

1 **Running head:** Crowdsourcing to unify disturbance ecology  
2 **Towards a generalizable framework of disturbance ecology through crowdsourced science**  
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46  
47

48 **Abstract.** Disturbances fundamentally alter ecosystem functions, yet predicting their impacts  
49 remains a key scientific challenge. While the study of disturbances is ubiquitous across many  
50 ecological disciplines, there is no agreed-upon, cross-disciplinary foundation for discussing or  
51 quantifying the complexity of disturbances, and no consistent terminology or methodologies  
52 exist. This inconsistency presents an increasingly urgent challenge due to accelerating global  
53 change and the threat of interacting disturbances that can destabilize ecosystem responses. By  
54 harvesting the expertise of an interdisciplinary cohort of contributors spanning 42 institutions  
55 across 15 countries, we identified an essential limitation in disturbance ecology: the word  
56 ‘disturbance’ is used interchangeably to refer to both the events that cause, and the consequences  
57 of, ecological change, despite fundamental distinctions between the two meanings. In response,  
58 we developed a generalized framework of ecosystem disturbances to reconcile this limitation,  
59 providing a well-defined lexicon for understanding disturbance across perspectives and scales.  
60 The framework results from ideas that resonate across multiple scientific disciplines and  
61 provides a baseline standard to compare disturbances across fields. This framework can be  
62 supplemented by discipline-specific variables to provide maximum benefit to both inter- and  
63 intra-disciplinary research. To support future synthesis or meta-analysis of disturbance research,  
64 we also encourage researchers to be explicit in how they define disturbance drivers and impacts,  
65 recommend minimum reporting standards that studies should detail about the magnitude,  
66 duration, and rate of change of driver and response variables of a disturbance, regardless of scale.  
67 We discuss the primary factors we considered when developing a baseline framework and  
68 propose four future directions to advance our interdisciplinary understanding of disturbances and  
69 their social-ecological impacts: integrating across ecological scales, understanding disturbance  
70 interactions, establishing baselines and trajectories, and developing process-based models and  
71 ecological forecasting initiatives. Our experience through this process motivates us to encourage

- 72 the wider scientific community to continue to explore new approaches for leveraging Open
- 73 Science principles in generating creative and multidisciplinary ideas.

74 **Introduction.**

75 Disturbances related to human activities, including both abrupt and long-term impacts of climate  
76 change, are predicted to continually intensify in the coming century (IPCC, 2019). For instance,  
77 wildfires have ravaged global landscapes over the last two decades, impacting human lives,  
78 crops, and biodiversity — highlighted by recent outbreaks in Australia, Brazil, California, and  
79 British Columbia (Cleatus and Mulik, 2014;Boer et al., 2020;Tedim et al., 2020). Twenty of the  
80 hottest years in history have occurred in the past 22 years (Organization, 2018), and extreme  
81 events like marine heat waves are projected to increase in frequency by more than an order of  
82 magnitude as climate change continues (IPCC, 2019). As well, long-term changes in temperature  
83 and moisture can lead to changes in ecosystem structure (e.g., species composition) and function  
84 (e.g., biogeochemical cycles). Such disturbances can radically alter trajectories of ecosystem  
85 dynamics, and importantly, they occur within a broader ecological context that can generate  
86 interactions among ecosystem processes and lead to unpredictable ecosystem responses (Paine et  
87 al., 1998;Calderón et al., 2018;Zscheischler et al., 2018;Knelman et al., 2019).

88 Despite increases in the frequency (e.g., return interval), duration (e.g., pulse vs. press events),  
89 and scale (e.g., severity, intensity, magnitude, extent, etc.) of disturbance events, predicting their  
90 onset, characteristics, and consequences remains difficult (Battisti et al., 2016). This is in part  
91 because of differences in conceptual models, scales of investigation, and language used across  
92 scientific disciplines (Salafsky et al., 2008;Battisti et al., 2016). Disturbances occur through  
93 space and time with different frequencies (number of occurrences per unit time), intensities  
94 (magnitude of the disturbance), and extents (spatiotemporal domain affected) (Sousa,  
95 1979;1984;Grimm and Wissel, 1997;Paine et al., 1998;Miller et al., 2011). Additionally, natural  
96 versus anthropogenic disturbances differ in their underlying causes and socio-ecological

97 implications, yet are commonly discussed with the same terminology (Salafsky et al., 2008).  
98 Inconsistencies in disturbance frameworks have long been noted by the ecological community  
99 (Rykiel Jr., 1985;Pickett et al., 1989;Poff, 1992;Peters et al., 2011;Gaiser et al., 2020), and the  
100 struggle to derive a common framework for understanding and predicting disturbances continues  
101 in modern literature (Smith, 2011b;Borics et al., 2013;Hobday et al., 2016;Jentsch and White,  
102 2019). Disturbances are often inferred to be synonymous with pulse events, perturbations,  
103 threats, and/or stressors, and thus, the concept of disturbance encapsulates phenomenon across a  
104 range of spatial and temporal conditions. However, these terms should not be used  
105 interchangeably, and have subtle and meaningful differences in specific fields of inquiry (Rykiel  
106 Jr., 1985;Lake, 2000;Borics et al., 2013;Jentsch and White, 2019;Keeley and Pausas,  
107 2019;Kempainen et al., 2019). For instance, Slette et al. (2019) argued that the plethora of  
108 literature on drought is generally based on loose descriptions rather than explicit definitions or  
109 quantitative metrics of drought, while Hobday et al. (2018) noted that other disturbances lack  
110 even basic quantitative categorization or naming schemes. Because of these inconsistencies,  
111 attempts to compare disturbances across types and ecosystems have resulted in few outcomes  
112 that can be generalized across fields (Peters et al., 2011). Definitions of disturbance originally  
113 focused on ‘discrete events’ that alter an ecosystem or its function (Pickett and White, 2013), but  
114 recent definition frameworks have incorporated aspects of drivers of disturbance, disturbance  
115 regimes, and scale of disturbances in time and space (Turner, 2010;Peters et al., 2011;Gaiser et  
116 al., 2020). Collectively, these shortcomings point to the need for an interdisciplinary  
117 understanding of disturbances.  
118  
119 Differences in how disturbances are studied are driven in part by their spatial and temporal  
120 heterogeneity and in part by differences in typical scales of investigation across scientific

121 disciplines. While some disturbance events have relatively discrete temporal and spatial  
122 boundaries (e.g., wildfires, hurricanes, earthquakes), others are diffuse or overlap in time and  
123 space (e.g., ocean acidification, overgrazing, nutrient loadings, droughts) (Godfrey and Peterson,  
124 2017). This makes it difficult to identify which events depart from ‘normal’ ecosystem  
125 processes, especially within the broader context of ongoing environmental change (Duncan et al.,  
126 2010;Mishra and Singh, 2010;Slette et al., 2019). Finally, because the impacts of disturbances  
127 are contingent on historical events and local socio-economic conditions (Duncan et al.,  
128 2010;Seidl et al., 2016;Dietze et al., 2018;Słowiński et al., 2019), a single type of disturbance  
129 can be perceived in different ways depending on the environment and species of interest.  
130 Dynamic hydrology, for example, is fundamental to floodplain wetland systems, which are  
131 adapted and shaped by flooding events, but flooding events are typically considered disturbances  
132 in upland contexts.

133

134 To facilitate interdisciplinary investigation and understanding of disturbances, we need a  
135 generalizable framework with which to talk about such events. Ideally, this framework would be  
136 able to manage the heterogeneity inherent in disturbances while providing consistency in how  
137 disturbances are defined and studied. Such a foundation would consist of shared goals and be  
138 built upon commonly agreed-upon terms and metrics. We propose a generalizable disturbance  
139 framework that builds upon earlier frameworks (Grimm and Wissel, 1997;Peters et al.,  
140 2011;Newman et al., 2020) by emphasizing drivers (also called ‘driving forces’ or ‘indirect  
141 threats’) vs. impacts of disturbance, acknowledging multiple system responses to disturbance,  
142 and enabling cross-ecosystem comparisons. We expand on existing frameworks by recognizing  
143 multiple scales of interactions over space and time and acknowledging disturbance legacies that  
144 may alter the vulnerability of an ecosystem (e.g., risk of organismal, elemental, or other losses)

145 to other drivers. The framework presented here provides a baseline of commonalities for  
146 interdisciplinary collaborations and communication; and it can be supplemented with discipline-  
147 specific variables for more in-depth investigation into particular aspects of disturbances.

148  
149 To address this challenge, we used an open call on social media to assemble a cross-disciplinary  
150 team of 50 collaborators at different career stages across 42 institutions in 15 countries with a  
151 diverse suite of scientific specialties (Graham and Krause, 2020;Graham and Smith, 2020). We  
152 used our collective expertise to propose a pathway towards a new conceptual model of ecological  
153 disturbances that integrates contributions across disparate disciplines. The range of disciplines and  
154 scale of research represented by contributors include microbial or plant ecology at the gene,  
155 population, community, and ecosystem level; biogeochemistry across freshwater, marine, and  
156 terrestrial ecosystems; ecology focused at soil pore scale all the way to organisms at the landscape  
157 and watershed scales; environmental social scientists; and conservation biologists. The project  
158 featured a flexible, collaborative, and iterative writing process (using Google docs), freely open  
159 authorship opportunities advertised via Twitter that recruited many early career scientists including  
160 graduate students and postdoctoral researchers. It was coordinated by a small international  
161 leadership team and broke down barriers between researchers at various career stages, institutions,  
162 and disciplines. By proposing a generalizable framework for disturbances, we strive towards a  
163 common currency to compare ecological drivers and responses across conditions and systems.

164  
165 We start by describing a generalizable framework derived from crowdsourced scientific  
166 knowledge, followed by an overview of factors that resonated across disciplines such as system  
167 stability theory and spatiotemporal considerations. The final sections propose minimum  
168 reporting standards for widely implementing this common framework and cross-disciplinary



169 approaches for addressing areas of need. This emergent framework is intended to help synthesize  
170 ideas among historically disparate events and disciplines, more rigorous tracking of events across  
171 space and time, and new ways of understanding disturbance impacts between fields. In turn,  
172 resulting knowledge can influence the ways in which humans manage ecosystems and their  
173 responses to disturbances by aiding managers in identifying slow-developing disturbances as  
174 they occur, referencing disturbances against historical events by comparing quantitative  
175 characteristics, and being able to better predict ecological impacts to define conservation  
176 strategies.

177

178 **A generalizable framework.**

179 One essential limitation in our understanding and managing of disturbances is that the word  
180 ‘disturbance’ is used interchangeably to describe two distinct processes—events that cause  
181 ecological change and consequences of extreme events—that are both termed disturbances  
182 despite fundamental distinctions between the two types of processes. Some researchers define  
183 disturbances by properties that describe an event (e.g., type, duration, frequency, intensity)  
184 (Hobday et al., 2016;Hobday et al., 2018), while others characterize disturbances by their  
185 impacts (e.g., ecological, or societal damages) (Smith, 2011a). Others try to integrate disturbance  
186 drivers and impacts by describing disturbances as a chain of events. For instance, the Driver-  
187 Pressure-State-Impact-Response (DPSIR), provides a structure in which a series of causal links  
188 from ‘driving forces’ (economic sectors, human activities) through ‘pressures’ (emissions, waste)  
189 to ‘states’ (physical, chemical and biological) and ‘impacts’ on ecosystems, human health and  
190 functions, leading to political ‘responses’ (prioritization, target setting, indicators) (Pirrone et al.,  
191 2005). Furthermore, many definitions of disturbances solely consider short-term events that  
192 represent rapid deviations from a biotic or abiotic background state without regard to historical

193 processes (Jentsch and White, 2019). Finally, solely defining disturbances by their impact size  
194 directly conflicts with the idea of ecological resistance and the vast amount of theory developed  
195 for this phenomenon. If we were to define a disturbance based only on its impact, highly resistant  
196 ecosystems would never be disturbed regardless of the prevalence of extreme events.

197 Disturbance theory lacks a one-size-fits-all approach due to the spatial, temporal, and cross-  
198 disciplinary complexities in studying disturbances. A key challenge in the development of such  
199 an approach is that individual disturbances operate within a broader context of historical events  
200 that cumulatively alter disturbance magnitude and impact. For instance, Ryo et al. (2019)  
201 describe the temporal dependency of interacting disturbances in terms of ‘nestedness’, wherein  
202 the complexity of interactions is dependent on the relative closeness of the events in question.  
203 Within this framework, a single event is a subset of multiple disturbances within a continuous  
204 trajectory. Importantly, there are carryover effects within trajectories in which disturbance  
205 impacts can accumulate and/or alter the internal mechanisms affecting responses through time,  
206 even for parts of an ecosystem not affected by earlier disturbances (Nowicki et al., 2019).  
207 Therefore, driver-response relations are dependent on both short- and long-term histories. While  
208 Ryo’s framework only considers temporal aspects of disturbances (Ryo et al., 2019), it highlights  
209 the need for a fluid framework to provide a common foundation for studying disturbances across  
210 scales and lines of inquiry—one that can adjust for variation between systems and research  
211 goals.

212

213 We propose a robust and tangible framework of disturbance that is applicable regardless of the  
214 line of inquiry and/or spatiotemporal scale of investigation (Figure 1). Specifically, we define a  
215 disturbance event as the occurrence of a driver whereby a force, either biotic or abiotic, generates

216 a deviation from the local, prevailing background conditions (i.e., a *disturbance driver*). In the  
217 proposed framework, a driver is characterized by its magnitude of deviation from an  
218 environmental baseline (low to high deviation describes weak to strong drivers). In contrast, a  
219 *disturbance impact* represents the social-ecological consequences of a driver relative to a scale-  
220 dependent baseline state. A key attribute of the framework is that drivers and impacts are both  
221 relative to a baseline state. Baselines are determined based on abiotic conditions that are relevant  
222 to the particular disturbance in question (e.g., moisture content and evapotranspiration, nutrient  
223 concentrations) as well as biotic factors such as population size, species composition, and life  
224 history dynamics. Using relationships between disturbance drivers and disturbance impacts, we  
225 generate *four universal types of disturbances* with variation within each type due to the strength  
226 of the driver and the size of the impact. Conceptualizing disturbance drivers, either abiotic or  
227 biotic, on an x-axis and the impact of disturbance impacts on a y-axis yields four quadrants:  
228 weak & positive, strong & positive, weak & negative, and strong & negative.

229

230 The position of drivers and impacts across and within the quadrants varies with the line of  
231 inquiry (Figure 1). For example, a 10-day drought is a severe disturbance for a drought-sensitive  
232 microorganism, but probably inconsequential for humans in urban environments (Figure 1).  
233 Additionally, the disturbance impact for a single driver could be simultaneously positive and  
234 negative, dependent on scale. For example, deforestation for agriculture could be positive from  
235 an immediate human perspective (food production) but negative from an ecological perspective  
236 (habitat loss). This allows for interacting and compounding disturbances to be viewed within the  
237 same framework as single events and for events that cause tipping points to be represented as  
238 weak driver-high impact events.

239

240 Spatial and temporal scales are also implicitly represented in the proposed framework, as  
241 historical exposures have direct effects on the impact of a given driver and the scale of interest  
242 defines the magnitude of both the driver and impact. Likewise, the ecological state of a system  
243 (e.g., its stability, resistance, resilience, and successional stage) also influences the ‘impact’ axis  
244 of disturbances through escalating or mediating the impact. Further explanation, definitions and  
245 examples of each quadrant are presented in more detail in Table 1 and Figure 1.

246

247 The advantage of conceptualizing and classifying disturbances into this inclusive framework is to  
248 increase interoperability of disturbance research across scientific domains. While the current  
249 framework is qualitative in nature and based upon discipline-specific expert knowledge of driver  
250 and impact magnitudes, there are further opportunities to develop quantitative thresholds to  
251 separate quadrants for particular lines of investigation. For example, a disturbance driver and  
252 impact size fall within a range of historical variation that is specific to the event type.

253 Quantitative thresholds for event types can then be developed to separate events along driver and  
254 impact axes based on the distribution of historical events along those axes.

255

### 256 **Common Factors to Consider in a Generalizable Framework.**

257 Below, we describe a subset of factors related to disturbance that most strongly resonated across  
258 researchers from different disciplines when putting together our generalizable framework.

259 Investigations of disturbances are vast and multifaceted, and we do not intend the sections below  
260 to be comprehensive reviews of the subjects mentioned. Rather, we present concepts that are  
261 most transferable and therefore able to underpin a common understanding. We point the reader  
262 to references within this section for more thorough reviews of the topics discussed.

263 *System Stability*

264 System stability has been a pillar of disturbance research across scientific domains as it can be  
265 used to describe a system's response to environmental change (Ives and Carpenter, 2007; Duncan  
266 et al., 2010; Hodgson et al., 2015; Seidl et al., 2016; Todman et al., 2016). Here, we review core  
267 and emerging system stability theory to form a conceptual basis for an interdisciplinary approach  
268 to understanding disturbance discussed in later sections. There are several aspects of system  
269 stability theory that are common throughout social-ecological domains, including the concepts of  
270 resilience, resistance, and redundancy. While these concepts often underlie hypothesis-testing in  
271 disturbance ecology, the exact nature of their relationships to disturbance impacts and recovery  
272 trajectories remains unknown. Resilience is commonly defined as the ability of a system to  
273 recover from disturbance, while resistance is the ability of an ecosystem to remain unchanged  
274 when being subjected to disturbance (Holling, 1973; Westman, 1978; Holling, 1996; McCann,  
275 2000; Gunderson et al., 2002; Griffiths and Philippot, 2013; Seidl et al., 2016; Lamentowicz et al.,  
276 2019). Resistance and resilience are functions of biodiversity and species traits. They are  
277 quantified using various metrics, including the time, slope/rate, and angle of recovery relative to  
278 a baseline state (Connell and Sousa, 1983; Shade et al., 2012). Additionally, functional  
279 redundancy and similarity are also used in ecology to describe the capacity of a system to resist  
280 and recover from a disturbance, whereby the presence of functionally redundant phenotypes  
281 enhances ecosystem stability (Naeem and Li, 1997; Allison and Martiny, 2008). Similarly,  
282 response diversity, which relies on differential responses among species to environmental  
283 fluctuations and disturbance, such that in fluctuating environments different species are favored,  
284 is another factor in determining ecosystem stability (Yachi and Loreau, 1999; Elmqvist et al.,  
285 2003). Functional redundancy, functional similarity, and response diversity are framed within the

286 'insurance hypothesis,' wherein higher biodiversity increases the likelihood that the community  
287 contains species with differential species functions or responses to the environment, providing  
288 "insurance" for aggregate properties of the community (Allison and Martiny, 2008;Mori et al.,  
289 2013). Indeed, biodiversity is an integral component of system stability theory and has been  
290 repeatedly linked to the capacity of ecosystems to be resistant to a diversity of disturbances, such  
291 as biological invasions, and climate fluctuations (Cardinale and Palmer, 2002;Isbell et al.,  
292 2009;Cardinale et al., 2012;Kardol et al., 2018).

293  
294 Another central paradigm in stability theory is that the intensity of the response to disturbances is  
295 often non-linearly related to the intensity of the disturbance itself. There is a growing  
296 understanding of the importance of tipping points that, when reached or exceeded, cause strongly  
297 non-linear system responses and potential sudden shifts in system behavior (Dai et al.,  
298 2012;Loecke et al., 2017). For instance, Scheffer et al. (2001) showed that ecosystems can  
299 deviate rapidly from their current state due to minor shifts in underlying biotic or abiotic drivers.  
300 Similarly, slow, and often undetectable changes can reduce ecosystem resilience, leading to  
301 unpredictable system collapses (Walker et al., 2012). When pressures exceed ecosystem tipping  
302 points, regime shifts can occur, and ecosystems are pushed into a different (alternative) state that  
303 is maintained by self-reinforcing feedbacks (Pausas and Bond, 2020). Identifying which  
304 disturbance regimes are susceptible to regime shifts that result in a switch between stable states  
305 or to a new alternative state remains a key obstacle. While there is growing capacity to predict  
306 regime shifts (e.g., by rising variance in ecosystem properties or by slow recovery rates), several  
307 difficulties remain in their prediction, in part due to the challenge of measuring appropriate  
308 indicators for resilience (Van Nes and Scheffer, 2007;Scheffer et al., 2009;Scheffer, 2010;Dai et  
309 al., 2012;Dai et al., 2013). While system stability is also criticized as ambiguous or difficult to

310 apply in practice (Grimm and Wissel, 1997), historical and emerging research on system stability  
311 pervades many disciplines and provides a common foundation for understanding disturbances  
312 across a broad suite of ecosystems and across lines of investigation with different underlying  
313 objectives.

314

### 315 *Spatiotemporal Considerations*

316 Spatial extent and temporal duration are integral components of disturbances, and quantifying  
317 these characteristics for individual disturbances is key to understanding the ecological impacts of  
318 those disturbances. Though the specific nature of disturbance extents and duration can vary  
319 greatly, all disturbances occur over space and time; any generalizable framework must, therefore,  
320 consider the spatiotemporal extent and variability of disturbances. This includes defining the  
321 baseline conditions relative to which a disturbance is assessed to build a set of domain-agnostic  
322 principles. These baselines may vary as a function of the spatiotemporal scale over which an  
323 analysis is being performed, and the deviation of a system from its baseline at a given scale can  
324 be used to assess a disturbance's intensity and impact (e Silva et al., 2013). As ecosystems  
325 change in response to climate, land-use change, and other human activities, conditions that were  
326 once considered disturbed against a static baseline may now shift into a new normal range of  
327 variation (Figure 2). In this section, we review key spatial and temporal perspectives that  
328 influence disturbances and that should underlie an interdisciplinary understanding of disturbance.

329

330 The drivers and impacts of disturbance are dependent on the spatial features of their broader  
331 landscapes and the spatial perspective of a given study's objectives. For example, pre-existing  
332 ecosystem characteristics, such as habitat connectivity and topography, influence the spatial  
333 structure of disturbance impacts by dictating its ability to spread as well as the ecosystem's

334 ability to be recolonized by surviving organisms in neighboring spaces (Turner et al.,  
335 1994;Drever et al., 2006;Buma, 2015). Further, the spatial perspective taken when studying a  
336 disturbance can also heavily influence conclusions drawn about its effects. Some disturbances  
337 (e.g., fine-scale temperature shifts) may be apparent only at local scales while others influence  
338 regional and coarser scales (Aalto et al., 2017;Lembrechts et al., 2019). In general, disturbances  
339 that directly affect species interactions tend to be observable at local scales (Mod et al., 2014),  
340 while disturbances related to habitat alterations are detectable at coarser spatial resolution  
341 (Hamer and Hill, 2000;Dumbrell et al., 2008;Chase, 2014).

342

343 Both the impact of a disturbance as well as ecosystem responses to one ultimately depends on  
344 how the disturbances modifies scale-specific factors that control ecosystem stability (Dobson et  
345 al., 1997;Dumbrell et al., 2008;Wei and Zhang, 2010;Cohen et al., 2016). For example, extant  
346 dispersal rates and disturbance scale can regulate the recovery of disturbed ecological  
347 communities and the spread of impacts across space to neighboring populations (Zelnik et al.,  
348 2019). Furthermore, the spatial extent and patterning of disturbances can influence disturbance  
349 impacts.

350

351 Additionally, impacts at multiple spatial scales can interact. Local disturbances can play an  
352 important role in maintaining regional biodiversity through patch dynamics mediated by intra-  
353 and inter- species dynamics such as competition and colonization trade-offs (Tilman,  
354 1994;Grime, 2006;le Roux et al., 2013;He et al., 2019), which may reduce vulnerability to larger  
355 scale disturbances. Local disturbances can also exacerbate the impacts of more widespread  
356 regional disturbances, placing ecosystems under increased threat of collapse (Kendrick et al.,  
357 2019). If resilience is overcome because of multiple disturbances, then compound disturbances



358 may cause a state change or ‘ecological surprise’ that is largely unpredictable (Paine et al.,  
359 1998).

360

361 For a common understanding of disturbances, we also need to acknowledge the central influence  
362 of time without explicitly defining a single general time scale of disturbances. Some disturbances  
363 impact ecosystem dynamics over short time scales (i.e., pulse events), whereas other  
364 disturbances operate over long time periods (i.e., ramp and press events) (Connell, 1997;Connell  
365 et al., 1997;Jentsch and White, 2019). Importantly, a single type of event may constitute a  
366 disturbance at one timescale, but not at another. While a forest fire may be a significant deviation  
367 from an environmental baseline considered on an annual or decadal scale (and therefore, a  
368 disturbance at this timescale), it may fall within the historical range of environmental variation at  
369 a centennial timescale (and therefore, not a disturbance at this timescale)(Turner, 2010).

370 Furthermore, the effects of slow increases in mean annual temperatures may be insignificant over  
371 the course of a few years when considering the background variation in mean annual  
372 temperatures (IPCC, 2019). However, at a centennial scale, the warming trend shifts the mean, as  
373 well as the frequency of extreme temperatures generating climates outside the range of historical  
374 variation. Similar arguments can be made for nitrogen deposition, chronic fertilization, pesticide  
375 applications, elevated CO<sub>2</sub>, and many other global disturbances (Jackson et al., 2001;Ferretti et  
376 al., 2010;Ripple et al., 2014).

377

378 This temporal perspective highlights that changing conditions through time (‘non-stationarity’  
379 (Milly et al., 2008;Vicente-Serrano and López-Moreno, 2008;Wolkovich et al., 2014) is also a  
380 central consideration for any conceptualization of disturbances to be applicable in the future.

381 Baseline conditions and driver–response relationships are dynamically conditioned by the

382 legacies of disturbance and ecological memory (Johnstone et al., 2016;Nowicki et al., 2019).  
383 Ecological succession is a classic example of ecosystem trajectories that interact with more  
384 discrete events to yield an aggregate disturbance impact. Disturbances can interrupt and  
385 potentially alter trajectories of succession through impacts on community dynamics that  
386 dramatically alter ecosystem functions (Ghoul and Mitri, 2016). For example, antibiotic  
387 administration and delivery mode can disrupt microbial community assembly and succession in  
388 the human infant gut microbiome that in turn can drive long-term impacts on host health (Koenig  
389 et al., 2011).

390

### 391 *Rising Importance of Interacting Disturbances*

392 An obstacle to historical paradigms of disturbance theory is that changes in environmental  
393 conditions will not only alter the frequency of disturbances, but also the potential for multiple  
394 interacting disturbances to impact system stability (Seidl et al., 2017). For example, drought may  
395 increase the vulnerability of wildfire in forests, or wildfire in forest enhances the probability of  
396 erosion and mudslides that affect ecosystems and communities downstream (Tiribelli et al.,  
397 2019). Multiple interacting disturbances can lead to novel ecosystem responses, sometimes  
398 impacting an ecosystem's resilience to the second disturbance (Folt et al., 1999;Darling and  
399 Côté, 2008;Buma, 2015;Burton and Boulanger, 2018) and compromising our abilities to  
400 understand disturbances in unknown future environments (Hobbs et al., 2009;Pidgen and Mallik,  
401 2013;Hobbs et al., 2014;Carlson et al., 2017;Mehran et al., 2017;Calderón et al.,  
402 2018;Zscheischler et al., 2018;Brando et al., 2019;Knelman et al., 2019;Ryo et al., 2019).

403

404 Given the variation that occurs both in disturbed systems and in the goals of disturbance studies  
405 and applications, we present a framework that describes a minimum foundation for best practices

406 for creating and sharing knowledge about disturbed systems in a novel and changing world.  
407 Towards this end, it then becomes necessary to follow standardized reporting practices to  
408 characterize the historical range of variation of disturbances and to classify individual events  
409 within a scale-flexible framework.

410

#### 411 **Minimum reporting standards.**

412 Because scales of investigation vary tremendously between disciplines, it is necessary for  
413 researchers to present sufficient data in publications and community repositories that capture  
414 complexity for other researchers to evaluate placement of their investigated disturbances within  
415 this framework (Slette et al., 2019). When possible, standardized indices are suggested to  
416 explicitly describe disturbance driver (e.g., Palmer Drought Severity Index, Standardized  
417 Precipitation Evapotranspiration Index, Normalized Burn Ratio) and impacts (e.g., quantifying  
418 the response of species and communities to disturbance) (Palmer, 1965; Van der Maarel,  
419 1975; Eidenshink et al., 2007; Vicente-Serrano and López-Moreno, 2008; Veraverbeke et al.,  
420 2010; Battisti and Fanelli, 2015). In ecological research, indices are most well-described for plot-  
421 scale studies and anthropocentric framings of scale that relate to our own human experiences  
422 rather than ecological processes (e.g., monetary losses from hurricanes), while they are more  
423 nascent for cross-scale disturbance work. Therefore, in addition to indices, it is necessary to  
424 report variables that describe the magnitude, duration, and rate of change of drivers and response  
425 variables in a consistent manner that is applicable regardless of scale. For example, Salafsky et  
426 al. (2008) propose a hierarchical lexicon for biodiversity conservation that divides elements of  
427 investigation threats vs. actions and a suite of nested variables beneath these categories. While  
428 this lexicon encompasses some aspects of disturbances described here, it is focused on one

429 aspect of disturbance impacts towards a specific end goal of determining priorities and resource  
430 allocations for conservation strategies.

431

432 We suggest three categories of variables for minimum reporting standards to facilitate a cohesive  
433 understanding of disturbances across scientific disciplines: (1) ecosystem properties, (2) driver  
434 descriptors, and (3) impact descriptors with suggested variables for each listed in Table 2. An  
435 integral distinction of these standards compared to previous efforts is the explicit recording of  
436 spatial and temporal scales needed for an interoperable understanding of disturbances (Peters et  
437 al., 2011). Ecosystem properties are foundational variables that provide context for disturbance  
438 interpretation (e.g., ecosystem type, successional state, and system stability). Driver and impact  
439 descriptors are each divided into three categories: reference, spatial, and temporal variables.  
440 These variables capture system stability and spatiotemporal dynamics that allow for multiscale  
441 comparisons including mild versus extreme intensity, acute versus chronic timescales, and abrupt  
442 versus gradual change (Ryo et al., 2019). Collectively, they allow for the placement of events on  
443 both the driver and impact axes of the proposed conceptual framework as well as providing  
444 context that describes the scale and scope of the investigation.

445

#### 446 **Promising Cross-Disciplinary Approaches to Address Areas of Need.**

447 While we address some challenges of disturbance research here, developments in technology,  
448 methodology, and cross-disciplinary approaches are necessary to close knowledge gaps. We  
449 highlight the need to *integrate disturbance responses across scales of ecological organization*,  
450 from genes to ecosystems. We expect that future studies should consider multiple scales of  
451 sampling and analysis that comprehensively evaluate disturbances and their effects across

452 spatial, temporal, and/or organismal scales. Ecological hierarchies underlie self-organized  
453 ecosystems and provide a structure for using information theory and other advanced statistical  
454 techniques to predict whole ecosystem impacts (Allen and Starr, 2017;Arora et al., 2019). Social-  
455 ecological applications of machine learning, graph theory, and information theory are  
456 exponentially increasing and can decipher complex relationships in multidimensional data  
457 streams as well as scale dynamics from pore-to-global scales (Peters et al., 2018;Weintraub et al.,  
458 2019). These approaches are used to collapse complex data types into tangible variables by  
459 deciphering classes of organisms and relationships among these classes through space or time.  
460 They reveal the organizational structure of a system through interaction networks that include  
461 both random and ordered processes (Ings et al., 2009). Remote sensing can also aid in evaluating  
462 the spatial extent and spatial patterning of disturbance, thereby defining the appropriate scale of  
463 sampling for these analyses (Shiklomanov et al., 2019). However, empirical tests on the potential  
464 for disturbance impacts to propagate through ecosystem hierarchies are lacking and is a major  
465 research need. One opportunity would be the use of paired experimental and modeling  
466 approaches to elucidate networked changes in ecological systems resulting from disturbance  
467 impacts. The use of experiments and clearly outlined hypotheses is increasingly argued as a core  
468 need for generating predictive understandings of ecosystem responses to disturbance (Spake et  
469 al., 2017;Currie, 2019).

470

471 Our second area of need also considers the broader issue of scale—understanding *how*  
472 *disturbances interact with each other* and potentially compound through space and time. Recent  
473 work has underscored interactions between extreme events occurring closely in space and time,  
474 for example by elucidating how the discrete effects of flooding on biogeochemistry are related to  
475 prior fire exposure (Knelman et al., 2019), and that the effects of a fire may depend on previous

476 droughts or insect outbreaks (Burton and Boulanger, 2018). The long-term processes of  
477 environmental change also have multifaceted impacts on ecosystems but are most frequently  
478 studied independently (Rillig et al., 2019; Song et al., 2019). Such work accentuates questions  
479 into ecosystem trajectories—as disturbances increase through time, are there thresholds beyond  
480 which ecosystems are irreversibly altered? Thus, a multivariate perspective is necessary to  
481 accurately assess the impact of interacting disturbances (Zscheischler and Seneviratne, 2017).

482

483 The evolutionary consequences of living in an environment with recurrent disturbances are also  
484 poorly understood (Pausas and Keeley, 2014; Pausas et al., 2017). Some species, for example,  
485 have evolved specific life-history adaptations that enable them to not only survive and exploit  
486 disturbances but even to require them for their persistence (e.g., savannas, Mediterranean  
487 shrublands, alpine vegetation, riparian cottonwoods) (Mahoney and Rood, 1998; Lytle and Poff,  
488 2004; Keeley et al., 2011; de L. Dantas et al., 2013; le Roux et al., 2013). Similarly, disturbances  
489 have countervailing effects on population dynamics in that they can cause immediate mortality  
490 of species, but also create new habitat, thereby increasing growth rate or increasing population  
491 size post-disturbance (Pausas and Keeley, 2014; McMullen et al., 2017). For instance, if the  
492 consequences of climate change-related disturbances are studied separately, the results may be  
493 greatly biased as compared to when the consequences are considered simultaneously (Niittynen  
494 et al., 2018). Therefore, the interactions between disturbances that change eco-evolutionary  
495 dynamics provide a relatively unexplored area for future research.

496

497 A third research need, *establishing appropriate baselines and trajectories* for different  
498 ecosystems, disturbance, and organism types is essential for evaluating disturbances that alter  
499 ecosystem structure and function. Paleoecological data can provide historical reference baselines,

500 help evaluate sensitivity to disturbances across different windows of space and time, and unveil  
501 past state changes that provide a foundation for understanding how ecological hierarchies will  
502 respond to future environmental changes (Lamentowicz et al., 2019). Time-series methods are  
503 also well-equipped to separate disturbances from long-term trends and evaluate changes in  
504 disturbance regimes through time (e.g. wavelet analysis) (Keitt, 2008;Tonkin et al., 2017). For  
505 instance, Sabo and Post (2008) developed tools based on Fourier analysis to disentangle the  
506 periodic (seasonal), stochastic (interannual), and catastrophic components of river flow regimes.  
507 Space-for-time approaches, in which distances from an event are used as a proxy for the time-  
508 since-event, can reveal long-term impacts without necessitating decades of monitoring (Pickett et  
509 al., 1989;Walker et al., 2010). Although space-for-time investigations require a correlation  
510 between the age of an ecosystem attribute and spatial structuring that may not be applicable to  
511 highly disturbed landscapes, chronosequences can be used to investigate plant and soil  
512 successional processes at decadal to millennial timescales (Walker et al., 2010;Laliberté et al.,  
513 2013;Sutherland et al., 2016;Fanin et al., 2018).

514  
515 Finally, we underline the need for enhancing predictive capabilities through *process-based*  
516 *models and ecological forecasting initiatives* that represent the disturbance drivers on ecosystem  
517 attributes, going beyond historical correlations that fail to represent causal relationships (Dietze  
518 et al., 2018;Tonkin et al., 2019). Generating a model robust to disturbance type, ecosystem, and  
519 scale that allows managers to detect disturbance drivers and predict disturbance impact sizes is  
520 one of the ultimate goals of disturbance ecology. Mechanistic models can further progress  
521 towards this goal by representing interactions among species through time (Tonkin et al., 2018).  
522 Process-based and forecasting models can be tailored to highly specific conditions and can  
523 provide managers with both a predicted outcome and a range of uncertainty based on the

524 underlying driver (Tonkin et al., 2019). They are commonly used to guide management practices  
525 in fisheries and conservation efforts (Tonkin et al., 2019). Collectively, process-based and  
526 forecasting models are potential tools developing mitigation strategies and informing how  
527 humans might intervene at individual, local, regional, and global scales to minimize social-  
528 ecological damages caused by disturbances (Dale et al., 1998;Berkes et al., 2000;Folke et al.,  
529 2005;D’Amato et al., 2011)(Dale et al. 1998, Berkes et (Goldstein et al., 2020)al. 2000, Folke et  
530 al. 2005, D’Amato et al. 2011).

531

### 532 **Conclusion.**

533 Our work synthesizes knowledge globally across institutions using crowdsourced open science  
534 and demonstrates that novel approaches can generate emergent ideas greater than the sum of  
535 their independent disciplinary parts. The integration of interdisciplinary contributions of over 50  
536 individuals, from 42 institutions - from academic, governmental, and non-governmental  
537 organizations - in 15 countries, into the novel conceptual framework presented here demonstrates  
538 the currently untapped potential for supporting collaborative co-creation of research, facilitated  
539 by social media and collaborative writing platforms. For a detailed description of the writing  
540 process and contributor demographics, see Graham and Smith (2020). Our experiences through  
541 this process motivate us to encourage the wider scientific community to continue to explore the  
542 suitability of similar approaches for facilitating collaborative research that benefits from a large  
543 interdisciplinary knowledge base and allows us to fully embrace Open Science principles in  
544 collaborative interdisciplinary research.

545

546 Using a completely open and crowdsourced scientific approach, we integrate insights from  
547 numerous scientific perspectives to present a generalizable framework for cross-disciplinary



548 disturbance investigations. We discuss and use ideas that are common across multiple disciplines  
549 to underlie the framework as a foundation for investigations into the causes and consequences of  
550 disturbances. Discipline-specific variables can supplement this framework to generate deeper  
551 insight into specific research questions. We highlight that the current lexicon used to discuss  
552 disturbances generates confusion by conflating events that drive ecological change with the  
553 impacts of extreme events. To overcome this challenge, we propose parsing disturbance theory  
554 between disturbance drivers and disturbance impacts and encourage researchers to be explicit  
555 about how they define their studied disturbance within this context.

556

557 Using drivers and impacts as axes of variation, the framework generates four universal  
558 disturbance types that are applicable regardless of the line of inquiry or its spatiotemporal scale  
559 (Figure 1). To provide consistency in comparing disturbances within this framework, we suggest  
560 three categories of variables for minimum reporting standards: i) ecosystem properties that  
561 provide context and ii) disturbance driver and iii) disturbance impact descriptors that capture  
562 system stability and spatiotemporal dynamics.

563

564 We also highlight promising lines of research to generate a more universal understanding of  
565 disturbance events and their impacts, including integrating scales of ecological research,  
566 understanding how disturbances interact with each other, establishing appropriate baselines and  
567 trajectories, and developing process-based models and ecological forecasting initiatives that will  
568 enable robust prediction capabilities and mitigation strategies. As global change accelerates the  
569 threats of disturbances, the framework presented here serves as a foundation for cross-  
570 disciplinary discussion of the complexities of understanding the causes and consequences of  
571 disturbances across studies with different scientific and management goals. We encourage

572 researchers to be explicit in how they define disturbance drivers and impacts and to continue to  
573 work towards interoperable terminology and knowledge of disturbances.

574

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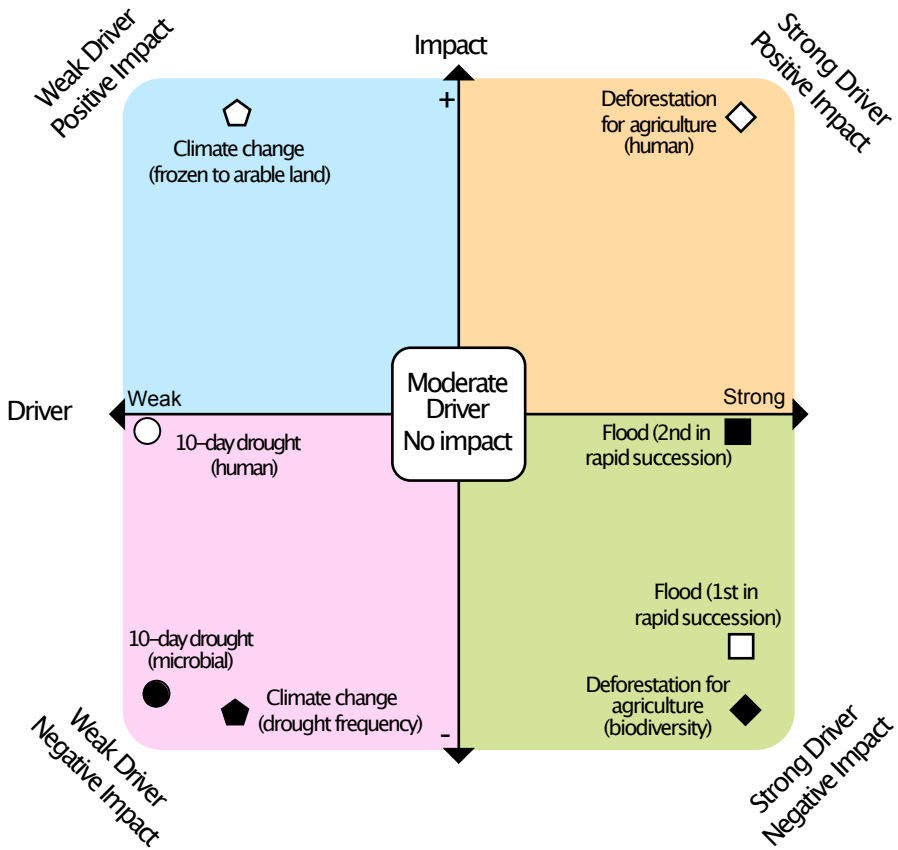


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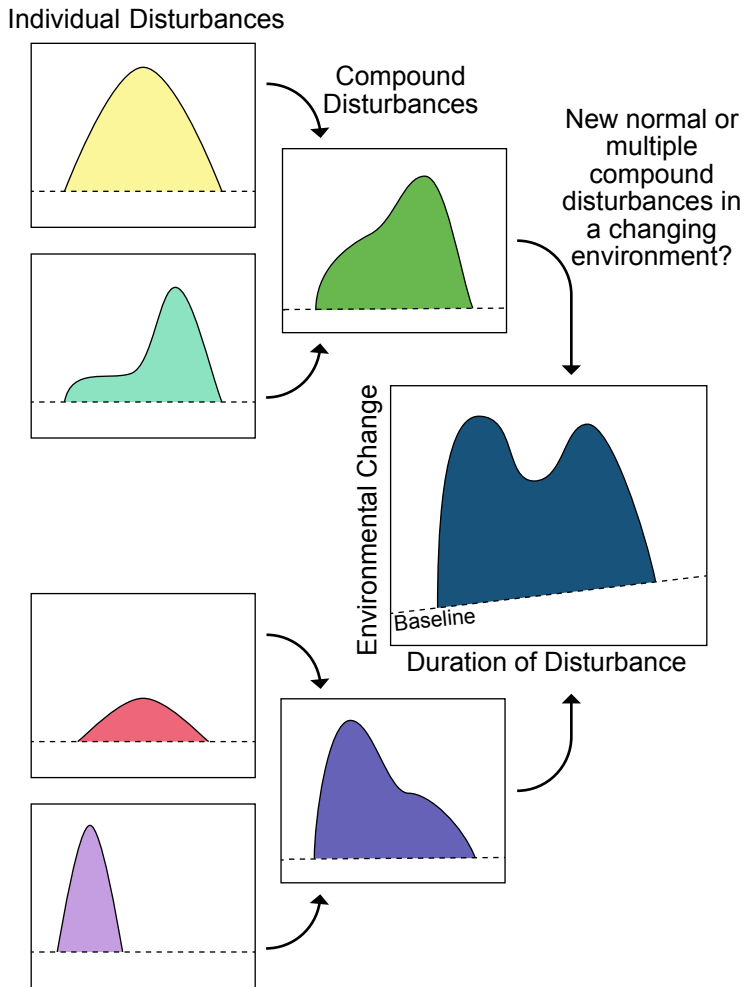
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**Figures and Tables.**



**Figure 1.** Current disturbance lexicon conflates two distinct processes—events that drive ecological change and impacts of extreme events—both interchangeably termed disturbances despite fundamental distinctions between the two types of processes. We disentangle these processes to derive four universal types of disturbances that are applicable regardless of the line of inquiry or its spatiotemporal scale. Drivers (x-axis) are defined as when a force, either biotic or abiotic, generates deviation from local, prevailing background conditions. A driver is characterized by its magnitude of deviation from an environmental baseline (low to high deviation denotes weak to strong driver). Impacts (y-axis) are defined as the impact of social-ecological consequences of a driver relative to a scale-dependent baseline state. Impacts can be positive or negative depending on the perspective of the study. Each quadrant is, therefore, a unique disturbance type defined in more detail in Table 1, and the position of drivers and impacts across and within the quadrants slides with the line of inquiry. Examples of disturbances across spatial and temporal scales are denoted within each quadrant.



**Figure 2.** An obstacle to historical paradigms of disturbance theory is that changes in environmental conditions will not only alter the frequency of disturbances, but also the potential for multiple interacting disturbances. As multiple disturbances compound through time, a crucial question emerges: “When does a disturbed state become normal?” Compound disturbances can take many forms and result in both linear and non-linear ecosystem responses. As an example, Figure 2 shows an additive trajectory of disturbances and resultant environmental change. The leftmost panels represent single disturbance events that have long been the targets of scientific research. As disturbances aggregate through time, a new class of ‘compound’ disturbances have been a rising topic (middle panels). With the continuing increases in the frequency and intensity of disturbances, a key challenge remains to disentangle multiple compounding disturbances from normal variability in ecosystem functions (rightmost panel). Another challenge is that environmental baselines (dashed line) shift through time, adding a chronic component to the study of short-term disturbance events.

**Table 1.** Description and examples of four universal disturbance types generated by proposed framework.

Quadrant	Description	Example
High Deviation-Negative Impact	Occur when large deviations from environmental baselines generate negative impacts on ecosystem functions.	Category 5 hurricane Mass wasting Oil spills Tornados Floods (human perspective) Wildfires (human perspective) Deforestation biodiversity impacts
High Deviation-Positive Impact	Occur when large deviations from environmental baselines generate positive impacts on ecosystem functions.	Deforestation for agriculture increasing crop production (human perspective) Floods (wetland ecosystems)
Low Deviation-Negative Impact	Occur when small deviations from environmental baselines generate negative impacts on ecosystem functions.	Short term drought-induced microorganism mortality Climate change-induced (i.e., temperature/CO <sub>2</sub> driven) drought impacts
Low Deviation-Positive Impact	Occur when small deviations from environmental baselines generate positive impacts on ecosystem functions.	Climate change (human societies in very cold environments) Small wildfires that prevent catastrophic megafires

**Table 2.** Proposed minimum reporting standards for interoperability of disturbance investigations.

Ecosystem Properties	Reference Disturbance Properties (Reported for both Driver and Impact)	Temporal Disturbance Properties (Reported for both Driver and Impact)	Spatial Disturbance Properties (Reported for both Driver and Impact)
Ecotone	Reference Baseline State	Duration	Coordinates
Successional State	Method for determining baseline state	Rate of onset	Scale of study
Resistance	Intensity (deviation from mean or baseline)	Rate of decline	Scale of disturbance
Resilience		Variability through time	Area of extent
Recovery		Variability through space	
Temporal Stability			
Method and input variables for determining resistance, resilience, recovery, and temporal stability (recommend Hillebrand et al 2018)			