

1 Cultural conformity and persistence in the context of differing site fidelity

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39 Abstract

40

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42 Animal culture often shows geographic structure, with individuals in close proximity sharing more
43 cultural features than individuals further apart. However, spatial extent of cultural features, along with
44 the degree of conformity to local cultures, vary within and among species. Further, rates of cultural
45 change presumably also vary, though documentation of temporal variability lags behind documentation
46 of spatial variability. Understanding both spatial and temporal variation is essential to understanding
47 cultural evolution, but mechanisms likely to be driving this variation have not been sufficiently explored.
48 We hypothesized that conformity to local culture in Oscine songbirds and the persistence of culture over
49 time and space are promoted by habitats that facilitate stable populations in which individuals show
50 relatively high site fidelity. In contrast, sites in which habitat features cause rapid population turnover
51 provide more vacant territories and so more opportunities for colonization. Colonization should drive
52 more rapid cultural change, either through adult colonists importing foreign cultural variants or young
53 colonists making errors as they learn the local song. To test this set of hypotheses, we examined
54 temporal and spatial variation in vocal culture in a songbird (dickcissel, *Spiza americana*) in two distinct
55 habitat types. As predicted, we found high site fidelity in relatively stable native grasslands and much
56 lower site fidelity in nearby cropland sites which were disturbed by farming practices during the
57 breeding season. We also found evidence of higher levels of song sharing and slower changes in vocal
58 culture in our grasslands relative to croplands, though we found no evidence of different spatial scales
59 of song sharing between these habitats. This is the first study we know of correlating the temporal rate
60 of cultural change to differences in a demographic factor between habitats. Although our conclusions
61 are based on many recorded songs, automated assessments of song similarity were imprecise and so
62 our results here underestimate the overall degree of song sharing and thus possibly the strength of the
63 effects of time and distance on this sharing. Further, because we examined song sharing at only seven
64 sites, firm conclusions about site fidelity and song sharing will require larger samples in the future.

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67 keywords:

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69 apparent survival, culture, dialect, dickcissel, site fidelity, song sharing

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72 Introduction

73

74 Culture consists of behaviors learned from conspecifics, and it is widespread in birds, mammals, and
75 some other taxa (Laland and Janik 2006). Because social interactions determine cultural transmission,
76 understanding the factors that influence the movements and interactions of individuals within
77 populations is essential to understanding temporal changes and geographic distributions of animal
78 culture (Fayet et al. 2014, Aplin 2018).

79

80 Some of the best-studied examples of animal culture include sharing of songs in birds. In many
81 songbirds, individuals in one location produce songs more similar to other males in the neighborhood
82 than to the songs of more distant conspecifics (Lemon 1975, Krebs and Kroodsma 1980, Baker and
83 Cunningham 1985). Conformity to the local culture is often striking (e.g., Baker and Thompson 1985,
84 Schook et al. 2008).

85

86 Patterns of vocal culture or song sharing in birds vary within and among species. In some species, song
87 types are shared by only a handful of neighbors (e.g., Payne et al. 1988), while in others, song sharing
88 extends across continents (e.g., Kroodsma et al. 1999). Degree of song sharing varies also, with
89 neighbors having almost indistinguishable vocalizations in some species (e.g., Schook et al. 2008) and
90 sharing only a portion of their song types in others (e.g., Beecher and Campbell 2005). In some species,
91 vocal traditions have been observed to evolve slowly from year to year (e.g., McGregor and Thompson
92 1988) or over decades (e.g., Derryberry 2009, Ju et al. 2019), while in others dramatic changes are
93 readily observed from one year to the next (Payne 1985). Some of this cross-species variation can be
94 explained, for instance by the tendency to improvise (Beecher and Brenowitz 2005) and the degree to
95 which species are sedentary versus migratory (Handley and Nelson 2005). However, there also appears
96 to be substantial variation in song sharing patterns within species. Evidence of such geographic
97 variability has not been widely reported, but it may be common. For instance in the house finch
98 (*Haemorhous mexicanus*), the average level of sharing by neighbors and the degree to which particular
99 song elements are shared by all local birds differs among locations (Mundinger 1975, Bitterbaum and
100 Baptista 1979, Tracy and Baker 1999) and within sites over time (Ju et al. 2019). In species such as the
101 red-winged blackbird (*Agelaius phoeniceus*; Kroodsma and James 1994), chaffinch (*Fringilla coelebs*;
102 Lachlan and Slater 2003), and dickcissel (*Spiza americana*; Schook et al. 2008), both the levels of local
103 song sharing and the spatial extent of song sharing vary geographically. Explaining this variation in song
104 sharing within species should be an important step in building our understanding of the emergence and
105 persistence of the diversity of vocal culture.

106

107 Some variability in vocal culture within species results from isolation of populations, for instance on
108 oceanic islands or on isolated 'islands' of suitable habitat, leading to allopatric emergence of distinct
109 vocal traditions (e.g., Baptista 1975, Baptista and Johnson 1982, Slabbekoorn et al. 2003, Gammon et al.
110 2005, Baker et al. 2006, Keighley et al. 2017). Most empirical studies that link population characteristics
111 and avian vocal traditions have focused on this phenomenon. However, geographic patterns of vocal
112 culture in many species show spatial turnover within a continuously inhabited area across distances well
113 within the range of individual dispersal (e.g., Baptista 1975, McGregor 1980, Lachlan and Slater 2003,
114 Shieh 2004, Schook et al. 2008). In other words, isolation is important where it occurs, but a great deal
115 of variation in vocal culture is not attributable to current patterns of isolation.

116

117 The simplest mechanism creating cultural variability over space in continuous populations is isolation by
118 distance (Laiolo and Tella 2006), to borrow terminology from genetics (Manel et al. 2003). As distances
119 between individuals grow, cultural similarity between individuals declines due to decreasing social

120 connectivity. Patterns consistent with isolation by distance have been qualitatively (e.g., Baptista 1975,
121 Munding 1975) or quantitatively (e.g., Bitterbaum and Baptista 1979, Laiolo and Tella 2005, 2006,
122 Schook et al. 2008) described in multiple species. If isolation by distance were sufficient to explain all
123 geographic patterns of vocal culture, we would expect to find relatively constant, species-specific levels
124 of both song sharing within localities and declines in song sharing with distance. Instead, in the few
125 cases where these have been examined, local levels of song sharing vary (Munding 1975, Bitterbaum
126 and Baptista 1979, Tracy and Baker 1999, Laiolo and Tella 2005, Schook et al. 2008, Ju et al. 2019), and
127 levels of song sharing are not a consistent function of distance between individuals (Laiolo and Tella
128 2005, 2006, Schook et al. 2008).

129
130 This variability in spatial or temporal patterns of song sharing might be influenced by variation in
131 population processes, such as the movement of individuals, which might differ across landscapes. For
132 instance, turnover in territory occupancy influences the extent of song sharing in simulation models
133 (Goodfellow and Slater 1986). Territory vacancies, created by emigration or by death, could influence
134 song sharing in multiple ways (Baker and Thompson 1985, Kroodsmas et al. 1985, Goodfellow and Slater
135 1986, Holland et al. 1996). In the most extreme cases where all local individuals either leave or die and a
136 site is colonized by a new set of individuals, we would expect dramatic and immediate changes in song
137 type and geographic patterns of song sharing (Holland et al. 1996). This extreme scenario may be rare,
138 but variation in rates of population turnover within a species is common, presumably because of
139 variation in features of the habitat (Winker et al. 1995, Newton 1998), and might have major
140 implications for vocal traditions. For instance, temporal change in song types within a neighborhood
141 could be rapid if site fidelity is low or mortality high since drift (elimination of song types due to chance)
142 will be more likely with smaller numbers of individuals persisting from one year to the next (Goodfellow
143 and Slater 1986, Nunn et al. 2009). Further, the larger the proportion of territories occupied by young
144 birds in a given year, the higher the proportion of individuals learning the local song that year. Making
145 the typical assumption of a constant probability of copying errors per bird learning new song (Williams
146 and Slater 1990, Lachlan and Slater 2003), the more individuals that are learning song each year, the
147 greater the number of errors that accumulate in the population (Goodfellow and Slater 1986). These
148 deviations from tutors' songs, whether due to improvisation or learning error (Beecher and Brenowitz
149 2005), are presumably the ultimate source of new song types (Slater 1986) and so more novel song
150 variants should occur in areas with low site fidelity, and this could lower the local level of song sharing.
151 Territories made available by death or emigration might also be filled by immigrant adults, and these
152 immigrants could introduce foreign song variants to the population, thus further lowering local sharing
153 (Goodfellow and Slater 1986, Williams and Slater 1990). It is clear that if we wish to understand
154 variation in song sharing patterns in continuously occupied landscapes that we should study variation in
155 territory turnover as a function of habitat. However, we know of only one study to do so to date (Ortega
156 et al. 2014).

157
158 An excellent species for studying the role of population processes in influencing patterns of song sharing
159 is the dickcissel, a migrant songbird breeding in grasslands of the Eastern Great Plains of North America
160 (Temple 2002) and wintering primarily in Northern South America. The vast majority of dickcissel males
161 produce a single song (Figure 1), at least while occupying the same breeding territory within and
162 between seasons (Schook et al. 2008; T. Parker unpublished data). This song typically consists of two or
163 three distinct types of phrases. These phrases include the introductory "dick" and often a "cissel", each
164 repeated multiple times, as well as sometimes a trill of rapidly repeated notes (Figure 1). Dickcissels
165 show striking geographic structure in their song culture, with most males in close proximity to each
166 other singing almost identical song phrases, and the degree of similarity of phrase types often declining
167 considerably as the distance between recording sites approaches 2 km (Schook et al. 2008). Individual

168 dickcissels' songs also differ from each other in the number of times phrases are repeated and the
169 cadence of these repetitions (Figure 1). However, unlike the content of the phrases, some of the
170 differences in the repetition of phrases are not consistent within and across individuals, and so they
171 have not been the focus of our research (Schook et al. 2008, Parra et al. 2017). Circumstantial evidence
172 suggests that male dickcissels learn their song from immediate adult neighbors after dispersal (Krebs
173 and Kroodsma 1980). First, song sharing by dickcissels declines with increasing distance on the scale of
174 100's of meters to several km (Schook et al. 2008), and so pre-migration song learning during the first
175 summer of life would require that, after migration, dickcissels return to a small area within a few
176 hundred meters of where they learned their song(s). However, when nestling dickcissels are banded,
177 they are almost never re-sighted near the location of banding in subsequent years (Temple 2002; B.
178 Sousa unpublished data), indicating that they recruit locally only rarely. Further, because these study
179 sites where nestlings have been banded are often much larger than the extent of a local song
180 neighborhood, we cannot explain this lack of recaptures by assuming that the young birds have
181 dispersed within the song neighborhood but outside the bounds of the study site. If they disperse
182 outside of the study area, they are dispersing into other cultural areas. It also appears unlikely that
183 recently fledged dickcissels disperse into other song neighborhoods beyond the scale of the typical
184 study area to learn a variety of songs prior to migration. This is because fledgling dickcissels typically
185 remain within <1 km of their nest in the month after fledging (Verheijen 2017). Taken together, this
186 evidence suggests that young adult male dickcissels are able to match the songs of their adjacent
187 neighbors only by learning their song when they settle on their first breeding territory after their first
188 winter in South America. An alternative scenario would require males to learn songs during migration or
189 in winter and then to follow a tutor so as to settle near him on the breeding ground, but this seems
190 unlikely, in part because dickcissels migrate at night in flocks of hundreds or thousands of individuals
191 (Temple 2002).

192
193 We had a strong a priori expectation that rates of territorial turnover in dickcissels, driven by site
194 fidelity, should differ between sites covered primarily by crops and sites covered primarily by native
195 grasslands. We expected the cropland sites to have low site fidelity for two reasons. First, the crops in
196 which we frequently observed dickcissels, winter wheat and alfalfa, are typically harvested during the
197 dickcissel nesting season (T. Parker, personal observation). These harvests dramatically reduce
198 vegetation height and also presumably destroy most dickcissel nests in those crop fields. Further, the
199 uncultivated grassy edges of these crops are themselves likely to have elevated rates of nest predation
200 relative to larger patches of intact grassland (Warner 1994). Both these features of crop landscapes, high
201 rates of reproductive failure and the loss of dense vegetation cover, appear to induce territorial
202 abandonment in dickcissels (Schartz and Zimmerman 1971, Zimmerman 1982, Zimmerman and Finck
203 1989, Frawley and Best 1991, Igl 1991). In contrast, large intact grasslands that are not cut for hay
204 should have higher rates of nest success and will lack the regular catastrophic loss of vegetation cover
205 that characterizes crop landscapes, and so we expect dickcissels to abandon their territories in grassland
206 landscapes at lower rates than in crop landscapes.

207
208 In this paper¹, we present tests of a suite of hypotheses associated with spatial and temporal change in
209 the vocal culture of dickcissels. In summary, we hypothesized that accumulation of imperfect song
210 copies and the importation of song variants from elsewhere change local song culture over time and
211 space in dickcissels, and that these changes are more rapid in crop landscapes than in grassland

¹ We describe the evolution of this project and the development of our hypotheses in the
Supplementary Materials, Appendix S1.

212 landscapes because population turnover is faster in cropland due to crop harvest during nesting and
213 nest predation in unharvested field edges. We derived several testable predictions from these
214 hypotheses:
215 - average pairwise song similarity within a site will decline as the number of years between recordings
216 increases
217 - average pairwise song similarity within a site will decline as the geographic distance between
218 recordings increases
219 - dickcissel males in croplands will show lower apparent survival (which is a combined function of site
220 fidelity and true survival), than those in grassland
221 - overall levels of song sharing will be higher in the habitat with the higher apparent survival of male
222 dickcissels
223 - habitat with lower apparent survival of male dickcissels will show steeper declines in average pairwise
224 song sharing as a function of the number of years between recordings relative to the habitat with
225 higher apparent survival
226 - habitat with lower apparent survival of male dickcissels will show steeper declines in average pairwise
227 song sharing as a function of geographic distance between recordings relative to the habitat with
228 higher apparent survival
229
230

231 **Methods**

232
233 We recorded songs and location data from both marked and unmarked dickcissels at seven sites in
234 northeastern Kansas between 2006 and 2014 (Tables 1 – 3). We did not visit all sites in all years. In
235 native perennial grassland we had four non-adjacent sites on the Konza Prairie Biological Station (KPBS).
236 We also worked at three cropland sites with a mix of crops (primarily winter wheat and alfalfa) and
237 weedy field edges within 40 km of KPBS. In some years we monitored sites throughout the breeding
238 season (from late May or early June into late July), but in some other years, we made only a single site
239 visit during the peak of breeding between late May and mid-June). Years with only one visit could not
240 contribute to our estimate of within-season apparent survival or to within-male comparisons of songs
241 across dates within years. We captured male dickcissels by playing dickcissel song in the vicinity of mist
242 nests placed near territorial song perches. Immediately upon capture we fitted male dickcissels with a
243 uniquely numbered USGS aluminum leg band and three (occasionally fewer) colored leg bands in a
244 unique combination for later identification without recapture. By color banding birds and then observing
245 these banded individuals we were able to determine the identity, presence, and distribution of
246 individual birds during and between years. After capture and at subsequent visits, we recorded the song
247 of banded males (typically onto digital media using Marantz PMD680 recorders and Sennheiser ME66
248 shotgun microphones [T. Parker field teams], though also onto analog tapes using a Sony TCM-5000EV
249 cassette recorder and Sennheiser ME67 shotgun microphone [B. Sousa field teams]). We attempted to
250 record at least one clear song with limited background noise. Because a complete song lasts between
251 one and two seconds and males sing only a single song type (Schook et al. 2008), recordings under good
252 conditions were sometimes as brief as 20 seconds, but usually lasted from one to five minutes. The
253 observers making the recording typically vocally identified each song by the target male on the
254 recording immediately after each song was sung so that it could be distinguished from any other nearby
255 conspecific songs. For all recorded birds, banded or unbanded, we recorded location coordinates with a
256 GPS within 10 m of the location where the focal individual had perched while singing. For banded birds,
257 we averaged multiple GPS locations across dates to generate an average location for a given year. Field
258 observers were not blinded to the hypothesis, but relevant measurements were not made in the field,
259 so the lack of blinding is unlikely to have introduced bias.

260
261 Mark-recapture

262
263 To test whether male dickcissels show higher site fidelity in perennial grassland sites relative to cropland
264 sites, we used data from re-sighting of our banded males within and across seasons (Table 2). We
265 analyzed encounter histories with Cormack-Jolly-Seber models for live encounter data available in MARK
266 (version 8.1)(White and Burnham 1999), and tested for differences between cropland habitat and
267 grassland habitats in both apparent survival (ϕ) and detection probability (p). We expect that apparent
268 survival over two-week periods (the typical duration between our site visits) within breeding seasons is
269 driven primarily by emigration, and so consider within-season apparent survival to represent site
270 fidelity. Apparent survival over the non-breeding season can also be influenced by site fidelity if
271 individuals survive but choose a different breeding location in the subsequent year. However, regardless
272 of the extent to which differences in apparent survival between sites are driven by within-season or
273 between-season emigration, or the extent to which they are driven by differences in emigration as
274 opposed to differences in true survival, reduced apparent survival drives an increase in turnover. With
275 greater turnover, there are more opportunities for cultural evolution with the copying of local song or
276 the introduction of foreign song by immigrants.

277
278 We derived our mark-recapture data from two separate studies, one conducted by the T. Parker team,
279 and the other by the B. Sousa team. Some individual birds were monitored by both teams. In these
280 cases, we included only the data from the T. Parker team so that we could test for differences in
281 estimates derived by the two different teams (see below) without double counting birds, and so that we
282 could maximize the sample size of the T. Parker data set, which included birds from both grassland and
283 cropland sites. Because these two studies differed in capture and re-sighting methods, we tested for
284 differences in their estimates of ϕ and p while simultaneously comparing estimates of ϕ and p between
285 cropland and grassland habitats. The T. Parker team monitored banded male dickcissels at both
286 grassland ($n = 102$ marked individuals, of which 81 were resighted at least once) and cropland ($n = 46$
287 individuals, 34 of which were resighted at least once) sites. In years when this team banded, they
288 captured and marked birds for about two weeks beginning the last week of May, which is typically at
289 least two weeks after male dickcissels began establishing territories. These researchers then attempted
290 to re-sight banded individuals approximately once every two weeks until late July. The B. Sousa team
291 monitored banded male dickcissels only in grasslands ($n = 109$ individuals, excluding individuals also
292 monitored by T. Parker team, of which 81 were resighted at least once), captured birds for banding
293 throughout the season, beginning in early May, and attempted to re-sight banded birds most days until
294 early August. To assess the differences between the two studies, we compared models that divided
295 capture histories among three categories (1. cropland, 2. T. Parker grassland, 3. B. Sousa grassland
296 [3hab]) vs. models that lumped data from the two studies and thus allocated all capture histories to
297 either 1. cropland or 2. grassland (2hab). We further assessed whether apparent survival differed
298 between breeding and non-breeding seasons (ssn). Last, to obtain an accurate estimate of site fidelity
299 within the breeding season, we tested for a difference in detection probability between the first and
300 subsequent survey rounds within the T. Parker dataset, to account for difference in the fact that an
301 experienced observer (T. Parker) always participated in the first round, but not in later rounds (rnd). We
302 did not assess among-year variation in detection probability or apparent survival in our data because the
303 mark-resighting data were too sparse. Our global model therefore contained a habitat and season effect
304 on apparent survival (3hab + ssn) and a habitat and a survey round effect on the detection probability
305 (3hab + rnd) (Table 4).

306

307 We were not able to test the global model for overdispersion via the \hat{c} procedure in MARK because
308 of missing survey rounds in certain years. Instead, we assessed the potential impact of controlling for
309 overdispersion by manually increasing the variance inflation factor (\hat{c}) and assessing the resulting model
310 ranking (Appendix S3: Table S1). Our top two models were the same for \hat{c} 1 – 3, equivalent to
311 overdispersion that is absent to moderate. The order of lower ranking models changed when we
312 increased \hat{c} , but these models all received almost no support with \hat{c} set at 1 and 2, and still low support
313 when we set \hat{c} to 3. When we set \hat{c} to 4, corresponding to high overdispersion (Burnham and Anderson
314 2002), the model containing a survey round effect of capture probability dropped in the rankings, but
315 the otherwise identical model excluding the survey round effect remained top ranked. It was not until
316 we set \hat{c} to 5, a level that would indicate major problems with model fit (Burnham and Anderson 2002),
317 that the top model no longer included an effect of habitat on apparent survival, though at both \hat{c} 5 and
318 6, at least one model with $\Delta QAIC < 2$ included this habitat effect. Thus, we conclude that our model
319 selection results were robust to overdispersion. We fitted all models with design matrices and a logit-
320 link function. To avoid bias in parameter estimates, we generated weighted average estimates and
321 associated standard errors across all models using Akaike weights, and thus accounted for both sampling
322 and model-selection uncertainty (Burnham and Anderson 2002).

323
324

325 Creating song similarity score

326

327 We generated song similarity scores between all possible pairs of analyzed songs within each of our
328 seven sites. We included one song per male per year, but if the same banded male was present at a site
329 in multiple years, we included songs from each year available. We chose to include the same males from
330 multiple years because (a) an important component of cultural stability is likely the persistence of
331 individuals across years, (b) individual males can change their song from one year to the next
332 (unpublished data) and so retention of males across years need not necessarily increase estimates of
333 song sharing across years, and thus (c) a valid description of the degree of song stability over time
334 should include as many males present in each year as possible.

335

336 In addition to banded males, we also recorded unbanded males at each site. However, we only analyzed
337 recordings of unbanded individuals made on a single date each year at each site. On any given day, it is
338 relatively straightforward to keep track of unbanded individuals and to avoid recording the same
339 individual twice. This is because male dickcissels spend the majority of their time singing, typically from
340 prominent, easily visible perches, and territories are small enough (approximately 1-2 ha) (Zimmerman
341 and Finck 1989) that an observer can keep track of the movements of several individual males at a time.
342 However, if we had included songs recorded from unbanded individuals on different dates over the
343 course of the season, we could not be certain which songs were from the same males and which were
344 from different males. It is possible that we recorded at least some unbanded males in multiple years.

345

346 Prior to assessing similarity among recorded songs, we processed songs in Raven (version 1.4)
347 (Bioacoustics_Research_Program 2011) to optimize comparability. We first identified a song in which all
348 components were evident and high amplitude background noise did not overlap the song's notes. If no
349 single song met these criteria, we isolated a sufficient number of songs (typically no more than 4) such
350 that we could obtain a clear and unobscured version of each separate song phrase from among the
351 selected renditions. We then filtered out noise below 2000 and above 20000 kHz. Most background
352 noise is low frequency, and so filtering below 2000 kHz dramatically reduced noise. However, in most
353 recordings, notes produced by other individual birds, by insects, or by various other sources in the
354 environment remained in the recording. We could readily distinguish the notes of our target bird from

355 these background notes by observing longer stretches of the recording to see the pattern produced by
356 the target bird and the sources of background noise. We filtered out any background noise that
357 approached the amplitude of the target bird's quietest notes by using the select option in Raven and
358 filtering out the content of the selections.

359
360 We then divided each song into component phrase types, 'dick', 'cissel', and 'trill' (Figure 1), for further
361 analysis. Dicks are short phrases of relatively low frequency at the start of the song and are repeated a
362 variable number of times. All songs include dicks. A dick series is sometimes preceded by a long series of
363 dick-like phrases that differ in form from the typical dick phrases sung by the male in question. These
364 series appear to be sung by any given individual only rarely, possibly when agitated. We excluded these
365 putative agitation phrases from our analyses because they are not a consistent component of a male's
366 song. Cissels are longer and more complex and contain some higher frequency components. Cissels are
367 usually repeated two to three times in a pattern that is relatively consistent within birds, though
368 individuals occasionally will not sing the full set, and there is often variation in pattern within a site even
369 among males who sing cissel phrases that have approximately identical internal structure. Some cultural
370 variants typically lack cissels. Trills consist primarily of a series of relatively few notes repeated multiple
371 times in rapid succession with the pause duration between repeated sets of notes no longer than the
372 pauses between notes within a set. Some trills have a short set of introductory notes. A full trill series is
373 rarely repeated. Cissels occasionally contain a series of notes that might qualify as a trill, but because
374 cissel phrases have substantial proportions of non-trilled notes and the full phrase is usually repeated in
375 a predictable pattern, cissel phrases and trill phrases cannot be mistaken for each other. In some cases
376 an individual male sings more than one version of a given phrase. Almost never does an individual male
377 sing >2 versions of a given phrase type, and in the very few cases where it would have been possible to
378 classify >2 versions, differences between two of the versions were quantitative rather than qualitative
379 and relatively small. For this reason, and to avoid bias in comparisons between pairs of birds (see
380 Supplementary Materials, Appendix S2), we constrained our classification scheme to two versions per
381 phrase type per individual.

382
383 We next created a song sharing index between all possible pairs of individuals within each site, including
384 across years. This index combined quantitative assessment of similarity between phrases of the same
385 type with an assessment of the proportion of phrase types (broadly defined as dick, cissel, and trill)
386 shared between individuals. The index ranged from 0 to 1, with 1 indicating the two individuals sang
387 exactly the same song types with exactly the same arrangement of notes.

388
389 The quantitative similarity scores were based on comparisons between males within a shared phrase
390 type. We generated scores comparing dicks to dicks, cissels to cissels, and trills to trills. We did not
391 compare whole songs to each other because there is moderate variability within males and even greater
392 variability among males within a local cultural area regarding the number and timing of phrase types,
393 but high repeatability regarding the internal structure of phrase types. Thus focusing on individual
394 phrases allows us to focus on the most consistent component of vocal culture, and it is this consistency
395 we are seeking to quantify and explain.

396
397 The quantitative assessment of similarity within a phrase type was based on two distinct methods
398 (implemented in MATLAB version R2016b; see <https://osf.io/em43d/> for code). The first involved
399 assessing the similarity of note positions as evident in sound spectrograms (Große Ruse et al. 2016,
400 Sandsten et al. 2016). This multi-tapered spectrogram method compromises time and frequency
401 resolution in order to reduce noise sensitivity. This is accomplished by using multiple windows or
402 'multitapers'. The second method, which is particularly useful in recognizing repeated patterns within

403 phrases, is referred to as a filtered ambiguity spectrum (Große Ruse et al. 2016, Sandsten et al. 2016).
404 This is derived with a Fourier transformation of the spectrogram in the two dimensions of time and
405 frequency. Instead of marking the position or number of notes on the spectrogram, this method
406 determines time and frequency differences between notes, and is thus robust to slight differences in the
407 number or timing of repeated elements that are known to reduce estimates of similarity when focusing
408 entirely on the spectrogram. Because these two methods provide different information regarding
409 quantitative similarity, as part of the process of generating an overall average similarity score, we
410 averaged the pairwise comparison values in the two similarity matrices (see Supplementary Materials,
411 Appendix S2 for more details, R versions 3.3.3 and 3.4.2, see <https://osf.io/em43d/> for code).

412
413 Because these quantitative measurements only reflect shared phrase types, we needed to account for
414 cases in which males did not sing the same broadly defined phrase types. For instance if one individual
415 only sang dicks and cissels, and another only sang dicks and trills, our quantitative comparison would be
416 based entirely on similarity of dicks, but would ignore the dramatic difference in the other types of
417 phrases produced. Thus we calculated Jaccard's Index of similarity ($JI = \text{number of phrase types sung by}$
418 $\text{both individuals} / (\text{number of phrase types sung by both individuals} + \text{number of phrase types unique to}$
419 $\text{bird 1} + \text{number of phrase types unique to phrase 2})$, in which 0 = no overlap in phrase types and 1 =
420 production of an identical set of phrase types. Because Jaccard's index represents the proportion of
421 overlapping categories, we were able to multiply this value by the quantitative similarity averaged
422 across shared phrase types (which had been converted to range from 0 [no overlap] to 1 [songs
423 identical], as described in the supplement) to generate an intuitive score of overall song similarity
424 ranging from 0 to 1 between all pairs of individuals at each site (R versions 3.3.3 and 3.4.2, see
425 <https://osf.io/em43d/> for code).

426
427

428 Reliability of song similarity score

429
430 To evaluate the effectiveness of our automated method for assessing song similarity, we compared the
431 levels of song sharing between different recordings of the same individual bird in the same year (Table
432 5; within-individual) with recordings across individuals (within and across years; between individual) at
433 the same site. Our expectation was that if our method were useful, song sharing values within
434 individuals should be higher than average song sharing values from across-individual comparisons.
435 Human observers can readily recognize that two songs recorded from the same individual in the same
436 year are nearly identical, and that songs from other individuals differ, sometimes by a small amount, and
437 sometimes radically. To assess the reliability of the differences between within and between-individual
438 song sharing levels, we used boot-strapping to calculate 95% confidence intervals around the difference
439 between within-individual and between-individual song similarity scores within each site (R version
440 3.4.4, see <https://osf.io/em43d/> for code). If song sharing scores from within individuals tended to be
441 higher than scores from between individuals, the mean difference should be consistently positive.

442
443 Our automated measure of song similarity was modestly effective at recognizing the similarity between
444 songs produced by the same individual bird within the same breeding season. Within each of the seven
445 sites, the mean song similarity score for comparisons between songs of the same individual produced on
446 different dates in the same year was higher than the mean for between-individual comparisons (Figure
447 2), although typically by only approximately 0.1 correlation units (Table 6). Despite the relatively small
448 average differences, the 95% CIs of these differences were narrow in proportion to the size of the
449 differences and in all but one case, the lower bound was closer to the mean difference than to zero. In
450 no case did the lower bound of the CI encompass zero (Table 6). Thus, we have a consistent and clear,

451 but relatively weak difference in average song sharing scores between within-individual and between-
452 individual song comparisons. In light of this, it was not uncommon for single between-individual scores
453 to be higher than single within-individual scores (Figure 2). Because human observers can readily
454 distinguish songs from the same individual from most other songs produced at a site in a given year, and
455 a human observer would almost never rank two songs from the same individual as less similar than two
456 songs from different individuals, this suggests that the automated method produces only modestly
457 reliable measures of song similarity. However, we opted to use the song similarity scores from the
458 automated analyses because quantifying song similarity between all relevant pairs of songs with human
459 judgement would be subjective and prohibitively time consuming due to both the large number of
460 comparisons we wished to make and the difficulty in quantifying degree of similarity in songs when
461 those songs differ in many distinct ways. Further, in a preliminary analysis, we experimented with a
462 common form of song similarity quantification known as cross-correlation (Raven 1.3; Appendix S5), and
463 concluded that it was unsuitable. The cross-correlation analyses, while also producing higher values of
464 song similarity on average for within-individual comparisons than across-individual comparisons,
465 produced a much higher range of variability for within-individual comparisons (Appendix S5: Figure S1)
466 than did the method we used here (Figure 2).

467 468 Song similarity in biological context

469
470 Two of our predictions are that song sharing declines across time and across space. To assess these
471 predictions, within each of our four grassland and three cropland focal sites, we assessed the
472 relationships between song similarity and both (a) the time (in years) between individual recordings and
473 (b) the geographic distance between individual recordings (in meters). Because our measures of song
474 similarity and temporal and geographic distance involved all possible pairwise comparisons within a site,
475 each individual song was included in many song similarity measures. This violates the assumption of
476 parametric statistics that measures are independent of each other. So, instead we used a permutation
477 method that does not make this assumption (R version 3.4.4, see <https://osf.io/em43d/> for code). We
478 used MRM (multiple regression on distance matrices) as described by Lichstein (2007) and as can be
479 implemented in the *ecodist* package (Goslee and Urban 2017) in R. That is, we used linear models in
480 combination with randomized procedures to test the linear association of song-similarities with space
481 and time. In these analyses, the matrix of song-similarities was the dependent matrix, while the matrices
482 of temporal and spatial distances were the independent matrices. MRM randomly permutes the rows
483 and columns of the dependent matrix while the independent matrices remain fixed. We ran one
484 thousand permutations to generate a null distribution of relationships between song similarity scores
485 and both time differences and spatial distances. We could then compare the observed regression
486 coefficients to this null distribution of coefficients to evaluate the strength of the evidence against the
487 null hypothesis that song sharing was unrelated to time differences or spatial distance between
488 recordings. We generated 1-sided p-values because our hypotheses had clear directional predictions.
489 The p-value was the proportion of the regression coefficients from the permutation analysis that were
490 more steeply negative than the observed coefficients. We predicted negative slopes for the
491 relationships between temporal differences and song similarity and between geographic distance and
492 song similarity. Further, we considered the possibility that as time between recordings increased, the
493 geographic distance between those recordings would become a less effective predictor of the similarity
494 between those songs since song features could shift on the landscape across years. To evaluate this, we
495 tested for a positive interaction term between time and distance. We evaluated the importance of this
496 interaction by comparing the change in the R^2 value between the model containing only main effects
497 and the model with the interaction term. We used the `lm` function in base R (version 3.4.4,
498 (R_Core_Team 2019) to generate parameter estimates for each permutation using the methods

499 underpinning the ecodist package (Goslee and Urban 2017) as a model. We did not use the ecodist
500 package itself because that package did not allow us to assess the interaction models, nor could it
501 produce 1-tailed p-values.

502
503 Two other predictions of our hypotheses are that song sharing should decline more steeply over time
504 and over distance in locations where site fidelity is lower (cropland, see results) than in locations where
505 site fidelity is higher (grassland, see results). The analyses described above provided slopes and p-values
506 for the relationships between time and song sharing and between distance and song sharing within each
507 of our seven sites. This allowed us to assess our prediction qualitatively by simply asking if the slopes
508 were steeper for the three cropland sites than for the four grassland sites. However, to estimate how
509 likely it was that we would observe a difference in slopes (for song sharing vs. time lag and for song
510 sharing vs. geographic distance) as extreme as this by chance alone, we conducted another round of
511 permutations (R version 3.4.4, see <https://osf.io/em43d/> for code). In this analysis, we permuted the
512 song-similarity matrix for each of the seven sites 1000 times, each time calculating the corresponding
513 slope coefficients for each site for time vs. song sharing and for distance vs. song sharing. Then,
514 separately for the time and distance slopes, we calculated the average slope for each of the two location
515 types (cropland and grassland) and then calculated the difference in average slope between the two
516 location types. That is, for each permutation we calculated the slope of the four grassland sites and
517 three croplands sites, took the average of the slopes for the grassland sites and the average of the
518 slopes for the cropland sites, then took the difference between average slopes of cropland and
519 grassland sites. We repeated this process 1000 times. If our hypothesis were correct, we expected our
520 observed difference in average slope between cropland and grassland sites to be greater than most of
521 the simulated differences. We assessed this prediction separately for the time vs. song sharing slopes
522 and the distance vs. song sharing slopes.

523
524 In all our statistical analyses we follow recommendations to avoid dichotomous determinations of
525 ‘significance’ and instead to focus on strength of evidence, statistical clarity, and on avoiding over-
526 confidence in results from single studies (Stoehr 1999, Amrhein et al. 2017, Amrhein et al. 2019, Dushoff
527 et al. 2019). This means that we evaluated p-values as continuous evidence bounded by 0, representing
528 “complete incompatibility” between the data and the null hypothesis, and 1, representing “perfect
529 compatibility” with the null hypothesis (Greenland et al. 2016). Small differences in p-values, such as
530 between 0.1 and 0.05 show low repeatability, and thus should not be interpreted as meaningfully
531 different (Amrhein et al. 2017). In cases of relatively large p-values, indicating that the data would not
532 be particularly unlikely given the null, we interpret the observed pattern with caution and withhold
533 strong conclusions until further data can be gathered (Amrhein et al. 2017).

534 535 **Results**

536 537 **Mark-Recapture Analyses**

538
539 We found strong support for a relationship between our habitats and both apparent survival and
540 detection probability (3hab: Relative Importance [RI; the summed AICc Weights of all models containing
541 this variable] = 1.00; Table 4). This variable (3hab) also reflects the influence of differences in field
542 methods between the T. Parker and B. Sousa grassland data sets on estimates of apparent survival and
543 detection probability. Ignoring the differences in field methods between the T. Parker and B. Sousa data
544 sets by combining the apparent survival and detection probability for both grassland datasets (2hab)
545 had no support (Table 4). In contrast, we found strong support for a difference in apparent survival
546 within a breeding season versus over the non-breeding season (ssn: RI = 1.00; Table 4). Although the

547 top-ranked model indicated a difference between the first and subsequent survey rounds within a
548 breeding season in the T. Parker dataset (rnd; included because of differences in observers between first
549 and later survey rounds), model certainty was lower (rnd: RI = 0.57; Table 4) for this variable than for
550 3hab or ssn. As a result, the inclusion of this effect (rnd) on detection probability had only a minor effect
551 on the estimates of apparent survival.

552
553 Rates of apparent survival were substantially lower on cropland sites than on grassland sites when
554 examining the T. Parker data (with standardized methods across grassland and cropland sites). Within a
555 breeding season, dickcissels that bred in croplands had a 19.8% (95% CI = 11.4 – 28.1%) chance of
556 disappearing (dying or emigrating) in the 14-day period between survey rounds. However, birds
557 breeding in grassland had only a 7.5% (T. Parker; 95% CI = 4.0 – 10.9%) and 8.3% (B. Sousa; 95% CI = 4.2
558 – 12.4%) chance of disappearing over a 14 day period (Table 7). Over the 310 day period between
559 breeding seasons, apparent survival estimates had larger confidence intervals, and these intervals
560 overlapped between the grassland and cropland (Table 7). When assessed across the entire year (Table
561 7), as is relevant for overall population turnover, we found annual apparent survival for cropland that
562 was less than half of that for grassland for the T. Parker data. The annual apparent survival for grassland
563 from the B. Sousa data was intermediate. The detection probability of dickcissels was highest on the
564 grassland habitats within the B. Sousa dataset (0.927; CI = 0.874 – 0.959) as expected given the more
565 intense survey methods, and was somewhat higher in grassland habitats (0.758; CI = 0.699 - 0.808) than
566 in cropland habitats (0.593; CI = 0.436 - 0.733) within the T. Parker dataset, but with overlapping
567 confidence intervals (Table 8; we report weighted average estimates for all variables included in any
568 candidate model in Appendix S4: Table S1.

569
570

571 Song Similarity Analyses

572

573 The median song sharing scores based on within site and within year comparisons, but averaged across
574 all years, were higher at our grassland sites than at our cropland sites (Figure 2). When we instead
575 looked at song sharing scores from individual years separately, in three of five years all grassland sites
576 had higher median song sharing than all cropland sites. In the other two years, one grassland site had a
577 median song sharing value lower than a cropland site, but all the other grassland sites had higher values
578 than all cropland sites (Supplementary Materials, Appendix S6, Figure S1). However, we observed
579 considerable variability in song sharing within both types of sites, and interquartile ranges of grassland
580 and cropland sites often overlapped, thus undermining statistical clarity. The intercepts from the linear
581 models predicting song sharing from time and distance were all larger from the grassland sites than the
582 cropland sites in both the model with only main effects of distance and time (Table 9) and the model
583 including an interaction term between distance and time (Table 10). These differences are consistent
584 with higher song sharing in grasslands than in croplands within a given year (time 0) and between
585 individuals in close proximity (distance near 0).

586

587 All the relationships between song sharing and time within sites had negative slopes (Table 9, 10),
588 though p-values for each site ranged from small ($p_{\text{rand}} < 0.001$), indicating clear evidence of the negative
589 slope to moderately large ($p_{\text{rand}} = 0.33$), providing much weaker evidence of a negative slope. Although
590 these variable p-values suggest the relationships varied in strength across sites, negative slopes are
591 consistent with song sharing values across years having declined as the number of intervening years
592 increased. The three steepest declines of song sharing over time in the main-effects model were in the
593 three cropland sites (Table 9), and in the interaction model, three of the four steepest declines in song
594 sharing with time were in the three cropland sites (Table 10). The difference between the average slope

595 for cropland sites and the average slope for grassland sites in the main effects model was larger than
596 88% of these differences when randomized through permutation analyses. This is equivalent to a p-
597 value of 0.12, which should be interpreted in light of our small sample size ($n = 7$) as only moderate
598 evidence of a more rapid decline in song sharing over time in croplands than in grasslands.
599

600 At six of the seven sites, relationships between song sharing and geographic distance had negative
601 slopes (Table 9, 10, p_{rand} for negative slopes ranged from $< 0.0001 - 0.260$), indicating with varying
602 degrees of confidence that song sharing values declined as the geographic distance between individuals
603 increased even on these relatively small sites. However, we could discern no clear pattern of differences
604 in these slopes between cropland and grassland sites, in part because the slopes varied substantially
605 among sites within habitat types (Table 9, 10). Consistent with this lack of evident pattern, when we
606 conducted a permutation test we found that differences between the average slope for cropland sites
607 and the average slope for grassland sites in the main-effects model was larger than only 30% of the
608 randomized differences. This corresponds to a p-value of 0.70, which is a lack of clear evidence
609 regarding any differences in the geographic decline in song sharing between cropland and grassland.
610

611 Surprisingly, we lacked consistent evidence for a strong positive interaction between time and distance.
612 For two of the seven sites, the interaction term was negative, counter to prediction. For the five sites
613 with positive interactions, including the interaction term increased the R^2 by $< 1\%$ over the main effects
614 model in two cases and $>5\%$ over the main effects model in only two cases (Table 11).
615

616 Discussion

617 Our results are consistent with the hypotheses that song sharing in dickcissels declines over time and
618 space and that vocal culture in dickcissels changes more quickly over time in cropland landscapes where
619 population turnover is faster than in grassland where populations are more stable. Within a given
620 breeding season, dickcissels were more than twice as likely to disappear from our cropland sites, where
621 crop harvest in mid-breeding season presumably often destroyed nests, than from our grassland sites.
622 At all sites, similarity between songs declined as the number of years between the recording of those
623 songs increased thus indicating cultural change over time, though these trends were not all equally
624 convincing. The slopes of these relationships between time and song sharing were steepest for the three
625 cropland sites, indicating fastest cultural change at those locations. The statistical evidence for this
626 pattern was only moderately clear, however, and so further work will be needed to assess it. Song
627 similarity between pairs of dickcissels also declined over geographic distance within 6 of the 7 sites, but
628 there were no obvious trends in the differences in this relationship between cropland and grassland
629 sites, and so it may be that geographic changes in vocal culture are not strongly driven by differences in
630 population turnover, at least at the scale of our sampling. It may also be that the spatial decline in song
631 sharing happens over shorter distances in grasslands than in croplands, counter to our prediction.
632 However, we have modest evidence of lower overall song sharing in cropland than in grassland sites and
633 so population turnover may influence variation in cultural conformity. It is notable that we were able to
634 detect these patterns despite having only a moderately reliable measure of song similarity. That said,
635 although these patterns are consistent with several of our hypotheses, we cannot consider our results
636 definitive because we studied only seven sites (4 grassland, 3 cropland), and so we lack a sufficient
637 sample for robust comparisons of grassland and cropland sites in general.
638
639

640 Our observation of change in vocal culture over time represents important additional empirical evidence
641 for a rarely studied phenomenon. Empirical evidence from only a handful other species demonstrates
642 that vocal culture in songbird populations can change over time (McGregor and Thompson 1988,

643 Holland et al. 1996, Derryberry 2009, Ju et al. 2019). However, we expect that in species where
644 individuals must learn songs from conspecifics, local culture should change whenever a non-local song
645 variant is imported by immigrants and copied by residents, or whenever a local song variant is copied
646 imperfectly (Williams and Slater 1990).

647
648 Another important observation was that the cropland sites, which had much lower site fidelity than the
649 grassland sites, showed evidence of faster change in vocal culture. Although this pattern is intuitive and
650 is consistent with simulation models (Goodfellow and Slater 1986, Nunn et al. 2009), we know of no
651 prior empirical study that has sought to link these two processes. It seems plausible that both
652 importation of new song features and the accumulation of learning errors drive change in our system
653 because we have observed both the appearance of dramatically different song components and slight
654 changes in note duration, frequency, and shape (Supplementary Materials, Appendix S7, Fig S1).
655 Regardless of which mechanism of song change is most important, opportunities for new males to
656 colonize sites were more than twice as frequent on our cropland sites than on our grassland sites.

657
658 Although it may be that reduced site fidelity drives lower levels of song sharing within years in our
659 system, it has been hypothesized that such an effect depends on features of song learning and culture
660 that vary among species (Ortega et al. 2014). In the dickcissel, dialect conformity is high, and each male
661 sings a single song. Thus when established males emigrate, this provides opportunities for new songs to
662 enter the local culture via immigration or learning errors. However, in species in which individuals each
663 sing multiple song types and these different song types are learned from different tutors, a reduction in
664 site fidelity could reduce the diversity of adult tutors, and thus the potential diversity of song types
665 available for copying. This is what Ortega et al. (2014) found in chipping sparrows (*Spizella passerina*),
666 where locations with lower site fidelity had lower diversity of song types. Thus to predict the role of site
667 fidelity on song sharing for any given species, we should know the degree of local conformity, number of
668 song types sung per male, and possibly also the number of tutors typically imitated by males learning
669 songs, especially if males sing multiple song types.

670
671 Knowledge of the number of tutors could potentially provide other insights into the development of
672 vocal culture as well. Unfortunately, we do not yet know the details of how dickcissel males learn song
673 in a vocal neighborhood or how many males in a neighborhood tend to converge on a single song type in
674 a stable population. It may be that males prefer to learn the most common song, but that low site
675 fidelity leads simultaneously to low abundance of even the most common songs and to frequent
676 introductions of new songs that are more likely to be copied than they might be if one existing song
677 were already widely sung. To understand why low song sharing is associated with low site fidelity in
678 dickcissels, we should investigate how dickcissels decide which songs to copy.

679
680 Although song sharing declined over geographic distance in all four grassland sites and two of three
681 cropland sites, the slopes of these relationships were not consistently steeper in either habitat type. This
682 suggests that low site fidelity may not lead to a faster rate of decline in song sharing over geographic
683 distance, or even that low site fidelity may reduce this decline in song sharing over space. However, the
684 intercept we observed in our models was consistently lower for cropland sites than for grassland sites,
685 and so if slopes of decline do not differ consistently between these two types of sites, at any given
686 distance, song sharing should on average be lower in our cropland sites than in our grassland sites. This
687 said, we should also consider the possibility that the geographic scale at which we sampled within these
688 sites may be too small for an ideal test of this effect. Earlier research on dickcissels observed declines in
689 song sharing even at relatively close distances, though the largest data set showed the steepest declines
690 in song sharing between 1 and 2 km (Schook et al. 2008). However, within our sites for the current

691 study, nearly all individuals were within 1 km of each other. Thus, before we dismiss the hypothesis that
692 reduced site fidelity promotes more rapid decline in song sharing with geographic distance in dickcissels,
693 we should examine song sharing patterns over longer distances, and of course with larger samples.
694

695 Habitat features can influence the evolution of avian vocal culture through direct selection on song
696 propagation (Ey and Fischer 2009), but we are reasonably confident that this mechanism is not driving
697 the patterns that we report here. We are confident because in all of our study sites, and across the
698 geographic range of this species, males sing from exposed perches in areas of either uniformly low
699 vegetation or with only widely scattered higher trees, and their songs propagate primarily through the
700 open air with few or no obstructions. Another possible source of selection on vocal culture may be
701 ambient noise (Slabbekoorn and den Boer-Visser 2006), and two of our grassland sites were close
702 enough to a major highway that, in places on these sites, ambient noise may have been somewhat
703 elevated. However, for most of our birds in both cropland and grassland sites, ambient noise did not
704 obviously differ. Thus, neither of the two environmental selection pressures demonstrated to influence
705 song culture evolution in birds appears likely to differ between our site types.
706

707 We generated two separate estimates of apparent survival of territorial male dickcissels occupying
708 grassland sites because we based these estimates on two data sets gathered with different methods.
709 These two data sets produced fairly different apparent survival estimates, and we think that this
710 represents differences in the set of individuals who were marked and observed. With one set of
711 methods, researchers (led by T. Parker) banded birds beginning approximately two weeks after male
712 dickcissels began settling on territories. These researchers banded birds for two weeks, and then
713 surveyed each site for banded birds once every two weeks, except in seasons where they surveyed sites
714 only once. With the other set of methods, researchers (led by B. Sousa) banded and surveyed birds
715 nearly daily throughout the season. This latter set of methods often captured birds who did not appear
716 to be long-term territory holders. Early in the season when male dickcissels were first arriving, these
717 briefly-present individuals may have been prospecting for territories during migration stopover. Later in
718 the season, such individuals likely were adults who had abandoned a territory elsewhere and were
719 wandering the landscape, possibly prospecting for territories for the following year. The inclusion of
720 these individuals that never established long-term territories seems likely to have led to the lower rates
721 of apparent survival estimates from the B. Sousa data. For this reason, we feel confident that the B.
722 Sousa grassland data are not as appropriate as the T. Parker grassland data for comparison to the T.
723 Parker cropland data, where the methods of capture and re-sighting were standardized across grassland
724 and cropland sites. Thus we also feel confident that the relative differences in apparent survival
725 between the grassland and cropland sites calculated from the T. Parker data represent real biological
726 differences in site fidelity between these habitats in the sites we studied.
727

728 We hope our results stimulate further research into the relationship between population turnover and
729 rates of cultural change. Assessment of the validity of our findings with a larger sample of sites would be
730 useful. As our ability to quantify differences between songs of pairs of birds continues to improve, we
731 will be able to estimate more precisely the strength of these effects. As we mention above, exploring
732 how individual colonists choose song tutors should help us understand the mechanistic links between
733 population turnover and cultural change. We also hope that researchers begin to test for relationships
734 between population processes and cultural change across a range of taxa and types of culture. Certainly
735 we hope that bird song researchers investigate this link, but it seems plausible that population turnover
736 could influence change in a wide range of cultural attributes. These questions may be especially
737 relevant given the growing recognition that cultural diversity is a component of biodiversity and is
738 worthy of conservation (Ryan 2006, Laiolo and Tella 2007).

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Author contributions are as follows:

Author	Study conception / design	data acquisition	data analysis	data interpretation	writing manuscript	editing manuscript
T. H. Parker	x	x	x	x	x	x
B. Sousa		x				
S. T. Leu			x			x
S. Edmondson			x	x		x
C. Foo		x	x	x		
A. Strauss		x		x		
H. Kahl		x		x		x
K. Ballinger		x		x		
E. Ross		x		x		
M. Große Ruse			x			
M. Sandsten			x			x
B. Verheijen			x	x		x
W. Jensen	x			x		x

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902 **Table 1.** Locations and sizes of study sites. Sites at Konza Prairie Biological Station (KPBS) are KPBS
 903 designated ‘watersheds’ and the number in the name refers to the number of years between controlled
 904 early spring burns (prior to settlement by dickcissels). We did not necessarily incorporate an entire
 905 watershed into the study, and the area measurement we present represents the portion of the
 906 watershed(s) we used.

Site	type	latitude	longitude	approx. area (ha)#
KPBS				
R20A / 2A	grass	39.06	-96.60	42
20C / 1D*	grass	39.06	-96.56	27
4F	grass	39.07	-96.57	16
C1B / C4A	grass	39.09	-96.54	37
North	crop	39.21	-96.59	44
Ashland	crop	39.12	-96.61	25
Strauss	crop	39.39	-96.85	18

908 # These areas do not match the area of the KPBS designated watersheds because in most cases we did
 909 not include entire watersheds in our study.

910 * All song recordings but one come from 20C, but birds banded and monitored for the mark-recapture
 911 study by B. Sousa were in both 20C and adjacent 1D.

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914 **Table 2.** Site by site summary of sampling. We did not visit all sites in all years. Number of visits varied
 915 among years. Sampling carried out by T. Parker and students identified by 'P'. Sampling carried out by B.
 916 Sousa and associates identified by 'S'. Song recording methods were similar between these two groups,
 917 though recording equipment differed. Monitoring protocols for mark/re-sighting differed substantially
 918 between P and S teams, and thus we considered these data separately for analyses. In 2008 and 2009
 919 both P and S teams monitored birds in 20C. For any birds monitored by both T. Parker and B. Sousa
 920 teams, we included only the T. Parker data in the mark-recapture analysis so that we could generate
 921 separate mark-recapture estimates for T. Parker and B. Sousa data.
 922

Site		year						
		2006	2007	2008	2009	2010#	2011	2014*
R20A / 2A	visits	4	5	1	0	4	1	1
	mark/re-sight	S	S	S		P	P	P
	song recording	P, S	S	S		P	P	P
20C / 1D	visits	1	1	3 (1)	5 (5)	4	1	0
	mark/re-sight	P	P	P (S)	P (S)	P	P	
	song recording	P	P	P, S	P, S	P	P	
4F	visits	1	0	3	4	4	1	0
	mark/re-sight	P		P	P	P	P	
	song recording	P		P	P	P	P	
C1B	visits	0	0	4	4	4	1	1
	mark/re-sight			P	P	P	P	
	song recording			P	P	P	P	P
North	visits	0	0	4	5	4	1	1
	mark/re-sight			P	P	P	P	
	song recording			P	P	P	P	P
Ashland	visits	0	0	0	5	4	1	1
	mark/re-sight				P	P	P	
	song recording				P	P	P	P
Strauss	visits	0	0	0	0	4	1	1
	mark/re-sight					P	P	
	song recording					P	P	P

923 # last year birds were banded

924 * data not used for mark-recapture analysis due to the lack of sampling over prior 2 years and the 4
 925 years since last birds were banded.

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928 **Table 3.** Number of songs analyzed per year for each site. Within a given year, each song was recorded
929 from a different individual male dickcissel.

Site	2006	2007	2008	2009	2010	2011	2014
2A-R20A	32	40	15	0	25	33	5
4F	15	0	9	12	11	10	0
20C	26	3	10	19	14	9	0
C1B	0	0	14	25	25	18	5
Ashland	0	0	0	11	5	6	5
North	0	0	17	16	10	9	7
Strauss	0	0	0	0	15	10	5

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935 **TABLE 4.** Cormack-Jolly-Seber (CJS) model selection results for dickcissels captured and resighted in
 936 northeastern Kansas, USA, between 2006 and 2011. CJS models estimate apparent survival (ϕ) and
 937 probability of capture (p). Model selection was based $\Delta AICc$ values and Akaike weights (w_i). Apparent
 938 survival was modeled separately for cropland and grassland habitats (2hab), for cropland and the two
 939 sources of grassland habitat data separately (3hab), for periods within and between breeding seasons
 940 (ssn), or kept constant (con). Detection probability was modeled separately for cropland and grassland
 941 habitats (2hab), for cropland and the two sources of grassland habitat data separately (3hab), for the
 942 first survey of each year and subsequent surveys rounds in the T. Parker dataset (rnd), or kept constant
 943 (con).
 944

Model Structure		Model Statistics				
ϕ	p	K	Deviance	AICc	$\Delta AICc$	w_i
3hab + ssn	3hab + rnd	11	302.40	1291.46	0.00	0.574
3hab + ssn	3hab	9	307.10	1292.06	0.60	0.426
2hab + ssn	3hab + rnd	9	324.17	1309.13	17.67	0.000
2hab + ssn	3hab	7	328.81	1309.69	18.23	0.000
3hab + ssn	2hab	8	330.04	1312.95	21.49	0.000
ssn	3hab + rnd	7	333.85	1314.73	23.27	0.000
3hab	3hab + rnd	8	333.18	1316.09	24.63	0.000
ssn	3hab	5	341.22	1318.03	26.57	0.000
3hab	3hab	6	342.91	1321.75	30.29	0.000
3hab + ssn	con	7	342.33	1323.21	31.75	0.000
2hab + ssn	2hab	6	350.70	1329.54	38.08	0.000
2hab	3hab + rnd	7	352.26	1333.14	41.68	0.000
con	3hab + rnd	6	358.29	1337.13	45.67	0.000
2hab	3hab	5	361.82	1338.63	47.17	0.000
ssn	2hab	4	364.68	1339.47	48.01	0.000
2hab + ssn	con	5	362.77	1339.58	48.12	0.000
con	3hab	4	369.51	1344.30	52.84	0.000
3hab	2hab	5	369.18	1345.99	54.53	0.000
2hab	2hab	4	381.07	1355.86	64.40	0.000
ssn	con	3	386.85	1359.62	68.16	0.000
con	2hab	3	389.28	1362.05	70.59	0.000
3hab	con	4	393.99	1368.78	77.32	0.000
2hab	con	3	405.45	1378.22	86.76	0.000
con	con	2	426.46	1397.22	105.76	0.000

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948 **Table 5.** Number of individuals from which two songs, each from a different date, were analyzed to
 949 assess the sensitivity of our song comparison method.

Site	2006	2007	2008	2009	2010	2011	2014
grassland:							
2A-R20A	12	0	0	0	14	0	0
4F	4	0	7	6	6	0	0
20C	6	0	3	10	6	0	0
C1B	0	0	0	4	2	0	0
cropland:							
Ashland	0	0	6	5	9	0	0
North	0	0	6	6	2	0	0
Strauss	0	0	0	0	5	0	0

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952 **Table 6.** Differences between within individual and across individual measures of song similarity (\pm 95%
 953 bootstrap confidence interval) by site. All differences are > 0 demonstrating that, on average, within-
 954 individual song comparisons generated higher song sharing values than among-individual song
 955 comparisons within the same site.
 956

site	lower confidence limit		average difference		upper confidence limit
grassland:					
2A-R20A	0.077	–	0.109	–	0.140
4F	0.069	–	0.101	–	0.131
20C	0.009	–	0.042	–	0.074
C1B	0.067	–	0.097	–	0.127
cropland:					
Ashland	0.055	–	0.105	–	0.150
North	0.110	–	0.162	–	0.214
Strauss	0.203	–	0.238	–	0.271

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962 **TABLE 7.** Estimates of daily apparent survival (ϕ) and standard errors for dickcissels monitored between
 963 2006 and 2011 in northeastern Kansas, USA, estimated separately for birds breeding in cropland and
 964 grassland habitats and for birds monitored by T. Parker’s team (P) and B. Sousa’s team (S). We estimated
 965 apparent survival for the periods within and between breeding seasons separately, and calculated
 966 apparent survival, standard errors, and 95% confidence intervals for a 14-day period between two
 967 survey rounds within each breeding season, and for a 310-day period between breeding seasons. These
 968 estimates and associated SE and 95%CI are model averages weighted by Akaike weights.
 969

Specific Interval	Habitat	Team	ϕ (daily)	SE	ϕ (period)	SE	95% CI
Within Breeding Seasons (14-Day Interval)	Cropland	P	0.984	0.004	0.802	0.043	0.719 - 0.886
	Grassland	P	0.994	0.001	0.925	0.018	0.891 - 0.960
	Grassland	S	0.994	0.002	0.917	0.021	0.876 - 0.958
Between Breeding Seasons (310-Day Interval)	Cropland	P	0.998	0.001	0.511	0.131	0.255 - 0.767
	Grassland	P	0.999	0.000	0.730	0.055	0.621 - 0.838
	Grassland	S	0.997	0.001	0.358	0.057	0.246 - 0.469
Between Years (365-Day Interval)	Cropland	P	0.995	0.001	0.202	0.061	0.095 - 0.335
	Grassland	P	0.998	0.000	0.537	0.048	0.443 - 0.631
	Grassland	S	0.996	0.000	0.254	0.044	0.168 - 0.341

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972 **TABLE 8.** Estimates of the detection probability (p), standard errors and 95% confidence intervals for
 973 dickcissels monitored between 2006 and 2011 in northeastern Kansas, USA. We estimated detection
 974 probabilities separately for birds breeding in cropland and grassland habitats, for birds monitored by T.
 975 Parker’s team (P) and B. Sousa’s team (S), and for the first survey of each year and subsequent surveys
 976 rounds in the P dataset to correct for a potential observer effect (see methods). These estimates and
 977 associated SE and 95%CI are model averages weighted by Akaike weights.
 978

Specific Interval	Habitat	Source	p	SE	95% CI
1st Survey of the Year	Cropland	P	0.713	0.088	0.516 - 0.853
Other Surveys	Cropland	P	0.593	0.078	0.436 - 0.733
1st Survey of the Year	Grassland	P	0.771	0.040	0.682 - 0.840
Other Surveys	Grassland	P	0.758	0.028	0.699 - 0.808
All Surveys	Grassland	S	0.927	0.021	0.874 - 0.959

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984 **TABLE 9.** The relationships between pairwise song similarity values and geographic distance between
 985 pairs of recordings and time difference (in years) between pairs of recordings, along with the overall
 986 model R² value (song sharing ~ geographic distance + time difference). Each estimate is accompanied by
 987 a one-tailed p-value.
 988

Site	intercept	<i>p</i> _{rand}	distance slope	<i>p</i> _{rand}	time slope	<i>p</i> _{rand}	overall R ²	<i>p</i> _{rand}
grassland:								
R20A-2A	0.618	0.001	-1.25E-05	0.260	-7.03E-03	0.000	0.022	0.000
20C	0.621	0.000	-6.64E-05	0.000	-1.95E-03	0.106	0.039	0.000
4F	0.654	0.180	-3.11E-05	0.205	-9.30E-04	0.334	0.003	0.602
C1B	0.626	0.008	-3.89E-05	0.011	-4.49E-03	0.138	0.016	0.070
cropland:								
Ashland	0.588	0.005	1.28E-05	0.658	-1.21E-02	0.002	0.085	0.009
Strauss	0.552	0.062	-3.26E-05	0.133	-8.49E-03	0.117	0.025	0.205
North	0.519	0.001	-5.16E-05	0.009	-8.10E-03	0.017	0.029	0.029

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992 **TABLE 10.** The relationships between pairwise song similarity values and geographic distance between
 993 pairs of recordings and time difference (in years) between pairs of recordings, an interaction term
 994 between distance and time, and the overall model R² value (song sharing ~ geographic distance + time
 995 difference + geographic distance*time difference). Each estimate is accompanied by a p-value, and p-
 996 values are one-tailed for the distance and time slopes and the distance x time interaction.
 997

Site	int.	<i>p</i> _{rand}	distance slope	<i>p</i> _{rand}	time slope	<i>p</i> _{rand}	distance x time	<i>p</i> _{rand}	overall R ²	<i>p</i> _{rand}
grassland:										
R20A-2A	0.624	0.001	-3.10E-05	0.076	-9.89E-03	0.000	8.00E-06	0.032	0.023	0.000
20C	0.621	0.000	-6.80E-05	0.000	-2.32E-03	0.180	1.00E-06	0.406	0.039	0.000
4F	0.655	0.193	-3.30E-05	0.208	-1.14E-03	0.354	1.00E-06	0.467	0.003	0.626
C1B	0.622	0.022	-2.70E-05	0.049	-1.76E-03	0.292	-8.00E-06	0.896	0.017	0.074
cropland:										
Ashland	0.580	0.193	3.90E-05	0.878	-7.12E-03	0.117	-1.40E-05	0.864	0.090	0.010
Strauss	0.555	0.098	-4.00E-05	0.162	-1.06E-02	0.131	5.00E-06	0.333	0.025	0.244
North	0.527	0.000	-7.50E-05	0.003	-1.27E-02	0.005	1.30E-05	0.030	0.031	0.005

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1000 **TABLE 11.** Comparison of the model fit, as measured by R^2 , between models that use only main effects
 1001 (geographic distance between pairs of recordings, time difference [in years] between pairs of
 1002 recordings) to explain pairwise song similarity and models that use both these main effects and an
 1003 interaction between the two to explain pairwise song similarity.
 1004

Site	main effects R^2	interaction R^2	absolute increase in R^2	percent increase in R^2
grassland:				
R20A-2A	0.02206	0.02295	0.00088	4.01 %
20C	0.03941	0.03943	0.00002	0.04 %
4F	0.00262	0.00262	0.00000	0.19 %
C1B	0.01652	0.01731	0.00080	4.83 %
cropland:				
Ashland	0.08547	0.08985	0.00438	5.12 %
Strauss	0.02504	0.02537	0.00033	1.32 %
North	0.02889	0.03109	0.00220	7.63 %

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1013 Figure headings

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1015 **Figure 1.** Songs of multiple male dickcissels from two sites (C1B, a grassland site; North Farm, a cropland
1016 site) in 2008. Panels A – G and M and N all depict songs with two phrase types: one brief ‘dick’ phrase
1017 often repeated several times followed by several (in this case, always three) very similar ‘cissel’ phrases.
1018 The songs in panels H – K also begin with ‘dick’ phrases, but then are followed by a ‘trill’ phrase and then
1019 by two ‘cissel’ phrases. Panels A and B depict songs produced by the same male on two different dates,
1020 as do panels C and D, H and I, and M and N, with each of these pairs of panels demonstrating the high
1021 degree of within-male song consistency. We can also see similarity between the songs of different males
1022 within a site, for instance the dick phrases in panels A – E and the cissel phrases in A – F are all very
1023 similar as are the dick phrases in panels H – N, the cissel phrases in H – K, and the trill phrases in H – L.
1024 However, even very similar phrases differ in consistent ways between individual birds. For instance, the
1025 arrows in panels A and B identify a small note that is shorter and of higher frequency than the
1026 corresponding note identified by the arrows in panels C and D. Similarly, the arrows in panels H and J
1027 indicate the final series of notes in the cissel phrase which show differences in the pattern of frequency
1028 modulation and duration. The arrow in panel I indicates a pure-tone note which remains on a single
1029 frequency, while the corresponding note in panel K descends in frequency. Also in panel J, notice that
1030 the trill has an introductory portion that differs from the trills in panels H, I, K, and L.

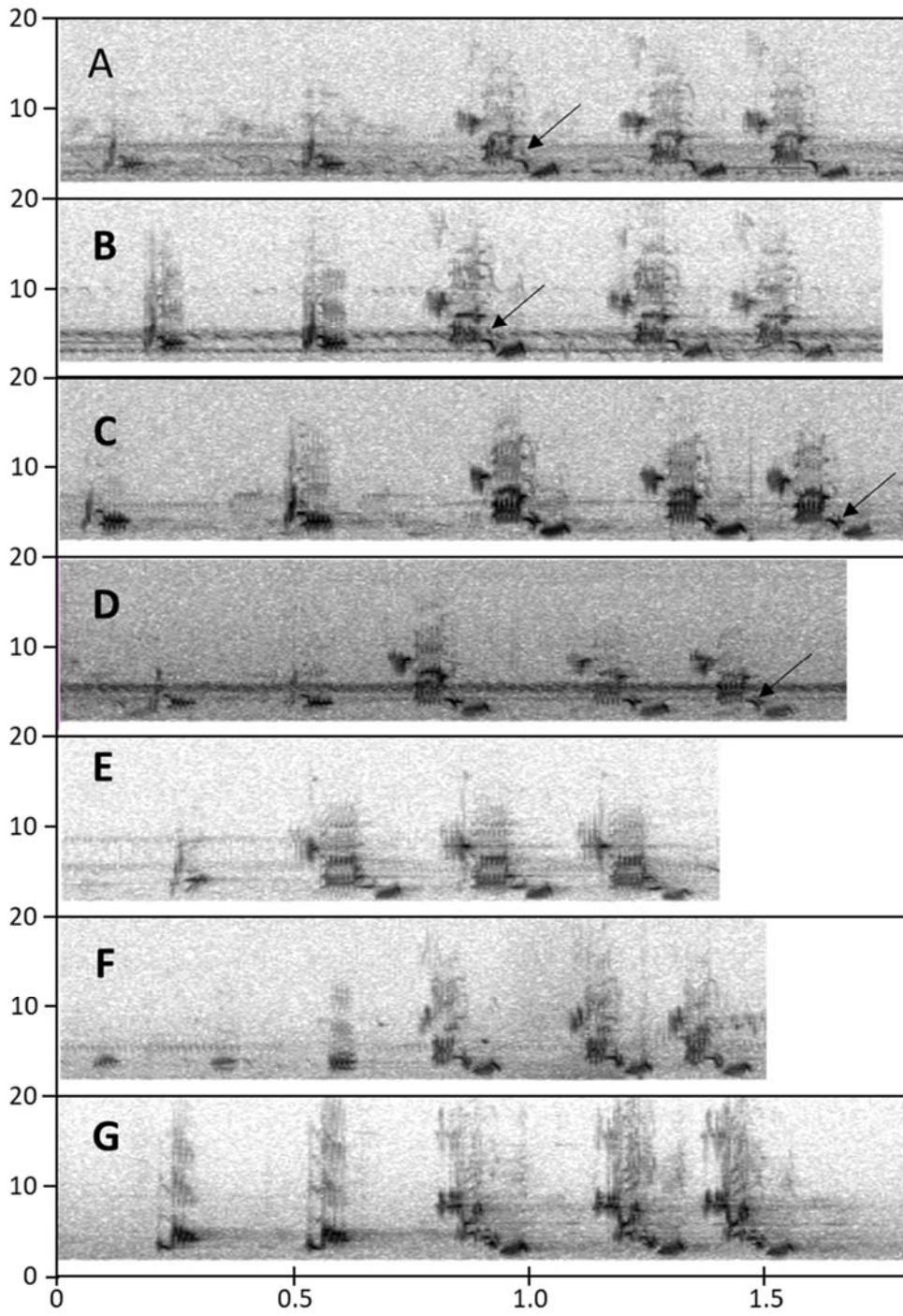
1031

1032 **Figure 2.** Plots of median song sharing values for comparisons of recordings of the same individual on
1033 different days within a given year (dark grey boxes) at each of the seven sites and for comparisons of
1034 recordings of songs different individuals at these same sites (light gray boxes). The central line is the
1035 median and the box represents the interquartile range (middle 50th percentile).

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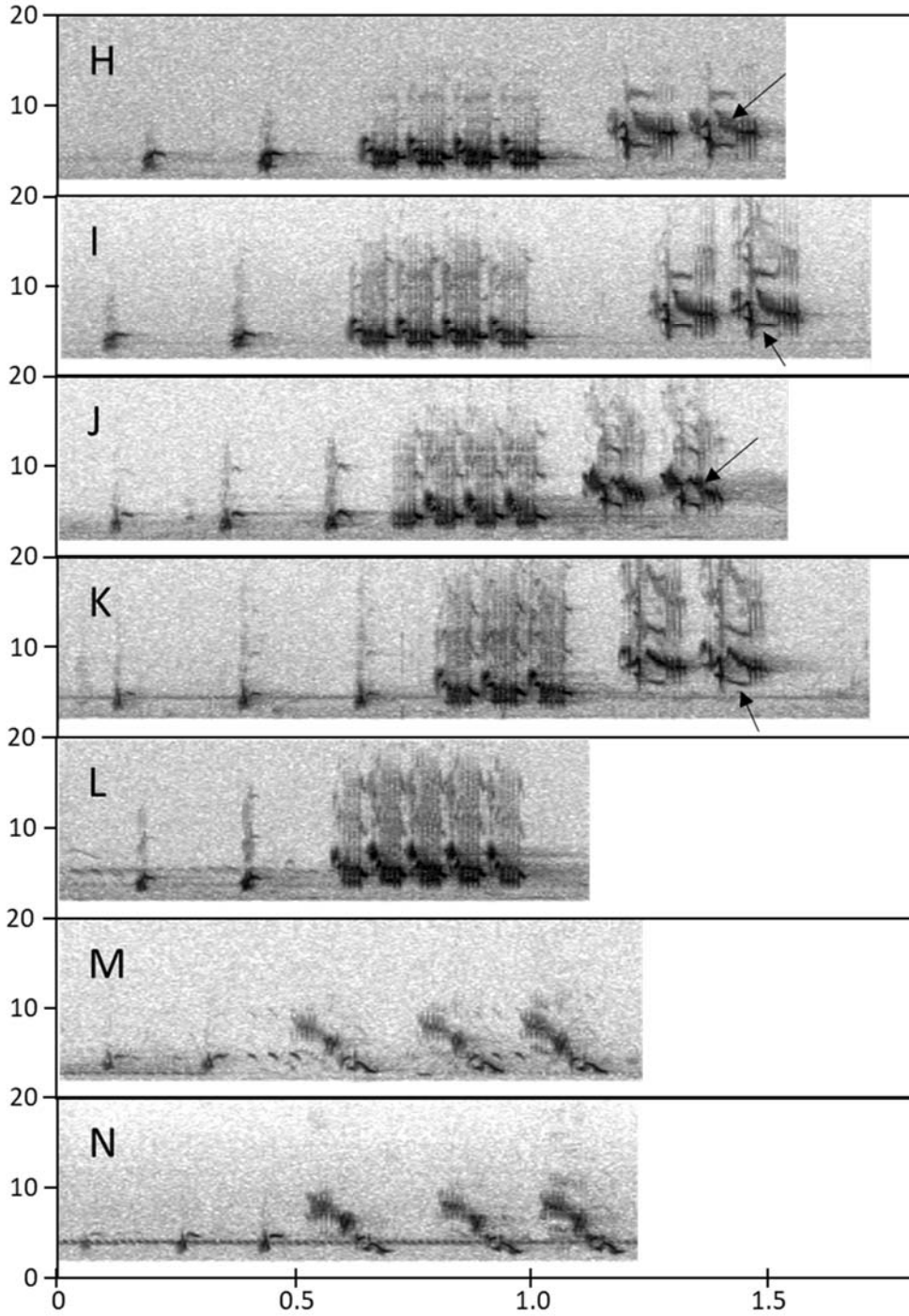
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1038 Figure 1
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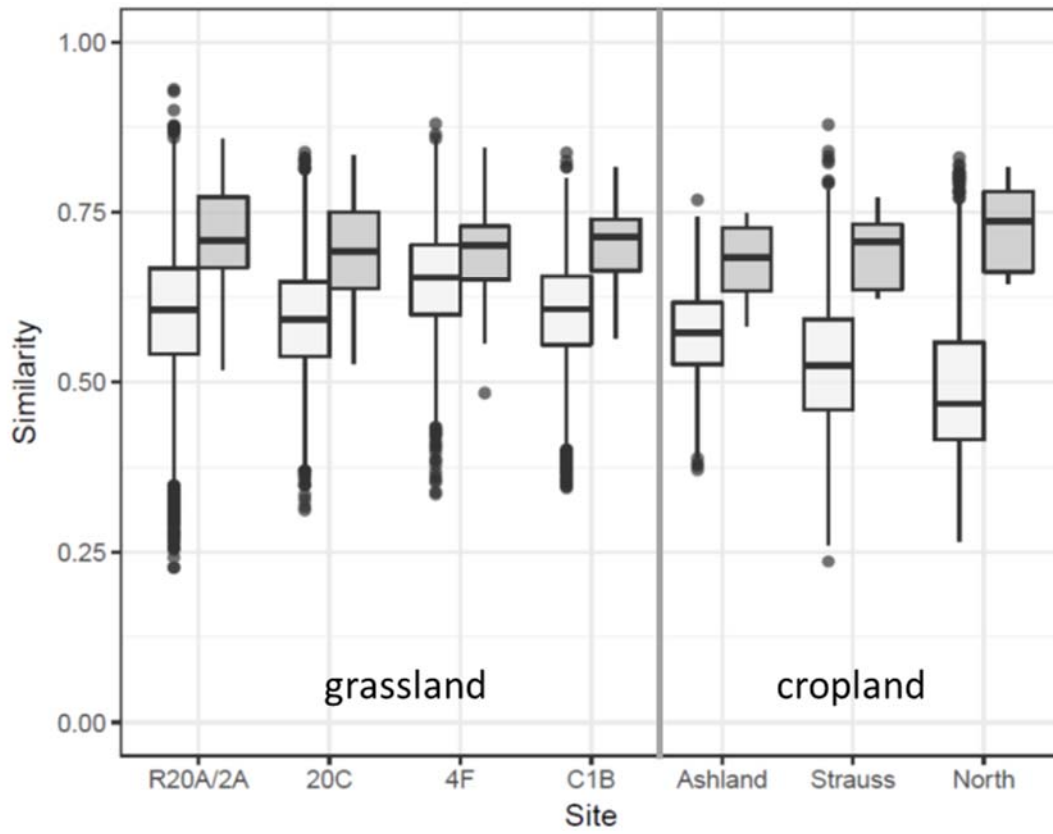
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1044 Figure 2
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1055 Supplementary materials

1056

1057 Appendix S1

1058

1059 Our interests in dickcissel song and our methods of study have evolved over time. We began recording
1060 dickcissel song in 2005 and 2006 to assess the geographic extent and pattern of song sharing (Schook et
1061 al. 2008). We were also curious about within-individual variation in song, and so in 2006 began color-
1062 banding individuals and returning to record individual birds within and across seasons. We continued
1063 recording marked birds in 2007 to explore the possibility of change in song across years. In 2008, we
1064 initiated the project that is the subject of this paper to study whether changes in population turnover
1065 would be associated with the rate of temporal or spatial change in song culture, and so we began a
1066 larger effort to mark birds at our existing grassland sites and at a cropland site. In both 2009 and 2010,
1067 we added a new cropland site and continued marking birds at all of our sites. We returned to record
1068 songs, but not to mark new birds, at most sites in 2011 and 2014. A parallel study recorded song and
1069 banded dickcissels at several grassland sites as part of an exploration of sexual selection in this species in
1070 2006-2009 (Sousa and Westneat 2013b, a), and we include these recordings in the analyses for this
1071 current paper.

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1076 Schook, D. M., M. D. Collins, W. E. Jensen, P. J. Williams, N. E. Bader, and T. H. Parker. 2008. Geographic
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1078 Sousa, B. F., and D. F. Westneat. 2013a. Positive association between social and extra-pair mating in a
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1080 67:243-255.

1081 Sousa, B. F., and D. F. Westneat. 2013b. Variance in mating success does not produce strong sexual
1082 selection in a polygynous songbird. *Behavioral Ecology* 24:1381-1389.

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1092 Appendix S2.

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1094 Methods - Additional details regarding calculation of song similarity scores

1095

1096 Our scheme of classifying songs as dicks, cissels, and trills works well, but it is not perfect. It appears, for
1097 instance, that short dick phrases, which are sometimes sung in rapid series, may evolve over time into
1098 trills (T. Parker, unpublished observations), and thus there are phrases which could be classified into
1099 either category. However, of the many hundreds of songs we have processed and classified, the vast
1100 majority required no subjective judgement regarding classification. In most cases, we could adopt a
1101 simple decision rule to ensure consistency. For instance, in cases where dick phrases could have been
1102 judged to be either dicks or part of a trill series, we considered them dicks if they matched other phrases
1103 at the site that were unambiguously classified as dicks.

1104

1105 When comparing song phrases with our quantitative methods, if males at a site sang two versions of the
1106 same phrase types, e.g., two distinct dick phrases, we classified the phrases of that type as 'one' and
1107 'two', e.g., 'dick1' and 'dick2', and compared all 'one' phrases to each other and all 'two' phrases to each
1108 other but not the 'one' phrases to the 'two' phrases. If two males at a site sang the same two versions of
1109 a phrase, they always sang them in the same order, so 'one' and 'two' were not arbitrary labels but
1110 reflected the order in which the two versions appeared. If a male sang one commonly shared version
1111 and one rare version or locally unique version, we assigned the commonly shared version to category
1112 'one' or 'two' to match the position in which that phrase version was typically sung at the site so that we
1113 could compare like to like. If some males sang two versions of a phrase type and other males sang only
1114 one version of that phrase type, for the males who sang only one version, we saved two copies of that
1115 version, one as version one and the other as version two. This allowed us to assess the quantitative
1116 similarity of the given phrase type between males even if they sang different numbers of versions. If a
1117 male singing only one version of a phrase type sang that version >1 time, we typically used different
1118 renditions of that version as version one and two. The benefit of this was that when comparing two
1119 males who both sang only one version of a phrase type, but for whom we had saved two copies of that
1120 version, we could increase the precision of our song sharing estimate by basing it on an average of the
1121 comparison between 'version one' and 'version two', which in this case were just two renditions of the
1122 same version.

1123

1124 We limited our analyses to considering two versions of a given phrase type per male because this
1125 appeared to be the best way to minimize bias in the calculation of average similarity. First, it was
1126 extremely rare for males to sing what might be interpreted as >2 versions of a given phrase type so
1127 there would have been very few cases where including a third phrase type would have provided any
1128 additional information. Second, in the much more frequent case when males sang two versions, we
1129 could match the versions between two males and so compare like versions of a phrase when quantifying
1130 song similarity. Third, if two males did not sing the same phrase version, we could quantify how
1131 different those two versions were from each other. Further, for all pairs of males regardless of whether
1132 either male sang one version or two versions of the phrase type, the average similarity based on two
1133 comparisons (one of phrase version 1, the other of phrase version 2) always provided an unbiased
1134 estimate of the similarity between the phrases. If instead we had allowed for a third version to account
1135 for the very few males who could be considered to sing a third version of a given phrase type, this would
1136 have complicated the calculation of averages for several reasons. For a male who sang two versions, we
1137 would have had to choose one of those versions to compare with the third version of males who sang
1138 three, and this choice would have been arbitrary. When comparing a male who sang two versions of a
1139 phrase with either a one-version or a two-version male, one of the versions would have contributed to

1140 2/3 of the weight of the average because it would have been used twice. The only ways around this
1141 added substantial complexity to already complex data management and analysis methods. Because this
1142 complexity addressed a very rare issue, and a relatively unimportant one where present, we opted for
1143 the solution just described.

1144
1145 Before we averaged the two quantitative measures (spectrogram and ambiguity spectrum) of similarity
1146 for each phrase type, we wanted to weight the information from both of these quantitative measures
1147 identically, and so each measure needed to be on the same scale. Although both of these methods
1148 create values that have the potential to range between 0 and 1 (with 0 = identical and 1 = complete lack
1149 of similarity; which we later reversed such that 1 = identical, see below), the variability of the filtered
1150 ambiguity spectrum method is in practice much lower in our data set. Because we wanted the two
1151 measures of similarity to be weighted as equally as possible, we standardized the two matrices to
1152 actually vary between 0 and 1 by dividing the value in each matrix cell by the highest value for that
1153 measurement type from across all comparisons. This was approximately equal to 1 for the multi-tapered
1154 spectrogram data (and so its transformed range \approx original range), but was 0.308 for the filtered
1155 ambiguity spectrum method. After the measurements were standardized, we averaged the values
1156 across the two measures. We created this average within each phrase type, or if two versions of a
1157 phrase type occurred at a site, within each version of each phrase type. We weighted each phrase type
1158 equally when averaging across types. If there were two versions of a phrase type at a given site, each
1159 was weighted 0.5 to continue to give equal weight to each broadly defined phrase type. Finally, to
1160 reverse the direction of the song similarity score so that it ranged from 0 (no overlap) to 1 (songs
1161 identical), we subtracted the song similarity averages from 1.

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1167 Appendix S3

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1169 Table S1. To assess the robustness of our apparent survival model ranking to possible over dispersion,

1170 we examined model rankings with \hat{c} set to whole numbers between 1 and 6. Note that the top two

1171 models remained the same for $\hat{c} = 1, 2,$ and $3.$ Only when \hat{c} reached 5 did the top model no longer

1172 include an effect of habitat on apparent survival, although this habitat effect remained in the set with

1173 Delta QAIC < 2 for both $\hat{c} = 5$ and $6.$ Thus our model is robust to over dispersion.

1174

Model Structure		Model Statistics				
ϕ	p	K	Deviance	AICc	Δ AICc	w_i
$\hat{c} = 1$						
3hab + ssn	3hab + rnd	11	302.40	1291.46	0.00	0.574
3hab + ssn	3hab	9	307.10	1292.06	0.60	0.426
2hab + ssn	3hab + rnd	9	324.17	1309.13	17.67	0.000
2hab + ssn	3hab	7	328.81	1309.69	18.23	0.000
3hab + ssn	2hab	8	330.04	1312.95	21.49	0.000
ssn	3hab + rnd	7	333.85	1314.73	23.27	0.000
3hab	3hab + rnd	8	333.18	1316.09	24.63	0.000
ssn	3hab	5	341.22	1318.03	26.57	0.000
3hab	3hab	6	342.91	1321.75	30.29	0.000
3hab + ssn	con	7	342.33	1323.21	31.75	0.000
2hab + ssn	2hab	6	350.70	1329.54	38.08	0.000
2hab	3hab + rnd	7	352.26	1333.14	41.68	0.000
con	3hab + rnd	6	358.29	1337.13	45.67	0.000
2hab	3hab	5	361.82	1338.63	47.17	0.000
ssn	2hab	4	364.68	1339.47	48.01	0.000
2hab + ssn	con	5	362.77	1339.58	48.12	0.000
con	3hab	4	369.51	1344.30	52.84	0.000
3hab	2hab	5	369.18	1345.99	54.53	0.000
2hab	2hab	4	381.07	1355.86	64.40	0.000
ssn	con	3	386.85	1359.62	68.16	0.000
con	2hab	3	389.28	1362.05	70.59	0.000
3hab	con	4	393.99	1368.78	77.32	0.000
2hab	con	3	405.45	1378.22	86.76	0.000
con	con	2	426.46	1397.22	105.76	0.000
$\hat{c} = 2$						
3hab + ssn	3hab	9	153.55	655.14	0.00	0.665
3hab + ssn	3hab + rnd	11	151.20	656.89	1.75	0.277
2hab + ssn	3hab	7	164.41	661.91	6.77	0.022
2hab + ssn	3hab + rnd	9	162.08	663.67	8.53	0.009
ssn	3hab	5	170.61	664.05	8.92	0.008
ssn	3hab + rnd	7	166.93	664.43	9.29	0.006
3hab + ssn	2hab	8	165.02	664.56	9.42	0.006

3hab	3hab + rnd	8	166.59	666.13	11.00	0.003
3hab	3hab	6	171.45	666.93	11.79	0.002
3hab + ssn	con	7	171.17	668.67	13.54	0.001
2hab + ssn	2hab	6	175.35	670.82	15.68	0.000
2hab	3hab + rnd	7	176.13	673.64	18.50	0.000
ssn	2hab	4	182.34	673.76	18.62	0.000
2hab	3hab	5	180.91	674.35	19.22	0.000
con	3hab + rnd	6	179.15	674.62	19.48	0.000
2hab + ssn	con	5	181.38	674.83	19.69	0.000
con	3hab	4	184.75	676.17	21.04	0.000
3hab	2hab	5	184.59	678.03	22.89	0.000
2hab	2hab	4	190.53	681.95	26.82	0.000
ssn	con	3	193.42	682.82	27.69	0.000
con	2hab	3	194.64	684.04	28.90	0.000
3hab	con	4	196.99	688.41	33.28	0.000
2hab	con	3	202.72	692.12	36.99	0.000
con	con	2	213.23	700.62	45.48	0.000

$\hat{c} = 3$

3hab + ssn	3hab	9	102.37	442.83	0.00	0.485
3hab + ssn	3hab + rnd	11	100.80	445.36	2.53	0.137
2hab + ssn	3hab	7	109.60	445.99	3.16	0.100
ssn	3hab	5	113.74	446.06	3.23	0.096
ssn	3hab + rnd	7	111.28	447.67	4.84	0.043
3hab + ssn	2hab	8	110.01	448.43	5.60	0.029
2hab + ssn	3hab + rnd	9	108.06	448.52	5.69	0.028
3hab	3hab	6	114.30	448.65	5.82	0.026
3hab	3hab + rnd	8	111.06	449.48	6.65	0.017
3hab + ssn	con	7	114.11	450.49	7.66	0.011
2hab + ssn	2hab	6	116.90	451.25	8.42	0.007
ssn	2hab	4	121.56	451.85	9.03	0.005
2hab	3hab	5	120.61	452.93	10.10	0.003
2hab + ssn	con	5	120.92	453.24	10.41	0.003
con	3hab	4	123.17	453.46	10.64	0.002
con	3hab + rnd	6	119.43	453.78	10.95	0.002
2hab	3hab + rnd	7	117.42	453.80	10.97	0.002
3hab	2hab	5	123.06	455.38	12.55	0.001
ssn	con	3	128.95	457.23	14.40	0.000
2hab	2hab	4	127.02	457.32	14.49	0.000
con	2hab	3	129.76	458.04	15.21	0.000
3hab	con	4	131.33	461.62	18.79	0.000
2hab	con	3	135.15	463.43	20.60	0.000
con	con	2	142.15	468.42	25.59	0.000

$\hat{c} = 4$						
3hab + ssn	3hab	9	76.78	336.68	0.00	0.244
ssn	3hab	5	85.30	337.06	0.39	0.201
2hab + ssn	3hab	7	82.20	338.02	1.35	0.124
ssn	3hab + rnd	7	83.46	339.28	2.61	0.066
3hab	3hab	6	85.73	339.51	2.84	0.059
3hab + ssn	3hab + rnd	11	75.60	339.60	2.93	0.057
3hab + ssn	2hab	8	82.51	340.37	3.69	0.039
ssn	2hab	4	91.17	340.90	4.23	0.029
2hab + ssn	3hab + rnd	9	81.04	340.94	4.27	0.029
3hab	3hab + rnd	8	83.29	341.15	4.48	0.026
3hab + ssn	con	7	85.58	341.40	4.73	0.023
2hab + ssn	2hab	6	87.67	341.46	4.78	0.022
con	3hab	4	92.38	342.11	5.43	0.016
2hab	3hab	5	90.45	342.21	5.54	0.015
2hab + ssn	con	5	90.69	342.45	5.77	0.014
con	3hab + rnd	6	89.57	343.36	6.68	0.009
2hab	3hab + rnd	7	88.07	343.88	7.21	0.007
3hab	2hab	5	92.29	344.05	7.38	0.006
ssn	con	3	96.71	344.43	7.75	0.005
2hab	2hab	4	95.27	345.00	8.32	0.004
con	2hab	3	97.32	345.03	8.36	0.004
3hab	con	4	98.50	348.23	11.55	0.001
2hab	con	3	101.36	349.08	12.40	0.001
con	con	2	106.62	352.32	15.64	0.000

$\hat{c} = 5$						
ssn	3hab	5	68.24	271.66	0.00	0.231
3hab + ssn	3hab	9	61.42	272.98	1.32	0.120
2hab + ssn	3hab	7	65.76	273.24	1.58	0.105
3hab	3hab	6	68.58	274.03	2.37	0.071
ssn	3hab + rnd	7	66.77	274.25	2.59	0.063
ssn	2hab	4	72.94	274.33	2.67	0.061
con	3hab	4	73.90	275.30	3.63	0.038
3hab + ssn	2hab	8	66.01	275.53	3.86	0.034
2hab + ssn	2hab	6	70.14	275.59	3.92	0.033
2hab	3hab	5	72.36	275.78	4.12	0.029
3hab + ssn	con	7	68.47	275.95	4.28	0.027
2hab + ssn	con	5	72.55	275.97	4.31	0.027
3hab + ssn	3hab + rnd	11	60.48	276.14	4.48	0.025
3hab	3hab + rnd	8	66.64	276.16	4.49	0.024
2hab + ssn	3hab + rnd	9	64.83	276.40	4.73	0.022

ssn	con	3	77.37	276.75	5.08	0.018
con	3hab + rnd	6	71.66	277.11	5.44	0.015
con	2hab	3	77.86	277.23	5.57	0.014
3hab	2hab	5	73.84	277.26	5.59	0.014
2hab	2hab	4	76.21	277.61	5.95	0.012
2hab	3hab + rnd	7	70.45	277.93	6.27	0.010
3hab	con	4	78.80	280.19	8.53	0.003
2hab	con	3	81.09	280.47	8.80	0.003
con	con	2	85.29	282.66	10.99	0.001
<hr/>						
$\hat{c} = 6$						
<hr/>						
ssn	3hab	5	56.87	228.06	0.00	0.219
ssn	2hab	4	60.78	229.95	1.89	0.085
2hab + ssn	3hab	7	54.80	230.06	1.99	0.081
3hab	3hab	6	57.15	230.38	2.31	0.069
3hab + ssn	3hab	9	51.18	230.52	2.46	0.064
con	3hab	4	61.58	230.76	2.69	0.057
ssn	3hab + rnd	7	55.64	230.90	2.83	0.053
2hab	3hab	5	60.30	231.50	3.43	0.039
ssn	con	3	64.47	231.63	3.56	0.037
2hab + ssn	con	5	60.46	231.66	3.59	0.036
2hab + ssn	2hab	6	58.45	231.67	3.61	0.036
con	2hab	3	64.88	232.03	3.97	0.030
3hab + ssn	2hab	8	55.01	232.30	4.24	0.026
3hab + ssn	con	7	57.06	232.31	4.25	0.026
2hab	2hab	4	63.51	232.68	4.62	0.022
3hab	2hab	5	61.53	232.72	4.66	0.021
3hab	3hab + rnd	8	55.53	232.82	4.76	0.020
con	3hab + rnd	6	59.72	232.94	4.87	0.019
2hab + ssn	3hab + rnd	9	54.03	233.37	5.30	0.015
3hab + ssn	3hab + rnd	11	50.40	233.84	5.77	0.012
2hab	3hab + rnd	7	58.71	233.97	5.90	0.011
2hab	con	3	67.57	234.73	6.66	0.008
3hab	con	4	65.66	234.84	6.77	0.007
con	con	2	71.08	236.22	8.15	0.004

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1180 Appendix S4
 1181

1182 Table S1. Model-averaged parameter estimates associated with all factor levels for variables included in
 1183 candidate models in mark-recapture analyses of male dickcissels.

factor and level	weighted average phi (daily)	SE
breeding season (3 hab models)		
Cropland	0.9843852	0.0036881
Prairie (T. Parker)	0.9944556	0.0013555
Prairie (B. Sousa)	0.9938236	0.0016063
non-breeding season (3 hab models)		
Cropland	0.9978365	0.0008221
Prairie (T. Parker)	0.9989836	0.0002436
Prairie (B. Sousa)	0.9966893	0.0005106
breeding season (2 hab models)		
Cropland breeding season	0.9837991	0.0036152
Prairie breeding season	0.9949143	0.0009709
non-breeding season (2 hab models)		
Cropland non-breeding season	0.9978667	0.0008041
Prairie non-breeding season	0.9981497	0.0002363
p	weighted average p	SE
(3 hab models)		
Cropland, first occasion of the year	0.7132702	0.0648546
Cropland, subsequent occasions of the year	0.5928664	0.0636747
Prairie (T. Parker), first occasions	0.7706777	0.0380641
Prairie (T. Parker), subsequent occasions	0.7577173	0.0276309
Prairie (B. Sousa), first occasions	0.9269718	0.0209690
Prairie (B. Sousa), subsequent occasions	0.9269718	0.0209690
(2 hab models)		
Cropland, all occasions	0.6447336	0.0546136

Prairie, all occasions

0.8265300

0.0172619

1184

1185

1186 Appendix S5

1187

1188 Exploratory analysis conducted with cross-correlation

1189

1190 We conducted a preliminary study to evaluate the suitability of the cross-correlation in Raven for
1191 quantifying song similarity in dickcissels. In this preliminary analysis, conducted in 2010, we used songs
1192 of color-banded territorial male Dickcissels recorded in the summers of 2006-2010 at our sites described
1193 in the main text of this paper. The within-bird portion of our analysis included 24 dickcissels present in
1194 multiple years of the study from four sites: two sites on the Konza Prairie Biological Station (KPBS) and
1195 two agricultural sites managed by Kansas State University (KSU). This portion included 54 song
1196 comparisons. The between-bird portion of our analysis included 24 birds from the same sites and 2851
1197 song comparisons.

1198

1199 *Evaluation Methods*

1200 We used RavenPro 1.3 (Cornell Lab of Ornithology 2008) to convert high-quality song recordings into
1201 visual spectrograms for analysis. We used a bandpass filter to eliminate frequencies above 20,000 Hz
1202 and below 2,000 Hz. We removed isolated insect noises and the songs of other birds using the select
1203 function where they did not interfere with the target bird's song.

1204

1205 Before evaluating song similarity, we focused on cissel phrases. For every date we visited a site, we
1206 chose the cleanest recording of each cissel phrase of each birds' song to include in cross-correlation
1207 analysis. Even if other phrases were present in the song, we only looked at dick and cissel phrases as
1208 they appeared in some variety in every bird's song.

1209

1210 Using Raven, we performed batch cross-correlations of each phrase within each site and between the
1211 sites producing a large matrix of cross-correlation values comparing all cissel phrases to all cissel
1212 phrases.

1213

1214 *Preliminary Findings: Within-bird v. Between Bird*

1215 The cross-correlation results comparing all birds present for multiple years showed a trend for higher
1216 cross-correlation scores within individual birds than between birds; but ultimately within-bird cross-
1217 correlation scores should have been consistently higher than between-bird cross-correlation scores.
1218 Both groups had substantial unexplained variability in cross-correlation score distribution, with the same
1219 range of scores for within-bird comparisons as in between-bird comparisons. Within-individual
1220 comparisons produced scores as low as 0.1 and many scores < 0.6 (Figures S1, S2).

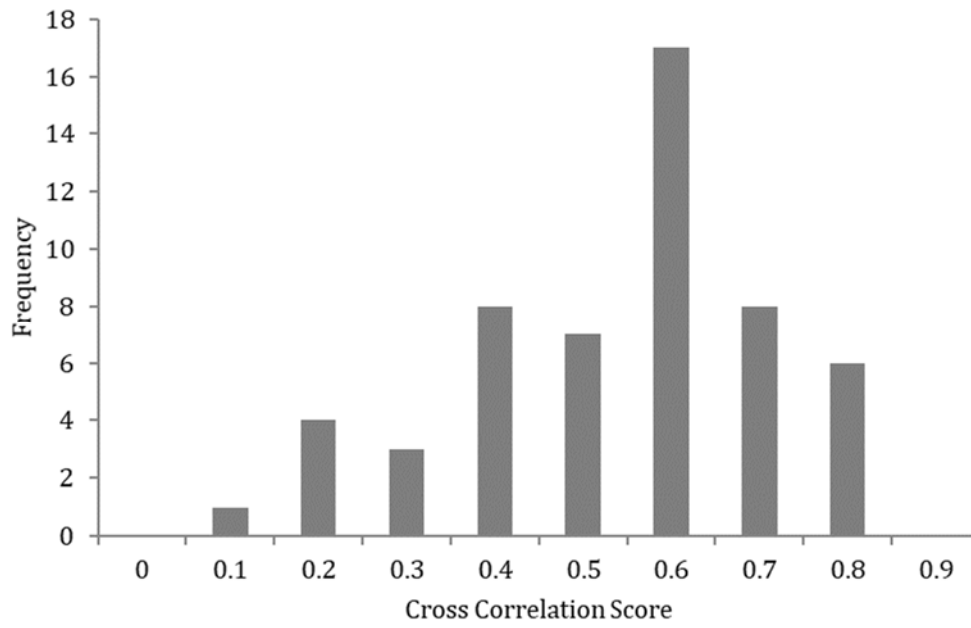
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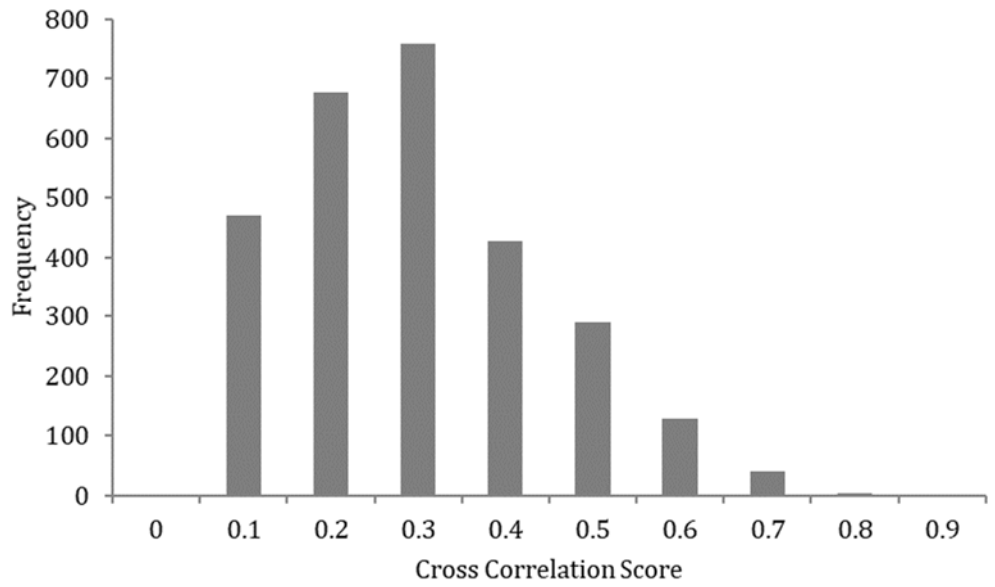
1225 Figure S1



1226 Figure S1. Cross correlation scores for comparisons between two songs produced by the same male
1227 dickcissel on different dates.
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1232 Figure S2
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1234 Figure S2. Cross correlation scores for comparisons between pairs of songs sung by different male
1235 dickcissel.
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1242 Appendix S6

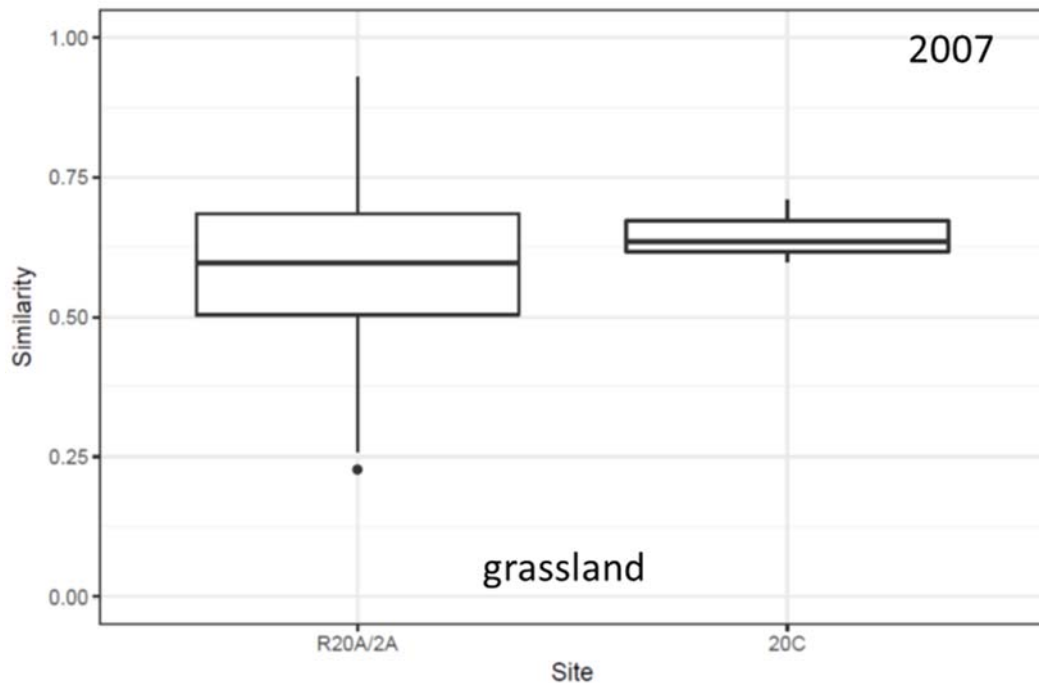
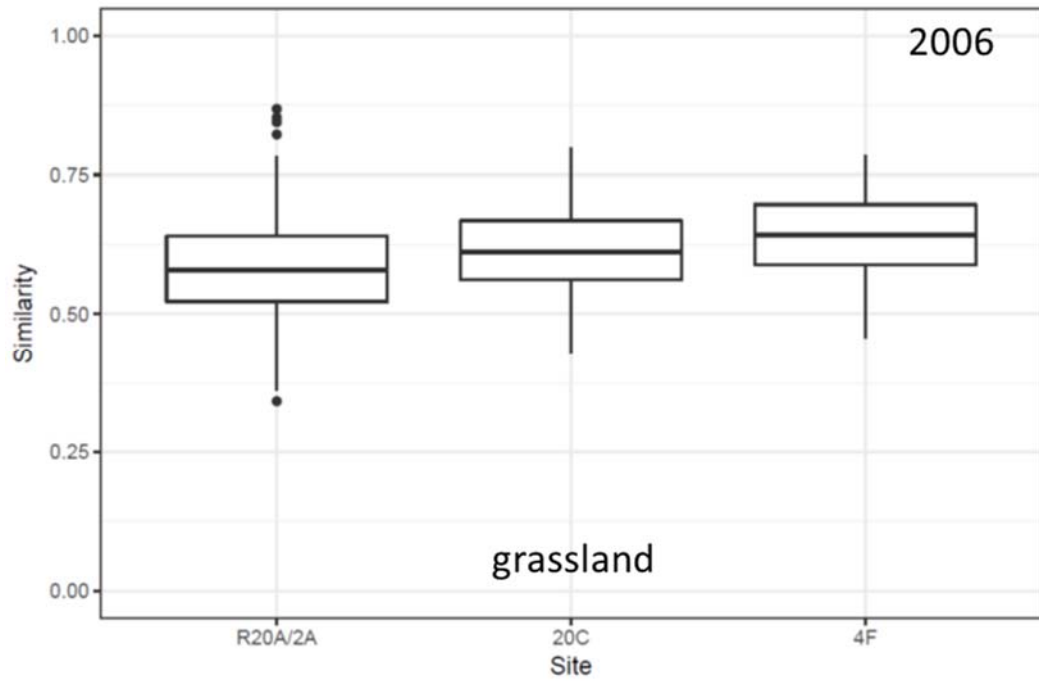
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1244 Figure S1. Plots of median song sharing values across the grassland sites (R20A/2A, 20C, 4F, C1B) and

1245 cropland sites (North, Ashland, Strauss) in different years. The central line is the median and the box

1246 represents the interquartile range (middle 50th percentile).

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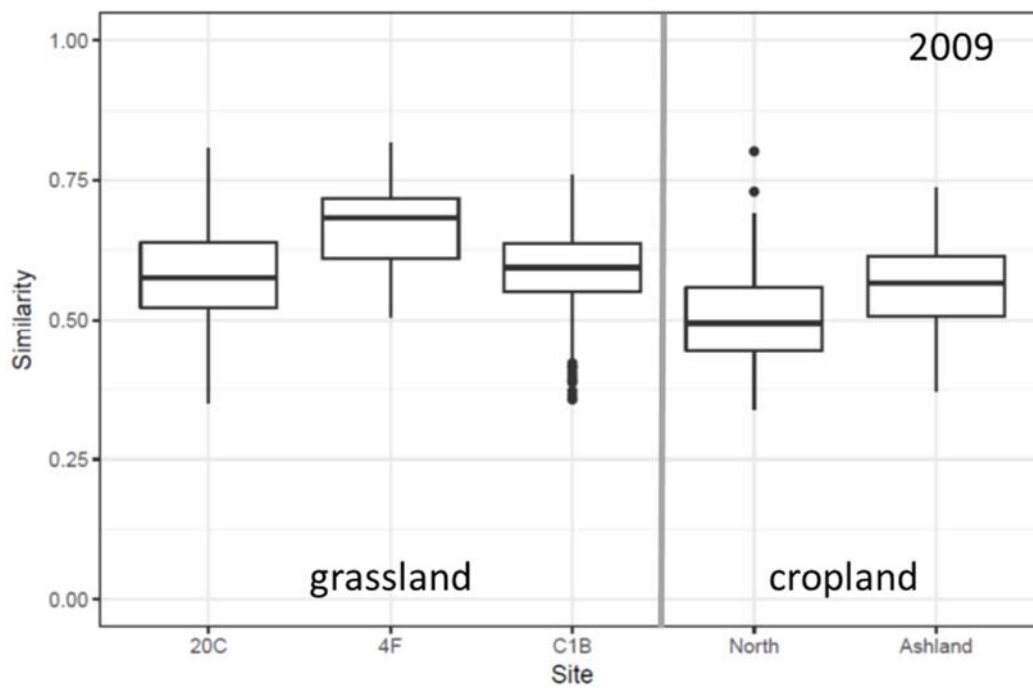
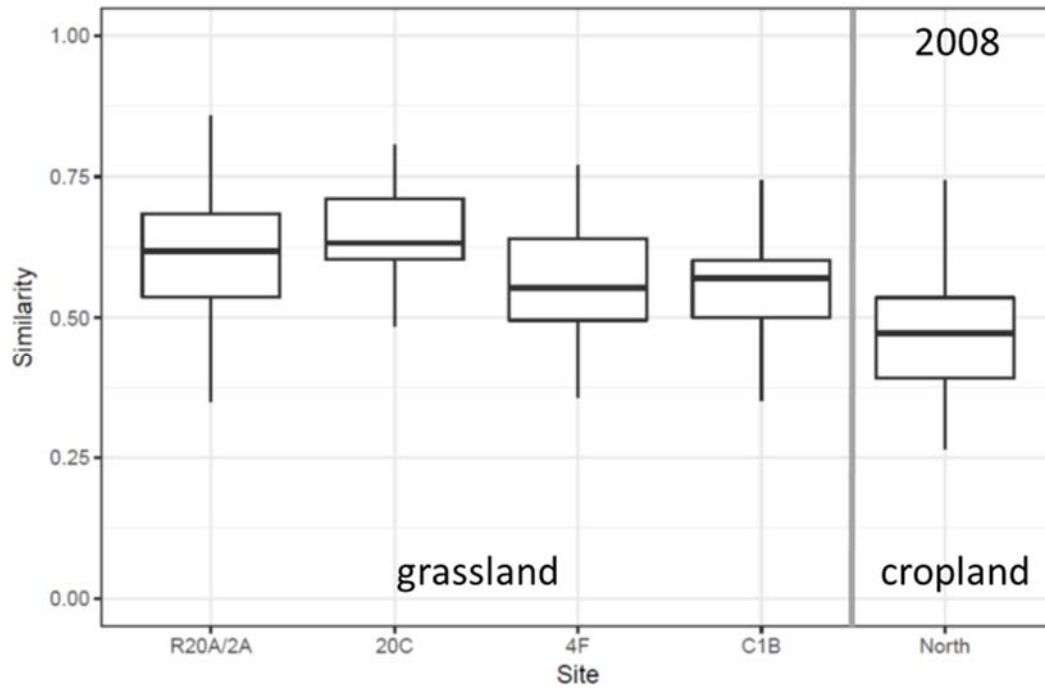
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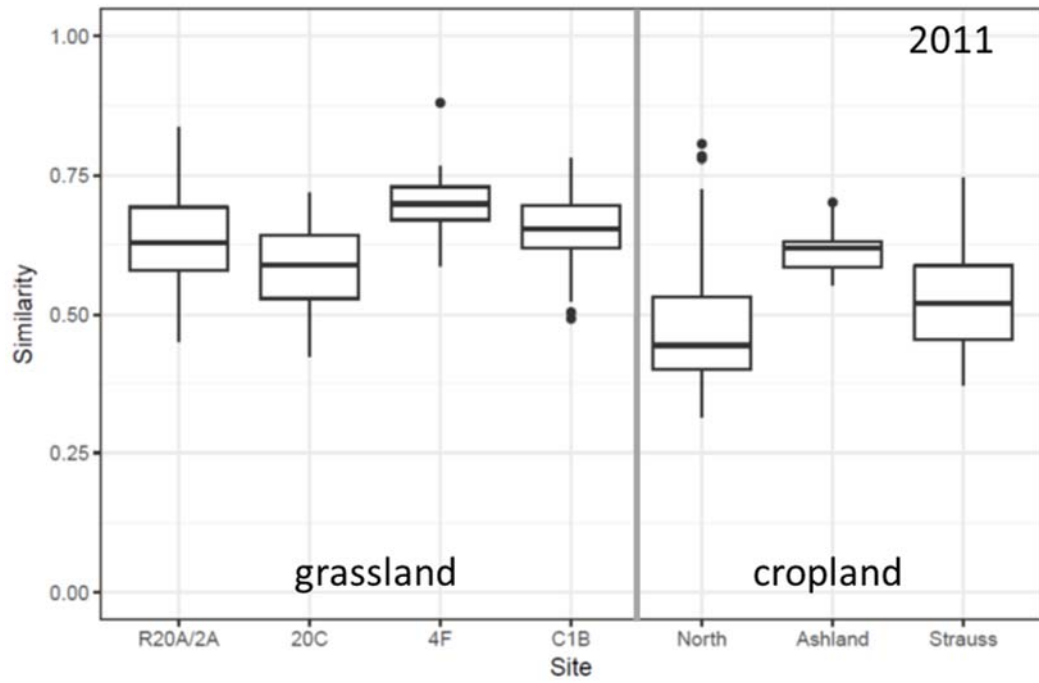
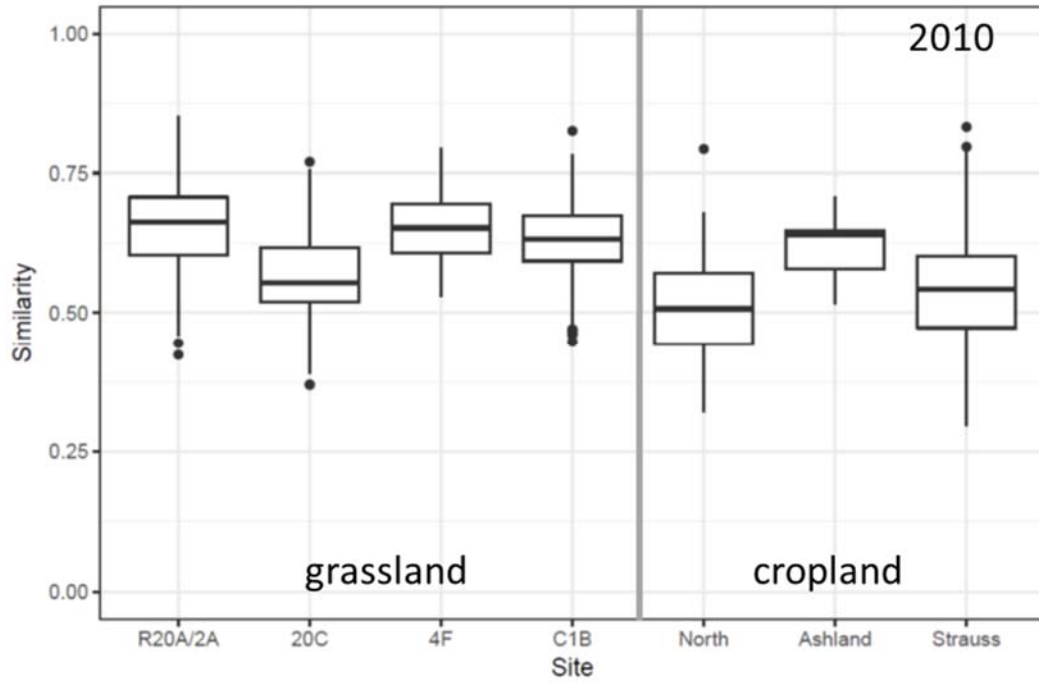
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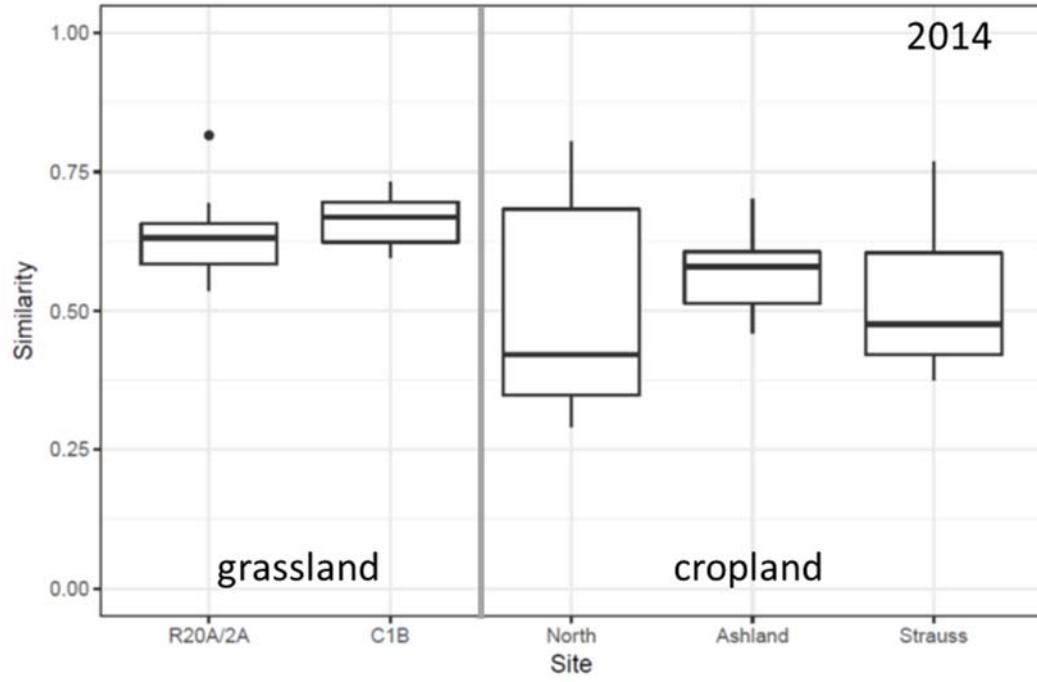


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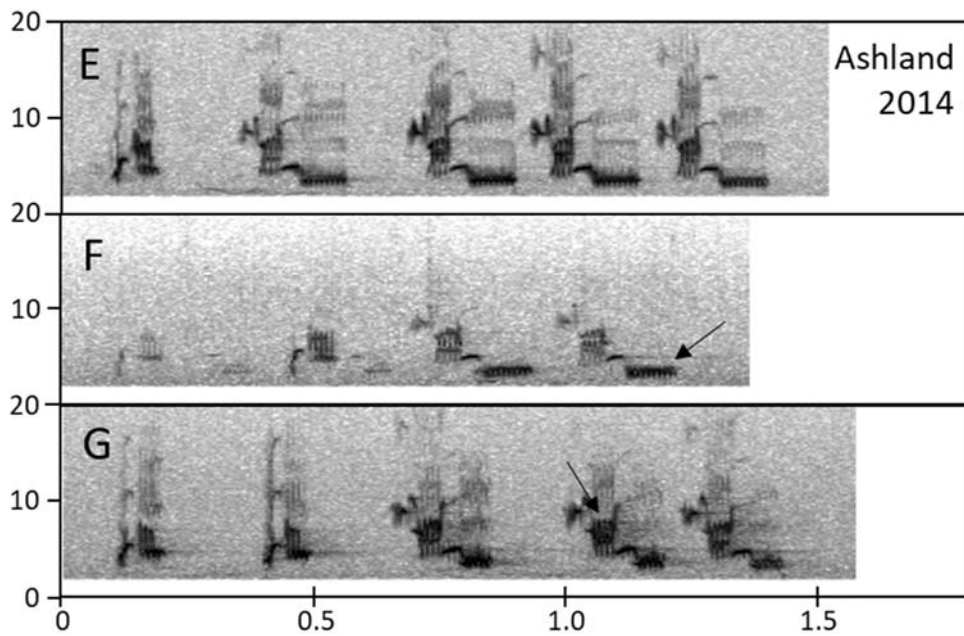
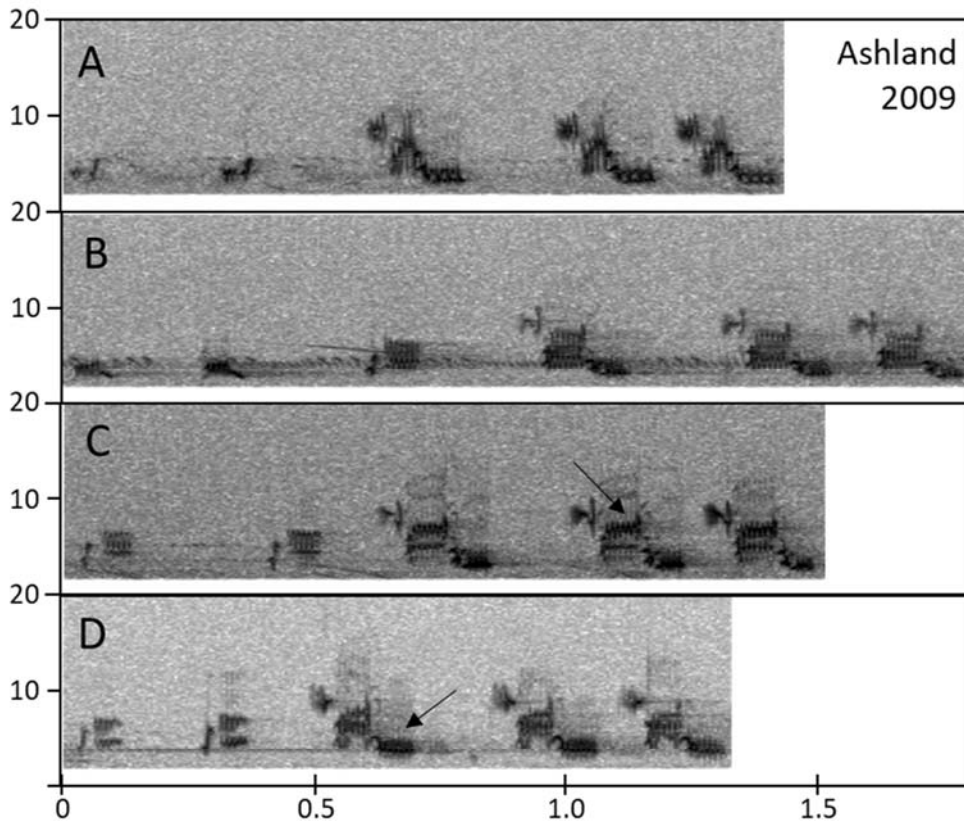
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1269 Appendix S7

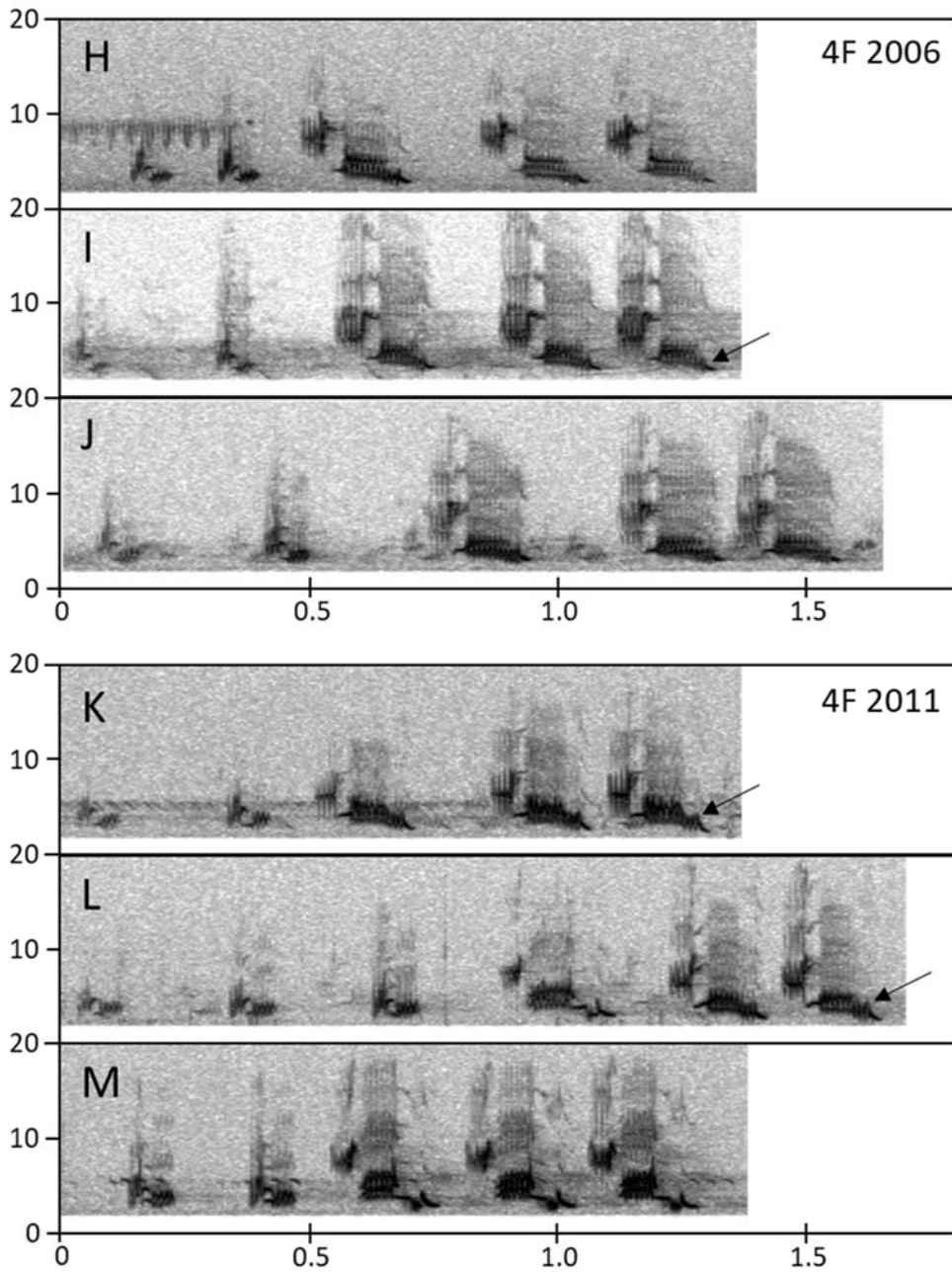
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1271 Figure S1. Changes in vocal culture of dickcissels over five years at two different sites (Ashland, a
1272 cropland site, 2009 and 2014; 4F a grassland site, 2006 and 2011). In most Ashland songs in 2014, the
1273 central portion of cissel phrase (arrow in panel G) was of shorter duration than the comparable portion
1274 in 2009 (arrow in panel C). Also, the low frequency terminal portion of the cissel phrase tended to be
1275 much longer and to have a narrower frequency range in 2014 (arrow in panel F) than in 2009 (arrow in
1276 panel D). Some of the evident differences in 4F songs are more subtle. For instance, the final note of the
1277 common cissel phrase in 2006 (arrow in panel I) was of slightly longer duration and greater frequency
1278 range in 2011 (arrow in panel K). However, a distinctly different cissel phrase can also be seen in some
1279 2011 songs. Some males in 2011 only produced this different type of cissel (panel M), but some other
1280 males produced both types of cissels (panel L).

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