

1 **High cryptic diversity in a caddisfly that co-occurs in lakes and streams**

2

3 Christine A. Parisek<sup>1\*</sup>, Michael P. Marchetti<sup>2</sup>, Matthew R. Cover<sup>1</sup>

4

5 <sup>1</sup> *Department of Biological Sciences, California State University Stanislaus, Turlock, CA 95382*

6 <sup>2</sup> *Department of Biology, Saint Mary's College of California, Moraga, CA 94575*

7 \* *Corresponding author: [caparisek@ucdavis.edu](mailto:caparisek@ucdavis.edu); Present Address: Department of Wildlife,*

8 *Fish, & Conservation Biology, University of California Davis, Davis, CA 95616*

9

10 **ORCID**

11 Christine Parisek – <https://orcid.org/0000-0002-7648-879X>

12 Michael Marchetti – <https://orcid.org/0000-0001-8574-6802>

13 Matthew Cover – <https://orcid.org/0000-0003-3315-1027>

14

15 **Author Contribution Statement:** CAP led study design, field and lab work, data analyses, and  
16 wrote the manuscript. MRC helped design the study, provided methods guidance, helped  
17 interpret data, and assisted in manuscript revisions. MPM provided feedback on field and lab  
18 procedures and helped interpret data. All authors revised the article critically for important  
19 intellectual content and approved the final version to be published.

20

21 **Competing Interests Statement:** The authors have no competing interests to disclose.

22

23 **Data Accessibility Statement:** Specimen COI sequences will be publicly accessible in the  
24 Barcode of Life Data Systems by the time of publication.

25

26

27 **This PDF file includes:** Main Text; Figures 1-11; Tables 1-2

**ABSTRACT**

28  
29  
30 Lake and stream fauna are frequently studied, yet surprisingly little is known about ecological  
31 and evolutionary dynamics of species that inhabit both lentic and lotic habitats. There are few  
32 examples of species co-occurring in the different habitat flow types, which raises questions on  
33 how this may impact their ability to adapt to changing climatic conditions. The aquatic insect  
34 *Limnephilus externus* (Trichoptera: Limnephilidae) is widely distributed in lakes of the Nearctic  
35 and Palearctic regions; in our study area of the northern Sierra Nevada mountains (California,  
36 USA), larval stages of this species co-occur in connected lakes and streams. We examined larval  
37 body and case morphology, interspecies phoretic associations, and the mitochondrial DNA  
38 cytochrome *c* oxidase I (COI) gene among lake and stream populations of *L. externus*. Further,  
39 we begin to explore potential morphologic differences in distinct *L. externus* haplogroups. We  
40 observed differences between lake and stream populations in abundance, phenology, some  
41 aspects of body and case morphology, and abdominal mite presence, indicating that lakes and  
42 streams may yield distinct ecological phenotypes for the species. We also observed distinct  
43 regional differences in caddisfly body condition and sturdiness of case construction, as well as  
44 distinct communities of micro-invertebrates associated with the caddisfly and cases. Lake-stream  
45 *L. externus* did not show genetic divergence; however, three potentially distinct haplogroups  
46 were present across the research sites, as well as in sequences from North America and Canada  
47 which were imported from BOLDSYSTEMS. *L. externus* appears to exhibit wide geographic  
48 range and low geographic sequence structure which could account for the species' large variation  
49 in phenology and morphology at the lake-stream level. As the Sierra Nevada faces warming  
50 temperatures, reduced snowpack, and flow cessation, sensitive high elevation species will face  
51 potentially detrimental consequences. Aquatic insect life history and phylogenetic structure  
52 provides valuable insight into the ecological and evolutionary dynamics that influence the  
53 adaptability of aquatic fauna to climatic change.

54

**Keywords**

56 Lentic-Lotic, Aquatic insect, DNA barcoding, phenology, morphology, phoresy

## INTRODUCTION

57  
58  
59  
60  
61  
62  
63  
64  
65  
66  
67  
68  
69  
70  
71  
72  
73  
74  
75  
76  
77  
78  
79  
80  
81  
82  
83  
84  
85  
86  
87

Lentic and lotic habitats are believed to differentially influence ecological and evolutionary dynamics. Indeed the distinction between these two hydraulic habitat types has been fundamental to the classification of aquatic ecosystems and has strongly influenced the way freshwater scientists conduct research and organize their disciplines (Wetzel 2001, Lottig et al. 2011, Allan et al. 2021). In riverine systems, mechanisms of upstream dispersal are a necessity for plant and animal species (Wubs et al. 2016), while dendritic network patterns create variation in metacommunities among headwater and mainstem habitats (Brown and Swan 2010). Lakes are commonly understood to favor greater dispersal traits, possibly because they are less stable over evolutionary timescales relevant to speciation. For example, lentic odonate species have larger latitudinal ranges than lotic species in the Nearctic and Palearctic (Hof et al. 2006). While some studies have found lotic species have greater genetic population differentiation and potential for cryptic diversity (Marten et al. 2006), this may not always be the case (Ribera et al. 2001).

There are few theoretical and empirical examples of studies on the ecological and evolutionary dynamics of individuals that can co-occur in both lentic and lotic habitats. The best examples of lake-stream eco-evolutionary comparisons thus far have come from fishes, especially work on three-spined stickleback (*Gasterosteus aculeatus*). In sticklebacks, co-occurrence seems possible due to morphologic variability and/or parapatric speciation (Thompson et al. 1997, Rennison et al. 2019, Paccard et al. 2020). Interestingly, in a case study transplanting lake-genotyped sticklebacks into streams, survival of lake-genotype fishes was poor and individuals with a hybrid lake-stream genotype had only moderately improved survival (Moser et al. 2016). In another case, freshwater drum (*Aplodinotus grunniens*) exhibited more robust bodies in rivers and reservoirs with lower retention time (more flow), yet interestingly this species can show amenability to both lentic and lotic habitats beyond the age of ~12 years (Rypel et al. 2006). Minnows (*Phoxinus*) from lakes and streams often also exhibit a similar morphologic pattern, though some evidence to the contrary suggests that in minnows this may be region-dependent (Ramler et al. 2017, Scharnweber 2020).

Species that co-occur in lotic and lentic systems may be especially common in high altitude, glaciated mountain landscapes, where lakes are often hydrologically linked in chains by

88 stream segments. High mountain lakes and streams are often oligotrophic, and wave action along  
89 rocky littoral zones of lakes produces microhabitats that can resemble headwater streams (Merritt  
90 and Cummins 1996, Baker et al. 2016). Stream-dwelling invertebrates have been observed to live  
91 in the inlet and outlet regions of high elevation lakes (Wissinger et al. 2016), yet the ecological  
92 and evolutionary dynamics of populations of aquatic organisms that co-occur in these  
93 mountainous lake and stream habitats remains poorly understood. Clarifying lentic-lotic  
94 population dynamics, especially in sensitive mountain ecoregions, would provide a basis to  
95 assess ecological and evolutionary behaviors of aquatic organisms and how these may alter in  
96 future climate change scenarios.

97 Here, we test whether populations of the caddisfly *Limnephilus externus* (Trichoptera:  
98 Limnephilidae) that co-occur in lakes and streams are evolutionarily and/or ecologically distinct.  
99 Specifically, we compare population genetic structure, abundance, larval phenology, larval body  
100 and case morphology, and interspecies phoretic interactions between lentic and lotic populations  
101 of *L. externus*. We follow this with a brief examination of morphologic differences between the  
102 three distinct *L. externus* haplogroups that emerged from this analysis.

103

104

## METHODS

105

106

### Study Organism

107

108

109

110

111

112

113

114

115

116

117

*Limnephilus externus* Hagen (Trichoptera: Limnephilidae) is a caddisfly whose larvae  
typically inhabit lentic habitats, such as lakes, permanent to semi-permanent shallow ponds, and  
wetlands (**Figure 1**) (Berté and Pritchard 1986, Wissinger et al. 2003, Jannot et al. 2008). The  
five larval instars and the pupa are aquatic; after pupation *L. externus* emerge as a terrestrial  
winged adult (**Figure 2**). Larvae create bulky cylindrical non-rigid cases, or “hedgehog cases”  
(Johansson and Johansson 1992), assembled from fragments of vegetation, detritus, and other  
organic matter (Berté and Pritchard 1986, Wiggins 2004). While *L. externus* flight duration is not  
well documented, adults of this species likely live less than 2 months (Berté and Pritchard 1986,  
Wissinger et al. 2003). *Limnephilus externus* is well documented in lake habitats throughout the  
western North America, Canada, and the Palearctic (Morse 1993, Ruiter et al. 2013, Mendez et  
al. 2019). There are very few records of larvae of *Limnephilus spp.* in streams; in California

118 *Limnephilus spp.* is widely known from lakes but outside of this study we are only aware of  
119 several documented stream site records (Pratha 2014).

120

121

### Study Area and Sampling

122 Sampling occurred in two regions in the northern Sierra Nevada mountain range,  
123 California, USA (**Figure 3**). The Lakes Basin, in the northern Sierra Nevada, is a high elevation  
124 (2000m) mountain region featuring a dendritic network of headwater streams and oligotrophic  
125 lakes. Six lakes and six streams of close proximity were selected from more than twenty glacial  
126 lakes and their connecting streams (**Figure 4**). These lakes occur in the headwaters of two  
127 adjoining watersheds: the Feather River (Silver, Little Bear, Big Bear, and Goose lakes) and the  
128 Yuba River (Upper and Lower Salmon lakes). To add context to the study, we also sampled *L.*  
129 *externus* populations from one additional lake (without inlet or outlet stream) in the Lakes Basin,  
130 Haven Lake (Feather River watershed), as well as a lake-stream pair in a second region ~100km  
131 south (Tamarack Lake and outlet stream, Upper Truckee River watershed); these contextual  
132 samples were only used in phenology and population genetics analyses.

133 In the winter preceding this study (2016 – 2017), California experienced above average  
134 rainfall and snowpack and thus above average streamflow (Guirguis et al. 2019). The first  
135 sampling event in late June 2017 occurred during peak snowmelt and streamflow. A second  
136 sampling event in July 2017 occurred after peak water levels had subsided.

137 Water quality parameters, measured as spot samples during population and habitat  
138 surveys, were similar across all lake and stream study sites and typical of water quality in the  
139 higher elevations of the Sierra Nevada mountains. Conductivity was consistently below 25  
140  $\mu\text{S}/\text{cm}$ , while pH in both lakes and streams was neutral (pH 6.1-7.6). Dissolved oxygen levels  
141 were typically near saturation (70-90%), with lower values occurring during early morning  
142 hours, reflecting some moderate diurnal fluctuations. Water temperatures were similar among  
143 lakes and streams, and were higher, on average, in July (20.5 °C) than June (18.0 °C).

144 Sampling for *L. externus* took place in both lotic (stream habitats within 100m of lake  
145 outlets or inlets) and lentic (at least 100m from the nearest inlet or outlet) habitats. Five 1 m<sup>2</sup>  
146 sampling areas were selected along the littoral zone of lakes and the benthic zones of streams in  
147 water depths of 5-50cm. Sampling areas were spaced at least 1 m apart. Population surveys were  
148 performed for a timed interval (12 minutes per 1 m<sup>2</sup> area) by sampling a combination of cobble,

149 boulder, and bedrock. At each site we examined and picked up 100-125 cobble-sized rocks to  
150 document the abundance of *L. externus*. All individual *L. externus* were preserved in 70%  
151 ethanol and taken to the lab for further analysis.

152 In the lab, *L. externus* larvae were roughly sorted into 5 instars based on case size. All  
153 subsequent analyses were performed using only individuals of the largest size class (presumed  
154 fifth instar). *A posteriori* measurements of head capsule width of the largest size class (mean  
155 1.57 mm) were similar to ranges for 5<sup>th</sup> instar *L. externus* larvae reported in other studies (mean  
156 1.62 mm, (Berté and Pritchard 1986); mean 1.60 mm, (Wissinger et al. 2003)).

157

158

### Population Genetics

159 We examined genetic variation among sampled *L. externus* populations through sequencing and  
160 analysis of the mitochondrial cytochrome *c* oxidase subunit 1 (COI) gene. We removed a single  
161 leg from twenty-nine individuals and placed each leg in a unique microplate well with 1-2 drops  
162 of 70% ethanol. Samples were sent to the Canadian Center for DNA Barcoding at the University  
163 of Guelph for standard DNA extraction, mtDNA COI gene isolation, and gene amplification,  
164 with established QA/QC standards.

165 To examine genetic variation in our populations in the context of populations collected  
166 elsewhere, we aligned and compared the returned COI sequences to those found in  
167 BOLDSYSTEMS (Ratnasingham and Hebert 2007). We searched the BOLDSYSTEMS Public  
168 Data Portal for nucleotide sequences belonging to “*Limnephilus externus*” and exported all 252  
169 matching records and their metadata; data came from ten institutions, spanned three countries,  
170 and broke into three Barcode Index Number (BIN) clusters (i.e., algorithm-generated operational  
171 taxonomic units that are performed once per month based on diverging sequences)  
172 (Ratnasingham and Hebert 2013). We removed twenty-four sequences without BIN information  
173 and forty sequences with invalid residues. All sequences were aligned using a global alignment  
174 with free end gaps and a 65% similarity cost matrix. Additional sequences were removed if they  
175 showed many gaps in the nucleotide alignment, were too short relative to the other aligned  
176 sequences, or were of duplicate locations with identical (or nearly identical (<0.002)) sequences  
177 congregated within the same haplogroup branch. The final nucleotide alignment comprised 29  
178 original sequences and 25 unique BOLDSYSTEMS sequences which may be found in **Dataset**  
179 **S1**.

180 Phylogenetic trees of the 54 COI sequences were constructed using both a distance-  
181 matrix method (UPGMA) and a Bayesian inference method (MrBayes (v3.2.6)). We built trees  
182 using UPGMA for three different pairwise genetic distance models (i.e., *Jukes-Cantor*, *HKY*,  
183 *Tamura-Nei*) using a bootstrap resampling method (100 replicates). Bayesian analyses used both  
184 the *JC69* (nst=1) and *HKY85* (nst=2) substitution models (Huelsenbeck and Ronquist 2001). We  
185 selected the only two imported BOLDSYSTEMS sequences available from the Palearctic  
186 (Finland) as outgroups. All trees produced with the UPGMA and Bayesian analyses contained  
187 similar distinct clades and haplogroups, thus we only present results from the UPGMA Jukes-  
188 Cantor model that assumes equal rates of nucleotide substitutions as an inferred phylogenetic  
189 relationship. Algorithm-generated BIN assignments from BOLDSYSTEMS are included in the  
190 branch label of exported BOLD sequences. We identified haplogroups using a criteria of  $\geq 0.01$   
191 (1%) dissimilarity between parallel branches that resulted in substantially larger variation  
192 between groups than within groups. Initial metadata review was performed in R. Nucleotide  
193 sequence alignments and phylogenetic tree construction used Geneious software (v 10.2.3).

194

195

### Morphology and Phoretic Associations

196 All collected *L. externus fifth-instar* larvae (n=44; 27 lake, 17 stream) and their associated cases  
197 were individually photographed, given unique identification codes, and examined under a  
198 dissection microscope at 10-20x magnification. Each individual was measured for head-capsule-  
199 width (HCW), body length, pronotum length, body width at both the pronotum and 2<sup>nd</sup>  
200 abdominal segment, case length, and case width at its widest point. Body morphology was  
201 measured using a micrometer ( $\pm 0.01$ mm) and case morphology was measured using calipers  
202 ( $\pm 0.1$ mm).

203 We qualitatively documented the following body and case morphologic features for each  
204 collected individual: abdominal condition, gill length, gill thickness, head capsule pigmentation,  
205 abdominal mites, case width type, presence of silt in the case, case material type, case sturdiness  
206 or fragility, case material length, lateral case extensions (Limm and Power 2011), case assembly  
207 uniformity, and case microinvertebrate hitchhikers (**Table 1**). Two distinct conditions of the  
208 ventral abdomen were also observed: even color tone with robust appearance, and black spotting  
209 with a transparent cuticle. Finally, a variety of microinvertebrates ( $< 500\mu\text{m}$ ; e.g., Chironomidae,  
210 Acari, Oligochaeta, *Hydra*) were found attached to or embedded in caddisfly cases, as well as

211 clinging to abdominal gills. These associated microinvertebrate taxa were coarsely identified,  
212 enumerated, and separately preserved from caddisfly larvae in 70% ethanol.

213 To examine possible differences between lake and stream populations, we performed  
214 two-tailed t-tests assuming equal variance for the seven quantitative variables, and Fisher's exact  
215 tests of independence on the qualitative nominal data. We also performed two-way ANOVAs  
216 and Tukey's post-hoc tests to determine differences in the same seven quantitative variables  
217 among the three haplogroups identified in phylogenetic analyses, and Fisher's exact tests were  
218 used for qualitative differences among haplogroups (n=29; 10 in clade one; 6 in clade two; 13 in  
219 clade three). We report all p-values less than 0.05. All analyses were performed in R (R version  
220 4.2.0, R Core Team 2022).

221

222

## RESULTS

223

224

### Distribution and population genetic structure

225

226

227

228

229

230

231

232

233

234

235

236

237

238

239

240

*Limnephilus externus* was widely distributed in both lakes and streams but was more abundant in lakes. Although *L. externus* is known primarily as a lake-dwelling caddisfly, we documented its presence in 7 of 7 lakes and 5 of 7 streams (**Table 2**). We regularly collected twenty individuals per hour at 4 lakes in June (larvae were not observed from Upper Salmon Lake and Little Bear Lake) and at 5 lakes in July (larvae were not observed from Lower Salmon Lake). In contrast, no larvae could be found in the streams experiencing high snowmelt flows in June. In July, one stream (Lower Salmon outlet) yielded at least 20 individuals per hour, while others had lower abundance (<10 could be attained per hour). Many empty cases were observed in both lakes and streams in July. In lakes, larvae were commonly found on or near submerged vegetation (e.g., aquatic grasses), while in streams larvae were found primarily attached to stable substrates (e.g., fallen logs) in pools. While 5th instar larvae were present among all *L. externus* populations in July 2017, the proportion of instars varied greatly among sites (**Figure 5**). Fifth instars were the dominant size class at Big Bear Lake and Upper Salmon Lake. Lakes had roughly equal proportion of 4th and 5th instars (40.5% and 44.6%, respectively), yet streams had more 5th instars (59.6%) than 4th instars (14.9%). Few of the individuals we collected were 1st-3rd instars (lakes 14.9%, streams 25.5%).



241 Analysis of the mtDNA COI gene sequences indicates a high degree of intraspecies  
242 variation, low geographic structure, and wide geographic distribution of haplogroups. Lake and  
243 stream individuals from the Lakes Basin formed three distinct haplogroups, and the three  
244 haplogroups comprised of Lakes Basin individuals correspond with the three BOLDSYSTEMS  
245 algorithm-generated BINS (**Figure 6**). Within group dissimilarity (haplogroup one: range 0.1-  
246 0.5%, mean 0.2%; haplogroup two: 0.1-0.5, mean 0.2%; haplogroup three: range 0.1-0.4%, mean  
247 0.2%) was much less than between-group dissimilarity (haplogroups one and two: range 0.8-  
248 1.1%, mean 0.9%; haplogroups one and three: range 0.8-1.4%, mean 1.1%; haplogroups two and  
249 three: range 0.8-1.2% mean 1.0% ). All three haplogroups included individuals from both the  
250 United States and Canada, indicating that the three genetically distinct haplogroups are widely  
251 distributed. The first haplogroup includes multiple individuals from Lakes Basin, all the sampled  
252 individuals from the Tamarack study site, as well as individuals collected outside this study from  
253 other parts of the Sierra Nevada (Mono County, CA), Washington (USA), and Manitoba  
254 (Canada). The second haplogroup includes individuals predominantly from the Upper and Lower  
255 Salmon lake watershed and one individual from Big Bear Lake (Lakes Basin), as well as from  
256 the Rocky Mountains (Colorado) and individuals from across Canada (Alberta, Manitoba, New  
257 Brunswick). The third haplogroup includes individuals from the hydrologically connected  
258 system that includes Silver, Little Bear, and Big Bear lakes and streams, as well as nearby Goose  
259 and Haven Lakes (Lakes Basin), plus one individual from Manitoba (Canada).

260

261

### Morphology

262

263

264

265

266

267

268

269

270

271

*L. externus* larvae exhibited significant differences in abdomen condition and gill thickness between lake and stream individuals (fisher test,  $p=0.0002$  and  $p=0.0008$ , respectively). Black-spotted, transparent abdomens with attenuated gills were more common in lake individuals (100%) than stream individuals (28.6%) (**Figure 7**). Thick abdominal gills were also more common among lake individuals (72.7%) than stream individuals (8.33%) (**Figure 8**).

Caddisfly case construction and materials varied substantially among habitats and over time (**Figure 9**). Cases were significantly longer in lakes than streams in the Lakes Basin (t-test,  $p=0.0001$ ). There were no other significant differences in cases among lake and stream individuals. Cases included more aquatic vegetation in June, while in July cases were constructed predominantly with twigs and bark. All cases from Tamarack Lake and outlet were

272 fragile, bulky, and frequently had lateral extensions made with thin twigs; in contrast, all Lakes  
273 Basin *L. externus* cases exhibited stronger construction and no lateral case extensions.

274 Among the three haplogroups, there were significant differences in pronotum length  
275 ( $F=6.31$ ,  $p=0.0068$ ), body length ( $F=4.64$ ,  $p=0.0208$ ), and case length ( $F=4.98$ ,  $p=0.0183$ ). A  
276 Tukey's post hoc test revealed pronotum length was shorter in haplogroup one compared to  
277 haplogroups two-three, and body length was shorter in haplogroup one than haplogroup two;  
278 haplogroup three exhibited similarities with haplogroup one and two in different characteristics  
279 (**Figure 10**). Head pigmentation and case structure sturdiness also were significantly, or nearly  
280 significantly, different across haplogroups (fisher test,  $p = 0.0626$  and  $p=0.0287$ , respectively).

281

### 282 **Phoretic Associations**

283 *Hydra*, nematodes, oligochaetes, chironomid midges (three morphospecies), and water  
284 mites (two morphospecies) were all found securely fastened to many caddisfly cases, either on  
285 the surface or buried into silt in cases (**Figure 9D & 11**). These microinvertebrates were  
286 phoretically associated with both lake (36.6%) and stream (50%) caddisfly cases. We did not  
287 observe differences in the microinvertebrate community composition between cases from lake  
288 and stream individuals.

289 Abdominal mite presence was significantly different between lake-stream habitats and  
290 among haplogroups (fisher test,  $p=0.013$  and  $p=0.0397$ , respectively). Mites were only found on  
291 the abdomen of individuals from lakes (40.9%), not streams (0%); however abdominal mite  
292 infestation was only observed at Upper Salmon and Big Bear Lakes, with all individuals  
293 belonging to haplogroup three. The highest abdominal mite infestation was 31 mites on a single  
294 individual; infested individuals had a mean of 4 mites. All individuals with water mites on their  
295 abdomen were observed to be less robust, had dark and transparent abdomens, and attenuated  
296 black spotted gills. However, nearly half of the larvae that lacked water mites at the time of  
297 collection also had some of these characteristics.

298 Water mites observed on the exterior of caddisfly cases were identified as adult oribatids  
299 (Acariformes: Sarcoptiformes: Oribatida), possibly in the family Trhypochthoniidae or  
300 Malaconothridae, while those clinging to the abdomen were identified as larval hygrobatoid  
301 water mites (Acariformes: Parasitengona: Hydrachnidia: Hygrobatoidea), possibly in the family

302 Hygrobatidae or Unionicolidae (Heather Proctor, University of Alberta, personal  
303 communication).

304

305

## DISCUSSION

306

307 We formally documented the presence of *Limnephilus externus*, a caddisfly widely  
308 known from lentic habitats throughout North America, in both lake and stream habitats in the  
309 Sierra Nevada. We examined the degree to which *L. externus* occurring in lakes and streams are  
310 evolutionarily and ecologically distinct by comparing: (1) population genetic structure, (2)  
311 abundance, (3) larval phenology, (4) larval body and case morphology, and (5) interspecies  
312 phoretic interactions. Further, we briefly explored the potential for morphologic differences  
313 between distinct haplogroups of *L. externus*.

314 Constructing phylogenetic relationships through the use of the mitochondrial cytochrome  
315 *c* oxidase I (COI) gene (i.e., DNA barcoding) helps to reveal patterns in biodiversity (Hebert et  
316 al. 2003). Studies connecting the techniques of DNA barcoding with traditional taxonomy have  
317 increasingly reported higher cryptic diversity than previously suspected (Sheth and Thaker 2017,  
318 DeSalle and Goldstein 2019). Indeed use of the cytochrome gene has revealed high genetic  
319 diversity and low geographic structure in other aquatic insect species (Ståhls and Savolainen  
320 2008, Zhou et al. 2010 p. 010, Sproul et al. 2014). This study found no genetic differences  
321 between lake and stream *L. externus*. Instead we found three geographically widespread and  
322 genetically distinct haplogroups, separated by 1-2% genetic difference; all three haplogroups  
323 were present in lakes and streams in the Sierra Nevada. These putative haplogroups may  
324 represent distinct subspecies (White et al. 2014), or may not be biologically meaningful without  
325 additional multilocus data (Dasmahapatra et al. 2010) as a minimum 2-3% genetic divergence is  
326 often used to distinguish haplogroups as separate subspecies or species. The three haplogroups in  
327 our analysis do match with the three algorithm-generated BINS identified by BOLDSYSTEMS,  
328 which are intended to closely approximate species. *L. externus*' three haplogroups are widely  
329 distributed throughout the United States and Canada. Our findings suggest *L. externus* has a wide  
330 geographic range and low geographic structure that could support phenotypic plasticity between  
331 habitat types, and possibly genotypic and phenotypic variation at the haplogroup level. These  
332 results also suggest *L. externus* exhibits potentially high cryptic biodiversity and may be well

333 adapted to disperse long distances. Relative to other insect species the COI mtDNA gene evolves  
334 quickly within the *Limnephilus* genus, supporting its use when exploring recent divergences  
335 (McCullagh et al. 2015, Steinke et al. 2022). Our results support and expand on the extensive  
336 genetic analysis and findings of *L. externus* in the Manitoba province of Canada (Zhou et al.  
337 2011, Ruiter et al. 2013). This rapid evolution of the COI gene, or the widespread gene flow  
338 hypothesis, could account for why three widely distributed haplogroups had large variation in  
339 morphology and may be found distributed across entire countries. This phylogenetic finding  
340 alludes to hidden biodiversity patterns and the need to further identify species boundaries in  
341 aquatic insect taxa.

342 Lake and stream populations exhibited distinct ecological phenotypes in abundance,  
343 phenology, some aspects of body and case morphology, and abdominal mite presence. Fifth  
344 instar *L. externus* were present at all lake-stream sites, while other instars varied in proportion.  
345 All lake individuals had abdomens that were transparent (tracheae were visible), black-spotted,  
346 and with more attenuated gills, whereas a small fraction of stream individuals had these  
347 characteristics. Lake individuals were also observed to have thicker abdominal gills. These  
348 morphological differences could represent adaptations resulting from several possible abiotic  
349 factors that differ between lakes and streams (e.g., lower levels of dissolved oxygen in lakes).  
350 Similarly, gill breadth and visible tracheae have been key factors in distinguishing the lentic  
351 *Baetis tracheatus* from the lotic *B. bundyae*, which has narrow gills and invisible tracheae (i.e.,  
352 abdomen not transparent) (Engblom 1996, Ståhls and Savolainen 2008). On the other hand,  
353 research has linked altered and atrophied tracheal gills (i.e., black speckling) in caddisflies to the  
354 introduction of pollutants or bacteria in a headwater stream (Simpson 1980).

355 Lake *L. externus* constructed cases using longer pieces of material than those in streams,  
356 and cases from the Tamarack region had weaker construction. Caddisfly case construction is  
357 highly dependent on the availability of materials in the surrounding habitat, yet the observed  
358 differences in case structure could also reflect adaptations to abiotic or biotic factors (i.e., flow,  
359 predator defense). For example, *L. externus*' stout cases are reported to be a better deterrent to  
360 predation by beetle larvae than some more tubular cases of other species (Wissinger et al. 2006),  
361 while another study reported that differences in case structure between two *Limnephilus* species.  
362 (*L. pantodapus* and *L. rhombicus*) affected the behavior of predaceous dragonfly larvae  
363 (Johansson and Johansson 1992). Indeed the construction of more protective cases has been

364 found to be a resource allocation trade-off inducible by predator chemical cues (Correa-Araneda  
365 et al. 2017).

366 Across haplogroups, pronotum length, total body length, case sturdiness, and presence of  
367 abdominal mites were significantly different between at least two haplogroups. Haplogroups also  
368 exhibited a nearly significant difference in head pigmentation, which has previously been used to  
369 distinguish between *Limnephilus* species (Ruiter et al. 2013). We consider these morphological  
370 haplogroup differences to suggest that real clade-level differences may exist and should be  
371 further studied. This study was designed to investigate differences between lakes and streams in  
372 one region, and therefore a representative sampling of each haplogroup may not have been  
373 achieved.

374 Finally, a collection of microinvertebrates (i.e., chironomid midges, water mites, hydrae,  
375 oligochaetes) were discovered buried within *L. externus* cases. In addition, water mites were  
376 found on the abdomen of only lake individuals. We observed at least three morphospecies of  
377 chironomid midge on the cases, suggesting that the microinvertebrate community on the cases  
378 may be diverse. Water mites observed on case exteriors were identified as adult oribatid  
379 (Acariformes: Sarcoptiformes: Oribatida) mites, possibly in the family Trhypochthoniidae  
380 (Heather Proctor, University of Alberta, personal communication). Oribatids commonly feed on  
381 detritus, algae, and occasionally macrophytes (Behan-Pelletier and Hill 1978, Proctor and  
382 Pritchard 1989). The association of Oribatid mites on the organic cases suggests a commensal  
383 relationship in which the mites could be benefiting by living in or feeding on the cases. The  
384 nature of these ecological associations at these locations is not known, however, *L. externus* did  
385 not appear to be negatively affected or parasitized by any of the microinvertebrates on the  
386 exterior of their cases. Therefore, in these instances, we suspect a phoretic (non-harmful)  
387 association. In contrast, mites found on the abdomen of *L. externus* larvae may pose greater  
388 threat. Abdominal water mites were identified as larval hygrobatoid water mites (Acariformes:  
389 Parasitengona: Hydrachnidiae: Hygrobatoida), possibly in the family Hygrobatidae or  
390 Unionicolidae (Heather Proctor, University of Alberta, personal communication). Hygrobatoid  
391 mites are known to engage in pre-parasitic attendance of caddisflies, remaining near the host  
392 until it is close to pupation and feeding on it when it emerges as an adult (Proctor and Pritchard  
393 1989).

394 The occurrence of phoretic and parasitic relationships is common among aquatic  
395 organisms. Other aquatic insects have been documented to play host to midge and water mite  
396 travelers in relationships that vary along the gradient of ectoparasitism, predation, and phoresy  
397 (Tracy and Hazelwood 1983, Henriques-Oliveira and Nessimian 2009, Buczyńska et al. 2015).  
398 In Quebec, Canada, *Limnephilus* has been documented to have water mite larvae  
399 (Hygrobratoidea), with prevalence ranging from 4-42% (Fairchild and Lewis 1987). Other aquatic  
400 organisms, like the fish *Ancistrus multispinis* in Atlantic forest streams in Southeastern Brazil,  
401 have chironomid larvae in phoretic association (Mattos et al. 2018). Understanding the role of  
402 associated macroinvertebrates on aquatic organisms is a challenging topic to study; (Grabner  
403 2017) found testing for parasitic taxa using PCR might be an efficient and cost-effective method  
404 to identifying links between host feeding type and prevalence. Additional studies would be  
405 needed to identify the nature of these associations and their consequences to *L. externus*.

406  
407

## 408 CONCLUSION

409

410 In this study, we documented the presence of *Limnephilus externus* in both lake and  
411 stream habitats. Lake populations had conspicuous abdominal tracheae, thicker gills, and black  
412 spotting. Lake populations exhibited longer case construction, and only caddisfly cases from the  
413 Tamarack region were significantly more fragile in construction. Microinvertebrate hitchhikers  
414 found on the cases of the caddisflies are presumed to maintain a phoretic relationship, while  
415 mites on the abdomen may be demonstrating pre-parasitic attendance behavior. Finally, while  
416 lake populations were not genetically different from stream populations, we did find three  
417 geographically widespread haplogroups present in the Sierra Nevada as well as throughout  
418 western North America and Canada. These putative haplogroups exhibited some significant  
419 morphological variation but further research is needed to validate these results. Overall, our  
420 observations and analyses suggest that environmental differences between lake and stream  
421 habitats may produce variation in plastic traits, but dispersal and gene flow are likely preventing  
422 genetic differentiation.

423 The frequency of aquatic invertebrate species that co-inhabit lentic and lotic ecosystems  
424 is unknown, and reflects the paucity of studies of aquatic fauna across habitat types. Our findings  
425 suggest that species with plastic traits amenable to both flow types may be overlooked in aquatic

426 research. As a result, we may be missing valuable information on ecological and evolutionary  
427 behaviors of aquatic organisms, especially in light of anticipated climatic changes.

428         While lotic and high elevation lake shoreline habitats have been recognized for their  
429 ecological similarities, the way these two distinct ecosystems will respond to climatic changes  
430 will be vastly different. Indeed (Wissinger et al. 2016) observed cold-water stream insects  
431 inhabiting rocky and wave-swept alpine lake shorelines of Colorado, Switzerland, and New  
432 Zealand, and evidence that freshwater fauna may be amenable to both hydraulic habitat types is  
433 growing (Yarnell et al. 2019). Mountain systems in particular face high stressors and are  
434 sensitive to environmental changes (Moser et al. 2019). Many of the aquatic habitats in the Sierra  
435 Nevada are dependent on snowmelt, yet California's increasingly common drought years and  
436 resulting low snowpack are anticipated to decrease snowmelt feeding into aquatic systems (Smits  
437 et al. 2020). With deteriorating snowpack and warming lakes, the adaptability of aquatic fauna to  
438 find refugia is expected to be a tremendous benefit to their survival (Birrell et al. 2020, Frakes et  
439 al. 2021). With this study, we hope to contribute to a larger body of knowledge and facilitate  
440 directions for future mountain aquatic research.

441

442

#### ACKNOWLEDGEMENTS

443 CAP was supported by the UC Davis Center for Watershed Sciences' Bechtel Next Generation  
444 Funds. We thank John Wheeler for assistance with DNA barcoding. We thank the biologists who  
445 contributed to fruitful discussion on *Limnephilidae* during the conceptualization of this project:  
446 Scott Wissinger, Dave Ruitter, Dave Herbst, Roland Knapp, Bob Wisseman, Mark Wetzel,  
447 Jonathan Lee, John Epler, Steve Cairns, Allen Collins, Richard Campbell, Patina Mendez;  
448 Heather Proctor for knowledgeable input and identification of the water mites; and Isaac  
449 Chellman (CDFW), Nick Russelson (FS), and James Matthew Johnson (FS) for support in field  
450 research discussion and permissions.

451

452

453

## LITERATURE CITED

- 454
- 455
- 456 Allan, J. D., M. M. Castillo, and K. A. Capps. 2021. *Stream Ecology: Structure and Function of*  
 457 *Running Waters*. Springer Nature.
- 458 Baker, M. A., C. D. Arp, K. J. Goodman, A. M. Marcarelli, and W. A. Wurtsbaugh. 2016.  
 459 Chapter 7 - Stream-Lake Interaction: Understanding Coupled Hydro-Ecological Systems.  
 460 Pages 321–348 in J. B. Jones and E. H. Stanley (editors). *Stream Ecosystems in a*  
 461 *Changing Environment*. Academic Press, Boston.
- 462 Behan-Pelletier, V., and S. B. Hill. 1978. Feeding habits and spore dispersal of arctic oribatid  
 463 mites. *Revue Ecologie Biologie du Sol* 15:497–516.
- 464 Berté, S. B., and G. Pritchard. 1986. The life histories of *Limnephilus externus* Hagen, *Anabolia*  
 465 *bimaculata* (Walker), and *Nemotaulius hostilis* (Hagen) (Trichoptera, Limnephilidae) in a  
 466 pond in southern Alberta, Canada. *Canadian Journal of Zoology* 64:2348–2356.
- 467 Birrell, J. H., A. A. Shah, S. Hotaling, J. J. Giersch, C. E. Williamson, D. Jacobsen, and H. A.  
 468 Woods. 2020. Insects in high-elevation streams: Life in extreme environments imperiled  
 469 by climate change. *Global Change Biology* 26:6667–6684.
- 470 Brown, B. L., and C. M. Swan. 2010. Dendritic network structure constrains metacommunity  
 471 properties in riverine ecosystems. *Journal of Animal Ecology* 79:571–580.
- 472 Buczyńska, E., P. Buczyński, A. Zawal, G. Michoński, and A. Szlauer-Łukaszewska. 2015. First  
 473 record of parasitism of water mite larva (Acari: Hydrachnidia) on the pupa of  
 474 Trichoptera. *Acta Parasitologica* 60:196–199.
- 475 Correa-Araneda, F., A. Basaguren, R. T. Abdala-Díaz, A. M. Tonin, and L. Boyero. 2017.  
 476 Resource-allocation tradeoffs in caddisflies facing multiple stressors. *Ecology and*  
 477 *Evolution* 7:5103–5110.
- 478 Dasmahapatra, K. K., M. Elias, R. I. Hill, J. I. Hoffman, and J. Mallet. 2010. Mitochondrial  
 479 DNA barcoding detects some species that are real, and some that are not. *Molecular*  
 480 *Ecology Resources* 10:264–273.
- 481 DeSalle, R., and P. Goldstein. 2019. Review and Interpretation of Trends in DNA Barcoding.  
 482 *Frontiers in Ecology and Evolution* 7.
- 483 Engblom, E. 1996. Ephemeroptera, Mayflies. Page 53 in A. Nilsson (editor). *Aquatic Insects of*  
 484 *North Europe—A Taxonomic Handbook*. Apollo Books, Stenstrup.
- 485 Fairchild, W. L., and D. J. Lewis. 1987. Parasitic Water Mite Larvae (Hydrachnida:  
 486 Hygrobatoidae) Associated With Caddisfly Larvae (Trichoptera: Leptoceridae,  
 487 Limnephilidae). *The Canadian Entomologist* 119:809–813.
- 488 Frakes, J. I., J. H. Birrell, A. A. Shah, and H. A. Woods. 2021. Flow increases tolerance of heat  
 489 and hypoxia of an aquatic insect. *Biology Letters* 17:20210004.
- 490 Grabner, D. S. 2017. Hidden diversity: Parasites of stream arthropods. *Freshwater Biology*  
 491 62:52–64.
- 492 Guirguis, K., A. Gershunov, T. Shulgina, R. E. S. Clemesha, and F. M. Ralph. 2019.  
 493 Atmospheric rivers impacting Northern California and their modulation by a variable  
 494 climate. *Climate Dynamics* 52:6569–6583.
- 495 Hebert, P. D. N., A. Cywinska, S. L. Ball, and J. R. deWaard. 2003. Biological identifications  
 496 through DNA barcodes. *Proceedings of the Royal Society of London. Series B:*  
 497 *Biological Sciences* 270:313–321.



- 498 Henriques-Oliveira, A. L., and J. L. Nessimian. 2009. Phoresy and commensalism of  
499 Chironomidae larvae (Insecta: Diptera) in the state of Rio de Janeiro, Brazil. *Lundiana:*  
500 *International Journal of Biodiversity* 10:11–18.
- 501 Hof, C., M. Brändle, and R. Brandl. 2006. Lentic odonates have larger and more northern ranges  
502 than lotic species. *Journal of Biogeography* 33:63–70.
- 503 Huelsenbeck, J. P., and F. Ronquist. 2001. MRBAYES: Bayesian inference of phylogenetic  
504 trees. *Bioinformatics* 17:754–755.
- 505 Jannot, J. E., S. A. Wissinger, and J. R. Lucas. 2008. Diet and a developmental time constraint  
506 alter life-history trade-offs in a caddis fly (Trichoptera: Limnephilidae). *Biological*  
507 *Journal of the Linnean Society* 95:495–504.
- 508 Johansson, A., and F. Johansson. 1992. Effects of two different caddisfly case structures on  
509 predation by a dragonfly larva. *Aquatic Insects* 14:73–84.
- 510 Limm, M. P., and M. E. Power. 2011. The caddisfly *Dicosmoecus gilvipes*: Making a case for a  
511 functional role. *Journal of the North American Benthological Society* 30:485–492.
- 512 Lottig, N. R., E. H. Stanley, P. C. Hanson, and T. K. Kratz. 2011. Comparison of regional stream  
513 and lake chemistry: Differences, similarities, and potential drivers. *Limnology and*  
514 *Oceanography* 56:1551–1562.
- 515 Marten, A., M. Brändle, and R. Brandl. 2006. Habitat type predicts genetic population  
516 differentiation in freshwater invertebrates. *Molecular Ecology* 15:2643–2651.
- 517 Mattos, T. M., D. R. Carvalho, M. S. de Brito, and F. G. Araújo. 2018. Occurrence of phoresy  
518 between *Ancistrus multispinis* (Actinopterygii: Siluriformes) and *Ichthyocladius* sp.  
519 (Diptera: Chironomidae) in Atlantic forest streams, Southeastern Brazil. *Zoologia*  
520 (Curitiba) 35.
- 521 McCullagh, B. S., S. A. Wissinger, and J. M. Marcus. 2015. Identifying PCR primers to facilitate  
522 molecular phylogenetics in Caddisflies (Trichoptera). *Zoological Systematics* 40:459.
- 523 Mendez, P. K., M. J. Myers, J. E. Damerow, C. Lew, and V. H. Resh. 2019. Species occurrence  
524 and distribution of Trichoptera (caddisflies) in California. *Zoosymposia* 14:113–133.
- 525 Merritt, R. W., and K. W. Cummins. 1996. *An Introduction to the Aquatic Insects of North*  
526 *America*. Kendall Hunt.
- 527 Morse, J. C. 1993. A Checklist of the Trichoptera of North America, including Greenland and  
528 Mexico. *Transactions of the American Entomological Society (1890-)* 119:47–93.
- 529 Moser, D., A. Frey, and D. Berner. 2016. Fitness differences between parapatric lake and stream  
530 stickleback revealed by a field transplant. *Journal of Evolutionary Biology* 29:711–719.
- 531 Moser, K. A., J. S. Baron, J. Brahney, I. A. Oleksy, J. E. Saros, E. J. Hundey, S. A. Sadro, J.  
532 Kopáček, R. Sommaruga, M. J. Kainz, A. L. Strecker, S. Chandra, D. M. Walters, D. L.  
533 Preston, N. Michelutti, F. Lepori, S. A. Spaulding, K. R. Christianson, J. M. Melack, and  
534 J. P. Smol. 2019. Mountain lakes: Eyes on global environmental change. *Global and*  
535 *Planetary Change* 178:77–95.
- 536 Paccard, A., D. Hanson, Y. E. Stuart, F. A. von Hippel, M. Kalbe, T. Klepaker, S. Skúlason, B.  
537 K. Kristjánsson, D. I. Bolnick, A. P. Hendry, and R. D. H. Barrett. 2020. Repeatability of  
538 Adaptive Radiation Depends on Spatial Scale: Regional Versus Global Replicates of  
539 Stickleback in Lake Versus Stream Habitats. *Journal of Heredity* 111:43–56.
- 540 Pratha, S. 2014. MAPIT - A Mapping Application for Freshwater Invertebrate Taxa. USU  
541 Graduate Reports 37.

- 542 Proctor, H., and G. Pritchard. 1989. Neglected Predators: Water Mites  
543 (Acari:Parasitengona:Hydrachnellae) in Freshwater Communities. *Journal of the North*  
544 *American Benthological Society* 8:100–111.
- 545 Ramler, D., A. Palandačić, G. B. Delmastro, J. Wanzenböck, and H. Ahnelt. 2017.  
546 Morphological divergence of lake and stream Phoxinus of Northern Italy and the Danube  
547 basin based on geometric morphometric analysis. *Ecology and Evolution* 7:572–584.
- 548 Ratnasingham, S., and P. D. N. Hebert. 2007. bold: The Barcode of Life Data System  
549 (<http://www.barcodinglife.org>). *Molecular Ecology Notes* 7:355–364.
- 550 Ratnasingham, S., and P. D. N. Hebert. 2013. A DNA-Based Registry for All Animal Species:  
551 The Barcode Index Number (BIN) System. *PLOS ONE* 8:e66213.
- 552 Rennison, D. J., Y. E. Stuart, D. I. Bolnick, and C. L. Peichel. 2019. Ecological factors and  
553 morphological traits are associated with repeated genomic differentiation between lake  
554 and stream stickleback. *Philosophical Transactions of the Royal Society B: Biological*  
555 *Sciences* 374:20180241.
- 556 Ribera, I., T. G. Barraclough, and A. P. Vogler. 2001. The effect of habitat type on speciation  
557 rates and range movements in aquatic beetles: Inferences from species-level phylogenies.  
558 *Molecular Ecology* 10:721–735.
- 559 Ruiter, D. E., E. E. Boyle, and X. Zhou. 2013. DNA barcoding facilitates associations and  
560 diagnoses for Trichoptera larvae of the Churchill (Manitoba, Canada) area. *BMC*  
561 *Ecology* 13:5.
- 562 Rypel, A. L., D. R. Bayne, and J. B. Mitchell. 2006. Growth of Freshwater Drum from Lotic and  
563 Lentic Habitats in Alabama. *Transactions of the American Fisheries Society* 135:987–  
564 997.
- 565 Scharnweber, K. 2020. Morphological and trophic divergence of lake and stream minnows  
566 (*Phoxinus phoxinus*). *Ecology and Evolution* 10:8358–8367.
- 567 Sheth, B. P., and V. S. Thaker. 2017. DNA barcoding and traditional taxonomy: An integrated  
568 approach for biodiversity conservation. *Genome* 60:618–628.
- 569 Simpson, K. W. 1980. Abnormalities in the tracheal gills of aquatic insects collected from  
570 streams receiving chlorinated or crude oil wastes. *Freshwater Biology* 10:581–583.
- 571 Smits, A. P., S. MacIntyre, and S. Sadro. 2020. Snowpack determines relative importance of  
572 climate factors driving summer lake warming. *Limnology and Oceanography Letters*  
573 5:271–279.
- 574 Sproul, J. S., Derek. D. Houston, N. Davis, E. Barrington, S. Y. Oh, R. P. Evans, and D. K.  
575 Shiozawa. 2014. Comparative phylogeography of codistributed aquatic insects in western  
576 North America: Insights into dispersal and regional patterns of genetic structure.  
577 *Freshwater Biology* 59:2051–2063.
- 578 Ståhls, G., and E. Savolainen. 2008. MtDNA COI barcodes reveal cryptic diversity in the *Baetis*  
579 *vernus* group (Ephemeroptera, Baetidae). *Molecular Phylogenetics and Evolution* 46:82–  
580 87.
- 581 Steinke, D., S. L. deWaard, J. E. Sones, N. V. Ivanova, S. W. J. Prosser, K. Perez, T. W. A.  
582 Braukmann, M. Milton, E. V. Zakharov, J. R. deWaard, S. Ratnasingham, and P. D. N.  
583 Hebert. 2022. Message in a Bottle—Metabarcoding enables biodiversity comparisons  
584 across ecoregions. *GigaScience* 11:giac040.
- 585 Thompson, C. E., E. B. Taylor, and J. D. McPhail. 1997. Parallel Evolution of Lake-Stream Pairs  
586 of Threespine Sticklebacks (*Gasterosteus*) Inferred from Mitochondrial Dna Variation.  
587 *Evolution* 51:1955–1965.

- 588 Tracy, B. H., and D. H. Hazelwood. 1983. The Phoretic Association of *Urnatella gracilis*  
589 (Entoprocta:Urnatellidae) and *Nanocladius downesi* (Diptera:Chironomidae) on  
590 *Corydalus cornutus* (Megaloptera:Corydalidae). *Freshwater Invertebrate Biology* 2:186–  
591 191.
- 592 Wetzel, R. G. 2001. *Limnology: Lake and River Ecosystems*. Gulf Professional Publishing.
- 593 White, B. P., E. M. Pilgrim, L. M. Boykin, E. D. Stein, and R. D. Mazor. 2014. Comparison of  
594 four species-delimitation methods applied to a DNA barcode data set of insect larvae for  
595 use in routine bioassessment. *Freshwater Science* 33:338–348.
- 596 Wiggins, G. B. 2004. *Caddisflies: The Underwater Architects*. University of Toronto Press.
- 597 Wissinger, S. a., W. s. Brown, and J. e. Jannot. 2003. Caddisfly life histories along permanence  
598 gradients in high-altitude wetlands in Colorado (U.S.A.). *Freshwater Biology* 48:255–  
599 270.
- 600 Wissinger, S. A., B. Oertli, and V. Rosset. 2016. Invertebrate Communities of Alpine Ponds.  
601 Pages 55–103 in D. Batzer and D. Boix (editors). *Invertebrates in Freshwater Wetlands:*  
602 *An International Perspective on their Ecology*. Springer International Publishing, Cham.
- 603 Wissinger, S. A., J. C. Whissel, C. Eldermire, and W. Brown. 2006. Predator defense along a  
604 permanence gradient: Roles of case structure, behavior, and developmental phenology in  
605 caddisflies. *Oecologia* 147:667–678.
- 606 Wubs, E. R. J., R. G. A. Fraaije, G. A. de Groot, R. H. J. Erkens, A. G. Garssen, E. Kleyheeg, B.  
607 M. Raven, and M. B. Soons. 2016. Going against the flow: A case for upstream dispersal  
608 and detection of uncommon dispersal events. *Freshwater Biology* 61:580–595.
- 609 Yarnell, S. M., R. A. Peek, N. Keung, B. D. Todd, S. Lawler, and C. Brown. 2019. A Lentic  
610 Breeder in Lotic Waters: Sierra Nevada Yellow-Legged Frog (*Rana sierrae*) Habitat  
611 Suitability in Northern Sierra Nevada Streams. *Copeia* 107:676–693.
- 612 Zhou, X., L. M. Jacobus, R. E. DeWalt, S. J. Adamowicz, and P. D. N. Hebert. 2010.  
613 Ephemeroptera, Plecoptera, and Trichoptera fauna of Churchill (Manitoba, Canada):  
614 Insights into biodiversity patterns from DNA barcoding. *Journal of the North American*  
615 *Benthological Society* 29:814–837.
- 616 Zhou, X., J. L. Robinson, C. J. Geraci, C. R. Parker, O. S. Flint, D. A. Etnier, D. Ruitter, R. E.  
617 DeWalt, L. M. Jacobus, and P. D. N. Hebert. 2011. Accelerated construction of a regional  
618 DNA-barcode reference library: Caddisflies (Trichoptera) in the Great Smoky Mountains  
619 National Park. *Journal of the North American Benthological Society* 30:131–162.  
620

621 **Figures**

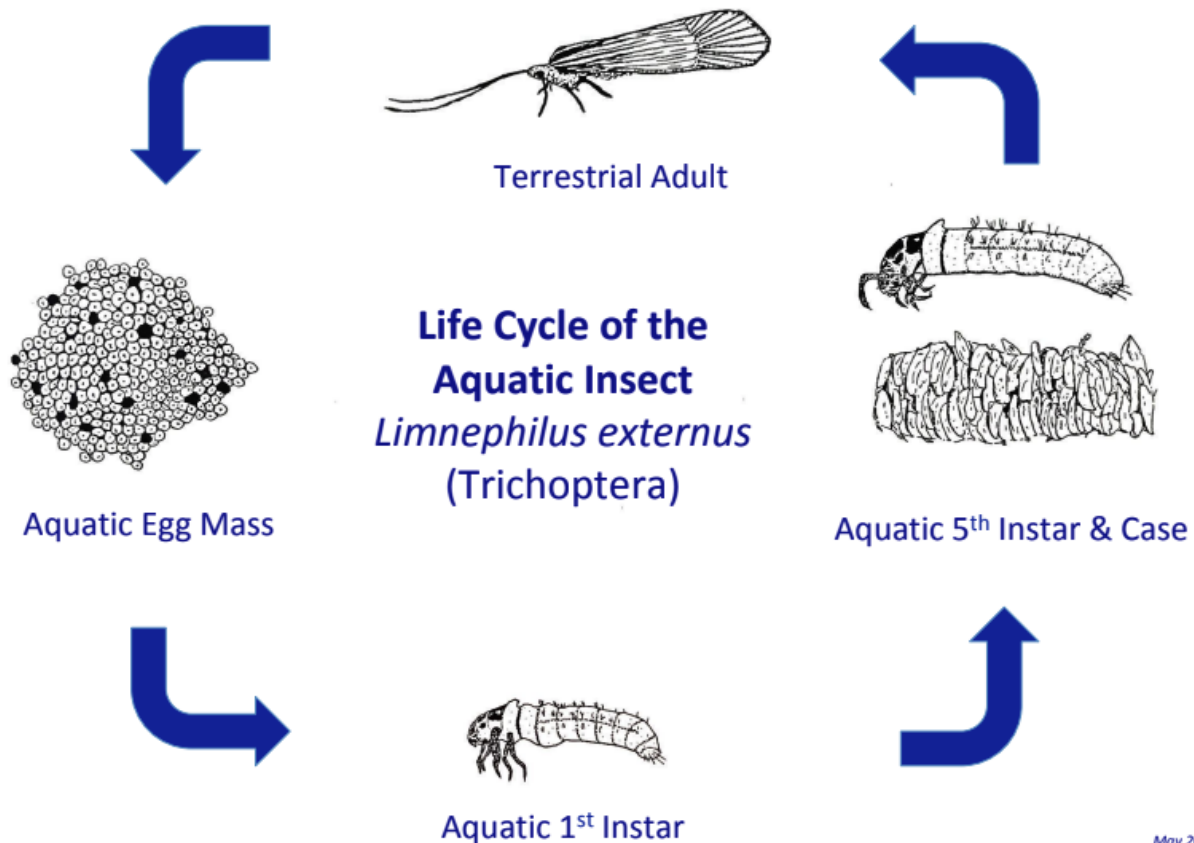
622



623 **Figure 1.** Lateral view of *Limnephilus externus* and its case shown overlaying a metric ruler.

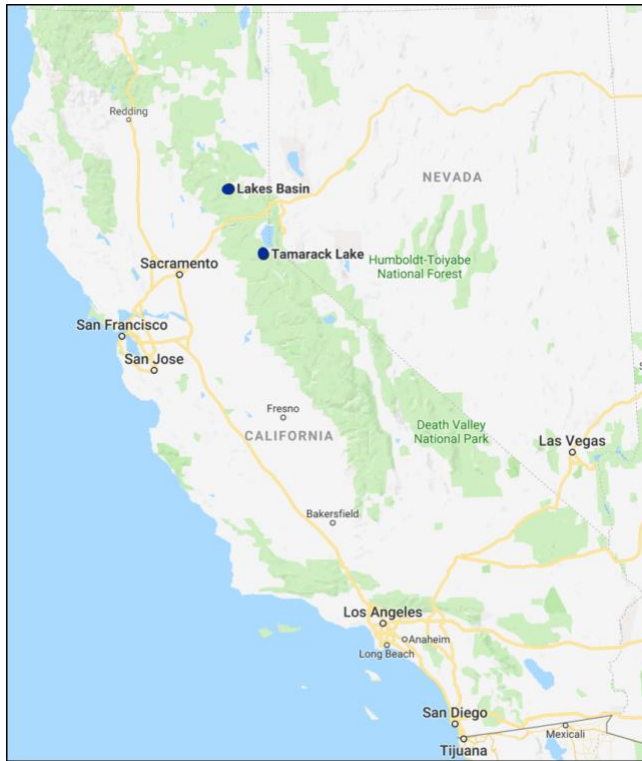
624

625  
626



627  
628  
629  
630

**Figure 2.** Illustration of the life cycle of the aquatic insect *Limnephilus externus*.



631

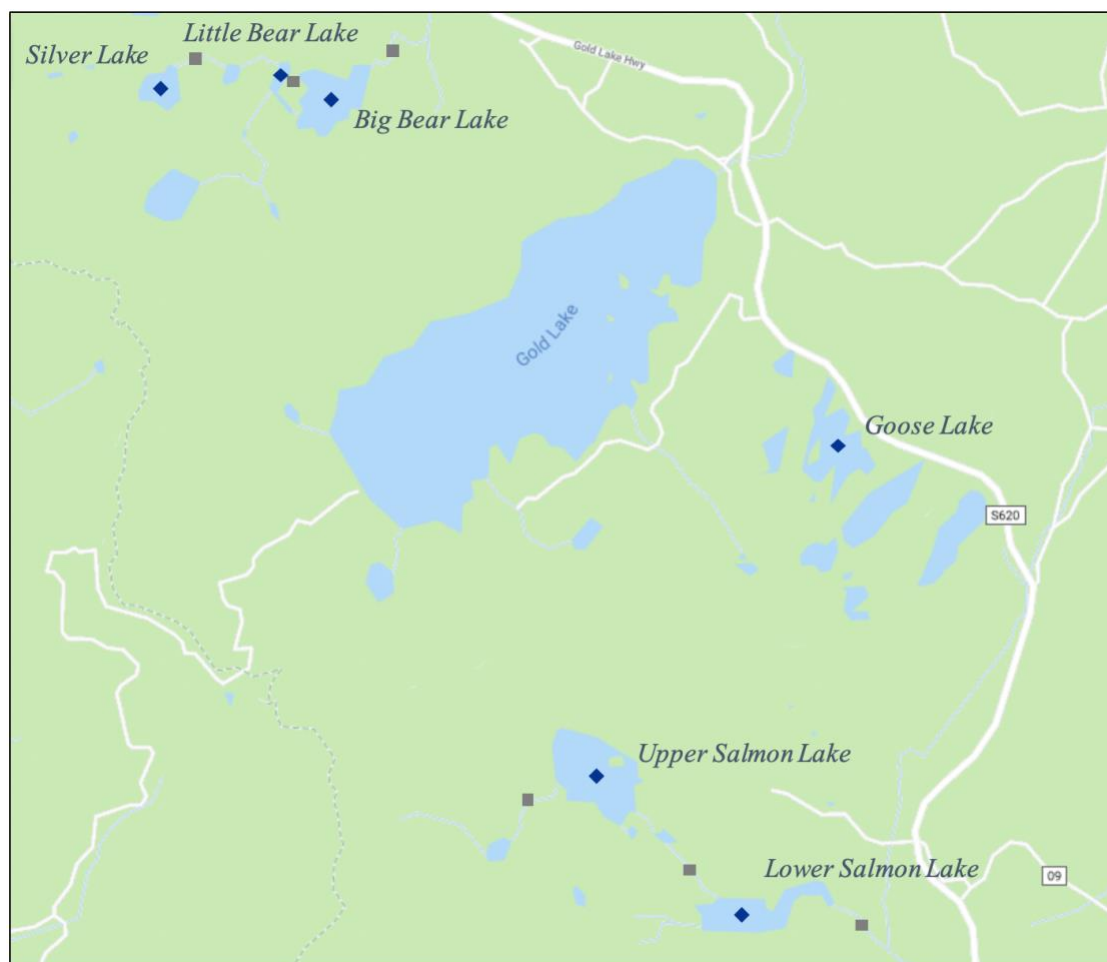
632 **Figure 3.** Map of California, USA showing the primary field location (Lakes Basin) and  
633 contextual site (Tamarack Lake) in the Sierra Nevada mountain range. Map data ©2019 Google,  
634 INEGI.

635

636

637

638



639

640 **Figure 4.** Sampling sites in the Lakes Basin, northern Sierra Nevada, CA. Lotic habitats (gray  
641 squares) and lentic habitats (blue diamonds) are shown. Silver, Little and Big Bear Lakes share  
642 connectivity. Upper and Lower Salmon Lakes share connectivity. Goose Lake has no inlet or  
643 outlet stream. Map data ©2019 Google, INEGI.

644

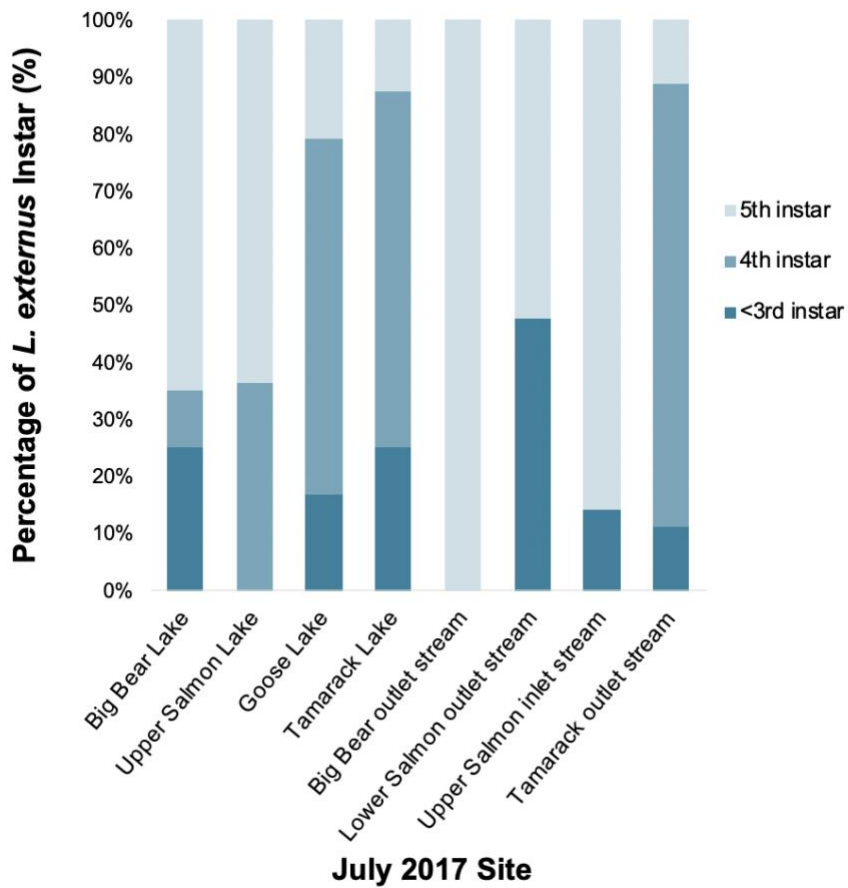
645

646

647

648

649



650 **Figure 5.** Proportion of *Limnephilus externus* individuals each instar per site in July 2017.

651

652

653

654

655

656

657

658

659

660

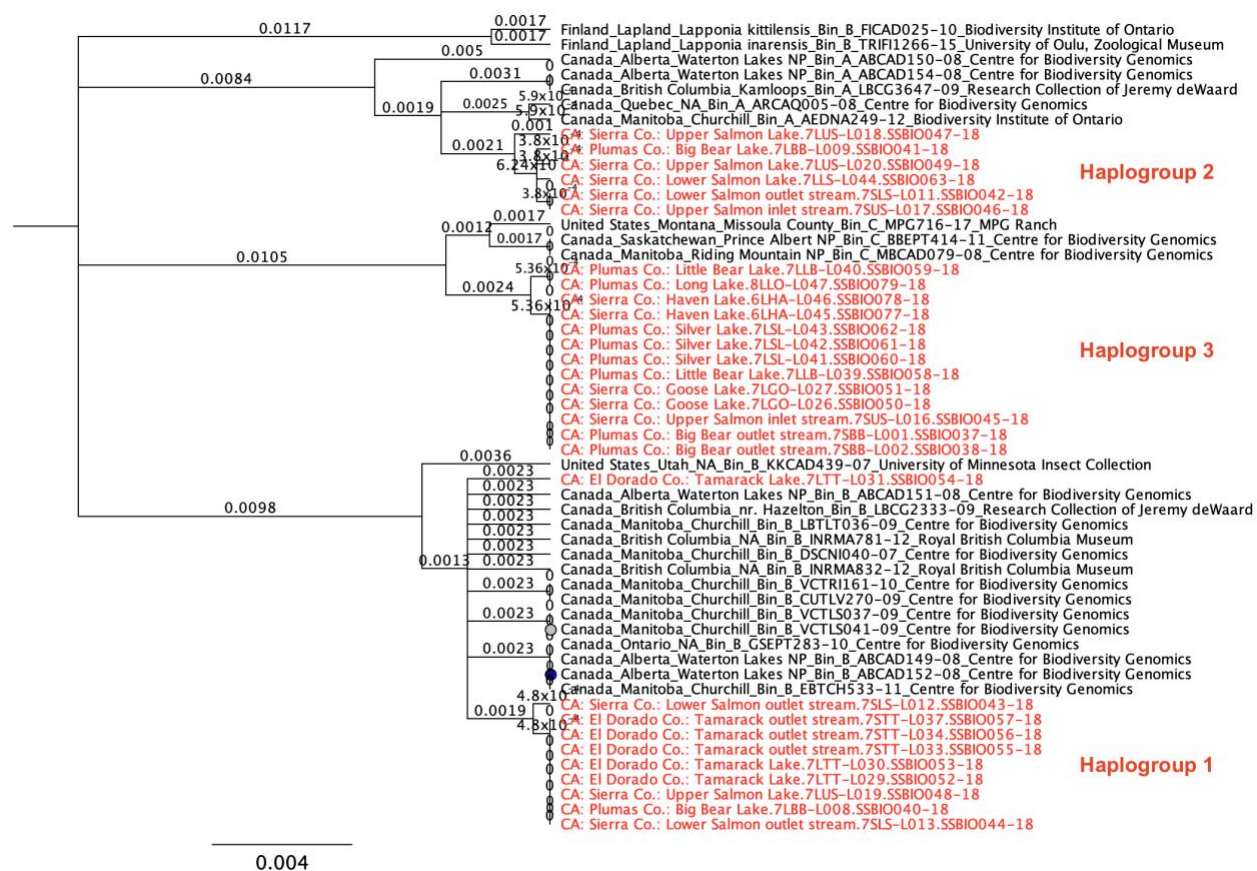
661

662

663



664



665

666

667

668 **Figure 6.** Phylogenetic tree of *Limnephilus externus* in the Sierra Nevada (red) using  
 669 mitochondrial COI gene data. Additional individuals throughout the United States, Canada, and  
 670 Finland (black) were included from publicly available data on BOLDSYSTEMS for contextual  
 671 support. Each imported specimen name includes: country, state or province, county (if available),  
 672 BOLDSYSTEMS BIN cluster (A, B, C), specimen ID, and voucher specimen location.  
 673 Haplogroups identified in the Sierra Nevada are in the order Two, Three, and One, from top to  
 674 bottom. The number of substitutions per site (number on horizontal branch) represents the  
 675 difference two parallel branches are from one another. Here three haplogroups exhibit a  
 676 minimum 0.01 = 1% additive difference from each other; a roughly 2-3% additive difference  
 677 between two parallel branches would be required for two halpogroups to be considered a distinct  
 678 species. Phylogenetic tree constructed using the Jukes-Cantor genetic distance model.

679

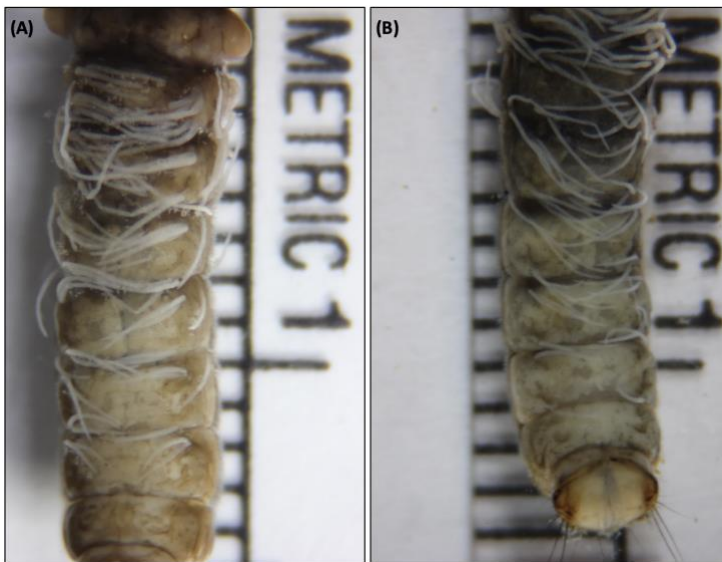
680

681

682  
683

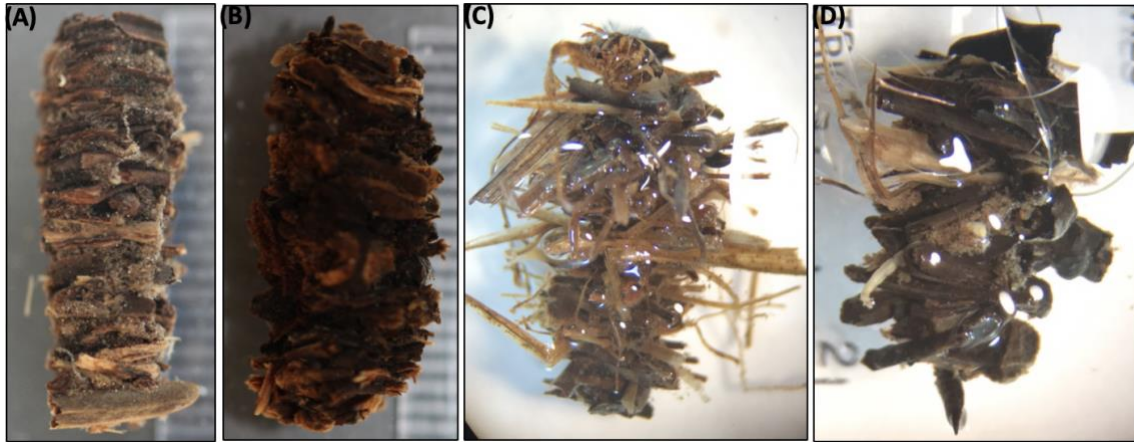


684 **Figure 7.** Ventral view of *Limnephilus externus* abdomen exhibiting (A) no spotting, even color  
685 tone, and robust appearance, and (B) black spotted, increased abdomen transparency resulting in  
686 more visible tracheae, splotchy color tone, and attenuated gill appearance.  
687



688  
689 **Figure 8.** Ventral view of *Limnephilus externus* abdomen with (A) thick and (B) thin gills.

690

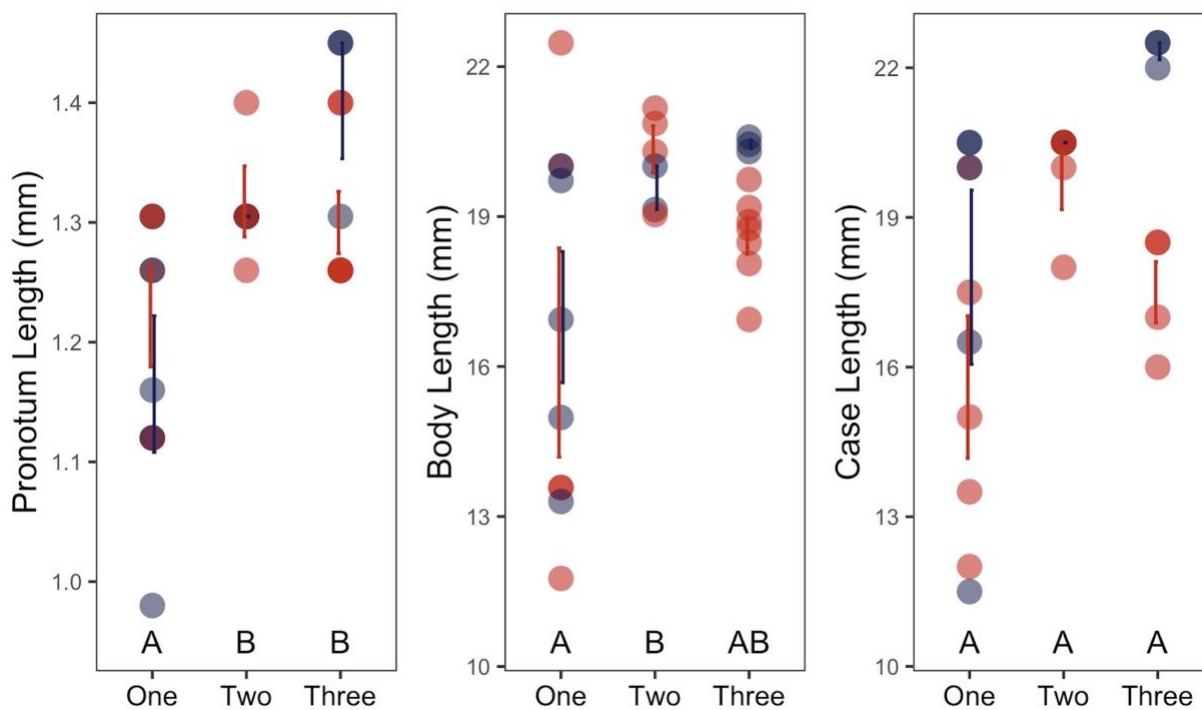


691

692 **Figure 9.** Variation in case types of *Limnephilus externus*: (A) narrow and sturdy, (B) bulky with  
693 twigs, (C) bulky with softer vegetation, (D) fragile with lateral extensions. Regional differences  
694 can be seen between Lakes Basin (A-C) and Tamarack (D). A midge can be seen embedded in  
695 the case in the lower left of D.

696

697



698

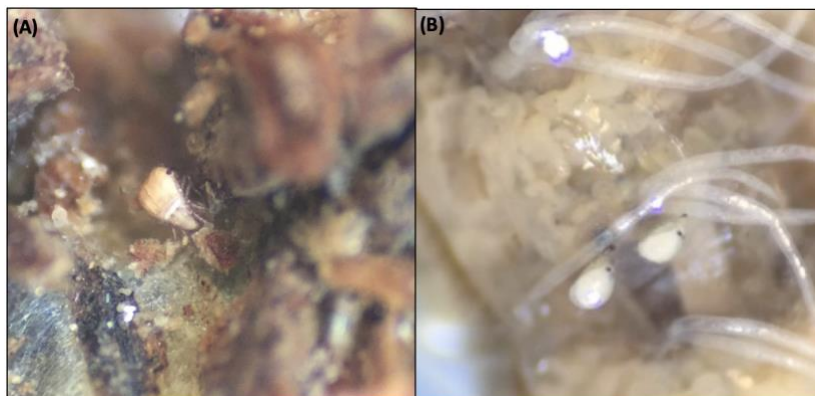
699

700 **Figure 10.** Differences among haplogroups (i.e., One, Two, Three) for each significant response  
 701 variable (i.e., pronotum length, body length, case length). Tukey's post-hoc test results are  
 702 represented above the x-axis. Lake (red) and stream (blue) individuals are illustrated. Error bars  
 703 represent the standard error of the mean for lake vs stream individuals in each haplogroup.

704

705

706



707

708 **Figure 11.** Acari (water mites) on *Limnephilus externus*' (A) case exterior and (B) abdomen.

709

710

711

712 **Table 1.** Qualitative data collected on body and case morphology.

Body	0	1
Abdominal condition	Robust appearance, even color tone, no spotting	Transparent (visible tracheae), black spotted, and attenuated gills
Gill length	Does not cross midline	Crosses ventral midline
Gill thickness	Thin	Thick
Posterior extension of head capsule pigmentation	Does not extend along coronal suture	Extends along coronal suture
Abdominal mites	Absent	Present
Case	0	1
Shape	Straight	Bulged
Presence of silt	Absent	Present in crevices
Primary material type	Bark	Soft aquatic vegetation
Structure sturdiness	Breaking/Fragile	Relatively strong/sturdy
Length of case material pieces	Short	Long
Lateral Extensions	Absent	Present
Assembly uniformity	Uniform	Variable
Microinvertebrate hitchhikers	Absent	Present

713

714

715

716 **Table 2.** Number of individuals collected per month per site during timed sampling.

Site	June 2017	July 2017
Silver Lake	25	18
Little Bear Lake	0	20
Big Bear Lake	19	20
Upper Salmon Lake	1	20
Lower Salmon Lake	23	0
Goose Lake	82	20
Tamarack Lake	NA	8
Upper Salmon inlet	0	7
Salmon Creek	0	1
Lower Salmon outlet	0	20
Silver outlet	0	0
Little Bear outlet	0	0
Big Bear outlet	0	10
Tamarack outlet	NA	9