

1 Complexity revealed in the greening of the Arctic

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77 **Abstract**

78 The “greening of the Arctic” is among the world’s most significant large scale ecological  
79 responses to global climate change<sup>1</sup>. The Arctic has warmed at twice the rate of the rest of  
80 the planet on average in recent decades<sup>2</sup> and satellite-derived vegetation indices have  
81 indicated widespread increases in productivity (termed “greening”) at high latitudes<sup>3–8</sup>.  
82 Greening trends have been attributed to *in situ* increases in vegetation biomass, cover and  
83 abundance<sup>9–11</sup> associated with warming trends<sup>12</sup>. Satellite observations allow for the  
84 quantification of vegetation change across northern biomes that are otherwise unevenly  
85 sampled by *in situ* ecological observations<sup>13</sup>. Satellite-derived data thus broadly inform  
86 predictions of large-scale climate feedbacks involving plant biomass, carbon storage, and  
87 surface energy budget<sup>14,15</sup>. Recently however, remotely-sensed Arctic greening trends have  
88 shown periods of slowing or even reversing in some regions (termed “Arctic browning”) <sup>16–19</sup>  
89 sometimes arising from acute declines in productivity<sup>18,20–23</sup> seemingly at odds with earlier  
90 responses to long-term warming trends<sup>24</sup>. Research now indicates substantial diversity in  
91 ecological responses to changing climate regimes in the Arctic<sup>25</sup>, but precise attribution of  
92 patterns and trends to ecological process remains a challenge due to conceptual and  
93 technical barriers in the analysis and combined interpretation of satellite and *in situ*  
94 observations<sup>3,20,26,27</sup>. An emerging consensus is that the underlying causes and future  
95 dynamics of Arctic greening and browning patterns and trends are complex, variable, and  
96 inherently scale dependent. Here, we review the complexities associated with observing and  
97 interpreting high-latitude greening to promote improved consensus, suggest a framework to  
98 focus future work, and identify these key research priorities that will advance applications of  
99 satellite and *in situ* observations to the study of past, present, and future Arctic vegetation  
100 change.

101

102 **The greening of the Arctic**

103 Over the past forty years, circum-Arctic measures of vegetation dynamics by satellites  
104 document widespread and long-term greening trends that are generally interpreted as signs

105 of increased *in situ* biomass and productivity of Arctic terrestrial vegetation<sup>3,5,6,12,23,28</sup>. Slowing  
106 or reversal of these trends in recent years suggests a greater diversity of ecological  
107 responses to regional climate change than previously assumed<sup>18,20,26,29,30</sup> (Fig. 1).  
108 Terminology is mixed, but ‘greening’ is commonly used as shorthand for describing multi-  
109 decadal increases in remotely-sensed proxies of vegetation productivity thought to represent  
110 increased vegetation biomass *in situ*. Less frequently, greening is also used to describe  
111 advances in the seasonal timing of these vegetation proxies<sup>29,31</sup>. ‘Browning’ has been used  
112 in different ways in the literature, either representing a slowdown in the rate of greening, a  
113 switch in trend direction, or a decrease in greenness due to plant dieback from disturbances  
114 such as fires, insect outbreaks or extreme weather events<sup>18</sup>. In the most recent  
115 Intergovernmental Panel on Climate Change report, tundra vegetation change was identified  
116 as one of the clearest examples of terrestrial impacts, with reported high confidence in both  
117 the detection and attribution of responses to climate change with evidence for change  
118 detection including greening trends derived from satellite observations<sup>1,16</sup>. Recent efforts to  
119 synthesize vegetation change in Arctic ecosystems – including changes in plant productivity,  
120 biomass, cover, composition or phenology over time and in response to warming – suggest  
121 that vegetation change is concurrent with greening observed by satellites<sup>9,32,33</sup>. However,  
122 whether and how *in situ* changes in tundra productivity and phenology are directly related to  
123 the widespread changes in vegetation indices measured by satellites remains unclear.

124

### 125 **Vegetation indices as proxies of vegetation productivity**

126 Long-term trends in global vegetation dynamics are most commonly quantified from time-  
127 series of spectral vegetation indices derived from optical satellite imagery. These indices are  
128 designed to isolate signals of vegetation productivity from background variation by  
129 emphasizing reflectance signatures associated with plant structure or physiology in discrete  
130 regions of the radiometric spectrum<sup>3,34–37</sup>. Common vegetation indices include the  
131 Normalized Difference Vegetation Index (NDVI, Fig. 2), Enhanced Vegetation Index (EVI),  
132 Soil Adjusted Vegetation Index (SAVI), and Green Chromatic Coordinate index (GCC),

133 among many others<sup>38–40</sup>. NDVI has been and continues to be the most widely used  
134 vegetation index, owing much to its simple ratio formula based on spectral bands monitored  
135 by early-generation earth observing satellites launched in the 1970s (Fig. 2). It is primarily for  
136 this historical continuity - rather than being best fit-to-purpose - that NDVI is the most  
137 commonly used index to quantify multidecadal Arctic greening. Most studies of long-term  
138 trends calculate annual measures of maximum NDVI to measure change over space and  
139 time, though time-integrated approaches are also used<sup>35,41–43</sup>. The longest-term freely-  
140 available NDVI datasets have been produced from several sensors with broad spatial  
141 coverages and different sampling frequencies, including primarily: 1) the Advanced Very-  
142 High-Resolution Radiometer (AVHRR – 1982 to present) on board NOAA satellites, 2) the  
143 Moderate-resolution Imaging Spectroradiometer (MODIS – 2000 to present) on board NASA  
144 satellites, and 3) NASA-USGS Landsat sensors (1972 to present). However, trends in NDVI  
145 data produced from different satellite datasets do not always correspond at a given location  
146 nor are dynamics of different greening metrics consistent across datasets<sup>44</sup> (Fig. 1). Thus, it  
147 can be challenging to distinguish ecological change from differences due to methods and  
148 sensor/platform-related issues when interpreting localized greening or browning signals  
149 (Table 1).

150

### 151 **The ecology of greening and browning *in situ***

152 The biophysical and ecological processes that drive greening or browning patterns  
153 measured by satellites are diverse and may unfold across overlapping scales, extents and  
154 timeframes. In tundra ecosystems, vegetation changes linked to greening include for  
155 example: encroachment of vegetation on previously non-vegetated land surfaces<sup>9</sup>,  
156 increasing biomass of previously existing vegetation<sup>45</sup>, changes in community composition –  
157 such as tundra shrub expansion<sup>9</sup>, and/or changes in plant traits such as height<sup>32</sup>, leaf area,  
158 or phenology<sup>46,47</sup>. Tall shrub tundra typically has a higher NDVI than other tundra plant  
159 types<sup>48–50</sup>, and bare ground<sup>34</sup> has a much lower NDVI than vegetated tundra (Fig. 2). Tundra  
160 without vascular plants, however, could have a substantial cover of biological soil crust

161 communities consisting of lichens, cyanobacteria, mosses and green algae that may  
162 influence NDVI<sup>51</sup>. Thus, heterogeneity in plant communities, land cover and topography can  
163 influence the greenness of landscapes<sup>52</sup> and likely greening trends over time.

164

165 Not all areas of the Arctic are greening (Fig. 1), and in recent years heterogeneity in the  
166 direction and magnitude of vegetation change has become more pronounced<sup>18–20,30</sup>.

167 Ecological explanations for vegetation browning include for example the sudden loss of living  
168 biomass due to extreme climatic events<sup>21,53–55</sup>, biological interactions (e.g., disease or  
169 herbivore outbreaks)<sup>56–58</sup>, permafrost degradation<sup>19,59</sup> (Fig. 1), increases in standing dead  
170 biomass<sup>60</sup>, coastal erosion<sup>61</sup>, salt inundation<sup>62</sup>, altered surface water hydrology<sup>63,64</sup> or  
171 fire<sup>6,65,66</sup>. Additionally, decreased rates of vegetation greening could also be attributed to  
172 reduced productivity, not necessarily indicating browning vegetation, but rather a decrease in  
173 annual greenness due to more adverse growing season conditions, shorter growing  
174 seasons<sup>20</sup> or nutrient limitation<sup>25</sup>. Despite these changing dynamics, long-term greening  
175 trends remain far more pervasive than browning in tundra ecosystems (ratio of 20:1 in Park  
176 et al 2016). At circum-Arctic scales, the magnitude, spatial variability, and proximal drivers of  
177 patterns and trends of greening and browning are not well understood.

178

### 179 **Correspondence between satellite and ground-based observations**

180 Evidence for correspondence among *in situ* vegetation change and changes in satellite-  
181 derived vegetation indices is mixed<sup>67–70</sup>. NDVI trends across satellite datasets do not  
182 necessarily directly correspond with one another<sup>3,6</sup>, nor does any one sensor or vegetation  
183 index combination correspond directly with *in situ* vegetation composition change<sup>69</sup>. For  
184 example, AVHRR NDVI greening trends did not correspond with the lack of change  
185 observed with Landsat NDVI data and *in situ* plant composition between 1984 and 2009 in  
186 North Eastern Alaska<sup>69</sup>. NDVI has been related to interannual variation in radial shrub  
187 growth<sup>7,10,71</sup>, yet how radial growth links to change in leaf area or aboveground biomass is  
188 not always clear<sup>72–74</sup>, let alone how it influences landscape measures of productivity (Fig. 4).

189 Making direct comparisons of productivity changes from vegetation cover estimates<sup>9,33</sup>,  
190 biomass harvests<sup>48</sup> or shrub growth<sup>75</sup> is complicated by the lack of annual-resolution data  
191 and low sampling replication across the landscape.

192

193 In addition to productivity analyses, growing season length<sup>17,76,77</sup> and plant phenology  
194 advance over time<sup>76,78–82</sup> have been quantified using both satellite and ground-based  
195 datasets, though paired comparisons do not always correspond (Fig. 5). Measures of longer  
196 growing seasons have been attributed to earlier snowmelt and/or earlier leaf emergence in  
197 spring<sup>83</sup>, and longer periods of photosynthetic activity or later snowfall in autumn<sup>84</sup>. However,  
198 the few studies that have monitored both localised leaf emergence and senescence of  
199 tundra plants have not found evidence for an increasing growing period at specific sites<sup>76,77</sup>.  
200 In addition, community-level analyses indicate shorter flowering season lengths at sites  
201 around the tundra biome<sup>85</sup>. Plant phenology changes with warming<sup>85,86</sup> could also be linked  
202 to changing species composition or diversity<sup>9,32,33</sup>, thus influencing the phenological diversity  
203 across the landscape<sup>87,88</sup>. However, for satellite observations may not capture whether  
204 photosynthetic activity begins earlier in the spring and/or continues later into the autumn in  
205 tundra ecosystems where deciduous vascular plants make up only a portion of the vegetated  
206 land cover. Taken together, whether circum-Arctic satellite observations across high  
207 latitudes represent either a longer snow-free period uncoupled from vegetation response or  
208 an actual realized longer growing season of plants remains uncertain<sup>76,89–91</sup>.

209

210 Explaining the lack of correspondence between *in situ* and satellite-derived measures of  
211 tundra vegetation change and greening is fraught with complexities of terminology,  
212 challenges of interpretation of spectral vegetation indices at high latitudes, and scaling  
213 issues (Fig. 4).

214

215 **Challenge 1: Terminology**



216 Although the terms ‘greening’ and ‘browning’ were first popularized in the context of boreal  
217 forest change<sup>8</sup> they have been adopted to describe widespread changes throughout all  
218 terrestrial Arctic systems<sup>4,5,7,20</sup>. Greening and browning trends refer to decadal phenomena  
219 that may operate at any spatial scale, from localized patches, to landscapes or even biome  
220 extents, while greening and browning events occur more rapidly (i.e., are short term) and,  
221 due to their mechanistic drivers, will often be restricted from patch to regional scales (the  
222 impacts of volcanic eruptions, such as Mount Pinatubo in 1991, are an exception<sup>92</sup>).  
223 Therefore, greening or browning events might be embedded within overall greening or  
224 browning trends without necessarily driving them (Fig. 6). In turn, greening or browning  
225 trends and events may also result in threshold changes where productivity does not return to  
226 the longer-term baseline (Fig. 6; e.g., pulse in recruitment at treeline<sup>93</sup> or shrubline<sup>94</sup> or a  
227 large fire<sup>66</sup>). The baseline to which we compare productivity change will influence our  
228 interpretation of trends<sup>95</sup>. In both satellite datasets and field observations, the baseline  
229 conditions are often constrained by the limitations of data availability rather than any  
230 ecologically meaningful starting point<sup>3</sup>. For these reasons, substantial uncertainty associated  
231 with ecological attribution of greening and browning could be reduced by more  
232 comprehensive descriptions of these time series beyond simply the direction of trends (Fig.  
233 6).

234

235 With a baseline and trend direction established, examining the trend magnitude and variance  
236 around the fit over time can aid ecological interpretation (Fig. 6). To distinguish greening and  
237 browning events from the longer-term trends, we propose defining events as “outliers in  
238 NDVI (or other spectral vegetation indices) that occur relative to the long-term mean or  
239 trend” using a Theil-Sen estimator or similar statistical test for robust trend analyses of  
240 satellite data<sup>35,96</sup>. Here, we define a *greening trend* as an increase in NDVI or other  
241 greenness-related indices over decadal time scales. When attributed to *in situ* vegetation  
242 change, we interpret this pattern as improved conditions for photosynthesis, reduced  
243 resource limitation, or responses to disturbance in plant communities, resulting in greater

244 aboveground biomass, leaf area, productivity or successional change. We define a *browning*  
245 *trend* as a decrease in NDVI or other greenness-related indices over decadal time scales.  
246 Browning trends may correspond with an *in situ* change in vegetation productivity due to  
247 plant dieback or loss of vegetation cover through biotic or abiotic disturbances.

248

249 We suggest avoiding definitions of browning that refer to a slowdown of positive greening  
250 trends because the relationship between vegetation indices and on-the-ground measures of  
251 vegetation productivity is non-linear and variable (Fig. 2 and 6). A slowdown in a  
252 positive vegetation index time series trend could therefore relate to a decline, no change, or  
253 even an increase in vegetation productivity on the ground purely due to statistical rather than  
254 ecological factors. To some degree any definition of greening or browning is arbitrary, but  
255 the purpose of the definitions we propose here is to draw a distinction between slower acting  
256 climatic or biotic drivers of greening or browning trends versus event-driven changes caused  
257 by weather, biotic pulses, or other regional events such as fire. Beyond advocating for  
258 clearly defined terms, challenges persist in the interpretation of physiologically meaningful  
259 parameters from the available long-term optical satellite data, and in overcoming the  
260 mismatch between observations and their potential drivers that operate across different  
261 spatial and temporal scales.

262

## 263 **Challenge 2: Understanding spectral vegetation indices**

264 Vegetation indices are proxies of photosynthetic activity rather than direct measurements of  
265 biological productivity<sup>38,97,98</sup>. The statistical relationship between a vegetation index and  
266 biomass, phenology, or any other measures of productivity can vary due to a suite of intrinsic  
267 (e.g., sensor design, quality flagging algorithms) and extrinsic (e.g., atmospheric conditions,  
268 sun angle) factors<sup>3,99</sup> (Table 1). For example, the centre wavelength and width of red or  
269 near-infrared or other spectral bands used to generate vegetation indices were designed for  
270 different purposes in different sensors (Fig. 2). While the formula for NDVI may be the same,  
271 the covered spectral wavelength ranges differ between different satellite datasets<sup>100</sup>(Fig. 2B),

272 and may be more or less sensitive to specific non-vegetative influences, such as  
273 atmospheric scattering or the magnitude of spectral mixing associated with non-vegetated  
274 surfaces. Widespread non-vegetative changes in high-latitude ecosystems could confound  
275 and decouple vegetation index time series from changes in plant productivity (Table 1). For  
276 example, changes in the extent of summer snow patches<sup>101</sup>, surface water<sup>102</sup> or surface soil  
277 moisture<sup>103</sup> that are often associated with landscape-scale topographic variation could  
278 influence greening patterns and trends. In addition, satellite data signal processing varies  
279 across available products. Thus, strong caution is warranted when comparing products or  
280 even versions of the same product with different atmospheric corrections, quality  
281 assessments, and spatial/temporal compositing approaches. The influences of non-  
282 vegetative geophysical and signal processing factors on NDVI are actively studied by the  
283 remote-sensing community (Table 1), but could be better accounted for or quantified in  
284 Arctic greening studies.

285

286 The potential for non-linear relationships between vegetation indices and measures of Arctic  
287 vegetation productivity presents further conceptual challenges in trend interpretation (Fig. 2).  
288 These arise from comparing a normalized ratio against a continuous productivity measure of  
289 interest, such as biomass changes or shrub ring width (Fig. 4). A linear trend in an NDVI  
290 time series (Fig. 1) does not necessarily mean linear changes in vegetation productivity<sup>34,104</sup>  
291 (Fig. 2). Because greening and browning terminology are tied to changes in vegetation  
292 proxies, such as NDVI, rather than direct measures of biological change, mismatches could  
293 occur between remotely-sensed vegetation proxies and *in situ* vegetation change (Fig. 3, 4  
294 and 5). These potential mismatches exemplify why caution should be used when interpreting  
295 linear trends in ratio-defined (i.e., potentially nonlinear) proxies.

296

297 Measuring landscape phenology with satellite data (phenometrics), especially at high  
298 latitudes, presents additional challenges to simple ecological interpretation that are  
299 associated with methodologies and seasonal variations in data quality (Table 1). For

300 example, vegetation metrics from early spring are much more likely to be influenced by  
301 snow, standing water or low sun angle than those closer to peak biomass in mid- to late-  
302 summer<sup>5,49,103</sup>, yet these are critical periods for establishing a baseline for curve fitting or  
303 thresholding used to derive phenometrics. Seasonal variation in cloud or fog cover, highly  
304 variable and sensitive to changing sea ice conditions<sup>105</sup>, further influences both data  
305 availability and image compositing approaches in many phenology products<sup>46</sup>. Use of time-  
306 integrated vegetation indices can reduce some of these signal to noise issues<sup>17,106</sup>, but  
307 ultimately no phenometric is best suited to all Arctic environments. Snow regimes and land  
308 cover variability differ annually and regionally and thus phenometrics using coarse-grain  
309 imagery can integrate different abiotic and biotic signals at different points in space and  
310 time<sup>107</sup>. Phenological differences of days to weeks or even months<sup>108</sup> can result from  
311 analyses using different methods and metrics for the same datasets at the same location, ,  
312 such relative differences are of substantial ecological importance given the short growing  
313 seasons of the Arctic<sup>17,107</sup>. Circum-Arctic analyses of vegetation indices generally agree that  
314 phenological shifts in the greenness of the landsurface are widespread<sup>17,109,110</sup>, but caution is  
315 warranted for local-scale comparisons or mechanistic interpretations of biome-scale trends.

316

### 317 **Challenge 3: Scaling issues**

318 Scale, and its influence on pattern, presents a longstanding challenge in the interpretation of  
319 remotely-sensed vegetation proxies<sup>111-113</sup>. All long-term vegetation proxy time series  
320 (Landsat, MODIS, AVHRR) spatially aggregate spectral data to pixels (i.e. grains) that span  
321 hundreds of square metres to tens of square kilometres, reducing the spectral signatures of  
322 a substantial number of individual plants and non-vegetative features in a landscape to a  
323 single numerical value. The loss of variability within pixels masks information useful for the  
324 attribution of greening signals to ecological processes (Table 1, Fig. 4). For example, within  
325 a single AVHRR GIMMS3g pixel (where a sub-selection of 1 km x 1 km pixels are upscaled  
326 to 8 x 8 km<sup>37</sup>), greening signals, such as increased shrub cover on south-facing slopes or re-  
327 vegetation of drained lake beds, may be mixed with browning signals, from disturbances

328 such as retrogressive thaw slumps or vegetation trampling by herbivores. The emergent time  
329 series from such a pixel describes no single vegetation dynamic, but rather their integrated  
330 spectral responses (Fig. 4). Broad-scale patterns of spatial variability in greening and  
331 browning across pixels are also influenced by grain size<sup>113</sup> (Figure 1). However, the extent to  
332 which the sometimes-contradictory greening and browning signals found across different  
333 datasets can be attributed to the influence of scale of measurement on pattern formation is  
334 poorly understood. Both spatial and temporal patterns in coarse-grained vegetation proxies  
335 capture signals of changing phytomass<sup>10,34,50,69,104</sup>, but lacking additional context, they are  
336 generally insufficient for the attribution of trends to specific ecological mechanisms of *in situ*  
337 vegetation change.

338

339 The low temporal sampling frequency of a few days to a few weeks of many remote-sensing  
340 datasets also introduces temporal scale-dependent effects that may be magnified in Arctic  
341 systems (Table 1). At high latitudes, optical satellite sensors are only effective for a short  
342 annual window due to prolonged polar night, with further data quality issues associated with  
343 low sun angle, and persistent cloud cover (Table 1). For example, comparisons of phenology  
344 across latitudes can be less reliable at higher versus lower latitudes due to shorter growing  
345 seasons and therefore fewer satellite data collection points for use in change detection  
346 analyses<sup>114</sup>. Metrics based on the annual maximum NDVI of a given pixel are also more  
347 likely to be influenced by temporal sampling artefacts at high latitudes than those that  
348 integrate productivity estimates through time, such as the growing season integrated NDVI  
349 (GSINDVI)<sup>41</sup>, time-integrated NDVI (TiNDVI)<sup>42</sup> or early growing season integrated NDVI  
350 indices<sup>43</sup>. The magnitude and extent of spatial and temporal scaling issues in high-latitude  
351 ecosystems warrant further consideration and research, both from remote sensing and field-  
352 based projects<sup>112</sup>.

353

## 354 **Emerging tools and observation networks**

355 Many factors need careful consideration in comparisons between *in situ* changes in plant  
356 biomass and coarse-grained satellite measures of productivity and phenology. Existing *in*  
357 *situ* observations from long-term ecological monitoring<sup>9,18,33,76,115</sup>, historical imagery<sup>116,117</sup>,  
358 phenocam networks<sup>118</sup> and high-resolution imagery such as from aircraft, flux towers, and  
359 drones<sup>119,120</sup> are not spatially or temporally comprehensive, yet provide invaluable context to  
360 the interpretation and modelling of ecological dynamics captured by existing decadal satellite  
361 observations. Recent and ongoing release of satellite datasets to the research community  
362 such as the privately owned Digital Globe and Planet constellations or the European Union  
363 funded Sentinel missions will provide higher spatial (2-10 m) and temporal resolution (1-5  
364 days) across the Arctic with spectral bands designed for the calculation of both widely-used  
365 and newly developed vegetation indices<sup>121–123</sup>. Reanalysis of existing datasets with improved  
366 atmospheric corrections, such as MODIS MAIAC<sup>124</sup>, will improve understanding of past  
367 changes. Data collection campaigns equipped with improved sensors, such as those that  
368 can measure solar-induced chlorophyll fluorescence (SIF) at high resolution<sup>125,126</sup>, and the  
369 increasingly widespread adoption of proximal remote-sensing platforms such as aircraft,  
370 drones and phenocam networks using standardized protocols will be required to better test  
371 the links between *in situ* vegetation dynamics and broader remotely sensed patterns and  
372 trends. In addition, data integration modelling approaches will be necessary to  
373 mechanistically link remote sensing observations with ecological change in high-latitude  
374 ecosystems<sup>14,19,127</sup>.

375

## 376 **Future research priorities**

377 We have identified three future areas for fundamental advances in our understanding of  
378 greening and browning dynamics at high latitudes, these include:

379

380 **1. Validation of existing observations** – Where are we confident in observed greening  
381 and browning trends? Where do we have less confidence in the interpretation of

382 patterns and trends in vegetation indices? How can local-scale information (e.g.,  
383 topographic and/or land-cover heterogeneity) inform the validation of existing  
384 observations?

385 **2. Integrated interpretations of change** – Can scaling issues be surmounted to find  
386 common signals of change across different observations? How can this information from  
387 various sources and scales (e.g., satellites, airborne, drone, phenocam and *in situ*  
388 records) be integrated to inform deeper ecological understanding of the drivers of  
389 greening and browning patterns and trends?

390 **3. Mechanistic understanding of observations** – Can we mechanistically test, model  
391 and hind cast patterns of vegetation proxy change? How can greening and browning  
392 observations be integrated into dynamic vegetation and Earth system models to improve  
393 our understanding of global climate feedbacks at high latitudes (e.g., carbon cycling and  
394 surface energy budget feedbacks)?

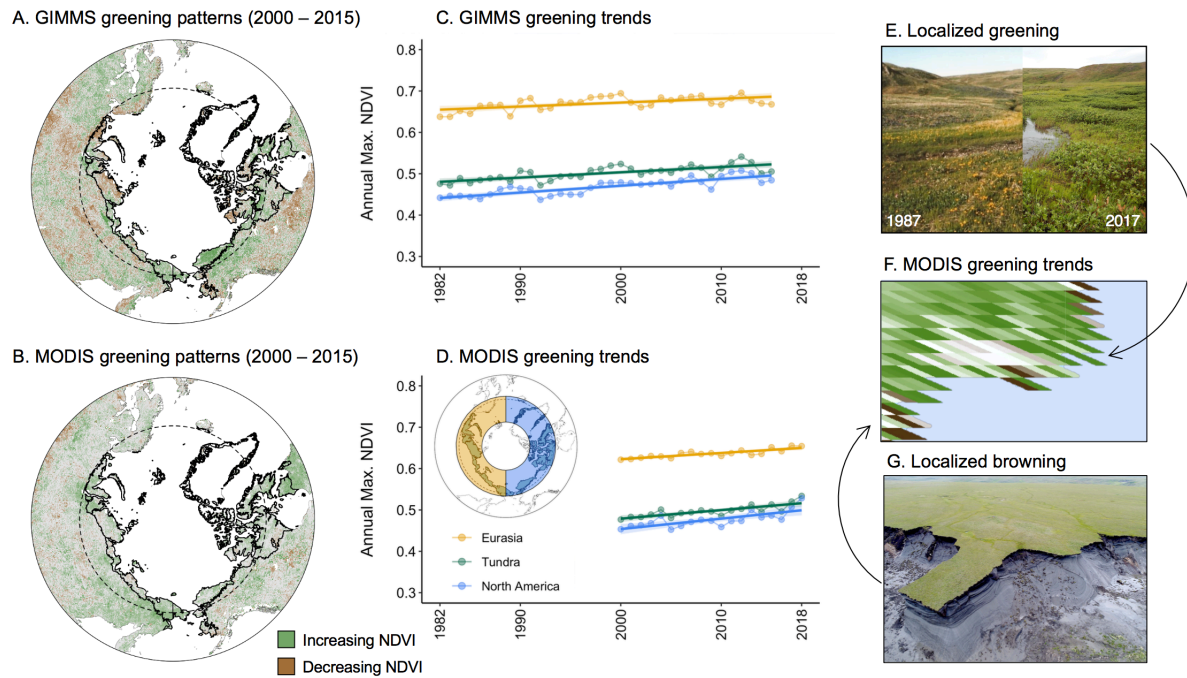
395

## 396 **Conclusions**

397 Recent research has highlighted the complexity in observed Arctic greening and browning  
398 trends and patterns. Although satellite data have been used to detect and attribute global  
399 change impacts and resulting climate feedbacks in Arctic ecosystems<sup>1,15</sup>, substantial  
400 questions and uncertainties remain. The three major challenges in resolving these  
401 uncertainties are: 1) improving the clarity of the definitions of widely used terminology  
402 associated with greening and browning phenomena, 2) promoting the understanding of the  
403 strengths and limitations of vegetation indices when making ecological interpretations and,  
404 3) better incorporating and accounting for different scales of observations and observation  
405 error into analyses of changing tundra productivity and phenology. New sensors and better  
406 access to legacy data are promising developments, but new data alone will not provide  
407 solutions to many of these longstanding conceptual and technical challenges. The  
408 complexity of Arctic greening patterns will only be fully understood through multidisciplinary  
409 efforts spanning the fields of ecology, remote sensing, climate science, Earth science and

410 computer science that look towards contemporary and future change, but also backwards by  
411 conducting re-analyses of historical data. Ultimately, we urgently need a deeper  
412 understanding of the relationships between patterns and processes in greening and  
413 browning dynamics to improve estimates of the globally-significant climate change  
414 feedbacks in high-latitude ecosystems<sup>1</sup>.





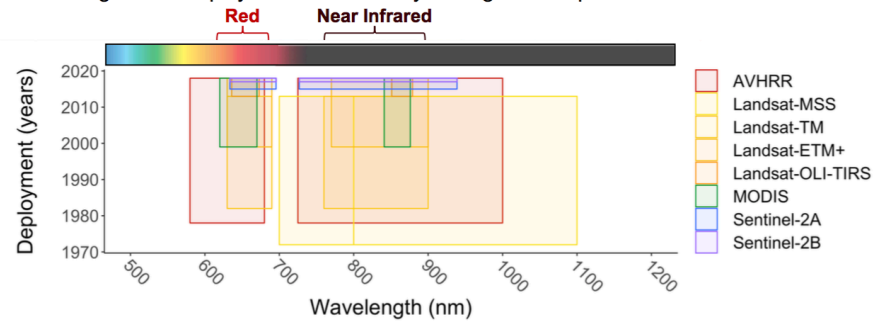
415

416 Figure 1. Arctic greening patterns vary across space and time and among satellite datasets  
 417 likely driven in part by actual *in situ* change and in part by challenges of satellite data  
 418 interpretation and integration. Trends in maximum NDVI are spatiotemporally variable across  
 419 the circum-polar North (A and B, data subsetted to temporally overlapping years), and  
 420 maximum NDVI varies by geographic region (C and D, full time series), expressed by  
 421 localized greening - for example shrub encroachment (E) - and browning such as this  
 422 retrogressive thaw slump (G) occurring at the pixel scale on Qikiqtaruk - Herschel Island in  
 423 the Canadian Arctic (F). NDVI trends were calculated using robust regression (Theil-Sen  
 424 estimator) in the Google Earth Engine for the GIMMS3gv1 (1982 to 2015) and MODIS  
 425 MOD13A1v6 (2000 to 2018) NDVI products. Dashed line indicates the Arctic Circle and the  
 426 black outlined polygon indicates the Arctic tundra region from the Circum-Arctic Vegetation  
 427 Map ([www.geobotany.uaf.edu/cavm/](http://www.geobotany.uaf.edu/cavm/)).

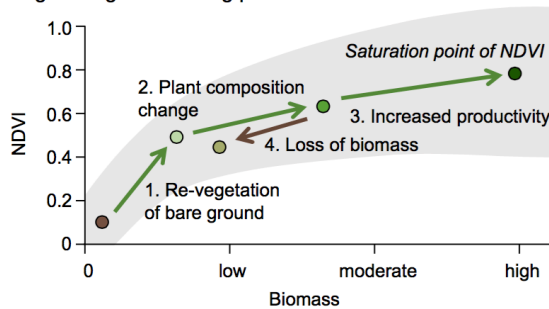
A. Ratio formula for NDVI

$$\text{NDVI} = \frac{(\text{NIR} - \text{Red})}{(\text{NIR} + \text{Red})}$$

B. Wavelengths and deployment duration vary among satellite platforms



C. Simplified non-linear relationship between NDVI and biomass and hypothetical productivity changes leading to greening or browning patterns or trends



**Greening:**

1. Re-vegetation of bare ground – e.g., Revegetation of bare ground after disturbance or glacial retreat (primary succession)
2. Plant composition change – e.g., increased growth and cover of shrubs in a tundra plant community
3. Increased productivity – e.g., Doubling of biomass due to increased vegetation growth

**Browning:**

4. Loss of biomass – e.g., Decreased plant productivity due to high levels of herbivory

428

429 Figure 2. The Normalized Difference Vegetation Index (NDVI is calculated by a simple ratio

430 formula of the red and near infrared bands (A). Different satellite sensors produce bands that

431 are nominally called ‘Red’ or ‘NIR’ (among others) but they can span substantially different

432 spectral widths even if they share a similar centre wavelength (B). Time series of high-

433 latitude NDVI greenness from different satellite datasets or changing sensors on the same

434 satellite platform may differentially respond to changes captured in these spectra. Different

435 satellite datasets have been deployed for longer or shorter durations introducing challenges

436 to cross sensor comparisons when also capturing longer-term vegetation change (Fig. 1)

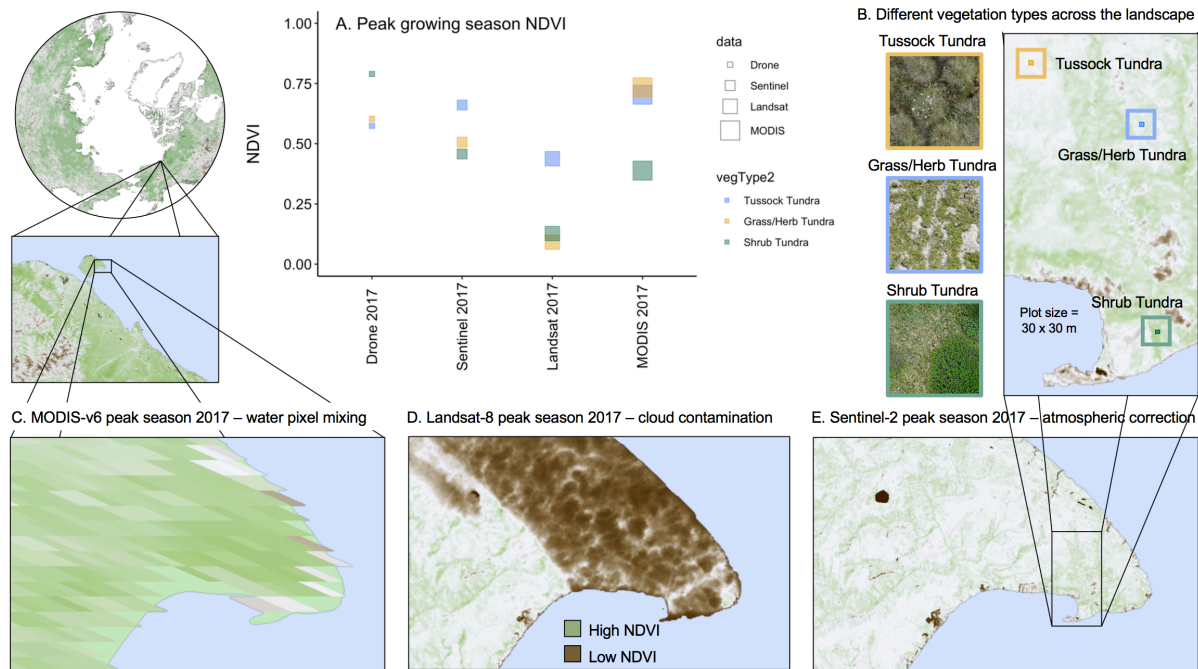
437 even among intercalibrations of the same sensor type on different generations of satellite

438 platforms. The relationship between biomass and NDVI is non-linear (C). Thus, different

439 ecological mechanisms (hypothetical here) could lead to very different magnitudes of

440 greening and browning change depending on the initial and final biomass of the changing

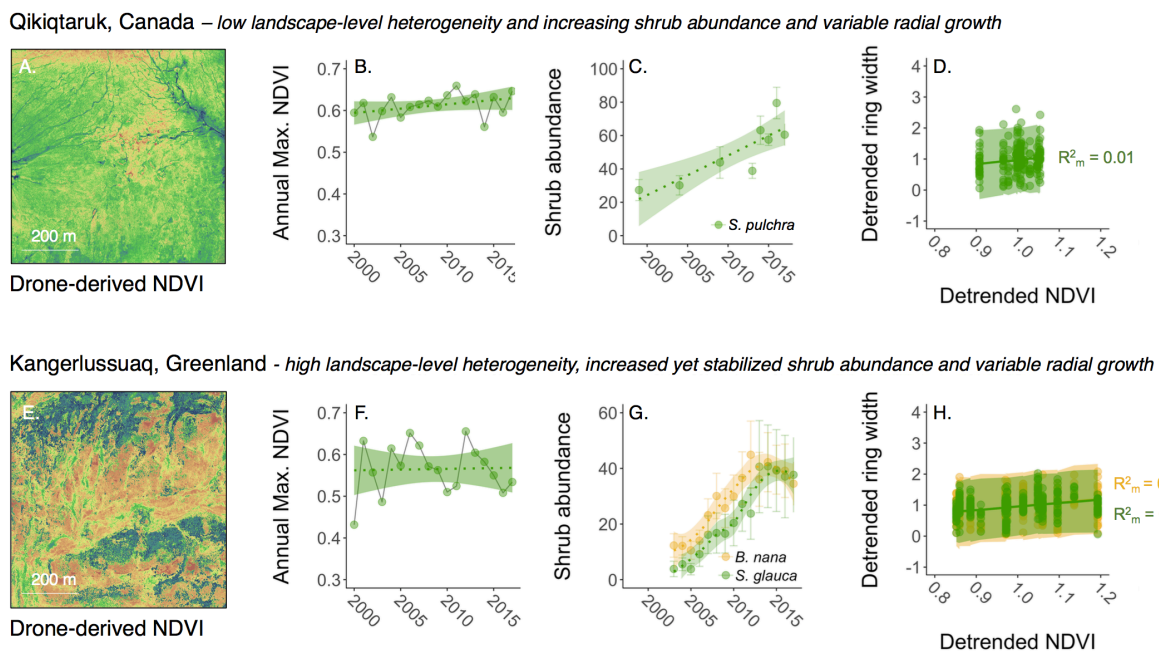
441 vegetation.



442

443 Figure 3. Localized interpretations and comparisons of NDVI 'greenness' are challenging to  
 444 make across data collected across different spatial scales (including grain sizes and  
 445 extents), landscape contexts, and periods within the growing season (A – E, Table 1). Here,  
 446 we plot NDVI patterns for peak season (derived from available cloud-free data between 13<sup>th</sup>  
 447 July to 4<sup>th</sup> August in 2017, but note in B that there were no cloud-/fog-free Landsat data  
 448 available). We purposefully present data with quality and processing issues above to  
 449 highlight the challenges in isolating scaling factors (e.g., timing of image acquisition and  
 450 grain size of imagery), data quality (e.g., cloud contamination and lack of atmospheric  
 451 corrections) from differences in ecological context (e.g., vegetation type) in quantifying NDVI  
 452 in regional to global studies where data quality issues maybe spatially or temporally variable  
 453 among locations. On Qikiqtaruk – Herschel Island in the Canadian Arctic during the period of  
 454 2017 peak biomass, NDVI values from commonly available satellite data products and drone  
 455 datasets (A) differed substantially across products and across 30 m x 30 m plots of three  
 456 different vegetation types (B). Here, factors such as sub-pixel mixing (C), cloud or fog  
 457 contamination (D), lack of atmospheric correction (E), different plot grain sizes of data in  
 458 more or less heterogeneous vegetation cover and timing of data collection could have all  
 459 influenced NDVI values. Data were analysed and extracted for 30 x 30 m plots using the

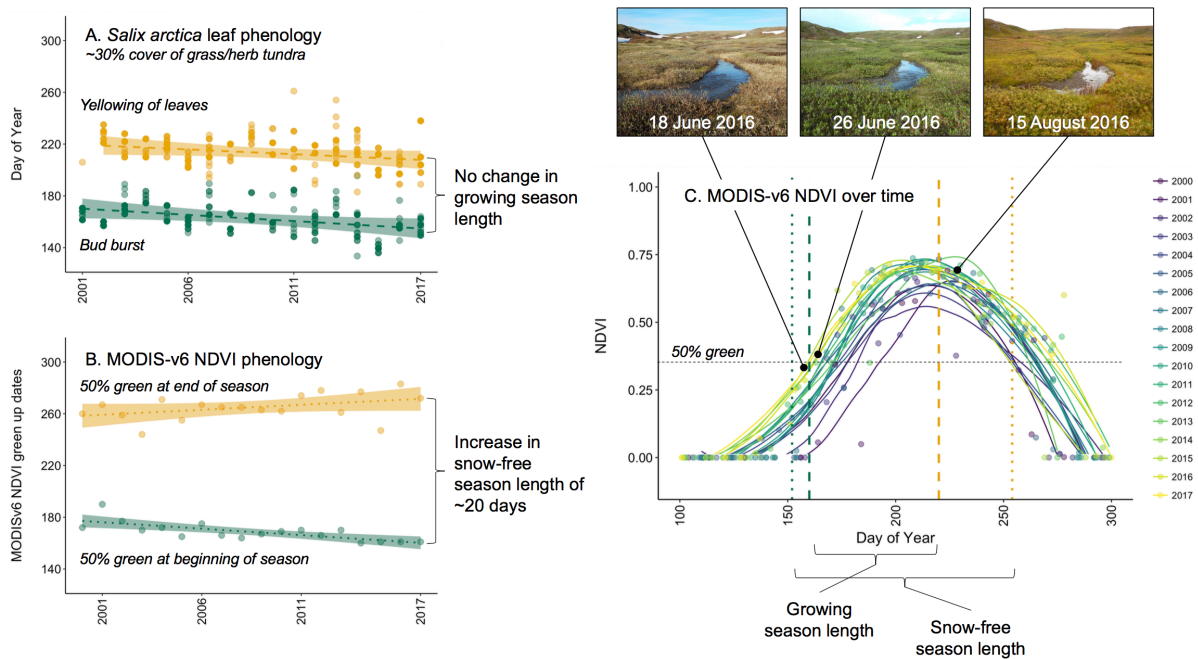
460 Google Earth Engine for the MODIS MYD13A1v6 (pixel size = 500 m x 500 m) and Landsat  
 461 8 (pixel size = 30 m x 30 m) NDVI product, and the top-of-atmosphere Sentinel-2 NDVI  
 462 product without atmospheric corrections (pixel size = 10 m x 10 m) NDVI, and Pix4D-  
 463 processed drone data collected using a radiometrically calibrated four-band multispectral  
 464 sensor (Sequoia, pixel size = 12 cm x 12 cm) on an FX-61 fixed-wing platform with the High-  
 465 latitude Drone Ecology Network protocols<sup>128</sup> (arcticdrones.org).  
 466



467  
 468 Figure 4. Sub-pixel spatial heterogeneity in greening and browning (A, E) can influence the  
 469 observed signal at coarser grains (B, F) and may or may not represent *in situ* observations  
 470 of vegetation change such as increases in shrub abundance (C, G) and interannual  
 471 variability in shrub growth (D, H, sample sizes: Yukon *Salix pulchra* = 21<sup>76,129</sup>, Greenland  
 472 *Betula nana* = 42<sup>73,130</sup>, *Salix glauca* = 32<sup>73,131</sup>). Error bars (C, G) are standard error around  
 473 mean values of shrub abundance derived from point framing in 12 1-m<sup>2</sup> plots at the  
 474 Qikiqtaruk site<sup>76,129</sup> and 13 0.25-m<sup>2</sup> plots at the Kangerlussuaq site<sup>132,133</sup>. Models error (D, H)  
 475 are credible intervals for a Bayesian hierarchical models of the relationship between annual  
 476 growth rings and NDVI with shrub individual and year as random effects. Detrending is using  
 477 a spline fit from the dpIR package in R. Credible intervals for model slopes overlapped with

478 zero indicating that the relationships in D and H are not statistically significant. Marginal  $R^2$   
 479 values indicate the variance in detrended ring widths explained by detrended NDVI (D, H).  
 480 Low heterogeneity (i.e. relatively homogenous land cover) sites might be more likely to  
 481 express clear greening (A) trends versus high-heterogeneity sites (with a variety of land-  
 482 cover types, each potentially responding differently) that might be more likely to have  
 483 variable NDVI among years (B). Landscape NDVI patterns (A and F) were measured using a  
 484 Parrot Sequoia and FX-61 fixed wing platform according to High-latitude Drone Ecology  
 485 Network protocols in the summer of 2017 (arcticdrones.org) and analysed using the Pix4D  
 486 software. Coarser-grain NDVI time series (MODIS MOD13A1v6, 500m pixels) were  
 487 calculated using Google Earth Engine and the Phenex package in R.

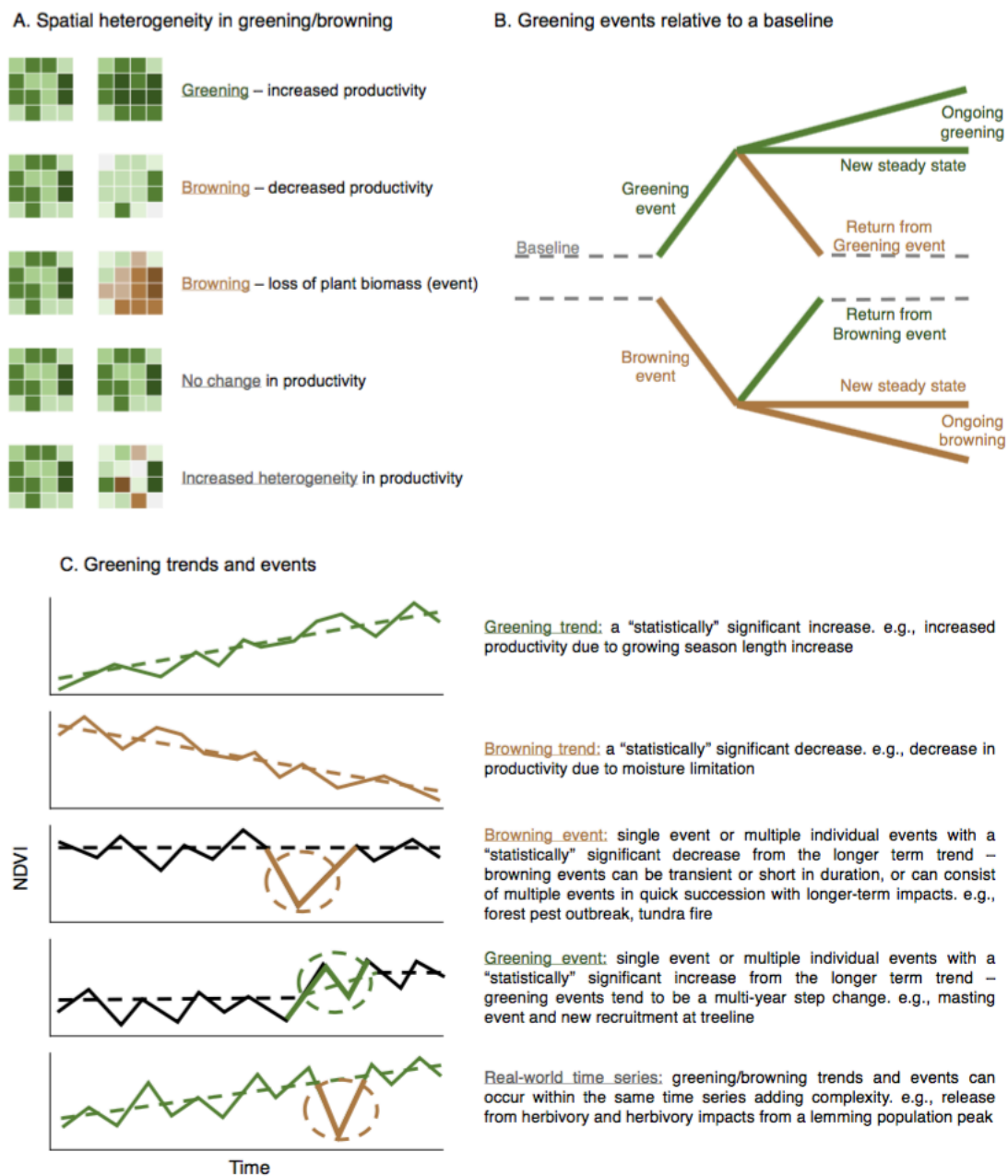
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489

490 Figure 5. Satellite observed snow-free season length of the land surface (B and C) might not  
 491 directly correspond to the growing season of plants in tundra ecosystems (A). Plant  
 492 phenology data are from 20 monitored plots on Qikiqtaruk-Herschel Island for the species  
 493 *Salix arctica*, which makes up approximately 30% of the cover in the grass- and forb-  
 494 dominated vegetation type (Fig. 3), indicate that both leaf emergence and senescence have  
 495 become earlier, resulting in no change in realized growing season length despite substantial

496 increases in the snow-free period of the land surface<sup>76</sup> (A – C). Plant phenology data are  
 497 from the Qikiqtaruk Ecological Monitoring program<sup>129</sup> (A), and satellite data are MODIS  
 498 MOD13A1v6 extracted for the pixel containing the phenology transects with the Google  
 499 Earth Engine and interpolated and smoothed using the Phenex library in the programming  
 500 language R (B and C).



501  
 502 Figure 6. Conceptual diagrams and definitions of greening/browning trends versus  
 503 greening/browning events. Five examples of local scale (smaller sub-units within a single  
 504 conceptual ‘pixel’) changes in plant productivity show how the combination of

505 greening/browning events/trends within a pixel can be reduced to a higher-level  
506 greening/browning pattern as their effects are scaled up (A). The ecological processes that  
507 comprise greening and browning trends include a combination of events, such as a pulse of  
508 plant recruitment, a dieback of plants due to an extreme winter climate event, herbivore or  
509 disease outbreak or other disturbance and the subsequent recovery, or longer-term change  
510 such as increasing shrub cover or progression of permafrost disturbances and periglacial  
511 processes (B and C). A combination of high-/low-frequency and high-/low-intensity events  
512 can result in, for example, a browning trend over time (see also<sup>18</sup>).

513 Table 1. A variety of factors can influence the magnitude and direction of change in  
 514 vegetation indices. These effects can be more or less important in coarse-grain imagery and  
 515 can be particularly problematic at high latitudes. The effects include: 1) radiometric effects -  
 516 differences among satellite datasets include band widths, atmospheric effects, cloud-  
 517 screening algorithms, sensor degradation, orbital shift and bidirectional reflectance  
 518 distribution functions originating from differences in field of view and sun geometries<sup>30,99,134</sup>;  
 519 2) spectral mixing - the reflectance of sub-pixel spatial heterogeneity that can influence the  
 520 overall pixel signal (Fig. 3); and, 3) adjacency effects - the reflectance of surrounding pixels  
 521 that can influence the signal of a given pixel (Fig. 3).

<b>Factors influencing vegetation indices</b>	<b>Specific effects</b>	<b>Influence on greening patterns and trends</b>
Low sun angle	Radiometric effects	At high latitudes, low sun angles and cloud shadows can have a greater influence on vegetation indices relative to lower latitudes <sup>106</sup> . Low sun angle reduces NDVI, an effect magnified in spring and autumn. Shadows also reduce NDVI and may be difficult to detect in coarse grained imagery <sup>43</sup> .
Cloud cover	Radiometric effects, Spectral mixing, Adjacency effects	Thin cloud, fog and smoke can influence imagery, reducing NDVI. Particularly problematic in coastal regions, cloud and fog can vary greatly between image acquisitions <sup>135</sup> . Cloud-screening algorithms differ among satellite datasets (partly as a function of available spectral bands), and partly cloudy or hazy conditions are particularly difficult for screening algorithms to detect consistently across different satellite products. The fogginess of Arctic locations can vary throughout time due to changing sea ice conditions <sup>105</sup> or increasing temperatures <sup>43</sup> .
Standing water	Spectral mixing, Adjacency effects	Standing water <sup>102</sup> can influence comparisons of vegetation indices across space and may not be detectable in coarse-grained imagery, despite influencing spectral signatures. NDVI values of water are generally low, however shallow water or standing water intermixed with vegetation or algal growth may not be identified as water by quality filters and may have higher NDVI. Water within a pixel may lead to artificially low NDVI values and can influence estimates of NDVI change over time. This is especially relevant to the Arctic during the spring and summer as snow melts and turns into numerous ephemeral ponds and lakes whose spectral signatures will be mixed with nearby vegetation. Changes in standing water over time associated with changing precipitation, permafrost conditions, and/or warming could drive NDVI signals rather than any changes in the plant biomass <sup>101,102,136,137</sup> .
Snow patches	Spectral mixing, Adjacency effects	Sub-pixel sized snow patches <sup>101</sup> will decrease the NDVI for a given tundra area. NDVI values of snow are strongly negative. Earlier snow loss may drive a strong positive trend in NDVI.



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		Longer persistence of snow on the landscape in patches may not be filtered by quality algorithms, but still lead to lower NDVI values.
Soil moisture	Spectral mixing	Soil moisture can influence the reflectance of vegetated tundra surfaces <sup>103,138,139</sup> . NDVI values are sensitive to soil moisture, which may or may not covary with vegetation changes. Furthermore, NDVI is relatively insensitive to changes in very sparsely vegetated (e.g., the High Arctic <sup>140</sup> ) and very densely vegetated (e.g., forest or shrubland <sup>141</sup> ) environments.
Short growing season	Timing of image acquisition	Trends in NDVI metrics and growing season length can be influenced by data acquisition and not only vegetation change. To compare spatial patterns in vegetation indices among sites, images are required from the same time within the growing season and the same time points within the day <sup>137</sup> . However, the short growing seasons at high latitudes make image acquisition a particularly important issue in these settings. Different datasets have different temporal frequencies for overpasses thus influencing comparisons. Growing season length decreases with higher latitudes, thus the impact of missing data is of a greater magnitude as latitude increases.
Rapid plant phenology	Chosen phenometric	The specific metrics used to quantify greening or browning will influence the resulting patterns observed <sup>108</sup> . Combining datasets with different spatial and temporal resolutions and/or using different phenometrics can limit comparisons when methodological signals overwhelm vegetation signals (Fig. 3). Variation in phenology metrics due to curve-fitting methods can exceed variation in measured phenology signals. Thus, using the same phenological functions across large geographic and ecological gradients, such as across the high latitudes, may introduce biases and/or errors.

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523 **Author Contributions**

524 IHM-S and JTK conducted the analyses and wrote the manuscript with contributions from all  
525 authors. GKP, JWB and HE contributed substantially to early versions of the manuscript.  
526 IHM-S, JTK, JJA, AMC, CJ, SA-B, HJDT and ESP collected drone and *in situ* data. This  
527 paper results from two collaborations: the sTundra working group led by IHM-S, SCE and  
528 ADB and the 'Event Drivers of Arctic Browning Workshop' at the University of Sheffield led  
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530

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554

555 **Data and code availability**

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558 Latitude Drone Ecology Network (arcticdrones.org/), shrub abundance<sup>129,132</sup>, annual growth  
559 ring<sup>129–131</sup> and phenology datasets<sup>129</sup>. Code is available in a GitHub repository  
560 (github.com/ShrubHub/GreeningHub).

561

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