

1 Title

2 Idea Paper: improving forecasts of community composition with lightweight biodiversity  
3 monitoring across ecological and anthropogenic disturbance gradients

4  
5 Authors

6 Jamie M. Kass (corresponding author)

7 ORCID: 0000-0002-9432-895X

8 Biodiversity and Biocomplexity Unit, Okinawa Institute of Science and Technology Graduate  
9 University, 1919-1 Tancha, Onna, Okinawa, 904-0495 Japan

10 email: jamie.m.kass@gmail.com

11

12 Nao Takashina

13 ORCID: 0000-0002-9594-9264

14 Department of International Studies, The University of Tokyo, 5-1-5 Kashiwa, Chiba 277-0459,  
15 Japan

16

17 Nicholas R. Friedman

18 ORCID: 0000-0002-0533-6801

19 Environmental Informatics Section, Okinawa Institute of Science and Technology Graduate  
20 University, 1919-1 Tancha, Onna, Okinawa 904-0495, Japan

21

22 Buntarou Kusumoto

23 ORCID: 0000-0002-5091-3575

24 Kasuya Research Forest, Faculty of Agriculture, Kyushu University, 394 Tsubakuro, Sasaguri,  
25 Fukuoka, 811-2415 Japan

26

27 Mary E. Blair

28 ORCID: 0000-0003-0139-020X

29 Center for Biodiversity and Conservation, American Museum of Natural History, New York,  
30 NY, USA

31

32 Abstract

33 Accurate and up-to-date biodiversity forecasts enable robust planning for environmental  
34 management and conservation of landscapes under a wide range of uses. Future predictions of  
35 the species composition of ecological communities complement more frequently reported species  
36 richness estimates to better characterize the different dimensions of biodiversity. The models that  
37 make community composition forecasts are calibrated with data on species' geographic patterns  
38 for the present, which may not be good proxies for future patterns. The future establishment of  
39 novel communities represents data on species interactions unaccounted for by these models.  
40 However, detecting them in a systematic way presents challenges due to the lack of monitoring  
41 data for landscapes with high environmental turnover, where such communities are likely to  
42 establish. Here, we propose lightweight monitoring over both ecological and anthropogenic  
43 disturbance gradients using passive sensors (i.e., those that operate continuously without much  
44 human input) to detect novel communities with the aim of updating models that make  
45 community composition forecasts. Monitoring over these two gradients should maximize  
46 detection of novel communities and improve understanding of relationships between community  
47 composition and environmental change. Further, barriers regarding cost and effort are reduced by  
48 using relatively few sensors requiring minimal upkeep. Ongoing updates to community  
49 composition forecasts based on novel community data and better understanding of the associated  
50 uncertainty should improve future decision-making for both resource management and  
51 conservation efforts.

52

53 Keywords

54 community, global change, ideas for fundamental questions, species distribution modeling,  
55 observation, species interactions

56

57 Research question

58 How can we improve forecasts of community composition under ongoing global change without  
59 prohibitive cost or effort?

60

61 Value

62 Where species will be in the future, which will be found together, and how they will respond to  
63 ongoing global change are foundational questions for ecosystem management and conservation.  
64 Accurate forecasts of biodiversity enable better planning for allocation of resource use,  
65 maintenance of ecosystem services, response to invasive species, and establishment of protected  
66 areas (Newbold et al. 2015). Particularly, forecasts made across environmental gradients can lead  
67 to more informed management for areas that are natural (e.g., set aside for conservation),  
68 unnatural (e.g., under resource management), or semi-natural (e.g., socio-ecological production  
69 landscapes such as *satoyama*). However, models that make biodiversity forecasts can estimate  
70 inaccurate future environmental responses for species and communities if current data on  
71 species' geographic patterns are not reasonable proxies for future patterns (Blois et al. 2013a).  
72 Future colonizations of newly suitable areas by species, leading to the establishment of  
73 communities with new species combinations (i.e., novel communities), represent heretofore  
74 unknown information about community composition and species interactions. Novel  
75 communities, which emerge more frequently with rapid landscape change (Finsinger et al. 2017),  
76 can expand or alter existing species interactions (Williams & Jackson 2007) and lead to  
77 disruptions in ecosystem services (Hobbs et al. 2006). Ongoing detection and documentation of  
78 novel communities is thus of vital importance to update community composition predictions for  
79 biodiversity forecasts.

80

81 Unfortunately, locating novel communities in a systematic way is a challenging exercise due in  
82 large part to the lack of monitoring data across landscapes with high environmental turnover,  
83 where such communities are likely to establish. Many efforts to catalog biodiversity focus on  
84 flagship species or single estimates of species richness. On the other hand, biodiversity  
85 monitoring systems provide high-frequency data streams and could also be harnessed to update  
86 community forecasts (Bush et al. 2017). However, the high economic and labor costs can be  
87 prohibitive, and existing systems often have bias in their spatial coverage and representation of

88 environmental heterogeneity (Metzger et al. 2013). Moreover, both conservation efforts (e.g.,  
89 establishing protected areas; Kusumoto et al. 2017) and monitoring systems tend to prioritize  
90 specific ecosystems rather than the gradients between them (Fig. 1). The resulting lack of  
91 regularly updated data on novel communities can be seen as a bottleneck for prediction accuracy  
92 of community composition forecasts.

93

#### 94 Relevant hypotheses

95 Community composition forecasts are typically made using space-for-time substitutions, which  
96 employ modeled relationships between the current geographic patterns of species and  
97 communities with environmental variables to predict future patterns (Blois et al. 2013b). These  
98 predictions are most often made using either stacked or joint species distribution models  
99 (SDMs). The stacked SDM approach involves combining the predictions of SDMs fit with  
100 abiotic variables for individual species to estimate the community composition per spatial grid  
101 cell, then applying a biotic filter by removing unlikely candidate species from predicted  
102 communities based on knowledge of species interactions (SESAM; Guisan & Rahbek 2011). The  
103 joint SDM approach models multiple species' distributions together in the same modeling  
104 framework and typically measures interaction strength based on co-occurrence correlations after  
105 accounting for shared environmental preferences (joint SDMs; Warton et al. 2015).

106

107 But as changes to known species interaction networks will alter community predictions for both  
108 approaches, information on novel communities is crucial to improve the accuracy of long-term  
109 community forecasts. Although some data exists on current novel community establishment,  
110 ongoing global change will result in further shifts to present-day community structure. As  
111 species turnover tends to increase with environmental heterogeneity (Buckley and Jetz 2008),  
112 monitoring over gradients between different environments should be ideal for detecting existing  
113 novel communities and their future emergence.

114

#### 115 New research idea

116 We propose monitoring over environmental gradients to detect novel communities and use this  
117 information to update community composition forecasts. Long-term monitoring over  
118 environmental gradients at a fine temporal resolution has a high likelihood of detecting novel

119 community establishment (Fig. 1). Such detections represent information that can be used to  
120 identify mismatches between observations and predictions, leading to changes in community  
121 composition forecasts that rely on known relationships between species. It is important to note  
122 that this approach cannot determine the causes of novel community emergence, as other casual  
123 factors that correlate with environmental gradients may exist.

124

125 Our proposal includes monitoring over two main gradients driving community change:  
126 ecological (e.g., forest to grassland, lowland to montane forest) and anthropogenic disturbance  
127 (e.g., forest to urban, primary to logged forest). Importantly, these two gradients have different  
128 temporal resolution: anthropogenic disturbance (e.g., land cover conversion) occurs on shorter  
129 timescales than natural ecosystem change. Monitoring over these gradients should both  
130 maximize detection of novel communities and improve understanding of community  
131 relationships with environmental change. Real-time predictions of change for heterogeneous  
132 landscapes using remote sensing data have exciting potential (e.g., Slingsby et al. 2020). There  
133 are also examples of in situ monitoring over natural gradients to increase coverage of conditions  
134 and detect species' niche shifts (Carvalho et al. 2016). However, to our knowledge this is the  
135 first proposal to use in situ monitoring data from both ecological and anthropogenic disturbance  
136 gradients to make real-time updates to community forecasts.

137

### 138 How to tackle the question through the proposed new idea

139 Here, we advocate for the implementation of lightweight, in situ monitoring systems composed  
140 of transects over ecological and anthropogenic disturbance gradients in representative ecoregions  
141 around the world to collect structured data on novel communities. To reduce cost and effort,  
142 these transects can be composed of strategically placed passive sensors (i.e., those that operate  
143 continuously without much human input) including camera traps or audio recorders (although we  
144 discuss monitoring of terrestrial, relatively motile animals, the concept could be applied to other  
145 systems). Large camera trap grids monitored over short time periods provide a good balance of  
146 cost and accuracy for modeling species occupancy and richness, though accounting for  
147 seasonality presents challenges (Kays et al. 2020). In contrast, the lightweight system we propose  
148 serves only to collect detection data rather than produce datasets for modeling, utilizes different  
149 sensors to improve detection and expand the taxa sampled, and addresses seasonality by

150 monitoring over annual cycles (Ross et al. 2018). Hence, fewer sensors in smaller grids or  
151 transects can be sufficient. Information on novel communities can directly improve community  
152 composition forecasts by updating species interaction information used to correct stacks of  
153 individual-species SDMs (Fig. 2), or by updating co-occurrence patterns used by joint SDMs.  
154 Additionally, these structured data can be combined with unstructured datasets (i.e.,  
155 opportunistic presence data) using data integration approaches to improve the accuracy of large-  
156 scale range models (Isaac et al. 2020).

157

158 As uncertainty remains an issue for community composition forecasts made over space and time,  
159 it must be recognized that model predictions need validation and should be interpreted with  
160 caution. Although the modeling approaches discussed above can serve as diagnostic tools to help  
161 generate hypotheses about community assembly processes, both have existing methodological  
162 limitations (Zurell et al. 2020). Further, presence of species in new areas could reflect itinerant  
163 dispersal or short-lived sink populations rather than self-sustaining populations, and models of  
164 global climate and anthropogenic change used to make community forecasts have inherently  
165 high uncertainty. Thus, field validation (i.e., community inventories to evaluate model  
166 performance) and consideration of ancillary data (i.e., population studies to confirm novel  
167 community establishment) are advisable before any management or policy decisions are made. It  
168 is important to note that uncertainty in community forecasts can also be harnessed to estimate  
169 different scenarios of community dynamics across a landscape.

170

171 The importance of updated forecasts of (novel) ecological communities and how best to apply  
172 them to improve future decision-making is crucial to demonstrate to managers and conservation  
173 practitioners. For example, we expect forecasts to show how ecological gradients themselves  
174 should be considered high priority areas for monitoring and conservation actions due to their  
175 ability to maintain processes that foster evolutionary potential and ecological resilience to change  
176 (Blair et al. 2013).

177

### 178 Motivation

179 The ideas we propose here originate from discussions between the authors about how to improve  
180 community composition forecasts using new biodiversity monitoring strategies, and we wanted

181 to share these ideas with the field to spark new ways forward for predicting community change.  
182 We intend to use the ideas in this paper as a foundation for future grant proposals.

183

#### 184 Acknowledgments

185 We thank Robert P. Anderson, Ana C. Carnaval, Damaris Zurell, and Roland Kays for critical  
186 comments that helped improve this manuscript. JMK is supported by the Japan Society for the  
187 Promotion of Science (JSPS) Postdoctoral Fellowships for Foreign Researchers Program. NRF  
188 and JMK are supported by subsidy funding to the Okinawa Institute of Science and Technology  
189 Graduate University. NT is supported by JSPS KAKENHI Grant Number 21K17913.

190

#### 191 Conflict of Interest

192 The authors have no conflicts of interest to report regarding this paper.

193

#### 194 References

- 195 1. Blair, M. E., Sterling, E. J., Dusch, M., Raxworthy, C. J., & Pearson, R. G. (2013).  
196 Ecological divergence and speciation between lemur (*Eulemur*) sister species in  
197 Madagascar. *Journal of Evolutionary Biology*, 26, 1790-1801. doi: 10.1111/jeb.12179
- 198 2. Blois, J. L., Zarnetske, P. L., Fitzpatrick, M. C., & Finnegan, S. (2013a). Climate change  
199 and the past, present, and future of biotic interactions. *Science*, 341, 499-504. doi:  
200 10.1126/science.1237184
- 201 3. Blois, J. L., Williams, J. W., Fitzpatrick, M. C., Jackson, S. T., & Ferrier, S. (2013b).  
202 Space can substitute for time in predicting climate-change effects on biodiversity.  
203 *Proceedings of the National Academy of Sciences*, 110, 9374-9379. doi:  
204 10.1073/pnas.1220228110
- 205 4. Buckley, L. B., & Jetz, W. (2008). Linking global turnover of species and environments.  
206 *Proceedings of the National Academy of Sciences*, 105, 17836-17841. doi:  
207 10.1073/pnas.0803524105
- 208 5. Bush, A., Sollmann, R., Wilting, A., Bohmann, K., Cole, B., Balzter, H., ... & Douglas,  
209 W. Y. (2017). Connecting Earth observation to high-throughput biodiversity data. *Nature*  
210 *Ecology & Evolution*, 1, 1-9. doi: 10.1038/s41559-017-0176

- 211 6. Carvalho, S. B., Gonçalves, J., Guisan, A., & Honrado, J. P. (2016). Systematic site  
212 selection for multispecies monitoring networks. *Journal of Applied Ecology*, 53, 1305-  
213 1316. doi: 10.1111/1365-2664.12505
- 214 7. Finsinger, W., Giesecke, T., Brewer, S., & Leydet, M. (2017). Emergence patterns of  
215 novelty in European vegetation assemblages over the past 15 000 years. *Ecology Letters*,  
216 20, 336-346. doi: 10.1111/ele.12731
- 217 8. Guisan, A., & Rahbek, C. (2011). SESAM—a new framework integrating macroecological  
218 and species distribution models for predicting spatio-temporal patterns of species  
219 assemblages. *Journal of Biogeography*, 38, 1433-1444. doi: 10.1111/j.1365-  
220 2699.2011.02550.x
- 221 9. Hobbs, R. J., Arico, S., Aronson, J., Baron, J. S., Bridgewater, P., Cramer, V. A., ... &  
222 Zobel, M. (2006). Novel ecosystems: theoretical and management aspects of the new  
223 ecological world order. *Global Ecology and Biogeography*, 15, 1-7. doi: 10.1111/j.1466-  
224 822X.2006.00212.x
- 225 10. Isaac, N. J., Jarzyna, M. A., Keil, P., Dambly, L. I., Boersch-Supan, P. H., Browning,  
226 E., ... & O'Hara, R. B. (2020). Data integration for large-scale models of species  
227 distributions. *Trends in Ecology & Evolution*, 35, 56-67. doi: 10.1016/j.tree.2019.08.006
- 228 11. Kays, R., Arbogast, B. S., Baker-Whatton, M., Beirne, C., Boone, H. M., Bowler, M., ...  
229 & Spironello, W. R. (2020). An empirical evaluation of camera trap study design: How  
230 many, how long and when?. *Methods in Ecology and Evolution*, 11, 700-713. doi:  
231 10.1111/2041-210X.13370
- 232 12. Kusumoto, B., Shiono, T., Konoshima, M., Yoshimoto, A., Tanaka, T., & Kubota, Y.  
233 (2017). How well are biodiversity drivers reflected in protected areas? A  
234 representativeness assessment of the geohistorical gradients that shaped endemic flora in  
235 Japan. *Ecological Research*, 32, 299-311. doi: 10.1007/s11284-017-1451-6
- 236 13. Metzger, M. J., Brus, D. J., Bunce, R. G. H., Carey, P. D., Gonçalves, J., Honrado, J.  
237 P., ... & Zomer, R. (2013). Environmental stratifications as the basis for national,  
238 European and global ecological monitoring. *Ecological Indicators*, 33, 26-35. doi:  
239 10.1016/j.ecolind.2012.11.009



- 240 14. Newbold, T., Hudson, L. N., Hill, S. L., Contu, S., Lysenko, I., Senior, R. A., ... &  
241 Purvis, A. (2015). Global effects of land use on local terrestrial biodiversity. *Nature*, 520,  
242 45-50. doi: 10.1038/nature14324
- 243 15. Ross, S. R. J., Friedman, N. R., Dudley, K. L., Yoshimura, M., Yoshida, T., & Economo,  
244 E. P. (2018). Listening to ecosystems: data-rich acoustic monitoring through landscape-  
245 scale sensor networks. *Ecological Research*, 33, 135-147. doi: 10.1007/s11284-017-  
246 1509-5
- 247 16. Slingsby, J. A., Moncrieff, G. R., & Wilson, A. M. (2020). Near-real time forecasting and  
248 change detection for an open ecosystem with complex natural dynamics. *ISPRS Journal*  
249 *of Photogrammetry and Remote Sensing*, 166, 15-25. doi: 10.1016/j.isprsjprs.2020.05.017
- 250 17. Warton, D. I., Blanchet, F. G., O'Hara, R. B., Ovaskainen, O., Taskinen, S., Walker, S.  
251 C., & Hui, F. K. (2015). So many variables: joint modeling in community ecology.  
252 *Trends in Ecology & Evolution*, 30, 766-779. doi: 10.1016/j.tree.2015.09.007
- 253 18. Williams, J. W., & Jackson, S. T. (2007). Novel climates, no-analog communities, and  
254 ecological surprises. *Frontiers in Ecology and the Environment*, 5, 475-482. doi:  
255 10.1890/070037
- 256 19. Zurell, D., Zimmermann, N. E., Gross, H., Baltensweiler, A., Sattler, T., & Wüest, R. O.  
257 (2020). Testing species assemblage predictions from stacked and joint species  
258 distribution models. *Journal of Biogeography*, 47, 101-113. doi: 10.1111/jbi.13608

259 Figure 1. Comparison of different passive data collection approaches in terms of cost (financial  
260 and labor) and detecting novel communities, designated on a scale of “low” (yellow), “medium”  
261 (orange), and “high” (red). Opportunistic occurrence data is low-cost to acquire, but represents  
262 an amalgam of various sampling areas, times, and methodologies, making it difficult to detect  
263 the confirmed establishment of novel communities. Grid arrays ( $n \times n$ ) of passive sensors for a  
264 single ecosystem collect structured data that can more reliably detect community occupancy yet  
265 are high-cost. Alternatively, a linear array ( $n \times 1$ ) of passive sensors collects structured data  
266 along a gradient and samples areas with greater environmental heterogeneity, which should  
267 result in higher detectability of novel communities with a lower cost.

268

269 Figure 2. A conceptual example of how novel community detection can change landscape-level  
270 community composition forecasts on short timescales (no significant climate change, but  
271 possible ecological or anthropogenic change). This example was created using range estimates of  
272 virtual species in R and real data on climatic variables and Japanese ecoregions (Supplemental  
273 Information 1, 2). Two linear arrays of passive sensors monitor surrounding areas (black circles)  
274 across the gradient between two ecoregions (west: dark gray, east: light gray) delineated by a  
275 white dotted line. Species A occurs across both the east and west ecoregions, while species C  
276 occurs only in the east due to abiotic constraints. In time 1 (Panel (a)), species B occurs only in  
277 the west due to biotic constraints. Thus, although the fundamental niche of species B includes  
278 areas in the east ecoregion, it is not included in the eastern community predictions. In time 2  
279 (Panel (b)), the biotic constraint is released and species B can colonize parts of the east  
280 ecoregion, resulting in the establishment of novel communities BC and ABC. This information is  
281 used to extend the range prediction of species B into the east ecoregion, resulting in an updated  
282 community composition forecast.

Novel  
community  
detection

Low

Medium

High

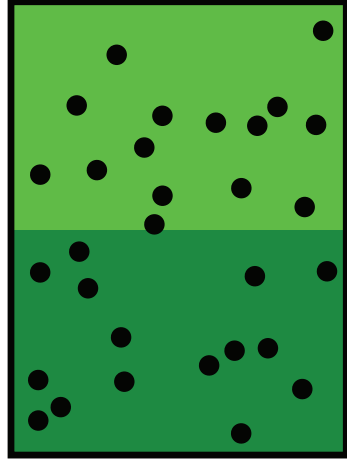
Cost  
(financial, labor)

Low

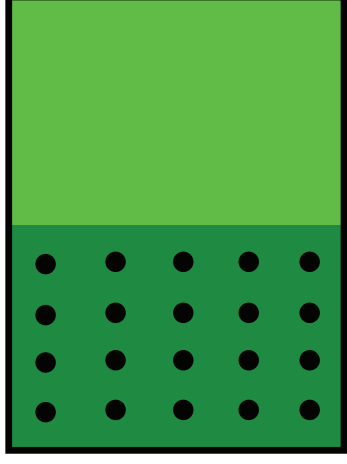
High

Medium

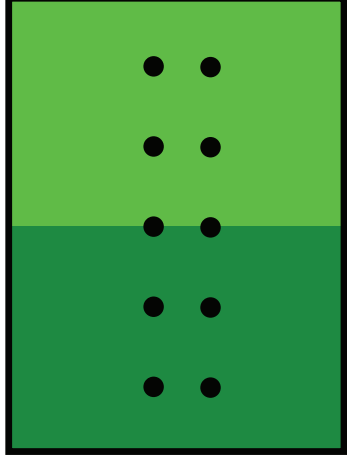
Strategy



Opportunistic  
occurrence data

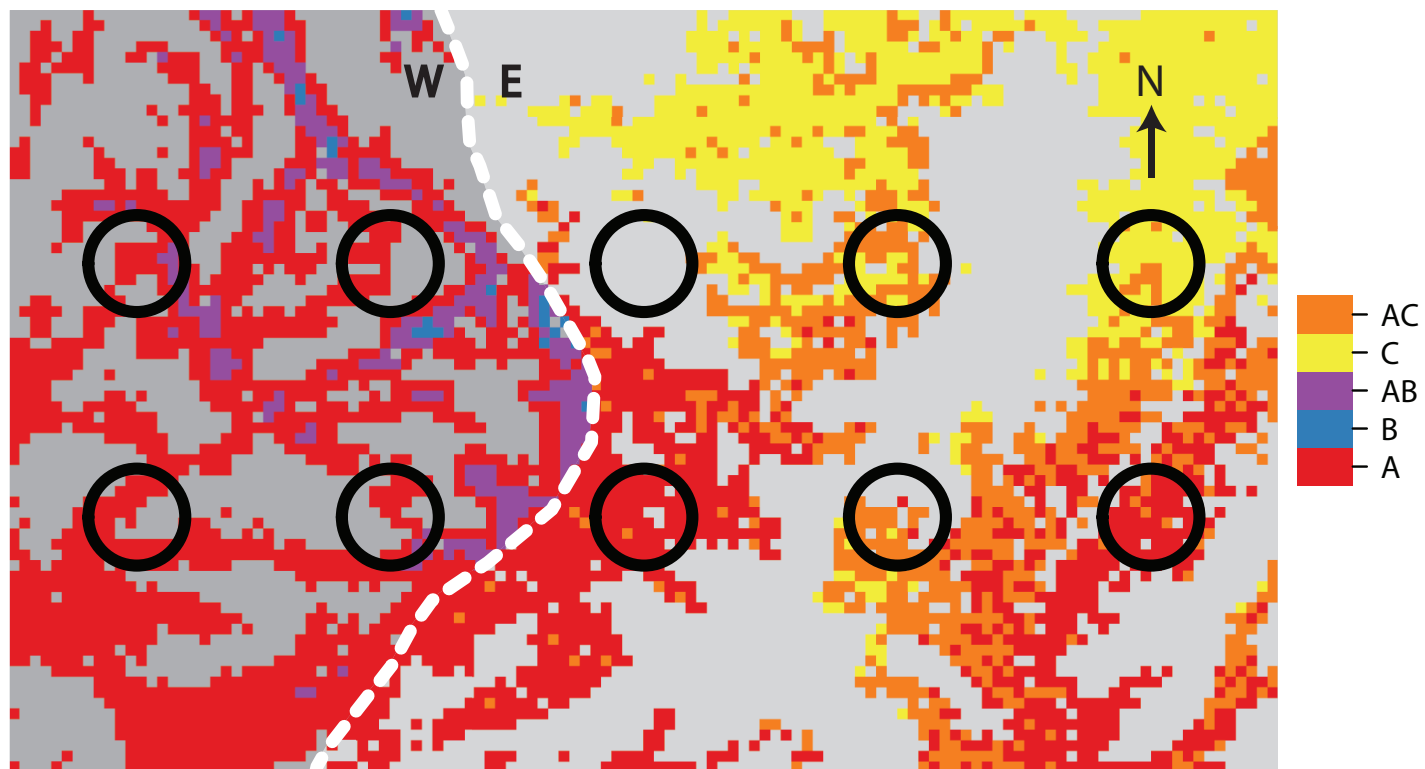


Passive sensor  
grid array  
(ecosystem-focused)



Passive sensor  
linear array  
(gradient-focused)

(a) Time 1



(b) Time 2

