

1 Shade-tree rehabilitation in vanilla agroforests is yield neutral and 2 may translate into landscape-scale canopy cover gains

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37 **Abstract**

38 Agroforestry can contribute to an increase in tree cover in historically forested tropical
39 landscapes with associated gains in biodiversity and ecosystem functioning, but only if
40 established on open land instead of inside forest. However, trade-offs between shade and
41 crop yields are common across many agroforestry crops, driving shade-tree loss in forest-
42 derived agroforests and hindering tree rehabilitation in open-land-derived agroforests. To
43 investigate whether this common dynamic plays a role in vanilla agroforests, we studied 209
44 vanilla agroforests along an 88-year chronosequence in Madagascar and used remotely-
45 sensed canopy cover data to investigate tree rehabilitation in the agricultural landscape. We
46 found yields to vary widely but independently of canopy cover and land-use history (forest-
47 vs. open-land-derived), averaging at $154.6 \text{ kg ha}^{-1} \text{ yr}^{-1}$ ($SD = 186.9$). Furthermore, we found
48 that open-land-derived vanilla agroforests gained 32.6% canopy cover over 60 years,
49 whereas forest-derived agroforests only gained 14.2%. Canopy cover increased also at the
50 landscape scale: Areas in the agricultural landscape with medium initial canopy cover gained
51 6.4% canopy cover from 2000 to 2010, but areas with high initial canopy cover lost canopy
52 cover. These opposing trends suggest tree rehabilitation across areas covered by vanilla
53 agroforests, whereas remnant forest fragments in the agricultural landscape were
54 transformed or degraded. Overall, forest-dependent ecosystem functions may thus suffer
55 while functions provided by areas with medium canopy cover may benefit. Our results
56 suggest that yield-neutral tree rehabilitation through agroforestry could, if coupled with
57 effective forest protection, provide a win-win situation for ecosystem functions and
58 agricultural production in smallholder-dominated agricultural landscapes.

59

60 **Justification statement**

61 Agroforestry is promoted as a way to restore trees in historically forested open landscapes,
62 but trade-offs between shade tree cover and yields are common across many tropical
63 agroforestry crops like coffee or cacao. Using data from an 88-year chronosequence with
64 209 agroforests, we show that such trade-offs do not exist for vanilla, an important cash
65 crop in the biodiversity hotspot Madagascar. This offers an opportunity for win-win
66 situations between ecosystem functions and biodiversity on the one hand, and farmers
67 income on the other. Furthermore, we show that vanilla agroforests gain canopy cover over
68 time as trees rehabilitate on land formerly used for shifting cultivation. Local-scale gains
69 may also translate to landscape-scale, where canopy cover increased within the agricultural
70 landscape. This finding highlights opportunities for tree rehabilitation on farmland that may
71 increase connectivity between increasingly fragmented Malagasy forests. Yield-neutral tree
72 rehabilitation in vanilla agroforests has thus the potential to restore ecosystem functions
73 and services on the plot- and landscape-scale while offering an economically viable option
74 for smallholder farmers.

75 **Introduction**

76 Rehabilitation of historically forested open land is widely advocated to re-establish
77 connectivity and ecosystem functions in tropical rainforest landscapes (Bastin et al., 2019;
78 Chazdon, 2003). To date, governments and institutions have made pledges to restore 140
79 million hectares of land in the tropics (Brancalion et al., 2019). However, realizing those
80 pledges could compete with food security, questioning how feasible (Eitelberg et al., 2016)
81 and indeed desirable (Holl & Brancalion, 2020) their fulfilment is. In this light, agroforests
82 may provide an opportunity to combine trees with agricultural production on the same land
83 (De Beenhouwer et al., 2016; FAO, 2017). Particularly agroforests that are established on
84 historically forested open-land hold a large potential, because open-land-derived
85 agroforests rehabilitate open land (Martin, Osen, et al., 2020). Thereby, the focus lies on the
86 rehabilitation of ecosystem functions, without necessarily restoring ecological integrity
87 (Chazdon et al., 2016). Agroforests that are established inside forests, on the other hand,
88 typically contribute to forest degradation, thus hampering ecosystem functioning (Martin,
89 Osen, et al., 2020).

90 Nonetheless, trade-offs between shade cover and yields are common across many key
91 agroforestry crops (Tscharntke et al., 2011), limiting their potential to contribute to tree
92 rehabilitation in tropical rainforest landscapes. Such shade-yield trade-offs are exemplified
93 in coffee and cacao agroforests (Blaser et al., 2018; Steffan-Dewenter et al., 2007), where
94 felling trees is typically beneficial to farmers aiming at optimizing yields. Finding a balance
95 between ecosystem services, biodiversity and profitability thus requires targeted incentives
96 or subsidies (Tscharntke et al., 2014). In their absence, a decrease in canopy cover and tree
97 height over time commonly occurs (Tscharntke et al., 2011), but time series or

98 chronosequences, which are necessary to identify trends, are rare (see Nijmeijer et al.
99 (2019) for an exception). Finding farming techniques or crops where such trade-offs do not
100 inherently occur would, on the other hand, offers an opportunity to profitably farm crops in
101 high-shade agroforestry systems without the need for further incentives.

102 One candidate crop where shade-yield trade-offs are currently unknown is vanilla. When
103 farmed in agroforestry systems, the vanilla orchid (*Vanilla planifolia*) is typically hung up on
104 'tutor trees' which give support to the non-woody vine (Correll, 1953). Vanilla flowers are
105 then hand pollinated and green pods are harvested nine months later. The green pods are
106 subsequently cured, thereby developing their distinct flavour and black colouration while
107 losing roughly 80% of their weight (Havkin-Frenkel & Belanger, 2018). The resulting black
108 vanilla has strongly increased in price over recent years, triggering the expansion of vanilla
109 farming (Hänke et al. 2018; Llopis et al. 2019, Supplementary Material Figure 1).

110 In north-eastern Madagascar, vanilla is the main cash crop for smallholder farmers (Hänke
111 et al., 2018) who produce 40% of the world's vanilla (FAO, 2020). Here, vanilla is almost
112 exclusively produced in rather extensively managed agroforestry systems, partly in contrast
113 to other production areas such as La Réunion or Mexico, where shade houses are common
114 (Havkin-Frenkel & Belanger, 2018). Other prominent land uses in the Malagasy vanilla
115 region include remnant forest fragments, irrigated rice paddies and hill rice fields with the
116 associated herbaceous and woody fallow, that form part of the shifting cultivation cycle
117 locally known as *tavy* (Martin, Andriafanomezantsoa, et al., 2020; Styger et al., 2007). The
118 first cycle of shifting cultivation, where fire is used to convert forest into hill rice fields, is the
119 main reason for forest loss in the region (Schüßler et al., 2020; Zaehring et al., 2015). This
120 dynamic is consistent with trends across most of Africa, but contrasts with trends in the
121 remaining tropics (Curtis et al., 2018; van Vliet et al., 2012).

122 Vanilla agroforests may be established inside forest fragments or on open fallow land,
123 thereby differing in land-use history (Martin, Osen, et al., 2020). Forest-derived vanilla
124 agroforests degrade the forest they are established in but will typically be superior to
125 shifting cultivation (Martin, Osen, et al., 2020), i.e. the replacement of forest with hill rice
126 cultivation, for ecosystem function and biodiversity. Open-land-derived agroforests may
127 instead rehabilitate land formerly under hill rice cultivation, as their establishment stops the
128 re-occurring fires which characterize the shifting hill rice cultivation system (Holloway, 2004;
129 Styger et al., 2007). In north-eastern Madagascar, 30% of vanilla agroforests are forest-
130 derived while 70% are open-land-derived (Hänke et al., 2018), further underlining the
131 rehabilitation opportunity offered by open-land-derived agroforestry. The potential of
132 Madagascar to contribute to tree rehabilitation is also recognized in the recent study by
133 Brancalion et al. (2019), who attribute the 4th largest restoration opportunity area (in terms
134 of benefits and feasibility) of lowland tropical rainforest to Madagascar. Simultaneously, the
135 country is a biodiversity hotspot (Myers et al., 2000), exacerbating the need for both
136 effective biodiversity conservation within the existing protected areas as well as restoration
137 within the agricultural landscape.



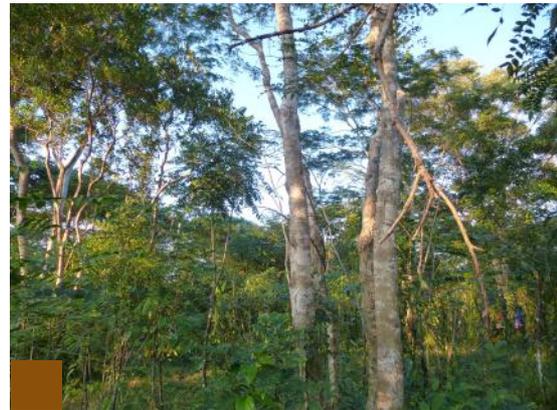
Tropical forest fragment



Forest-derived vanilla agroforest



Open land (Woody fallow)



Open-land-derived vanilla agroforest



Vanilla pied (unit of vanilla vine and tutor tree) with flower and green pods



Agricultural landscape in north-eastern Madagascar

138

139 *Figure 1: Top row: Forest-derived vanilla agroforests are directly established inside forest.*
 140 *Middle row: Open-land-derived vanilla agroforests are established on open land, typically*
 141 *woody fallow. Bottom row: Vanilla pied and agricultural landscape in north-eastern*
 142 *Madagascar where the study took place. Colour labels indicate contrasting land-use history*
 143 *of vanilla agroforests and are used throughout the manuscript. All photos by the authors.*

144 In this study, we aim to 1) understand how land-use history, canopy cover, agroforest age,
145 planting density and precipitation influence vanilla yields, 2) assess tree rehabilitation
146 dynamics across vanilla agroforests of different age and of contrasting land-use history, and
147 3) investigate how tree rehabilitation within vanilla agroforests may transform the
148 landscape as a whole. To this end, we assessed vanilla yields, canopy cover and canopy
149 height in 209 vanilla agroforests of contrasting land-use history and of different age (0 – 88
150 years), thus representing an 88-year chronosequence. Subsequently, we used remotely
151 sensed canopy cover data to study canopy cover change on the landscape-scale.

152 **Methods**

153 **Study region**

154 The SAVA region of north-eastern Madagascar is the historic (Correll, 1953) and current
155 (Hänke et al., 2018) center of global vanilla production and a biodiversity hotspot (Brown et
156 al., 2016; Myers et al., 2000). Mean annual temperature is 23.7 °C and annual rainfall
157 averages at 2238 mm (Mean across 209 focal agroforests; data from CHELSA Climatologies
158 (Karger et al., 2017)). The potential natural vegetation is tropical rainforest (Vieilledent et
159 al., 2018), but only 35% forest cover remains in the SAVA region (Ferreira Arruda, 2018).

160 **Data collection in agroforests**

161 We conducted our field studies in a total of 115 forest-derived and 94 open-land-derived
162 agroforests (209 in total) owned by 152 households across 14 villages (see Supplementary
163 Materials for village and agroforest selection). We collected field data between July and
164 October 2018 after the 2018 vanilla harvest.

165 During visits to the agroforest, we asked vanilla agroforest owners in native Malagasy about
166 1) the realized yield of green vanilla in 2017 and 2018 [kg agroforest⁻¹], 2) estimated green
167 vanilla theft from the agroforest before harvest in 2017 and 2018 [kg agroforest⁻¹], 3) the
168 number of *pieds* (combination of vanilla vine and tutor tree; Figure 1) in the agroforest, 4)
169 the year in which the agroforest was established, and 5) whether the agroforest was forest-
170 or open-land-derived (sensu Martin, Osen, et al. 2020). Vanilla yields are commonly
171 reported as the weight of green rather than black pods, since green pod weight is
172 independent of the curing technique. We subsequently added estimated theft to the
173 realized yields as we were interested in the productivity of the agroforests rather than the
174 farmers' income. We measured agroforest size using handheld GPS devices and applied a
175 slope correction (based on the digital surface model 'ALOS World 3D' (Japan Aerospace
176 Exploration Agency, 2018)) to account for different steepness of the terrain. By combining
177 yield data and the slope-corrected agroforest size, we calculated mean green vanilla yield
178 per hectare [kg ha⁻¹ year⁻¹] across the two years for further analysis. Based on slope-
179 corrected agroforest size and number of *pieds*, we calculated planting density [*pieds* ha⁻¹].
180 We used tablets to assess canopy cover as photos from mobile devices have been found to
181 be an adequate, cheap and fast technique to assess canopy cover (Bianchi et al., 2017).
182 Observers held a tablet (Lenovo YT3-850F) above their head (circa 190 cm) and used the
183 built-in camera (Lenovo 5C28C02840) with the standard lens and auto-exposure to take a
184 photo in azimuthal direction. We repeated this procedure at nine locations per plot (see
185 Supplementary Materials), resulting in 1881 photos from 209 agroforests. We then
186 classified all photos into vegetation/sky using the R-Package *caiman* (Diaz and Lencinas
187 2015; more details on canopy cover classification in Supplementary Materials) and
188 calculated mean canopy cover across all 9 photos to derive one value per agroforest.

189 Additionally, the observer estimated the highest point of vegetation above each camera
190 position, enabling us to calculate the mean canopy height across 9 locations for each
191 agroforest.

192 Some farmers did not know the number of *pieds* respectively the year of establishment of
193 their agroforest, leading to missing data for planting density and agroforest age in 8
194 respectively 3 cases (out of 209). We imputed this data for the linear mixed effect models
195 using the mean of each respective variable.

196 **Data extraction from raster layers**

197 To investigate how precipitation and temperature influenced vanilla yields, we extracted
198 annual mean temperature and annual precipitation for each agroforest from the CHELSA
199 climatologies (Karger et al., 2017) using the plot center as a reference point. Due to the
200 strong correlation of annual mean temperature and annual precipitation (-0.76, Pearson
201 correlation coefficient), we only used annual precipitation for further analysis. Analogously,
202 we obtained the elevation of each agroforest from the digital surface model 'ALOS World
203 3D' (Japan Aerospace Exploration Agency, 2018). Lastly, we extracted the percentage
204 landscape forest cover in a radius of 250 m around plot centres using published binary
205 forest cover data for the year 2017 (Vieilledent et al., 2018).

206 **Analysis of vanilla yields, canopy cover and canopy height**

207 We used linear-mixed effects models to analyse variation in vanilla yields, canopy cover and
208 canopy height, with 'household' (owner of agroforest, N= 152) and 'village' (N=14) as
209 random effects in all models. In a first model, we assessed the variation of green vanilla
210 yield [kg ha^{-1}] in relation to land-use history (forest vs. open-land-derived; coded as 1 vs. 0),
211 canopy cover, age of agroforest, planting density and annual precipitation. To reach

212 normality of model residuals, we applied a Box-Cox transformation to the response variable
213 (Box & Cox, 1964). We determined a lambda of 0.25 to be suitable for the transformation
214 using the *boxcox* function of the R-package *mass* version 7.3.51.4 (Ripley et al., 2013). Due
215 to the highly right-skewed nature of the age and planting density data, we square root
216 transformed these two variables.

217 In a second and third model, we assessed factors influencing canopy cover (untransformed)
218 and canopy height (Box-Cox-transformed with lambda 0.35), respectively. We used land-use
219 history, age of agroforest, elevation, landscape forest cover and planting density as
220 explanatory variables. Again, we square root transformed the age and planting density data.
221 In all models, we additionally included interactions between land-use history and all
222 explanatory variables to test whether responses would differ between forest- and open-
223 land-derived agroforests. In the yield model, none of the interactions were significant,
224 prompting us to present a reduced model without interactions. In the canopy height model,
225 only the interaction between age and land-use history was significant at the $p < 0.05$ level.
226 We thus only kept this interaction in the reduced model. In the canopy cover model, none
227 of the interactions were significant but we kept the interaction between land-use history
228 and age despite a p-value of 0.109 because we aimed at comparable and similarly complex
229 models for canopy height and canopy cover. All models are presented in full and reduced
230 (i.e. final) form in the Supplementary Materials (Supplementary Materials Table 1-3).

231 We fitted all models using the R-Package *lme4* version 1.1.21 (Bates, 2014) and scaled all
232 explanatory and response variables to zero mean and unit variance, allowing for direct
233 comparison of effect sizes within and across models (Harrison et al., 2018). We calculated
234 marginal and conditional R^2 -values for all models (Nakagawa & Schielzeth, 2013). We used

235 QQ-plots to assess normality of model residuals and tested for variable inflation; none of
236 the models had significant deviations in the QQ-plots or variable inflation values above 1.5.
237 To visualize the models, we calculated estimated marginal means and their 95% confidence
238 intervals using the R-Package *emmeans* version 1.4.5 (Length et al., 2018). We further back-
239 transformed the estimated marginal means to the original distributions to facilitate the
240 interpretation of model results.

241 **Analysis of canopy cover dynamics in the agricultural landscape**

242 We used remotely sensed canopy cover data to explore how observed tree rehabilitation
243 within agroforests translated to the landscape scale. We obtained canopy cover data for the
244 year 2000 and 2010 from a Landsat-derived product of continuous canopy cover values with
245 30 m resolution (Hansen et al., 2013). Using the *raster* R-package version 3.0.12 (Hijmans et
246 al., 2019), we subtracted the 2000 layer from the 2010 layer to obtain a new raster layer
247 with tree cover gains and losses, respectively (change of canopy cover between 2000 and
248 2010 [%]). We then excluded areas that fell in the sea and restricted both layers to an area
249 of 2 km around the centres of 60 focal villages for which we knew that vanilla farming was
250 common and from which we selected the villages for the plot-based part of this study
251 (Village selection described in Hänke et al., 2018). We chose 2 km because agroforests in
252 this range will typically belong to the focal village (*personal observation*). We then fitted a
253 generalized additive mixed model (GAMM) using the R-package *mgcv* version 1.8-28 (Wood,
254 2012) to evaluate how initial canopy cover in the year 2000 determined the change in
255 canopy cover from 2000 to 2010. We included 'village' as a random effect and also included
256 longitude and latitude of each raster cell as random effects to control for spatial
257 autocorrelation.

258 We analysed all data in R version 3.6.0 (R Core Team, 2019). The underlying data and R-code
259 are publicly available (see Data availability statement).

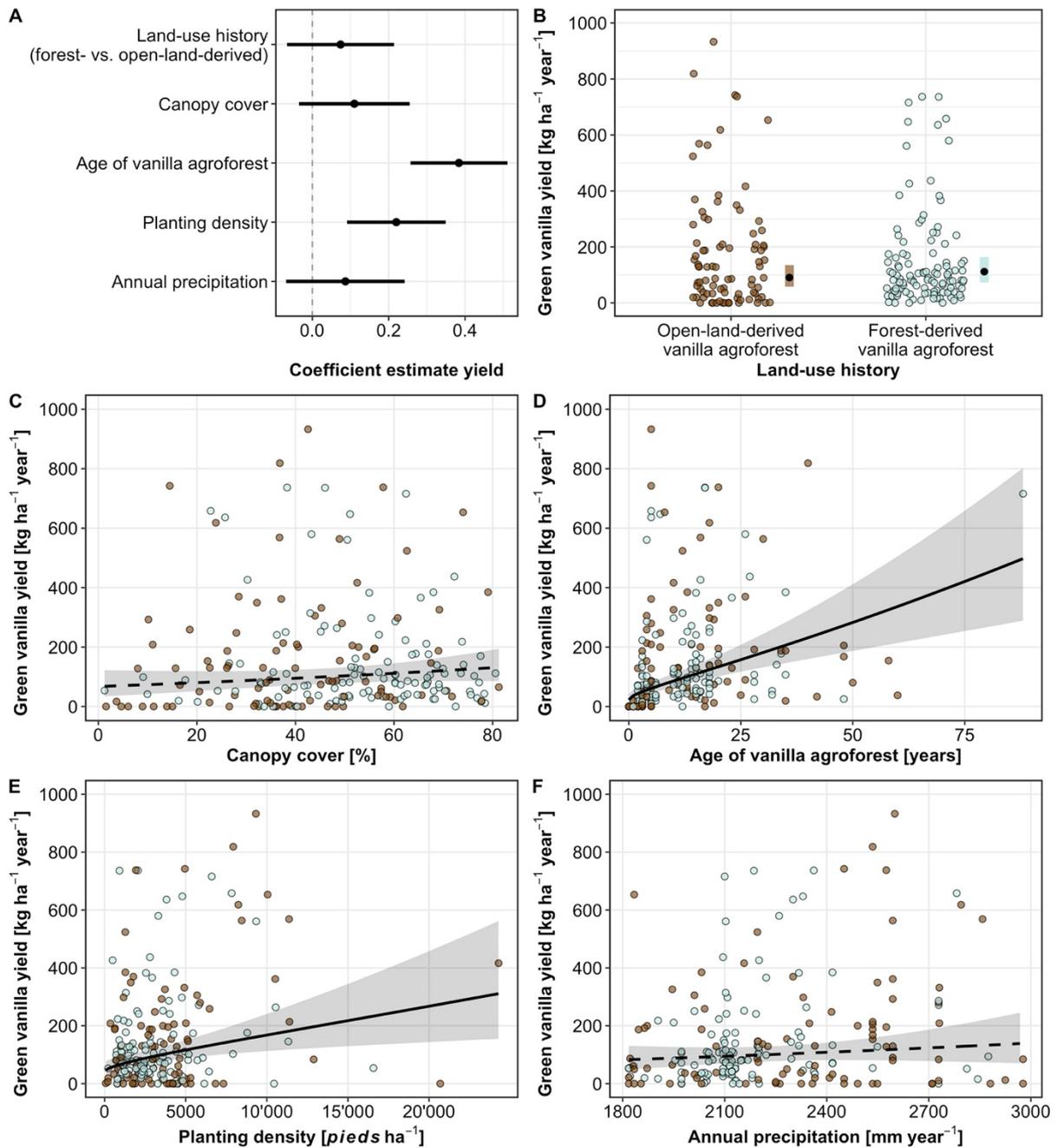
260 **Results**

261 **Determinants of vanilla yields**

262 Green vanilla yield per *pied* varied ranged from 0 – 860 g *pied*⁻¹ year⁻¹ with an average of
263 69.9 g *pied*⁻¹ year⁻¹ (*SD* = 112.3; N = 209 agroforests; mean from 2017 and 2018). Note that
264 this estimate includes *pieds* without any yield as it is calculated by dividing the total yield by
265 the number of *pieds* in each agroforest. Similarly, green vanilla yields differed strongly
266 across agroforests, ranging from 0 - 932.7 kg ha⁻¹ year⁻¹ with an average of 154.6 kg ha⁻¹
267 year⁻¹ (*SD* = 186.9; N= 209 agroforests; mean of 2017 and 2018). Using farmgate vanilla
268 prices for the year 2017 (Hänke et al., 2018), this average yield translates into gross earnings
269 of 4684 € ha⁻¹. The difference in green vanilla yield per ha between the two years was small
270 (2017: 158.8 kg ha⁻¹ (*SD* = 200.1); 2018: 150.2 kg ha⁻¹ (*SD* = 202.6; N = 209; p = 0.6423).
271 Farmers reported green vanilla theft in 26 agroforests (12.4%) for 2017 and in 25
272 agroforests (12.0%) for 2018. Farmers who reported theft, stated that they lost on average
273 9.15 kg (*SD* = 15.3) green vanilla per agroforest in 2017 and 8.72 kg (*SD* = 8.7) per agroforest
274 in 2018.

275 Our yield model (Figure 2, SM Table 1) revealed that vanilla yields varied independently of
276 land-use history, i.e. whether an agroforest was forest- or open-land-derived. Yields were
277 furthermore not significantly correlated to canopy cover and annual precipitation. Yields
278 were positively correlated with agroforest age and planting density. Overall, the marginal
279 R²-value of the model was 0.216 while the conditional R²-value was 0.450. The difference

280 between the two values was mainly driven by the random intercept variance for the random
 281 effect 'household' ($\tau_{00} = 0.26$); the random intercept variance for the random effect 'village'
 282 was negligible ($\tau_{00} = 0.03$)



283

284 *Figure 2: Results of a linear mixed effect model explaining green vanilla yield [kg ha⁻¹] across*
 285 *209 agroforests. A: Scaled effect plot of the reduced yield model for all five predictors. B-F:*
 286 *Green vanilla yields as a function of land-use history (B), canopy cover [%] (C), age of vanilla*
 287 *agroforest [years] (D), planting density [pieds ha⁻¹] (E) and annual precipitation [mm year⁻¹]*
 288 *(F). Green vanilla yields were independent of land-use history and positively associated with*

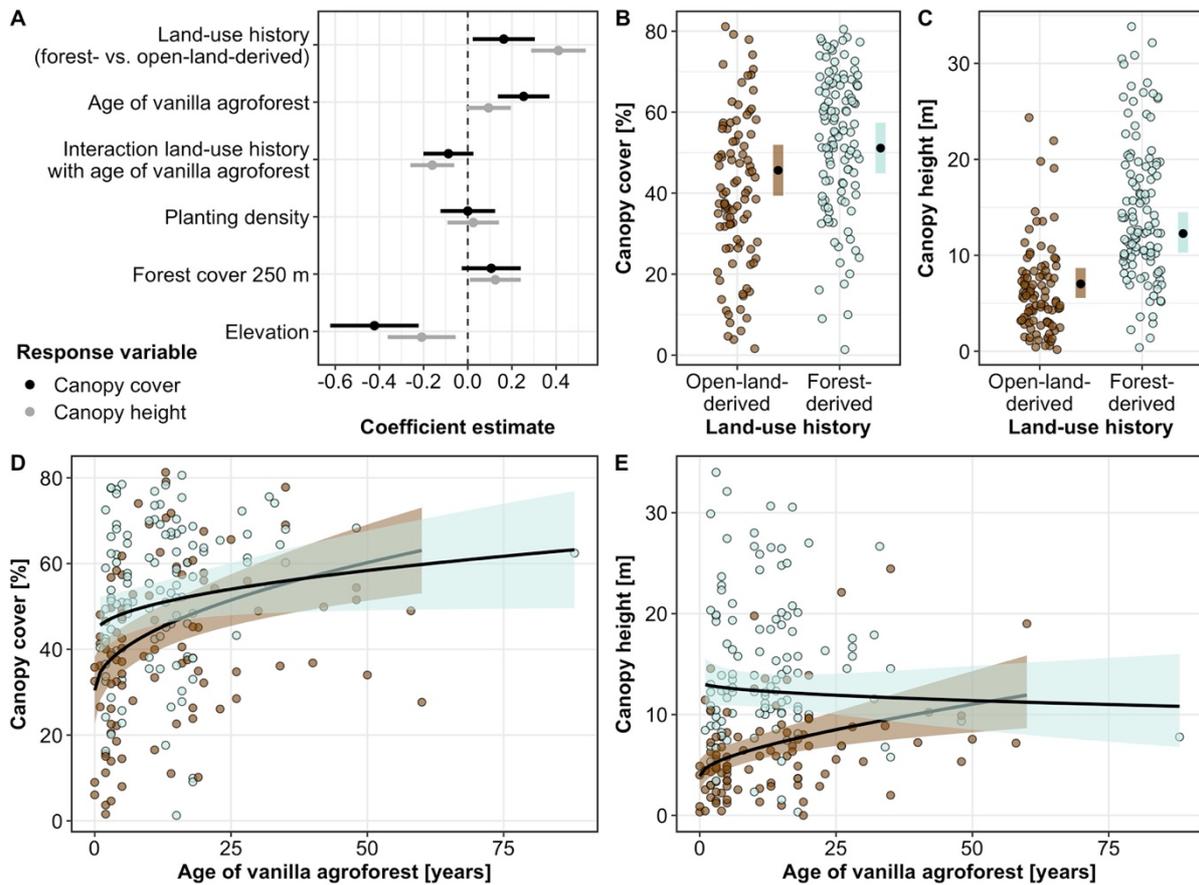
289 *all four continuous variables, but the relationships between canopy cover and yields as well*
290 *as annual precipitation and yields were not significant. Lines respectively black dots show*
291 *back-transformed estimated marginal means based on the linear mixed-effect model and*
292 *shaded areas depict 95% confidence intervals. Points are raw data separated in forest-*
293 *derived (blue) and open-land-derived (brown) agroforests. A table with model results can be*
294 *found in the supplementary materials (SM Table 1).*

295 **Determinants of canopy cover and canopy height**

296 Canopy cover was 12.9% (estimated marginal means 4.9%) higher in forest-derived vanilla
297 agroforests compared to open-land-derived agroforests (Figure 3). The age of the
298 agroforests differed along the chronosequence between 1 and 88 years in forest-derived
299 vanilla agroforests and between 0 and 60 years in open-land-derived agroforests. Age
300 positively correlated with canopy cover, in both forest- and open-land derived agroforestry:
301 In open-land-derived agroforests, canopy cover increased by 32.6% over 60 years, while
302 canopy cover only increased by 14.2% in forest-derived agroforests over 60 years (17.6%
303 over 88 years). Similarly, canopy height was 8.2 m (estimated marginal means 5.2 m) higher
304 in forest-derived agroforests compared to open-land-derived agroforests. The age of the
305 agroforest positively affected canopy height in open-land-derived agroforests where canopy
306 height increased on average by 8 m over 60 years while canopy height decreased by 1.7 m
307 in forest-derived agroforests over 60 years (2.1 m over 88 years).

308 Vanilla planting density did not correlate with canopy cover or height. Furthermore,
309 agroforests with more surrounding forest cover had higher trees and higher canopy cover,
310 but confidence intervals overlapped zero for the latter. Elevation was negatively associated
311 with both tree height and canopy cover. The canopy cover model had a marginal R^2 -value of
312 0.34 and a conditional R^2 -value of 0.56, while the canopy height model had a marginal R^2 -
313 value of 0.35 and a conditional R^2 -value of 0.74. The substantial difference between
314 conditional and marginal R^2 -values stemmed from the strong explanatory power of the

315 random effect 'household' (canopy cover model: $\tau_{00} = 0.16$ / canopy height model: $\tau_{00} =$
 316 0.37); the random intercept variance for the random effect 'village' was small (canopy cover
 317 model: $\tau_{00} = 0.07$ / canopy height model: $\tau_{00} = 0.02$).



318
 319 *Figure 3: Results of two linear mixed effect models explaining canopy cover [%] and canopy*
 320 *height [m] across 209 vanilla agroforests. A: Scaled effect plot of the reduced canopy cover*
 321 *model (black) and the reduced canopy height model (grey) for all five predictors, including*
 322 *the interaction between land-use history and age [years]. B & C: Forest-derived agroforests*
 323 *(blue) had both higher canopy height and higher canopy cover compared to open-land-*
 324 *derived agroforests (brown). D: Older forest- and open-land-derived agroforests had higher*
 325 *canopy cover, an effect that was stronger in open-land-derived agroforests. E: Older*
 326 *agroforests also had higher canopies, but only if open-land-derived. Lines respectively black*
 327 *dots show back-transformed estimated marginal means based on linear mixed-effect models*
 328 *and shaded areas depict 95% confidence intervals. Points are raw data separated in forest-*
 329 *derived (blue) - and open-land-derived (brown) agroforests. Tables with the results of both*
 330 *models can be found in the supplementary materials (SM Table 2 and 3).*

331

332 **Canopy cover dynamics in the agricultural landscape**

333 Areas within the agricultural landscape around villages that had low initial canopy cover in
334 the year 2000 experienced little change from 2000 to 2010 (Figure 4 C in the discussion, SM
335 Table 4). Areas with medium to high initial canopy cover experienced an increase in canopy
336 cover of up to 6.4% at 68.3% initial canopy cover (Figure 4 A). Areas with very high initial
337 canopy cover lost 4.4% of canopy cover (Figure 4 B, D). Overall, canopy cover increased by
338 2.7%. The general additive model explained 8.0% of the variation in the data.

339 **Discussion**

340 Across an 88-year chronosequence of 209 agroforests in the SAVA region of north-eastern
341 Madagascar, we found vanilla yields to vary widely and to be positively affected by planting
342 density and agroforest age, while canopy cover and precipitation had no effects on yields.
343 Older vanilla agroforests had higher canopy cover, and, if open-land-derived, also greater
344 canopy height. On the landscape-scale, areas within the agricultural landscape with medium
345 canopy cover gained canopy cover between the years 2000 and 2010.

346 **Determinants of vanilla yields**

347 We found vanilla yields to be hugely variable across agroforests, ranging from 0 - 932.7 kg
348 ha⁻¹. This variability was driven by variable yields per *ped* (unit of tutor tree and vanilla vine)
349 and planting densities. Such variability is typical for smallholder agroforests in tropical
350 countries (Clough et al., 2011) and points towards large yield gaps caused by sub-optimal
351 management practices (Lobell et al., 2009). This also suggests a large intensification
352 potential in existing agroforests and opportunities for sustainable intensification (Tilman et
353 al., 2011). Our yield estimate of 154.6 kg green vanilla per hectare is lower than most other

354 vanilla yield estimates, but published studies cover a large range of rather intensive
355 systems, including plantations with artificial shade (Supplementary Material Table 1),
356 potentially explaining lower yields in rather extensively managed Malagasy agroforests. The
357 here-reported yield estimate of 154.6 kg ha⁻¹ translates into gross earnings of annually 4684
358 € ha⁻¹, exhibiting the exceptional income opportunity vanilla provides under the high prices
359 of the year 2017 (Hänke et al., 2018). However, an average rural household in the study
360 region only sells 51.6 kg of green vanilla per year (Hänke et al. 2018; also including
361 households which did not sell any vanilla) and labour demands for the crop are high (Correll,
362 1953). Furthermore, high vanilla prices have led to a surge in local living costs, which are
363 estimated at 5751 € per household and year (Hänke & Fairtrade International, 2019), and
364 vanilla theft is commonplace (Neimark et al., 2019), further impairing the situation for
365 farmers.

366 In contrast to other studies (Havkin-Frenkel & Belanger, 2018; Santosa et al., 2005), we do
367 not see yield declines after a certain plantation age. The explanation for this is twofold:
368 Farmers constantly establish new *pieds*, resulting in old agroforests that still contain vanilla
369 vines of young and medium age (*DAM personal observation*). Furthermore, constant
370 'looping' of vines on the same *ped* is common: Hereby, vanilla vines are guided back down
371 to the soil where new roots establish (Fouché & Jouve, 1999). The originally planted part of
372 the vine may thus die at some point, but the vine survives thanks to the secondary access to
373 water and soil nutrients. Given that new *pieds* are also propagated by vine-cuttings (Fouché
374 & Jouve, 1999; Havkin-Frenkel & Belanger, 2018), planting of new *pieds* and looping are
375 comparable processes. In combination with the relatively short time to first produce (Circa 3
376 years; Havkin-Frenkel and Belanger 2018), the looping of vanilla vines may lead to stable
377 yields over time and could thus avoid boom and bust cycles. Such cycles are a common

378 occurrence in other agroforestry crops like cacao (Clough et al., 2009) and refer to farmers
379 realising short-term increases in yields through shade trees removal at the expense of
380 associated biodiversity and ecosystem functions (Tscharntke et al., 2011). The resulting yield
381 increase may be followed by a decrease, caused by elevated pest pressure and dwindling
382 soil fertility (Clough et al., 2009). Falling yields prompt the abandoning of plantations and
383 further forest conversion to agroforestry elsewhere (Clough et al., 2009). The likely absence
384 of these busts in vanilla agroforests does hence point towards the stability of the
385 agroforestry system.

386 Despite methodological improvements over, to our knowledge, all previous studies (SM
387 Table 1), this study lacks detail on many potential drivers of vanilla yields. This is highlighted
388 by the weight of the random effects. The random effect 'household' might reflect
389 differences in management practices between households (Hänke et al., 2018), while
390 'village' might represent biotic or abiotic village-level effects, such as different soil
391 properties. We thus call for more research on vanilla yield determinants that may generate
392 more applicable management advice for farmers.

393 **Increasing vanilla yields without impairing canopy cover**

394 We show that vanilla yields vary independently of canopy cover suggesting that no trade-
395 offs exist between yields and maintaining or restoring trees, much in contrast to
396 comparable crops where yields typically decline above 40% canopy cover, for example in
397 cacao (Blaser et al., 2018; Clough et al., 2011) or coffee (Jezeer et al., 2017). The here-shown
398 independence of yields and canopy cover enables farmers to maintain remnant forest trees
399 in forest-derived agroforests, which are highly beneficial for ecosystem services and
400 biodiversity (Tscharntke et al., 2011), at no direct cost. Furthermore, tree and canopy cover

401 rehabilitation in open-land-derived agroforests is also possible without compromising on
402 yields. The independence of vanilla yields and shade is supported by plant-physiological
403 experiments which show that vanilla performs well under various light regimes (Díez et al.,
404 2017).

405 Interestingly, vanilla planting density was independent of canopy cover and canopy height.
406 This suggests that closing yield gaps is possible by planting vanilla *pieds* more densely and
407 that doing so does not *per se* impair canopy cover or height within the currently existing
408 planting density range. Given the benefits of trees for biodiversity and ecosystem functions
409 and services (Leakey, 2014; Tschardt et al., 2011), this further strengthens the case for
410 sustainable intensification opportunities in vanilla agroforestry.

411 **Increasing canopy cover and tree height over time**

412 Tree rehabilitation in agroforestry systems is a global priority (FAO, 2017). However, many
413 tropical agroforests of key cash crops like cacao or coffee are forest-derived, thus typically
414 contributing to forest degradation rather than tree rehabilitation (Martin, Osen, et al.,
415 2020). Here, open-land derived agroforests may contribute to tree rehabilitation, but
416 empirical chronosequences that document tree recovery in open-land derived agroforests
417 are rare (but see Nijmeijer et al. (2019)). Here we show that canopy cover is higher in older
418 forest- and open-land-derived agroforests than in younger ones. Furthermore, trees were
419 higher in older open-land-derived agroforests, but not in older forest-derived agroforest.
420 This suggests that open-land-derived agroforests can play a key role in tree rehabilitation,
421 given that they originate from open fallow land. They could thus contribute to increased
422 carbon storage (Nair et al., 2009) and the restoration of ecosystem services (De
423 Beenhouwer et al., 2013) while providing new habitat for tree-dependent taxa (Clough et

424 al., 2011). The transformation of land under shifting cultivation into cash cropping systems
425 is furthermore in line with regional (Andriatsitohaina et al., 2020) and global trends (van
426 Vliet et al., 2012).

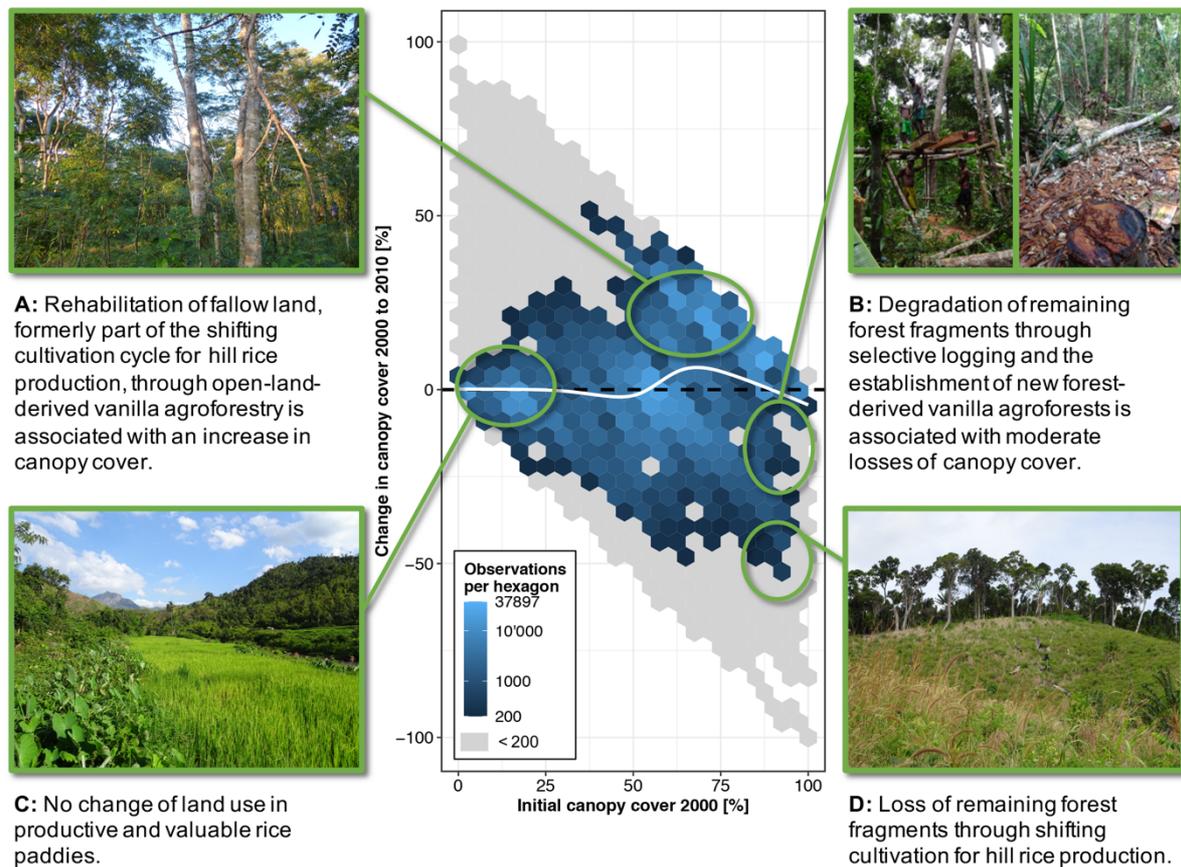
427 In contrast to open-land-derived agroforests, canopy cover in forest-derived agroforests will
428 likely only recover after an initial drop at time of establishment (Martin, Osen, et al., 2020),
429 which is not covered here as our chronosequence does not include forest fragments. The
430 stable tree height is in line with this interpretation, as the removal of single trees at time of
431 establishment may not reduce mean tree height at the plot level. Alternatively, the resulting
432 chronosequence could also stem from a change of practices over time, i.e. farmers today
433 cut more trees at time of establishment than they did in the past, resulting in recently
434 established forest-derived agroforests with low canopy cover in the chronosequence.

435 Taken together, our results highlight the value of open-land-derived agroforests for tree
436 rehabilitation and shows that forest-derived vanilla agroforests may have relatively stable
437 canopy cover over time.

438 **Canopy cover dynamics in the agricultural landscape**

439 We used remotely sensed canopy cover data to explore how observed plot-scale tree
440 rehabilitation translates to the landscape-scale. Comparing canopy cover changes between
441 2000 and 2010, we found that areas with lowest initial canopy cover, probably mostly rice
442 paddies, had stable canopy cover. This is to be expected, given the high productivity of
443 irrigated rice and its local importance for food security (Hänke et al., 2018; Laney & Turner,
444 2015), which make a conversion of rice paddies to other land uses unlikely. Areas with very
445 high canopy cover, i.e. forest fragments around villages, lost canopy cover over time. Here,
446 small losses may represent forest degradation through selective logging for timber or

447 through the establishment of new forest-derived vanilla agroforests. Some of these areas
 448 also showed large losses, likely reflecting shifting cultivation, where forest is cut and burned
 449 for hill rice cultivation (Figure 4).



450

451 *Figure 4: Canopy cover dynamics in the agricultural landscape in a 2 km circle around centres*
 452 *of 60 focal villages between 2000 and 2010 using canopy cover raster data with 30 m*
 453 *resolution (Hansen et al., 2013). Canopy cover increased overall by 2.7%, driven by canopy*
 454 *cover increase in areas with medium to high initial canopy cover (e.g. vanilla agroforests; A).*
 455 *Canopy cover did, however, decrease in areas with very high initial canopy cover (e.g. forest;*
 456 *B, D) and was stable in areas with little initial canopy cover (e.g. rice paddies; C). The central*
 457 *plot shows hexagon bins of bin-width 5% which are coloured according to the number of*
 458 *30x30 m raster cells (i.e. observations) within each hexagon bin. Hexagon bins with less than*
 459 *200 observations are grey. The white line depicts predicted outcomes of a general additive*
 460 *model explaining change in canopy cover (SM Table 4). All photos by the authors.*

461 Areas with medium to high initial canopy cover showed increases in canopy cover, most
 462 likely representing fallows that were transformed to open-land-derived vanilla agroforests.
 463 Here, the cessation of repeated burning for shifting cultivation, that comes with the

464 establishment of permanent agroforestry, may have enabled tree rehabilitation on the land,
465 as observed inside the plots.

466 Overall, these dynamics resulted in a net increase in canopy cover on the landscape scale,
467 as observed for agricultural landscapes across Madagascar (Zomer et al., 2016). The
468 combination of canopy cover gains and losses may be positive for species and ecosystem
469 services that can be provided by areas with medium canopy cover, such as the provision of
470 fruit or firewood. Forest-dependent species and ecosystem services that depend on high
471 canopy cover, as found in forest, will suffer. Conservation of remaining forests is thus
472 necessary to conserve the large share of Malagasy biodiversity that cannot persist outside
473 forest (Irwin et al., 2010; Martin, Andriafanomezantsoa, et al., 2020). Furthermore, the
474 forests of north-eastern Madagascar have some of the highest carbon stocks of all Malagasy
475 forest (Vieilledent et al., 2016), underlining the importance of forest conservation in light of
476 climate change mitigation.

477 Importantly, these findings are limited to the agricultural landscape around 60 focal villages
478 that are predominantly not at the deforestation frontier. Canopy cover dynamics might be
479 different in villages closer to large connecting forest blocks, where an overall increase in
480 canopy cover seems unlikely given the ongoing deforestation trend in Madagascar
481 (Vieilledent et al., 2018).

482 **Conclusion**

483 Our main finding, that yields and canopy cover in vanilla agroforests of north-eastern
484 Madagascar varied independently, suggests the possibility to combine high vanilla yields
485 with a high cover of trees. This has potential benefits for ecosystem services and
486 biodiversity in a globally important biodiversity hotspot. Our finding contrasts with other

487 agroforestry crops for which higher canopy cover typically impairs yields. Furthermore, the
488 higher canopy cover in older compared to younger vanilla agroforests suggests tree
489 rehabilitation opportunities in open-land-derived agroforests. If coupled with effective
490 protection of remaining forests, yield-neutral tree recovery in agroforestry systems could
491 provide a win-win situation for ecosystem functions, such as carbon storage, and
492 agricultural production in smallholder-dominated agricultural landscapes.

493 **Authors' contributions**

494 All authors conceived ideas and planned data collection and analysis. DAM, AW, and KO
495 coordinated the data collection; TR collected field data; DAM analysed and visualized the
496 data; DAM led the writing of the manuscript. All authors contributed to the writing and gave
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510 **Data availability statement**

511 Supplementary Materials, Data, and R-Code are available within the Open Science
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