

1 **Interpopulation differences in developmental reaction norms of both predator and prey**
2 **determine their trophic interaction**

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15 Running head: Reaction norms in trophic interaction

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18

19 **Abstract**

20 Phenotypic plasticity as a type of reaction norm creates variation in ecological interaction
21 strengths across different environmental conditions. If populations of both interacting species
22 (e.g., predator and prey) differ in the plasticity of their ecological traits, populations of both
23 interacting species and environmental conditions can jointly determine their interaction strength.
24 To examine this untested prediction, we experimentally investigated how geographic populations
25 of both predator and prey species with differential reaction norms of offensive and defensive
26 morphological traits, respectively, establish their trophic interactions, using salamander larvae
27 (*Hynobius retardatus*) and their prey frog tadpoles (*Rana pirica*). Past studies showed that
28 gigantism of salamander larvae as a result of cannibalism in their hatchling period triggers a
29 salamander–tadpole trophic interaction because only gigantic salamanders with remarkably
30 enlarged gape can swallow the large prey, frog tadpoles with an inducible morphological
31 defense. By manipulating combinations of two geographic populations of the salamanders and
32 tadpoles (*i.e.* Erimo and Chitose) and cannibalistic conditions of salamander hatchlings, we
33 examined how developmental reaction norms of both predator and prey shape this trophic
34 interaction. We found that geographic population identity of both salamander and tadpoles and
35 the cannibalistic condition of salamanders interactively determined the trophic interaction
36 strength between salamanders and tadpoles. Under cannibalistic conditions, giant salamander
37 larvae emerged, and gigantism was more prominent in Erimo than Chitose salamanders. While
38 the greater emergence of cannibalistic giant salamander larvae in Erimo salamanders resulted in
39 significant predatory impacts on the prey tadpoles, the predatory impacts by cannibalistic Erimo
40 salamander was larger for Chitose tadpoles than Erimo tadpoles because of Chitose tadpoles'
41 lesser ability to develop morphological defenses. This study demonstrates that developmental

42 reaction norms of interacting partners interactively determine ecological interaction, and
43 therefore suggests that genetic differences in reaction norms among geographic populations of
44 both interacting partners jointly shape variation in the interaction strength among geographic
45 regions.

46

47 *Key words: Genotype×Genotype×Environment interplay; inducible offense; inducible defense;*
48 *morphological plasticity; intraspecific variation; β -diversity; arms race; coevolution*

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50

51 **Introduction**

52 The factors determining the strength of ecological interactions have been a major interest in
53 ecology because ecological interactions are principal mechanisms determining demography of
54 the species involved in the interactions, and its effects propagate through the ecosystem (Nakano
55 et al. 1999, Persson et al. 2007, Ushio et al. 2018). Although interaction strengths depend on the
56 density of the interacting species, it is also strongly affected by the functional traits of the
57 interacting species (Sinclair et al. 2003, Schmitz et al. 2015). Since expression of functional traits
58 vary among species, species-specific values of functional traits (i.e. mean) have been a focus
59 when disentangling community structure and dynamics (McGill et al. 2006, Schmitz et al. 2015).
60 However, values in some key functional traits can vary considerably within species and this
61 intraspecific trait variation can be involved in generation of complicated variation in the strength
62 of ecological interactions (Miner et al. 2005, Miller and Rudolf 2011). Therefore, intraspecific
63 trait variation is increasingly recognized as an important factor in determining community
64 structure and dynamics (Bolnick et al. 2011, Des Roches et al. 2018, Raffard et al. 2018).

65 The effects of intraspecific trait variation on ecological interactions have been mainly
66 investigated in terms of three interacting elements: size, phenotypic plasticity, and genotype. The
67 first aspect is size variation among individuals. Because multicellular organisms generally shift
68 their interacting partners as they grow larger, a species' size composition is a factor determining
69 the types and strength of their ecological interactions (Persson et al. 2007, Miller and Rudolf
70 2011, Yamaguchi et al. 2016). Second is phenotypic plasticity as a form of reaction norm. Since
71 individuals change their functional traits (e.g., behavior, life history and morphology) in response
72 to different environmental conditions, interaction strengths can be altered by plastic responses of
73 individuals to the environmental conditions they face (Miner et al. 2005, Winterhalter and
74 Mousseau 2007, Kishida et al. 2014). For example, while many prey species enhance expression
75 of defensive traits in the presence of a specific predator to effectively protect them against
76 predators (*i.e.* inducible defense), some predator species enhance expression of offensive traits in
77 the presence of particular prey items to effectively catch and consume them (*i.e.* inducible
78 offense, Kishida et al. 2010). Both the expression of defensive traits of prey and offensive traits
79 of predators can strongly determine the sign and strength of predator-prey interactions (Kishida
80 et al. 2009b, 2014). The third context for intraspecific variation effects on ecological interactions
81 is genetic trait variation. Expression of functional traits itself is under genetic control (Miner et
82 al. 2005, Pigliucci 2005, Winterhalter and Mousseau 2007). If the value of a functional trait is
83 different among genetic populations, these populations can differ in the strength of their
84 ecological interactions (Yoshida et al. 2003, Hiltunen and Becks 2014, Bassar et al. 2017).

85 The importance of each of the entities to shape ecological interactions has been evidenced
86 by numerous studies (Bolnick et al. 2011, Miller and Rudolf 2011, Des Roches et al. 2018).
87 Although the effects of these entities have been considered independently, they also tightly link

88 to each other. For morphological traits, in particular, we can acknowledge clear linkage among
89 size variation, phenotypic plasticity and genetic variation. Changes in whole body size (i.e.,
90 ordinal isometric growth) or the size of organ parts (i.e., allometric growth) are sometimes
91 facultative rather than constitutive (Kishida et al. 2006, Gerber et al. 2008), and such plastic
92 change in morphological traits itself is under genetic control (i.e., reaction norm perspective,
93 Miner et al. 2005, Pigliucci 2005). Hence, the ability to develop functional morphological traits
94 can genetically differ among geographic populations (Kishida et al. 2007, Winterhalter and
95 Mousseau 2007). Consequently, the strength of ecological interactions is determined by a
96 combination of geographic populations and environmental conditions (i.e., $G \times E$ effect on
97 interaction strengths) (Yamamichi et al. 2019). Although this integrative view is quite intuitive,
98 there is little evidence of the expected effects of genetic variation in reaction norms of
99 morphological traits on ecological interactions (Kasada et al. 2014). The first objective of this
100 study is to test this prediction.

101 When investigating ecological interactions, focusing on the functional traits of both
102 interacting partners rather than either one alone is essential because the sign and strength of the
103 interaction should be determined by the performance of both interacting partners (Hiltunen and
104 Becks 2014, Bassar et al. 2017). For example, in predator-prey relationships, the consequence of
105 trophic interactions is determined by a balance of the predator's foraging performance and prey's
106 defensive performance (Takatsu & Kishida 2013). If populations of both predator and prey differ
107 in their functional trait reaction norms, populations of both interacting species and environmental
108 condition can jointly determine their interaction strength. Furthermore, the reaction norms of
109 both predator and prey may interactively operate on the interaction strength because functional
110 trait values of an interacting partner affect expression of functional traits of the opponent

111 (Kishida et al. 2010). For example, some prey species intensify expression of their defensive
112 traits when exposed to dangerous predators with more offensive traits (Kopp and Tollrian 2003,
113 Kishida et al. 2006). This implies that populations of both interacting species and environmental
114 conditions can interactively shape geographic variation in interaction strength (i.e., $G \times G \times E$
115 interactive effects on ecological interaction). Although testing this prediction is important to
116 advance our integrative understanding of the mechanisms shaping geographic variation in
117 ecological interactions in natural systems, to our knowledge, this remains untested. We thus
118 aimed to additionally test this prediction.

119 To test both predictions, we conducted an experiment using a trophic relationship between
120 predatory larval salamanders (*Hynobius retardatus*) and prey frog tadpoles (*Rana pirica*) as a
121 model system. Their trophic interaction is an excellent system to accomplish this objective,
122 because both predator and prey species have remarkable phenotypic plasticity in their functional
123 morphological traits (i.e., size of gape and body parts for salamanders and tadpoles, respectively)
124 in the trophic interactions. While *H. retardatus* salamander larvae can exhibit gigantism
125 characterized as enlarged gape in cannibalistic conditions (inducible offense), prey tadpoles can
126 exhibit body enlargement by thickening epithelium tissue as an inducible defense. Importantly,
127 expression ability of the antagonistic phenotypes vary among the geographic populations of both
128 species (Michimae and Wakahara 2002, Kishida et al. 2007). To examine whether and how
129 genetic differences in developmental reaction norms of predatory salamander and prey tadpoles
130 shape their trophic interactions, we conducted an experiment in which combinations of two
131 populations of the predatory salamanders and prey tadpoles and initial condition relevant to the
132 emergence of gigantism of salamanders were manipulated.

133

134 **Materials and Methods**135 *Study System*

136 *Hynobius retardatus* salamanders and *Rana pirica* frogs usually spawn in small ponds in
137 early spring in Hokkaido, Japan. Although salamander larvae are carnivores, the trophic
138 relationship with tadpoles is not always established even if the larvae of both species cohabit the
139 same ponds. Since tadpoles typically hatch 3–4 weeks earlier than salamanders, tadpoles are too
140 large to be consumed by salamander hatchlings (Nosaka et al. 2015). Therefore, a predator-prey
141 interaction between two species occurs when salamander larvae grow rapidly. Typically, the
142 rapid growth of salamander larvae is caused by cannibalism during their hatchling stage (Takatsu
143 and Kishida 2015). Salamander hatchlings that successfully consume conspecifics tend to grow
144 rapidly and become ‘giants’ with much larger body and gape. The emergence of cannibalistic
145 giants occurs more frequently in the presence than absence of tadpoles, since disturbance effects
146 of tadpoles enhance cannibalism among salamander hatchlings (Takatsu and Kishida 2020).
147 Importantly, relative gape to body size of the cannibalistic giants is greater than that of the non-
148 cannibalistic salamanders. That is, the cannibalistic salamanders have an extremely large gape
149 (i.e., inducible offensive phenotype) and, hence, individuals with the offensive phenotypes can
150 consume tadpoles as alternative large prey items (Takatsu and Kishida 2015).

151 Antagonistically, frog tadpoles exhibit inducible defense in this system. In the presence of
152 the salamander larvae, frog tadpoles enlarge body and tail by thickening their epithelium tissue.
153 This ‘bulgy’ phenotype makes it harder for the salamander larvae to swallow them (Kishida and
154 Nishimura 2004). Importantly, expression of the defensive bulgy phenotype depends on
155 predation risk. The tadpoles exhibit bulgier phenotype in the presence of offensive giant
156 salamanders compared to non-offensive ones (Takatsu and Kishida 2015, Takatsu et al. 2017).

157 Thus, larvae of the two amphibian species exhibit antagonistic morphological plasticity,
158 represented by condition-dependent allometric growth in functional traits (i.e. gape of
159 salamanders and body of frog tadpoles), that greatly affects trophic interactions between them
160 (Kishida and Nishimura 2004, Takatsu et al. 2017).

161 The potential to become an offensive giant salamander and to express the defensive bulgy
162 phenotype in tadpoles differ among their geographic populations (Michimae and Wakahara
163 2002, Kishida et al. 2007). In particular, there is large variation in the emergence of offensive
164 giants among the geographic populations of salamanders (e.g., Michimae 2006). Based on
165 knowledge of the geographic variation in inducible offense of salamanders, we selected two
166 localities of amphibians as experimental model populations: Erimo and Chitose. Compared to
167 Chitose population of salamander (hereafter Chitose salamander), Erimo population (hereafter
168 Erimo salamander) is characterized by higher frequency in the emergence of offensive giants
169 (Michimae 2006; Atsumi and Kishida, unpublished data). Using these two populations of
170 salamanders allows us to test our predictions, even though we had no prior knowledge about the
171 geographic variation in the expression ability of defensive phenotype in tadpoles between these
172 two geographic populations. We conducted the following experiment using the two geographic
173 populations of salamanders and tadpoles by collecting their eggs from several ponds located in
174 the Erimo (seven ponds around 42°6' N; 143°16'E) and Chitose (three ponds around 42°48' N;
175 141°35' E) regions. Collection and keeping methods of experimental animals are described in
176 Appendix S1.

177

178 ***Experimental Setting***

179 The experiment was conducted in an experimental room in Tomakomai Experimental
180 Forest, using semi-transparent polypropylene 22-L tanks (51.3 cm × 37.2 cm × 16.6 cm) of
181 which the bottom was covered by sand to a depth of 2 cm as benthic substrate. Minimum natural
182 water (ca. 10 ml per minute) was constantly supplied into each tank by using thin polypropylene
183 hose and flowed out into an overflow pipe ($\phi = 20\text{mm}$, 4cm height) set inside of the tank. The
184 overflow pipe was covered with mesh net (mesh size 1mm) to prevent the experimental animals
185 flowing out. Natural water was drawn using a pump from a well that is 5m away from a natural
186 river. The overflow pipe kept water depth (from sandy bottom to water surface) at 4 cm. Two
187 leaves of Japanese bigleaf magnolia (*Magnolia obovata*) (dry weight: 5 g) were provided as
188 biotic structures.

189 To examine how genetic variation in developmental reaction norms of predatory
190 salamander and prey tadpoles shape their trophic interactions, we conducted an three-way
191 factorial experiment in which combinations of the two geographic populations of the predatory
192 salamanders and prey tadpoles (i.e., Erimo and Chitose populations) and cannibalistic conditions
193 that possibly affecting emergence of the offensive giant salamanders were manipulated (i.e.,
194 Cannibalism and No-cannibalism treatments). We haphazardly placed 30 three-week-old
195 tadpoles that originated from either the Erimo (mean \pm SD snout-vent length, 7.32 ± 0.60 mm, $N =$
196 20) or Chitose population (7.10 ± 0.53 mm, $N = 20$) into each of all 80 tanks on May 17, 2018
197 (defined as day 1).

198 Cannibalistic conditions were controlled by manipulating size structure of salamander
199 hatchlings; greater size heterogeneity among salamander hatchlings facilitates cannibalism
200 (Kishida et al. 2015). Following a well-established method (Takatsu & Kishida 2015, Takatsu et
201 al. 2017), we manipulated size structure of salamander hatchlings by using individuals that

202 hatched at different times (i.e., early- or late-hatchlings) while keeping the total initial density
203 constant across treatments. We set the following two treatments: (1) the “Cannibalism” treatment
204 received 5 early- and 10 late- salamander hatchlings to create initial size heterogeneity and (2)
205 the “No-cannibalism” treatment received either 15 early- or 15 late- salamander hatchlings. For
206 this manipulation, we obtained the early- and late- salamander hatchlings by controlling the
207 water temperature experienced by the embryos from a single egg cluster. The difference in hatch
208 timing between early- (16th May) and late- (24rd May) salamander hatchlings was 8 days. The
209 salamander hatchlings were assigned to the relevant treatments 1 day after they hatched (total
210 length at day 8, $N = 20$ each: 25.36 ± 2.40 mm and 20.33 ± 1.16 mm in Erimo early and late
211 hatchlings; 20.97 ± 2.08 mm and 17.08 ± 1.30 mm in Chitose early and late hatchlings). The
212 method for obtaining the early and late- hatchlings is described elsewhere (Takatsu and Kishida
213 2015, Takatsu et al. 2017, Takatsu and Kishida 2020).

214 Although No-cannibalism treatment consisted of the two specific conditions (i.e., either
215 early hatchlings only or late hatchlings only), we pooled the data of the two alternative hatch
216 timing conditions, because our preliminary analyses showed non-significant effects of hatch
217 timing of salamanders on survival of both species in the No-cannibalism treatment (Appendix
218 S2) . Note that to avoid excessive use of the animals, we did not prepare a tadpole-only treatment
219 for estimating background mortality of tadpoles. This is acceptable since previous studies
220 repeatedly showed that mortality of tadpoles in the absence of predators was negligibly low in
221 similar experimental settings (Nosaka et al. 2015; Takatsu & Kishida, 2015; Yamaguchi et al.
222 2016; Takatsu et al. 2017; Takatsu and Kishida 2020). Throughout the experiment, we added a
223 piece of rabbit chow (dry weight: 0.2 g) and 100 frozen *Chironomid* larvae to all tanks on every
224 Monday, Wednesday and Friday as an alternative food for the tadpoles and the salamanders,

225 respectively. The experimental room was maintained under natural light-dark (ca. 14h/10h)
226 conditions and water temperature ranged from 13 to 20 °C. At day 11, 18, 25, 32 and 39, we
227 counted all surviving tadpoles and salamanders. We defined the period between successive days
228 on which we counted surviving individuals as the experimental period: 1st (day 1-11), 2nd (day
229 11-18), 3rd (day 18-25), 4th (day 25-32) and 5th period (day 32-39). From the count data, we
230 calculated survival rate of the salamanders and frogs in each experimental period by dividing the
231 number of survivors at the end of each period by that at the beginning of the period. In addition,
232 we photographed the dorsal side of surviving animals using a digital camera (Panasonic Lumix
233 DC-TZ90) at the end of 2nd period (day 18) just before predation by salamanders on tadpoles
234 began. The experiment was ended on day 39 because metamorphosis of tadpoles began in some
235 tanks. Our conditions for hatch timing and density of the two amphibians species are relatively
236 high but within a range in the natural habitats (Michimae 2006; Nosaka et al. 2015).

237

238 ***Statistical analyses – cannibalism and interaction strength***

239 We investigated how the time-trajectory of the strength of salamander cannibalism and
240 predator-prey interaction between salamander and tadpole was shaped by the three factors:
241 Cannibalism condition, and populations of salamander and tadpole. Previous studies have shown
242 that temporal change in the strength of cannibalism and tadpole consumption depends on the
243 initial size structure of salamanders; when salamander hatchlings are uniform in their size,
244 cannibalism and tadpole consumption rarely occurs and thus survival rate of two species are
245 constantly high. In contrast, when size variation among salamander hatchlings is large,
246 cannibalism occurs and subsequently tadpole consumption occurs as a result of emergence of
247 offensive giant salamanders (Takatsu and Kishida 2015). Once offensive giant salamanders

248 consume tadpoles, their predation become intense through time because consumption and growth
249 of salamanders positively feedback (Nosaka et al. 2015). This process of the trophic interaction
250 has been shown in experiments by using some geographic populations of salamanders and
251 tadpoles which were different from those used in the present study (Takatsu and Kishida 2015,
252 2020, Takatsu et al. 2017). If cannibalism occurrence and emergence of offensive giants of
253 salamanders and/or defensive performance of tadpoles are not uniform among the geographic
254 populations, the temporal dynamics of salamanders' cannibalistic interaction and trophic
255 interaction with tadpoles should vary depending on a combination of cannibalism condition and
256 the populations of salamanders and tadpoles. We therefore predicted that salamander size
257 structure and populations of two species jointly determine the time-trajectory of cannibalism and
258 tadpole consumption strength. To investigate this prediction, we fit a linear mixed model with
259 normal errors (LMM) to natural logarithms of salamander and tadpole survival rate in each
260 experimental period (see below). The explanatory variables were the three factors above (i.e.,
261 cannibalism condition of salamanders, salamander population, and frog population),
262 experimental period (centered [Schielzeth 2010]), and their interaction. Experimental tank
263 identity was included as a random effect because we measured survival rate five times in each
264 tank. We thus expect a significant effect of the interaction among cannibalism condition (i.e.
265 $E_{\text{cannibalism}}$ representing Cannibalism and No-cannibalism treatment), the population of two
266 species (i.e. $G_{\text{salamander}}$ and G_{tadpole} representing Erimo and Chitose populations) and experimental
267 period. We firstly examined the significance of this interaction term, and then separatory
268 analyzed Cannibalism and No-cannibalism treatment to interpret the biological meanings of
269 complex interaction terms.

270 We assumed that the mortality of salamanders and tadpoles represent the strength of

271 cannibalism and trophic interaction between two species, because previous experiments
272 conducted in similar experimental settings showed that the mortality of two species is quite low
273 in the absence of the trophic interactions (Kishida et al. 2011; Nosaka et al. 2015; Takatsu and
274 Kishida 2015; Takatsu et al. 2017). Since our study investigated the effect of multiple interacting
275 species on survivorship represented as proportion data, the multiplicative risk model was
276 employed (Sih et al. 1998). In the multiplicative risk model, proportion data is log-transformed
277 so that independent effects of different species can be assessed (Soluk and Collins 1988).
278 Therefore, we used natural logarithms of survival rates of salamanders and tadpoles as response
279 variables for the analyses of interaction strength. The statistical significance of each parameter in
280 mixed models was computed via the Satterthwaite approximation (Satterthwaite 1946). All
281 analyses in this study were performed in R 3.6.2 (R Development Core Team 2019). We used R
282 package *lme4* (v. 1.1; Bates et al. 2015) to construct LMMs, and package *lmerTest* (3.1;
283 Kuznetsova et al. 2017) to evaluate the significance of parameters within LMMs.

284

285 ***Phenotyping***

286 To dissect mechanistic details of the trophic interactions, we assessed morphological traits
287 of salamanders and tadpoles. From dorsal side photographs of the surviving amphibians at the
288 end of 2nd period, we measured heart-vent length and gape width of salamanders and maximum
289 body width and snout-vent length of tadpoles, using Image J software (Schneider et al. 2012).
290 For tadpoles, as many individuals were measured as possible per tank (i.e., moving or inclining
291 individuals on the photograph were not measured). For salamanders, the individuals with the
292 four largest body length values were measured in each tank, because very few salamanders can
293 become offensive giants as potential predators for tadpoles. Gape width of salamanders and

294 maximum body width of tadpoles are considered as functional traits that are critically important
295 for their trophic relationship, because size balance between gape of salamander and the largest
296 body part of tadpoles strongly determines predation success (Kishida and Nishimura 2004,
297 Takatsu and Kishida 2013). Heart-vent length of salamanders and snout-vent length of tadpoles
298 are considered as body size (Takatsu and Kishida 2013, Kishida et al. 2015). In addition, we
299 calculated relative gape width of salamanders (i.e., gape width / heart-vent length) and relative
300 body width of tadpoles (i.e., maximum body width / snout-vent length) to represent the degree of
301 salamander's offensive phenotype expression and that of tadpole's defensive phenotype
302 expression (Kishida et al. 2009a, Takatsu and Kishida 2013). By focusing on morphological
303 variables at the end of 2nd period (just before intense predation by salamanders on tadpoles
304 started [see Results]), we dissected mechanistic causes underlying differentiated trophic
305 interactions between salamanders and tadpoles among the treatments. We calculated mean values
306 of the tadpole variables in each tank for use in the statistical analyses. We used variables of the
307 salamanders whose body size was largest in each tank in the statistical analysis because the
308 number of offensive giants was very few (0-2) in each tank if emerged. In addition, using the
309 morphological data at the end of 2nd period, we examined the number of "potential predators" in
310 each tank, which is defined as the salamander whose gape width exceeded the mean tadpole
311 body width by 1.1 times (Takatsu and Kishida 2015). This count data is useful because it enables
312 us to test the importance of size balance of gape of salamander and body of tadpoles for their
313 trophic interaction.

314

315 *Statistical analyses – morphology of both species*

316 We conducted three-way analysis of variance (ANOVA) on mean body width, mean body

317 size (snout-vent length) and mean relative body width of tadpoles and gape width, body size
318 (heart-vent length) and relative gape width of the largest salamanders at the end of 2nd period,
319 considering cannibalistic condition of salamanders, salamander population identity, frog
320 population identity and their interaction as explanatory variables. Then, we conducted a Tukey
321 post hoc test to examine how variables differ among treatments.

322

323 **Results**

324 *The trajectory of salamander survival*

325 Analyses of data including all of the treatments showed that the interaction between
326 salamander cannibalism condition and salamander population identity altered the trajectory of
327 salamander survival ($G_{\text{salamander.Erimo}} \times E_{\text{cannibalism}} \times \text{Period}$, $\beta = 0.037 \pm 0.011$ SE, $P = 0.001$, Fig. 1a
328 and 1b, Table S2a in Appendix S3). To biologically interpret the effect of this interaction, we
329 separately analyzed the No-cannibalism and Cannibalism treatment. Analysis on the No-
330 cannibalism treatment detected weak effects of period and an interactive effect of period and
331 salamander population identity. This indicates that survival rate was slightly increased in the late
332 experimental periods (Period, $\beta = 0.005 \pm 0.002$, $P = 0.012$), and that this tendency was slightly
333 weaker in Erimo salamanders ($G_{\text{salamander.Erimo}} \times \text{Period}$, $\beta = 0.005 \pm 0.002$, $P = 0.045$, Table S2b in
334 Appendix S3). The survival rate of salamanders remained constantly high under the No-
335 cannibalism treatment in both salamander populations regardless of tadpole population (Fig. 1b,
336 39-day survival > 94 %, Table S4 in Appendix S4).

337 For the Cannibalism treatment, salamander population identity and period jointly shaped
338 survival trajectory; survival of Erimo salamanders was lower in the early period but higher in the
339 later experimental periods, compared to that of Chitose salamander ($G_{\text{salamander.Erimo}} \times \text{Period}$, $\beta =$

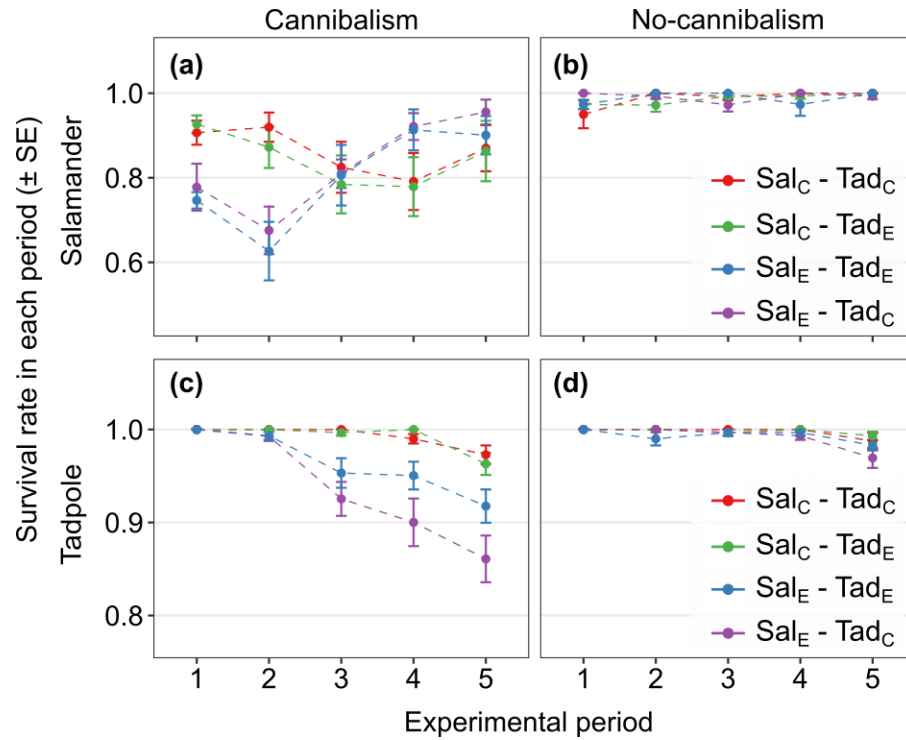
340 0.047±0.015, $P = 0.002$, Table S2c in Appendix S3, Fig. 1a). Tadpole population identity and
 341 their interactive terms had no effect on salamander survival ($P > 0.794$). Hence, irrespective of
 342 the tadpole population with which they coexist, the survival of Erimo salamanders was steeply
 343 reduced during the second period, whereas survival of Chitose salamanders was gradually
 344 reduced from the 1st to 4th period (Fig. 1a). The inter-population difference in survival trajectory
 345 resulted in a significant difference in net survival at the end of 2nd period, just before tadpole
 346 consumption started; in the Cannibalism treatment, the survival was 48.5±19.2 % in Erimo
 347 salamanders whereas 81.6±14.8 % in Chitose (pooled data for tadpole population, $P < 0.001$,
 348 Wilcoxon test). These results indicate that Erimo salamanders cannibalize conspecifics more
 349 frequently than Chitose salamanders particularly in the early experimental periods (Fig. 1a).

350

351 *The trajectory of tadpole survival*

352 Analyses of data including all treatments showed that the trajectory of tadpole survival was
 353 determined by the interaction among tadpole population identity, salamander cannibalism
 354 condition and salamander population identity (i.e. $G_{\text{tadpole.Erimo}} \times G_{\text{salamander.Erimo}} \times E_{\text{cannibalism}} \times$
 355 Period , $\beta = 0.005 \pm 0.002$, $P = 0.032$, Table S3a in Appendix S3, Fig. 1c and 1d). To dissect the
 356 higher-order interaction, we separately analyzed the No-cannibalism and Cannibalism treatment.
 357 The analysis of the No-cannibalism treatment detected a weak effect of the interaction between
 358 period and salamander population identity. Tadpole survival was slightly reduced in the presence
 359 of Erimo salamanders in the later experimental periods ($G_{\text{salamander.Erimo}} \times \text{Period}$, $\beta = -$
 360 0.002 ± 0.001 , $P = 0.035$, Table S3b, Fig. 1d). Still, tadpole survival was high in all treatments (>
 361 96%, see Table S4). In the Cannibalism treatment, interaction between salamander population
 362 identity and tadpole population identity influenced the survival trajectory. Tadpole survival was

363 lower when facing with Erimo salamanders ($G_{\text{salamander.Erimo}}$, $\beta = -0.027 \pm 0.005$, $P < 0.001$) and
 364 Erimo tadpole had higher survival when facing with Erimo salamanders ($G_{\text{tadpole.Erimo}} \times$
 365 $G_{\text{salamander.Erimo}}$, $\beta = 0.014 \pm 0.007$, $P = 0.044$); survival was greatly reduced through time when
 366 facing with Erimo salamanders ($G_{\text{salamander.Erimo}} \times \text{Period}$, $\beta = -0.015 \pm 0.002$, $P < 0.001$) but this
 367 reduction was weaker in Erimo tadpoles ($G_{\text{tadpole.Erimo}} \times G_{\text{salamander.Erimo}} \times \text{Period}$, $\beta = 0.009 \pm 0.003$,
 368 $P = 0.006$, Table S3c, Fig. 1c). In the Cannibalism treatment of Chitose salamander (where the
 369 potential predators did not appear at the end of 2nd period, see Results), tadpole survival was
 370 consistently high in both tadpole populations until the end of 4th period (i.e., day 32) (Fig. 1c). In
 371 contrast, in the Cannibalism treatment of Erimo salamander (where the potential predators
 372 appeared at the end of 2nd period, see Discussion), tadpole survival was continuously reduced
 373 after 2nd period (since day 18). Further, survival reduction was more severe in Chitose tadpoles
 374 than in Erimo (Fig. 1c). Additional analysis between the two treatments with Erimo salamanders
 375 under the Cannibalism treatment showed that tadpole survival was commonly reduced through
 376 time (Period, $\beta = -0.018 \pm 0.002$, $P < 0.001$), but this tendency was weaker in Erimo tadpoles than
 377 Chitose tadpoles ($G_{\text{tadpole.Erimo}} \times \text{Period}$, $\beta = 0.008 \pm 0.003$, $P = 0.007$, Appendix S5). The complex
 378 interaction in survival trajectory resulted in the variation in net survival of tadpole at the end of
 379 experiment across the treatments. Under the Cannibalism treatment in Erimo salamanders,
 380 tadpole survival was markedly reduced, especially for Chitose tadpoles (39 days survival rate:
 381 83.3 ± 9.1 % and 71.9 ± 13.8 % for Erimo and Chitose tadpole, respectively), whereas tadpole
 382 survival of both populations remained high in the other treatments (> 96 %, Table S4).



383

384 Fig. 1. Mean (\pm SE) of salamander (a and b) and tadpole (c and d) survival rates in each
 385 one-week experimental period. Salamander and tadpole mortality largely reflect
 386 cannibalism and tadpole consumption by salamanders, respectively. Trophic
 387 interactions occurred in Cannibalism treatment (a and c) while rarely occurred in No-
 388 cannibalism treatment (b and d). Colors of dots represent combination of salamander
 389 and tadpole population: Sal_C and Sal_E, Chitose and Erimo salamander; Tad_C and Tad_E,
 390 Chitose and Erimo tadpole.

391

392 *Morphology of salamanders and tadpoles*

393 We show results of three-way ANOVA on morphological variables of salamanders and
 394 tadpoles, while presenting the results of Tukey post hoc tests in Figures. Three-way ANOVA
 395 revealed that gape width of the largest salamander was larger in the Cannibalism treatment than

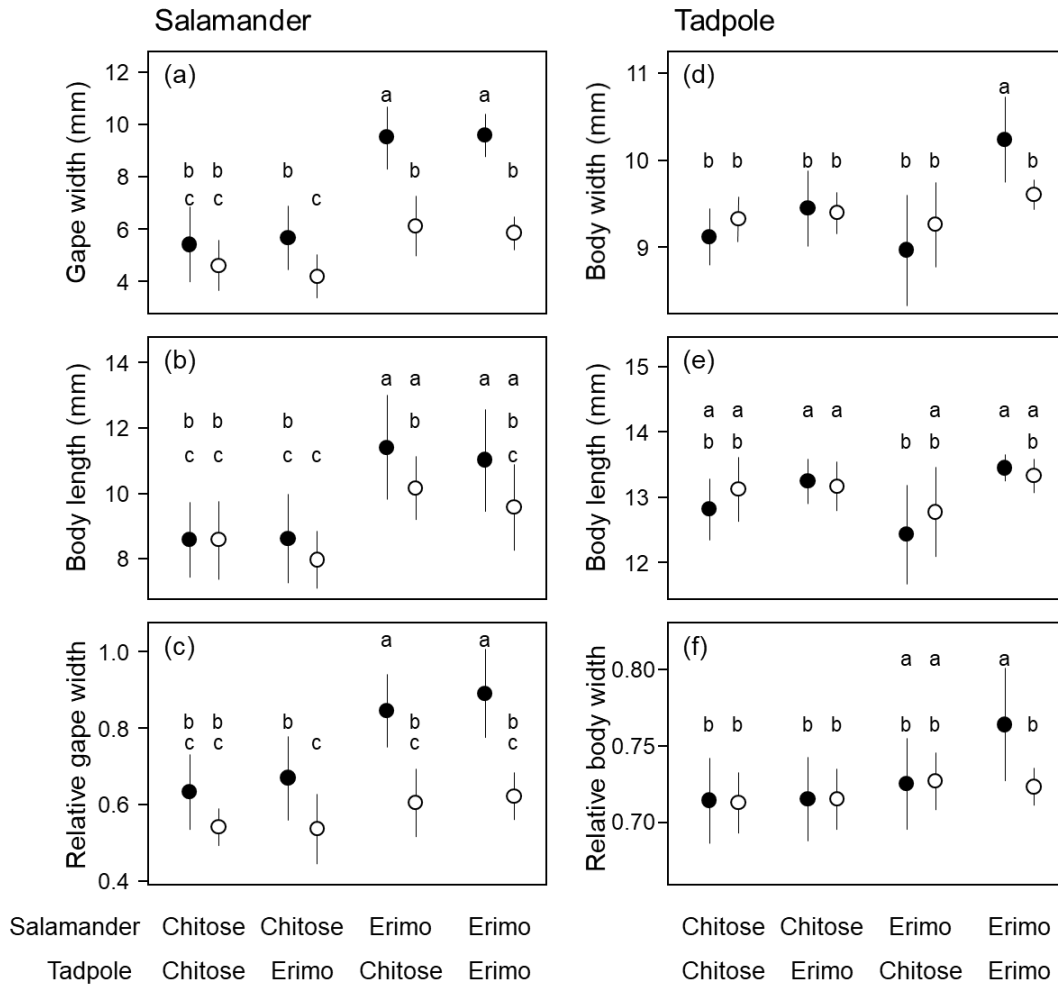
396 No-cannibalism treatment ($E_{\text{cannibalism}}$, $P < 0.001$, details are in Table S6a in Appendix S6) and
 397 was larger in the Erimo than the Chitose population ($G_{\text{salamander}}$, $P < 0.001$). Importantly, the
 398 interpopulation difference in gape width was enlarged in the Cannibalism treatment ($G_{\text{salamander}} \times$
 399 $E_{\text{cannibalism}}$, $P < 0.001$). Thus, Erimo salamander had the largest gape width in the Cannibalism
 400 treatment, regardless of the tadpole population with which they coexisted (Table S6a, Fig. 2a).
 401 Body length (heart-vent length) of the largest salamander was larger in the Cannibalism
 402 treatment than No-cannibalism treatment ($E_{\text{cannibalism}}$, $P = 0.006$, Table S6b) and also larger in
 403 Erimo population than Chitose population ($G_{\text{salamander}}$, $P < 0.001$). Hence, regardless of the
 404 tadpole populations, body length was largest in the Erimo population under cannibalistic
 405 condition (Fig. 2b). Relative gape width of the largest salamanders was larger in the Cannibalism
 406 treatment than No-cannibalism treatment ($E_{\text{cannibalism}}$, $P < 0.001$, Table S6c) and in Erimo
 407 population than Chitose population ($G_{\text{salamander}}$, $P < 0.001$) and, importantly, the interpopulation
 408 difference in relative gape width was greater when the salamanders were in the Cannibalism
 409 treatment than No-cannibalism treatment ($G_{\text{salamander}} \times E_{\text{cannibalism}}$, $P < 0.001$). Thus, irrespective
 410 of the tadpole populations, morphology of the largest salamanders was the most offensive in the
 411 Erimo population under cannibalistic condition (Fig. 2c). These results clearly indicated that
 412 offensive giant salamanders emerged more easily in Erimo population than Chitose population if
 413 the salamander hatchlings grew under cannibalistic condition.

414 Erimo tadpoles had a greater body width than Chitose tadpoles (G_{tadpole} , $P < 0.001$, Table
 415 S7a in Appendix S6). The interpopulation difference in body width of tadpoles was more
 416 apparent under cannibalistic salamander conditions than non-cannibalistic conditions ($G_{\text{tadpole}} \times$
 417 $E_{\text{cannibalism}}$, $P = 0.005$) and was also more apparent in the presence of Erimo salamanders than
 418 Chitose salamanders ($G_{\text{salamander}} \times G_{\text{tadpole}}$, $P = 0.004$). As a result, Erimo tadpoles exposed to

419 Erimo salamanders that grew under cannibalistic conditions had the widest body among
420 treatments (Fig. 2d). Erimo tadpoles had larger body length (Snout-vent length) than Chitose
421 tadpoles (G_{tadpole} , $P < 0.001$, Fig. 2e, Table S7b). The interpopulation difference in body length
422 of tadpoles was significantly larger in the presence of Erimo salamanders than Chitose
423 salamanders ($G_{\text{salamander}} \times G_{\text{tadpole}}$, $P = 0.021$) and tended to be larger when facing salamanders
424 under cannibalistic condition than non-cannibalistic condition ($G_{\text{tadpole}} \times E_{\text{cannibalism}}$, $P = 0.072$). In
425 the presence of Erimo salamanders, tadpoles exhibited bulgier phenotype (i.e., greater ratio of
426 body width to body length) than in the presence of Chitose salamanders ($G_{\text{salamander}}$, $P = 0.003$,
427 Table S7c). Erimo tadpoles tended to express the bulgy phenotype more strongly when Erimo
428 salamanders were under cannibalistic condition than under non-cannibalistic condition ($G_{\text{salamander}}$
429 $\times G_{\text{tadpole}} \times E_{\text{cannibalism}}$, $P = 0.086$). Thus, Erimo tadpoles subjected to the presence of cannibalistic
430 Erimo salamander exhibited the greatest defensive bulgy morph (Fig. 2f). These results indicate
431 that Erimo tadpoles attained the most defensive state (i.e., largest body width) when exposed to
432 Erimo salamanders that grew under cannibalistic situation. This was achieved by strongly
433 accelerating both isometric growth (i.e., body width increase proportional to body length) and
434 allometric growth (i.e., expression of bulgy phenotype) of body width.

435 At the end of the 2nd period, salamanders with the potential to consume tadpoles
436 (salamanders with gape width exceeding mean tadpole body width at 1.1 times [Nosaka et al.
437 2015]) appeared only in the Erimo population under cannibalistic condition with Chitose
438 tadpoles (appeared in 4 of 8 tanks, one individual on average). This result explains why the
439 strongest tadpole consumption by salamanders occurred in that treatment.

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Fig. 2. Functional trait size (a. gape width and d. body width), body length (b and e) and functional trait size relative to body size (c and f) of predator salamander and prey tadpole in each combinations of geographic populations (Chitose and Erimo) and cannibalism treatments (black and white dots represents Cannibalism and No-cannibalism treatment, respectively). The largest salamander in each tank, and all measurable tadpoles were measured at the end of 2nd period (day18), just before the tadpole consumption started.

450

451 **Discussion**

452 Since interaction strength depends on the functional traits of both interacting partners,
453 identifying factors affecting trait values is imperative to better understand the mechanisms
454 shaping geographic patterns of ecological interactions (Miner et al. 2005, Hendry 2016, Bassar et
455 al. 2017). Since individuals change their trait values according to their external and internal
456 conditions (Miner et al. 2005) and such reaction norms are under the genetic controls (Pigliucci
457 2005, Winterhalter and Mousseau 2007), trait values expressed by individuals is determined by
458 the specific combination of genotypes of the individuals and environmental conditions to which
459 the individuals are subjected. Importantly, trait values of an interacting species sometimes affects
460 the traits of their interacting partner, operating as an external factor selecting the latter (Kopp and
461 Tollrian 2003, Kishida et al. 2006, see also Kishida et al. 2010). This implies that interspecific
462 interaction strengths can be determined not only by the reaction norms of both interacting
463 partners but also through their interaction. Consequently, genotypes of both interacting species
464 and environmental conditions jointly and interactively shape geographic variation in interaction
465 strength (i.e., $G \times G \times E$ interactive effects on ecological interaction). Our experiment using
466 predatory salamander larvae (*H. retardatus*) and prey frog tadpoles (*R. pirica*) provided
467 experimental evidence supporting this prediction; temporal changes in survival of prey tadpoles
468 were determined by the combinations of geographic population of both tadpole and salamander
469 and growth condition of the predatory salamanders (Fig. 1).

470 Though in our experiment survival rate of salamanders remained quite high across all No-
471 cannibalism treatments for the whole experimental period (>94% at the end of experiment), it
472 was significantly reduced in the Cannibalism treatments. This result indicates that our

473 manipulation of size structure of salamander hatchlings successfully controlled a cannibalistic
474 interaction and therefore allows us to conclude that mortality of salamanders in this experiment
475 was largely caused by cannibalism. Importantly, intensity of salamander cannibalism differed
476 between the two salamander populations. In the Cannibalism treatment, although survival of
477 Chitose salamanders remained high, survival of Erimo salamanders strongly decreased in the
478 early periods. As a result, in the Cannibalism treatment, more than half of Erimo salamanders
479 died but only 20% of Chitose salamanders died due to cannibalism by the end of 2nd period. The
480 interpopulation difference in salamander cannibalism in the early period translated into
481 interpopulation variation in size of a functional trait of the cannibalistic salamander. At the end
482 of 2nd period, the largest individuals of Erimo salamanders in Cannibalism treatment had the
483 widest gape, regardless of the tadpole populations with which they coexisted.

484 Because Erimo salamanders had larger body length than Chitose salamanders across
485 treatments, the widest gape of the cannibalistic Erimo salamanders is partly explained by their
486 larger body size. In addition, the widest gape of the cannibalistic Erimo salamander is achieved
487 through greater allocation to enlargement of their gape than Chitose salamanders. The widest
488 gape of Erimo salamander in the Cannibalism treatment likely determined their significant
489 predatory effects on tadpoles in the subsequent period, because only the tadpoles in these
490 treatments suffered significant mortality. By the end of experiment, both Erimo and Chitose
491 tadpoles in the Cannibalism treatment of Erimo salamanders suffered significant mortality; while
492 tadpole mortality of both populations was negligibly low in the other treatments (less than 4%,
493 Table S4).

494 We found that predatory impacts of salamanders on tadpoles differed between tadpole
495 populations. Comparison of tadpole mortality between the two Cannibalism treatments of Erimo

496 salamander (i.e., Erimo tadpole-Erimo salamander-Cannibalism treatment and Chitose tadpole-
497 Erimo Salamander-Cannibalism treatment) revealed that Chitose tadpoles suffered predation
498 mortality 1.7 times higher than Erimo tadpoles (i.e., mortality rate of Chitose and Erimo tadpoles
499 were 28.1% and 16.6 %, respectively. Fig. 1). The largest Erimo salamanders had similar
500 morphology between the two treatments and Erimo tadpoles had 1.11 times wider body than
501 Chitose tadpoles in the Cannibalism treatment of Erimo salamanders just before salamander
502 predation on tadpoles started (i.e., end of 2nd period). Hence, differences in the mortality of
503 tadpoles between treatments were likely caused by variation in defensive performance between
504 tadpole populations. In support of this, the potential predatory salamanders whose gape was large
505 enough to swallow the coexisting tadpoles appeared only in the Cannibalism treatments of Erimo
506 salamanders faced with Chitose tadpoles. The difference in body width were the results of
507 greater isometric and allometric growth of Erimo tadpoles than Chitose tadpoles under the risky
508 situation with cannibalistic Erimo salamanders. Although body length of tadpoles as indicator of
509 body size was similar between the two populations at the beginning of the experiment, Erimo
510 tadpoles had longer body than Chitose tadpoles at the end of 2nd period and this trend was
511 intensified in the presence of Erimo salamander. This means that Erimo tadpoles accelerated
512 their ordinal growth in the riskier situation than Chitose tadpoles, implying that the Erimo
513 tadpoles more enlarged their body width isometrically with increase in their body size.
514 Furthermore, the fact that Erimo tadpoles in the Cannibalism treatment of Erimo salamanders
515 had the largest relative body width among the treatments (Fig. 2d) represents that Erimo tadpoles
516 expressed defensive bulgy phenotype more strongly than Chitose tadpoles under the riskiest
517 situation (facing with cannibalistic Erimo salamanders), implying that Erimo tadpoles more
518 enlarged their body width allometrically than Chitose tadpoles.

519 As discussed above, we argue that population specific reaction norms in morphological
520 traits of both predator and prey are the mechanisms underlying $G \times G \times E$ interplay in their
521 trophic interactions. Although population specific behavioural plasticity (e.g., development of
522 aggressiveness) can be an alternative mechanism (Bell and Stamps 2004), its importance is
523 negligible in our study since densities of experimental animals are relatively high. In the high-
524 density situation, the salamanders and tadpoles should have encountered frequently and thus
525 opportunity of salamanders to attack tadpoles would be frequent. In this case, occurrence of
526 successful predation events should have exclusively depended on the size balance between
527 salamander gape and tadpole body, since salamander is a representative swallowing-type
528 predator and the tadpoles are their large prey items (Takatsu and Kishida 2013). Importance of
529 predator-prey size balance in the trophic interaction was also evidenced by the additional
530 analysis (Appendix S7). Although effects of the treatment on tadpole mortality was significant in
531 our original analysis, the treatment effects became no longer significant by including the number
532 of potential predators as an additional predictor (Appendix S7). This result strongly suggests that
533 size balance between salamander gape and tadpole body is the exclusive mediator of the
534 treatment effects on the trophic interaction of salamanders on tadpoles.

535 Our experimental result predicts multiscale variation in interaction strength across the
536 geographic regions. In the comparison of interaction strength among the treatments with natural
537 population pairs of salamander and tadpole, Erimo pair (i.e., Erimo salamanders and Erimo
538 tadpoles) more differed in the interspecies interaction strength (tadpole survival) between
539 Cannibalism and No-cannibalism treatment than Chitose pair (i.e., Chitose salamanders and
540 Chitose tadpoles) (Appendix S8). That is, while Erimo and Chitose pairs showed similarly weak
541 interaction in No-cannibalism treatment, Erimo pair much more intensively interacted than

542 Chitose pair in Cannibalism treatment (Fig. 1c and d, see also Fig. S1). As a conditional factor
543 involving salamander cannibalism, our study featured size structure (i.e., hatch timing variation)
544 of salamander hatchlings, which should vary among ponds within geographic regions. Likewise,
545 other prospective conditional factors affecting cannibalism (e.g., conspecific density, alternative
546 prey abundance, structural complexity and predator presence: Fox 1975, Polis 2003, Kishida et
547 al. 2011) are generally spatially heterogenous within geographic regions. If within-region spatial
548 heterogeneity in those factors is equal across the geographic regions, interaction strength should
549 more vary in Erimo (no to strong interaction) than Chitose region (no to weak interaction).
550 Interaction strength variability across ponds may further create variability in pond communities
551 because amphibian larvae can strongly influence pond communities due to the far the largest
552 biomass among pond animals. Trophic interactions between the two larval amphibians can
553 impacts other species through alterations in their density and individual traits. Hence, population-
554 specific reaction norms of the two amphibian species can create regional variation in the
555 heterogeneity of their trophic interaction, and potentially further shape regional variation in β -
556 diversity of pond communities in nature.

557 More broadly, our study illuminates the importance of the interplay between environmental
558 conditions and genotypes of both interaction partners as the factors causing heterogeneity in the
559 strength of ecological interactions. If environmental heterogeneity is similar across geographic
560 regions, individuals with higher trait plasticity can exert more variable impact on interaction
561 strength than those with lower trait plasticity. In fact, compared to Chitose salamanders (i.e., less
562 plastic inducible offense — non- to poorly-offensive), Erimo salamanders (i.e., highly plastic
563 inducible offense — non- to highly-offensive) exerted more variable effects to intensify the
564 trophic interaction with tadpoles. Similarly, compared to Chitose tadpoles (i.e., less plastic

565 inducible defense — non- to poorly-defensive), Erimo tadpoles (i.e, greater plasticity — non- to
566 highly-defensive) exerted more variable effects to weaken the trophic interaction with
567 salamanders. As a result, while Erimo salamanders experienced larger variation in the
568 interspecies trophic interaction strength across the treatments than Chitose salamanders, Erimo
569 tadpoles experienced smaller variation in the interspecies trophic interaction strength across the
570 treatments. In trophic interactions, we can generally expect that higher plasticity in predator
571 inducible offense will increase the variation of trophic interaction strength, but higher plasticity
572 in prey inducible defense will decrease the variation of trophic interaction strength. This suggests
573 that the combination of reaction norms of both interaction partners can shape multiscale spatial
574 variation of interaction strength (i.e., heterogeneity in interaction strength within a geographic
575 region vary among different geographic regions). For example, if predators with highly plastic
576 inducible offense (i.e., from non- to highly-offensive) and prey with less plastic inducible
577 defense (i.e., from non- to poorly-defensive) cohabit in a geographic region, heterogeneity in
578 interaction strength within the region is expected to be considerably large (i.e., from no to quite
579 strong interaction). Conversely, if predators with less plastic inducible offense (i.e., from non- to
580 poorly-offensive) and prey with highly plastic inducible defense (i.e., from non- to highly-
581 defensive) cohabit in the other region, heterogeneity in interaction strength within this region is
582 expected to be quite small (i.e., from no to quite weak interaction).

583 Developmental reaction norms can be the target of natural selection (Urban 2008, 2010).
584 Geographic variation of differential developmental reaction norms of both salamander and
585 tadpole may be a result of differential coevolution history of the predator and prey amphibians.
586 While larger body size is necessary for tadpoles to avoid predation by giant salamanders, much
587 larger gape is required for salamanders to consume the defensive tadpoles with larger body

588 (Takatsu and Kishida 2015, Takatsu et al. 2017). Although only two geographic populations
589 were examined, the intrapopulation pattern of developmental reaction norms of predator and
590 prey follows an arms race scenario; Erimo salamanders had a higher ability to develop the
591 offensive phenotype (i.e., the salamander became giant with an enlarged gape) than Chitose
592 salamanders, and Erimo tadpoles grew more rapidly and expressed more defensive phenotypes
593 than Chitose tadpoles. This phenotypic pattern implies a geographic mosaic in coevolution with
594 Erimo as a coevolutionary hotspot (a region where coevolution is escalated) and Chitose as a
595 coldspot. In this coevolutionary scenario where evolutionary enhancement of antagonistic
596 phenotype expression imposes stronger selective pressure on the opponent, ecological
597 interactions may be stronger in coevolutionary hotspots than in coldspots. Therefore, describing
598 geographic patterns of developmental reaction norms of the two amphibians and testing the
599 coevolutionary hypothesis should advance our understanding of the mechanisms promoting
600 variation in the strength of ecological interactions.

601 Our study suggests that genetic variations in reaction norms of both species shape regional
602 variation in heterogeneity of the interaction strength within regions. Under different selection
603 regimes, populations can have evolved different reaction norms for their functional traits (e.g.,
604 plastic or fixed phenotypes along environmental gradient) (Kishida et al. 2007, Winterhalter and
605 Mousseau 2007). Investigating how reaction norms of functional traits for interacting partners
606 vary geographically and how the combination of reaction norms of the interacting partners
607 control their interactions is fruitful to disentangle complex geographic mosaics of ecological
608 interactions around the globe.

609

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617

618 **Literature cited**

619 Bassar, R. D., T. Simon, W. Roberts, J. Travis, and D. N. Reznick. 2017. The evolution of
620 coexistence: Reciprocal adaptation promotes the assembly of a simple community.
621 *Evolution* 71:373–385.

622 Bell, A. M., and J. A. Stamps. 2004. Development of behavioural differences between
623 individuals and populations of sticklebacks, *Gasterosteus aculeatus*. *Animal Behaviour*
624 68:1339–1348.

625 Bolnick, D. I., P. Amarasekare, M. S. Araújo, R. Bürger, J. M. Levine, M. Novak, V. H. W.
626 Rudolf, S. J. Schreiber, M. C. Urban, and D. A. Vasseur. 2011. Why intraspecific trait
627 variation matters in community ecology. *Trends in Ecology and Evolution* 26:183–192.

628 Fox, L. R. 1975. Cannibalism in natural populations. *Annual Review of Ecology and Systematics*
629 6:87–106.

630 Gerber, S., G. J. Eble, and P. Neige. 2008. Allometric space and allometric disparity: A
631 developmental perspective in the macroevolutionary analysis of morphological disparity.
632 *Evolution* 62:1450–1457.

633 Hendry, A. P. 2016. *Eco-evolutionary Dynamics*. Princeton University Press, Princeton, NJ.

- 634 Hiltunen, T., and L. Becks. 2014. Consumer co-evolution as an important component of the eco-
635 evolutionary feedback. *Nature Communications* 5:1–8.
- 636 Kasada, M., M. Yamamichi, and T. Yoshida. 2014. Form of an evolutionary tradeoff affects eco-
637 evolutionary dynamics in a predator–prey system. *Proceedings of the National Academy of*
638 *Sciences* 111:16035–16040.
- 639 Kishida, O., Z. Costa, A. Tezuka, and H. Michimae. 2014. Inducible offences affect predator-
640 prey interactions and life-history plasticity in both predators and prey. *Journal of Animal*
641 *Ecology* 83:899–906.
- 642 Kishida, O., Y. Mizuta, and K. Nishimura. 2006. Reciprocal phenotypic plasticity in a predator-
643 prey interaction between larval amphibians. *Ecology* 87:1599–1604.
- 644 Kishida, O., and K. Nishimura. 2004. Bulgy tadpoles: Inducible defense morph. *Oecologia*
645 140:414–421.
- 646 Kishida, O., A. Tezuka, A. Ikeda, K. Takatsu, and H. Michimae. 2015. Adaptive acceleration in
647 growth and development of salamander hatchlings in cannibalistic situations. *Functional*
648 *Ecology* 29:469–478.
- 649 Kishida, O., G. C. Trussell, A. Mougi, and K. Nishimura. 2010. Evolutionary ecology of
650 inducible morphological plasticity in predator-prey interaction: Toward the practical links
651 with population ecology.
- 652 Kishida, O., G. C. Trussell, and K. Nishimura. 2007. Geographic variation in a predator-induced
653 defense and its genetic basis. *Ecology* 88:1948–1954.
- 654 Kishida, O., G. C. Trussell, and K. Nishimura. 2009a. Top-down effects on antagonistic
655 inducible defense and offense. *Ecology* 90:1217–1226.

- 656 Kishida, O., G. C. Trussell, K. Nishimura, T. Ohoushi, and T. Ohgushi. 2009b. Inducible
657 defenses in prey intensify predator cannibalism. *Ecology* 90:3150–3158.
- 658 Kishida, O., G. C. Trussell, A. Ohno, S. Kuwano, T. Ikawa, and K. Nishimura. 2011. Predation
659 risk suppresses the positive feedback between size structure and cannibalism. *Journal of*
660 *Animal Ecology* 80:1278–1287.
- 661 Kopp, M., and R. Tollrian. 2003. Reciprocal phenotypic plasticity in a predator-prey system:
662 Inducible offences against inducible defenses? *Ecology Letters* 6:742–748.
- 663 McGill, B. J., B. J. Enquist, E. Weiher, and M. Westoby. 2006. Rebuilding community ecology
664 from functional traits. *Trends in Ecology and Evolution* 21:178–185.
- 665 Michimae, H. 2006. Differentiated phenotypic plasticity in larvae of the cannibalistic salamander
666 *Hynobius retardatus*. *Behavioral Ecology and Sociobiology* 60:205–211.
- 667 Michimae, H., and M. Wakahara. 2002. Variation in cannibalistic polyphenism between
668 populations in the Salamander *Hynobius retardatus*. *Zoological science* 19:703–707.
- 669 Miller, T. E. X., and V. H. W. Rudolf. 2011. Thinking inside the box: Community-level
670 consequences of stage-structured populations. *Trends in Ecology and Evolution* 26:457–
671 466.
- 672 Miner, B. G., S. E. Sultan, S. G. Morgan, D. K. Padilla, and R. A. Relyea. 2005. Ecological
673 consequences of phenotypic plasticity. *Trends in Ecology and Evolution* 20:685–692.
- 674 Nakano, S., H. Kuhara, and N. Miyasaka. 1999. Terrestrial-aquatic linkages : Riparian arthropod
675 inputs alter trophic cascades in a stream food web. *Ecology* 80:2435–2441.
- 676 Nosaka, M., N. Katayama, and O. Kishida. 2015. Feedback between size balance and
677 consumption strongly affects the consequences of hatching phenology in size-dependent
678 predator-prey interactions. *Oikos* 124:225–234.

- 679 Persson, L., P. A. Amundsen, A. M. De Roos, A. Klemetsen, R. Knudsen, and R. Primicerio.
680 2007. Culling prey promotes predator recovery - Alternative states in a whole-lake
681 experiment. *Science* 316:1743–1746.
- 682 Pigliucci, M. 2005. Evolution of phenotypic plasticity: Where are we going now? *Trends in*
683 *Ecology and Evolution* 20:481–486.
- 684 Polis, G. A. 2003. The evolution and dynamics of intraspecific predation. *Annual Review of*
685 *Ecology and Systematics* 12:225–251.
- 686 R Development Core Team. 2019. R: A language and environment for statistical computing. R
687 Foundation for Statistical Computing, Vienna, Austria.
- 688 Raffard, A., F. Santoul, J. Cucherousset, and S. Blanchet. 2018. The community and ecosystem
689 consequences of intraspecific diversity: a meta-analysis. *Biological Reviews* 1:648–661.
- 690 Des Roches, S., D. M. Post, N. E. Turley, J. K. Bailey, A. P. Hendry, M. T. Kinnison, J. A.
691 Schweitzer, and E. P. Palkovacs. 2018. The ecological importance of intraspecific variation.
692 *Nature Ecology and Evolution* 2:57–64.
- 693 Satterthwaite, F. E. 1946. An approximate distribution of estimates of variance components.
694 *Biometrics Bulletin* 2:110–114.
- 695 Schielzeth, H. 2010. Simple means to improve the interpretability of regression coefficients.
696 *Methods in Ecology and Evolution* 1:103–113.
- 697 Schmitz, O. J., R. W. Buchkowski, K. T. Burghardt, and C. M. Donihue. 2015. Functional traits
698 and trait-mediated interactions: Connecting community-level interactions with ecosystem
699 functioning. Pages 319–343 *Advances in Ecological Research*. First edition. Elsevier Ltd.
- 700 Schneider, C. A., W. S. Rasband, and K. W. Eliceiri. 2012. NIH Image to ImageJ: 25 years of
701 image analysis. *Nature Methods* 9:671–675.

- 702 Sih, A., G. Englund, and D. Wooster. 1998. Emerging impacts of multiple predators on prey.
703 *Trends in Ecology and Evolution* 13:350–355.
- 704 Sinclair, A. R. E., S. Mduma, and J. S. Brashares. 2003. Patterns of predation in a diverse
705 predator-prey system. *Nature* 425:288–290.
- 706 Soluk, D. A., and N. C. Collins. 1988. Synergistic interactions between fish and stoneflies:
707 facilitation and interference among stream predators. *Oikos* 52:94–100.
- 708 Takatsu, K., and O. Kishida. 2013. An offensive predator phenotype selects for an amplified
709 defensive phenotype in its prey. *Evolutionary Ecology* 27:1–11.
- 710 Takatsu, K., and O. Kishida. 2015. Predator cannibalism can intensify negative impacts on
711 heterospecific prey. *Ecology* 96:1887–1898.
- 712 Takatsu, K., and O. Kishida. In press. Enhanced recruitment of larger predators in the presence
713 of large prey. *Journal of Animal Ecology*.
- 714 Takatsu, K., V. H. W. Rudolf, and O. Kishida. 2017. Giant cannibals drive selection for
715 inducible defense in heterospecific prey. *Biological Journal of the Linnean Society*
716 120:675–684.
- 717 Urban, M. C. 2008. Salamander evolution across a latitudinal cline in gape-limited predation
718 risk. *Oikos* 117:1037–1049.
- 719 Urban, M. C. 2010. Microgeographic adaptations of spotted salamander morphological defenses
720 in response to a predaceous salamander and beetle. *Oikos* 119:646–658.
- 721 Ushio, M., C. H. Hsieh, R. Masuda, E. R. Deyle, H. Ye, C. W. Chang, G. Sugihara, and M.
722 Kondoh. 2018. Fluctuating interaction network and time-varying stability of a natural fish
723 community. *Nature* 554:360–363.

- 724 Winterhalter, W. E., and T. A. Mousseau. 2007. Patterns of phenotypic and genetic variation for
725 the plasticity of diapause incidence. *Evolution* 61:1520–1531.
- 726 Yamaguchi, A., K. Takatsu, and O. Kishida. 2016. Contacts with large, active individuals
727 intensify the predation risk of small conspecifics. *Ecology* 97:3206–3218.
- 728 Yamamichi, M., T. Klauschies, B. E. Miner, and E. van Velzen. 2019. Modelling inducible
729 defenses in predator–prey interactions: assumptions and dynamical consequences of three
730 distinct approaches. *Ecology Letters* 22:390–404.
- 731 Yoshida, T., L. E. Jones, S. P. Ellner, G. F. Fussmann, and N. G. Hairston. 2003. Rapid
732 evolution drives ecological dynamics in a predator-prey system. *Nature* 424:303–6.
- 733