

1 Caddisfly co-occurs in lakes & streams

2 **Morphological plasticity in a caddisfly that co-occurs in lakes and streams**

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4 Christine A. Parisek^{1*}, Michael P. Marchetti², Matthew R. Cover¹

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6 ¹ *Department of Biological Sciences, California State University Stanislaus, Turlock, CA 95382*

7 ² *Department of Biology, Saint Mary's College of California, Moraga, CA 94575*

8 * *Corresponding author: caparisek@ucdavis.edu; Present Address: Department of Wildlife,*

9 *Fish, & Conservation Biology, University of California Davis, Davis, CA 95616*

10

11 **ORCID**

12 Christine Parisek – <https://orcid.org/0000-0002-7648-879X>

13 Michael Marchetti – <https://orcid.org/0000-0001-8574-6802>

14 Matthew Cover – <https://orcid.org/0000-0003-3315-1027>

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16 **Competing Interests Statement:** The authors have no competing interests to disclose.

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18 **Data Accessibility Statement:** Specimen COI sequences are publicly accessible in the Barcode

19 of Life Data Systems (<http://dx.doi.org/10.5883/DS-LIMNEPH>). Morphological data and color

20 figures are provided in the supplement.

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22 **Preprint Server:** EcoEvoRxiv, CC-BY SA 4.0, <https://doi.org/10.32942/X24S3X>

23 **This PDF file includes:** Main Text; Figure legends 1-11; Tables 1-2

24 **ABSTRACT**

25 Lake and stream fauna are frequently studied, yet surprisingly little is known about ecological
26 and evolutionary dynamics of species that inhabit both lentic and lotic habitats. There are few
27 examples of species co-occurring in the different habitat flow types, which raises questions on
28 how this may impact their ability to adapt to changing climatic conditions. The aquatic insect
29 *Limnephilus externus* Hagen (Trichoptera: Limnephilidae) is widely distributed in lakes of the
30 Nearctic and Palearctic regions; in our study area of the northern Sierra Nevada mountains
31 (California, USA), larval stages of this species co-occur in connected lakes and streams. We
32 examined larval body and case morphology, interspecies phoretic associations, and the
33 mitochondrial DNA cytochrome *c* oxidase I (COI) gene among lake and stream populations of *L.*
34 *externus*. Further, we begin to explore potential morphological differences in distinct *L. externus*
35 haplogroups. We observed differences between lake and stream populations in abundance,
36 phenology, some aspects of body and case morphology, and abdominal mite presence, indicating
37 that lakes and streams may yield distinct ecological phenotypes for the species. We also
38 observed distinct regional differences in caddisfly body condition and sturdiness of case
39 construction, as well as distinct communities of micro-invertebrates associated with the caddisfly
40 and cases. Lake-stream *L. externus* did not show genetic divergence; however, three potentially
41 distinct haplogroups were present across the research sites, as well as in sequences from North
42 America and Canada which were imported from BOLDSYSTEMS. *Limnephilus externus*
43 appears to exhibit wide geographic range and low geographic sequence structure which could
44 account for the species' large variation in phenology and morphology at the lake-stream level.
45 As the Sierra Nevada faces warming temperatures, reduced snowpack, and flow cessation,
46 sensitive high elevation species will face potentially detrimental consequences. Aquatic insect

47 life history and phylogenetic structure provides valuable insight into the ecological and
48 evolutionary dynamics that influence the adaptability of aquatic fauna to climatic change.

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50 **Keywords**

51 Lentic-Lotic, Aquatic insect, DNA barcoding, Phenology, Morphology, Phoresy

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INTRODUCTION

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Lentic and lotic habitats are believed to differentially influence ecological and evolutionary dynamics. Indeed the distinction between these two hydraulic habitat types has been fundamental to the classification of aquatic ecosystems and has strongly influenced the way freshwater scientists conduct research and organize their disciplines (Wetzel 2001, Lottig et al. 2011, Allan et al. 2021). In riverine systems, mechanisms of upstream dispersal are a necessity for plant and animal species (Wubs et al. 2016), while dendritic network patterns create variation in metacommunities among headwater and mainstem habitats (Brown and Swan 2010). Lakes are commonly understood to favor greater dispersal traits, possibly because they are less stable over evolutionary timescales relevant to speciation. For example, lentic odonate species have larger latitudinal ranges than lotic species in the Nearctic and Palearctic (Hof et al. 2006). While some studies have found lotic species have greater genetic population differentiation and potential for cryptic diversity (Marten et al. 2006), this may not always be the case (Ribera et al. 2001).

There are few theoretical and empirical examples of studies on the ecological and evolutionary dynamics of individuals that can co-occur in both lentic and lotic habitats. The best examples of lake-stream eco-evolutionary comparisons thus far have come from fishes, especially work on three-spined stickleback (*Gasterosteus aculeatus* Linnaeus). In sticklebacks, co-occurrence seems possible due to morphologic variability and/or parapatric speciation (Thompson et al. 1997, Rennison et al. 2019, Paccard et al. 2020). Interestingly, in a case study transplanting lake-genotyped sticklebacks into streams, survival of lake-genotype fishes was poor and individuals with a hybrid lake-stream genotype had only moderately improved survival (Moser et al. 2016). In another case, freshwater drum (*Aplodinotus grunniens* Rafinesque)

75 exhibited more robust bodies in rivers and reservoirs with lower retention time (more flow), yet
76 interestingly this species can show amenability to both lentic and lotic habitats beyond the age of
77 ~12 years (Rypel et al. 2006). Minnows (*Phoxinus* spp. Rafinesque) from lakes and streams
78 often also exhibit a similar morphologic pattern, though some evidence to the contrary suggests
79 that in minnows this may be region-dependent (Ramler et al. 2017, Scharnweber 2020).

80 Species that co-occur in lotic and lentic systems may be especially common in high
81 altitude, glaciated mountain landscapes, where lakes are often hydrologically linked in chains by
82 stream segments. High mountain lakes and streams are often oligotrophic, and wave action along
83 rocky littoral zones of lakes produces microhabitats that can resemble headwater streams (Merritt
84 and Cummins 1996, Baker et al. 2016). Stream-dwelling invertebrates have been observed to live
85 in the inlet and outlet regions of high elevation lakes (Wissinger et al. 2016), yet the ecological
86 and evolutionary dynamics of populations of aquatic organisms that co-occur in these
87 mountainous lake and stream habitats remains poorly understood. Clarifying lentic-lotic
88 population dynamics, especially in sensitive mountain ecoregions, would provide a basis to
89 assess ecological and evolutionary behaviors of aquatic organisms and how these may alter in
90 future climate change scenarios.

91 Here, we test whether populations of the caddisfly *Limnephilus externus* Hagen
92 (Trichoptera: Limnephilidae) that co-occur in lakes and streams are evolutionarily and/or
93 ecologically distinct. Specifically, we compare population genetic structure, abundance, larval
94 phenology, larval body and case morphology, and interspecies phoretic interactions between
95 lentic and lotic populations of *L. externus*. We follow this with a brief examination of
96 morphologic differences between the three distinct *L. externus* haplogroups that emerged from
97 this analysis.

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METHODS

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Study Organism

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Study Area and Sampling

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Limnephilus externus Hagen (Trichoptera: Limnephilidae) is a caddisfly whose larvae typically inhabit lentic habitats, such as lakes, permanent to semi-permanent shallow ponds, and wetlands (**Figure 1**) (Berté and Pritchard 1986, Wissinger et al. 2003, Jannot et al. 2008). The five larval instars and the pupa are aquatic; after pupation *L. externus* emerge as terrestrial winged adults (**Figure 2**). *Limnephilus externus* larvae create bulky cylindrical non-rigid cases (“hedgehog cases”; Johansson and Johansson 1992) assembled from fragments of vegetation, detritus, and other organic matter (Berté and Pritchard 1986, Wiggins 2004). While *L. externus* flight duration is not well documented, adults of this species likely live less than two months (Berté and Pritchard 1986, Wissinger et al. 2003). *Limnephilus externus* is well documented in lake habitats throughout the western North America, Canada, and the Palearctic (Morse 1993, Ruitter et al. 2013, Mendez et al. 2019). There are very few records of larvae of *Limnephilus* spp. in streams; in California *Limnephilus* spp. is widely known from lakes and outside of this study we are only aware of several documented stream site records (Pratha 2014).

Sampling occurred in two regions in the northern Sierra Nevada mountain range, California, USA. The Lakes Basin, in the northern Sierra Nevada, is a high elevation (2000 m) mountain region featuring a dendritic network of headwater streams and oligotrophic lakes. Six lakes and six streams of close proximity were selected from > 20 glacial lakes and their connecting streams (**Figure 3**). These lakes occur in the headwaters of two adjoining watersheds:

121 the Feather River (Silver, Little Bear, Big Bear, and Goose lakes) and the Yuba River (Upper
122 and Lower Salmon lakes). To add context to the study, we also sampled *L. externus* populations
123 from one additional lake (without inlet or outlet stream) in the Lakes Basin, Haven Lake (Feather
124 River watershed), as well as a lake-stream pair in a second region ~100km south (Tamarack
125 Lake and outlet stream, Upper Truckee River watershed); these contextual samples were only
126 used in phenology and population genetics analyses.

127 In the winter preceding this study (2016 – 2017), California experienced above average
128 rainfall and snowpack and thus above average streamflow (Guirguis et al. 2019). The first
129 sampling event in late June 2017 occurred during peak snowmelt and streamflow. A second
130 sampling event in July 2017 occurred after peak water levels had subsided.

131 Water quality parameters, measured as spot samples during population and habitat
132 surveys, were similar across all lake and stream study sites and typical of water quality in the
133 higher elevations of the Sierra Nevada mountains. Conductivity was consistently below 25
134 $\mu\text{S}/\text{cm}$, while pH in both lakes and streams was neutral (pH 6.1-7.6). Dissolved oxygen levels
135 were typically near saturation (70-90%), with lower values occurring during early morning
136 hours, reflecting some moderate diurnal fluctuations. Water temperatures were similar among
137 lakes and streams, and were higher, on average, in July (20.5 °C) than June (18.0 °C).

138 Sampling for *L. externus* took place in both lotic (stream habitats within 100 m of lake
139 outlets or inlets) and lentic (lake habitats at least 100 m from the nearest inlet or outlet) habitats.
140 Five 1 m² sampling areas were selected along the littoral zone of lakes and the benthic zones of
141 streams in water depths of 5-50 cm. Sampling areas were spaced at least 1 m apart. Population
142 surveys were performed for a timed interval (12 minutes per 1 m² area) by sampling a
143 combination of cobble, boulder, and bedrock. At each site we examined and picked up 100-125

144 cobble-sized rocks to document the abundance of *L. externus*. All collected *L. externus* were
145 preserved in 70% ethanol and taken to the lab for further analysis.

146 In the lab, *L. externus* larvae were roughly sorted into 5 instars based on case size. All
147 subsequent analyses were performed using only individuals of the largest size class (presumed
148 5th instar). *A posteriori* measurements of head capsule width of the largest size class (mean 1.57
149 mm) were similar to ranges for 5th instar *L. externus* larvae reported in other studies (mean 1.62
150 mm, Berté and Pritchard 1986; mean 1.60 mm, Wissinger et al. 2003).

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152 **Population Genetics**

153 We examined genetic variation among sampled *L. externus* populations through
154 sequencing and analysis of the mitochondrial cytochrome *c* oxidase subunit 1 (COI) gene. We
155 removed a single leg from twenty-nine individuals and placed each leg in a unique microplate
156 well with 1-2 drops of 70% ethanol. Samples were sent to the Canadian Center for DNA
157 Barcoding at the University of Guelph for standard DNA extraction, mtDNA COI gene isolation,
158 and gene amplification, with established QA/QC standards. Forward primer C_LepFolF and
159 reverse primer C_LepFolR were used to conduct PCR amplification on the marker COI-5P, a
160 standard insect barcoding marker.

161 To visualize haplotype relationships and their spatial distribution for the Lakes Basin and
162 Tamarack specimens, a haplotype network was created using the Minimum Spanning Network
163 algorithm on twenty-eight sequences that had been trimmed to 633 base pairs with no missing
164 data (Bandelt et al. 1999, Posada and Crandall 2001). To further understand these relationships
165 and to examine genetic variation in our populations in the context of populations collected
166 elsewhere, we aligned and compared the returned COI sequences to those found in

167 BOLDSYSTEMS (Ratnasingham and Hebert 2007). We searched the BOLDSYSTEMS Public
168 Data Portal for nucleotide sequences belonging to “*Limnephilus externus*” and exported all 252
169 matching records and their metadata; data came from ten institutions, spanned three countries,
170 and broke into three Barcode Index Number (BIN) clusters (i.e., algorithm-generated operational
171 taxonomic units that are performed once per month based on diverging sequences)
172 (Ratnasingham and Hebert 2013). We removed twenty-four sequences without BIN information
173 and forty sequences with invalid residues. All sequences were aligned using a global alignment
174 with free end gaps and a 65% similarity cost matrix. Additional sequences were removed if they
175 showed many gaps in the nucleotide alignment, were too short relative to the other aligned
176 sequences, or were of duplicate locations with identical (or nearly identical (< 0.002)) sequences
177 congregated within the same haplogroup branch. The final nucleotide alignment comprised
178 twenty-nine original sequences and twenty-five unique BOLDSYSTEMS sequences which may
179 be found in **Dataset S1**.

180 Phylogenetic trees of the fifty-four COI sequences were constructed using both a
181 distance-matrix method (UPGMA) and a Bayesian inference method (MrBayes version 3.2.6).
182 We built trees using UPGMA for three different pairwise genetic distance models (i.e., *Jukes-*
183 *Cantor*, *HKY*, *Tamura-Nei*) using a bootstrap resampling method (100 replicates). Bayesian
184 analyses used both the *JC69* (nst = 1) and *HKY85* (nst = 2) substitution models (Huelsenbeck
185 and Ronquist 2001). We selected the only two imported BOLDSYSTEMS sequences available
186 from the Palearctic (Finland) as outgroups. All trees produced with the UPGMA and Bayesian
187 analyses contained similar distinct clades and haplogroups, thus we only present results from the
188 UPGMA Jukes-Cantor model that assumes equal rates of nucleotide substitutions as an inferred
189 phylogenetic relationship. Algorithm-generated BIN assignments from BOLDSYSTEMS are

190 included in the branch label of exported BOLD sequences. We identified haplogroups using a
191 criteria of ≥ 0.01 (1%) dissimilarity between parallel branches that resulted in substantially larger
192 variation between groups than within groups.

193 Initial sequence metadata review was performed in R (version 4.2.0, R Core Team 2022).
194 Haplotype networks were created using using PopART (Population Analysis with Reticulate
195 Trees; Leigh and Bryant 2015). Nucleotide sequence alignments and phylogenetic tree
196 construction used Geneious software (version 10.2.3). Original COI sequence data from this
197 study are publicly accessible in the Barcode of Life Data Systems ([http://dx.doi.org/10.5883/DS-](http://dx.doi.org/10.5883/DS-LIMNEPH)
198 [LIMNEPH](http://dx.doi.org/10.5883/DS-LIMNEPH)).

199 Morphology and Phoretic Associations

200 *Limnephilus externus* fifth-instar larvae (n = 44; 27 lake, 17 stream) and their associated
201 cases were individually photographed, given unique identification codes, and examined under a
202 dissection microscope at 10-20x magnification. Each individual was measured for head-capsule-
203 width (HCW), body length, pronotum length, body width at both the pronotum and second
204 abdominal segment, case length, and case width at its widest point. Body morphology was
205 measured using a micrometer (± 0.01 mm) and case morphology was measured using calipers (\pm
206 0.1 mm).

207 We also qualitatively documented the following body and case morphologic features for
208 each collected individual: abdominal condition, gill length, gill thickness, head capsule
209 pigmentation, abdominal mites, case width type, presence of silt in the case, case material type,
210 case sturdiness or fragility, case material length, lateral case extensions (Limm and Power 2011),
211 case assembly uniformity, and case microinvertebrate hitchhikers (**Table 1**). Two distinct
212 conditions of the ventral abdomen were also observed: even color tone with robust appearance,

213 and black spotting with a transparent cuticle. Finally, a variety of microinvertebrates (< 500 μm ;
214 e.g., Chironomidae, Acari, Oligochaeta, *Hydra*) were found attached to or embedded in
215 caddisfly cases, or clinging to abdominal gills from within the case. These associated
216 microinvertebrate taxa were coarsely identified, enumerated, and preserved in 70% ethanol.

217 To examine possible differences between lake and stream populations, we performed
218 two-tailed t-tests assuming equal variance for the seven quantitative variables, and Fisher's exact
219 tests of independence on the qualitative nominal data. We also performed two-way ANOVAs
220 and Tukey's post-hoc tests to determine differences in the same seven quantitative variables
221 among the three haplogroups identified in phylogenetic analyses, and Fisher's exact tests were
222 used for qualitative differences among haplogroups (n = 26; 10 in group one; 6 in group two; 10
223 in group three). Six sequenced specimens did not have gill health criteria available. We report all
224 p-values ≤ 0.05 . Analyses and mapping were performed in R (R version 4.2.0, R Core Team
225 2022). Quantitative and qualitative data are provided with open access as part of this manuscript
226 (**Dataset S2**). See supplemental material (**Figures S3**) for colored versions of all figures.

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RESULTS

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Distribution and population genetic structure

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Limnephilus externus was widely distributed in both lakes and streams but was more abundant in lakes. Although *L. externus* is known primarily as a lake-dwelling caddisfly, we documented its presence in 7 of 7 lakes and 5 of 7 streams (**Table 2**). We regularly collected twenty individuals per hour at 4 lakes in June (larvae were not observed from Upper Salmon Lake and Little Bear Lake) and at 5 lakes in July (larvae were not observed from Lower Salmon Lake). In contrast, no larvae could be found in the streams experiencing high snowmelt flows in

236 June. In July, one stream (Lower Salmon outlet) yielded at least 20 individuals per hour, while
237 others had lower abundance (< 10 could be attained per hour). Many empty cases were observed
238 in both lakes and streams in July. In lakes, larvae were commonly found on or near submerged
239 vegetation (e.g., aquatic grasses), while in streams larvae were found primarily attached to stable
240 substrates (e.g., fallen logs) in pools. While 5th instar larvae were present among all *L. externus*
241 populations in July 2017, the proportion of instars varied greatly among sites (**Figure 4**). Fifth
242 instars were the dominant size class at Big Bear Lake and Upper Salmon Lake. Lakes had
243 roughly equal proportion of 4th and 5th instars (40.5% and 44.6%, respectively), yet streams had
244 more 5th instars (59.6%) than 4th instars (14.9%). Few of the individuals we collected were 1st-
245 3rd instars (lakes 14.9%, streams 25.5%).

246 Analysis of the mtDNA COI gene sequences indicates moderate intraspecies variation,
247 low geographic structure, and wide geographic distribution of haplogroups. A haplotype network
248 revealed eight unique haplotypes present between the Lakes Basin and Tamarack regions
249 (**Figure 5**). Unique haplotypes are present at the following sites: H1 (Big Bear, Lower Salmon,
250 Upper Salmon, Tamarack), H2 (Lower Salmon, Upper Salmon), H3 (Big Bear, Upper Salmon,
251 Goose, Silver, Haven), H4 (Little Bear, Long), H5 (Big Bear), H6 (Lower Salmon), H7 (Upper
252 Salmon), and H8 (Tamarack). There were no apparent lake-stream differences among haplotype
253 groups. The three largest haplogroups (H1, H2, H3), comprising individuals from both Lakes
254 Basin and Tamarack, also correspond with the three BOLDSYSTEMS algorithm-generated
255 BINS (**Figure 5-6**). The haplotype network supported findings clarified by the phylogenetic
256 analysis, confirming three primary haplogroups with the smaller sized groups nesting into one of
257 these three primary groups on the tree. Within group dissimilarity (haplogroup one: range 0.1-
258 0.5%, mean 0.2%; haplogroup two: 0.1-0.5, mean 0.2%; haplogroup three: range 0.1-0.4%, mean

259 0.2%) was much less than between-group dissimilarity (haplogroups one and two: range 0.8-
260 1.1%, mean 0.9%; haplogroups one and three: range 0.8-1.4%, mean 1.1%; haplogroups two and
261 three: range 0.8-1.2% mean 1.0%). In the phylogenetic approach, all three haplogroups included
262 individuals from both the United States and Canada, indicating that the three genetically distinct
263 haplogroups are widely distributed. The first haplogroup includes multiple individuals from
264 Lakes Basin, all the sampled individuals from the Tamarack study site, as well as individuals
265 collected outside this study from other parts of the Sierra Nevada (Mono County, CA),
266 Washington (USA), and Manitoba (Canada). The second haplogroup includes individuals
267 predominantly from the Upper and Lower Salmon lake watershed and one individual from Big
268 Bear Lake (Lakes Basin), as well as from the Rocky Mountains (Colorado) and individuals from
269 across Canada (Alberta, Manitoba, New Brunswick). The third haplogroup includes individuals
270 from the hydrologically connected system that includes Silver, Little Bear, and Big Bear lakes
271 and streams, as well as nearby Goose and Haven Lakes (Lakes Basin), plus one individual from
272 Manitoba (Canada).

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Morphology

275 *Limnephilus externus* larvae exhibited significant differences in abdomen condition and
276 gill thickness between lake and stream individuals (fisher test, $p = 0.0002$ and $p = 0.0008$,
277 respectively). Black-spotted, transparent abdomens with attenuated gills were more common in
278 lake individuals (100%) than stream individuals (28.6%) (**Figure 7**). Thick abdominal gills were
279 also more common among lake individuals (72.7%) than stream individuals (8.33%) (**Figure 8**).

280 Caddisfly case construction and materials varied substantially among habitats and over
281 time (**Figure 9**). Cases were significantly longer in lakes than streams in the Lakes Basin (t-test,

282 $p = 0.0001$). There were no other significant differences in cases among lake and stream
283 individuals. Cases included more aquatic vegetation in June, while in July cases were
284 constructed predominantly with twigs and bark. All cases from Tamarack Lake and outlet were
285 fragile, bulky, and frequently had lateral extensions made with thin twigs; in contrast, all Lakes
286 Basin *L. externus* cases exhibited stronger construction and no lateral case extensions.

287 Among the three haplogroups, there were significant differences in pronotum length ($F =$
288 6.31 , $p = 0.0068$), body length ($F = 4.64$, $p = 0.0208$), and case length ($F = 4.98$, $p = 0.0183$). A
289 Tukey's post hoc test revealed pronotum length was shorter in haplogroup one compared to
290 haplogroups two-three, and body length was shorter in haplogroup one than haplogroup two;
291 haplogroup three exhibited similarities with haplogroup one and two in different characteristics
292 (**Figure 10**). Head pigmentation and case structure sturdiness also were significantly, or nearly
293 significantly, different across haplogroups (fisher test, $p = 0.0626$ and $p = 0.0287$, respectively).

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Phoretic Associations

297 *Hydra*, nematodes, oligochaetes, chironomid midges (three morphospecies), and water
298 mites were found securely fastened to many caddisfly cases, either stuck to the surface or buried
299 into case silt (**Figure 9D & 11**). These case-associated microinvertebrates were phoretically
300 associated with both lake (36.6%) and stream (50%) caddisfly cases across the three primary
301 haplogroups. We did not observe differences in the microinvertebrate community composition
302 between cases from lake and stream individuals.

303 Abdominal mite presence was significantly different between lake-stream habitats and
304 among haplogroups (fisher test, $p = 0.013$ and $p = 0.0397$, respectively). Abdominal mites were

305 only found on individuals from lakes (40.9%), not streams (0%); however abdominal mite
306 association was only observed at Upper Salmon and Big Bear Lakes which make up haplogroup
307 one and two. The highest abdominal mite association was 31 mites on a single individual;
308 inflicted individuals had a mean of 4 mites. All individuals with water mites on their abdomen
309 were observed to be less robust, had dark and transparent abdomens, and attenuated black
310 spotted gills. However, nearly half of the larvae that lacked water mites at the time of collection
311 also had some of these characteristics.

312 Water mites observed on the exterior of caddisfly cases were identified as adult oribatids
313 (Acariformes: Sarcoptiformes: Oribatida), possibly in the family Trhypochthoniidae or
314 Malaconothridae, while those clinging to the abdomen were identified as larval hygrobatoid
315 water mites (Acariformes: Parasitengona: Hydrachnidia: Hygrobatoidea), possibly in the family
316 Hygrobatidae or Unionicolidae (Heather Proctor, University of Alberta, personal
317 communication).

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DISCUSSION

320 We formally documented the presence of *Limnephilus externus*, a caddisfly widely
321 known from lentic habitats throughout North America, in both lake and stream habitats in the
322 Sierra Nevada. We examined the degree to which *L. externus* occurring in lakes and streams are
323 evolutionarily and ecologically distinct by comparing: (1) population genetic structure, (2)
324 abundance, (3) larval phenology, (4) larval body and case morphology, and (5) interspecies
325 phoretic interactions. Further, we briefly explored the potential for morphological differences
326 between distinct haplogroups of *L. externus*.

327 The use of the mitochondrial cytochrome *c* oxidase I (COI) gene (i.e., DNA barcoding)
328 helps to reveal patterns in biodiversity (Hebert et al. 2003). Studies connecting the techniques of
329 DNA barcoding with traditional taxonomy have increasingly reported higher cryptic diversity
330 than previously suspected (Sheth and Thaker 2017, DeSalle and Goldstein 2019). Indeed in some
331 cases, use of the COI gene has revealed relatively high genetic diversity and low geographic
332 structure in other aquatic insect species (Heilveil and Berlocher 2006, Ståhls and Savolainen
333 2008, Pessino et al. 2014). This study found no genetic differences between lake and stream *L.*
334 *externus*. Instead, we found eight unique haplotypes, at least three of which are geographically
335 widespread and distinct, separated by 1-2% genetic difference. These putative haplogroups may
336 have potential to represent distinct subspecies (White et al. 2014), but likely are not biologically
337 meaningful without additional multilocus data (Dasmahapatra et al. 2010) as a minimum 2-3%
338 genetic divergence is often used to distinguish haplogroups as distinct subspecies or species. We
339 note that the three primary haplogroups in our analysis do match with the three algorithm-
340 generated BINS identified by BOLDSYSTEMS, which are intended to nearly approximate
341 species, suggesting further work would be valuable to explore these relationships. *Limnephilus*
342 *externus*' three primary haplogroups are widely distributed throughout the United States and
343 Canada. Our findings suggest *L. externus* has a wide geographic range and low geographic
344 structure that could support phenotypic plasticity between habitat types, and possibly genotypic
345 and phenotypic variation at the haplogroup level. These results also suggest *L. externus* exhibits
346 potentially high morphological plasticity and may be well adapted to disperse long distances.
347 Relative to other insect species the COI mtDNA gene evolves quickly within the *Limnephilus*
348 genus, supporting its use when exploring recent divergences (McCullagh et al. 2015, Steinke et
349 al. 2022). Our results support and expand on the extensive genetic analysis and findings of *L.*

350 *externus* in the Manitoba province of Canada (Zhou et al. 2011, Ruiter et al. 2013). This rapid
351 evolution of the COI gene, or the widespread gene flow hypothesis, could account for why three
352 widely distributed haplogroups had large variation in morphology and may be found distributed
353 across entire countries. This phylogenetic finding alludes to hidden biodiversity patterns and the
354 need to further identify species boundaries in aquatic insect taxa.

355 Lake and stream populations exhibited distinct ecological phenotypes in abundance,
356 phenology, some aspects of body and case morphology, and abdominal mite presence. Fifth
357 instar *L. externus* were present at all lake-stream sites, while other instars varied in proportion.
358 All lake individuals had abdomens that were transparent (tracheae were visible), black-spotted,
359 and with more attenuated gills, whereas a small fraction of stream individuals had these
360 characteristics. Lake individuals were also observed to have thicker abdominal gills. These
361 morphological differences could represent adaptations resulting from several possible abiotic
362 factors that differ between lakes and streams (e.g., lower levels of dissolved oxygen in lakes).
363 Similarly, gill breadth and visible tracheae have been key factors in distinguishing the lentic
364 *Baetis tracheatus* Keffer Müller & Machel from the lotic *B. bundyae* Lehmkuhl, which has
365 narrow gills and invisible tracheae (i.e., abdomen not transparent) (Engblom 1996, Ståhls and
366 Savolainen 2008). On the other hand, research has linked altered and atrophied tracheal gills
367 (i.e., black speckling) in caddisflies to the introduction of pollutants or bacteria in a headwater
368 stream (Simpson 1980).

369 Lake *L. externus* constructed cases using longer pieces of material than those in streams,
370 and cases from the Tamarack region had weaker construction. Caddisfly case construction is
371 highly dependent on the availability of materials in the surrounding habitat, yet the observed
372 differences in case structure could also reflect adaptations to abiotic or biotic factors (i.e., flow,

373 predator defense). For example, *L. externus*' stout cases are reported to be a better deterrent to
374 predation by beetle larvae than some more tubular cases of other species (Wissinger et al. 2006),
375 while another study reported that differences in case structure between two *Limnephilus* species
376 (*L. pantodapus* McLachlan and *L. rhombicus* Linnaeus) affected the behavior of predaceous
377 dragonfly larvae (Johansson and Johansson 1992). Indeed the construction of more protective
378 cases has been found to be a resource allocation trade-off inducible by predator chemical cues
379 (Correa-Araneda et al. 2017).

380 Across haplogroups, pronotum length, total body length, case sturdiness, and presence of
381 abdominal mites were significantly different between at least two haplogroups. Haplogroups also
382 exhibited a nearly significant difference in head pigmentation, which has previously been used to
383 distinguish between *Limnephilus* species (Ruiter et al. 2013). We consider these morphological
384 haplogroup differences to suggest that real clade-level differences may exist and should be
385 further studied. This study was designed to investigate differences between lakes and streams in
386 one region, and therefore a representative sampling of each haplogroup may not have been
387 achieved.

388 Finally, a collection of microinvertebrates (i.e., chironomid midges, water mites, hydrae,
389 oligochaetes) were discovered buried within *L. externus* cases. In addition, water mites were
390 found on the abdomen of only lake individuals. We observed at least three morphospecies of
391 chironomid midge on the cases, suggesting that the microinvertebrate community on the cases
392 may be diverse. Water mites observed on case exteriors were identified as adult oribatid
393 (Acariformes: Sarcoptiformes: Oribatida) mites, possibly in the family Trhypochthoniidae
394 (Heather Proctor, University of Alberta, personal communication). Oribatids commonly feed on
395 detritus, algae, and occasionally macrophytes (Behan-Pelletier and Hill 1978, Proctor and

396 Pritchard 1989). The association of Oribatid mites on the organic cases suggests a commensal
397 relationship in which the mites could be benefiting by living in or feeding on the cases. The
398 nature of these ecological associations at these locations is not known, however, *L. externus* did
399 not appear to be negatively affected or parasitized by any of the microinvertebrates on the
400 exterior of their cases. Therefore, in these instances, we suspect a phoretic (non-harmful)
401 association. In contrast, mites found on the abdomen of *L. externus* larvae may pose greater
402 threat. Abdominal water mites were identified as larval hygrobatooid water mites (Acariformes:
403 Parasitengona: Hydrachnidiae: Hygrobatoidea), possibly in the family Hygrobatidae or
404 Unionicolidae (Heather Proctor, University of Alberta, personal communication). Hygrobatooid
405 mites are known to engage in pre-parasitic attendance of caddisflies, remaining near the host
406 until it is close to pupation and feeding on it when it emerges as an adult (Proctor and Pritchard
407 1989).

408 The occurrence of phoretic and parasitic relationships is common among aquatic
409 organisms. Other aquatic insects have been documented to play host to midge and water mite
410 travelers in relationships that vary along the gradient of ectoparasitism, predation, and phoresy
411 (Tracy and Hazelwood 1983, Henriques-Oliveira and Nessimian 2009, Buczyńska et al. 2015).
412 In Quebec, Canada, *Limnephilus* has been documented to have water mite larvae
413 (Hygrobatoidea), with prevalence ranging from 4-42% (Fairchild and Lewis 1987). Other aquatic
414 organisms, like the fish *Ancistrus multispinis* Regan in Atlantic forest streams in Southeastern
415 Brazil, have chironomid larvae in phoretic association (Mattos et al. 2018). Understanding the
416 role of associated macroinvertebrates on aquatic organisms is a challenging topic to study;
417 (Grabner 2017) found testing for parasitic taxa using PCR might be an efficient and cost-
418 effective method to identifying links between host feeding type and prevalence. Additional

419 studies would be needed to identify the nature of these associations and their consequences to *L.*
420 *externus*.

421

422 CONCLUSION

423 In this study, we documented the presence of *Limnephilus externus* in both lake and
424 stream habitats. Lake populations had conspicuous abdominal tracheae, thicker gills, and black
425 spotting. Lake populations exhibited longer case construction, and only caddisfly cases from the
426 Tamarack region were significantly more fragile in construction. Microinvertebrate hitchhikers
427 found on the cases of the caddisflies are presumed to maintain a phoretic relationship, while
428 mites on the abdomen may be demonstrating pre-parasitic attendance behavior. Finally, while
429 lake populations were not genetically different from stream populations, we did find eight unique
430 haplotypes present. Of these eight haplotypes, three are distinct and geographically widespread
431 in the Sierra Nevada as well as throughout western North America and Canada. These putative
432 haplogroups exhibited some significant morphological variation but further research is needed to
433 validate these results. Overall, our observations and analyses suggest that environmental
434 differences between lake and stream habitats may produce variation in plastic traits, but dispersal
435 and gene flow are likely preventing genetic differentiation.

436 The frequency of aquatic invertebrate species that co-inhabit lentic and lotic ecosystems
437 is unknown, and reflects the paucity of studies of aquatic fauna across habitat types. Our findings
438 suggest that species with plastic traits amenable to both flow types may be overlooked in aquatic
439 research. As a result, we may be missing valuable information on ecological and evolutionary
440 behaviors of aquatic organisms, especially in light of anticipated climatic changes.

441 While lotic and high elevation lake shoreline habitats have been recognized for their
442 ecological similarities, the way these two distinct ecosystems will respond to climatic changes
443 will be vastly different. Indeed (Wissinger et al. 2016) observed cold-water stream insects
444 inhabiting rocky and wave-swept alpine lake shorelines of Colorado, Switzerland, and New
445 Zealand, and evidence that other freshwater fauna may be amenable to both hydraulic habitat
446 types is growing (Yarnell et al. 2019). Mountain systems in particular face high stressors and are
447 sensitive to environmental changes (Moser et al. 2019). Many of the aquatic habitats in the Sierra
448 Nevada are dependent on snowmelt, yet California's increasingly common drought years and
449 resulting low snowpack are anticipated to decrease snowmelt feeding into aquatic systems (Smits
450 et al. 2020). With deteriorating snowpack and warming lakes, the adaptability of aquatic fauna to
451 find refugia is expected to be a tremendous benefit to their survival (Birrell et al. 2020, Frakes et
452 al. 2021). With this study, we hope to contribute to a larger body of knowledge and facilitate
453 directions for future mountain aquatic research.

454

455

456

AUTHOR CONTRIBUTION STATEMENT

457 CAP led study design, field and lab work, data analyses, and wrote the manuscript. MRC helped
458 design the study, provided methods guidance, helped interpret data, and assisted in manuscript
459 revisions. MPM provided feedback on field and lab procedures and helped interpret data. All
460 authors revised the article critically for important intellectual content and approved the final
461 version to be published.

462

463

ACKNOWLEDGEMENTS

464 CAP was supported by the UC Davis Center for Watershed Sciences' Bechtel Next Generation
465 Funds. This project was supported by the Society for Freshwater Science Boesel-Sanderson
466 Endowment, Sierra Nevada Aquatic Research Lab - Valentine Eastern Sierra grant, and
467 Department of Biological Sciences Biology Research Fund at CSU Stanislaus. We thank John
468 Wheeler for assistance with DNA barcoding and Sean Canfield for generously assisting with the
469 haplotype network analysis. We thank the biologists who contributed to fruitful discussion on
470 *Limnephilidae* during the conceptualization of this project: Scott Wissinger, Dave Ruitter, Dave
471 Herbst, Roland Knapp, Bob Wisseman, Mark Wetzell, Patina Mendez, Jonathan Lee, John Epler,
472 Steve Cairns, Allen Collins, Richard Campbell; Heather Proctor for knowledgeable input and
473 identification of the water mites; and Isaac Chellman (CDFW), Russell Nickerson (FS), and
474 James Matthew Johnson (FS) for support in field research discussion and permissions. Finally,
475 we thank our two anonymous reviewers for their constructive and thoughtful feedback.

476

477

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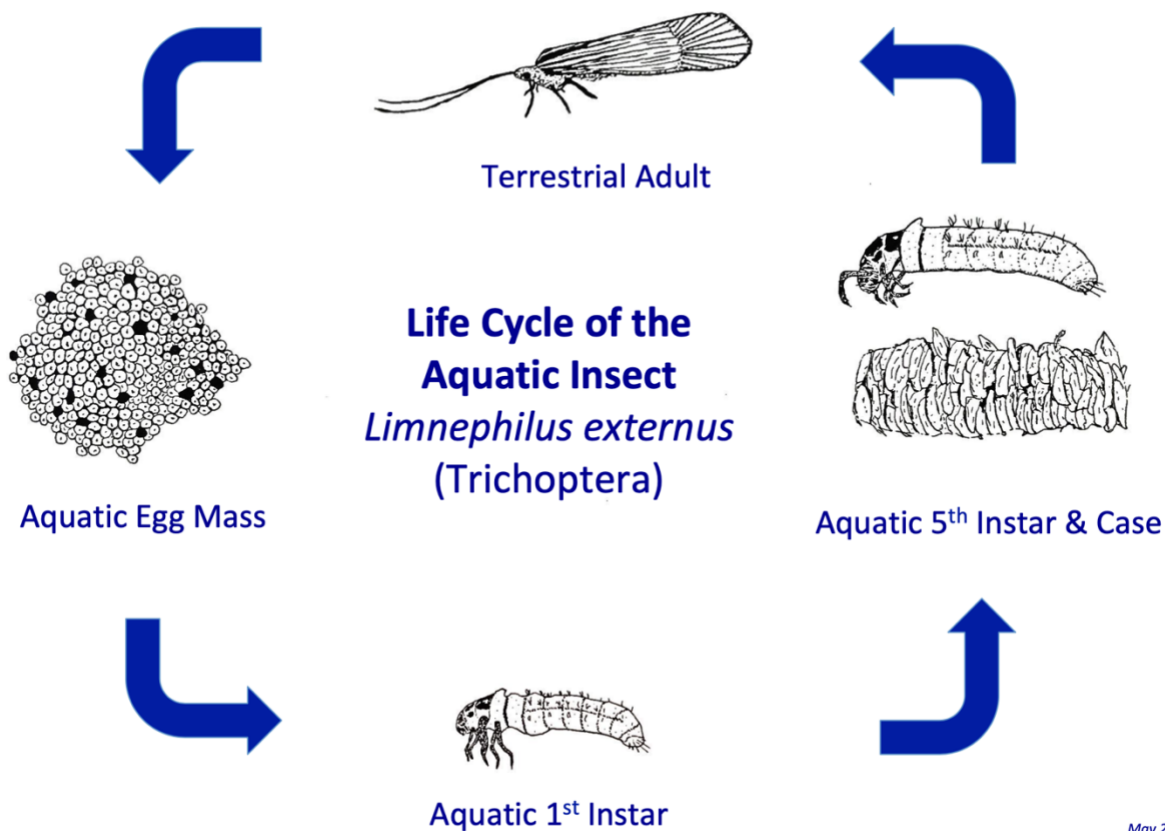
647

FIGURES



648

649 **Figure 1.** Lateral view of *Limnephilus externus* and its case shown overlaying a metric ruler.

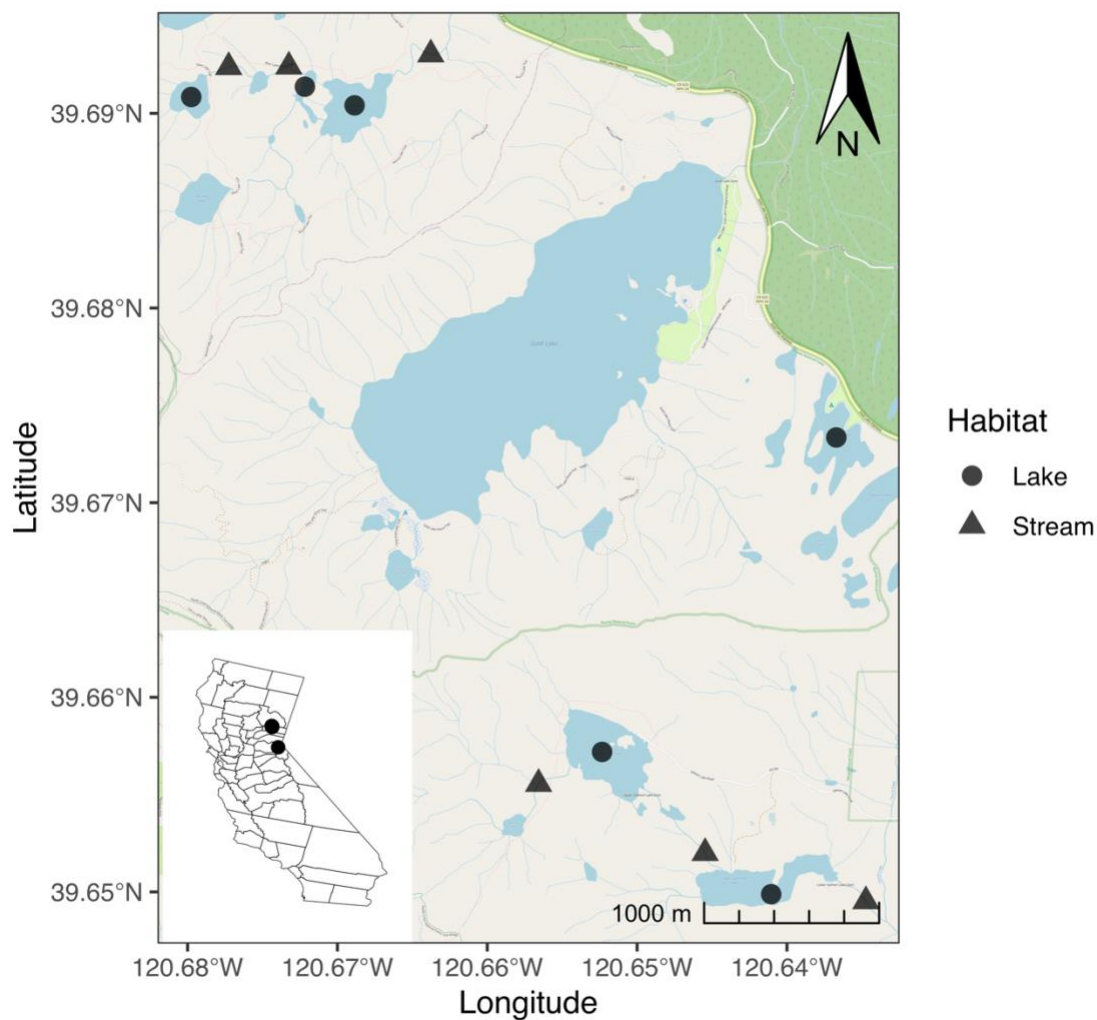


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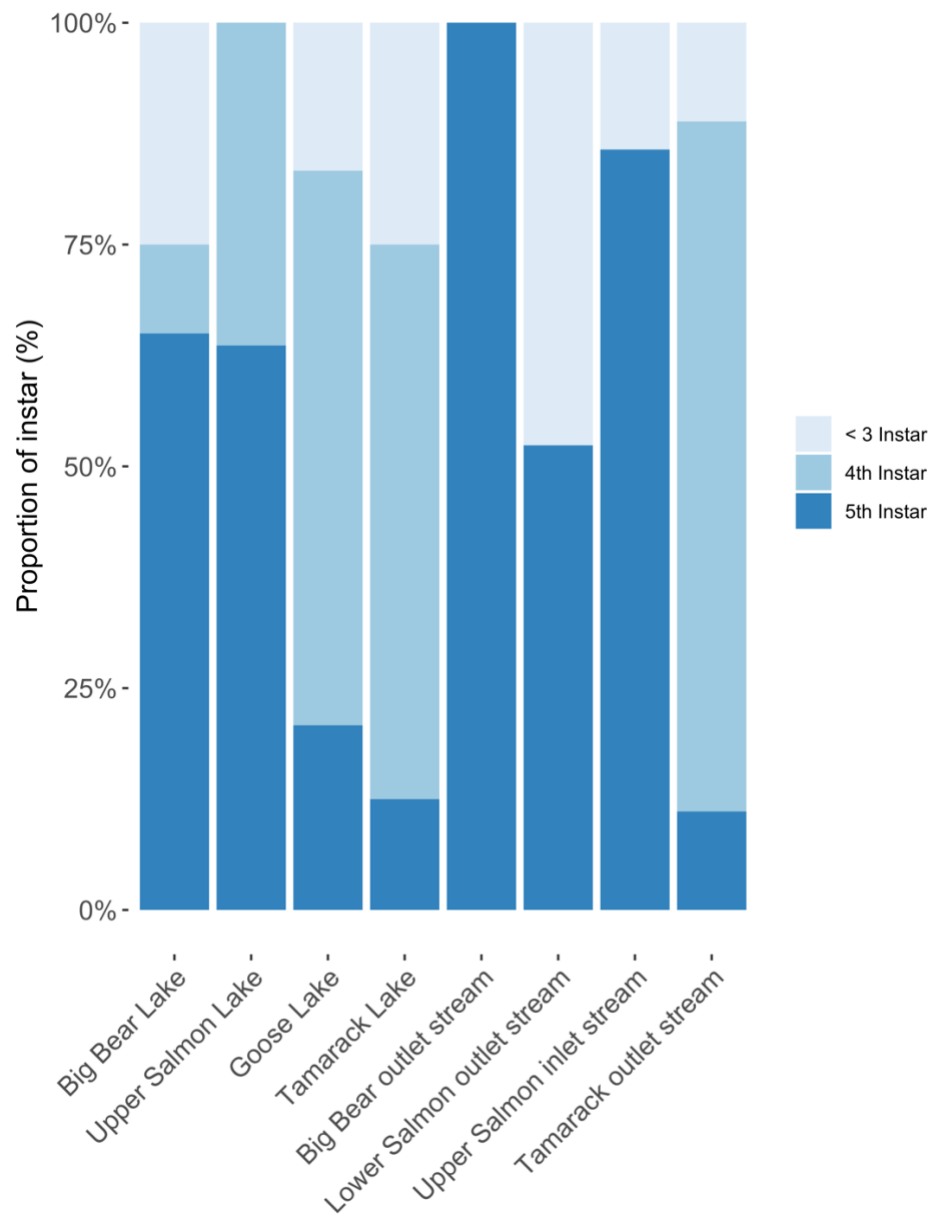
651 **Figure 2.** Illustration summarizing key aspects of the life cycle of the aquatic insect, *Limnephilus*

652 *externus*. Illustration by Christine Parisek.



653
 654 **Figure 3.** Lotic (triangle) and lentic (circle) sampling sites in Lakes Basin, northern Sierra
 655 Nevada, CA. Silver, Little and Big Bear Lakes share connectivity. Upper and Lower Salmon
 656 Lakes share connectivity. Goose Lake has no inlet or outlet stream. Inset map of California, USA
 657 displays primary field location (Lakes Basin) and contextual site (Tamarack Lake) in the Sierra
 658 Nevada mountain range.
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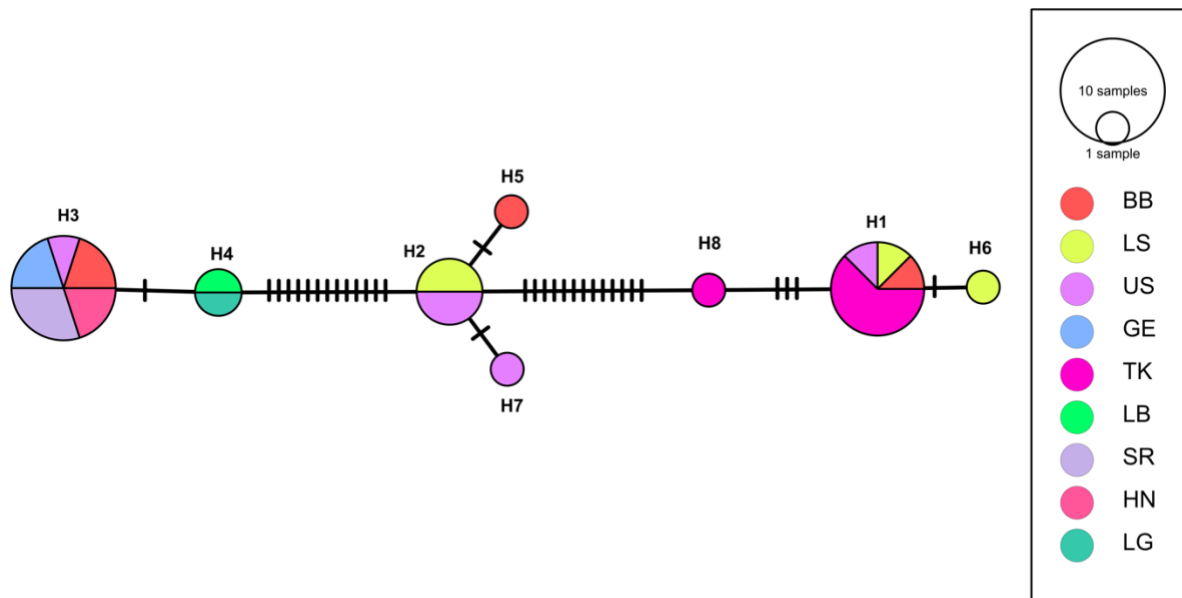


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662 **Figure 4.** Proportion of *Limnephilus externus* individuals of each instar per site in July 2017.

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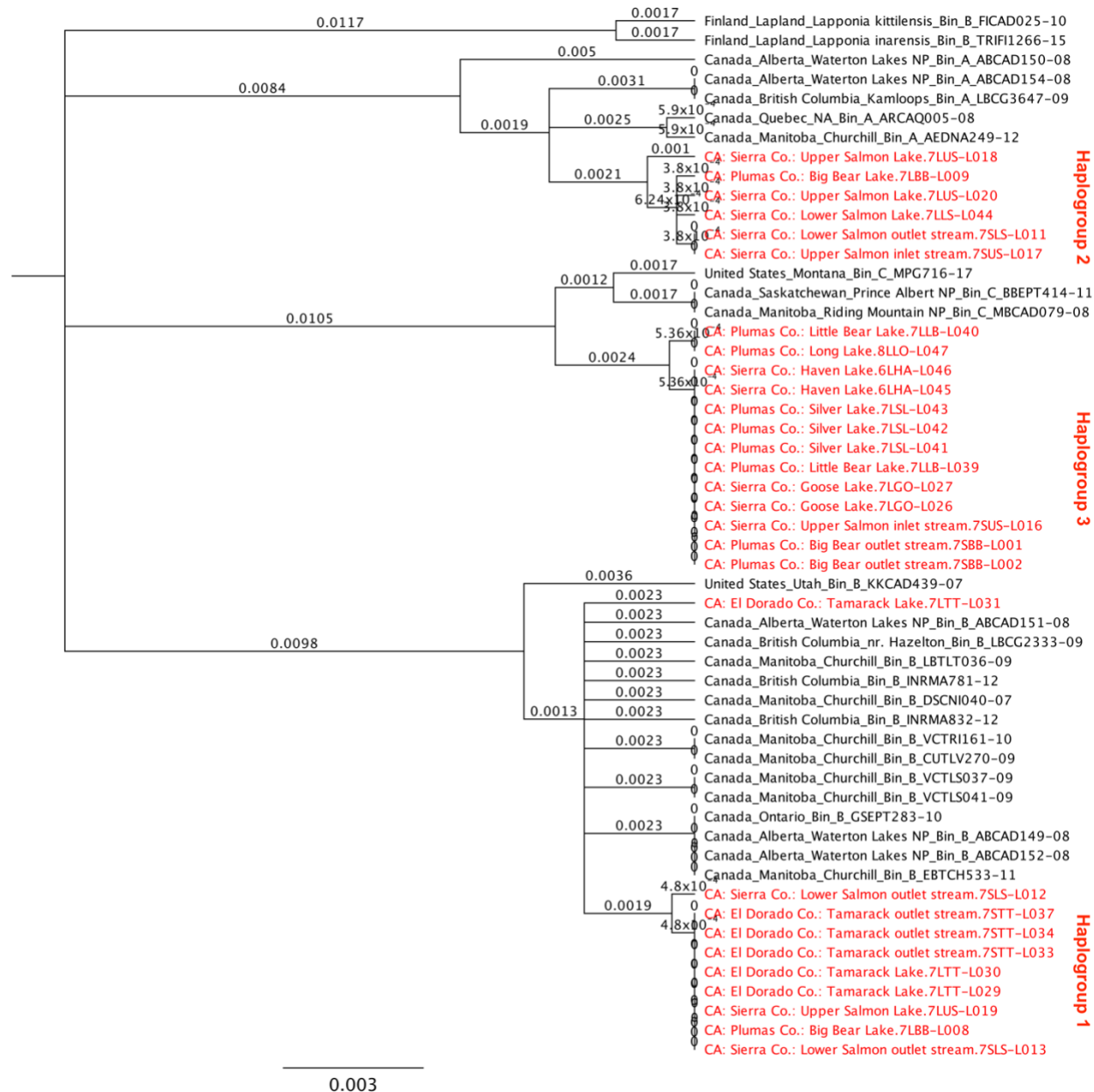


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666 **Figure 5.** Haplotype network showing the number of pairwise differences between groups (1 tick
 667 mark = 1 mutational difference). Eight haplotypes were present in this study. Haplotypes are
 668 present at the following sites: **H1** (Big Bear, Lower Salmon, Upper Salmon, Tamarack), **H2**
 669 (Lower Salmon, Upper Salmon), **H3** (Big Bear, Upper Salmon, Goose, Silver, Haven), **H4**
 670 (Little Bear, Long), **H5** (Big Bear), **H6** (Lower Salmon), **H7** (Upper Salmon), and **H8**
 671 (Tamarack). Site codes are: Big Bear (BB), Lower Salmon (LS), Upper Salmon (US), Goose
 672 (GE), Tamarack (TK), Little Bear (LB), Silver (SR), Haven (HN), Long (LG).

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675

676 **Figure 6.** Phylogenetic tree of *Limnephilus externus* in the Sierra Nevada (light gray) using
 677 mitochondrial COI gene data. Additional individuals throughout the United States, Canada, and
 678 Finland (black) were included from publicly available data on BOLDSYSTEMS for contextual
 679 support. Each imported specimen name includes: country, state or province, county (if available),
 680 BOLDSYSTEMS BIN cluster (A, B, C), and specimen ID. Haplogroups identified in the Sierra

681 Nevada are in the order Two, Three, and One, from top to bottom. The number of substitutions
682 per site (number on horizontal branch) represents the difference two parallel branches are from
683 one another. Here three haplogroups exhibit a minimum $0.01 = 1\%$ additive difference from each
684 other; a roughly 2-3% additive difference between two parallel branches would be required for
685 two haplogroups to be considered a distinct species. Phylogenetic tree constructed using the
686 Jukes-Cantor genetic distance model.

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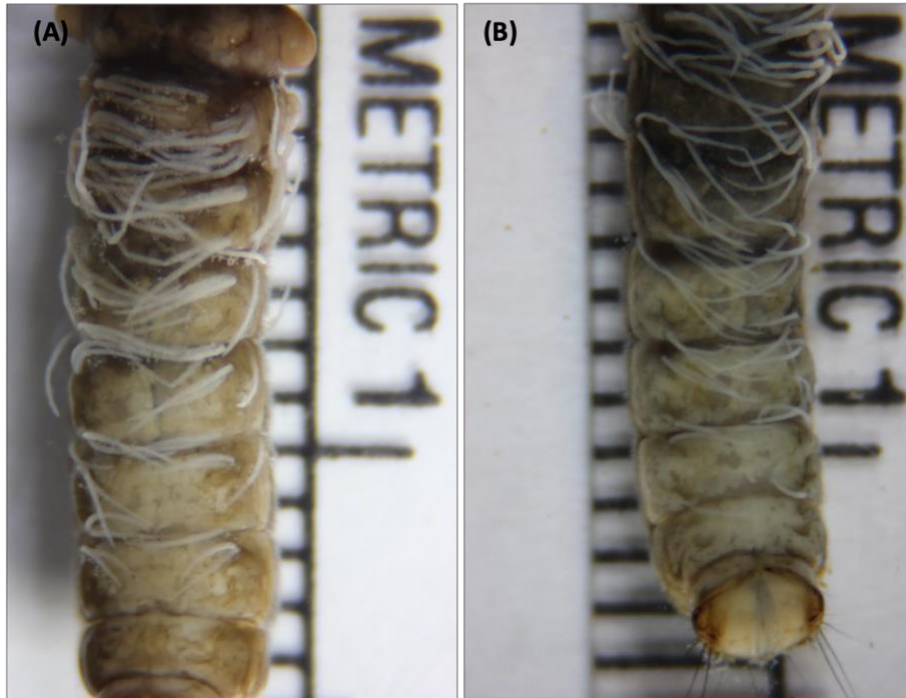
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691

692 **Figure 7.** Ventral view of *Limnephilus externus* abdomen exhibiting (A) no spotting, even color
693 tone, and robust appearance, and (B) black spotted, increased abdomen transparency resulting in
694 more visible tracheae, splotchy color tone, and attenuated gill appearance.

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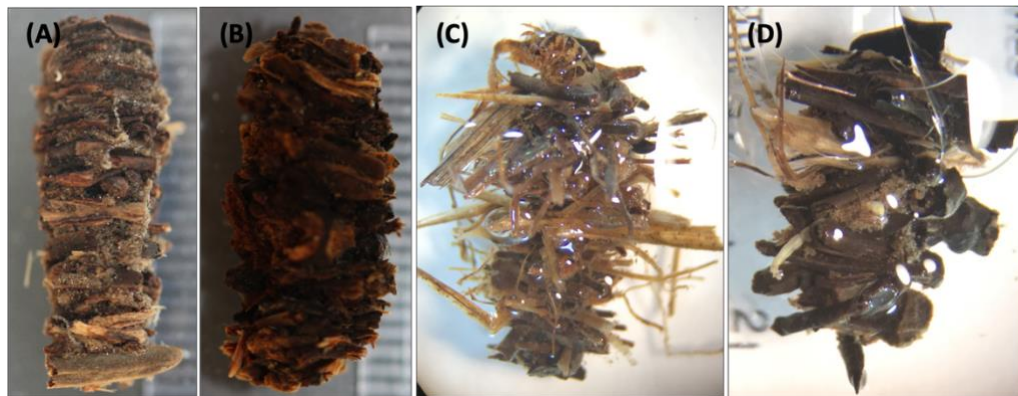


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697 **Figure 8.** Ventral view of *Limnephilus externus* abdomen with (A) thick and (B) thin gills.

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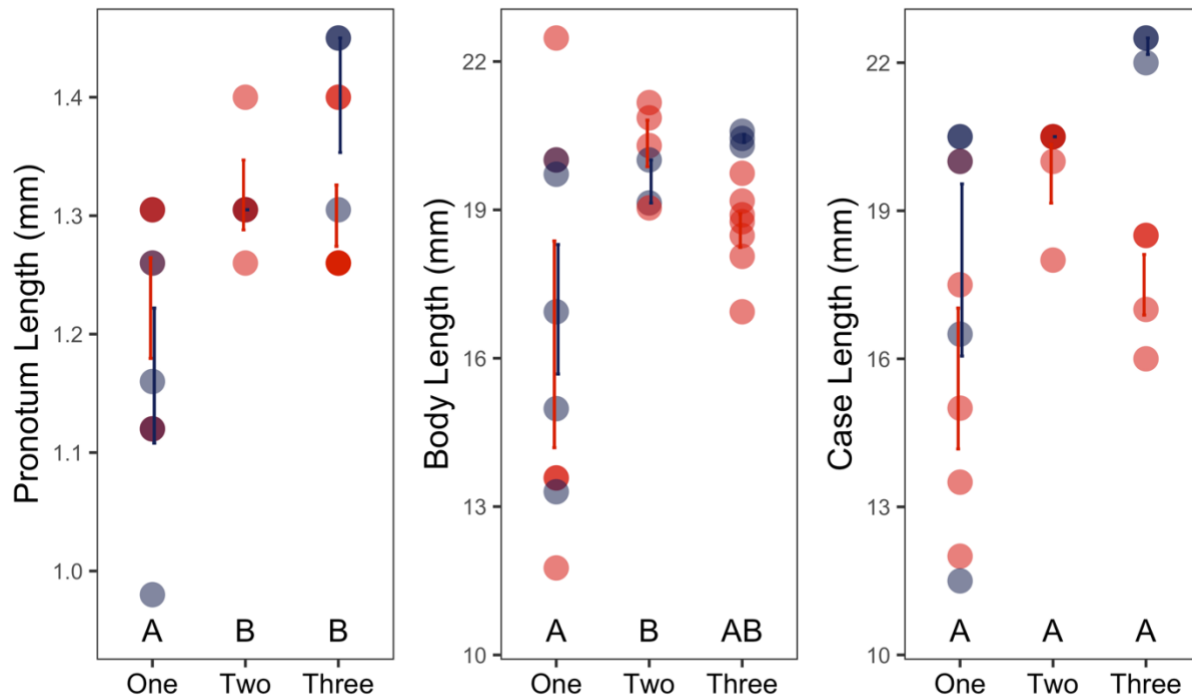
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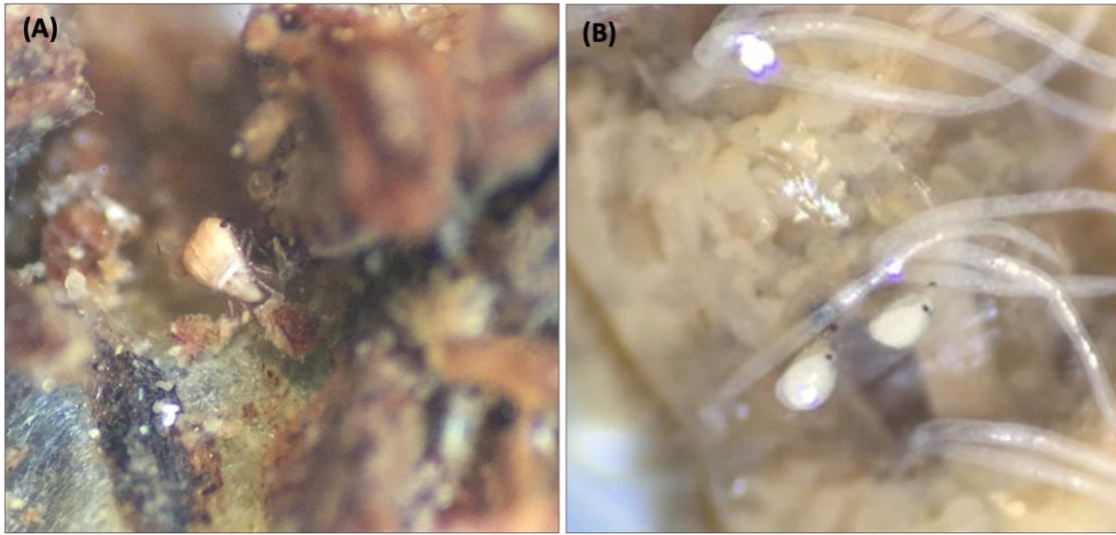
701 **Figure 9.** Variation in case types of *Limnephilus externus*: (A) narrow and sturdy, (B) bulky with
 702 twigs, (C) bulky with softer vegetation, (D) fragile with lateral extensions. Regional differences
 703 can be seen between Lakes Basin (A-C) and Tamarack (D). A midge can be seen embedded in
 704 the case in the lower left of (D).

705



706
 707 **Figure 10.** Differences among haplogroups (i.e., One, Two, Three) for each significant response
 708 variable (i.e., pronotum length, body length, case length). Tukey's post-hoc test results are
 709 represented above the x-axis. Lake (semi-transparent circle) and stream (semi-transparent
 710 triangle) individuals are distinguished. Error bars represent the standard error of the mean for
 711 lake vs stream individuals in each haplogroup.
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715 **Figure 11.** Acari (water mites) on *Limnephilus externus*' (A) case exterior and (B) abdomen.

716 **Table 1.** Qualitative data collected on body and case morphology.

Body	0	1
Abdominal condition	Robust appearance, even color tone, no spotting	Transparent (visible tracheae), black spotted, and attenuated gills
Gill length	Does not cross midline	Crosses ventral midline
Gill thickness	Thin	Thick
Posterior extension of head capsule pigmentation	Does not extend along coronal suture	Extends along coronal suture
Abdominal mites	Absent	Present
Case	0	1
Shape	Straight	Bulged
Presence of silt	Absent	Present in crevices
Primary material type	Bark	Soft aquatic vegetation
Structure sturdiness	Breaking/Fragile	Relatively strong/sturdy
Length of case material pieces	Short	Long
Lateral Extensions	Absent	Present

Assembly uniformity	Uniform	Variable
Microinvertebrate hitchhikers	Absent	Present

717

718 **Table 2.** Number of individuals collected per month per site during timed sampling.

Site	June 2017	July 2017
Silver Lake	25	18
Little Bear Lake	0	20
Big Bear Lake	19	20
Upper Salmon Lake	1	20
Lower Salmon Lake	23	0
Goose Lake	82	20
Tamarack Lake	NA	8
Upper Salmon inlet	0	7
Salmon Creek	0	1
Lower Salmon outlet	0	20
Silver outlet	0	0
Little Bear outlet	0	0
Big Bear outlet	0	10
Tamarack outlet	NA	9