

1 ***Survival and migration of rock ptarmigan in central Scandinavia***

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14  
15 **Abstract**

16 In a world undergoing massive declines in the distribution and abundance of many wildlife  
17 species, documenting basic ecological characteristics is often needed to be able to understand  
18 and potentially mitigate current and future pressures. Species living in alpine areas might be  
19 particularly vulnerable to climate change, in part because they are less likely to be able to  
20 migrate to new suitable areas. Here we report from a two year case study of rock ptarmigan  
21 (*Lagopus muta*) in central Scandinavia. Ptarmigan were captured in winter (n=84), and fitted  
22 with radio collars. We estimated the natural survival from mid-winter to late summer to be  
23 0.55 (SE: 0.07), with no distinct differences between juveniles and adults, sex, or between the  
24 two years. Natural survival through late winter (February – April) was estimated at 0.77 (SE:  
25 0.05), survival through breeding season May-July at 0.65 (SE: 0.08), and harvest mortality  
26 through the February winter harvest at 9% (SE: 3%). Moreover, we documented large scale  
27 movement from the wintering grounds before the breeding season in the spring. The longest  
28 recorded movement was 79.5 km, and the mean distance from the capture site for birds still  
29 in the sample in May-July was 20.3 (SD: 18) km. We discuss the implications of the results in  
30 terms of ongoing climate change.

31  
32  
33 **Running title:** Rock ptarmigan demography in Scandinavia

## 34 **1. Introduction**

35 Species inhabiting alpine or polar habitats are expected to be particularly affected by ongoing  
36 climate change (Post et al., 2009; Revertmann et al., 2012). This is partly because these areas  
37 are likely to become much warmer in the future, but also because species inhabiting such  
38 areas have less possibilities to find new locations to thrive (Sirami et al., 2017). Among the few  
39 resident birds in these habitats, the rock ptarmigan (*Lagopus muta*) has a wide, circumpolar  
40 distribution across the northern tundra and high-mountain regions (Storch, 2007). Rock  
41 ptarmigan as a species are expected to be strongly affected by climate change (Booms et al.,  
42 2012; Revertmann et al., 2012; Hansen et al., 2013), and are therefore a suitable model species  
43 for examining climate change effects on alpine wildlife populations. Globally rock ptarmigan  
44 is at present considered least concern in the international Red List of Species  
45 (BirdLife\_International, 2016), but locally there are concerns about declining populations. In  
46 2015 the species was classified as near threatened (NT) in the Norwegian Red List of Species  
47 (Henriksen and Hilmo, 2015).

48

49 Compared to other grouse species, the rock ptarmigan are among the least studied species in  
50 terms of number of scientific publications (Moss et al., 2010). Rock ptarmigan ecology and  
51 demography has earlier been studied in e.g. UK (e.g. Watson et al. 1998), North-America  
52 (Wilson and Martin, 2008; 2010), Russia (e.g. Potapov and Potapov, 2012), Japan (e.g. Suzuki  
53 et al., 2013), high-alpine areas in the Italian Alps, French Alps and Pyrenees (Scherini et al.,  
54 2003; Novoa et al., 2008; Novoa et al., 2011), Iceland (e.g. Nielsen, 1999) and the archipelago  
55 of Svalbard (e.g. Pedersen et al., 2012; Unander et al., 2016). However, from an important  
56 part of their distribution across mainland Scandinavia, only aspects relating to population  
57 genetics (e.g. Costanzi and Steifetten, 2019), habitat use (Pedersen et al., 2013) and  
58 population dynamics based on harvest bag data (Kvasnes et al., 2010) has been studied.  
59 Baseline demography and ecology, including spatial behavior, is largely unknown.

60

61 In this case study, we provide insight into some key aspects of rock ptarmigan demography  
62 and ecology in a study area in central Scandinavia (**Figure 1**). The case study is based on 84  
63 birds marked with VHF collars in 2012-2013, and followed until they were either recorded as  
64 dead, or until we lost contact (i.e censored). Although our study is completely descriptive and  
65 does not intend to test any specific hypothesis about rock ptarmigan ecology or demography,  
66 there are two specific aspects that we focus on:

- 67 1) First, we use known fate models to estimate survival probabilities. We investigate to  
68 which extent survival probability differs between years, age (juvenile vs adult) and sex  
69 (females vs males). We estimate survival probabilities during a six month period from  
70 February-July, as well as for late winter (settlement period) and in the breeding period.  
71 We also report harvest mortality rates for the winter harvest season in February.
- 72 2) Second, we estimate movement rates away from the wintering grounds, using mean  
73 displacement rates as our estimator of interest. The rationale for this was to identify  
74 to which extent the birds captured at their wintering grounds were stationary or  
75 moved to other areas in the breeding season.

76

77

## 78 **2. Materials and Methods**

### 79 **2.1. Study area**

80 The current study was conducted in central Norway, in the municipality Lierne (central  
81 location for our core study area: 64°25'N, 13°59'E), partly within the Lierne National Park  
82 (**Figure 1**). The study area cover both the northern boreal, low- and mid-alpine ecoregions.  
83 Poor vegetation types dominate the area, but there are also features of richer types such as  
84 intermediate marshes. The lower parts is dominated by willow (*Salix* spp.) and scattered  
85 forests of mountain birch (*Betula pubescens*), the mid areas of sedges, grasses, patches of  
86 dwarf birch (*B. nana*) and snowbed communities, while the highest parts lack continuous  
87 vegetation cover. Yearly precipitation normal is equivalent to 675 millimeter per year, while  
88 the temperature normal is -10°C for January and 12°C for July. Snowcover at the study area  
89 is 1–3 m deep during winter and persists from early October to late May. The main predators  
90 on juvenile and adult rock ptarmigan in the study area were gyrfalcon (*Falco rusticolus*) and  
91 golden eagle (*Aquila chrysaetos*), with occasional observations of red foxes (*Vulpes vulpes*),  
92 arctic foxes (*V. lagopus*) and wolverines (*Gulo gulo*). Potential additional predators on egg  
93 and chicks includes raven (*Corvus corax*) and hooded crow (*C. cornix*).  
94

### 95 **2.2. Field data collection**

96 Rock ptarmigan were captured at night in the winter of 2012-2013, with handheld spotlights  
97 and long-handled dip nets from snowmobiles. Similar approach has been used in previous  
98 studies on willow ptarmigan (*L. lagopus*) in Norway (Sandercock et al., 2011). At capture,  
99 birds were aged as juveniles or adults and sexed, and morphometric measures (weight (g)  
100 and wing length (mm)) were taken. We sexed and aged (juvenile: born the proceeding  
101 summer, adults: all older birds) the birds based on plumage coloration, wing length, and  
102 patterns of pigmentation of the outer primaries (Bergerud et al., 1963). In total, we captured  
103 and radio collared 84 rock ptarmigan (n= 44 males, n=40 females, n=50 juveniles, n=34  
104 adults) across the two seasons. Note that there was a marked difference in the proportion of  
105 juveniles among captured birds in 2012 (78%) and 2013 (24%), corresponding well with a  
106 large difference in breeding success in 2011 (high) and 2012 (low) respectively (E. B. Nilsen,  
107 pers com based on line transect data on willow ptarmigan from the study area).  
108

109 Each bird was marked with a uniquely numbered leg ring, and instrumented with a necklace  
110 VHF-radio transmitter (Holohil, 12g) with a 24 months expected battery life. The collars had  
111 mortality switches, so that we could detect when a bird was dead. Previous studies have  
112 shown that necklace radios of this size have little to no effect on the demographic  
113 parameters or movements of ptarmigan under natural conditions (see Sandercock et al.,  
114 2011 and citations therein).  
115

116 Radio collared birds were tracked from the ground or air at irregular intervals. When tracking  
117 from ground (using ski or snowmobile during winter, and on foot during spring/summer), we  
118 either conducted radio-triangulation at relatively close distances (from 50 m – to a few  
119 hundred meters) to obtain precise positions, or obtained just one signal so we could  
120 determine the state of the bird based on the mortality switch. When tracking from the air,  
121 we used either a small winged aircraft or a small helicopter. Whenever we obtained a  
122 mortality signal, we tried to locate the bird on the ground to determine the cause of death.  
123 In addition, we were notified by small game hunters in the region when a marked bird was  
124 shot (hunting is allowed from August/September-February in this part of Scandinavia).

125 Hunters cannot normally see the collars during a hunting situation, and there were no  
126 particular restrictions for hunters regarding shooting marked birds. We believe that all or  
127 most shot birds were reported, but do not have any independent data to back up this  
128 statement.

129

130 At the onset of the study, we opted to follow the birds throughout their full annual cycle, as  
131 has been done in previous studies on willow ptarmigan (Sandercock et al., 2011). However,  
132 because of the high combined loss-rates (i.e. combination of censoring and mortality), and  
133 the fact that the birds were spread across a huge area (**Figure 2**) during summer and  
134 fall/early winter, we obtained relatively few locations in that period. We therefore here  
135 restrict our analysis to the time between February 1<sup>st</sup> and July 31<sup>st</sup> each year. In the study  
136 periods, we conducted 3 and 4 flights in the years 2012 and 2013, respectively. For each  
137 individual, we had between zero and 16 relocations. Birds that were never relocated after  
138 marking (n=2) were not included in the analysis.

139

140 All data used in this study is published and openly accessible through GBIF, located here:  
141 <https://www.gbif.org/dataset/b848f1f3-3955-4725-8ad8-e711e4a9e0ac> (Nilsen et al., 2017).

142

143

### 144 **2.3 Statistical analysis**

145 Based on the data described above, we examined the two aspects of rock ptarmigan ecology  
146 and demography as outlined above. To examine survival during late winter, spring and  
147 summer (February 1<sup>st</sup> – July 31<sup>st</sup>), we first determined the entry – and exit time for each bird  
148 into the sample, at a monthly basis for each of the two years of the study period. Entry point  
149 was determined as either i) the month of capture, or ii) February for birds that survived the  
150 first year and entered into their second study year. Exit from the sample were again  
151 determined on a monthly basis, and were coded as either mortality or censored. Censored  
152 birds were either i) those that we lost contact with, or ii) those that survived until the end of  
153 the annual study period. Because we did not have continuous follow-up times, we adjusted  
154 our data to a monthly schedule. For birds of which mortality was recorded, exit was defined  
155 to occur in the mid-point between the last live contact and the first mortality signal. For  
156 birds that were censored because we lost contact, we censored the birds the first month  
157 after the last detection. We used Kaplan-Meier models (Pollock et al., 1989; Murray, 2006)  
158 to estimate survival probabilities, the non-parametric cumulative incidence function  
159 (NPCIFE) to estimate cause-specific mortality (Heisey and Patterson, 2006), and cox-  
160 proportional hazard models (Murray, 2006; Murray and Patterson, 2006) to test for  
161 differences among years (2012 vs 2013), age (juveniles vs adults) and sex (males vs females).  
162 We could not consider interactions between independent variables, because of relatively  
163 low sample sizes. We used Akaike's Information Criterion (AIC) to guide the model selection  
164 procedures (Burnham and Anderson, 2002). All analysis were conducted in program R  
165 version 3.6.0 (R Development Core Team, 2019), and survival analysis were conducted using  
166 add-on library survival (Therneau, 2010).

167

168 The R-code for performing the analysis reported here is available at Open Science  
169 Framework through an add-on connection with GitHub (<https://osf.io/gcm3u/>).

170

171 **3. Results**

172 Out of 84 radio marked rock ptarmigans, 9 were shot by hunters in the study area, and 33  
173 were recorded as dead due to other causes (mostly predation). The birds were marked in  
174 February and March, and we estimated the harvest mortality during the four weeks of  
175 February to be 0.09 (SE: 0.03).

176 Based on cox proportional hazard models, we did not find any signs that natural survival  
177 differed between years, or between age- and sex categories (AIC for null model: 224.59; AIC  
178 for model including year: 225.73; AIC for model including sex: 223.74; AIC for model  
179 including age: 226.02). For all models, the proportional hazards assumption was met (year-  
180 model:  $\chi^2=0.5$ ,  $p=0.48$ ; sex-model:  $\chi^2=0.02$ ,  $p=0.9$ ; age-model:  $\chi^2=0.93$ ,  $p=0.33$ ).  
181 Based on the pooled sample, overall probability to survive from February 1st and through  
182 July was estimated at 0.45 (SE: 0.07) (**Figure 2**), with natural survival (i.e. disregarding birds  
183 that were shot by hunters during the harvest season) estimated at 0.55 (SE: 0.07). Survival  
184 probability for late winter to start of the breeding season (February 1st - April 30th) was  
185 estimated at 0.7 (SE: 0.05), with natural survival estimated at 0.77 (SE: 0.05). Survival  
186 probability for the breeding season (May 1st -July 31st) was estimated at 0.65 (SE: 0.08).

187 Most of the birds captured and marked with radio collars left the area where they had been  
188 captured before the breeding season started in the spring (**Figure 3**). Based on the maximum  
189 displacement for birds still alive and in the sample ( $n=36$ ), mean displacement distance was  
190 estimated at 7.8 (SD: 12.2)km in April. In a pooled sample for May-July, the displacement for  
191 birds still alive and in the sample ( $n=25$ ) was estimated at 20.3 (SD: 18)km. We note that this  
192 is probably an underestimation, because we are more likely to have lost contact with birds  
193 that moved long distances. The maximum recorded movement from the capture site was  
194 79.5 km. A total of 6 birds moved longer than 50 km, of which 4 were males and 2 were  
195 females. For 16 birds, we detected movements longer than 25 km from the capture site,  
196 including 7 males and 9 females.

197

198 **4. Discussion**

199 A main contribution of our study is to provide updated information about some central  
200 aspect of rock ptarmigan biology within a central part of its distribution. Although our study  
201 is purely exploratory in nature, we assessed some basic ecological aspects of a rock  
202 ptarmigan population in Norway. We conclude that i) overwinter survival is relatively low  
203 but with no marked differences between years, age classes or sexes, and ii) that there are  
204 distinct movements between winter and summer areas, and that few birds remained  
205 resident in the wintering area.

206 Survival is a key demographic rate, and variation in survival rates will influence both short-  
207 and long term population growth rate (Caswell, 2001). In our study, overwinter (February –  
208 July) natural survival was estimated at 55%. Assuming similar mortality risk for the remaining  
209 part of the annual cycle (which might not be justified), annual survival can be extrapolated to  
210 be around 30% ( $0.55 \times 0.55$ ). Compared to annual survival probabilities for rock ptarmigan in  
211 Japan (estimated at 44-74% for birds of different ages: Suzuki et al. (2013)), France (61% and  
212 70% in Haut Giffre and Canigou Massif, respectively: Novoa et al. (2011)) and Svalbard (40-  
213 50% for males and females, respectively: Unander et al. (2016)) this is very low survival,  
214 bearing in mind that we only estimated survival for a part of the year. At Island, rock

215 ptarmigan survival was shown to be highly variable, varying between 36-65% for adult birds  
216 and constant at 19% for juveniles (Sturludottir et al., 2018). Because our study only lasted  
217 two years, we are not able to estimate robustly any between year variation due to stochastic  
218 environmental factors or variation in harvest pressure. Often, populations inhabiting alpine  
219 areas (i.e. high altitudes) have higher survival and lower reproduction compared to  
220 populations/species at lower elevations (Sandercock et al., 2005a; b). In our study area, this  
221 model would predict that rock ptarmigan should have higher survival and lower reproductive  
222 output than the willow ptarmigan inhabiting lower elevations. Our results from this short  
223 term study is not consistent with a “high survival strategy” for rock ptarmigans in our study  
224 area, but we can not conclude if this inconsistency arise due to the short term nature of the  
225 field study, or represent a more general life history strategy for rock ptarmigan in  
226 Scandinavian mountains.

227 Most birds that were captured during winter moved out of the wintering area before the  
228 breeding season in the spring. Such movements have been discussed in the literature for  
229 willow ptarmigan in central Scandinavia, with slightly different conclusions from different  
230 study areas (Brøseth et al., 2005; Hornell-Willebrand et al., 2014). For rock ptarmigan we  
231 have no comparable data from Scandinavia, but long distance movements have been  
232 documented (Gardarsson and Bossert, 1997) with flights >300km. Examining the genetic  
233 structure of rock ptarmigan in southern Scandinavia, Costanzi (Costanzi, 2019) reported low  
234 genetic differentiation among rock ptarmigan in a large contiguous mountain habitat in  
235 southern Norway, but with considerable differentiation between neighboring populations  
236 separated by unsuitable rock ptarmigan habitat. Because climate change is likely to result in  
237 reduced area occupancy and suitable habitat for rock ptarmigan in alpine areas (see e.g.  
238 Revermann et al., 2012 for an example from the Swizz alps), a plausible prediction is that the  
239 alpine islands that rock ptarmigan inhabit in the Scandinavian landscape will become both  
240 smaller and more isolated in the future. Understanding their spatial behavior, including their  
241 propensity to long scale dispersal, seasonal migration and the landscape effects on  
242 movement behavior will be key when designing future conservation plans for rock ptarmigan  
243 in Scandinavia.

244

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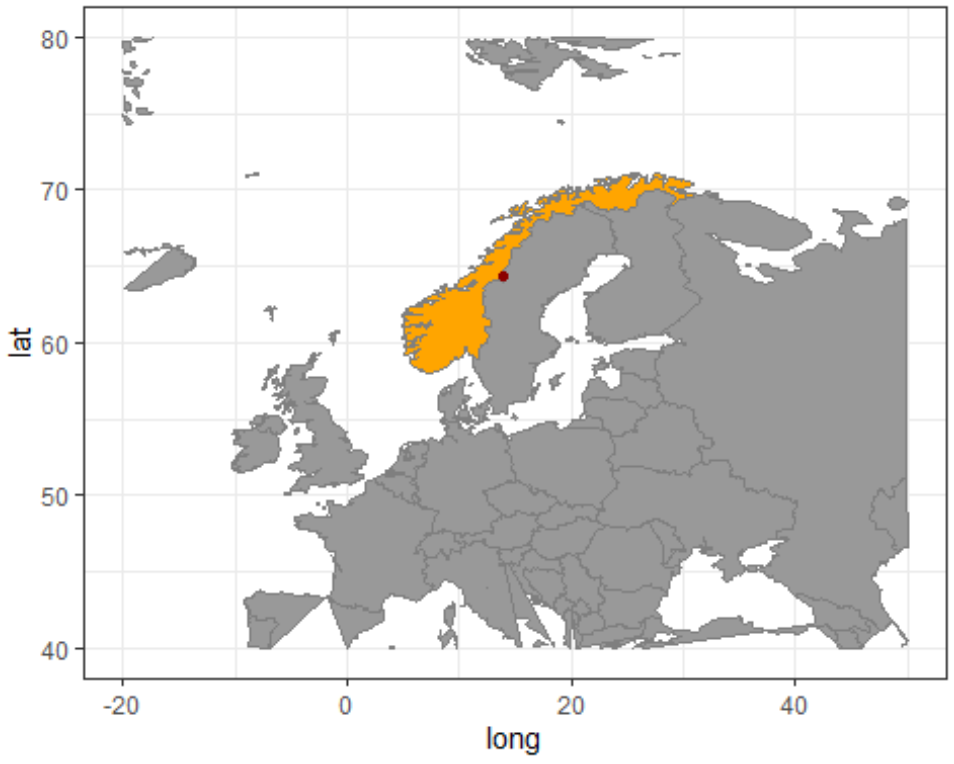
357 **Figure Legends**

358 *Figure 1:* Map of the study area, with Norway marked in orange and the study area marked  
359 with a red circle.

360 *Figure 2:* Kaplan-Meyer survival curve for a sample of rock ptarmigans radio collared in  
361 Central Norway in 2012-2013. The time on the x-axis is shifted so that week 1 represent the  
362 first week in February each year, whereas week 31 is the last week in July. Lower table  
363 depicts the number at risk throughout the study period.

364 *Figure 3:* Linear displacement from the capture site plotted against week since February 1st.  
365 Each line represent the trajectory for one individual bird, and red crosses indicates death  
366 events.

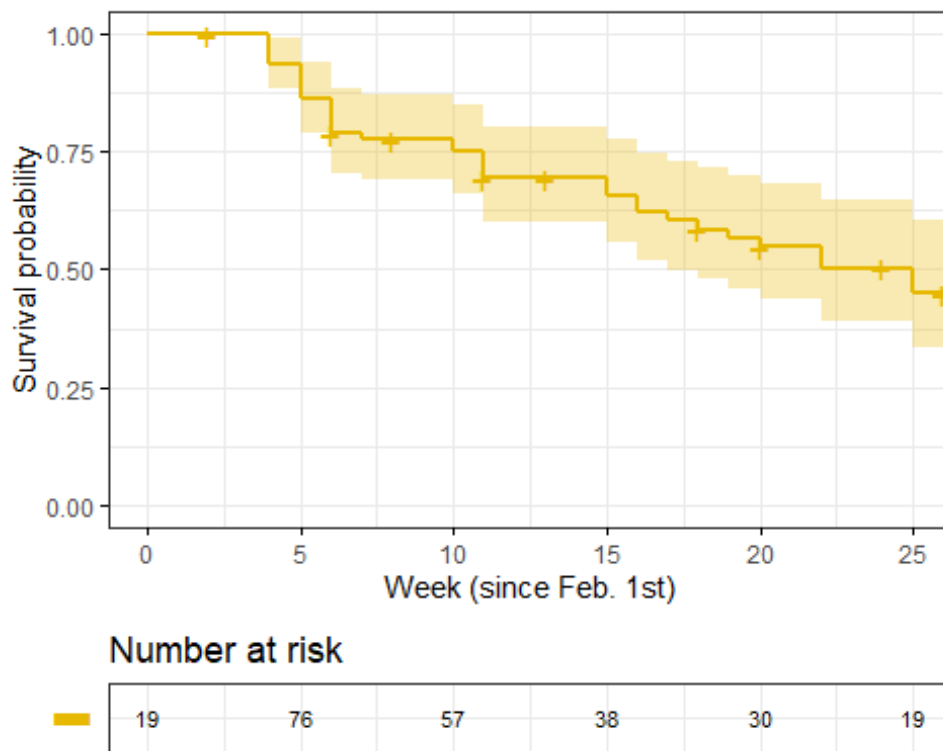
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369 **FIGURE 1**

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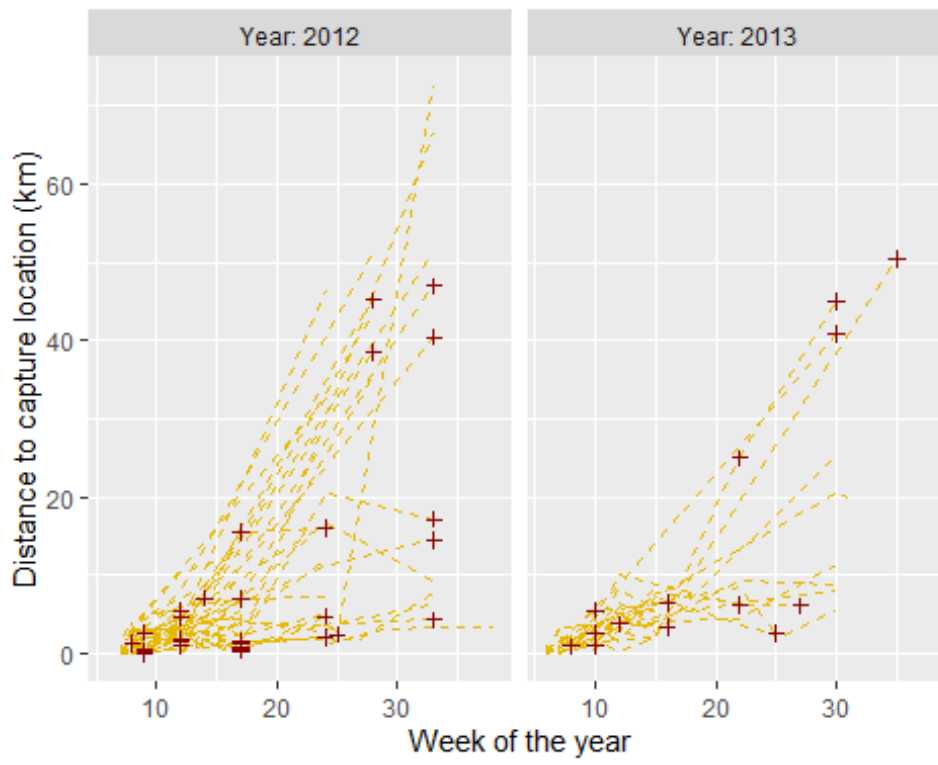


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372 **FIGURE 2**

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376 **FIGURE 3**

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