

1 **Assessing symbiont extinction risk using cophylogenetic data**

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3 Jorge Doña<sup>1</sup> and Kevin P. Johnson<sup>1</sup>

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5 1. Illinois Natural History Survey, Prairie Research Institute, University of Illinois at Urbana-Champaign,  
6 1816 S. Oak St., Champaign, IL 61820, USA

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8 \*Corresponding authors: Jorge Doña & Kevin Johnson; e-mail: jorged@illinois.edu & kpjohnso@illinois.edu

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10 **Abstract:** Symbionts have a unique mode of life that has attracted the attention of ecologists and  
11 evolutionary biologists for centuries. As a result of this attention, these disciplines have produced  
12 a mature body of literature on host-symbiont interactions. In contrast, the discipline of symbiont  
13 conservation is still in a foundational stage. Here, we aim to integrate methodologies on symbiont  
14 coevolutionary biology with the perspective of conservation. We focus on host-symbiont  
15 cophylogenies, because they have been widely used to study symbiont diversification history and  
16 contain information on symbiont extinction. However, cophylogenetic information has never been  
17 used nor adapted to the perspective of conservation. Here, we propose a new statistic,  
18 “cophylogenetic extinction rate” (*Ec*), based on coevolutionary knowledge, that uses data from  
19 event-based cophylogenetic analyses, and which could be informative to assess relative symbiont  
20 extinction risks. Finally, we propose potential future research to further develop estimation of  
21 symbiont extinction risk from cophylogenetic analyses and continue the integration of this existing  
22 knowledge of coevolutionary biology and cophylogenetics into future symbiont conservation  
23 studies and practices.

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25 **Keywords:** coevolution, coextinction risk, conservation biology, cophylogenies, host-symbiont  
26 interactions, parasites.

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28 **Highlights:**

- 29 - We apply coevolutionary knowledge to symbiont conservation biology.  
30 - We propose a new statistic to assess symbiont extinction risk from cophylogenies.  
31 - We suggest potential future opportunities for advancing the field.

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## 34 1. Introduction

35 Symbionts, defined as parasites, mutualists, and commensals that interact intimately with their  
36 hosts (Leung and Poulin, 2008), are highly relevant components of ecosystems, comprising up to  
37 75% of all ecological interactions (Lafferty et al., 2006; Dobson et al., 2008). A major concern is  
38 that conservative estimates predict that a high percentage of symbiont species are expected to go  
39 extinct due to climate change (e.g., 10 % of parasite species by 2070, Carlson et al., 2017a; Dunn  
40 et al., 2009; Koh et al., 2004). However, despite their relevance and conservation status, symbiont  
41 conservation biology is still in a foundational stage (e.g., Windsor, 1995; Dougherty et al., 2016;  
42 Rocha et al., 2016; Cizauskas et al., 2017; this special issue).

43 Many studies of symbionts have covered various aspects of their ecology and evolution  
44 (Poulin, 2011; Clayton et al., 2015), some of which may be useful for assessing symbiont  
45 extinction risks (Soulé, 1980; Groom et al., 2012). However, these studies have not considered  
46 the implications or relevance of their findings for conservation biology. Host-symbiont  
47 cophylogenies are one example of this rationale. Host-symbiont cophylogenies have been used to  
48 uncover the diversification history of many symbiont groups. However, though never used for  
49 assessing symbiont extinction risk, they might also provide valuable information on symbiont  
50 extinctions across phylogeny.

51 Symbiont (i.e. affiliated, associated, or dependent species) extinction risk has been  
52 commonly assessed in a coextinction framework, using host information (e.g., host extinction  
53 probabilities) to calculate symbiont extinction probability (Colwell et al., 2012; Dunn et al., 2009;  
54 Koh et al., 2004; Moir et al., 2014, 2010). Overall, current available approaches to estimate  
55 coextinction rates can be summarized as follows (Colwell et al., 2012; Moir et al., 2010): 1)  
56 Probabilistic models based on host likelihood of becoming extinct (e.g., estimated from historical  
57 data) that may also include some symbiont-specific traits, such as host specificity (Campião et al.,  
58 2015; Carlson et al., 2017a; Dunn et al., 2009; Koh et al., 2004). 2) Host-breadth models, which  
59 can handle uncertainty in estimating host-breadth (Vesk et al., 2010), and may be used along with  
60 decision protocols not only to estimate coextinction rates but also identify actions that may  
61 increase the persistence of threatened species (Moir et al., 2012, 2011). 3) Models based on  
62 ecological networks, which are more focused on identifying overall consequences in network  
63 structure after species removal (Bascompte and Stouffer, 2009). These models can handle more  
64 than two trophic levels, and have been recently improved to model more realistic situations (e.g.,  
65 to integrate the variation in the interaction dependence, compensation after species extinctions,  
66 and the formation of new interactions; Baumgartner et al., 2020; Dunne et al., 2002; Vieira and  
67 Almeida-Neto, 2015). While these approaches have significantly advanced our knowledge of  
68 coextinction of interacting species (Bascompte and Stouffer, 2009; Carlson et al., 2017a; Dunn et  
69 al., 2009; Koh et al., 2004; Taylor and Moir, 2014; Thacker et al., 2006; Vieira and Almeida-Neto,  
70 2015), some improvements in assessing coextinction rates are still needed (see Carlson et al., 2019;  
71 Colwell et al., 2012; Moir et al., 2010; Strona, 2015).

72 Important gaps have been identified in the ability to estimate coextinction risk (Colwell et  
73 al., 2012; Moir et al., 2010). For instance, Moir et al., (2010) identified three categories of gaps:

74 accuracy in 1) host data (e.g., threat of the host status), 2) dependent data (host specificity), and  
75 the 3) interactions between these two components (e.g., differences in the interactions across  
76 geographic regions). In addition, Colwell et al. (2012) call for incorporating demographic and  
77 evolutionary dynamics, host switching, affiliate phylogeny, and risk factors for affiliate extinction.  
78 Overall, while some of the current gaps in assessing coextinction risk have already been overcome  
79 (e.g., better estimates of host specificity, Vesk et al., 2010; or allowing new interactions to emerge,  
80 Baumgartner et al., 2020), most of these aspects are yet to be implemented in current  
81 methodologies.

82 One example of a gap in assessing coextinction risk is that correlated extinctions of host  
83 and symbionts are not currently implemented in coextinction estimation (Colwell et al., 2012; Moir  
84 et al., 2010; Rezende et al., 2007). Particular groups of species might be more prone to extinction  
85 due to different factors, for example shared life-history traits due to phylogenetic relatedness.  
86 Indeed, current methodologies to assess coextinction risk generally do not consider symbiont  
87 evolutionary history (but see Hoyal Cuthill et al., 2016; Moir et al., 2016). This lack of  
88 consideration is especially important for traits that might be related to the likelihood of a symbiont  
89 group becoming extinct (Cizauskas et al., 2017; Colwell et al., 2012; Moir et al., 2014). These  
90 traits could include host-switching capability (Clayton et al., 2015), effective population sizes  
91 (Allendorf, 2017), potential for rapid evolution (Villa et al., 2019), or hybridization dynamics  
92 (Detwiler and Criscione, 2010; Doña et al., 2019c; Vallejo-Marín and Hiscock, 2016).  
93 Accordingly, new approaches able to obtain estimates of symbiont extinction rates, including more  
94 aspects of symbiont evolutionary history, might enhance our ability in assessing symbiont  
95 extinction risk (Colwell et al., 2012; Moir et al., 2010; Strona, 2015). However, to our knowledge,  
96 these approaches are still non-existent.

97 Current cophylogenetic methodologies might offer a possibility to include evolutionary  
98 history in symbiont extinction risk assessments (Section 2). Cophylogenetic methods have been  
99 useful to disentangle which ecological and evolutionary traits drive the evolutionary history of  
100 symbionts (Clayton et al., 2015, and references therein). For instance, from cophylogenetics, we  
101 know that at an evolutionary scale, lower dispersal capabilities are associated with higher  
102 cospeciation rates and lower host-switching rates (Clayton et al., 2015; Doña et al., 2017b; Sweet  
103 and Johnson, 2018). Similarly, cophylogenetic methods could be used to increase our knowledge  
104 of which symbiont traits are behind particular extinction rates (e.g., transmission, aggregation, or  
105 geographic patchiness, see Box 1). Also, among the most remarkable advantages of a  
106 cophylogenetic-based statistic would be that this statistic would not directly derive from host  
107 extinction probabilities, and thus might allow uncovering symbiont groups with high extinction  
108 probabilities inhabiting hosts with low extinction probabilities, in contrast to current approaches  
109 to estimate coextinction.

110 In this overview, we propose a new statistic derived from cophylogenetic analyses  
111 “cophylogenetic extinction rate” (*Ec*), to assess symbiont extinction risks (Section 2). As this  
112 statistic derives largely from coevolutionary theory, we list different ecological and evolutionary  
113 variables that are expected to influence this statistic and might be useful to consider when

114 interpreting *Ec* values (Box 1, Figure 1). Moreover, given the novelty of the approach, we propose  
115 potential future research to further develop estimation of cophylogenetic symbiont extinction risks  
116 and the inclusion of these data into current modeling practices.

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119 **Box 1. Ecological and evolutionary variables influencing cophylogenetic extinction rate (*Ec*)**

120 The variables listed below, along with their predictions regarding extinction risk, are derived from  
121 well-studied topics in coevolutionary, coextinction, and conservation biology. Note that we place  
122 the most emphasis on those variables that are the most unique and relevant to symbionts (e.g., we  
123 do not include a "habitat" variable, apart from the host, because it might be less relevant for  
124 symbionts than for other interacting systems, such as plant-pollinator or predator-prey systems).  
125 Some of these variables have already been reviewed in depth (Colwell et al., 2012; Moir et al.,  
126 2014, 2010; Strona, 2015). Cophylogenetic extinction risk (*Ec*) derives mostly from  
127 coevolutionary biology theory; therefore, the value of this parameter is expected to be congruent  
128 with predictions from this theory.

129 *1. Host specificity:* In the most basic sense, host specificity can be defined as the number of  
130 recorded hosts for a given symbiont species (Lymbery, 1989). Also, it can include the  
131 phylogenetic relationships among hosts (phylogenetic host specificity) and the variation across the  
132 geographic range (geographic host-specificity) (Poulin et al., 2011; Wells and Clark, 2019). This  
133 variable has been widely integrated in coextinction models (Colwell et al., 2012; Moir et al., 2010;  
134 Strona, 2015). In short, highly host-specific symbionts are expected to have a higher extinction  
135 risk. However, Strona et al., (2013) found that host-stability instead of host-specificity was the  
136 main determinant of the risk of becoming extinct.

137 *2. Mode of transmission:* The mode of transmission of a symbiont species (e.g., vertically: from  
138 parents to offspring vs. horizontally: between individual hosts that are not parents and offspring)  
139 is associated with major ecological and evolutionary aspects that may be relevant to consider in  
140 symbiont extinction risk assessments (Lipsitch et al., 1995; Whiteman et al., 2004; Huyse et al.,  
141 2005; Barrett et al., 2008; Clayton et al., 2015; Doña et al., 2017b; Sweet and Johnson, 2018;  
142 Poulin, 2011; Antonovics et al., 2017; Fisher et al., 2017; Doña et al., 2019c). All else being equal,  
143 vertically-transmitted symbionts would be predicted to have a higher probability of extinction  
144 given their level of specialization and other relevant features that increase extinction risk (e.g.,  
145 high virulence or low levels of genetic diversity).

146 *3. Virulence:* Virulence can be defined as the reduction of host fitness caused by the symbiont  
147 (Cressler et al., 2016; Herre, 1993; Read, 1994). More virulent parasites might increase the risk  
148 of extinction of the host, because of the morbidity of host individuals that harbor virulent parasites.  
149 This process will lead to a greater chance of coextinction of the host and parasite, because there  
150 may be no time for natural selection to adjust virulence.

151 *4. Straggling and host-switching:* Straggling (i.e., symbiont dispersal to a novel host) seems to be

152 frequent at an ecological scale, whereas successful host-switches (i.e., not only reaching a new  
153 host but also reproducing on the new host) are comparatively rare (Whiteman et al., 2004; Rivera-  
154 Parra et al., 2017; Doña et al., 2019b, 2018). Higher straggling and host-switching capabilities  
155 might be associated with a higher likelihood of escaping from host extinctions (Agosta et al., 2010;  
156 Agosta and Klemens, 2008; Clayton et al., 2015; Engelstädter and Fortuna, 2019). However, the  
157 low rate of successful host colonization suggests that it is probably unrealistic to think that host-  
158 switching may save symbiont species from becoming extinct (Settele et al., 2014; Carlson et al.,  
159 2017a; Cizauskas et al., 2017).

160 *5. Symbiont population genetic structure:* Intermediate degrees of population subdivision  
161 generally yield the highest adaptive potential (Allendorf et al., 2007; Futuyma, 2013). Most  
162 symbiont species have levels of gene flow between populations that are often higher than between  
163 host populations, and therefore extreme levels of population subdivision are not expected to be the  
164 norm (Clayton et al., 2015; Doña et al., 2019a; Huysse et al., 2005; Mazé-Guilmo et al., 2016;  
165 McCoy et al., 2003; Poulin, 2011). However, symbiont populations are theoretically expected to  
166 become more fragmented due to anthropogenic causes (Pickles et al., 2013; Carlson et al., 2017a),  
167 leading to situations in which worrisome levels of subdivision can become more frequent.

168 *6. Aggregation:* Symbionts are generally aggregated among the available hosts so that most host  
169 individuals are inhabited by few to no symbionts, while many symbionts inhabit just a few host  
170 individuals (Poulin, 2011; Rózsa et al., 2000). Aggregation can be an important parameter for  
171 symbiont persistence (Clayton et al., 2015; Cornell et al., 2003; Criscione and Blouin, 2005;  
172 Dhamarajan, 2015; Montarry Josselin et al., 2019), affecting levels of genetic diversity and  
173 increasing the probability of stochastic extinction.

174 *7. Host population size:* Symbionts from hosts with small population sizes (and low-density) are  
175 expected to be more vulnerable because they typically have lower abundances (Arneberg Per et  
176 al., 1998; Ellis et al., 2017), higher coextinction risks (Strona, 2015), and even higher probabilities  
177 of extinction than that of their hosts because of the aggregated distributions of symbionts.

178 *8. Symbiont effective population size:* Symbionts often have life-history features that may reduce  
179 the effective population size ( $N_e$ ) (Criscione and Blouin, 2005; Dabert et al., 2015; Dobson, 1986;  
180 Doña et al., 2015; Monsion et al., 2008). Obtaining precise estimates of symbionts  $N_e$  is a  
181 complicated task (Crellen et al., 2016; Thiele et al., 2018; Criscione, 2013; Strobel et al., 2019).  
182 Large and small  $N_e$  may be expected in symbionts depending on conditions (Seger et al., 2010;  
183 Criscione, 2013; Strobel et al., 2019; Hughes and Verra, 2001), and factors such as aggregation,  
184 bottlenecks, prevalence, and intensity influence  $N_e$  (Criscione and Blouin, 2005; Dabert et al.,  
185 2015; Dobson, 1986; Doña et al., 2015; Monsion et al., 2008). Knowledge of these parameters  
186 may provide useful insights regarding symbiont  $N_e$ .

187 *9. Geographic patchiness:* The geographic ranges of symbionts and hosts do not always match  
188 perfectly, with some symbionts almost mirroring the whole distribution of their hosts and others  
189 restricted to some small areas of host distribution (Krasnov et al., 2004; Bush et al., 2009; Poulin,  
190 2011; Clayton et al., 2015; Wells and Clark, 2019; Bush and Kennedy, 1994; Bush et al., 2013).



191 Symbionts restricted to reduced areas of host distribution may be expected to have a higher  
192 vulnerability of becoming extinct.

193 *10. Host effects:* Not all hosts are equally suitable for the symbionts. Apart from host extinction  
194 risk (which may lead to coextinction processes), several host features also have a substantial effect  
195 on symbiont traits when there is an intimate association (Clayton et al., 2015). This effect might  
196 derive from coadaptative (e.g., arm-races dynamics) or non-coadaptive processes (e.g., host-  
197 density influencing parasite abundance) (Clayton and al, 2010; Clayton et al., 2015; Bush and  
198 Clayton, 2018; Villa et al., 2016, 2018; Bush et al., 2019; Arneberg Per et al., 1998; Ellis et al.,  
199 2017; Clausen, 1939; Hall et al., 2014). Furthermore, hosts may possess traits that hamper  
200 symbiont colonization (Poulin et al., 2012). Taken together these host features may be related to  
201 the likelihood of a symbiont becoming extinct.

202 *11. Symbiont body size and life-cycle:* Large bodied symbionts tend to have smaller effective  
203 population sizes and to depend on larger hosts, which are also more vulnerable to extinction (Bush  
204 and Clayton, 2006; Clayton et al., 2015; Ripple et al., 2017). Also, symbionts with complex life-  
205 cycles and limited climatic tolerances (e.g., ectotherms, Cizauskas et al., 2017) might be at an  
206 increased risk of extinction (Colwell et al., 2012; Koh et al., 2004; Lafferty, 2012; Poulin and  
207 Morand, 2004).

208 *12. Trait matching:* Due to the coevolutionary process, symbionts tend to possess traits that match  
209 very tightly those of their hosts (Clayton et al., 2015). All else being equal, highly host-specific,  
210 and specialized symbionts might be predicted to be more endangered because they depend more  
211 upon their hosts, and their odds of successfully colonizing a new host species in ecological time  
212 are lower than that of multi-host and often more generalist species (Agosta et al., 2010; Remold,  
213 2012; Thompson, 1994). However, factors such as overestimates of the level of host specificity  
214 (Braga et al., 2018; Dallas et al., 2017; Doña et al., 2019b), ability for rapid evolution (Bush et al.,  
215 2019; Koch et al., 2014; Villa et al., 2019), and host stability (Strona et al., 2013) may counteract  
216 this prediction.

217 *13. Inter- and intraspecific competition:* Symbionts not only interact with their hosts, but they also  
218 interact with diverse communities (including other symbionts) with whom they share their host  
219 (Bush and Malenke, 2008; Clayton et al., 2015; Doña et al., 2017a; Harbison et al., 2008; Johnson  
220 et al., 2009; Perez and Atyeo, 1984; Proctor, 2003). Overall, competitively superior species are  
221 expected to be less vulnerable to extinction than competitively inferior species (Clayton et al.,  
222 2015). However, due to asymmetries in resource distribution and competition, competitively  
223 superior species, counterintuitively, may not necessarily be the winners in a rapidly changing  
224 climate scenario (Northfield and Ives, 2013; Van Den Elzen et al., 2017).

## 225 **2. Obtaining symbiont extinction rates from cophylogenies**

226 In cophylogenetic analyses, host and symbiont evolutionary trees are compared to uncover the  
227 processes driving symbiont diversification (Page, 2003; De Vienne et al., 2013; Clayton et al.,  
228 2015; Martínez-Aquino, 2016). Several cophylogenetic methods exist, and these can be classified

229 into two main categories: distance-based and event-based methods (Page, 2003; De Vienne et al.,  
230 2013; Martínez-Aquino, 2016). In short, distance-based methods (e.g., ParaFit; Legendre et al.,  
231 2002) measure the topological distance between host and symbiont trees and statistically evaluate  
232 whether the congruence is higher than expected by chance (Huelsenbeck et al., 2003; De Vienne  
233 et al., 2013; Martínez-Aquino, 2016). In these methods, significantly high levels of congruence  
234 are generally assumed to be the result of codivergence between host and symbionts (Huelsenbeck  
235 et al., 2003; De Vienne et al., 2013; Martínez-Aquino, 2016). Event-based methods (e.g., Jane;  
236 Conow et al., 2010) use costs for macroevolutionary events (i.e., events such as cospeciation, host-  
237 switches, and losses that occur at a macroevolutionary scale, see below) which must be previously  
238 specified by the user, to reconcile host and symbiont phylogenetic trees (De Vienne et al., 2013;  
239 Charleston and Libeskind-Hadas, 2014; Martínez-Aquino, 2016). The result of an event-based  
240 cophylogenetic analysis generally includes the optimal solution to reconcile both trees (i.e., given  
241 the costs specified) and the corresponding number of macroevolutionary evolutionary events of  
242 each category that were needed to reach that solution. These events typically include cospeciation,  
243 duplication, host-switching, loss, and failure to diverge (De Vienne et al., 2013; Charleston and  
244 Libeskind-Hadas, 2014; Martínez-Aquino, 2016).

245 Here, we propose a way to obtain a rough estimate of the relative extinction rate of a  
246 particular symbiont lineage using the proportional number of losses (i.e. “sorting events”) from an  
247 event-based cophylogenetic reconstruction (Fig. 2). This approach is similar to the current practice  
248 of using the percentage of cospeciation events as a measure of the relative importance of  
249 cospeciation in a symbiont lineage (Johnson and Clayton, 2003; De Vienne et al., 2013; Doña et  
250 al., 2017b). In cophylogenetic reconstructions, the inferred losses can be interpreted as the  
251 consequence of two distinct processes (Fig. 2; Clayton et al., 2015): (1) as genuine events of  
252 parasite extinctions; or (2) as sorting events (e.g., ‘missing the boat’), when a symbiont fails to  
253 disperse along with one host lineage. Note that even though sorting events are not directly  
254 indicative of symbiont species extinctions, they do contain information regarding symbiont  
255 transmission efficiency and reflect the probability of stochastic extinction, and therefore might be  
256 valuable for assessing symbiont extinction at a species scale (Paterson et al., 1999; MacLeod et  
257 al., 2010; Poulin, 2011; Clayton et al., 2015).

258 The estimation of  $Ec$  (i.e., cophylogenetic extinction rate) would be as follows:  $Ec =$   
259  $\left(\frac{L}{E+2S}\right)$ ; where  $L$  represents the number of losses,  $E$  the total number of macroevolutionary events  
260 other than host-switches (i.e., cospeciation + duplication + losses), and  $S$  the number of host-  
261 switches. The number of host-switches ( $S$ ) is included twice because, in contrast to the other  
262 macroevolutionary events, it should effectively lower the extinction risk (Carlson et al., 2017a;  
263 Cizauskas et al., 2017; Clayton et al., 2015; Colwell et al., 2012; Dunn et al., 2009; Koh et al.,  
264 2004; Moir et al., 2010, 2014). Consequently,  $Ec$  can discriminate between two symbiont groups  
265 with the same ratio of losses ( $L$ ) vs. all the other events ( $E+S$ ) but differing in the number of host-  
266 switches; so that the lowest  $Ec$  will be that of the group with a higher number of host-switches.  
267 The parameter  $Ec$  increases linearly with the number of losses, and decreases as host-switching

268 increases (Fig 3). We encourage accompanying  $Ec$  with a confidence interval to show the level of  
269 precision in the estimate (e.g., the modified Wilson confidence interval for a binomial proportion;  
270 (Brown et al., 2001; Signorell et al., 2019). To aid in calculating  $Ec$  and confidence intervals  
271 (modified Wilson), we provide a Shiny app (<https://jdona.shinyapps.io/extinction/>).

272 As a proof of concept of this approach, we calculated  $Ec$  for two symbiont lineages; the  
273 feather mite genera *Proctophyllodes* and *Trouessartia* (Acari: Astigmata: Analgoidea and  
274 Pterolichoidea). As input for the calculations, we used the results of event-based cophylogenetic  
275 reconstructions from Doña et al., (2017b). In this study, *Trouessartia* mites were found to have 1  
276 loss and 9 host-switches out of 14 events, and *Proctophyllodes* mites 1 loss and 32 host-switches  
277 out of 42 events. From these values, the estimated  $Ec$  is slightly higher for *Trouessartia* ( $Ec =$   
278  $0.04$ ;  $CI = 0-0.21$ ) than for *Proctophyllodes* ( $Ec = 0.01$ ;  $CI = 0-0.07$ ) mites. This result agrees with  
279 existing comparative knowledge from the ecology and evolution of these mites. Specifically,  
280 *Trouessartia* mites are known to have: 1) a lower species diversity on Passeriformes (Doña et al.,  
281 2016, 2018), 2) lower prevalence (i.e., the proportion of individuals inhabited by a symbiont  
282 species within a host sample; Reiczigel et al., 2019) and intensity (i.e., the number of individual  
283 symbionts inhabiting a particular host; Reiczigel et al., 2019) (Fernández-González et al., 2018;  
284 Doña et al., 2019b), 3) lower genetic diversity (Fernández-González et al., 2018; Doña et al.,  
285 2019b), and 4) infrapopulations genetically more structured (i.e., with lower gene flow among  
286 infrapopulations—all the individual symbionts inhabiting an individual host—, than  
287 *Proctophyllodes* species (Doña et al., 2019a).

### 288 3. Future opportunities

289 Global symbiont diversity is at a high risk of extinction (Carlson et al., 2017a; Dunn et al., 2009;  
290 Koh et al., 2004). The situation is highly concerning because symbionts have not received  
291 appropriate scientific and public attention, especially when compared to free-living species (see  
292 this special issue; Carlson et al., 2020). Here, we have proposed the first cophylogenetic-based  
293 statistic ( $Ec$ ) for assessing symbiont extinction risk. Previous studies on coextinctions have  
294 repeatedly called for including evolutionary information in symbiont extinction risk assessments,  
295 and  $Ec$  allows one to do so. Also, in contrast to most methods to estimate coextinction,  $Ec$  does  
296 not rely upon host extinction probabilities, and therefore, it allows uncovering symbionts with a  
297 high extinction risk associated with hosts with a low extinction risk. Overall, we believe that  $Ec$   
298 represent a highly valuable addition to current methods in assessing symbiont extinction rates.  
299 Moreover, the calculation of this parameter can act as a first step, thus stimulating further advance  
300 in symbiont conservation biology and cophylogenetics. In this vein, we anticipate the following  
301 three areas as valuable for future research efforts to advance symbiont conservation theory towards  
302 better extinction risk assessments:

303 3. 1. *Meta-analyses of current cophylogenetic data:* To date, several cophylogenetic studies on  
304 symbionts have been carried out (De Vienne et al., 2013; Clayton et al., 2015). These studies  
305 usually report the number of reconstructed macroevolutionary events of each type. Accordingly,



306 *Ec*, may be calculated from already published studies and compared between studies in a meta-  
307 analytic framework. These studies might allow obtaining new information on the drivers of  
308 symbiont extinction (Cizauskas et al., 2017; Moir et al., 2014), similarly as how it has been done  
309 to uncovers the drivers of cospeciation and host-switching (Clayton et al., 2015 and references  
310 therein). Moreover, by doing so, lineage-specific *Ec* will be generated and available for symbiont  
311 initiatives that list their conservation status (e.g., PEARL; (Carlson et al., 2017b), and future  
312 conservation-focused studies.

313 *3. 2. Integrating symbiont conservation biology into coevolutionary and cophylogenetic research*  
314 *agenda:* Future studies are encouraged to improve how extinction rates are estimated from  
315 cophylogenetic comparisons. For example, some types of host-switches often imply that an  
316 extinction event has happened on the old host (i.e., host-switching with extinction or host-  
317 switching with speciation and extinction). On the other hand, host-switching can save a symbiont  
318 from extinction (i.e., if the host goes extinct after the switch). To our knowledge, current event-  
319 based cophylogenetic methods do not allow computing these types of host-switches. Future  
320 research is needed here. Lastly, given that the extinction rate is likely not constant throughout the  
321 evolution of a lineage, providing information on the variation of the extinction rate through time  
322 would be useful. However, to our knowledge, no event-based cophylogenetic method allows for  
323 use as an input fully-dated phylogenies nor produces as a result dated macroevolutionary events.  
324 Future improvements in cophylogenetic methods may allow providing extinction rate estimations  
325 through time.

326 *3. 3. Expanding modeling of symbiont extinction (and coextinction) rates:* As stated above,  
327 cophylogenies might offer an opportunity to include data from evolutionary history in coextinction  
328 assessments. One way to do so would be, for example, to incorporate cophylogenetic extinction  
329 rate data as a proxy of the evolutionary propensity of extinction of a given group of symbionts into  
330 current coextinction models. Indeed, species distribution and evolutionary models that are often  
331 used to calculate species extinction risk of free-living species, already accommodate different  
332 variables (e.g., dispersal, demography, genetic data) to improve their predictions (Carlson et al.,  
333 2019). On the other hand, because cophylogenetic methods still do not allow using time-calibrated  
334 phylogenies (see section 3.2), another related avenue for improvement would be to also incorporate  
335 lineage-specific extinction rates from time-calibrated phylogenies (Beaulieu and O'Meara, 2015;  
336 Rabosky, 2016). These methods are based on the notion that speciation and extinction processes  
337 leave distinct signatures on the branching structure of a phylogeny (Nee et al., 1994). While these  
338 approaches have been extensively used in free-living species, to our knowledge, extinction rates  
339 from phylogenies have almost not been yet used in symbiont studies (but see Alcalá et al., 2017).  
340 Thus, even though these estimates on extinction risk from phylogenies should be treated with  
341 caution (particularly when including non-sequenced species; Rabosky, 2016), the increasing  
342 availability of robust-comprehensive phylogenetic trees offers an opportunity to use these  
343 phylogenetically based methods in symbionts (Johnson, 2019).

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347 particular, to this contribution.

#### 348 **Appendix A. Supplementary data**

349 A shiny application to directly calculate *Ec* and confidence intervals from the number of  
350 macroevolutionary events estimated from an event-based cophylogenetic reconstruction can be  
351 found here (<https://jdona.shinyapps.io/extinction/>). The R function is also available at GitHub  
352 ([https://github.com/Jorge-Dona/cophylogenetic\\_extinction\\_rate](https://github.com/Jorge-Dona/cophylogenetic_extinction_rate)).

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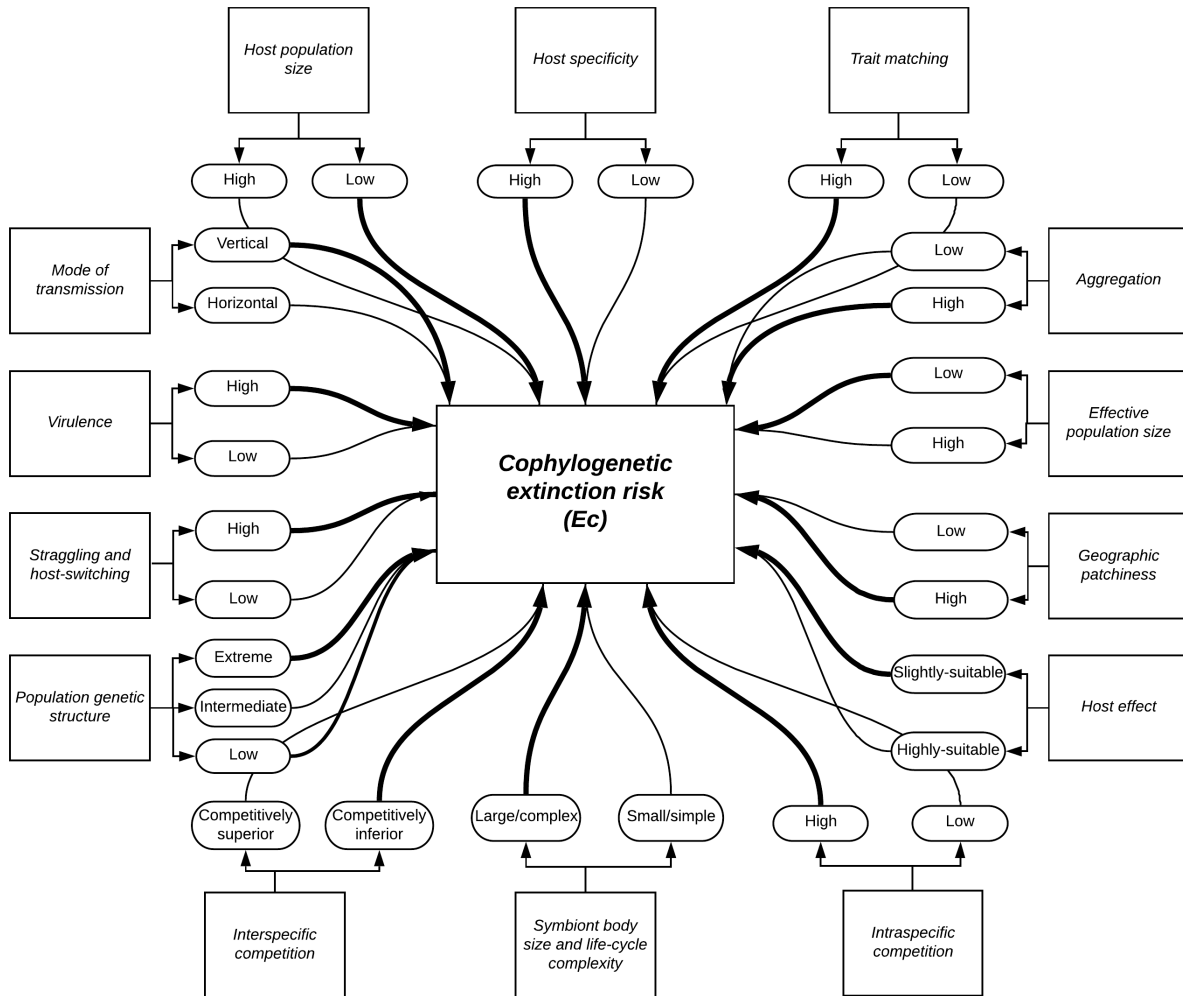
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727 **Fig. 1.** Diagram depicting predictions of variables influencing *Ec* (see Box 1). Note that this is a  
 728 highly simplified summary, with greater detail provided in Box 1 and references therein. Also, to  
 729 ease rapid access to these predictions, we have set line thickness to represent different extinction  
 730 risk levels (thicker = higher; thinner = lower) according to our interpretation of current knowledge  
 731 on these topics.

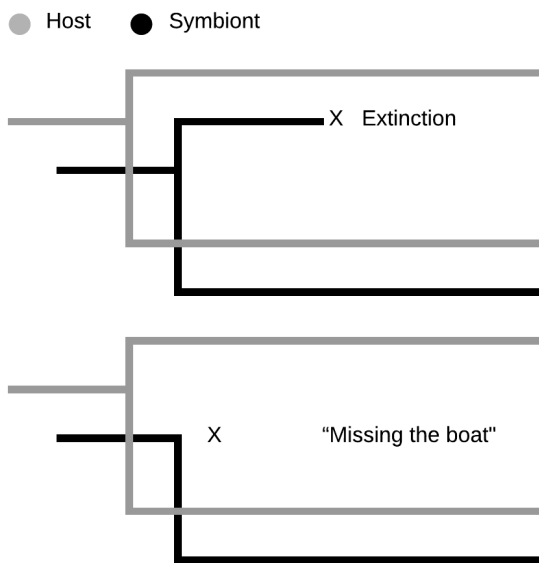


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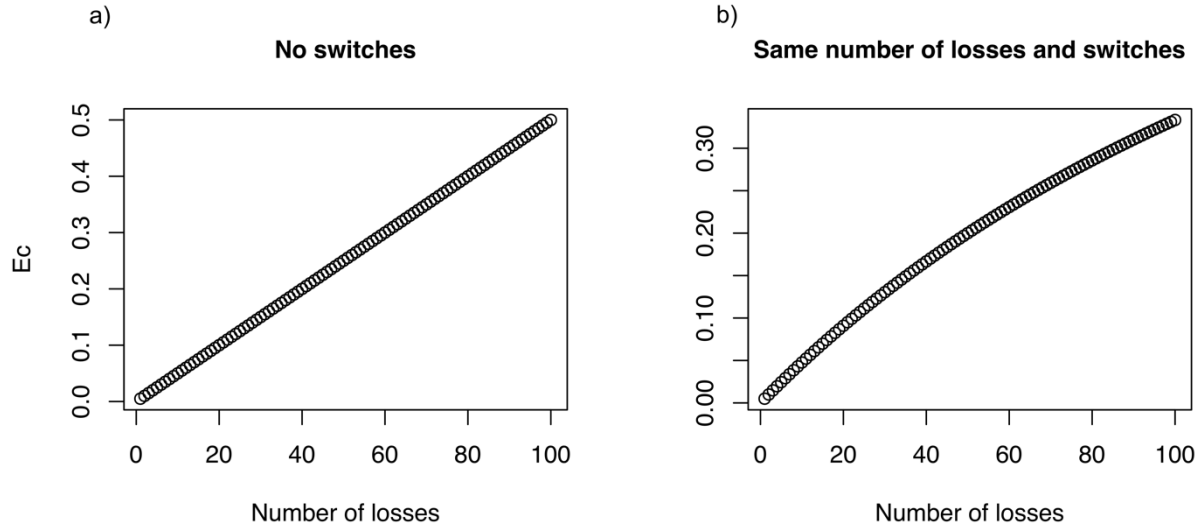
734 **Fig. 2.** Diagram depicting symbiont losses in an event-based cophylogenetic reconstruction.  
735 Inferred losses from a cophylogenetic reconstructions can be interpreted as the consequence of  
736 two distinct processes (Clayton et al., 2015): (1) as genuine events of parasite extinctions (top  
737 figure); or (2) as sorting events (e.g., ‘missing the boat’, bottom figure), when a symbiont fails to  
738 disperse with one host lineage. Note that even though sorting events are not directly indicative of  
739 symbiont species extinctions, they do provide information about symbiont transmission efficiency  
740 and reflect the probability of stochastic extinction, and therefore might be valuable for assessing  
741 symbiont extinction at a species scale (Paterson et al., 1999; MacLeod et al., 2010; Poulin, 2011;  
742 Clayton et al., 2015).

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757 **Fig. 3.** Results of simulations showing the behavior of  $E_c$  under increasing numbers of losses and  
758 host-switches. The number of total events ( $E$ ) is 200 in both plots. Note that in a) there were no  
759 switches in any iteration while in b), the same number of losses and switches is held across all the  
760 iterations (e.g., one loss and one switch or two losses and two switches)



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