

1 **Title:** Drone data reveal heterogeneity in tundra greenness and phenology not captured by
2 satellites

3

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17

18 **Abstract:**

19

20 Data across scales are required to monitor ecosystem responses to rapid warming in the
21 Arctic and to interpret tundra greening trends. Here, we tested the correspondence among
22 satellite- and drone-derived seasonal change in tundra greenness to identify optimal spatial
23 scales for vegetation monitoring on Qikiqtaruk - Herschel Island in the Yukon Territory,
24 Canada. Combining time-series of the Normalised Difference Vegetation Index (NDVI) from
25 ultra-fine-grain multispectral drone imagery and satellite data (Sentinel-2 and MODIS) with
26 ground-based observations for two growing seasons (2016 and 2017), we found high
27 cross-dataset correspondence in peak season greenness (Spearman's $\rho > 0.77$) and
28 cross-season greenness changes (drone-sentinel $R^2 = 0.69$) for eight one-hectare plots, with
29 drones capturing lower NDVI values relative to Sentinel-2 satellites. We identified a plateau
30 in the spatial variation of tundra greenness at distances of around half a metre in the plots,
31 suggesting that these grain sizes are optimal for monitoring such variation in the two most
32 common vegetation types on the island. We further observed a notable loss of seasonal
33 variation in the spatial heterogeneity of landscape greenness (46.2 - 63.9%) when
34 aggregating from ultra-fine-grain drone pixels (approx. 0.05 m) to the size of medium-grain
35 satellite pixels (10 – 30 m). Finally, seasonal changes in drone-derived greenness were
36 highly correlated with measurements of leaf-growth for focal deciduous species in the
37 ground-validation plots (mean Spearman's $\rho = 0.68$). These findings indicate that
38 multispectral drone measurements can capture temporal plant growth dynamics across
39 tundra landscapes. Overall, our results demonstrate that novel technologies such as drone
40 platforms and compact multispectral sensors allow us to study ecological systems at
41 previously inaccessible scales and fill gaps in our understanding of tundra ecosystem
42 processes. Capturing fine-scale variation across tundra landscapes will improve predictions
43 of the ecological impacts and climate feedbacks of environmental change in the Arctic.

44

45 **Keywords:** Arctic tundra, vegetation monitoring, landscape phenology, satellite, drones,
46 UAV and RPAS, NDVI, scale

47 Introduction

48

49 Identifying the scales at which ecological processes operate is a fundamental, yet often
50 neglected element of ecological research (1–3). Cross-scale ecological information can
51 inform our understanding of the causes and consequences of global change (2). In tundra
52 ecosystems, vegetation responses triggered by rapid Arctic warming could influence
53 ecosystem functions through altered carbon and nutrient cycles with potential feedbacks to
54 the global climate system (4–8). Yet, challenging logistics have limited the extent of
55 field-based observations in Arctic ecosystems (9–11). The grain sizes of global-extent
56 satellite products (tens of meters to kilometres) are too coarse to capture the fine-scale
57 dynamics of tundra plants (12–14) and to link vegetation change to key ecosystem functions
58 (13). Thus by bridging this “scale-gap”, we can transform our understanding of pan-Arctic
59 tundra vegetation change and associated global-scale climate feedbacks.

60

61 *Satellites show greening of the tundra*

62

63 Satellite observations indicate a ‘greening’ of tundra ecosystems (13,15–20) and shifts in
64 growing season phenology over recent decades (13,21–24). Observations of increasing
65 tundra greenness are often reported from surface-reflectance-derived Normalised Difference
66 Vegetation Index (NDVI) (13,16,25,26). Satellite-observed tundra greening has been
67 concurrent with ground-based observations of vegetation change in Arctic ecosystems (27)
68 including increasing shrub cover (28–31) and taller community level plant height (32), as well
69 as earlier leaf emergence and flowering at some (33–36), but not all tundra sites (37–39).
70 However, mismatches between ground and satellite-based observations suggest the
71 potential for an observational scale gap (13).

72

73 *Arctic vegetation change and phenology have been linked to warming*

74

75 Satellite-observed Arctic greening trends have been linked directly to warming air
76 temperatures (13,19,20,40–46) and indirectly to sea-ice declines (17,47–51). Ground-based
77 observations of tundra vegetation change correspond with warming (27,32,52) but do not
78 always co-occur with satellite greening trends in the regions around the ecological
79 monitoring sites (13,53). While satellite-based phenology observations from the Arctic have
80 been mainly linked to temperature (22,54,55), *in situ* phenology in the tundra has been
81 shown to be influenced by a suite of interacting factors including, but not limited to:
82 snowmelt, temperature, day length, and the proximal influences of sea-ice on localised
83 climate affect (34–36,38,56,57). Thus, ecological studies indicate greater complexity of
84 drivers that analyses of satellite-derived greening trends to date.

85

86 *Inconsistencies amongst satellite platforms and heterogenous greening trends*

87

88 Greening trends and phenology measures derived from different satellite platforms do not
89 always correspond (13,18) and greening trends vary at global (18), continental (42,58–60)
90 and regional scales (46–48,61–64). Many areas of the Arctic show no trends in NDVI, with
91 only around 20% of the Arctic spectrally greening and around 1 - 4% of the Arctic spectrally
92 browning (13,62,65,66). Recent analyses suggest a slowdown of the Arctic-wide spectral

93 greening trend over the past decade (43,67). Furthermore, despite NDVI being related to the
94 photosynthetically active biomass in the tundra (14,68–70), geophysical, environmental and
95 ecological factors, in addition to the non-linearity of NDVI-biomass relationships, complicate
96 the interpretation of satellite-derived NDVI time-series at high-latitudes (13,71). The growing
97 complexity highlighted in Arctic greening trends has led to repeated calls for ground
98 validation of satellite observations (11,13,18,59,60,66,72,73).

99

100 *The scale discrepancy problem in Arctic greening*

101

102 A major problem in linking satellite-derived trends of tundra spectral greenness and
103 phenology to *in situ* observations of ecological processes is the discrepancy in observational
104 scales (13,29,61,72,74). Satellite datasets with long-term records are limited by their
105 moderate- to coarse-grain sizes, ranging from 30 m (Landsat) to 250 m (MODIS) and 8 km
106 (AVHRR-GIMMS3g). *In situ* ecological monitoring in the Arctic is logistically challenging and
107 therefore restricted in extent to a limited number of sites and often metre-squared plots
108 (10,75). Only a few studies have linked on-the-ground vegetation or phenology change to
109 satellite trends in NDVI in Arctic tundra (13,14,47,48,53,76–78). However, drones equipped
110 with compact sensors now allow for the collection of ultra-fine-grain multispectral imagery at
111 landscape extents that can potentially bridge the scale-gap between satellite and
112 ground-based observations (14,79–82).

113

114 *Novel drone data to study variation in greenness*

115

116 Here, we set out to test whether drones can be used to identify the key ecological scales for
117 studying tundra greenness on Qikiqtaruk by bridging the scale gap between satellite and *in*
118 *situ* data. First, we tested whether satellite- and drone-derived measures of mean
119 landscape-scale greenness (NDVI) agree across two growing seasons while controlling for
120 the potentially confounding effects of topography and land cover. Second, we identified the
121 key spatial scales for ecological variation in landscape greenness within the two most
122 common vegetation types at our study site using variograms. Third, we tested how the
123 magnitude of seasonal variation in tundra greenness scales across grain sizes from
124 fine-resolution drone imagery to medium-grain satellite imagery. Finally, we assessed
125 whether drone-derived NDVI corresponds with on-the-ground measures of within growing
126 season change in plant growth frequently measured by long-term field-based monitoring
127 networks. Thus, in our analysis we validated satellite-derived landscape estimates of
128 vegetation greenness with ultra-fine-grain drone data and described spatial and temporal
129 variation in tundra productivity at grain sizes and extents that were not previously accessible.

130

131 **Methods**

132

133 *Site description: Qikiqtaruk - Herschel Island*

134

135 Qikiqtaruk (69.57 N, 138.91 W) is located in the Beaufort Sea along the coastline of the
136 North Slope of the Yukon Territory, Canada. The vegetation is characteristic moist acidic
137 shrub tundra (83) found in the Western Arctic regions of North America that has experienced
138 strong spectral greening in recent decades (13). The two most common plant communities

139 on the island are the tussock sedge (“Herschel”) and Dryas-vetch (“Komakuk”) vegetation
140 types (84,85). We established four study areas on the east end of the island, each with two
141 co-located one-hectare plots in the two key vegetation cover types (Figure 1, Table S1). We
142 selected plots with homogenous terrain and land cover to represent the two key vegetation
143 types and to control for the potentially confounding effects of terrain and cover heterogeneity.

144

145 *Multispectral drone time-series*

146

147 We analysed 62 drone surveys across four research areas with one-hectare plots in the two
148 vegetation types at the site (Table S3). We collected multispectral drone imagery using
149 Parrot Sequoia (Paris, France) compact multispectral sensors mounted on multi-rotor drone
150 platforms in June to August in 2016 and 2017. We used three different drone platforms: a
151 Tarot 680 Pro hexacopter with camera sensor stabilisation in 2016, and a 3DR Iris+ and a
152 DJI Phantom 4 Pro without sensor stabilisation in 2017. Surveys were flown in a lawn-mower
153 flight pattern at an altitude of ca. 50 m, giving ground-sampling distances of 0.04 m to 0.06
154 m. Images were acquired with 75% front- and side-lap as close as possible to solar noon
155 (mean absolute difference to solar noon 2.16 h, maximum 6-7 h). See Table S3 for further
156 details.

157

158 We processed the Sequoia imagery using Pix4D Mapper v4.0.21 (Lausanne, Switzerland)
159 with the *agMultispectral* template and the ‘merge map tiles’ option set to true to generate
160 co-registered single-band surface reflectance maps. Radiometric calibration was carried out
161 in Pix4D Mapper using pre- or post-flight imagery of calibrated reflectance panels; in 2016
162 we used a MicaSense (Seattle, USA) panel and in 2017 a SphereOptics (Herrsching,
163 Germany) Zenith Lite panel. We measured panel reflectance pre- and post- season and
164 used the mean values for radiometric calibration. We also calibrated for sensor properties
165 and sun irradiance measured by the incident light sensor. We used four to six ground control
166 points per survey to geolocate the imagery in Pix4D Mapper with an estimated accuracy of
167 1-2 pixels between bands and 2-6 pixels between surveys (81). We calculated the Sequoia
168 NDVI as the normalised difference between the near-infrared (770 nm – 810 nm) and red
169 (640 nm – 680 nm) bands of sensor.

170

171 *Satellite time-series*

172

173 We obtained MODIS NDVI values for the time period between May and September 2016
174 and 2017 for all 250 m MODIS pixels that contained the survey plots. NDVI values were
175 retrieved from the 16-day MOD13Q1 v6 Terra product (86) using the Google Earth Engine
176 (87). We discarded all values with a ‘Summary QA’ score of -1 (no data) or 3 (cloudy). Table
177 S4 lists the resulting MODIS-pixel-date pairs. The MODIS NDVI is calculated as the
178 normalised difference between bands 1 (841 nm – 876 nm) and band 2 (620 nm – 670 nm).

179

180 For the Sentinel-2 time-series, we gathered all Sentinel-2 MSI L1C scenes containing the tile
181 covering Qikiqtaruk (T07WET) that were available on the Copernicus Open Access Hub
182 (<https://scihub.copernicus.eu/>) for the same time period as the MODIS pixels. We processed
183 all scenes to L2A using Sen2Cor 2.4.0 (88), retained all bands with 10 m resolution (2-4 &
184 8), applied the cloud mask and generated a true-colour image. We inspected all scenes

185 visually and discarded all imagery with cloud contamination over the study area (78% of
186 scenes for 2016 and 74% of scenes for 2017). The resulting set contained nine cloud-free
187 Sentinel-2 L2A scenes of the study area from 2016 and fifteen scenes from 2017 (Table S5).
188 Finally, the Sentinel NDVI was calculated as the normalised difference between band 8
189 (784.5 nm - 899.5 nm) and band 4 (650 nm - 680 nm).

190

191 *Ground-based plant phenology measurements*

192

193 We carried out ground-based phenology monitoring in eight 2 m x 2 m plots (Table S2), one
194 adjacent to each one-hectare plot (mean distance = 23 m, max distance = 52 m). Within
195 these plots we monitored six individual plants from the most common species: *E. vaginatum*,
196 *D. integrifolia*, *S. pulchra* and *A. latifolia* in tussock sedge tundra; *D. integrifolia*, *S. arctica*
197 and *A. latifolia* in Dryas-vetch tundra. We measured the length of the longest leaf on each
198 individual on the survey date to the nearest millimetre. This approach is widely used in
199 field-based phenology monitoring protocols (89), and will allow for NDVI to be directly related
200 to phenological changes in plant traits. The majority of ground-based phenology surveys
201 were carried out on the same day as the drone surveys (mean difference = 0.3 days,
202 maximum difference = 3 days, Table S6).

203

204 *Cross-sensor correspondence*

205

206 To test cross-sensor correspondence, we first plotted the mean NDVI for all plots,
207 time-points and sensors available (MODIS = single pixel) across both growing seasons. We
208 then calculated the mean difference and Spearman's rank correlation of the peak-season
209 NDVI for 2017 amongst the sensors (mean 20 July - 10 August). We matched all drone and
210 Sentinel-2 scenes that were less than two days apart, resampled the drone bands to the
211 Sentinel-2 grid, calculated the NDVI and tested the predictive relationship between the
212 resampled drone and sentinel NDVI pixel-pairs for a random sub sample (10% of total, n =
213 700) with Bayesian linear models (Table S9 and S10) from the MCMCglmm v.2.29 package
214 (Hadfield 2010).

215

216 *Spatial autocorrelation*

217

218 To assess the spatial autocorrelation of variation in tundra greenness within the eight plots,
219 we sampled variograms and fitted variogram models using the gstat v. 2.0-5 package
220 (Pebesma 2004, Gräler et al 2016). All variograms were sampled with a bin width of 0.15 m
221 from 0 to 15 m and a bin width of 3 m from 0 to 45 m.

222

223 *Grain size and phenology*

224

225 We tested the influence of grain-size on observations of tundra greenness phenology by
226 fitting simplified growing season curves to the raster stacks for each plot and season. We
227 first resampled the drone bands for all time-points to grids with grain sizes of 0.5, 1, 5, 10, 20
228 and 33.33 m. We then calculated the NDVI and fitted simple quadratic models to each pixel
229 in the growing season stacks ($y = ax^2 + bx + c$, where x is the day of year and y the pixel
230 NDVI, a the quadratic coefficient, b the linear coefficient and c the constant term). We found

231 a strong negative correlation between the quadratic and linear coefficients of the models
232 (Figure S5), we selected only the quadratic coefficient for further analysis.

233

234 *Ground validation*

235

236 To test the correspondence between our ground-based phenology measurements and the
237 drone observations, we derived time-series of the mean longest leaf length and greenness
238 for each 2 m x 2 m ground-based monitoring plot. For each drone survey, we calculated the
239 mean-NDVI of the 2 m x 2 m monitoring plot and matched this with the mean longest leaf
240 length values derived from the corresponding ground-based surveys (Table S6). We then
241 calculated the Spearman's rank correlation between mean-NDVI and mean longest feaf
242 length for each plot and season.

243

244 *Statistical analyses*

245

246 We conducted statistical analyses using R v. 3.6.0 (90). We used the 'resample' function of
247 the raster package in R for resampling from finer to coarser resolutions (91). See extended
248 methods for further information.

249

250 **Results**

251

252 *Landscape greenness corresponded among sensors*

253

254 Landscape greenness corresponded among drone, Sentinel-2 and MODIS across both the
255 2016 and 2017 growing seasons. Growing season curves of the plot mean NDVI were
256 similar (Figure 1) and peak-season plot mean NDVI values for 2017 were highly correlated
257 across sensors (Spearman's $\rho > 0.77$, Table S7). However, we observed an offset between
258 drone and satellite plot-mean NDVI of around 0.08 absolute NDVI that was consistent for
259 both MODIS and Sentinel platforms (Table S8). Resampled 10 m drone pixels and the
260 corresponding spatially co-located Sentinel-2 pixels were highly correlated (marginal $R^2 =$
261 0.69, see Figure 2 and Table S9). We found that vegetation type, the specific Sentinel
262 platform (Sentinel-2A / Sentinel-2B), and the time-difference between Sentinel scene and
263 drone data acquisition influenced the relationship between Sentinel-2 pixel NDVI and
264 drone-derived NDVI (marginal $R^2 = 0.87$, Table S10).

265

266 *Spatial variation in landscape greenness peaked at approx. 0.5 m*

267

268 We observed a peak in spatial variation in the NDVI values of pixel pairs for distances below
269 0.5 meter (mean range 0.46 m) during the peak-season of 2017 (26-28 July) and little
270 additional spatial variation was found between pixel pairs for distances of up to 45 m
271 thereafter (Figure 3). This pattern was consistent across vegetation types in seven out of our
272 eight plots (Figure 3, S2 and S3). The only exception is the Dryas-vetch plot in Area 3, which
273 showed the same patterns for distance below 10 m, but thereafter spatial variation steadily
274 increased (Figure S3).

275

276 *Seasonal-variation was lost when aggregating to medium grain sizes*

277

278 We observed a notable loss in the amount of seasonal variation in tundra greenness when
279 aggregating grain size from ultra-fine-grain drone to medium-grain satellite data. The loss
280 was particularly pronounced at grain-sizes above 10 m – the grain size of Sentinel-2 MSI
281 pixels (46.2 - 63.9%) (Figure 4). The variation in the quadratic coefficient of the simple
282 growing season curves (Figure 4b and S6) decayed logarithmically with grain size (Figure
283 4a), while no change occurred in the mean tendency of the coefficient (Figure S4). The
284 quadratic and linear coefficients of the growing season curves were strongly correlated
285 (Spearman's $\rho = -0.999$), thus the same pattern holds true for the linear component of the
286 curve fit (Figure S5).

287

288 *Drone-derived spectral greenness correlated well with leaf measurements*

289

290 Drone-derived spectral greenness correlated well (mean $\rho = 0.66 - 0.71$) with ground-based
291 measurements of cross-season phenology for graminoids and deciduous plants (Figure 5).
292 The mean Spearman's correlation coefficient of the measured mean leaf length and the
293 mean NDVI values in the ground-based phenology plots was 0.68 across all species and
294 time-series (Table S11 and Figure 5a). The graminoids and deciduous shrub species
295 followed this mean tendency well across all time-series, while the partially-evergreen *D.*
296 *integrifolia* showed mixed responses between plots and years (mean $\rho = 0.22$, Figure 5a).
297 The drone-based greenness time-series of the 2 m x 2 m ground-phenology plots highlight
298 fine-scale differences in phenology such as the continuous greening of tussocks that was
299 visible at the tussock sedge tundra plot in Area 2 (Figure 5c).

300

301 **Discussion**

302

303 Our analysis of cross-scale time-series of landscape greenness on Qikiqtaruk highlights four
304 main findings: 1) Measures of mean tendency in landscape greenness were consistent, but
305 slightly offset with drones capturing lower NDVI values than Sentinel-2 and MODIS satellites
306 (Figures 1 and 2). 2) The majority of variation in landscape greenness was contained at
307 scales of around half-a-metre, and thus not captured by medium-grain satellites such as
308 Sentinel-2 (Figure 3). 3) When aggregating growing season curves from ultra-fine-grain
309 drone to medium-grain satellite pixel sizes, a notable amount (46.2 - 63.9%) of variation in
310 greenness phenology was lost (Figure 4). 4) Drone-based measures of landscape greenness
311 correlated well with ground-based measurements of leaf length (Figure 5). Taken together,
312 our results highlight that drone platforms and compact multispectral sensors can capture key
313 ecological processes such as vegetation phenology and bridge the existing scale gap
314 between satellite and ground-based monitoring in tundra ecosystems.

315

316 Our study indicates cross-platform agreement, yet a positive offset, in mean landscape
317 greenness. The correspondence between drone and satellite-derived NDVI has yet to be
318 tested across Arctic sites (13,14). Some studies of natural or agricultural systems have
319 reported similar or higher levels of agreement between multispectral reflectance products
320 from drones and satellites (14,92,93), while others reported mixed or poor agreement
321 (94–96). As in Franzini et al. (2019), we observed a positive offset between drone and
322 satellite NDVI (Figure 2a and Table S8) that warrants further investigation. Possible

323 ecological interpretations of this offset are that drone data better capture low NDVI values.
324 Possible technical explanations for this offset include: differences in viewing geometries
325 between drones (highly variable) and satellites (relatively consistent), distinct sensor
326 properties of drone and satellite sensors, influences of the atmosphere between the sensor
327 and the land surface and a variety of other factors influencing the estimated reflectance.
328 Siewert and Olofson (2020) do not report this offset in the more heterogenous tundra of
329 Arctic Sweden, raising the possibility that within-landscape variation in land cover or
330 topography may influence correspondence between vegetation greenness across scales.
331 The homogeneity of the landscape within our survey plots likely contributes to the strong
332 correlation between drone- and satellite-derived NDVI that we have observed (13).
333 Additional research is needed to evaluate how other scale-varying landscape characteristics
334 like land cover (including non-vegetative surfaces like water, rocks, snow, etc.) and
335 topography affect drone and satellite correspondence across the diverse and structurally
336 complex tundra biome.

337

338 In our study, ecological information was lost when upscaling from ultra-fine-grain (~ 0.05 m)
339 drone to moderate grain (~ 10 – 30m) satellite resolutions. Even the most recent generation
340 of freely-available multispectral satellite products can be limited in their ability to capture
341 fine-grain ecological processes of tundra vegetation change (13). Information transfer during
342 upscaling leads to the loss of more information in tundra ecosystems compared to other
343 biomes (14,97) as land cover and vegetation structure are fragmented at finer scales (98).
344 However, exactly how spatial aggregation influences the loss in observed ecological
345 variability across the diversity of Arctic landscapes remains poorly quantified (11,13). Yet,
346 this variability is critical to understanding climate-driven changes in vegetation phenology
347 (35,36,99), plant-pollinator interactions (100), and trophic interactions (101). With
348 drone-based monitoring, we observed a decrease in magnitude of the spatial variability in
349 landscape-level phenology throughout the growing season (Figures 4 and Figure S2), while
350 aggregation to moderate satellite grains obscured both the magnitude and timing of
351 phenological heterogeneity (Figures 4 and S6). Thus, time-series of fine-grain
352 remotely-sensed observations will be critical for answering key research questions about
353 tundra ecosystem functioning in a warming Arctic (102).

354

355 Our results indicate that drone-based greenness time-series captured variation in plot-level
356 leaf-growth of deciduous tundra plant species. We demonstrate how drones can be used to
357 measure variation in tundra plant phenology of metre-scale patches at landscape extents.
358 Drones have been successfully used to monitor phenology of individual plants (trees) in
359 temperate forest ecosystems (103–105), and our study indicates that individual plant-level
360 phenology monitoring of sub-decimeter variability from drones could also be carried out in
361 tundra ecosystems. Future studies that quantify plant growth or phenology events such as
362 leaf emergence and flowering across the landscape could provide key information on
363 resource availability for plant-consumer interactions (100,101). Our findings also highlight
364 known limitations of NDVI to track phenology in evergreens or other non-deciduous taxa (*D.*
365 *integrifolia*, Figure 5), suggesting that tests of alternative vegetation index - plant growth
366 relationships (105) are needed to capture cross-season variation in tundra evergreen and
367 moss species. Combining drone-based time-series with observations from phenocams,
368 satellite and ground-based study plots has the potential to revolutionise our understanding of

369 landscape-scale phenology (13) by moving beyond the previously small samples of
370 individuals monitored in the Arctic tundra (36,37,39,106).

371

372 Our study highlights limitations and challenges associated with the collection of multispectral
373 drone time-series in Arctic ecosystems. Recent studies have discussed challenges
374 associated with the use of compact multispectral drone sensors including radiometric
375 consistency and repeatability (81,107,108). Due to logistic constraints, we were not able to
376 always conduct surveys under optimal conditions due to sun angle or cloud cover nor as
377 frequently as planned due to wind or precipitation (Table S3), which likely introduced bias
378 and/or noise into our drone data (e.g., Figure 4b). Access limitations meant that we could not
379 capture spring and autumn on Qikiqtaruk. As an early-generation multispectral drone sensor,
380 the Parrot Sequoia was tailored for deriving the NDVI, which despite being the legacy
381 workhorse of tundra remote-sensing has limitations (11,13). In particular, NDVI can be
382 confounded by moisture and surface water (11,13,73,109), complicating interpretation in wet
383 tundra and particularly in fine-grain size data. However, the rapid technological development
384 of drones and sensors, as well as further consolidation and standardisation of methods
385 (110), will allow for pan-Arctic syntheses of fine-grain data to resolve the uncertainty and
386 complexity of Arctic greening patterns trends (13,14,81) (see also the High Latitude Drone
387 Ecology Network - <https://arcticdrones.org/>).

388

389 Our study demonstrates that drones can fill the scale-gap between satellite and field studies
390 in the observation of terrestrial Arctic vegetation change (13,111). Rather than investigating
391 and explaining patterns at scales pre-defined by satellite datasets or field-based networks,
392 researchers can use drones to identify scale-domains that are most closely associated with
393 ecological processes of interest. Field ecologists can now use scaling theory provided by the
394 remote sensing community (74,112–115) at scales and temporal intervals that will allow for
395 hypothesis testing about what mechanisms are driving landscape-level ecological change.
396 Drone imagery will also allow the remote sensing community to track the effects of sub-pixel
397 heterogeneity on satellite products down to the grain of individual plants and communities
398 (14) that have been long studied by field-based monitoring networks, like the International
399 Tundra Experiment (75). Only by improving our understanding of how ecologically important
400 information is captured across grain sizes can we reduce uncertainties in the medium- and
401 coarse-grain satellite observation that feed into Earth system models and shape their
402 predictions (4,8). Fine-scale remote sensing from drones and aircraft provides key tools for
403 disentangling the drivers behind the greening of the Arctic (13,14,79,102).

404

405 **Conclusions**

406

407 Novel remote-sensing technologies such as drones now allow us to study ecological
408 variation in landscapes continuously across scales. Fine-grain ecological observation is of
409 particular importance where variation in plant growth happens at very small spatial scales
410 such as in tundra ecosystems (13,71). Our finding of a peak in spatial variation found at
411 distances of ~0.5 m in the plots on Qikiqtaruk shows the grain size at which phenological
412 information within the plant communities is best captured at this site. We demonstrate that
413 key ecological information is lost when observing the tundra at even decimeter or coarser
414 scales, such as those of medium grain satellites (~ 10 – 30m). Despite the methodological

415 challenges of collecting multispectral drone imagery in remote environments (81), our
416 time-series of vegetation greenness correlated well with ground-based measurements of leaf
417 growth in the validation plots. Drones now enable cross-scale studies that fill scale gaps
418 between satellite and ground-based observations facilitating the identification of key drivers
419 of vegetation change to inform projections of climate change impacts and feedbacks in the
420 tundra biome.

421

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456

457 **Author Contributions**

458

459 JJA and IMS conceived the study with input from JTK and AMC. JJA carried out data
460 processing and analysis. JJA and IMS led the drone and ground-validation field work in

461 2016. AMC led the drone field surveys with input from JTK and GD led the ground-validation
462 for 2017 with input from JTK. JJA, IMS and JTK wrote the manuscript with input from AMC
463 and GD. IMS supervised and acquired funding for the research.

464

465 **Data availability**

466

467 All processed drone and Sentinel imagery is available via a data repository on Zenodo
468 (embargoed till publication of this manuscript).

469 Should the reviewers wish to access the data prior publication, a mirror of the Zenodo
470 repository can be accessed via this confidential link:

471



472

473 All code used to conduct the analysis, produce figures and as well as summary data
474 outputs and MODIS pixel values can be found on this GitHub repository (already openly
475 available):

476 https://github.com/jakobjassmann/qhi_phen_ts

477

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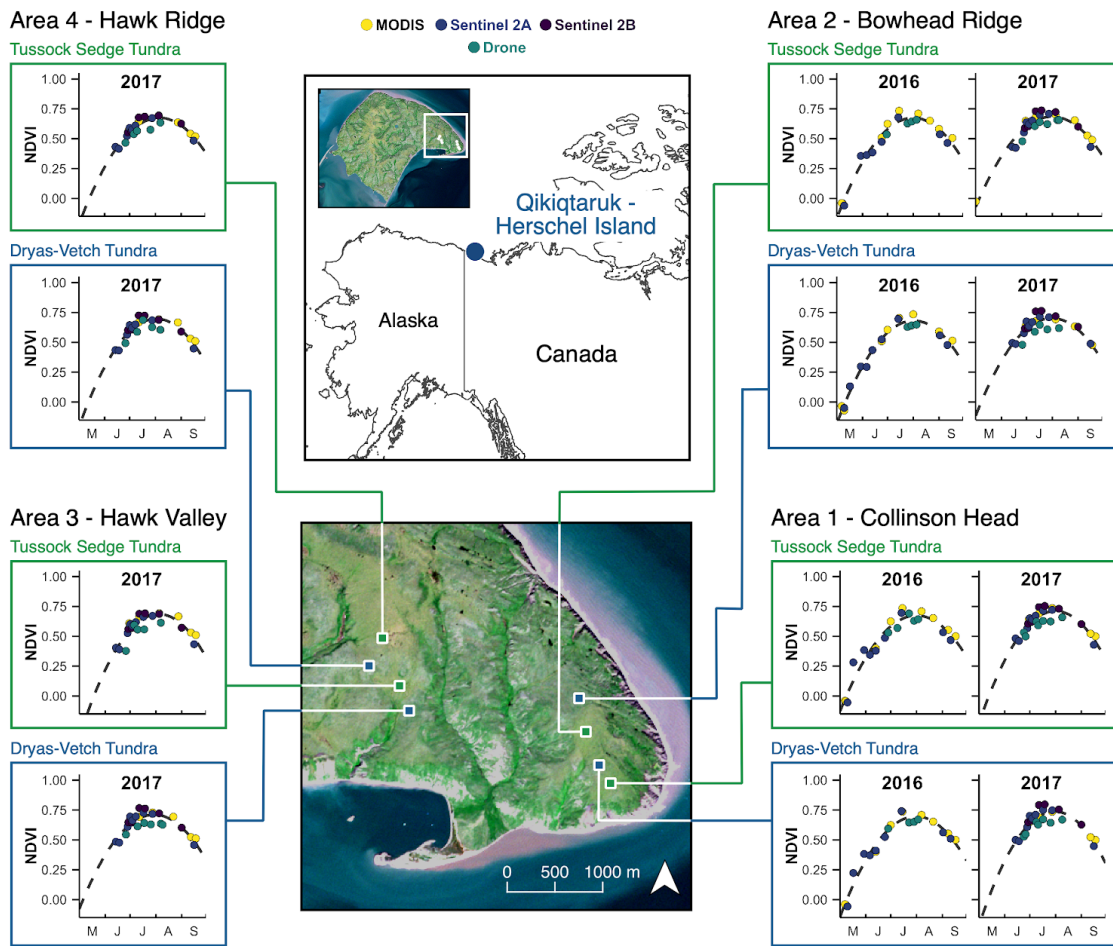
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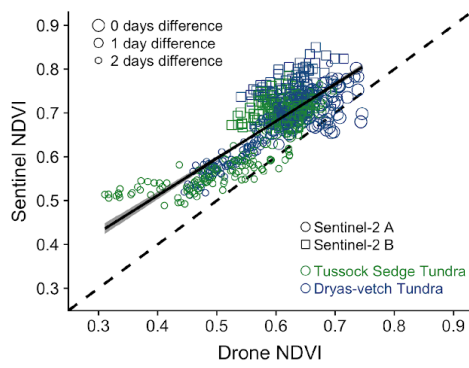
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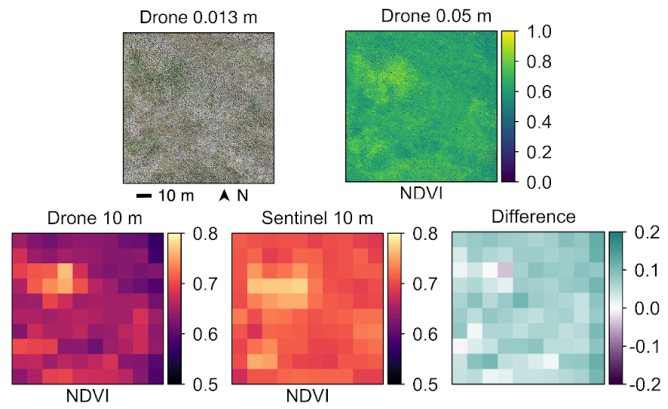
820 **Figure 1:** Drone-data captured the temporal variation in satellite data across vegetation communities, areas and
 821 years. This figure showcases variation in landscape greenness (NDVI) across the one-hectare sampling plots on
 822 Qikiqtaruk and outlines cross-season agreement amongst drone, Sentinel-2 and MODIS sensors. Map sources:
 823 North America (116,117) and Qikiqtaruk, Copernicus Sentinel-2 true colour image July 2017.

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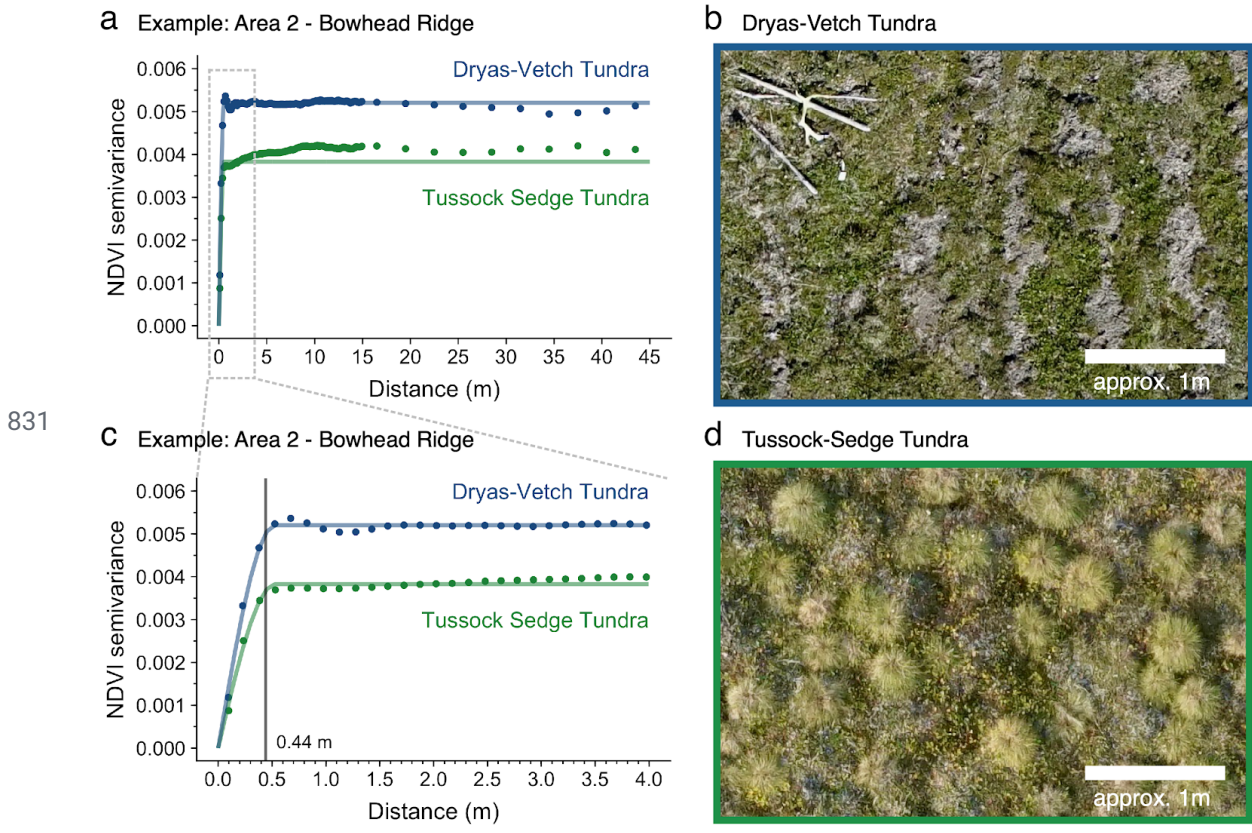
a Drone vs. Sentinel-2 NDVI



b Example: Area 2 - Dryas-vetch Tundra

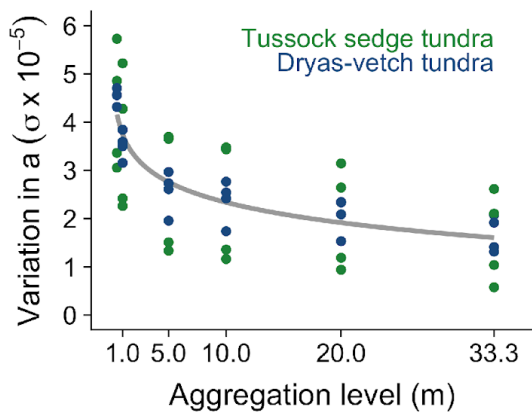


825 **Figure 2:** Drone-data better captured spatial heterogeneity in NDVI relative to Sentinel-2 MSI. a) Pixel by pixel
826 correlations between 10 m aggregated drone NDVI and native 10 m Sentinel-2 NDVI for a random sample of
827 pixels (10% of total pixels, n = 700) across all drone-sentinel image pairs for the 2017 growing season. b)
828 Example visualisations from the Dryas-vetch tundra at Area 2 - Bowhead Ridge showing ultra-fine-grain 0.013 m
829 true colour RGB imagery, 0.05 m native-scale drone NDVI, 10 m resampled drone NDVI, 10 m native Sentinel-2
830 NDVI and the absolute difference between resampled drone and Sentinel-2 NDVI.

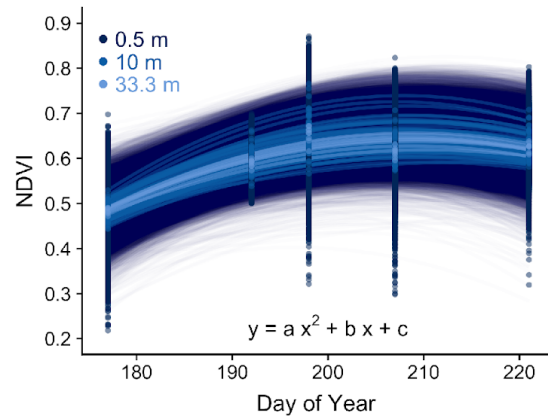


832 **Figure 3:** Spatial variation of vegetation greenness peaked at distances of ~0.5 m in both studied vegetation
 833 types, with little or no increase in the spatial dependence of greenness at distances above ~0.5 m. Figure shows
 834 example variograms. Overall spatial variation in greenness is higher in the Dryas-Vetch Tundra when compared
 835 to the Tussock-Sedge Tundra (a and c). Left panels: variograms for the Dryas-vetch and tussock sedge tundra
 836 plots in Area 2 for distances up to 5 m (a) and 45 m (c) at peak season in 2017. The dark grey line in (c) indicates
 837 the mean range estimated from the variogram models of both vegetation types from Areas 1, 2, and 4 during
 838 peak-season (26 and 28 July) in 2017 (see also Figure S1). Right panels: Dryas-vetch tundra with bare ground
 839 patches caused by cryoturbation and solifluction (c) and tussocks sedge tundra with distinctive patterns of
 840 tussocks interspersed by patches of willows and herbs (d).

a Variation in quadratic coef. with agg. level

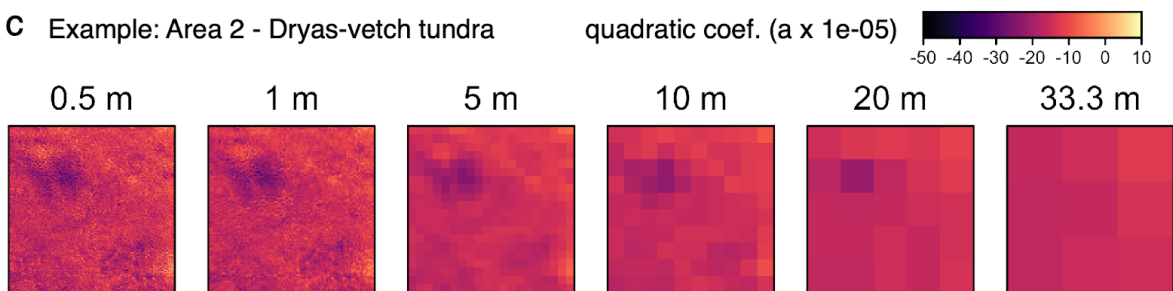


b Example curves: Area 2 - Dryas-vetch tundra

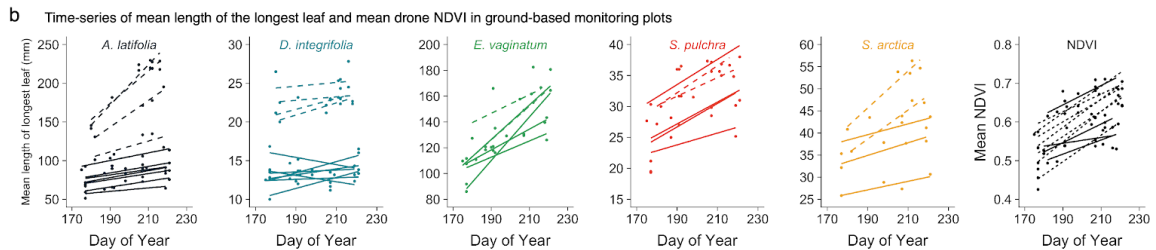
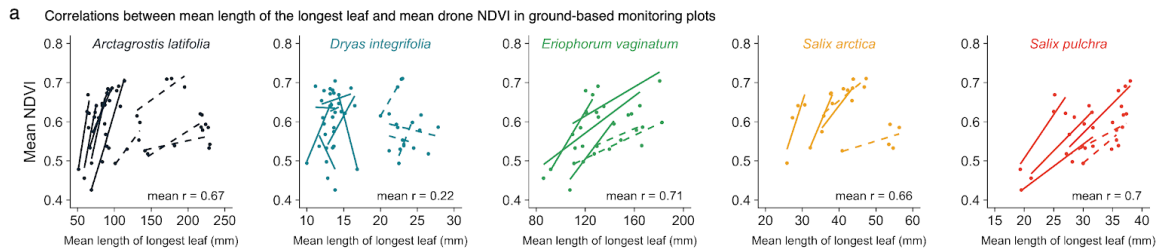


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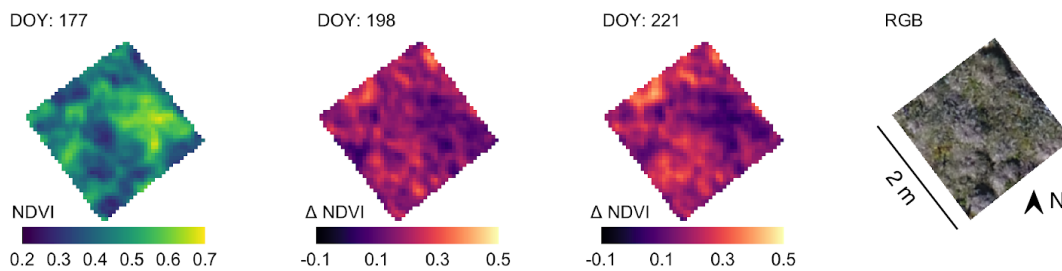
c Example: Area 2 - Dryas-vetch tundra



842 **Figure 4:** Fine-scale variation representing key ecological heterogeneity in tundra phenology was lost when
843 aggregating from ultra-fine-grain drone to medium-grain satellite pixel sizes. When aggregating the drone data
844 across grain sizes, we observed a logarithmic decay in variation (standard deviation) in the quadratic coefficient
845 (shown in a) of simple growing season curves fitted to the eight vegetation plots in the 2017 season. As
846 examples we visualise all curve-fits for the dryas-vetch tundra plot in Area 2 fitted to the time-series of the
847 minimum, central and maximum grain-sizes tested (0.5 m, 10 m and 33.3 m) in (b) and show the spatial
848 distribution of the quadratic coefficient for each grain size for dryas-vetch tundra plot in Area 2 in panel c), similar
849 patterns are found across all areas (a). See Figure S7 for example curves from the dryas vetch tundra plot in
850 Area 2 from all tested grain sizes.



c Example: Drone time-series of the ground-based monitoring plot in Area 2 - Tussock Sedge Tundra in 2017



851

852

853 **Figure 5:** Time-series of ground-based leaf length measurements correlated with drone-derived mean NDVI
 854 across four of five dominant plant species on Qikiqtaruk. Correlation between mean longest leaf length and NDVI
 855 in each 2 m x 2 m ground-phenology plot across all species, areas and seasons (a) and the corresponding
 856 time-series (b). Dashed lines indicate time-series from 2016 and solid lines indicate time-series from 2017. Lines
 857 represent least-square regressions to illustrate the relationships of longest-leaf length and NDVI (a), as well as
 858 day of year (DOY) and the time-series variable (b) for each plot and year combination. Uneven numbers of
 859 time-series between species result as the set of species monitored varied between vegetation types. (c) As an
 860 example, we illustrate the drone-based NDVI observations by showing the start, midpoint and end of the
 861 timeseries for the 2 m x 2 m ground-validation plot in the tussock sedge tundra of Area 2 in 2017. The first
 862 time-point in (c) represents the greenness in the plot at the beginning of the time-series, the two subsequent plots
 863 show the relative difference in greenness to this first observation at the given DOY, and the final plot shows a
 864 true-colour image of the plot taken by drone on the 17 July 2017 (DOY 198).