- 1 Relationships between mycorrhizal type and leaf flammability in the Australian
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- 24 Highlights
 - Variation in morphological and chemical leaf traits is an important driver of fire regimes.
 - Many leaf and root traits are likely to evolve in a coordinated fashion.
 - Mycorrhizal type is a root trait that is also associated with variation in leaf traits.
- Arbuscular mycorrhizal plants, on average, express leaf traits that support
 rapid leaf ignition.
- Non-mycorrhizal and dual ecto- and arbuscular-mycorrhizal plants
 express leaf traits associated with longer burn duration.
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- 35 Abstract
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37 Mycorrhizal fungi have been linked to fire processes in natural ecosystems via their effects on litter decomposability but, to our knowledge, relationships 38 39 between mycorrhizal fungi and leaf traits directly associated with aspects of 40 flammability have not been studied. Here, we assessed the relationships among leaf traits and host mycorrhizal type for 77 species of Australian trees and 41 shrubs to determine whether mycorrhizal type can explain variation in three 42 43 aspects of leaf flammability (ignitability, fire duration, and smoulder duration). 44 Several associations were observed between mycorrhizal type and leaf traits 45 directly linked to flammability measures, including specific leaf area, leaf mass, 46 leaf moisture content, and leaf chemistry. The observed patterns suggest that 47 interactions between mycorrhizal fungi and their host plants during the growth 48 and senescence of leaves may have subsequent effects on fire processes. 49 However, further work is necessary to evaluate the importance of these effects in 50 real ecosystems, including whether plants or fungi are responsible for these 51 patterns, and we propose four questions that will further progress in this area. 52 53 54 55 Main text 56 57 Substantial functional variation exists among groups of mycorrhizal fungi, 58 particularly in the ability to directly access nutrients from organic matter. Some, 59 such as ectomycorrhizal (ECM) and ericoid mycorrhizal (ERM) fungi, express 60 enzymes that allow access to organic nitrogen (N; Chalot and Brun, 1998) and phosphorus (P; Cairney, 2011), and possibly even carbon (C; Talbot et al., 2008). 61 Arbuscular mycorrhizal (AM) fungi facilitate increased access to inorganic forms 62 of N and P but generally do not express enzymes that allow access to organic 63 forms (although some have proposed that acid phosphatase is potentially 64 expressed by some AM fungi; reviewed by Joner et al., 2000). Nonmycorrhizal 65 66 (NM) plant species have alternative strategies for obtaining nutrients such as 67 proteoid cluster roots (Shane and Lambers, 2005) and symbiosis with N-fixing organisms (Peoples et al., 2009). This variation and observed stoichiometric 68 69 relationships between carbon and nutrients in soil have led to hypotheses that 70 ecosystems vary substantially in their capacities to store carbon in soil based on 71 the dominant mycorrhizal type and its effects on soil nutrient pools (Averill et al., 72 2014; Phillips et al., 2013). 73 74 Fire is another important driver of the capacity of an ecosystem to store carbon 75 and much work has been done in attempts to understand the important drivers 76 of fire frequencies, intensities, and durations (Cornwell et al., 2009; Grootemaat

et al., 2015a). Mycorrhizal fungi may directly influence on the likelihood of fire

within an ecosystem due to their effects on standing litter (Gadgil and Gadgil,
1971), either via its quantity (through changes in decomposability, due to

increasing C:N and C:P ratios during nutrient uptake; Bending and Read, 1995)

81 or its flammability (through reductions in litter moisture content; Koide and Wu,

82 2003). Thus, functional variation among mycorrhizal types in their nutrient

83 uptake capacities from litter leads to predictions that fire-related processes are

- 84 related to the dominant mycorrhizal association.
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Even more interesting is the extent to which the activity of mycorrhizal fungi 86 87 influences the decomposition rate via their effect on the traits of their plant 88 partner. Mycorrhizal fungi form an important part of many species' mineral nutrient uptake strategy, and this is linked to the species economic strategy with 89 90 special implications for the both leaf and root traits (Reich et al. 1999, Hobbie 91 etc.). The traits of the plant while alive have important implications for the 92 characteristics of the dead and dying organic matter produced: many living traits 93 have subsequent effects following leaf abscission, affecting important processes 94 in the litter layer (Cornwell et al. 2008). For instance, Cornellissen et al. (2001) 95 linked variation in ecosystem carbon cycling to differences among plants of 96 differing mycorrhizal types, largely via effects on plant growth rates and leaf 97 nutrient contents. In addition, leaf traits have been hypothesised to have co-98 evolved in a coordinated fashion with root traits according to the plant economic 99 spectrum (Tjoelker et al., 2005) suggesting a further link between mycorrhizal 100 type and leaf traits linked to flammability. 101 102 Here, we assess the relationships among leaf traits and host mycorrhizal type for 103 77 species of Australian trees and shrubs to determine whether mycorrhizal type 104 is associated with variation in three aspects of leaf flammability (ignitability, fire 105 duration, and smoulder duration). Leaf trait and flammability data were obtained from two studies: a survey of tree and shrub species growing in gardens in the 106 107 Australian Capital Territory (Gill and Moore, 1996), and a survey of tree and 108 shrub species from four vegetation types in two rainfall zones in New South 109 Wales (Grootemaat et al., 2015a, 2015b); data represent means for each species in each study, calculated from at least ten leaves per sample type. Grootemaat et 110 111 al. measured time to ignition (seconds), flame duration (seconds), and smoulder duration (seconds) on fresh green leaves ('fresh'), and on green ('dried') and 112 113 senesced ('senesced') leaves dried at 37 °C for 72 h. Gill and Moore measured 114 time to ignition (seconds) on 'fresh' and 'dried' (at 95 °C for at least 22 h). 115 Mycorrhizal type was assigned to each species based on classifications at the 116 genus level in Brundrett (2009), as depicted at 117 http://mycorrhizas.info/ozplants.html (accessed 5 April 2016) and additional 118 sources (Supplementary Table 1). All but one species in these data were 119 classified as either 'AM', 'ECM-AM' or 'NM'; for this analysis, Dodonaea viscosa 120 was reclassified from 'AM(NM)' to 'AM'. 121

Trait and flammability variables were log-transformed as performed by 122 123 Grootemaat et al. (2015a), where appropriate. Structural equation models 124 (SEMs) where used to assess whether mycorrhizal type was associated with leaf 125 flammability measures via leaf trait associations or via direct paths (indicating 126 associations mediated via unmeasured variables). Four models (Supplementary 127 Figure 1) were compared for each combination of sample type and leaf 128 flammability measure, each differing in the direct effects on flammability that 129 were included: (i) leaf traits and both the ECM-AM and NM types; (ii) leaf traits 130 and the ECM-AM (but not NM) type; (iii) leaf traits and the NM (but not ECM-AM) 131 type, and; (iv) only leaf traits. All models included indirect effects (via leaf traits) 132 of mycorrhizal type on flammability; we used a binary variable to represent each

133 of the ECM-AM and NM types in the models, including an independent path from each binary variable to each leaf trait and (where appropriate) flammability 134 135 measure. A variable representing the AM type was omitted from each model, 136 thus NM and ECM-AM paths are expressed relative to this missing class. We chose this approach due to the ancestral status of the AM type (Wang and Qiu, 137 138 2006) and evidence that the ECM/ECM-AM type and NM type likely represent 139 independent evolutionary events (Maherali et al., 2016). SEMs were fit and 140 evaluated in 'R' version 3.3.0 (R Core Team, 2016) using the 'lavaan' package 141 (Rosseel, 2012). For each combination of sample type and leaf flammability 142 measure, the model with the lowest Bayesian Information Criterion (BIC) was 143 selected for estimation of mycorrhizal type effects, presented below. Further 144 details of the methods used during model selection and model evaluation are 145 given in the supplemental online materials, with model fit statistics presented in 146 Table 1 and Supplementary Table 2.

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148 Several significant relationships were observed among leaf traits and 149 mycorrhizal type. These were weak for the leaf traits most strongly associated 150 with flammability ($R^2_{SLA^*[fresh]} = 0.10$, $R^2_{SLA^*[dried]} = 0.11$, $R^2_{moisture[fresh]} = 0.06$, R^2 151 mass[dried] = 0.06; Figure 1, Supplementary Table 3). SLA*, which is negatively 152 correlated with all three measures of leaf flammability, was generally 32% lower 153 for NM species than for AM species. However, this was only significant in models 154 using data from both studies, where sample sizes were much larger [n > 70]; in these studies, standardised partial regression coefficients (± standard error) for 155 156 the NM type were -0.35 ± 0.11 (fresh leaves; *P* = 0.002) and -0.38 ± 0.12 (fresh dried leaves; P = 0.001) for the NM type. SLA* was 19% lower for the ECM-AM 157 158 type relative to the AM type in these same models, but these relationships were 159 marginally nonsignificant: -0.23 ± 0.13 (fresh leaves: P = 0.059) and -0.22 ± 0.12 (fresh dried leaves; P = 0.081). Leaf moisture content, which is positively 160 correlated with time to ignition, was 25% lower in fresh leaves for ECM-AM 161 species than for AM species (-0.29 ± 0.12 , P = 0.014). Leaf mass, which is 162 positively associated with time to ignition but is negatively associated with both 163 164 flame duration and smoulder duration, was 82% higher for ECM-AM species than 165 AM species (0.25 ± 0.13) . However, this latter relationship was of marginal 166 statistical significance (P = 0.049), highly variable (Figure 1d-f), and only 167 observed when data from both studies were included in the model. 168

169 Significant associations between mycorrhizal type and tissue chemistry were 170 also observed and were generally stronger in the amount of variation explained 171 relative to the other leaf traits (R^2 N[dried] = 0.64, R^2 P[senesced] = 0.54, R^2 lignin[senesced] =

0.18; Figure 2, Supplementary Table 3). Flame duration is negatively associated
with concentrations of both P and N within leaves (Scarff et al., 2012;

Grootemaat et al., 2015a). Leaves from NM species were reduced in N (by 50%)

175 and P (by 67%) compared with AM species (N: -0.90 \pm 0.07, *P* < 0.001; P: -0.83 \pm

176 0.09, P < 0.001); note that all NM species in the subset for this analysis were from

177 the Proteaceae. ECM-AM species were 16% lower in N compared with AM

species (-0.36 \pm 0.12, *P* = 0.003) and also 27% lower in P (-0.25 \pm 0.14), but this latter effect was marginally nonsignificant (*P* = 0.079). Lignin concentrations are

179 latter effect was marginally nonsignificant (P = 0.079). Lignin concentrations are 180 positively associated with smoulder duration (Grootemaat et al. 2015a) and

were higher in senesced leaves of ECM-AM (by 36%; 0.44 ± 0.16 , P = 0.007) and

NM (by 35%; 0.39 ± 0.17 , P = 0.019) species than in AM species. Effects of 182 mycorrhizal type on tannin concentrations were not observed. 183 184 185 These analyses do not exclude the possibility that relationships between leaf traits and mycorrhizal types are simply due to the shared influence of 186 187 evolutionary history on each character, and not due to any causal relationship (regardless of the direction). To address this, we assessed the extent to which 188 189 mycorrhizal type explained variation in leaf traits after accounting for the effect 190 of shared evolutionary history in four partial Mantel tests, one for each 191 combination of trait type (associated with leaf ignitability or burn duration) and 192 shift in mycorrhizal type (AM to ECM-AM or AM to NM). Pairwise phylogenetic 193 distances were obtained from the plant phylogeny published in Zanne et al. 194 (2014), using Phylomatic Version 3 (http://phylodiversity.net/phylomatic/, 195 accessed 20 June 2017). Analyses were performed in R using functions from 'ape' 196 (Paradis et al., 2004) and 'vegan' (Oksanen et al., 2016) libraries; more details of 197 the analysis are provided in the supplemental online materials. We found 198 evidence for the relationship between mycorrhizal type and leaf traits associated 199 with ignitability to be at least partially unrelated to shared evolutionary history 200 for both the ECM-AM type (r = 0.06, P = 0.008) and the NM type (r = 0.10, P =201 0.004). The same was the case for leaf traits associated with burn duration for 202 the NM type (r = 0.39, P = 0.001) but not for the ECM-AM type (r = 0.06, P = 0.13). 203 Thus, in most cases, there does appear to be a direct relationship between mycorrhizal type and leaf traits associated with flammability. 204 205 206 It is unlikely that measurement of other leaf traits would improve predictions of 207 the relationship between leaf flammability and mycorrhizal type. This is because the SEMs were generally not improved by including paths directly linking 208 209 mycorrhizal type to any of the leaf flammability measures, with the exception of flame duration for dried leaves (see outcomes of model comparisons in Table 1). 210 211 In this latter case, nonmycorrhizal plants produced leaves that burned more 212 rapidly independently of other leaf traits (leaf mass, leaf [P], SLA; -0.26 ± 0.10 , P 213 = 0.009). However, all of the nonmycorrhizal hosts in this subset of the data 214 belonged to the Proteaceae, raising the question of whether this is an effect of 215 mycorrhizal type or plant taxon. Taken together, these results suggest that 216 mycorrhizal type does not influence leaf flammability, during growth and 217 senescence, independently of the leaf traits analysed here. 218

219 These observations suggest the possibility that mycorrhizal associations, during 220 leaf growth and senescence, may influence leaf flammability as a result of their 221 link with leaf traits that have been identified as determinants of leaf flammability 222 (Figure 3). If so, their effects may mediate or exacerbate fire likelihoods if not 223 accounted for, affecting our ability to predict ecosystem carbon balances. The patterns observed here also suggest that mycorrhizal type may play a larger role 224 225 in influencing fire-related processes via litter decomposability and burn duration than via ignitability (Figure 3). However, this study is only a first step towards 226 227 the development of a comprehensive understanding of mycorrhizal fungal 228 contributions to fire-related processes in ecosystems. Below, we propose four 229 questions that will further progress toward this understanding. 230

231 (1) How reproducible are these patterns for non-Australian plant species? – The 232 observations here are limited to a set of tree and shrub species sampled from 233 eastern Australia. Fire-related processes are understandably at the forefront of 234 ecosystem management and monitoring issues in Australia, resulting in a 235 relatively rich availability of data to address these questions. However, these 236 patterns may not be generalisable. For instance, dual ECM and AM associations 237 are relatively common in Australia but occur less frequently on other continents 238 (Brundrett, 2009), and it remains to be seen whether solely ECM trees exhibit 239 flammability-associated leaf traits that are similar to ECM-AM trees. There were 240 also few ERM plant species and no representatives from other mycorrhizal types 241 represented in the data, preventing us from estimating the contribution of other 242 mycorrhizal types to leaf flammability, even though they are frequently observed 243 in some fire-prone ecosystems (e.g., South African fynbos, van de Venter and 244 Esterhuizen, 1988; California chapparal, Keeley, 1987). In addition, the 245 mechanisms that affect evolution of plant traits that affect flammability are still 246 an area of controversy (e.g., Bond and Keeley, 2005; Schwilk and Ackerly, 2001). 247 However, the occurrence of relationships between mycorrhizal type and traits 248 associated with flammability does not require leaf flammability to be directly 249 under selection. Instead, this variation might arise as a result leaf traits evolving 250 in response to other agents of selection, with indirect effects on fire regimes. 251

252 (2) Does fire frequency and intensity determine the predominant mycorrhizal association within an ecosystem? – The primary assumption behind the analyses 253 254 described above is that the mycorrhizal type of the host plant is contributing toward the suite of leaf traits observed for that tree or shrub species. However, 255 256 an argument could be made that the direction of causality is actually reversed: that leaf traits associated with variation in fire frequency and intensity have 257 258 selected for stronger associations with one type of mycorrhizal fungus, leading to 259 that strategy being more successful in particular ecosystems. Studies of mycorrhizal fungal recolonisation of ecosystems following fire events suggest 260 that AM fungal communities may have some level of resistance or may recover 261 rapidly following fire (e.g., Bellgard et al., 1994; Treseder et al., 2004), although 262 fires that raise soil temperatures substantially and/or for prolonged periods of 263 time are likely to negatively affect AM fungi (Pattinson et al., 1999). Studies of 264 ECM systems reveal relatively slow recoveries of ECM fungal communities post-265 fire (Stendell et al., 1999; Treseder et al., 2004; but see Baar et al., 1999 and Peay 266 et al., 2009). In ecosystems where fire frequency is high, the AM symbiosis may 267 268 be a successful strategy for mycorrhizal-mediated nutrient uptake in the post-269 fire soil environment, particularly for resprouting trees and shrubs that might 270 benefit from interactions with intact networks of mycorrhizal fungal hyphae, but 271 much more work is needed to evaluate this hypothesis.

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(3) To what degree does the prevalence of crown-fire and surface-fire within
ecosystems affect the importance of mycorrhizal type in explaining fire-related
processes? - Fire behaviours are likely linked to a complex interaction of the
environment and leaf traits. Simply looking at leaf traits associated with species
classified into different mycorrhizal types might not provide the best insight into
how mycorrhizal fungi mediate fire behaviour or plant nutrition in a post-fire
environment. For instance, leaves that produce litter with low decomposability,

280 low packing ratios, and high fire sustainability may be the most important components of surface fires (Cornwell et al., 2009; Grootemaat et al., 2017). 281 282 while transitions to crown fires may be more strongly associated with high leaf 283 ignitability, linked with leaf moisture content and leaf mass (Grootemaat et al., 2015a; Belcher and Hudspith, 2017). In the analyses reported here, associations 284 285 between mycorrhizal type and different leaf traits (specific leaf area and leaf moisture content) sometimes had conflicting effects that largely negated the 286 287 overall indirect effect on leaf flammability (Figure 3). Thus, modelling fire 288 behaviour using mycorrhizal type requires additional knowledge regarding the 289 fire system that is common to the ecosystem being examined, since the degree 290 that flammability changes will depend on which trait exhibits the strongest 291 association with flammability in that system. Fire-prone ecosystems 292 characterised by surface fires following the build-up of a flammable litter bed 293 may be more affected by the decomposability of litter from mycorrhizal plants. 294 Alternatively, mycorrhizal fungal effects may play a larger role than previously 295 appreciated for initiation of crown fires given effects on fresh leaf moisture 296 content.

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298 (4) How important is variation within fungal communities of different 299 mycorrhizal types for leaf and litter flammability? – This analysis does not 300 exclude the potential for variation in the composition of mycorrhizal fungal 301 communities from influencing leaf flammability via effects on host fitness and leaf moisture content during, for example, drought stress (Augé, 2001). 302 303 Substantial leaf trait variation was observed within each mycorrhizal type, which 304 might be associated with variation in the composition or functionality of 305 mycorrhizal fungal communities associated with each host tree. Different 306 mycorrhizal communities are associated with variation in plant nutritional status 307 (Johnson et al., 2010), which is a key component driving the plant economic trait spectrum. More work is also necessary to address the potential for mycorrhizal 308 309 fungal community variation to affect litter flammability given the potential for 310 mycorrhizal fungal species to vary in their abilities to extract nutrients and water 311 from standing litter (possibly linked to, for example, exploration types; Agerer, 312 2001) or to resist or recover from fire events (Baar et al., 1999; Peay et al., 2009). 313 Taking a broader view, variation exists among different mycorrhizal types in the 314 strength of observed plant-soil feedbacks, with consequences for the 315 composition of plant and fungal communities (Bennett et al., 2017; Teste et al., 316 2017). Variation in the strengths of these feedbacks could further influence 317 spatial and temporal variation in leaf and litter traits, with consequences for 318 flammability at the ecosystem level. 319 320 To summarise, we observed that classification of Australian tree and shrub 321 species into one of three mycorrhizal types could be used to predict variation in

leaf traits associated with measures of leaf flammability. This observation opens
the door to mycorrhizal fungi being linked to feedbacks in fire frequency and
intensity via direct relationships with their host plants, in addition to previous
hypotheses regarding how they may alter flammability of standing litter. We

have not demonstrated that mycorrhizal fungi are, in reality, an important

327 component of these feedbacks, but we have suggested further research that

328 could help to determine the role that they play in fire-prone ecosystems.

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341	References
342	
343	Agerer, R., 2001. Exploration types of ectomycorrhizae. Mycorrhiza 11, 107–114.
344	doi:10.1007/s005720100108
345	Augé, R.M., 2001. Water relations, drought and vesicular-arbuscular mycorrhizal
346	symbiosis. Mycorrhiza 11, 3–42. doi:10.1007/s005720100097
347	Averill, C., Turner, B.L., Finzi, A.C., 2014. Mycorrhiza-mediated competition
348	between plants and decomposers drives soil carbon storage. Nature 505,
349	543–545. doi:10.1038/nature12901
350	Baar, J., Horton, T.R., Kretzer, A.M., Bruns, T.D., 1999. Mycorrhizal colonization of
351	<i>Pinus muricata</i> from resistant propagules after a stand-replacing wildfire.
352	New Phytol. 143, 409–418. doi:10.1046/j.1469-8137.1999.00452.x
353	Belcher, C.M., Hudspith, V.A., 2017. Changes to Cretaceous surface fire behaviour
354	influenced the spread of the early angiosperms. New Phytol. 213, 1521–
355	1532. doi:10.1111/nph.14264
356	Bellgard, S.E., Whelan, R.J., Muston, R.M., 1994. The impact of wildfire on
357	vesicular-arbuscular mycorrhizal fungi and their potential to influence the
358	re-establishment of post-fire plant communities. Mycorrhiza 4, 139–146.
359	doi:10.1007/BF00203532
360	Bending, G.D., Read, D.J., 1995. The structure and function of the vegetative
361	mycelium of ectomycorrhizal plants. New Phytol. 130, 401–409.
362	dol:10.1111/j.1469-813/.1995.tb01834.x
363	Bennett, J.A., Manerall, H., Keinnart, K.O., Lekberg, Y., Hart, M.M., Kiironomos, J.,
265	2017. Plant-soli leeubacks and mycorrnizal type initience temperate
266 266	doi:10.1126 / science.aci9212
267	Rond WI Koolow IF 2005 Fire as a global "horbiyore", the acology and
268	ovolution of flammable access tome. Trands Ecol. Evol. 20, 297, 204
360	doi:10.1016/i troo 2005.04.025
370	Brundrett MC 2009 Mycorrhizal associations and other means of nutrition of
371	vascular plants: understanding the global diversity of host plants by
371	resolving conflicting information and developing reliable means of
373	diagnosis Plant Soil 320 37–77 doi:10.1007/s11104-008-9877-9
374	Cairney IWG 2011 Ectomycorrhizal fungi: the symbiotic route to the root for
375	phosphorus in forest soils. Plant Soil 344, 51–71. doi:10.1007/s11104-
376	011-0731-0
377	Chalot, M., Brun, A., 1998. Physiology of organic nitrogen acquisition by

378	ectomycorrhizal fungi and ectomycorrhizas. FEMS Microbiol. Rev. 22, 21–
379	44. doi:10.1111/j.1574-6976.1998.tb00359.x
380	Cornelissen, J., Aerts, R., Cerabolini, B., Werger, M., Heijden, M. van der, 2001.
381	Carbon cycling traits of plant species are linked with mycorrhizal strategy.
382	Oecologia 129, 611–619. doi:10.1007/s004420100752
383	Cornwell, W.K., Cornelissen, J.H.C., Allison, S.D., Bauhus, J., Eggleton, P., Preston,
384	C.M., Scarff, F., Weedon, J.T., Wirth, C., Zanne, A.E., 2009. Plant traits and
385	wood fates across the globe: rotted, burned, or consumed? Glob. Change
386	Biol. 15, 2431–2449. doi:10.1111/j.1365-2486.2009.01916.x
387	Gadgil, R.L., Gadgil, P.D., 1971, Mycorrhiza and Litter Decomposition, Nature 233,
388	133–133. doi:10.1038/233133a0
389	Gill, A.M., Moore, P.H., 1996. Ignitibility of leaves of Australian plants. Australian
390	Flora Foundation. CSIRO.
391	Grootemaat S. Wright II. van Bodegom P.M. Cornelissen I.H.C. Cornwell W.K.
392	2015a Burn or rot: leaf traits explain why flammability and
393	decomposability are decoupled across species Funct Ecol 29 1486-
394	1497 doi:10 1111/1365-2435 12449
395	Grootemaat S Wright II van Bodegom PM Cornelissen IHC Cornwell WK
396	2015h Data from: Burn or rot: leaf traits explain why flammability and
397	decomposability are decoupled across species doi:10.5061/drvad m41f1
398	Grootemaat S Wright II van Bodegom PM Cornelissen IHC 2017 Scaling
399	un flammability from individual leaves to fuel beds. Oikos, in press
400	doi 10 1111 /oik 03886
400	Johnson N.C. Wilson G.W.T. Bowker M.A. Wilson I.A. Miller R.M. 2010
401	Resource limitation is a driver of local adaptation in mycorrhizal
402	symbioses PNAS 107 2093–2098 doi:10.1073/nnas.0906710107
403	Joner FL Aarle IM van Vosatka M 2000 Phosphatase activity of extra-radical
404	arhuscular mycorrhizal hynhae. A review Plant Soil 226 199–210
405	doi:10.1023/A:1026582207192
407	Keeley J.F. 1987 Role of fire in seed germination of woody taxa in California
407	chanarral Ecology 68 434-443 doi:10.2307/1939275
400	Koide RT Wu T 2003 Ectomycorrhizes and retarded decomposition in a
410	Pinus resinosa plantation New Phytol 158 401-407 doi:10.1046/i.1469-
410	$8137\ 2003\ 00732\ v$
412	Maherali H. Oherle R. Stevens P.F. Cornwell W.K. McGlinn D.L. 2016
412 /13	Mutualism Persistence and Abandonment during the Evolution of the
413	Mutualish refisice and Abandonment during the Evolution of the Mycorrhizal Symbiosis Am Nat 188 F113-F125 doi:10.1086/688675
415	Oksanen I. Blanchet F.C. Friendly, M. Kindt R. Legendre, P. McGlinn, D.
415	Minchin P.R. O'Hara R.R. Simnson G.I. Solymos P. Stevens M.H.H.
410	Szoecs F. Wagner H. 2016 yegan: Community Ecology Package version
417	2 <i>A</i> _1
410	Daradis F. Claudo I. Strimmor K. 2004 ADE: Analyses of Dhylogonotics and
419	Fundution in D language, Picinformatics 20, 200, 200
420	doi:10.1002 /bioinformatics /btg/12
421	Dattingon C.S. Hammill V.A. Sutton P.C. Magoo D.A. 1000 Simulated fire
422	raunson, G.S., Hannin, K.A., Suuon, D.G., MCgee, F.A., 1999. Sinulated life
423 424	Mucol Dog 102 401 406 doi:10.1017/S00E27E6200007412
+∠4 125	Wy UUI. Res. 103, 471-470. UUI. 101.101/30733/3027000/412 $Popu KC Carbolotto M Prupa TD 2000 Shore hast resistance playe or$
425	i cay, K.u., Gai Delotto, M., Di ulis, T.D., 2009. Spore field resistance plays an
7 20	important role in disturbance-inculated assemblage sinit of

427	ectomycorrhizal fungi colonizing <i>Pinus muricata</i> seedlings. J. Ecol. 97.3,				
428	537-547. doi:10.1111/j.1365-2745.2009.01489.x				
429	Peoples, M.B., Brockwell, J., Herridge, D.F., Rochester, I.J., Alves, B.J.R., Urquiaga,				
430	S., Boddey, R.M., Dakora, F.D., Bhattarai, S., Maskey, S.L., Sampet, C.,				
431	Rerkasem, B., Khan, D.F., Hauggaard-Nielsen, H., Jensen, E.S., 2009. The				
432	contributions of nitrogen-fixing crop legumes to the productivity of				
433	agricultural systems. Symbiosis 48, 1–17. doi:10.1007/BF03179980				
434	Phillips, R.P., Brzostek, E., Midgley, M.G., 2013. The mycorrhizal-associated				
435	nutrient economy: a new framework for predicting carbon-nutrient				
436	couplings in temperate forests. New Phytol. 199, 41–51.				
437	doi:10.1111/nph.12221				
438	R Core Team, 2016. R: A language and environment for statistical computing. R				
439	Foundation for Statistical Computing, Vienna, Austria. URL				
440	https://www.R-project.org/				
441	Rosseel, Y., 2012. lavaan: An R package for structural equation modeling. J. Stat.				
442	Softw. 48, 1–36.				
443	Scarff, F.R., Gray, B.F., M. Westoby, M. 2012. Exploring phosphate effects on leaf				
444	flammability using a physical chemistry model. Int. J. Wildland Fire 21,				
445	1042-1051. doi:10.1071/WF09065				
446	Schwilk, D.W., Ackerly, D.D., 2001. Flammability and serotiny as strategies:				
447	correlated evolution in pines. Oikos 94, 326–336. doi:10.1034/j.1600-				
448	0706.2001.940213.x				
449	Shane, M.W., Lambers, H., 2005. Cluster Roots: A Curiosity in Context. Plant Soil				
450	274, 101–125. doi:10.1007/s11104-004-2725-7				
451	Stendell, E.R., Horton, T.R., Bruns, T.D., 1999. Early effects of prescribed fire on				
452	the structure of the ectomycorrhizal fungus community in a Sierra Nevada				
453	ponderosa pine forest. Mycol. Res. 103, 1353–1359.				
454	doi:10.1017/S0953756299008618				
455	Talbot, J.M., Allison, S.D., Treseder, K.K., 2008. Decomposers in disguise:				
456	mycorrhizal fungi as regulators of soil C dynamics in ecosystems under				
457	global change. Funct. Ecol. 22, 955–963. doi:10.1111/j.1365-				
458	2435.2008.01402.x				
459	Teste, F.P., Kardol, P., Turner, B.L., Wardle, D.A., Zemunik, G., Renton, M.,				
460	Laliberte, E., 2017. Plant-soli feedback and the maintenance of diversity in				
461	Mediterranean-climate shrublands. Science 355, 1/3–1/6.				
462	dol:10.1126/science.aal8291 Tiaallaar M.C. Craina I.M. Wadin D. Daiah D.P. Tilman D. 2005 Linking loof				
403	I JOEIKER, M.G., Craine, J.M., Wedin, D., Reich, P.B., Thiman, D., 2005. Linking leaf				
404	The and root trait synuromes among 59 grassiand and savannan species. New				
405	Tracador KK Mack MC Cross A 2004 Polationships among fires fungi and				
400	soil dynamics in Alaskan baroal forests Ecol Appl 14 1926 1929				
407	doi:10.1890/03-5133				
469	van de Venter H Δ Esterbuizen Δ D 1988 The effect of factors associated with				
470	fire on seed germination of <i>Frica</i> sessiliflora and <i>F</i> hehecally (Fricaceae)				
471	South Afr I Bot 54 301–304 doi:10.1016/S0254-6299(16)31330-8				
472	Wang, B., Oiu, YL., 2006. Phylogenetic distribution and evolution of mycorrhizas				
473	in land plants. Mycorrhiza 16, 299–363. doi:10.1007/s00572-005-0033-6				
474	Zanne, A.E., Tank, D.C., Cornwell, W.K., Eastman, I.M., Smith, S.A., FitzIohn, R.G.,				
475	McGlinn, D.J., O'Meara, B.C., Moles, A.T., Reich, P.B., Royer, D.L., Soltis, D.E.,				

476	Stevens, P.F., Westoby, M., Wright, I.J., Aarssen, L., Bertin, R.I., Calaminus,
477	A., Govaerts, R., Hemmings, F., Leishman, M.R., Oleksyn, J., Soltis, P.S.,
478	Swenson, N.G., Warman, L., Beaulieu, J.M., 2014. Three keys to the
479	radiation of angiosperms into freezing environments. Nature 506, 89–92.
480	doi:10.1038/nature12872
481	
482	

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485	Table 1. Comparisons of model fit (using the Bayesian Information Criterion
486	scores) among models with and without direct paths linking mycorrhizal type
487	with measures of leaf flammability, representing statistical associations
488	mediated via unmeasured variables. The best model for each response variable
489	within each material type is that with the lowest score (indicated in bold). All
490	models included paths representing direct effects of mycorrhizal type on leaf
491	traits and, thus, indirect effects of mycorrhizal type on leaf flammability
492	measures. Additional fit measures and model parameters are included in
493	Supplementary Table 2 of the supporting online material.
494	

	Material type		
Direct effects	Fresh	Dried	Senesced
	Log_{10} time to ignition		
leaf traits + ECM-AM + NM	1018.04	266.46	87.59
leaf traits + ECM-AM	1013.69	263.04	84.69
leaf traits + NM	1013.98	265.13	84.62
leaf traits	1009.62	260.87	81.37
	Log ₁₀ flame duration		
leaf traits + ECM-AM + NM	85.86	49.95	78.27
leaf traits + ECM-AM	82.41	52.85	75.17
leaf traits + NM	82.44	47.23	74.96
leaf traits	79.03	49.94	72.32
	Log ₁₀ smoulder duration		
leaf traits + ECM-AM + NM	293.45	469.88	495.79
leaf traits + ECM-AM	290.36	468.08	493.56
leaf traits + NM	292.91	467.05	492.46
leaf traits	289.50	464.69	490.19





Figure 1. Violin plots demonstrating average trait values associated with leaves sampled from tree and shrub species, classified by mycorrhizal association. Traits were measured on fresh green leaves (a, d, g), dried green leaves (b, e, h), and dried senesced leaves (c, f, i). Width of the plots represent the calculated density at a particular trait value and the horizontal line represents the median observation within each group.



516 Figure 2. Violin plots demonstrating average trait values associated with leaves 517 sampled from tree and shrub species, classified by mycorrhizal association.

- 518 Traits were measured on dried green leaves (a, c, e, g) and dried senesced leaves
- 519 (b, d, f, h). See figure 1 caption for further details.
- 520



Figure 3. Conceptual figure summarising the relationships between differences in tree and shrub mycorrhizal type, leaf traits, and leaf flammability. Evolution of the ECM-AM and NM association from AM ancestors are related to shifts in leaf traits that increase burn duration, but have weak or uncertain effects on leaf ignitability. These trait shifts also have consequences for litter decomposability (e.g., Cornelissen et al. 2001) and litter bed packing, which influences litter

- 529 (e.g., Cornelissen et al. 2001) and litter bed packing, which influences litt
- 530 flammability (e.g., Grootemaat et al. 2017).