deploy evidence-based criteria for the selection of areas for potential conservation and restoration using a combination of field surveys, remote-sensing, soil maps and machine learning algorithms<sup>47</sup>. Rewarding companies and communities for positive change requires a robust monitoring, evaluation and reporting framework.

In the process of land evaluation, representative plots are randomly surveyed by accredited 178 institutions using standardised procedures that are verified independently (Fig. 2)<sup>48</sup>. The 179 representative plots for monitoring should be of sufficient size and number, and located 180 randomly in the survey area, avoiding edge effects. The sampling standards may follow well-181 protocols for large-scale sampling schemes, such SoilBON 182 elaborated as (https://www.globalsoilbiodiversity.org/soilbon), but these may differ by project, considering 183 representativeness of sampling, type of habitat, target organisms, etc. Sampling should capture 184 a significant proportion of biodiversity, and optimal sampling intensity should be determined 185 186 based on pilot studies or information from scientific literature. It is important to consider time of sampling (in growing season) and storage of samples to avoid loss of diversity through DNA 187 degradation<sup>26</sup> to enhance comparability across time. Molecular analysis of as many taxonomic 188 groups as possible – plants, animals, fungi, micro-eukaryotes and prokaryotes – offers the most 189 accurate views on overall biodiversity, reducing taxon-specific biases<sup>42</sup>. Additional 190 standardised semi-automated technologies may be used for recording images and/or sounds of 191 mammals and birds, followed by identification using machine learning techniques (e.g., 192 https://www2.helsinki.fi/en/projects/lifeplan). 193

The carbon-biodiversity co-benefits can be calculated based on temporal changes relative to 194 control plots and near-natural reference plots (endpoints) to account for climatic effects and 195 196 batch effects (i.e., temporal sampling effects). The control plots should occur in comparable vegetation in nearby lands not affected by the interventions being evaluated, and reflect a 197 situation of average management intensity or the "business as usual" scenario - how carbon 198 and biodiversity would have changed without the intervention applied (Fig. 3). It is important 199 to perform temporal sampling in the same representative and control plots to minimise 200 analytical error and provide feedback about the best and worst performing areas (while keeping 201 plot localities undisclosed). Along with remote sensing-based measurements of aboveground 202 203 carbon, soil carbon can be additionally estimated using deep cores to include subsoil. Biodiversity assessments are best performed from topsoil that contains high biomass and 204 205 highest biodiversity of most soil organisms. Biodiversity monitoring can be performed in five to ten year intervals, which is a typical time frame in carbon crediting businesses<sup>48</sup>. Biodiversity 206 and carbon crediting mechanisms should secure longevity, i.e. potential to prolong contracts 207 208 for decades or centuries - instead of the mere 30 years currently used under most carbon crediting schemes. Upgrading initial contracts to higher-value contracts should also be 209 210 considered, for example for young forests that become more highly valued when they become

old and support more biodiversity and soil carbon. To maintain such a long-term monitoring process, project managers should take care of proper storage of materials and data. Carefully preserved DNA samples can be reused decades later when better DNA sequencing methods emerge, or when additional taxonomic groups or markers are added for more comprehensive analysis of biodiversity. Currently, molecular analysis of soil samples from 100 plots (corresponding to a medium-size project) samples cost 3000 euros upwards, while sequencing costs per unit data continue to decline.

We propose that the metric for biodiversity crediting should include positive change in a unit 218 of time (e.g., 5 years) over a certain area (e.g., 1 ha). So, the tokens are related to both time 219 220 interval and area, which may differ across projects but not magnitude of change. The Wallacea Trust<sup>21</sup> suggested that 1% uplift of biodiversity relative to reference sites represents a suitable 221 metric. However, we find that reliable detection of 1% difference requires prohibitively large 222 sample sizes, e.g. several hundred composite soil samples in our case<sup>49</sup> and perhaps more for 223 the stochastic macroorganism inventories. Furthermore, combined animal and plant 224 inventories, metabarcoding surveys of specifically captured invertebrates and measures of 225 several ecosystem services, as advocated by The Wallacea Trust<sup>21</sup>, render biodiversity 226 estimates more time consuming and expensive compared with simple soil metabarcoding 227 228 surveys. Crediting for carbon sequestration should follow widely accepted protocols and units; 229 inclusion of soil carbon should offer better revenues compared with standard aboveground carbon projects. 230

231 Crediting institutions should release credits after data analysis and verification rather than based on future pledges, although ex-ante payments should be considered to reduce poverty in 232 developing countries<sup>20</sup>. Occasional verification must be performed by independent assessors to 233 secure transparency and validity of approaches and measurements. The weighting of carbon 234 235 and biodiversity components should remain flexible because of the potentially changing stakeholder expectations over time. Complementarity of the co-crediting components is 236 237 crucial; for example, a project that maintains or restores a natural savanna may, for instance, not capture as much carbon as a new tree plantation, but help preserve highly threatened 238 239 biodiversity.

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# 241 Trading parties

While carbon credits are issued by international and national governmental organisations, 242 biodiversity credits can be issued by parties that own or lease land and are interested in long-243 term conservation and income, for example governments, private and corporate landowners 244 and indigenous communities. Hence, for co-crediting, private issuers should collaborate with 245 local governments or buy carbon credits to sell co-credits. Buyers of these credits include 246 conservation-aware companies and persons such as the tourism sector, philanthropists, as well 247 as private resellers, i.e. parties acting in the carbon and developing biodiversity markets. To 248 avoid offsetting and greenwash and hence bad reputation, companies with harmful actions on 249 250 climate and nature could be excluded from this trade by project rules. Trading can be performed via tokens or cryptocurrency in banks as implemented for carbon credits and offsets. 251

Since much of the conserved and restorable land is available for biodiversity-carbon cocrediting, it will be of great importance to share benefits with local communities, including indigenous people. This should include both a part of monetary revenues and involvement through performing sustainable management practices, guarding of project areas and avoiding certain unsustainable practices, such as slash-and-burn agriculture and landscape burning to ease hunting.

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### 259 Conclusions

The urgency for our societies to reach Net Zero in the shortest possible amount of time is 260 triggering vast investments and initiatives around the world. However, trying to combat one 261 major challenge - climate change - while making another one worse (biodiversity loss), would 262 represent a huge opportunity loss. The inclusion, valuation and validation of biodiversity and 263 other functional and ecosystem services-related co-benefits within carbon crediting and 264 offsetting schemes will help reduce global biodiversity loss by incentivising carbon crediting 265 beneficiaries to account for biodiversity in their carbon sequestration projects<sup>28</sup>. Given the 266 increasing pressure on biodiversity, it is likely that the relative valuation in biodiversity 267 increases compared with carbon, especially when developed countries reach their emissions 268 reduction goals. Such monetary biodiversity benefits will promote environmentally sustainable 269 stewardship of land globally and contribute much of the global financing for conservation<sup>14</sup>. 270

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### 272 **Resource availability**

#### 273 *Lead contact*

274 Technical questions and requests of images should be addressed to Leho Tedersoo275 (leho.tedersoo@ut.ee), the lead contact.

276 Materials availability

- 277 This article produced no unique materials.
- 278 Data and code availability
- 279 This study did not generate data or code

280

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# 289 Author Contributions

- 290 LT and JS generated the initial concept; ASM, MK, KE, RR, KH, EB, RP, BG, FV, TK and
- AA further developed the concept from various perspectives; LT, TK and AA wrote the
- 292 manuscript, with input from all authors.

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# 294 Declaration of Interests

295 The authors declare no conflict of interest.

### 297 References

- 1. Urban, M.C. (2015). Accelerating extinction risk from climate change. Science 348, 571574.
- Bradshaw, C.J., Ehrlich, P.R., Beattie, A., Ceballos, G., Crist, E., Diamond, J., Dirzo, R.,
  Ehrlich, A.H., Harte, J., Harte, M.E., Pyke, G., and Blumstein, D.T. (2021).
- 302 Underestimating the challenges of avoiding a ghastly future. Front. Conserv. Sci. 2, 9.
- 303 3. IUCN. (2021). The IUCN Red List of Threatened Species. Version 2021-2. International
  304 Union for Conservation of Nature (IUCN), Gland, Switzerland.
- 4. Diaz, S., Settele, J., Brondízio, E.S., Ngo, H.T., Agard, J., Arneth, A., and Garibaldi, L.A.
- 306 (2019). Pervasive human-driven decline of life on Earth points to the need for
- transformative change. Science 366, eaax3100.
- 5. IPBES. (2019). Global assessment report on biodiversity and ecosystem services.
  Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services
  (IPBES) Secretariat, Bonn, Germany.
- 6. Lal, R. (2008). Carbon sequestration. Phil. Trans. R. Soc. B 363, 815-830.

312 7. Lindenmayer, D.B., Hulvey, K.B., Hobbs, R.J., Colyvan, M., Felton, A., Possingham, H.,

- and Gibbons, P. (2012). Avoiding bio-perversity from carbon sequestration solutions.
- 314 Conserv. Lett. 5, 28-36.
- 8. Hua, F., Bruijnzeel, L.A., Meli, P., Martin, P.A., Zhang, J., Nakagawa, S., and Brancalion,
  P.H. (2022). The biodiversity and ecosystem service contributions and trade-offs of forest
  restoration approaches. Science, 2022, eabl4649.
- 9. Camia, A., Giuntoli, J., Jonsson, R., Robert, N., Cazzaniga, N.E., Jasinevičius, G., and
- 319 Mubareka, S. (2021). The use of woody biomass for energy purposes in the EU.
- Publications Office of the European Union, Luxembourg. doi:10.2760/831621
- 10. Hong, P., Schmid, B., De Laender, F., Eisenhauer, N., Zhang, X., Chen, H., and Reich,
- P.B. (2022). Biodiversity promotes ecosystem functioning despite environmental change.
- 323 Ecol. Lett. 25, 555–569.
- 11. Yang, G., Wagg, C., Veresoglou, S.D., Hempel, S., and Rillig, M.C. (2018). How soil
  biota drive ecosystem stability. Trends Plant Sci. 23, 1057-1067.
- 12. Enquist, B.J., Feng, X., Boyle, B., Maitner, B., Newman, E.A., Jørgensen, P.M.,
- 327 Roehrdanz, P.R., Thiers, B.M., Burger, J.R., Corlett, R.T., and Couvreur, T.L., 2019. The

- commonness of rarity: Global and future distribution of rarity across land plants. Sci. Adv.
  5, p.eaaz0414.
- 13. Dee, L.E., Cowles, J., Isbell, F., Pau, S., Gaines, S.D., and Reich, P.B. (2019). When do
  ecosystem services depend on rare species? Trends Ecol. Evol. 34, 746-758.
- 14. Deutz, A., Heal, G.M., Niu, R., Swanson, E., Townshend, T., Zhu, L., Delmar, A.,
- Meghji, A., Sethi, S.A., Tobinde la Puente, J. (2020). Financing Nature: Closing the
  Global Biodiversity Financing Gap (Paulson Institute).
- 15. Trexler, M.C. (1991). Minding the Carbon Store: Weighing U.S. Forestry Strategies to
  Slow Global Warming. World Resources Institute, Washington, DC.
- 16. BBOP. (2012). Biodiversity Offset Design Handbook-Updated (Business and
  Biodiversity Offsets Programme, BBOP).
- 17. Pagiola, S. (2008). Payments for environmental services in Costa Rica. Ecol. Econ. 65,
  712-724.
- 18. Xu, Z., Xu, J., Deng, X., Huang, J., Uchida, E., and Rozelle, S. (2006). Grain for green
  versus grain: conflict between food security and conservation set-aside in China. World
  Devel. 34, 130-148.
- 344 19. Grimm, M. (2020). Conserving biodiversity through offsets? Findings from an empirical
  345 study on conservation banking. J. Nature Conserv. 57, 125871.
- 20. Porras, I. and Steele, P. (2020). Making the market work for nature: how biocredits can
- 347 protect biodiversity and reduce poverty. International Institute for Environment and348 Development, London, UK.
- 349 https://www.iied.org/sites/default/files/pdfs/migrate/16664IIED.pdf
- 21. The Wallacea Trust. (2022). Methodology for awarding biodiversity credits v. 1.5.
- 351 https://wallaceatrust.org/wp-content/uploads/2022/08/Biodiversity-credit-methodology352 1.5.pdf
- 22. Chao, A., Chiu, C.H., and Jost, L. (2014). Unifying species diversity, phylogenetic
- diversity, functional diversity, and related similarity and differentiation measures through
  Hill numbers. Annu. Rev. Ecol. Evol. Syst. 45, 297-324.
- 23. Lammerant, J., Starkey, M., de Horde, A., Bor, A.M., Driesen, K., and Vanderheyden, G.
- 357 (2021). Assessment Of Biodiversity Measurement Approaches For Business And
- 358 Financial Institutions (Update Report 3). EU Commission.
- 359 https://knowledge4policy.ec.europa.
- eu/sites/default/files/EU%20B%40B%20Platform%20Update%20Report%203 FINAL 1
- 361 March2021.pdf.

- 362 24. Mehrabi, Z., and Naidoo, R. (2022). Shifting baselines and biodiversity success stories.
  363 Nature 601, E17-E18.
- 25. Bull, J.W., Taylor, I., Biggs, E., Grub, H.M., Yearley, T., Waters, H., and Milner-
- Gulland, E.J. (2022). Analysis: the biodiversity footprint of the University of Oxford.
  Nature 604, 420-424.
- 26. Taberlet, P., Bonin, A., Zinger, L., and Coissac, E. (2018). Environmental DNA: For
  Biodiversity Research and Monitoring. Oxford University Press, London, UK.
- 369 27. Ji, Y., Ashton, L., Pedley, S. M., Edwards, D. P., Tang, Y., Nakamura, A., and Yu, D. W.
- 370 (2013). Reliable, verifiable and efficient monitoring of biodiversity via metabarcoding.
  371 Ecol. Lett. 16, 1245-1257.
- 28. Thomas, C.D., Anderson, B.J., Moilanen, A., Eigenbrod, F., Heinemeyer, A., and Gaston,
  K.J. (2013). Reconciling biodiversity and carbon conservation. Ecol. Lett. 16, 39-47.
- 29. Bryan, B.A., Runting, R.K., Capon, T., Perring, M.P., Cunningham, S.C., Kragt, M.E.,
- and Wilson, K.A. (2016). Designer policy for carbon and biodiversity co-benefits under
  global change. Nature Clim. Change 6, 301-305.
- 30. Di Sacco, A., Hardwick, K.A., Blakesley, D., Brancalion, P.H., Breman, E., Cecilio
  Rebola, L., and Shaw, K. (2021). Ten golden rules for reforestation to optimize carbon
  sequestration, biodiversity recovery and livelihood benefits. Glob. Change Biol. 27, 13281348.
- 31. Otero, I., Farrell, K.N., Pueyo, S., Kallis, G., Kehoe, L., Haberl, H., Plutzar, C., Hobson,
  P., García-Márquez, J., Rodríguez-Labajos, B., and Martin, J.L. (2020). Biodiversity
  policy beyond economic growth. Conservation Letters 13, e12713.
- 32. Scharlemann, J.P., Tanner, E.V., Hiederer, R., and Kapos, V. (2014). Global soil carbon:
  understanding and managing the largest terrestrial carbon pool. Carbon Manage. 5, 81-91.
- 386 33. Schroth, G., D'Angelo, S.A., Teixeira, W.G., Haag, D., and Lieberei, R. (2002).
- 387 Conversion of secondary forest into agroforestry and monoculture plantations in
- Amazonia: consequences for biomass, litter and soil carbon stocks after 7 years. For. Ecol.
  Manage. 163, 131-150.
- 34. Chaudhary, S., Dheri, G.S., and Brar, B.S. (2017). Long-term effects of NPK fertilizers
  and organic manures on carbon stabilization and management index under rice-wheat
  cropping system. Soil Till. Res. 166, 59-66.
- 552 cropping system. Son Tin. Res. 100, 57 00.
- 393 35. Jackson, R.B., Lajtha, K., Crow, S.E., Hugelius, G., Kramer, M.G., and Piñeiro, G.
- 394 (2017). The ecology of soil carbon: pools, vulnerabilities, and biotic and abiotic controls.
- 395 Annu. Rev. Ecol. Evol. Syst. 48, 419-445.

- 396 36. Wardle, D.A. (2002). Communities and Ecosystems: Linking the Aboveground and
  Belowground Components. Princeton University Press, New York, NY.
- 37. Orgiazzi, A., Bardgett, R.D., and Barrios, E. (2016). Global Soil Biodiversity Atlas
  (European Commission).
- 38. Bahram, M., Espenberg, M., Pärn, J., Lehtovirta-Morley, L., Anslan, S., Kasak, K., and
  Mander, Ü. (2022). Structure and function of the soil microbiome underlying N<sub>2</sub>O
  emissions from global wetlands. Nature Commun. 13, 1430.
- 403 39. Guerra, C.A., Berdugo, M., Eldridge, D.J., Eisenhauer, N., Singh, B.K., Cui, H., Abades,
- S., Alfaro, F.D., Bamigboye, A.R., Bastida, F., Blanco-Pastor, J.L. et al. (2022). Global
  hotspots for soil nature conservation. Nature 610, 693-698.
- 406 40. Ritter, C.D., Häggqvist, S., Karlsson, D., Sääksjärvi, I.E., Muasya, A.M., Nilsson, R.H.,
- and Antonelli, A (2019). Biodiversity assessments in the 21st century: the potential of
- 408 insect traps to complement environmental samples for estimating eukaryotic and
- 409 prokaryotic diversity using high-throughput DNA metabarcoding. Genome 62, 147-159.
- 41. Jiao, S., Chen, W., Wang, J., Du, N., Li, Q., and Wei, G. (2018). Soil microbiomes with
  distinct assemblies through vertical soil profiles drive the cycling of multiple nutrients in
  reforested ecosystems. Microbiome 6, 146.
- 413 42. Allan, E., Bossdorf, O., Dormann, C.F., Prati, D., Gossner, M.M., Tscharntke, T.,
- Blüthgen, N., Bellach, M., Birkhofer, K., Boch, S., and Böhm, S. (2014). Interannual
- variation in land-use intensity enhances grassland multidiversity. Proc. Natl. Acad. Sci.
  111, 308-313.
- 417 43. Cadotte, M.W., Carscadden, K., and Mirotchnick, N. (2011). Beyond species: functional
  418 diversity and the maintenance of ecological processes and services. J. Appl. Ecol. 48,
  419 1079-1087.
- 420 44. Brooks, T.M., Mittermeier, R.A., Da Fonseca, G.A., Gerlach, J., Hoffmann, M.,
- 421 Lamoreux, J.F., and Rodrigues, A.S. (2006). Global biodiversity conservation priorities.
  422 Science 313, 58-61.
- 423 45. Kõljalg, U., Nilsson, R. H., Jansson, A. T., Zirk, A., and Abarenkov, K. (2022). A price
  424 tag on species. ARPHA Preprints 3, e86743.
- 425 46. Põlme, S., Abarenkov, K., Nilsson, R.H., Lindahl, B.D., Clemmensen, K.E., Kauserud,
- 426 H., and Tedersoo, L. (2020). FungalTraits: a user-friendly traits database of fungi and
- 427 fungus-like stramenopiles. Fung. Divers. 105, 1–16.
- 428 47. Silvestro, D., Goria, S., Sterner, T., and Antonelli, A. (2022). Improving biodiversity
- 429 protection through artificial intelligence. Nature Sustain. 5, 415-424.

- 430 48. Michaelowa, A., Shishlov, I., Hoch, S., Bofill, P., and Espelage, A. (2019). Overview and
- 431 comparison of existing carbon crediting schemes. Nordic Environment Finance432 Corporation, Helsinki, Finland.
- 433 49. Tedersoo, L., Anslan, S., Bahram, M., Drenkhan, R., Pritsch, K., Buegger, F., Padari, A.,
- Hagh-Doust, N., Mikryukov, V., Kõljalg, U., and Abarenkov, K. (2020). Regional-scale
- 435 in-depth analysis of soil fungal diversity reveals strong pH and plant species effects in
- 436 Northern Europe. Front. Microbiol. 11, 1953.
- 437 50. Pitman, N. (2011). Social and Biodiversity Impact Assessment Manual for REDD+
- 438 Projects: Part 3 Biodiversity Impact Assessment Toolbox. Forest Trends, Climate,
  439 Community. Washington, DC.
- 440 51. James, J., and Harrison, R. (2016). The effect of harvest on forest soil carbon: A meta441 analysis. Forests 7, 308.
- 52. Lewis, S.L., Wheeler, C.E., Mitchard, E.T., and Koch, A. (2019). Regenerate natural
  forests to store carbon. Nature 568, 25-28.
- 444 53. Andres, S.E., Standish, R.J., Lieurance, P.E., Mills, C.H., Harper, R.J., Butler, D.W.,
- Adams, V.M., Lehmann, C., Tetu, S.G., Cuneo, P., and Offord, C.A. (2022). Defining
  biodiverse reforestation: Why it matters for climate change mitigation and biodiversity.
  Plants, People, Planet, in press.
- 448
- 449





Figure 1. Schematic representation of A) relative above- and belowground carbon and
biodiversity net benefits from protection and aff/reforestation of various ecosystem types and
B) win-win and lose-lose situations of these from the biodiversity and carbon perspectives.
Based on data and interpretations from references 9, 30, 38 and 49-53.

457





459 Figure 2. Conceptual scheme of carbon-biodiversity co-crediting.



- 461 Figure 3. Analytical workflow of above- and belowground carbon and soil eDNA analysis.
- 462 Table 1. Comparison of the proposed carbon-biodiversity co-crediting scheme to other
- 463 biodiversity or carbon projects.

Certification	Carbon and	Methodology for	Biodiversity	Comments	Reference
scheme	biodiversity	biodiversity	measurement	on	
	assessment		units	biodiversity	
Proposed	soil	soil DNA	multidiversity	established	This study
carbon-	biodiversity,	metabarcoding	increase	sampling	
biodiversity	aboveground	(prokaryotes,	(relative to	and	
co-credits	and soil	microeukaryotes,	reference) /	analytical	
	carbon	plants and	ha over 5	protocols	
		animals)	years;		
			carbon: 1 t		
			CO <sub>2</sub> -eq;		
The	biodiversity	plant releves,	1% net	no available	https://wallaceatrust.org/wp-
Wallacea	only (incl.	camera/audio	biodiversity	protocols,	content/uploads/2022/08/Biodi
Trust	ecosystem	recordings,	increase or	expensive	versity-credit-methodology-
biodiversity	services)	remote sensing,	avoided loss /	monitoring,	1.5.pdf
credits		ecosystem	ha (relative to	microbiome	
		services,	reference)	not	
		invertebrate	over 5 years	considered	
		metabarcoding			
WWF	mammals	sightings of	change in	local	https://wildlifecredits.com/
Namibia	only	indicator species	wildlife	communities	
Wildlife			presence in	gain	
"credits"			habitat	monetary	
			corridors	benefits, no	
				tradeable	
				credit or	
				reference	
Terrasos	protected	NA	$10 \text{ m}^2$ for $30$	offsetting	https://climatetrade.com/climat
biodiversity	habitat only		years	for habitat,	etrade-and-terrasos-jointly-
"credits"				no	promote-voluntary-
				monitoring	biodiversity-credits-to-support-
					biodiversity-conservation/
South Pole	aboveground	NA; carbon:	1.5 m <sup>2</sup>	offsetting	https://www.southpole.com/su
EcoAustralia	carbon (Gold	Gold Standard	protected	for habitat,	stainability-
carbon+	Standard)		Australian		solutions/ecoaustralia

biodiversity	and		vegetation;	no	
"credits"	Australian		carbon: 1 t	monitoring	
	native habitat		CO <sub>2</sub> -eq;		
REDD+	aboveground	carbon projects	NA	carbon	https://www.redd.plus/
carbon	carbon only;	must follow		project	
projects	biodiversity	SBIA <sup>b</sup> standards		revenues	
	not	for biodiversity		used for	
	developed <sup>1</sup>			rainforest	
				protection	
Verra	aboveground	new carbon	NA	ca. 75%	https://verra.org/programs/ccbs
carbon	carbon only;	projects must		carbon	/
projects	biodiversity	follow CCB <sup>c</sup>		projects	
	under	standards for		follow CCB	
	development <sup>a</sup>	biodiversity		standards	
				for	
				biodiversity	
Gold	aboveground	NA	NA	safeguarding	https://www.goldstandard.org/
Standard	carbon only;			principles	our-story/sector-land-use-
carbon	biodiversity			for	activities-nature-based-
projects	not			ecosystem	solutions
	developed <sup>a</sup>			services	
CORSIA	aboveground	NA	NA	biodiversity-	https://www.icao.int/environm
carbon	carbon only;			rich areas	ental-
projects	biodiversity			not included	protection/CORSIA/Pages/def
	not			in carbon	ault.aspx
	developed			projects	
Plan Vivo	aboveground	carbon projects	NA	biodiversity-	https://www.planvivo.org/
carbon	carbon only;	must assess		friendly	
projects	biodiversity	biodiversity		carbon	
	not	risks and do no		projects	
	developed <sup>a</sup>	harm			
Climate	aboveground	carbon projects	NA	ecologically	https://www.climateactionreser
Action	carbon only;	must follow		sustainable	ve.org/
Reserve	biodiversity	sustainable		carbon	
carbon	not	forestry practices		projects	
projects	developed				

464 <sup>a</sup>Confirmed by personal communication; <sup>b</sup>SBIA, Social and Biodiversity Impact Assessment,

465 https://s3.amazonaws.com/CCBA/SBIA\_Manual/SBIA\_Part\_3.pdf; <sup>c</sup>CCB, Climate, Community and

466 Biodiversity Alliance standards, https://verra.org/wp-content/uploads/2017/12/CCB-Standards-

467 v3.1\_ENG.pdf