

1 **Inferred molecular bioavailability of pyrogenic organic matter**
2 **compared to natural organic matter from global sediments and**
3 **surface waters**

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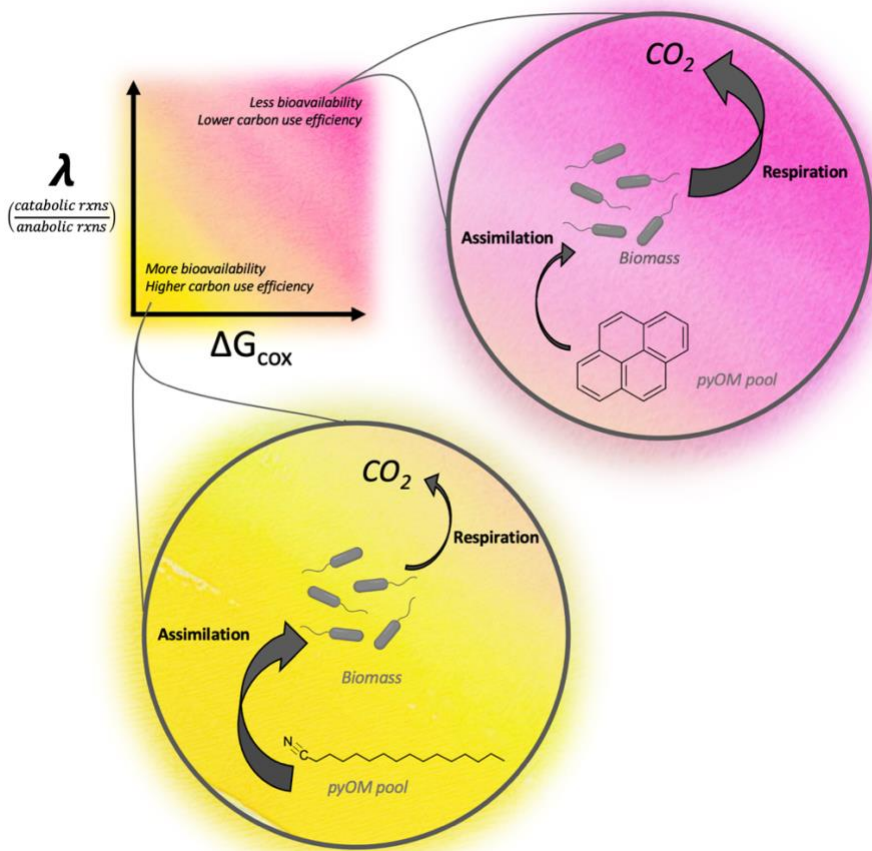
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20 GRAPHICAL ABSTRACT



21

22

23 ABSTRACT

24 Pyrogenic materials generated by wildfires are transported across terrestrial landscapes
25 into inland waters, creating many consequences for aquatic ecosystems. At least ~10% of
26 dissolved organic matter pools may be comprised of pyrogenic organic matter (PyOM), but
27 heterogeneous PyOM chemistry complicates efforts to understand its bioavailability. We used a
28 substrate-explicit model to predict the energy content, metabolic efficiency, and rate of aerobic
29 decomposition of representative PyOM compounds from primary literature. This enabled us to
30 systematically evaluate the bioavailability of a full spectrum of PyOM chemistries – which
31 would be unfeasible with laboratory experiments – and compare it to measured natural organic
32 matter (NOM) pools from global aquatic ecosystems. Overall, we found the range of predicted
33 bioavailability of PyOM was similar to NOM. Thermodynamic properties and carbon use
34 efficiencies of PyOM and NOM were statistically indistinguishable. Within PyOM, phenolic and
35 BC molecules had lower metabolic efficiency than other PyOM and NOM compounds, and the
36 impact of oxygen limitation on BC metabolism was lower than for other PyOM classes. In total,
37 our work supports the recent paradigm shift regarding PyOM bioavailability, highlighting its
38 potential role in global C emissions as the prevalence of wildfires increases.

39 INTRODUCTION

40 Wildfires have burned an average of 1.8-million ha year⁻¹ in the US alone over the past
41 80 years¹ with dramatic impacts on terrestrial and aquatic ecosystem functions.²⁻⁵ The frequency,
42 extent, and severity of wildfire activity in many regions has increased rapidly in response to
43 pervasive hotter and drier conditions associated with climate change,⁶⁻⁹ a trend that is expected to
44 continue.¹⁰ The impact of wildfires on river corridor biogeochemistry has received significant
45 recent attention and is mediated by the loading of sediment and pyrogenic materials into
46 waterways.^{11, 12} In particular, pyrogenic organic matter (PyOM) is a complex continuum of
47 materials generated by thermal alteration of biomass^{13, 14} and can have substantial impacts on
48 river corridor biogeochemistry due to the importance of organic matter as a carbon (C) and
49 energy source in rivers. Globally, wildfires generate 116–385 Tg per year of the most commonly
50 measured constituent of PyOM—black carbon (BC, defined herein as the condensed aromatic
51 fraction of PyOM typically measured by benzene polycarboxylic acid (BPCA) method^{11, 15}). This
52 wildfire generated BC provides a continuous source of PyOM to inland waters that totals 300 to
53 500 giga-metric tons of C stored in sediments, soils, and waters.¹⁶⁻²⁰ Within river corridors, BC
54 alone constitutes ~10% of dissolved organic C pools.¹⁶ Given that organic matter drives
55 biogeochemical cycles in most aquatic ecosystems, the loading of PyOM into river corridors has
56 the potential to produce substantial impacts on ecosystem functions and downstream drinking
57 water treatability.^{21, 22}

58 Historically, PyOM has been considered refractory, passively transported and deposited
59 throughout landscapes. Some estimates of PyOM residence times in aquatic landscapes extend
60 thousands of years.²³⁻²⁶ Yet, simple mass balances suggest depositional reservoirs, including soils
61 and sediments, do not fully account for PyOM production via biomass burning.²⁷ This implies

62 biotic and/or abiotic loss from source to sink.²⁸ This paradigm shift is highlighted by recent work
63 that has shown PyOM may be more bioavailable than previously thought²⁹⁻³¹, and is also
64 supported by biochar research highlighting a variety of reactivity of combustion by-products.^{32, 33}
65 PyOM may constitute a significant and unconstrained contributor to global biogeochemical
66 cycles and climate feedbacks. To address this knowledge gap, we used a substrate-explicit model
67 to predict organic matter energy content, metabolic efficiency, and rate of aerobic decomposition
68 to compare the bioavailability of PyOM to natural organic matter (NOM) in global surface
69 waters and sediments. We hypothesized that the continuum of chemistries in PyOM would show
70 roughly comparable bioavailability to diverse NOM pools.

71 We leverage a newly developed substrate-explicit model³⁴ that parameterizes OM-
72 specific oxidative degradation pathways and reaction rates based on the thermodynamic
73 properties of OM molecules. It relies on the elemental composition of individual OM molecules
74 to provide a systematic way to formulate reaction kinetics and enable flexible application across
75 environments, as it is agnostic of many factors that have complicated a universal understanding
76 of OM bioavailability, including molecular structure (e.g., double bonding and aromatic rings),
77 priming dynamics, chemical inhibition, mineral-associations and physical protection, terminal
78 electron acceptor concentrations, microbial community composition and accessibility, and
79 abiotic reactions (reviewed in Arndt et al.³⁵). The model requires two simple parameters (U_{max}
80 and V_h , see Supporting Information), which focus on the chemical composition of OM as a
81 governing factor in metabolism and overcome a key limitation in modelling aquatic OM
82 decomposition, that variation in parameters across environments limits transferability of model
83 predictions.³⁵ This ecosystem-agnostic approach allows us to examine the molecular
84 bioavailability of individual OM molecules, and it provides a baseline for future work

85 investigating pool-scale dynamics that require a broader range of parameters calibrated for
86 environmental context and targeted laboratory experiments that a examine the bioavailability of a
87 narrower range OM across environmental contexts. In total, our results support an emerging
88 paradigm in wildfire science in which PyOM are more bioavailable than historically considered.

89

90 **MATERIALS AND METHODS**

91 An extended version of our methods is available in the Supporting Information.

92 To assess the bioavailability of PyOM, we searched primary literature for representative
93 compounds of the PyOM continuum. Specifically, we targeted characteristic organic compounds
94 from field and laboratory burns of various fuel types representing a range of moisture,
95 temperature, and oxygen conditions (Table S1). The chosen compounds focused on biomass
96 burning alteration products, which are often used to characterize PyOM in the environment. This
97 included compounds such as theoretical BC compounds (defined here as condensed aromatic
98 core structures polysubstituted with O-containing functionalities³⁶), anhydrosugars, and
99 polycyclic aromatic hydrocarbons (PAHs). The list also included compounds created and/or
100 transformed from biomass burning, such as those derived from biopolymers like lignin (e.g.,
101 methoxyphenols), waxes (e.g., n-alkenes from thermal dehydration of n-alkanols), and resins
102 (e.g., thermally oxidized diterpenoids).^{37, 38} In total, our literature search for PyOM chemistries
103 yielded 389 compounds with 207 unique chemical formulae.

104 After generating a set of representative compounds, we used a substrate-explicit
105 modelling framework developed by Song et al.³⁴ to characterize the bioavailability of each
106 compound and predict its rate of decomposition. Importantly, the model utilizes molecular
107 formulae to predict energetic content, metabolic efficiency, and rates of aerobic metabolism,

108 while it does not account for structural components of organic molecules (e.g., double bonds,
109 folding patterns, cross-linkages). This enabled flexibility in application to high-throughput mass
110 spectrometry techniques that yield chemical formulae but not structural information (e.g., Fourier
111 Transform Ion Cyclotron Resonance Mass Spectrometry, FTICR-MS). Despite its limitations,
112 the substrate-explicit model used here has proven useful in linking NOM composition to aerobic
113 metabolism in natural environments^{34, 39, 40}, and its structure is consistent with Harvey et al.⁴¹
114 who argued for the importance of thermodynamic estimates of PyOM bioavailability that
115 underlie this model. It was chosen to allow for comparison of PyOM to the most comprehensive
116 assessment of global aquatic NOM pools to date.⁴²

117 Briefly, the substrate-explicit model uses the elemental stoichiometry of organic
118 molecules, based on molecular formulae, to predict the number of catabolic reactions that must
119 occur to provide the energy required for the synthesis of one mole of biomass carbon. This
120 quantity is described by the parameter lambda (λ) in which lower λ values denote more efficient
121 energetics of catabolism in producing biomass through anabolism. The model also predicts the
122 Gibbs free energy of C oxidation (ΔG_{Cox}), under standard conditions with a modification to pH 7
123 adjusted from LaRowe and Van Cappellen⁴³ by Song et al.³⁴, as well as C use efficiency (CUE)
124 as defined by Saifuddin et al.⁴⁴ Lower ΔG_{Cox} denotes higher thermodynamic favorability in an
125 electron donor half reaction associated with organic matter, and higher CUE reflects more C
126 assimilated into biomass per unit C respired. We also predicted the rate of aerobic metabolism
127 (as oxygen consumed per mol-C biomass produced) under three scenarios commonly observed in
128 aquatic ecosystems: (a) C-limitation, (b) oxygen (O_2) limitation, and (c) both C and O_2 -
129 limitation. For more details of the substrate-explicit modelling approach used, please see Song et
130 al.³⁴ Each metric (λ , ΔG_{Cox} , CUE, metabolic rates) denotes a different aspect of bioavailability.

131 Though the relative magnitude of the metrics in comparison to each other will vary based on the
132 specific stoichiometry of a molecule, highly bioavailable compounds are indicated by low λ and
133 ΔG_{Cox} coinciding with high CUE and metabolic rates.

134 Three sets of organic molecules were used as model inputs: global dissolved (1) surface
135 water NOM and (2) sediment NOM pools, measured by FTICR-MS as described by Garayburu-
136 Caruso et al.⁴²; and (3) literature-derived PyOM compounds as described above. Inputs to the
137 model from the PyOM compounds were unique molecular formulae, grouped in subsequent
138 analysis by their corresponding compound classes (Table S1). If one molecular formula was
139 represented by several PyOM compounds (e.g., $C_{10}H_{16}O_2$, which corresponds to the
140 sesquiterpenoid cis-Thujan-10-oic acid and 3-, 4- substituted methylcatechol phenols), we
141 assigned multiple compound classes to that molecular formula. Surface water and sediment
142 NOM pools were filtered to compounds occurring in 95% of samples to yield a dataset of
143 globally ubiquitous NOM. Formulae assignment and inferred chemical classes via Van Krevelen
144 diagrams in NOM pools are described by Garayburu-Caruso et al.⁴² We compared modelling
145 outputs from representative PyOM to outputs of ubiquitous NOM pools to infer relative
146 bioavailability using ANOVA and Tukey HSD statistical tests with R software. All model
147 outputs are available in Tables S2-S4. Code is available at:
148 <https://github.com/hyunseobsong/lambd>. Data describing NOM pool chemistry are published as
149 a data package⁴⁵ (available at: doi:10.15485/1729719) and are discussed in more detail by
150 Garayburu-Caruso et al.⁴².

151

152 **RESULTS AND DISCUSSION**

153 We used a substrate-explicit model to evaluate the emerging paradigm of PyOM
154 bioavailability and compared model outputs to comprehensive measurements of global NOM
155 pool composition.^{34, 39, 40, 42} In contrast to previous characterizations of PyOM bioavailability,^{29,}
156 ^{30, 41, 46} the model-based approach enabled us to directly compare known combustion products to
157 thousands of ubiquitous NOM compounds, which would have been unfeasible to directly assess
158 in a laboratory setting.

159

160 *Inferred Molecular Bioavailability of Pyrogenic Organic Matter.*

161 Though previous work has shown that OM chemistry of sediment and surface waters is
162 altered by wildfires,^{11, 16, 47, 48} our results suggest that the chemically distinct pools of OM altered
163 by pyrolysis may have similar overall bioavailability to NOM. We found that the ranges of
164 ΔG_{Cox} , λ , and CUE were similar between PyOM and NOM pools in sediments and surface
165 waters (Figure 1 and 2a). Predicted CUE of PyOM classes was also comparable to literature
166 values reported by others.^{44, 49, 50} Similarly, λ did not vary across all groups of organic molecules
167 or in post-hoc pairwise comparisons (ANOVA $p = 0.09$, and Tukey HSD p (PyOM-sediment) =
168 0.92, p (PyOM-water) = 0.40, p (water-sediment) = 0.10). While ΔG_{Cox} and CUE were
169 significantly different when comparing all three groups (ANOVA, $p < 0.001$), surface water and
170 sediment NOM had greater dissimilarity in these parameters than any comparison involving
171 PyOM. For example, the mean difference in NOM between surface water and sediment was 7.34
172 kJ/mol-C for ΔG_{Cox} and 0.058 for CUE. Comparatively, the differences between PyOM and both
173 surface water and sediment were less than 7.4 kJ/mol-C for ΔG_{Cox} and 0.058 for CUE. Further,
174 there was no evidence that CUE was different between PyOM and sediment NOM (Tukey HSD,
175 $p = 0.20$). These results signal a strong overlap between PyOM bioavailability and NOM pools;

176 however, within PyOM compounds, there was variability in ΔG_{Cox} , λ , and CUE consistent with a
177 heterogeneous continuum of organic matter (Figure 1 and S1). This is not surprising, given the
178 diversity of PyOM chemistries generated by wildfires of different burn severities and source
179 materials,^{11, 51, 52} some of which overlap with chemical classes in unaltered NOM.

180 Interestingly, the relative equivalence of predicted CUE across PyOM and NOM pools
181 revealed that PyOM decomposition in rivers could emit comparable amounts of CO₂ per mole of
182 C to extant NOM pools. CUE is used in many microbially-explicit decomposition models to
183 constrain organic matter bioavailability,⁵³⁻⁵⁹ and predicted CUE from PyOM pools can be easily
184 assimilated in microbially-explicit model predictions that rely on CUE as an input parameter.
185 Such an approach could be used to directly evaluate the impact of PyOM on global C cycles,
186 leading to a better incorporation of PyOM impacts in model predictions.⁶⁰

187 Within PyOM (Figure 1c), two clusters of compounds were distinctly separated from the
188 energetic and metabolic properties of most PyOM. For example, phenols had greater λ values
189 than the majority of PyOM and NOM compounds (all $\lambda > 0.039$, mean $\lambda = 0.084$), while the
190 electron donor half reaction involving BC molecules were less energetically favorable than other
191 PyOM classes (all $\Delta G_{Cox} > -143.42$, mean $\Delta G_{Cox} = -136.40$). However, both were within the
192 range of variability in NOM pools (Figure 1). Phenols and BC also had among the lowest CUE
193 values (BC mean = 0.47 and phenols mean = 0.54). Phenols are traditionally associated with
194 refractory organic matter, such as lignin and tannins, that exhibit long residence times in soils,⁶¹
195 although they have also been reported to be bioavailable in soils and waters in recent years (e.g.,
196 ^{61, 62}). Additionally, BC in this study is defined by inferred aromaticity from ultrahigh resolution
197 mass spectroscopy (i.e., the presence of condensed aromatic structures), which is also considered
198 to have low reactivity.^{26, 36} Although the bioavailability of the phenols and BC classes is

199 consistent with traditionally defined refractory PyOM pools, it is within the bioavailability range
200 observed in NOM, and these compound classes represent only a small portion of the PyOM
201 continuum.^{11, 14} We note that the comparatively low predicted CUE of phenols and BC indicates
202 that, if metabolized, their decomposition could have a greater impact on river corridor CO₂
203 emissions than other PyOM and NOM compounds. As a result, current understanding may
204 substantially underestimate the size, reactivity, and hydrobiogeochemical role of PyOM.¹¹

205

206 *Inferred Metabolism of Pyrogenic Organic Matter.*

207 Predicted rates of PyOM metabolism were also similar to NOM pools (Figure 2b),
208 reinforcing comparable bioavailability between the two pools of organic matter. Pairwise
209 comparison of metabolic rates revealed no evidence for differences between PyOM and sediment
210 NOM under oxygen limitation (Tukey HSD, $p = 0.23$) and without carbon or oxygen limitations
211 (Tukey HSD, $p = 0.34$). However, there was strong evidence that metabolic rates of both PyOM
212 and sediment NOM were different than surface water NOM (Tukey HSD, all $p < 0.001$). Aquatic
213 sediments can reach anoxia within millimeters of the sediment-water interface such that model
214 predictions under oxygen limitation may translate to no meaningful difference between PyOM
215 and NOM in natural sediments. Under C-limitation, PyOM had a statistically elevated rate of
216 metabolism relative to both surface water and sediment NOM (Tukey HSD, all $p < 0.001$).
217 However, we noted only small differences in rate values (means, surface water: 0.13, sediment:
218 0.13, PyOM: 0.17), with a similar range of values in sediment NOM (0.0008–0.45) and PyOM
219 (4.75e-08–0.45). Statistical differences were not surprising given an extremely large sample size
220 for NOM (sediment $n = 398$, surface water $n = 811$), and the low effect sizes denote that overall

221 differences in metabolisms between PyOM and NOM were minimal despite statistical
222 separation.

223 When considering the impact of various elemental limitations on PyOM metabolism, rate
224 predictions were strongly inhibited under low carbon and oxygen conditions. Predicted rates of
225 PyOM metabolism were approximately six times lower when C or O₂ was scarce. Low
226 decomposition rates such as those predicted here under oxygen and C limitation could be one
227 reason for the observed persistence of PyOM in depositional features that tend to be anoxic with
228 refractory C pools.²³⁻²⁵ Still, it is worthwhile to note that metabolism of all PyOM classes under
229 low O₂ or C was predicted to be substantially slower than without elemental limitations,
230 indicating PyOM compounds may both actively cycle in well-oxygenated surface waters with
231 fresh C inputs²⁸ and persist over long periods of time in O₂-limited sediments.

232 Among PyOM compound classes, BC was less negatively impacted by oxygen limitation
233 than any other group (Figure S2). Previous work has demonstrated that microorganisms are
234 capable of decomposing chemically complex organic molecules, such as long-chained and/or
235 aromatic hydrocarbons under low oxygen availability and/or anaerobia.⁶³⁻⁶⁶ Similar microbial
236 metabolic pathways may also be capable of degrading BC molecules in natural settings and
237 could be investigated with future laboratory work. Notably, our work also supports the notion
238 that black nitrogen could be more bioavailable than BC. We posit this may be due, in part, to its
239 chemical structure that includes pyrrole-type moieties, which are relatively biodegradable.^{67, 68}
240 While we only examined one class of PyOM molecules containing nitrogen (n-alkylnitriles), it
241 had among the highest predicted CUE and rates of aerobic metabolism.

242

243 *Correspondence to Empirical Investigations.*

244 While the substrate-explicit modelling approach used here has been validated in natural
245 settings and enabled comparison to a rich suite of ubiquitous NOM molecules, one key limitation
246 is its inability to account for structural characteristics of organic matter. Indeed, some aspects of
247 model predictions are inconsistent with experimental evidence, highlighting the role of
248 laboratory studies in evaluating PyOM bioavailability. For instance, n-alkenes and related
249 compounds tended to be most favorable for metabolism, despite these compound classes having
250 high stability in the environment and common usage as paleo-proxies in soils and sediments
251 (e.g., ^{69, 70}). These compounds are characterized by carbon-carbon double bond functional
252 groups, which may decrease bioavailability and are not considered in the model predictions.
253 Fatty acids and n-alkanes biomass burning by-products are generally reduced in chain length in
254 comparison to their un-burned counterparts⁷¹, and thus may be relatively bioavailable compared
255 to other lipids. Additionally, we note that previous work has shown fast degradation of
256 combustion-derived lipids in soils;⁷¹ as well as high n-alkene metabolism under anaerobic
257 conditions (in particular by sulfate-reducing bacteria⁷²⁻⁷⁵), n-alkene metabolism in natural
258 sediments,^{75, 76} and a range in lipid reactivities at the sediment-water interface.⁷⁷ While work on
259 n-alkene metabolism in aerobic settings is limited, the comparative bioavailability of n-alkenes
260 and known microbial degradation pathways suggests diverse microbiomes in sediments may
261 metabolize these compounds as part of natural biogeochemical cycles. Another notable
262 discrepancy is the relatively low bioavailability of anhydrosugars when compared to other PyOM
263 compounds. Experimentally, anhydrosugars are highly bioavailable in oxic conditions, with a
264 half-life of less than seven days.³⁰ The model may therefore not adequately account for enzyme-
265 catalyzed reactions, such as levoglucosan kinase or levoglucosan dehydrogenase, which cleave
266 and phosphorylate the 1,6-intramolecular linkage,⁷⁸ and could be potentially common enzymes

267 utilized by aquatic microorganisms.⁷⁹ Because of these nuances, the PyOM compound classes
268 presented here are best used as bounding estimates for experimental validation, and for holistic
269 comparison to NOM bioavailability. Still, the span of compounds investigated here, and their
270 comparison to NOM pools, provides a breadth of PyOM investigation that is unfeasible without
271 model-based approaches.^{29, 30, 41, 46}

272

273 *Conclusions.*

274 Our data supports the recent paradigm shift towards high PyOM bioavailability and
275 provides a foundation for detailed laboratory experiments investigating specific components of
276 the PyOM continuum. Intensifying wildfire regimes in many parts of the world are increasing the
277 production of PyOM with potential implications for source water supplies, which are critical for
278 domestic, industrial, agricultural, and ecological needs. Yet, many fundamental questions, such
279 as “how much” PyOM exists in ecosystems, “how fast” it cycles, and “how old” it is remain
280 largely unknown, and are an active area of research.⁸⁰ Furthermore, there has been no systematic
281 evaluation of the bioavailability of different constituents within the heterogeneous pools that
282 comprise the PyOM continuum.⁸¹ Our work provides the first comprehensive computational
283 assessment of the bioavailability of various chemical classes of PyOM in comparison to NOM
284 pools. The comparable bioavailability to NOM revealed that PyOM may be actively transformed
285 within the river corridor and may be an increasing source of C emissions to the atmosphere in the
286 future as the prevalence of wildfires increases.

287

288 **AUTHOR INFORMATION**

289

290 EBG conceived of the manuscript and was responsible for writing the manuscript and drafting all
291 figures. HSS performed all modelling. SG determined PyOM compounds for modelling based on
292 extensive literature review, with guidance from AMP. VGC and JCS contributed data and insight
293 on NOM pool chemistry. All authors contributed to revisions.

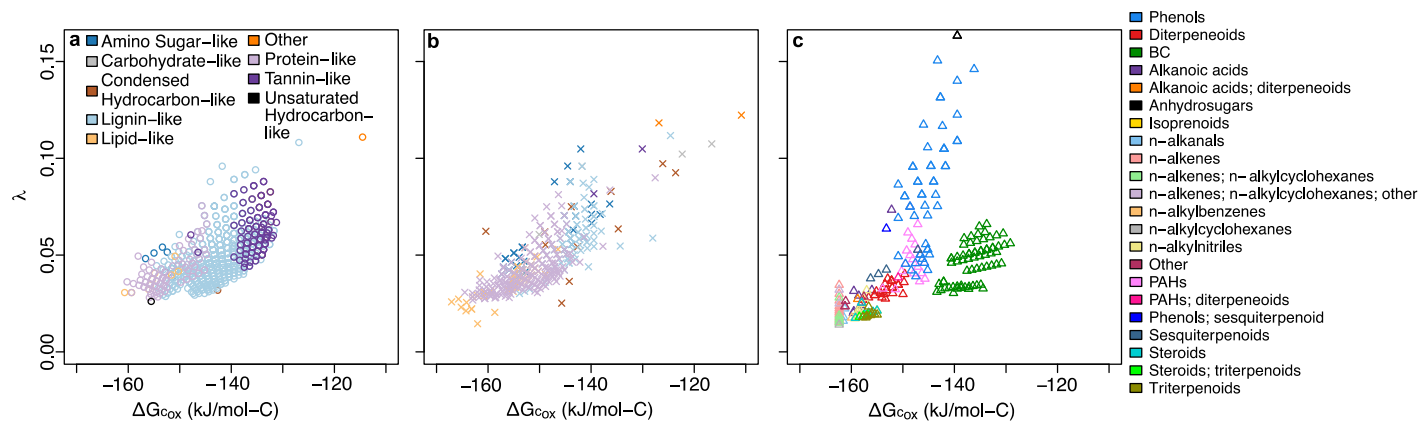
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304 Environmental Molecular Science Laboratory User Facility.

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307

308 **Figure 1.** Comparison of PyOM energy content (ΔG_{Cox}) and metabolic efficiency (λ) to global

309 NOM. Ubiquitous NOM molecules detected via FTICR-MS in global (a) surface water and (b)

310 sediment are colored by inferred chemical class. (c) Representative PyOM molecules are colored

311 by known chemical properties. Because PyOM molecules were derived from primary literature,

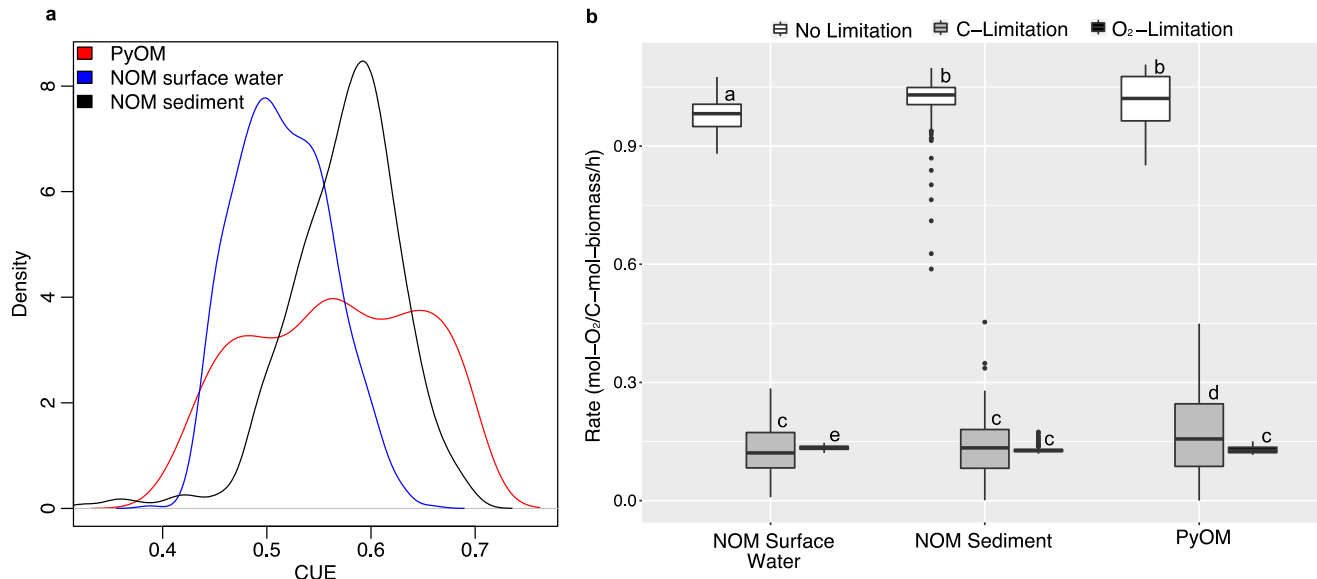
312 we are able to assign chemical properties at higher resolution than inferred classes from

313 measured NOM pools. Details on inferred chemical class assignment are provided in the

314 Supporting Information. Legends are inset in (a) for (a) and (b), and to the right of (c).

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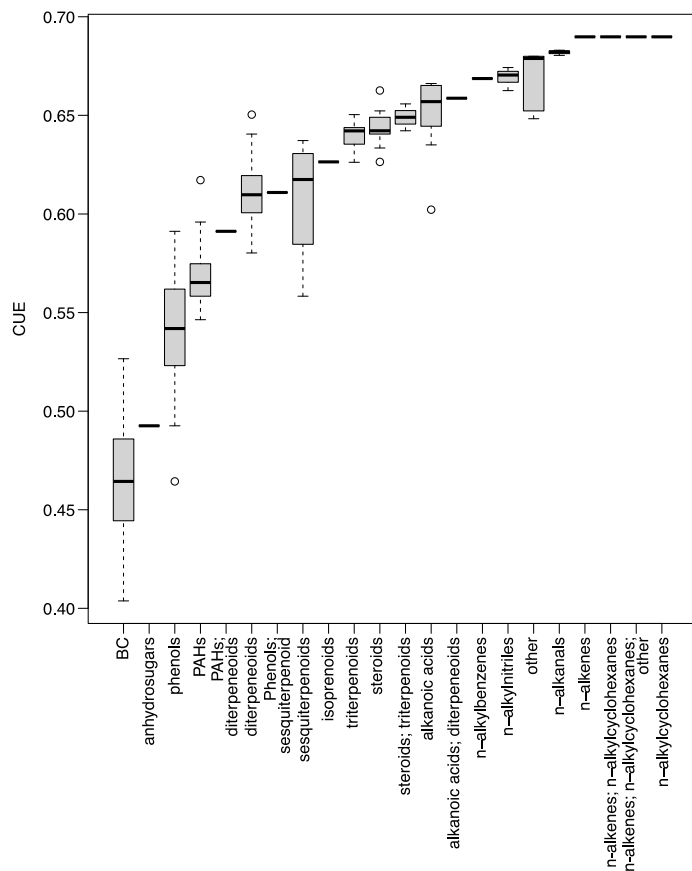
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317

318 **Figure 2.** Carbon use efficiency (CUE) and predicted metabolism of NOM and PyOM. (a)
 319 shows the distribution of CUE in PyOM (red) and ubiquitous surface water (blue) and sediment
 320 (black) NOM molecules rates. It shows a probability density function (PDF) that reflects the
 321 relative likelihood that value of a random sample drawn from a particular group would equal the
 322 value on the x-axis. (b) depicts the predicted metabolism of surface water NOM, sediment NOM,
 323 and PyOM. Letters in (b) denote statistical groups. Median values are denoted by a bar, the
 324 lower and upper hinges correspond to the first and third quartiles (the 25th and 75th percentiles),
 325 and the upper and lower whiskers extend from the hinge to the largest/smallest value no further
 326 than 1.5 * IQR from the hinge (where IQR is the inter-quartile range, or distance between the
 327 first and third quartiles), and data beyond the end of the whiskers are plotted individually.

328 SUPPLEMENTAL FIGURES



329

330 **Figure S1.** Carbon use efficiency (CUE) of PyOM, grouped by known chemical properties. High

331 CUE values indicate more C incorporated into biomass vs. respired per unit C. Median values

332 are denoted by a bar, the lower and upper hinges correspond to the first and third quartiles (the

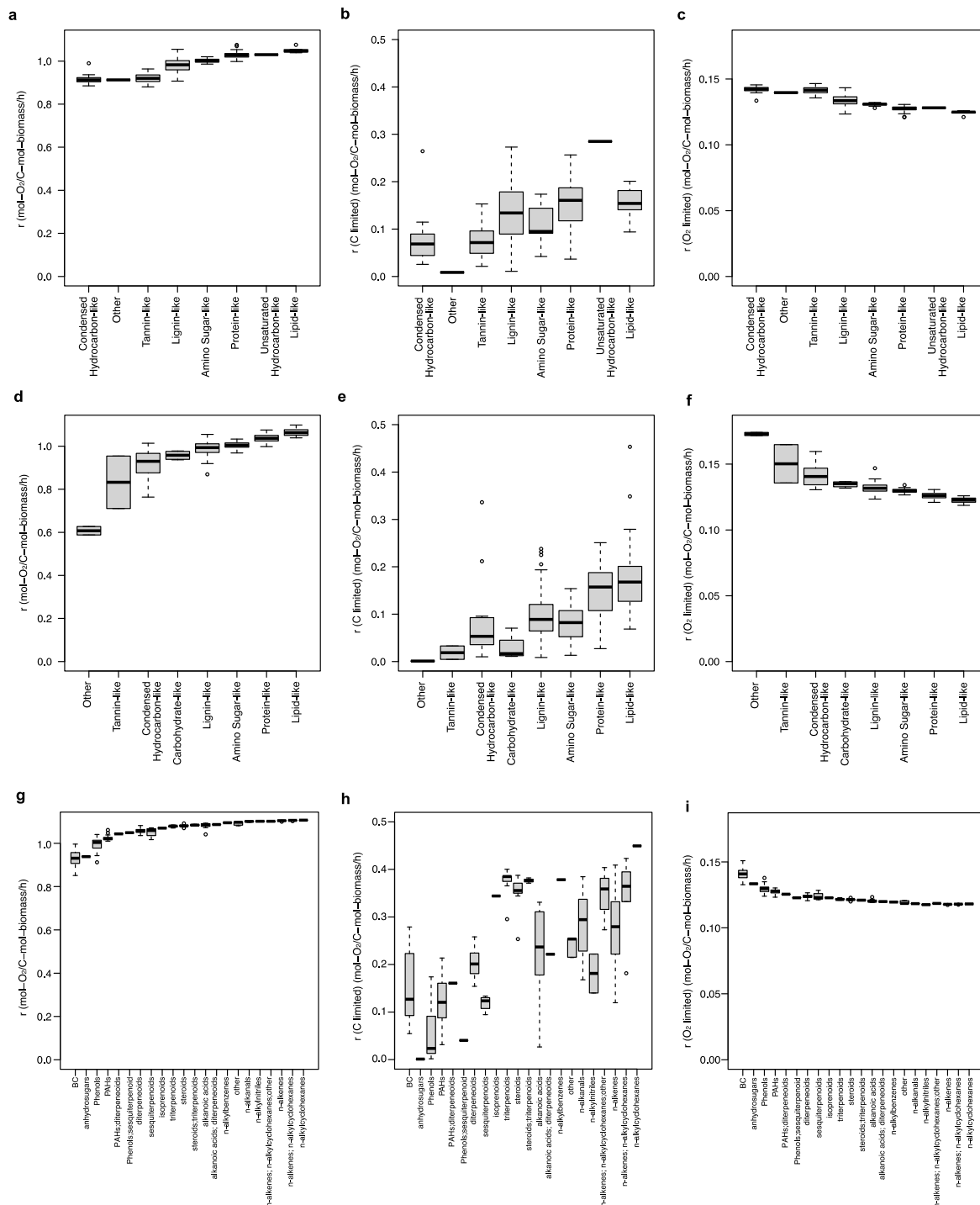
333 25th and 75th percentiles), and the upper and lower whiskers extend from the hinge to the

334 largest/smallest value no further than $1.5 * \text{IQR}$ from the hinge (where IQR is the inter-quartile

335 range, or distance between the first and third quartiles), and data beyond the end of the whiskers

336 are plotted individually.

337



338

339 **Figure S2.** Predicted metabolism of ubiquitous (a–c) surface water NOM, (d–f) sediment NOM,
 340 and (g–i) PyOM molecules, grouped by inferred chemical class or known chemical properties.

341 The first column shows predicted rates of aerobic metabolism without any elemental limitations.

342 C-limited and O₂-limited scenarios are shown in the second and third columns, respectively.

343 Median values are denoted by a bar, the lower and upper hinges correspond to the first and third

344 quartiles (the 25th and 75th percentiles), and the upper and lower whiskers extend from the hinge

345 to the largest/smallest value no further than $1.5 * IQR$ from the hinge (where IQR is the inter-

346 quartile range, or distance between the first and third quartiles), and data beyond the end of the

347 whiskers are plotted individually.

348

349 **Table S1.** PyOM molecules and chemical properties.

350

351 **Table S2.** Substrate-explicit model outputs for each PyOM compound.

352

353 **Table S3.** Substrate-explicit model outputs for each sediment NOM compound.

354

355 **Table S4.** Substrate-explicit model outputs for each surface water NOM compound.

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