

1 Title: Improving species conservation plans under IUCN's One Plan Approach using quantitative
2 genetic methods

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14 **Abstract**

15 Human activities are resulting in altered environmental conditions that are impacting the
16 demography and evolution of species globally. If we wish to prevent anthropogenic extinction
17 and extirpation, we need to improve our ability to restore wild populations. *Ex situ* populations
18 can be an important tool for species conservation. Quantitative genetic analysis can improve
19 management of these populations and thus the success of *in situ* population management actions
20 that they support. In this review we outline methods that could be used to improve the
21 management of *in situ* and *ex situ* populations in a One Plan Approach. We discuss how
22 quantitative genetic models can help measure genetic variation, phenotypic plasticity, and social

23 effects on phenotypes. Finally, we discuss how phenotypic change can be predicted using
24 measurements of additive genetic variance and selection. While previous work has highlighted
25 the value of *ex situ* populations for the field of quantitative genetics, we argue that quantitative
26 genetics can, in turn, offer opportunities to improve management and consequently conservation
27 of populations of species at risk. We show that quantitative genetic analyses are a tool that could
28 be incorporated into and improve *ex situ* management practices.

29 **Introduction**

30 Widespread human landscape transformations are resulting in changing conditions for species
31 across the globe (Parmesan 2006). Biodiversity is decreasing due to habitat loss, pollution,
32 disease, and climate change and most countries have not achieved biodiversity targets for 2020
33 set to slow rates of species declines (United Nations Environment Program Convention on
34 Biological Diversity, Aichi Target 12). This lack of progress calls for new approaches. In 2020,
35 the IUCN World Conservation Congress passed a resolution promoting the integration of *in situ*
36 (within a species' natural habitat) and *ex situ* (in human care outside a species' natural habitat)
37 conservation interventions by applying the One Plan Approach (OPA; WCC-2020-Res-079n;
38 Byers et al. 2013). Traditionally, species conservation planning has followed parallel but
39 separate tracks: field biologists and wildlife managers' efforts to address conservation needs *in*
40 *situ*, zoo, aquarium, and species-specific breeding centres (e.g. the United States Fish and
41 Wildlife Service Black-footed Ferret Conservation Center), efforts to develop sustainable *ex situ*
42 populations. Under the OPA developed by the IUCN's Conservation Planning Specialist Group
43 (CPSG), species conservation planning is conducted in an integrated manner by all responsible
44 parties, whether inside or outside of the natural habitat (Byers et al. 2013).

45 As recognized by the World Conservation Congress's 2020 Resolution 079, zoos and
46 aquariums can be an essential component of efforts to reduce the rate of species loss and to
47 improve the status of at risk species (Che-Castaldo, Grow, & Faust 2018;). However, *in situ*
48 recovery efforts that rely on source animals from *ex situ* conservation breeding programs can
49 face difficulties (Fischer & Lindenmayer 2000; Godefroid et al. 2011; Soorae 2021). The
50 reproductive fitness of individuals released to the wild can be reduced because of genetic drift,
51 inbreeding, and adaptation that might occur in captivity (Frankham 2008). Adaptation to captive
52 conditions could result in phenotypes that are maladaptive in the wild, resulting in lower survival
53 upon release, and adversely affect reintroduction efforts (Baskett, Burgess, & Waples 2013).
54 Additionally, gene flow via introduced individuals may alter evolutionary processes in the wild
55 resulting in negative effects on wild populations. We argue that some of these challenges can be
56 addressed— through the incorporation of quantitative genetic management techniques— to
57 improve *ex situ* population management, similar to that used to disentangle causes of phenotypic
58 change in wild populations (Pelletier et al. 2009; Chargé et al. 2014). Monitoring phenotypic and
59 genetic characteristics of *ex situ* populations would help to ensure their suitability for
60 conservation efforts, in particular under the OPA, in which captive and wild populations are
61 managed as a type of metapopulation (Byers et al. 2013).

62 Tracking the phenotypic dynamics of captive populations, and quantifying underlying
63 processes leading to change could be an effective management tool to ensure *ex situ* populations
64 will have a positive conservation impact (Princée 2016, Chapter 16). Many breeding programs
65 follow a mate pairing method based on matching mean kinship derived from pedigrees in an
66 effort to minimize genetic drift, inbreeding, and selection pressure while maintaining genetic
67 diversity (Montgomery et al. 1997; Ralls et al. 2000; Willoughby et al. 2014; Ballou et al. 2020).

68 However, the realities of captive management (e.g. the unequal reproductive success of mate
69 pairs and small effective population sizes) mean that evolutionary change can still occur
70 (Schulte-Hostedde & Mastromonaco 2015). Optimal breeding designs will not always be feasible
71 given a breeding program's resources and outcomes of any given captive management plan
72 could deviate from expectations because of unaccounted for influences. Deviation from an
73 optimal design either because it is not feasible or because of unaccounted factors could lead to
74 evolutionary change. For example, a study of Houbara Bustards (*Chlamydotis undulata*) revealed
75 evolutionary change in gamete production, courtship display rate, and body mass caused by
76 unintentional selection in captivity over just 5 generations (14 years) despite a breeding
77 management strategy based on mean kinship (Chargé et al. 2014).

78 Conservation breeding programs could be improved in many cases through analysis of
79 phenotypes. Herein, we undertake a review of quantitative genetics tools that we suggest can be
80 incorporated into *ex situ* population management, thereby improving the success of OPA
81 conservation efforts by quantifying, and ultimately preventing genetic adaptation to captivity
82 (Williams & Hoffman 2009). We describe methods that have been used in the study of
83 ecological and evolutionary dynamics in wild populations, expanding upon a previous review by
84 (Pelletier et al. 2009), including updated information on available tools, and suggesting how they
85 can be extended to *ex situ* populations (Fig. 1). First, we review why it is valuable for breeding
86 managers to monitor phenotypic dynamics (Section 1). Next, we describe how the plastic and
87 evolutionary dynamics of traits in captivity can be measured and we discuss how these
88 measurements can be used to improve the success of OPA conservation programs. We focus on
89 three major areas of consideration, including the measurement of evolutionary change (Section
90 2), phenotypic plasticity (Section 3), and parental and social effects (Section 4). Finally, we

91 describe the integration quantitative genetic information into current conservation breeding
92 practices to help inform *ex situ* and *in situ* conservation management and conclude with tools
93 that could be used to try to measure and predict adaptation (Section 5). We provide introductory
94 papers to allow managers to begin to monitor these processes in their breeding programs (Fig. 2).

95 **1. Phenotypic change in captivity**

96 Phenotypic differences in behaviour and morphology can occur between wild and captive-bred
97 individuals (O'Regan & Kitchener 2005; McDougall et al. 2006), which could decrease fitness
98 in the wild (Jolly & Phillips 2021). Differences between captive and wild phenotypes can be
99 caused by phenotypic plasticity, changes in demographic structure, evolutionary change, or all
100 processes. Phenotypic plasticity is the range of phenotypes an individual (or genotype) expresses
101 across a range of environmental conditions, while evolutionary change is a change in allele
102 frequencies underlying phenotypes caused by mutation, gene flow, genetic drift, and selection
103 (West-Eberhard 2003; Walsh & Lynch 2018).

104 The captive environment can potentially alter a broad range of traits. Morphological
105 changes have been commonly observed to change due to the captive environment (Courtney
106 Jones, Munn, & Byrne 2018; Fischer & Romero 2019). For example, differences in nutritional
107 environment and a change in physical activity can alter tissue development (Harbers et al. 2020).
108 Cues or social interactions that prompt development may also be altered in captivity (Monaghan
109 2008; Sultan 2015). Additionally, capture biases and the captive environment can inadvertently
110 select for specific behavioural temperaments resulting in differences between the temperament of
111 wild and captive individuals (McDougall et al. 2006; Monk et al. 2021).

112 Measurements of plastic trait responses and the genetic variation present in traits can both
113 provide information on the adaptive potential of the population and alert managers to potentially

114 unwanted evolutionary change (Section 2; Section 5). Even if captive and wild individuals
115 exhibit the same average phenotype, phenotypic plasticity could be masking evolutionary change
116 (e.g. Bonnet et al. 2017). For example, smaller individuals might be selected for in a captive
117 environment but better nutrition could result in increases in size that would mask this
118 evolutionary change. Only after being released into the wild where food resources are limited or
119 more difficult to acquire, would the evolutionary change towards a smaller size become
120 apparent.

121 Monitoring and quantifying evolutionary processes is of interest to *ex situ* population
122 managers because phenotypic change induced by captivity has been observed to reduce survival
123 and reproduction in the wild (Sundström et al. 2016; Cox and Lima 2006; Blumstein et al.
124 2002; Griffin et al. 2001). Further, change in captivity could alter the ecological role of the
125 organism or the societal value of organism. As the goal of *ex situ* populations is the restoration of
126 viable self-sustaining populations, we argue it is useful to understand environmental and genetic
127 contributions to phenotypes in captivity. Quantitative genetics provides a toolset for
128 disentangling the processes of evolutionary change and phenotypic plasticity. Quantitative
129 genetics is routinely used in breeding programs for domestic livestock (Walsh & Lynch 2018).
130 This methodology has also led to insight into the evolutionary dynamics in wild populations
131 (Charmantier, Garant, & Kruuk 2014) and it has been highlighted that zoo populations may
132 provide datasets, in the form of studbooks, well suited to quantitative genetic analysis (Pelletier
133 et al. 2009). Further, while molecular methods can help to track or identify loss of diversity in
134 genetic markers, changes in neutral genetic diversity do not always correspond well to changes
135 in adaptive genetic variation (Reed & Frankham 2001; Mittell, Nakagawa, & Hadfield 2015;
136 Lacy, Malo, & Alaks 2018). Thus, ideally, *ex situ* populations are managed through monitoring

137 of overall molecular genetic variation, quantitative genetic variation (the phenotypic variation
138 ascribed to molecular genetic variation), and the non-genetic causes of phenotypic variation.

139 **2. Evolutionary change**

140 **2.1 Trends in breeding values**

141 Quantitative genetic approaches use statistical tools to separate measured phenotypes into
142 genetic and environmental components, allowing the statistical quantification of potential
143 evolutionary change. Using a quantitative genetics approach, those managing *ex situ* populations
144 need information on pairwise additive relatedness (acquired through a pedigree, partial kinship
145 information, or molecular markers) and phenotypic data, combined in statistical models to
146 evaluate whether evolutionary change might be occurring in their captive population (Fig. 1).
147 Historically, quantitative genetic analysis was focused on laboratory and agricultural studies
148 where experimental breeding crosses were possible, but statistical techniques developed in the
149 1950s (Henderson 1950) and computational advances in the late 1990s allowed widespread use
150 of the “Animal Model.” The Animal Model is a form of mixed model that uses relatedness
151 among individuals to estimate the additive genetic variation of a trait (Wilson et al. 2010); it
152 models an individual’s phenotype as a function of the population mean phenotype plus an
153 additive genetic value and residual error. The additive genetic value, or the breeding value,
154 represents the additive genetic difference of an individual and the population average, or the sum
155 of the average effects of all the alleles the individual carries (Falconer & Mackay 1996; Lynch &
156 Walsh 1998). Changes in the average breeding value of a trait over time in a population can be
157 an indication of evolutionary change (Hadfield et al. 2010). Livestock producers are often
158 interested in changing the average breeding value of a population so that it is better for
159 production, for example in milk yield (Rendel & Robertson 1950), while evolutionary ecologists

160 are interested in determining how and whether evolutionary change is occurring in a wild
161 population (Walsh & Lynch 2018). In contrast, those maintaining *ex situ* populations for
162 conservation purposes will probably be interested in maintaining the average breeding value of a
163 trait in the captive population and the variance of the breeding values (the additive genetic
164 variance) in the interest of avoiding evolutionary change and maintaining adaptive potential
165 (Williams & Hoffman 2009). Minimizing mean kinship will reduce allele frequency change and
166 depending on the kinship matrix used managers can maximize the amount genetic variation or
167 maintain allele frequencies closer to the base population (Meuwissen et al. 2020; Morales-
168 González; Saura et al. 2008). However, monitoring and controlling breeding values for specific
169 traits could be combined with management plans to identify and control potential evolutionary
170 change. There is often uncertainty associated with each estimate of a breeding value, and
171 ignoring this error in the analysis of trends in breeding values can lead to an incorrect analysis
172 (Hadfield et al. 2010; Houslay & Wilson 2017; Princée 2016, Chapter 16) however, there are
173 techniques such as multivariate statistics or Bayesian analysis that can help with some of these
174 issues (Fig. 2).

175 When working with a captive population that is maintained across multiple facilities,
176 managers will also want to account for differences in phenotype between facilities and
177 understand how much of any observed variance different management practices among facilities.
178 Shared environmental effects such as year, rearing location, and parental effects should also be
179 accounted for in any estimation of the additive genetic variance because these values can inflate
180 similarity among relatives and bias estimates of the additive genetic variance. The same tools
181 that estimate additive genetic variance can also be used to account for such groupings in the data.
182 The use of mixed or hierarchical models in quantitative genetics is used to disentangle

183 components of variance beyond just components of genetic variance (Fig. 2). Given the proper
184 grouping (*e.g.* cohort year or rearing facility) is included in the data, we can estimate the
185 contribution of such a grouping to the total phenotypic variance. In some cases, the variance
186 associated with different people taking phenotypic measurements can be quantified and
187 accounted for in the measurement of heritability or repeatability of a trait (Ponzi et al. 2018).
188 Because of the relatively small size of captive populations, non-additive genetic variation and
189 increased inbreeding could also contribute to variation in traits (Wade & Goodnight 1998).
190 Quantitative genetics provides useful tools for measuring the impact of these genetic effects on
191 observed phenotypes and may help quantify evolutionary changes in captivity more accurately
192 (Pelletier et al. 2009; Wolak & Keller 2014). Our review is timely because recent genomic tools
193 will make quantitative genetic analyses possible in a broader range of species and populations
194 (Gienapp et al. 2017; *e.g.* Gervais et al. 2019). Genomic relatedness matrices can now be used in
195 lieu of a pedigree derived relatedness and implemented in an Animal Model approach to estimate
196 the additive genetic variances of traits in species where it previously was not possible. Further,
197 genomic tools can help to clarifying relationships among founding individuals in a population
198 and connect descendants of released individuals to lineages in the captive population.

199 Building an Animal Model to estimate evolutionary change using breeding values will
200 require a significant up-front time investment, but analysis can provide invaluable information
201 for management of quantitative genetic variation that cannot easily be estimated by other
202 methods. Further, once a suitable model has been developed it can be updated annually to
203 monitor any potential evolution occurring in traits of interest in the captive population over time.
204 Managers could then try to alleviate known or likely drivers of evolutionary change (see section
205 5). If changes in the average breeding values are determined to be of concern, managers could

206 increase gene flow from wild populations or to drive breeding values in a desired direction
207 through selective breeding. Increasing gene flow and selective breeding comes with difficulties
208 and depends on sampling individuals from the wild that have breeding values that can alter the
209 average captive breeding value in a desired direction. Knowledge of the wild population will
210 help inform strategies that use gene flow to alleviate evolutionary change in captivity (e.g.
211 sampling relatives from families with estimated breeding values in captivity). Selective breeding
212 should be done with caution because it could reduce genetic diversity and have unintended
213 consequences through selection on correlated traits (Ralls et al. 2000; Lande & Arnold 1983;
214 Arnold & Wade 1984a, 1984b).

215 Quantitative genetic analyses will be limited by the amount of data available for a
216 managed population. In some cases, an additive genetic variance estimate will be possible with
217 100 or fewer animals, but statistical power in these analyses also depend on the number of
218 relatives in a pedigree, the structure of the pedigree, and covariation of relatives with
219 confounding variables (e.g. maternal effects, rearing facility). Given a specific studbook
220 pedigree, a manager could conduct a simple power analysis to try to determine the heritability
221 they would be able to estimate with their specific pedigree structure (Hadfield et al. 2010;
222 Morrissey & Wilson 2010).

223 **2.2 Genetic Groups**

224 Founders in a population might come from populations with different genetic backgrounds that
225 might have traits with different average breeding values. Using genetic groups, Animal Model
226 methodology can account for known or assumed genetic structuring in a studied population
227 (Wolak & Reid 2017; Lacy 2012). Genetic groups are researcher defined groupings that are
228 ideally informed by knowledge of assumed or known genetic structuring in the wild (founders

229 from distant populations or molecular marker informed population structuring). One valuable
230 approach for joint *ex situ* and *in situ* management could be to assign founding individuals, and
231 progeny produced in the first few years of a conservation breeding program to one group, and
232 later immigrants brought into captivity as a second group. The proportion of each offspring's
233 genome attributed to the *ex situ* versus *in situ* population can then be determined using the
234 studbook pedigree. Beyond just accounting for biases, partitioning individuals among genetic
235 groups in this way allows explicit measurement of the effects of wild population gene flow on an
236 average trait value in the captive population (Wolak & Reid 2017). A difficult decision for
237 managers will be to determine the number of genetic groups to use for a given conservation
238 program. For example, after how much time should new individuals brought into captivity be
239 considered a new genetic group? Analysis of molecular markers could possibly help inform the
240 number of groups to use in a genetic group analysis. If enough data are available in the wild, trait
241 values could also be monitored and quantified for the *in situ* population, which would provide
242 comparisons to help determine the extent to which captive individuals differ from a baseline
243 (Fig. 1). Additionally, recent advances in analytical methods allow for the measurement of
244 different additive genetic variances between groupings and extend genetic group methods to
245 genomic relatedness, which may be useful for comparing the adaptive potential of a trait in the
246 wild or captive population (Muff et al. 2019; Aase et al. 2022). A study of song sparrows
247 (*Melospiza melodia*) on Mandarte Island, Canada provides an empirical example of a genetic
248 group model that mirrors an *ex situ* breeding program (i.e. a focal study population with
249 measured and periodic gene flow). In this case, the analysis used a genetic group model to
250 determine that gene flow to the island population is preventing local adaptation (Reid et al.
251 2020).

252 3. Plasticity and changes in plasticity

253 Phenotypic plasticity is the range of phenotypes that a single genotype, and in some cases
254 individual, can express across a range of environmental conditions (Sultan 2015; West-Eberhard
255 2003). Individuals can differ in their plastic responses to the same environmental gradient (Box
256 1; Fig. 3). Like variation in a phenotype, the variation in an individual's plastic response to
257 environmental conditions can be decomposed into environmental and genetic contributions
258 (Gienapp & Brommer 2014). If individuals differ in their plastic responses because of genetic
259 differences, plastic responses themselves could evolve. Therefore, captivity might influence
260 plastic responses through evolutionary change or environmental/developmental effects that alter
261 an individual's plastic response. Most importantly, an altered plastic response might affect the
262 fitness of an individual or family in captivity or the wild, which is why managers must be
263 concerned with the response.

264 To measure individual (combined environmental and additive genetic response;
265 individual by environment reaction norms; IxE; Fig. 3 B, C) plastic responses to captivity,
266 repeated measures on previously-captive individuals in wild environment are required (Nussey,
267 Wilson, & Brommer 2007; Box 1). This approach highlights the benefits of and need for an OPA
268 management strategy when *ex situ* populations are incorporated into species conservation. For
269 non-clonal species, we can most easily measure the individual level plastic responses (IxE; Fig. 3
270 B, C) of labile traits that are expressed multiple times in an individual's life (annual fecundity,
271 timing of breeding, migratory urge). The genotypic component (genotype by environment
272 interaction; GxE; Fig. 3D, E) of a response to captivity might be more easily measured and
273 relevant to managers. Measuring GxE interactions will require the measurement of phenotypes

274 from groups of relatives in the wild and captivity. GxE interactions could inform managers how
275 a group of related individuals might perform in the wild and captivity (Fig. 3E).

276 Understanding how captivity shapes plastic responses to environmental conditions
277 individuals will encounter *in situ* may be one of the most important considerations in a
278 reintroduction program. The captive environment differs in many ways from the wild
279 environment, and both genetic and environmental differences between individuals may cause
280 them to respond differently. Captivity could affect the plasticity of traits and the ability of
281 individuals to plastically respond to environmental variation. Some traits might revert to wild
282 values post-release, while others may not (Fig. 4). For example, plastic responses may be
283 adaptive in natural environmental conditions, and plasticity is now increasingly recognized as a
284 primary response to changing climatic conditions (Bonamour et al. 2019) . Early-life stages are
285 particularly sensitive to environmental conditions (English et al. 2016; West-Eberhard 2003).
286 Consequently, development during early-life in a captive environment could affect the way an
287 individual responds to environmental variation once released (Munch et al. 2018), and thus its
288 fitness. Finally, anti-predator behaviours will be valuable to monitor as they are sometimes, but
289 not always, observed to disappear over time in captivity (Cox & Lima 2006; Blumstein et al.
290 2002) and anti-predator behavioural training may help improve survival upon release (Reading et
291 al. 2013; Griffin et al. 2001; but see Moseby et al. 2012)

292 The consequences of changes to plasticity depend on whether the ability to plastically
293 respond to environmental conditions affects fitness for a given species in the wild. For example,
294 if there is a positive association between how quickly an individual responds to environmental
295 variation (the slope of the plastic response) and fitness (Fig. 5A), reduced plastic responses
296 caused by captivity could negatively impact the success of reintroduction or supplementation

297 efforts. That said, if there is no relationship observed between fitness and the plastic response
298 (Fig. 5D) it may not be as important to monitor or put effort into determining how to prevent the
299 loss of this response during captive management. While likely challenging to measure, it may be
300 worthwhile to investigate if and how (and how commonly) captivity alters plastic responses in
301 wild conditions and how to create environmental conditions in captivity that can maintain
302 appropriate plastic responses in the wild.

303 A sampling design challenge will be to measure plastic responses of 1) wild individuals
304 to captivity, 2) wild individuals to natural environmental variation, and 3) previously captive
305 individuals to natural environmental variation (Fig. 4). Often hundreds of individuals are
306 required for statistical power and each of these individuals needs to be repeatedly measured
307 across environmental contexts (Dingemanse & Dochtermann 2013). Power analysis could be
308 used to design data collection protocols that will ensure results can help improve a management
309 programs ability to detect plasticity or whether an existing data set is adequate to statistically
310 detect plasticity (Allegue et al. 2017).

Box 1 Measuring plasticity

Quantifying plasticity allows us to try to measure the contribution of non-genetic responses to environmental change to overall population level phenotypic change. Individual responses can be measured as a straight line connecting an individual's average phenotype in the captive and wild environment (Fig. 3). When the environmental variable in such an analysis is mean-centred the intercept of such a line indicates the average trait value of an individual and the slope connecting the environment-specific trait values indicates the individual's response to captivity (Fig. 3). Individual plastic responses are usually measured in multilevel/hierarchical/mixed models (Martin et al. 2011). Within the studied population,

individuals could have the same response (Fig. 3B) or might vary in their response to captivity (Fig. 3C). Differences among individuals could be caused by genetic or permanent environmental differences (environmental effects that have a persistent effect on an individual's phenotype; see (Kruuk 2004; Wilson et al. 2010). Like individual responses, family groups might have similar (Fig. 3D) or different responses (Fig. 3E) to captivity (Gienapp & Brommer 2014).

311

312 **4. Parental and indirect genetic effects**

313 Both parental effects and social interactions (*i.e.* indirect genetic effects on an individual caused
314 by the expression of genes in another individual, either a parent or conspecific) can have effects
315 on the phenotype of an individual. Indirect effects can be heritable and could impact the adaptive
316 potential of a trait (*e.g.* Moiron et al. 2020). Because captivity could alter both parental effects
317 and social interactions, the impacts of indirect genetic effect could vary drastically between wild
318 and captive populations. Monitoring wild and captive social networks can allow measurement of
319 the variance in a trait explained by interactions among individuals (Thomson et al. 2018).
320 Detecting differences among social networks of captive and wild populations is important
321 because of 1) the direct impacts a change in network might have on fitness or fitness related
322 traits; and 2) the potential effects of an altered network on the rate of evolutionary change in
323 captive versus wild environments.

324 In many species parents provide cues or care for offspring that can be altered by changes
325 in environmental conditions which are likely to result from captivity (Munch et al. 2018).
326 Because of the potential long-term impacts of an altered developmental environment, especially
327 for captive-reared animals, it may be particularly important to study how the captive

328 developmental environment affects offspring phenotypes (English et al. 2016). For example, in
329 common marmosets (*Callithrix jacchus*) early life exposure to higher fat diets increases the
330 probability of post-weaning obesity, and the milk from captive marmosets tends to have higher
331 fat content than wild marmosets (Power et al. 2008; Tardif et al. 2013). Further, mother
332 marmosets in captivity varied in their milk composition, suggesting that genetic and/or
333 environmental differences exist among mothers that have health consequences for their offspring
334 (Power, Oftedal, & Tardif 2002).

335 Beyond parental effects, social interactions among individuals can affect the phenotypes
336 expressed in a population (Fisher, Haines, et al. 2019; Fisher, Wilson, et al. 2019; Laskowski,
337 Wolf, & Bierbach 2016). For example, mates and neighbours can affect an individual's breeding
338 time (Fisher & McAdam 2019). The impact of this social interaction has been observed in
339 common terns (*Sterna hirundo*), where the breeding time of females is affected by their mate,
340 and in North American red squirrels (*Tamiasciurus hudsonicus*), where breeding time can be
341 influenced by neighbouring squirrels (Moiron et al. 2020; Fisher, Wilson, et al. 2019). Further,
342 impacts of indirect genetic effects likely depend on the number of conspecifics an individual
343 interacts with (Fisher & McAdam 2019), which has the potential to be altered by captivity.

344 **5. Putting it all together: combining quantitative genetic analyses with conservation**

345 **management tools**

346 Application of quantitative genetics to *ex situ* and *in situ* conservation programs will be limited
347 by the quality and amount of data available. Here we provide additional guidance for managers
348 interested in collecting the data required to conduct quantitative genetic analyses, including
349 available software, and standardized data collection. It may be most worthwhile for managers to

350 begin with a trait that has changed over generations in captivity or is known (or hypothesized) to
351 hamper breeding or reintroduction success (Fig. 6).

352 Studbooks for conservation breeding are routinely maintained in a variety of platforms,
353 from Excel and Access databases to dedicated software such as Poplink (Faust et al. 2019).
354 Currently approximately 1400 conservation studbooks are maintained in the web-based portal
355 ZIMS for Studbooks (Species360 Zoological Information Management System. Retrieved
356 from <http://zims.Species360.org>). Platforms provide varying options for data storage,
357 manipulation, and export.

358 Regardless of format, studbooks typically include basic data that is needed for
359 quantitative genetic analysis, in the form of pedigrees and life history events. Studbook pedigrees
360 can be simple pedigrees noting discrete parentage but also allow for the incorporation of
361 parentage “assumptions” that can be used to assign animals to groups in cases where pedigree is
362 unknown or to create cohorts for the study of gene flow. Additionally, the commonly used
363 studbook applications include an option to incorporate User Defined Data Fields (UDFs). UDFs
364 can be used to record phenotypic data or quantitative genetics output such as breeding value.
365 UDFs are flexible and can be updated which will be invaluable for estimated breeding values
366 that will change and need to be updated every time a new analysis is conducted. Studbooks are
367 databases commonly exported into analytical softwares (e.g. PMx, Ballou et al. 2020) that are
368 used to determine mate-pairings through a mean kinship list. PMx can also be used to compile
369 life history events, generate demographic life tables, determine fecundity rates, estimate breeding
370 seasonality, and other metrics of interest in the study of phenotypic change. As with studbook
371 softwares, PMx allows for the importation of UDF fields that can be added to mean kinship lists,
372 such as breeding value, which can then be considered in constructing pairing decisions.

373 Therefore, the estimates of gene flow, social management, and breeding strategies which
374 incorporate quantitative genetics analyses can be modeled and considered alongside gene
375 diversity (probability-based estimate of heterozygosity) retention and inbreeding coefficients to
376 improve management.

377 The challenge remains, however, of how quantitative genetics can be incorporated into
378 management paradigms for *ex situ* populations. Studbooks and associated analytical software
379 including PMx and Vortex allow managers to explore how manipulating social groupings,
380 housing conditions, husbandry methods, setting informed schedules of geneflow, and adjusting
381 pair selection might impact current management (Lacy & Pollak. 2021).

382 Accurate studbook records and standardization of trait measurements are crucial for the
383 preservation of a long-managed species; incorrect registration, administration errors, and limited
384 founder information will compromise pedigree authenticity. Lineages and pedigree data must be
385 accurate for effective application of quantitative genetic analyses; although some genetic
386 variances can still be estimated without bias if errors in paternity assignment are random
387 (Charmantier & Réale 2005; Firth et al. 2015). Pairing recommendations, either using
388 quantitative genetics or traditional pedigree-based inbreeding coefficients, will always be
389 presented with logistical and statistical limitations. Despite these limitations, the use of
390 quantitative genetics in study systems with adequate data and with proper acknowledgement of
391 uncertainty present the potential to improve management of *ex situ* and *in situ* recovery
392 programs.

393 In our view, the key promise that quantitative genetics provides to conservation breeding
394 programs is the ability to disentangle the processes that lead to phenotypic change in captivity.
395 Quantifying the relative contribution processes to phenotypic changes will enable adaptive

396 management and a prioritization of resources to the processes that most contribute to changes in
397 captivity. Quantitative genetic techniques provide a set of tools that allow us to try to determine
398 if more (or less) effort is needed to prevent causes of phenotypic change in captivity (plasticity,
399 evolution, social environment), in addition to current best practices mate-pairing based on mean
400 kinship. We emphasize that the OPA recommended by the IUCN is cohesive with quantitative
401 genetic tools because the effectiveness of quantitative genetic tools will improve with increasing
402 data gathered jointly from *in situ* and *ex situ* populations.

403 If restoring previous ecological conditions for a species at risk is impossible,
404 conservation must necessarily focus on maintaining or improving the adaptive potential of
405 populations (Chevin & Lande 2010). As the goal of *ex situ* populations is, ultimately, the
406 conservation of the species in the wild, their management must ensure that supported populations
407 can adapt to changing conditions in the wild. Predicting such adaptation will depend on
408 understanding how selection operates and is changing in the wild, how much additive genetic
409 variance is present for selected traits, and the suite of plastic responses available to a population
410 (Sultan 2015; Gienapp & Brommer 2014).

411 Determining whether and how any evolutionary or plastic responses result in
412 demographic changes remains a challenge for population biologists (Hendry 2016; Janeiro et al.
413 2017). However, some models have been developed that try to predict when plasticity or
414 evolution might prevent the extinction of a population (Vedder, Bouwhuis, & Sheldon 2013;
415 Chevin & Lande 2010). A particularly important parameter is the additive genetic variance of
416 fitness. The additive genetic variance of fitness should be equivalent, in theory, to the rate of
417 adaptive genetic evolution (Bonnet, Morrissey, & Kruuk 2019; Fisher 1930; de Villemereuil et
418 al. 2016). Thus, comparison of the additive genetic variance of fitness might indicate how

419 quickly adaptive genetic evolution is occurring in wild versus captive populations. The goal of *ex*
420 *situ* populations is ultimately to directly support conservation efforts for wild populations, for
421 example through population augmentation. As such, *ex situ* and *in situ* partners should work
422 together to quantify the wild population as changes due to captivity will directly impact program
423 success, which is the intent of the OPA.

424 **Conclusions**

425 Integrated planning and management of wild and captive populations in an OPA can
426 improve the impact of conservation efforts for species at risk (Lees et al. 2021). Here, we present
427 and provide support for the argument that quantitative genetic analysis is a powerful tool that can
428 be used to enhance *ex situ* population management, and help to integrate *ex situ* and *in situ*
429 activities. Several examples exist demonstrating how phenotypes have come to differ between
430 captive and wild populations, despite best management practices for *ex situ* populations that
431 include efforts to reduce the loss of diversity. The consequences of these differences are not
432 always known, but, may impact the fitness of individuals that are used to directly support *in situ*
433 conservation efforts. Using existing pedigrees and phenotypic data in the Animal Model
434 approach, managers can disentangle the causes of these differences and understand their
435 consequences. By extending the approach to include genetic groups, analyses can both quantify
436 the effects of gene flow on phenotypes. Finally, these models can help managers to measure rates
437 of adaptation in captivity or predict whether captive populations are maintaining the adaptive
438 potential necessary to persist under changing conditions in the wild. Since the data to run
439 quantitative genetics analyses often already exists (i.e. in studbooks), we see quantitative genetic
440 analysis as a promising tool for conservation breeding that can likely be integrated with existing
441 management methods used for maintaining genetic diversity. In doing so, *ex situ* populations will

442 ensure they are as effective as possible in supporting *in situ* conservation efforts and managers
443 can better identify where to direct limited resources to answer questions critical to improving the
444 management of a species.

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745 **Figure Captions**

746 **Figure 1:** Key questions that may arise in a conservation breeding program and the data and
747 models that can be used in a quantitative genetic and One Plan Approach framework to answer

748 them. For each question references are provided that either provide code to run similar analyses
749 or provide guides for the suggested model.

750 **Figure 2:** Introductory papers and resources for conservation managers looking to make use of
751 quantitative genetic analyses for breeding programs.

752 **Figure 3:** Variation in plastic responses to captivity. If there is a plastic response at the
753 population level (A) individuals might all have the same plastic response (B) or they could differ
754 in their responses to captivity (C). If individuals differ in their responses, these differences could
755 be caused completely by environmental differences and we would not see differences among
756 family groups (D) or genetic differences might also have differences among families
757 contributing to observed differences among individuals (E). We illustrate differences in
758 responses as if they were completely caused by environmental (D) or genetic differences (E), but
759 they can be caused by a combination of both environmental and genetic differences.

760 **Figure 4:** Three individual (or average family) responses to captivity. Responses to captivity
761 between individuals might differ because of genetic or environmental effects. Individuals might
762 not change a trait value to captivity at all (blue solid line), they may respond to captivity but then
763 return to wild trait values when released (purple dashed line), or individuals might maintain the
764 same captive phenotype despite returning to the wild environment (red dotted line).

765 **Figure 5:** Potential effects of captivity on the plastic response of a trait in the wild. Because of
766 evolutionary or environmental effects in captivity the plastic response to environmental
767 conditions post-release might be reduced or eliminated (A), or plastic responses post-release
768 might remain similar to those in the wild (B). The consequences of changes in plasticity will
769 depend on the relationship between plasticity and fitness in the wild. If plasticity is adaptive it

770 might play an important role for population persistence (C) or plasticity might not be important
771 under wild environmental conditions (D).

772 **Figure 6:** A decision tree for determining the steps in an analysis aimed at disentangling the
773 various causes in captivity that could contribute to changes in a trait.

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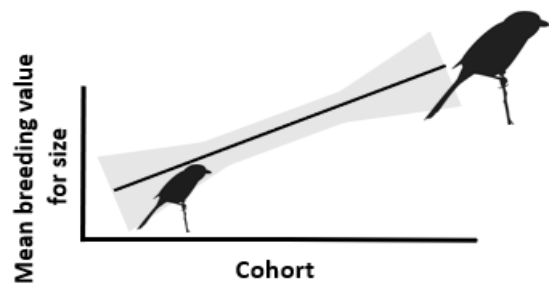
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785 **Figures**

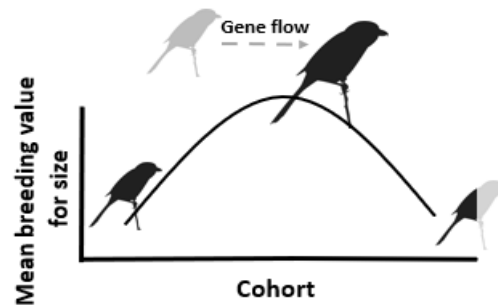
Are there signs of evolution in captivity or the wild?

Data: Phenotypes, pedigree
Model: Animal Model
References: Postma 2006; Hadfield et al. 2010; Wilson et al. 2010; Bonnet et al. 2019



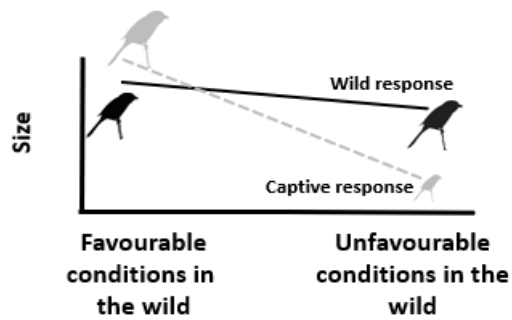
Does gene flow affect wild or captive phenotypes?

Data: Phenotypes, pedigree
Model: Animal Model
References: Wolak and Reid (2017); Muff et al. (2019); Reid et al. (2020)



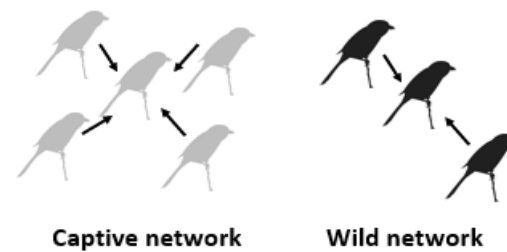
Does captivity alter plastic responses?

Data: Repeated phenotypic measures, pedigree (optional)
Model: Random regression models
References: Nussey, Wilson, and Brommer (2007); Gienapp and Brommer (2014); Houslay and Wilson (2017)



Do altered interaction networks impact individuals in captivity?

Data: Phenotypes, interacting individuals, pedigree (optional)
Model: Mixed model with neighbour or parental groups
References: Thomson et al. (2018); Fisher et al. (2019); Moiron et al. (2020)



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787 **Fig. 1**

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Getting started with Animal Models

Many of the papers on how to use an Animal Model have been written for ecologists studying populations in the wild. (i.e. *in situ* populations). Many of the concepts and tools will be similar for an analysis of a conservation breeding (*ex situ*) population.

- Wilson et al. (2010) is a good starting point reference that includes software that can be used for quantitative genetic analyses
- Pelletier et al. (2009) provide a perspective on using zoo populations to answer questions in quantitative genetics.
- Kruuk (2004) provides an in-depth overview of Animal Models

Books

- Charmantier, Garant, & Kruuk (2014) provide an overview of quantitative genetic techniques for wild populations.
- Falconer & Mackay (1996) provide an in-depth overview of the theory and analysis.
- Lynch & Walsh (1998) and Walsh & Lynch (2018) write detailed backgrounds on many quantitative and population genetic topics and are helpful references for more details on many concepts.
- Chapter 16 of Princée (2016) is an overview of quantitative genetic concepts and helpful guide for using studbooks to do quantitative genetic work.

Breeding Values

- Postma (2006) and Hadfield (2010) discuss statistical concerns in the analysis of breeding values
- Work on big horned sheep (Pigeon et al. 2016), snow voles (Bonnet et al. 2017), and red deer (Bonnet et al. 2019) include example R code for examining trends in breeding values

Measuring Plasticity

There is broad literature on measuring plastic responses in the context of behaviour, climate change, and experimental studies that will be useful to managers trying to use existing data to measure plastic responses to captivity (Gienapp & Brommer 2014).

- Houslay & Wilson (2017) provide useful tutorials for quantifying individual level plasticity and measuring selection on plasticity using the R package MCMCgllmm.
- From a perspective of behavioural traits detailed guidelines on the sampling schemes needed to measure different components of plasticity are provided by Dingemanse & Doctermann (2013) and Allegue et al. (2017) provide education software and a guide to the statistical quantification of individual differences.

Genetic Groups, Inbreeding, and Dominance Genetic Variance

- Wolak & Reid (2017) provide an explanation of genetic groups and code for running a basic analysis and Muff et al. (2019) extends their analysis to allow for different additive genetic variances among groups.
- Tools, software, and tutorials for examining inbreeding effects and estimating dominance genetic variance are provided by Wolak & Keller (2014) and Wolak (2012).

Measuring Selection and Adaptive Potential

Using associations between fitness and our trait of interest we can try to determine if our studied trait is under selection captivity.

- Lande and Arnold (1983) provide the classic and widely used multiple regression method for measuring selection
- Estimates of selection combined with information on heritability of traits can be used to predict responses to selection and compared to the observed trends to evaluate a hypothesis of selection in captivity causing evolutionary change (Queller 2017; Price 1970; Walsh & Lynch 2018)
- Bonnet et al. (2019) measure selection to compare observed and predicted evolutionary changes

Social networks

Differences in social networks over time and between wild and captive environments might contribute to phenotypic changes observed.

- Thomson et al. (2018) provide a tutorial on using Animal Models with multiple matrices to estimate non-genetic contributions to phenotypic variance (including social interactions)

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791 **Fig. 2**

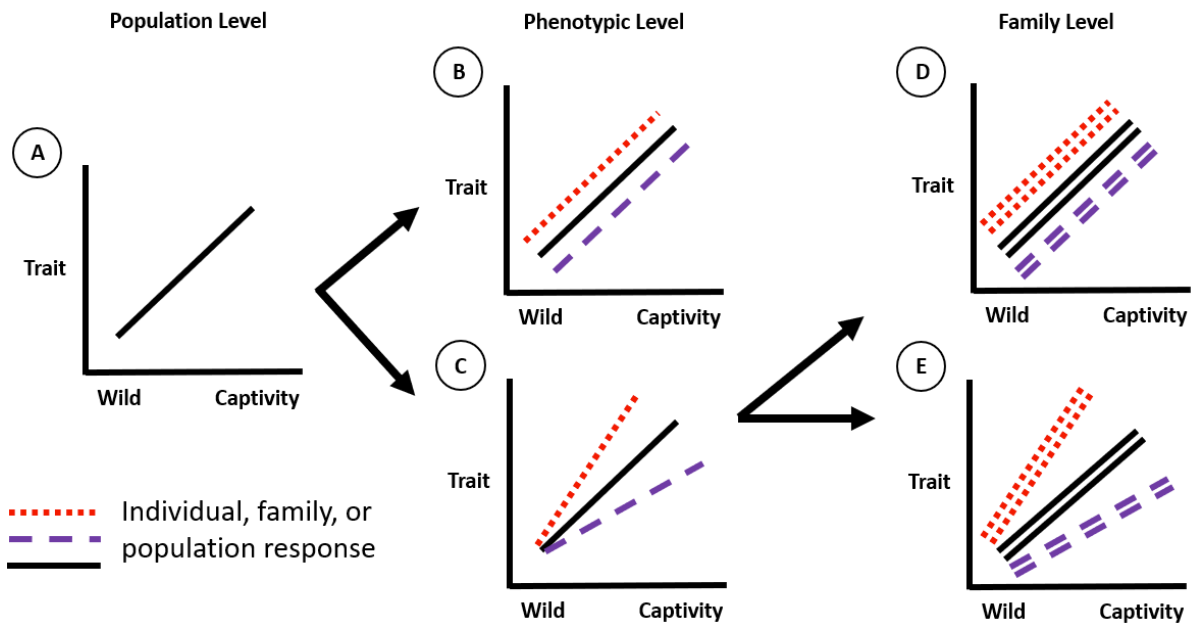
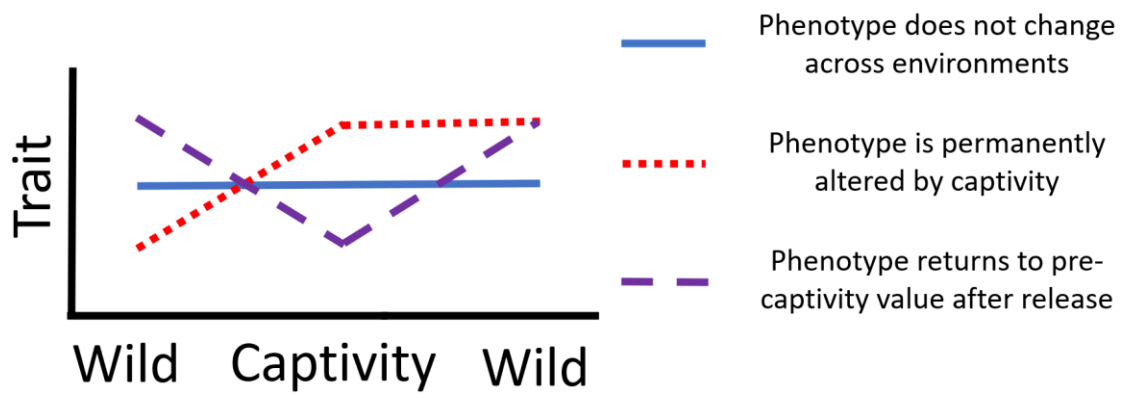


Fig. 3



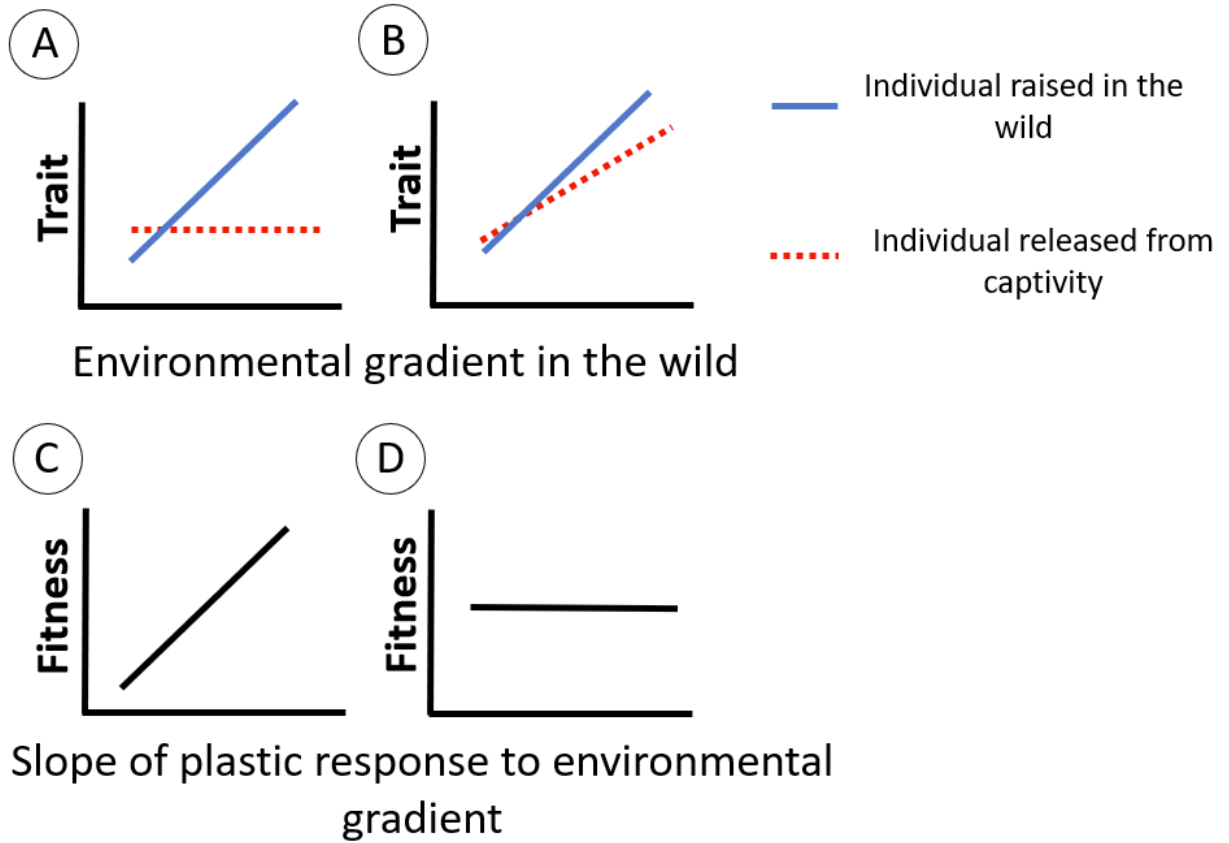
799 **Fig. 4**

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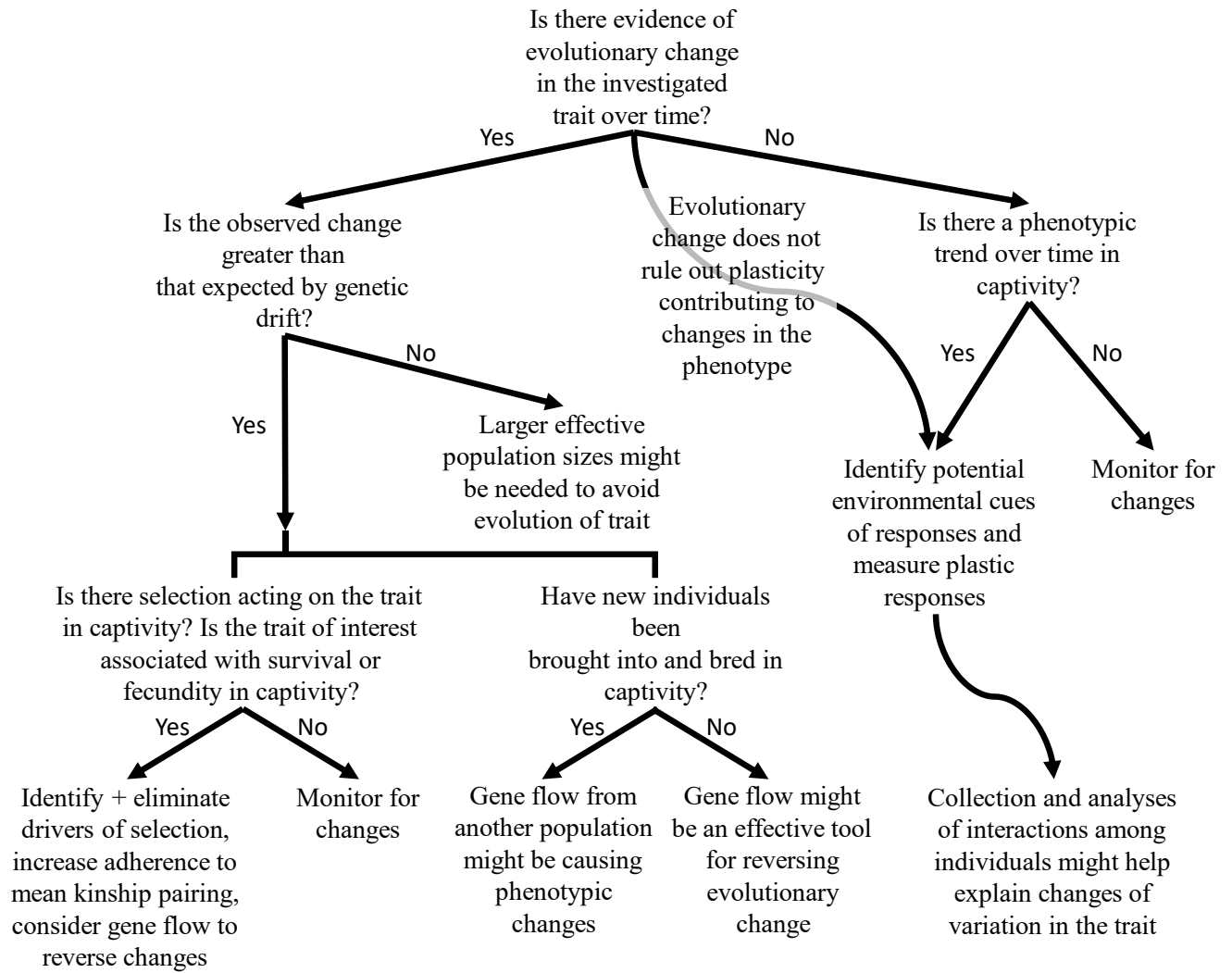
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805 Fig. 5



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807 **Fig. 6**

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