

1 **Effect of forest understorey stand density on woodland caribou (*Rangifer tarandus caribou*)**

2 **habitat selection**

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10 **Abstract:** Woodland caribou (*Rangifer tarandus caribou* Gmelin,1788) use older forests that  
11 provide abundant terrestrial lichen forage and refuge from predators. However, forest structural  
12 characteristics vary widely, differing in forage availability but also, perhaps, in the ability of  
13 caribou to move freely to access forage or to escape predation. We conducted a multivariate  
14 analysis of habitat in two geographically and biophysically distinct regions to identify the  
15 independent effects of various attributes, including forest understorey stand density, defined as  
16 standing and downed biomass, on caribou habitat selection. We developed Bayesian network  
17 models to predict the probability of habitat selection based on a set of remotely sensed habitat  
18 inputs. Caribou in the Bistcho range (northwestern Alberta) selected non-forest/sparsely forested  
19 areas while caribou in the Trout Lake region (northwestern Ontario) selected primarily forested  
20 habitats, nevertheless consistent with selection for reduced predation risk in both cases. Caribou  
21 also selected forest stands with lower understorey stand density in both regions, consistent with  
22 selection for stands that would allow greater ease of movement. The high-resolution satellite data  
23 resolved habitat characteristics more consistently and in greater detail than standard forest cover  
24 datasets that are most often used for these analyses, and led us to conclude that habitat  
25 management may require different treatments in different parts of the species' range to address  
26 what are nevertheless common pathways to decline.

27 *Key words:* woodland caribou, boreal forest, Canada, Bayesian networks, remote sensing

## 28 **Introduction**

29 Current policy and management for the boreal population of woodland caribou (*Rangifer*  
30 *tarandus caribou* Gmelin, 1788) in Canada is informed by the species' use of largely  
31 undisturbed, old stands of conifer forest (Environment and Climate Change Canada 2020).  
32 Specifically, black spruce (*Picea mariana* (Mill.) BSP), jack pine (*Pinus banksiana* Lamb.), and  
33 tamarack (*Larix laricina* (Du Roi) K. Koch)-leading forests and adjacent treed peatlands,  
34 muskegs, and bogs are cited as important habitats to restore and maintain to ensure the species'  
35 recovery. These forests are associated with abundant terrestrial lichens, on which caribou largely  
36 subsist during winter (Webber et al. 2022). Diets are broader in the snow-free season and forage  
37 quality is better in more productive forests (Denryter et al. 2022) where wolves (*Canis lupus*  
38 Linnaeus, 1758) and their primary prey, primarily moose (*Alces alces* Linnaeus, 1758), mule  
39 deer (*Odocoileus hemionus* Rafinesque, 1817), and white-tailed deer (*Odocoileus virginianus*  
40 Zimmerman, 1780) are more abundant (Latham et al. 2011; DeMars and Boutin 2018; Serrouya  
41 et al. 2021). Caribou generally forego opportunities to forage in these productive forests because  
42 of the elevated risk of predation (Briand et al. 2009; Thompson et al. 2015).  
43 Ecological characteristics of boreal forests vary widely (Pojar 1996) but forest conditions  
44 considered in studies of caribou habitat selection generally include only stand age and/or stand  
45 type, often because these are the only consistent data layers available at spatial scales typical of  
46 such studies. Other, more finely resolved forest characteristics, however, may play functional  
47 roles in the behavioural decisions that shape habitat selection by caribou.  
48 Here, we characterize the landscapes selected by caribou in two, geographically and  
49 biophysically distinct regions in central and western Canada, using several remotely sensed  
50 structural variables. We include for the first time a measure of understorey forest stand

51 conditions that is assumed to affect the mobility of caribou and therefore influence energetic  
52 trade-offs in the context of predation risk (Fryxell et al. 2020; Keim et al. 2021).

## 53 **Materials and Methods**

### 54 **Study areas**

55 Our study was conducted in the Bistcho boreal caribou range of northwestern Alberta  
56 (Environment and Climate Change Canada, 2020) and in the Trout Lake region of northwestern  
57 Ontario, Canada (Figure 1). The Bistcho range covers 14,366 km<sup>2</sup> and is contiguous with the  
58 Yates range to the east, the Calendar range in northeastern British Columbia, and the Cameron  
59 Hills region of southern Northwest Territories. Caribou move extensively among these ranges  
60 (Wilson et al. 2022). The range is located within the Northern Alberta Uplands and Hay River  
61 Plain ecoregions (Strong and Leggat 1992) and is composed primarily of lowland black spruce  
62 bogs and fens, as well as upland conifer, trembling aspen (*Populus tremuloides* Michx.) and  
63 mixedwood forests. Elevations vary between approximately 350 and 735 m above sea level.  
64 The Trout Lake region covers 16,476 km<sup>2</sup> and overlaps the Berens, Churchill, Kinloch, and  
65 Sydney caribou ranges, which comprise a generally continuous distribution of caribou in  
66 northwestern Ontario (Ministry of Natural Resources and Forestry 2014). The region is located  
67 within the Lake St. Joseph and Lake Nipigon ecoregions (Crins et al. 2009) on the Canadian  
68 Shield, which is characterized by exposed bedrock with shallow and coarse soils in the uplands,  
69 and a high density of small-medium sized lakes and wetland complexes in lowland areas. Black  
70 spruce and jack pine are the leading forest species. Elevations in the Trout Lake region vary  
71 between approximately 350 and 450 m.

72 Forest fires are a source of frequent natural disturbances in both regions and climates are  
73 similarly continental, with cold and relatively dry winters and short, warm summers. Mean

74 January and July temperatures in the Bistcho region (-20.4° C and 16.5° C) are similar to those  
75 of Trout Lake (-18.3° C and 18.1° C), but Bistcho receives about half the mean annual  
76 precipitation of Trout Lake (372 mm versus 686 mm; Environment Canada climate normals  
77 1981-2010, High Level, AB versus Red Lake, ON;  
78 [https://climat.meteo.gc.ca/climate\\_normals/index\\_e.html](https://climat.meteo.gc.ca/climate_normals/index_e.html)).

### 79 **Habitat variables**

80 We used SkyForest™ mapping products (First Resource Management Group Inc. [FRMG],  
81 North Bay, ON, Canada) to provide consistent, seamless, and detailed habitat mapping of forest  
82 stand conditions throughout both the Bistcho range and Trout Lake region. SkyForest™ uses  
83 open source and commercial satellite data, including optical data and synthetic aperture radar  
84 (SAR) data to produce raster products at 5-20 m resolutions. FRMG has been continuously  
85 developing these products since 2013, iteratively testing and applying proprietary indices of  
86 earth observation data against field data.

87 We defined *understorey stand density* as standing and downed biomass that could impede the  
88 movement of animals. We modelled this from backscatter data from synthetic aperture radar  
89 satellites at different bandwidths and reported for each pixel the number of modelled steps  
90 required for a field crew member to traverse plot transects (Figure S1, Figure S2, Figure S3,  
91 Table S1, Table S2, Figure S4).

92 We estimated *canopy height* by the difference in elevation between a digital surface model  
93 derived from WorldDEM™ elevation data (Airbus Defence and Space SAS, Ottobrunn,  
94 Germany) and the elevations from FRMG's patented Digital Terrain Model (DTM; US patent  
95 10,095,995 B2. Canadian patent 2,930,989 and patent pending). We derived the DTM from a  
96 data fusion of multiple SAR, optical and lidar satellites.

97 *Conifer basal area* is the percentage of total basal area of each 10-m grid cell that is composed of  
98 conifer species. This definition differs from the standard measure of basal area, which is a  
99 volumetric measure (m<sup>2</sup>) and does not indicate the relative composition of conifers versus  
100 hardwoods. *Crown coverage* is the percentage of the ground covered by a vertical projection of  
101 the forest canopy (Figure S5). We generated these both from a proprietary processing of  
102 Sentinel-2 optical satellite data and calibrated using data collected at field plots.  
103 *Terrain elevation* was estimated from the FRMG DTM as described above.

#### 104 **Habitat disturbance**

105 We defined habitat disturbances as anthropogenic features visible on 30-m Landsat imagery  
106 buffered by 500 m, as well as areas burned by wildfire within the past 40 years, current to 2015  
107 (<https://open.canada.ca/data/dataset/a71ab99c-6756-4e56-9d2e-2a63246a5e94>). This is the same  
108 definition developed for the federal recovery strategy for the boreal population of woodland  
109 caribou in Canada (Pasher et al. 2013; Environment and Climate Change Canada 2020). This  
110 corresponded to the general vintage of the telemetry data but there were probably some points in  
111 areas classified as disturbed that weren't disturbed when the caribou were there.

112 From this disturbance mapping we stratified sources stratified into: linear features (e.g., roads,  
113 seismic lines; all buffered by 500 m), polygonal anthropogenic features (e.g., recent forest  
114 cutblocks, well pads; all buffered by 500 m), and recent fires (unbuffered). Because these  
115 disturbances often overlapped, we assigned the following priority: linear features, otherwise  
116 polygonal anthropogenic features, otherwise recent fires.

117 The two study areas differed in their habitat disturbance profiles (ignoring overlapping  
118 disturbance features). Buffered linear features covered 57% of the Bistcho range but only 6% of  
119 the Trout Lake region, while 15% of the Trout Lake region was covered by buffered polygonal

120 disturbance and 13% of Bistcho. Recent fires covered 44% of Bistcho and 17% of the Trout  
121 Lake region.

122 We also included time since disturbance as a predictor, based on Landsat data from 1985-2018  
123 and from provincial fire databases for older disturbances. We stratified fire disturbance in the  
124 analysis into the following states:  $\leq 40$  years, 40-80 years, and  $> 80$  years.

### 125 **Ground calibration of landscape variables**

126 We calibrated remotely sensed estimates of understorey stand density, crown coverage, and  
127 conifer basal area coverages using field data collected at 107 plots in the Bistcho range and 109  
128 plots in the Trout Lake region. We present methods to determine sample plot locations and  
129 details of the data collected and calibration in the Supplementary Material.

### 130 **Caribou habitat use**

131 We acquired caribou telemetry data from Alberta and Ontario government databases for the most  
132 recent, approximately 5-year, periods available. All were GPS locations collected on adult  
133 female caribou collared by net-gunning individuals from helicopters in late winter. We assigned  
134 seasons to each location based on the following: snow-free: May-October and snow-covered:  
135 November-April.

136 There were 90,540 telemetry locations available for the analysis that fell within the bounds of the  
137 Bistcho range, collected on 31 collared caribou between 1 January 2015 and 5 November 2019.  
138 Within the Trout Lake region, there were 102,667 locations collected from 60 caribou between  
139 22 February 2010 and 8 July 2015.

### 140 **Analysis and modelling**

141 We overlaid telemetry points on each landscape habitat variable to assemble the dataset for the  
142 analysis. We then generated random points from the 100% minimum convex polygons to

143 represent habitats “available” to individual caribou. We used a number of random points equal to  
144 the number of observations to prevent overfitting to an oversampled class. We removed  
145 telemetry and random points that were located within mapped lakes and double-lined rivers,  
146 along with any caribou with <200 telemetry points. The resulting dataset for Bistcho was 57,076  
147 telemetry points collected from 20 caribou and for Trout Lake was 98,590 telemetry points  
148 collected from 50 caribou.

149 Using the binary target variable *Location*, consisting of both random and telemetry points, we fit  
150 a Bayesian Network model to the data. A Bayesian Network is a directed acyclic graph  
151 consisting of nodes (random variables) and edges (arrows between nodes) that represent  
152 probabilistic relationships among variables, in this case various landscape predictors and the  
153 target node, *Location*. Each variable is assigned two or more “states” that represent the range of  
154 values that the variable can take. States can be categorical or ordinal, with continuous values  
155 stratified or “discretized” into ordinal bins. The probabilistic relationships are encoded in either  
156 marginal (for nodes with no incoming edges) or joint (for nodes with one or more incoming  
157 edges) probability tables associated with each node in the graph.

158 We generated model structures using the Sons and Spouses structural learning algorithm  
159 (Costello et al. 2020) and fit parameters by expectation maximization (Bilmes 1998). The  
160 resulting networks predicted the probability of a location being a telemetry or random point,  
161 based on evidence provided by the values of the habitat predictors at the location. For example,  
162 in the case of a habitat vector (i.e., a set of habitat predictors and their values) within which an  
163 equal number of observations and random points are located, the probability of a location being  
164 classified as an observation would be 50%, indicating no selection by caribou. Therefore we  
165 interpreted a probability of >50% of being classified as a telemetry point to be evidence of



166 selection by caribou and <50% as evidence of avoidance (Wilson and DeMars 2015). Note that  
167 this differs from the definition of selection typically applied in resource selection functions (Lele  
168 et al. 2013) and that the reported inferences are exact probabilities and have no confidence  
169 intervals.

170 While states were discretized, we used “virtual evidence” (Bilmes 2004; Mrad et al. 2015) to  
171 interpolate continuous response curves for predictor variables. We generated response curves for  
172 each predictor in turn, holding all other predictors constant, by employing Jouffe’s likelihood  
173 matching (Conrady and Jouffe 2015). Matching ensures that the multivariate distributions of the  
174 subsamples being compared are as similar as possible, except for the predictor of interest, to  
175 isolate its independent effect. Matching is a common statistical technique that usually relies on  
176 subsetting samples to achieve similar distributions; however, likelihood matching achieves the  
177 same effect on the basis of the joint probability distribution represented by the Bayesian network.  
178 We assessed the fit of the final models using k-fold ( $k = 10$ ) cross-validation (Fielding and Bell  
179 1997), resulting confusion matrices, and by receiver-operator characteristic curves (Metz 1978).  
180 We measured the relative contribution of each predictor to the target node by the mutual  
181 information shared by the predictor and target (Scutari and Denis 2021). We used BayesiaLab  
182 10.2 (Bayesia SAS, Laval, France) for all analyses.

## 183 **Results**

184 For the Bistcho range, model edges linked all but the individual and season predictors to the  
185 target variable (Figure 2). There were also strong associations among several predictors; notably,  
186 understorey stand density, conifer basal area, crown coverage, and canopy height (Table S3).

187 The predictors with the strongest relative associations with the target node were crown coverage,

188 followed by elevation and understorey stand density. Collectively these three factors alone  
189 explained >75% of the mutual information with the target node described by all of the predictors.  
190 Model structure was similar for Trout Lake, with the learned model structure excluding links  
191 between individual and season with the target node, and similar correlations among the  
192 predictors (Figure 2, Table S4). The predictors with the strongest relative associations with the  
193 target node were habitat disturbance, conifer basal area, and understory stand density (explaining  
194 >67% of the total mutual information).

195 K-fold cross-validation indicated a reasonable fit of the final Bistcho model with an ROC index  
196 of 77.1% (Figure 3) and a mean precision (percentage of *actual* telemetry or random points  
197 *predicted* by the model to be telemetry or random points, respectively) of 76.7% for telemetry  
198 points and 63.5% for random points. The mean reliability (percentage of *predicted* telemetry or  
199 random points that were *actual* telemetry or random points, respectively) was 67.7% for  
200 telemetry points and 73.1% for random points. Fit of the Trout Lake model was similar to that  
201 for the Bistcho, with an ROC index of 76.0% (Figure 4) and a mean precision of 75.6% for  
202 telemetry points and 61.5% for random points. The mean reliability was 66.3% for telemetry  
203 points and 71.6% for random points.

204 At the home range scale modelled in this study, caribou responded similarly to understorey stand  
205 density in both study areas by selecting lower densities (Figure 4). However, caribou in the  
206 different regions responded differently to the other modelled predictors. Specifically, Bistcho  
207 caribou preferred non-forested or sparsely forested areas as indicated by low crown coverage and  
208 moderate stand canopy heights, while Trout Lake caribou selected moderate crown coverages  
209 and taller canopies, indicating selection for denser forests than caribou in the Bistcho range. Both  
210 Bistcho and Trout Lake caribou favoured purer conifer stands over mixedwood forests, with

211 Bistcho caribou avoiding forested stands altogether (i.e., selection consistently  $<0.5$ ), but  
212 avoiding conifer stands less than mixedwood (i.e., still a positive slope with increasing conifer  
213 basal area)  
214 Caribou in the Trout Lake region selected moderate elevations, while in the Bistcho they avoided  
215 uplands. In both study areas, caribou selected undisturbed habitat and avoided buffered linear  
216 and polygonal disturbances (Figure 5). Trout Lake caribou avoided recently burned areas within  
217 their home ranges but caribou in the Bistcho did not.

## 218 **Discussion**

219 This is the first study to measure understorey stand density based on remotely sensed data and to  
220 estimate its effect on habitat selection by caribou. We demonstrate the consistent effect of  
221 understorey stand density in two regions otherwise differing in the way caribou responded to  
222 other biophysical characteristics of the forests. Research has linked small-scale, cognitive  
223 foraging behaviour by caribou faced with predation risk (Avgar et al. 2013, 2015) to the viability  
224 of entire caribou populations (Fryxell et al. 2020) under large-scale habitat change (McGreer et  
225 al. 2015; Mallon et al. 2016). These studies suggest that habitat suitable for caribou depends not  
226 only on the type and quality of available food, but also on the energy costs of movement to  
227 obtain that food and to avoid predators. That caribou in both of our study areas selected areas of  
228 lower understorey stand density is consistent with this.

229 Some ground lichens on which caribou depend thrive on low productivity sites that are often  
230 associated with open forest canopies (Brodo et al. 2001; Lesmerises et al. 2011; Silva et al. 2019;  
231 Hämäläinen et al. 2020), perhaps coincident with lower understorey stand densities. We were not  
232 able to link a commensurate estimate of lichen abundance to our remotely sensed estimate of  
233 understorey stand density that would enable an adjustment for a potential forage effect.

234 However, we were able to adjust for crown coverage and found that caribou still selected stands  
235 with lower understorey stand density. This provided some confidence that our results were not  
236 confounded by the unobserved abundance of ground lichens, but the addition of reliable lichen  
237 mapping might improve model performance. Regardless, management interventions that reduce  
238 understorey stand density could be neutral or positive for caribou on sites that are otherwise  
239 favourable for lichens, if such treatments were sufficient to improve caribou energy balance.  
240 Lamont et al. (2019) recommended removing standing dead and downed trees in stands killed by  
241 mountain pine beetle (*Dendroctonus ponderosae*) to reduced locomotion costs of elk (*Cervus*  
242 *canadensis*), after observing that elk avoided beetle-killed areas. Nobert et al. (2020) suggested  
243 that mountain caribou populations might benefit from similar treatments where infestations  
244 affected pine-lichen winter ranges but cautioned that wolves might also benefit from such  
245 clearing.

246 The risk of confounding by the abundance of forest shrubs is more difficult to estimate. Open  
247 forest canopies can promote shrub growth on productive sites (e.g., Paulson et al. 2021), and  
248 caribou have broad diets during the snow-free season (Denryter et al. 2017); however, we found  
249 little evidence of seasonal variation in habitat selection by caribou on either study area.

250 Understorey density might reduce the travel speed of caribou. Dickie et al. (2022) found that  
251 caribou movements on seismic lines slowed when lines were subject to various restoration  
252 treatments intended to impede the movements of wolves, including the roll-back of coarse  
253 woody debris and the felling of trees.. There was no indication from our analysis that caribou  
254 travelling more slowly through denser understorey was sufficient to bias our estimates of  
255 selection in favour of these habitats.

256 The response by caribou to understorey density was remarkably similar in both study areas,  
257 given the wider variation in their response to other habitat predictors. In the Bistcho, caribou  
258 avoided forested uplands while caribou in the Trout Lake region were less discriminating,  
259 generally selecting forests at moderate elevations with moderate crown coverages.

260 In northwestern Alberta, lowland treed bogs and fens are low productivity environments that are  
261 generally avoided by moose and their main predator, wolves (Latham et al. 2011; DeMars and  
262 Boutin 2018; Serrouya et al. 2021). Moose in that region are more common in productive upland  
263 forests and, in particular, those with a significant deciduous component (Routh and Nielsen  
264 2021). That caribou are largely segregated spatially from moose via their habitat preferences is  
265 hypothesized to be key to sustaining caribou populations, due to their susceptibility to apparent  
266 competition with moose (James et al. 2004; DeCesare et al. 2009). On the other hand, in  
267 northwestern Ontario, the exposed bedrock and shallow soils of the Canadian Shield can limit  
268 the productivity of upland coniferous forests and caribou select low-volume jack pine and black  
269 spruce forests with abundant lichen and few shrubs (Antoniak and Cumming 1998). In contrast,  
270 moose in this region use lowland aquatic areas and more productive deciduous and mixedwood  
271 forests (Street et al. 2015). Nevertheless, the result is spatial segregation between caribou and  
272 moose (Cumming et al. 1996) in a manner similar to the Bistcho.

273 While caribou in Bistcho and Trout Lake regions pursued different tactics with respect to  
274 uplands and lowlands, and forested versus open habitats, overall, the results suggested caribou  
275 were following a similar strategy: selection for forest stands with higher conifer components and  
276 away from areas with significant deciduous components – a known habitat feature favoured by  
277 moose. We contend that both the differences and similarities in habitat selection exhibited by  
278 caribou in the two study areas were consistent with respect to seeking refuge from predators.

279 Habitat disturbance caused by anthropogenic activity and fire is correlated with demographic  
280 decline among woodland caribou subpopulations in Canada (Johnson et al. 2020) and recovery  
281 from disturbance is a focus of the national recovery strategy (Environment and Climate Change  
282 Canada 2020). The disturbance profiles of the two study areas differed due to differences in land  
283 use. The Bistcho range has experienced significant oil and gas exploration and development  
284 while forestry is restricted to the productive uplands of the southeastern portion of the range. As  
285 a result, the Bistcho range is associated with a high density of seismic lines, pipeline corridors,  
286 and industrial roads. Well pads are common but relatively small clearings (<2 ha), and the  
287 limited spatial extent of forestry means that there is little anthropogenic polygonal disturbance.  
288 In contrast, forestry is the main industrial activity in the Trout Lake region, which has resulted in  
289 less linear development, but a higher proportion of recent forestry cutblocks than in the Bistcho.  
290 High densities of linear features provide efficient travel corridors for wolves (Dickie et al. 2017)  
291 and can lead to the loss of the predation refugia thought necessary to sustain caribou (DeMars  
292 and Boutin 2018). Consistently, juvenile recruitment rates in the Bistcho range are less than half  
293 the estimates for subpopulations overlapping the Trout Lake study area (Johnson et al. 2020).  
294 Caribou in both study areas selected undisturbed habitat, but avoidance of fire was evident only  
295 in the Trout Lake region. In contrast, both buffered linear and polygonal disturbance were  
296 avoided in both areas. This builds on recent evidence that the relationship between caribou and  
297 fire is complex (Dalerum et al. 2007; Skatter et al. 2017; DeMars et al. 2019; Konkolics et al.  
298 2021) and our study suggest that different habitat characteristics may lead to different responses  
299 to fire.

300 This study underlines the importance of addressing the functional basis for caribou habitat  
301 selection behaviour when planning recovery actions. While analyses revealed both similarities

302 (i.e., stand density, conifer basal area) and differences (i.e., upland versus lowland, open versus  
303 forested habitats) in habitat selection patterns between geographically and biophysically distinct  
304 regions, caribou appeared to pursue similar strategies (i.e., avoiding predators, selecting for ease  
305 of movement), albeit with different tactics. Using high-resolution satellite data provided the  
306 opportunity to resolve habitat characteristics more consistently, in greater detail, and over larger  
307 areas than previous studies and allowed us to link structural elements of the forest to the  
308 functional requirements of caribou. We conclude that applying coarse-scale policies based  
309 simply on stand age or stand type may not be appropriate in different parts of caribou range, and  
310 that prescriptions necessary to restore or sustain caribou habitat will need to be adapted to local  
311 conditions, despite caribou facing common pathways to decline.

312 As recommended for mountain caribou (Nobert et al. (2020)), forest management prescriptions to  
313 address ease of movement by caribou might be appropriate in boreal ranges. Best practices could  
314 include harvesting strategies such as thinning, log processing and brush piling at roadsides to  
315 avoid high volumes of on-site coarse woody debris, burning and light scarification for site  
316 preparation (except where lichen mats are intact), and replanting at low stocking densities.  
317 Further work is required to understand how wolf mobility may also be enhanced by these  
318 treatments and whether interventions can be designed to avoid enhancing habitat suitability for,  
319 and mobility of, primary prey and wolves.

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### 328 **Competing Interests Statement**

329 Phil Green is CEO of First Management Resource Group, Inc., which was contracted to produce  
330 the SkyForest™ digital mapping products for the study. The other authors declare that they have  
331 no known competing financial interests or personal relationships that influenced the work  
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### 333 **Author Contribution Statement**

334 **SFW:** Conceptualization, data curation, methodology, formal analysis, writing, editing. **TDN:**  
335 Conceptualization, methodology, writing, editing. **PG:** Conceptualization, data curation,  
336 methodology, writing, editing. **ADV:** Funding acquisition, supervision, conceptualization,  
337 reviewing.

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### 341 **Data Availability Statement**

342 Use of caribou telemetry data was governed by limited agreements with the Province of Ontario  
343 and the Province of Alberta. Mapping products were developed commercially by First Resource  
344 Management Group, Inc. but coverages for the study areas, as well as plot data and model script  
345 files are available for download (<https://doi.org/10.17605/OSF.IO/GDNC9>).

### 346 **References**



347 Antoniak, K., and Cumming, H.G. 1998. Analysis of forest stands used by wintering woodland  
348 caribou in Ontario. *Rangifer* **18**(5): 157. doi:10.7557/2.18.5.1553.

349 Avgar, T., Baker, J.A., Brown, G.S., Hagens, J.S., Kittle, A.M., Mallon, E.E., McGreer, M.T.,  
350 Mosser, A., Newmaster, S.G., Patterson, B.R., Reid, D.E.B., Rodgers, A.R., Shuter, J.,  
351 Street, G.M., Thompson, I., Turetsky, M.J., Wiebe, P.A., and Fryxell, J.M. 2015. Space-  
352 use behaviour of woodland caribou based on a cognitive movement model. *J. Anim.*  
353 *Ecol.* **84**(4): 1059–1070. doi:10.1111/1365-2656.12357.

354 Avgar, T., Mosser, A., Brown, G.S., and Fryxell, J.M. 2013. Environmental and individual  
355 drivers of animal movement patterns across a wide geographical gradient. *J. Anim. Ecol.*  
356 **82**(1): 96–106. doi:10.1111/j.1365-2656.2012.02035.x.

357 Bilmes, J. 2004. On virtual evidence and soft evidence in Bayesian networks. University of  
358 Washington, Seattle, WA.

359 Bilmes, J.A. 1998. A gentle tutorial of the EM algorithm and its application to parameter  
360 estimation for Gaussian mixture and hidden Markov models. *Int. Comput. Sci. Inst.*  
361 **4**(510): 126.

362 Briand, Y., Ouellet, J.-P., Dussault, C., and St-Laurent, M.-H. 2009. Fine-scale habitat selection  
363 by female forest-dwelling caribou in managed boreal forest: Empirical evidence of a  
364 seasonal shift between foraging opportunities and antipredator strategies. *Écoscience*  
365 **16**(3): 330–340. doi:10.2980/16-3-3248.

366 Brodo, I.M., Sharnoff, S.D., and Sharnoff, S. 2001. *Lichens of North America*. Yale University  
367 Press, New Haven.

368 Conrady, S., and Jouffe, L. 2015. *Bayesian networks and BayesiaLab: a practical introduction*  
369 *for researchers*. Bayesia USA, Franklin, TN.

370 Costello, F.J., Kim, C., Kang, C.M., and Lee, K.C. 2020. Identifying high-risk factors of  
371 depression in middle-aged persons with a novel sons and spouses Bayesian network  
372 model. *Healthcare* **8**(4): 562. doi:10.3390/healthcare8040562.

373 Crins, W.J., Gray, P.A., Uhlig, P.W.C., and Wester, M.C. 2009. The ecosystems of Ontario, part  
374 1: Ecozones and ecoregions. Technical Report, Ontario Ministry of Natural Resources,  
375 Peterborough, ON.

376 Cumming, H.G., Beange, D.B., and Lavoie, G. 1996. Habitat partitioning between woodland  
377 caribou and moose in Ontario: the potential role of shared predation risk. *Rangifer* **16**(4):  
378 81. doi:10.7557/2.16.4.1224.

379 Dalerum, F., Boutin, S., and Dunford, J.S. 2007. Wildfire effects on home range size and fidelity  
380 of boreal caribou in Alberta, Canada. *Can. J. Zool.* **85**(1): 26–32. doi:10.1139/z06-186.

381 DeCesare, N.J., Hebblewhite, M., Robinson, H.S., and Musiani, M. 2009. Endangered,  
382 apparently: the role of apparent competition in endangered species conservation:  
383 Apparent competition and endangered species. *Anim. Conserv.* **13**(4): 353–362.  
384 doi:10.1111/j.1469-1795.2009.00328.x.

385 DeMars, C.A., and Boutin, S. 2018. Nowhere to hide: Effects of linear features on predator–prey  
386 dynamics in a large mammal system. *J. Anim. Ecol.* **87**(1): 274–284. doi:10.1111/1365-  
387 2656.12760.

388 DeMars, C.A., Serrouya, R., Mumma, M.A., Gillingham, M.P., McNay, R.S., and Boutin, S.  
389 2019. Moose, caribou, and fire: have we got it right yet? *Can. J. Zool.* **97**(10): 866–879.  
390 doi:10.1139/cjz-2018-0319.

391 Denryter, K., Cook, R.C., Cook, J.G., and Parker, K.L. 2022. Animal-defined resources reveal  
392 nutritional inadequacies for woodland caribou during summer–autumn. *J. Wildl. Manag.*  
393 **86**(2). doi:10.1002/jwmg.22161.

394 Denryter, K.A., Cook, R.C., Cook, J.G., and Parker, K.L. 2017. Straight from the caribou’s (  
395 *Rangifer tarandus* ) mouth: detailed observations of tame caribou reveal new insights into  
396 summer–autumn diets. *Can. J. Zool.* **95**(2): 81–94. doi:10.1139/cjz-2016-0114.

397 Dickie, M., Serrouya, R., McNay, R.S., and Boutin, S. 2017. Faster and farther: wolf movement  
398 on linear features and implications for hunting behaviour. *J. Appl. Ecol.* **54**(1): 253–263.  
399 doi:10.1111/1365-2664.12732.

400 Dickie, M., Sherman, G.G., Sutherland, G.D., McNay, R.S., and Cody, M. 2022. Evaluating the  
401 impact of caribou habitat restoration on predator and prey movement. *Conserv. Biol.* **37**:  
402 e14004. doi:10.1111/cobi.14004.

403 Environment and Climate Change Canada. 2020. Amended recovery strategy for the woodland  
404 caribou (*Rangifer tarandus caribou*), boreal population, in Canada. Environment and  
405 Climate Change Canada, Ottawa, ON. Available from  
406 [http://publications.gc.ca/collections/collection\\_2021/eccc/En3-4-140-2020-eng.pdf](http://publications.gc.ca/collections/collection_2021/eccc/En3-4-140-2020-eng.pdf)  
407 [accessed 22 April 2021].

408 Fielding, A.H., and Bell, J.F. 1997. A review of methods for the assessment of prediction errors  
409 in conservation presence/absence models. *Environ. Conserv.* **24**(1): 38–49.  
410 doi:10.1017/S0376892997000088.

411 Fryxell, J.M., Avgar, T., Liu, B., Baker, J.A., Rodgers, A.R., Shuter, J., Thompson, I.D., Reid,  
412 D.E.B., Kittle, A.M., Mosser, A., Newmaster, S.G., Nudds, T.D., Street, G.M., Brown,  
413 G.S., and Patterson, B. 2020. Anthropogenic disturbance and population viability of

414 woodland caribou in Ontario. *J. Wildl. Manag.* **84**(4): 636–650.  
415 doi:10.1002/jwmg.21829.

416 Hämäläinen, A., Strengbom, J., and Ranius, T. 2020. Low-productivity boreal forests have high  
417 conservation value for lichens. *J. Appl. Ecol.* **57**(1): 43–54. doi:10.1111/1365-  
418 2664.13509.

419 James, A.R.C., Boutin, S., Hebert, D.M., and Rippin, A.B. 2004. Spatial separation of caribou  
420 from moose and its relation to predation by wolves. *J. Wildl. Manag.* **68**(4): 799–809.  
421 doi:10.2193/0022-541X(2004)068[0799:SSOCFM]2.0.CO;2.

422 Johnson, C.A., Sutherland, G.D., Neave, E., Leblond, M., Kirby, P., Superbie, C., and  
423 McLoughlin, P.D. 2020. Science to inform policy: Linking population dynamics to  
424 habitat for a threatened species in Canada. *J. Appl. Ecol.* **57**(7): 1314–1327.  
425 doi:10.1111/1365-2664.13637.

426 Keim, J.L., DeWitt, P.D., Wilson, S.F., Fitzpatrick, J.J., Jenni, N.S., and Lele, S.R. 2021.  
427 Managing animal movement conserves predator–prey dynamics. *Front. Ecol. Environ.*:  
428 fee.2358. doi:10.1002/fee.2358.

429 Konkolics, S., Dickie, M., Serrouya, R., Hervieux, D., and Boutin, S. 2021. A burning question:  
430 What are the implications of forest fires for woodland caribou? *J. Wildl. Manag.*:  
431 jwmg.22111. doi:10.1002/jwmg.22111.

432 Lamont, B.G., Monteith, K.L., Merkle, J.A., Mong, T.W., Albeke, S.E., Hayes, M.M., and  
433 Kauffman, M.J. 2019. Multi-scale habitat selection of elk in response to beetle-killed  
434 forest. *J. Wildl. Manag.* **83**(3): 679–693. doi:10.1002/jwmg.21631.

435 Latham, A.D.M., Latham, M.C., Mccutchen, N.A., and Boutin, S. 2011. Invading white-tailed  
436 deer change wolf-caribou dynamics in northeastern Alberta: Deer change wolf-caribou  
437 dynamics. *J. Wildl. Manag.* **75**(1): 204–212. doi:10.1002/jwmg.28.

438 Lele, S.R., Merrill, E.H., Keim, J., and Boyce, M.S. 2013. Selection, use, choice and occupancy:  
439 clarifying concepts in resource selection studies. *J. Anim. Ecol.* **82**(6): 1183–1191.  
440 doi:10.1111/1365-2656.12141.

441 Lesmerises, R., Ouellet, J.-P., and St-Laurent, M.-H. 2011. Assessing terrestrial lichen biomass  
442 using ecoforest maps: a suitable approach to plan conservation areas for forest-dwelling  
443 caribou. *Can. J. For. Res.* **41**(3): 632–642. doi:10.1139/X10-229.

444 Mallon, E.E., Turetsky, M.R., Thompson, I.D., Fryxell, J.M., and Wiebe, P.A. 2016. Effects of  
445 disturbance on understory succession in upland and lowland boreal forests and  
446 implications for woodland caribou (*Rangifer tarandus caribou*). *For. Ecol. Manag.* **364**:  
447 17–26. doi:10.1016/j.foreco.2015.12.001.

448 McGreer, M.T., Mallon, E.E., Vander Vennen, L.M., Wiebe, P.A., Baker, J.A., Brown, G.S.,  
449 Avgar, T., Hagens, J., Kittle, A.M., Mosser, A., Street, G.M., Reid, D.E.B., Rodgers,  
450 A.R., Shuter, J., Thompson, I.D., Turetsky, M.J., Newmaster, S.G., Patterson, B.R., and  
451 Fryxell, J.M. 2015. Selection for forage and avoidance of risk by woodland caribou  
452 (*Rangifer tarandus caribou*) at coarse and local scales. *Ecosphere* **6**(12): art288.  
453 doi:10.1890/ES15-00174.1.

454 Metz, C.E. 1978. Basic principles of ROC analysis. *Semin. Nucl. Med.* **8**(4): 283–298.  
455 doi:10.1016/S0001-2998(78)80014-2.

456 Ministry of Natural Resources and Forestry. 2014. State of the woodland caribou resource report.  
457 Species at Risk Branch, Thunder Bay, Ontario.

458 Mrad, A.B., Delcroix, V., Piechowiak, S., Leicester, P., and Abid, M. 2015. An explication of  
459 uncertain evidence in Bayesian networks: likelihood evidence and probabilistic evidence:  
460 Uncertain evidence in Bayesian networks. *Appl. Intell.* **43**(4): 802–824.  
461 doi:10.1007/s10489-015-0678-6.

462 Nobert, B.R., Larsen, T.A., Pigeon, K.E., and Finnegan, L. 2020. Caribou in the cross-fire?  
463 Considering terrestrial lichen forage in the face of mountain pine beetle (*Dendroctonus*  
464 *ponderosae*) expansion. *PLOS ONE* **15**(4): e0232248. doi:10.1371/journal.pone.0232248.

465 Pasher, J., Seed, E., and Duffe, J. 2013. Development of boreal ecosystem anthropogenic  
466 disturbance layers for Canada based on 2008 to 2010 Landsat imagery. *Can. J. Remote*  
467 *Sens.* **39**(1): 42–58. doi:10.5589/m13-007.

468 Paulson, A.K., Peña, H., Alexander, H.D., Davydov, S.P., Loranty, M.M., Mack, M.C., and  
469 Natali, S.M. 2021. Understory plant diversity and composition across a postfire tree  
470 density gradient in a Siberian Arctic boreal forest. *Can. J. For. Res.* **51**(5): 720–731.  
471 doi:10.1139/cjfr-2020-0483.

472 Pojar, J. 1996. Environment and biogeography of the western boreal forest. *For. Chron.* **72**(1):  
473 51–58. doi:10.5558/tfc72051-1.

474 Routh, M.R., and Nielsen, S.E. 2021. Dynamic patterns in winter ungulate browse succession in  
475 the Boreal Plains of Alberta. *For. Ecol. Manag.* **492**: 119242.  
476 doi:10.1016/j.foreco.2021.119242.

477 Serrouya, R., Dickie, M., Lamb, C., van Oort, H., Kelly, A.P., DeMars, C., McLoughlin, P.D.,  
478 Larter, N.C., Hervieux, D., Ford, A.T., and Boutin, S. 2021. Trophic consequences of  
479 terrestrial eutrophication for a threatened ungulate. *Proc. R. Soc. B Biol. Sci.* **288**(1943):  
480 20202811. doi:10.1098/rspb.2020.2811.

481 Silva, J., Nielsen, S., Lamb, C., Hague, C., and Boutin, S. 2019. Modelling lichen abundance for  
482 woodland caribou in a fire-driven boreal landscape. *Forests* **10**(11): 962.  
483 doi:10.3390/f10110962.

484 Skatter, H.G., Charlebois, M.L., Eftestøl, S., Tsegaye, D., Colman, J.E., Kansas, J.L., Flydal, K.,  
485 and Balicki, B. 2017. Living in a burned landscape: woodland caribou (*Rangifer tarandus*  
486 *caribou*) use of postfire residual patches for calving in a high fire – low anthropogenic  
487 Boreal Shield ecozone. *Can. J. Zool.* **95**(12): 975–984. doi:10.1139/cjz-2016-0307.

488 Street, G.M., Vander Vennen, L.M., Avgar, T., Mosser, A., Anderson, M.L., Rodgers, A.R., and  
489 Fryxell, J.M. 2015. Habitat selection following recent disturbance: model transferability  
490 with implications for management and conservation of moose (*Alces alces*). *Can. J. Zool.*  
491 **93**(11): 813–821. doi:10.1139/cjz-2015-0005.

492 Strong, W.L., and Leggat, K.R. 1992. Ecoregions of Alberta. Alberta Forestry, Lands and  
493 Wildlife, Edmonton, AB.

494 Thompson, I.D., Wiebe, P.A., Mallon, E., Rodgers, A.R., Fryxell, J.M., Baker, J.A., and Reid, D.  
495 2015. Factors influencing the seasonal diet selection by woodland caribou (*Rangifer*  
496 *tarandus tarandus*) in boreal forests in Ontario. *Can. J. Zool.* **93**(2): 87–98.  
497 doi:10.1139/cjz-2014-0140.

498 Webber, Q.M.R., Ferraro, K.M., Hendrix, J.G., and Vander Wal, E. 2022. What do caribou eat?  
499 A review of the literature on caribou diet. *Can. J. Zool.* **100**(3): 197–207.  
500 doi:10.1139/cjz-2021-0162.

501 Wilson, S.F., Crosina, W., Dzus, E., Hervieux, D., McLoughlin, P.D., Trout, L.M., and Nudds,  
502 T.D. 2022. Nested population structure of threatened boreal caribou revealed by network  
503 analysis. *Glob. Ecol. Conserv.* **40**: e02327. doi:10.1016/j.gecco.2022.e02327.

504 Wilson, S.F., and DeMars, C.A. 2015. A Bayesian approach to characterizing habitat use by, and  
505 impacts of anthropogenic features on, woodland caribou (*Rangifer tarandus caribou*) in  
506 northeast British Columbia. *Can. Wildl. Biol. Manag.* 4: 107–118.

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508



509 Figure 1. Bistcho woodland caribou range (Alberta) and Trout Lake study area (Ontario). Figure  
510 was created with QGIS version 3.32.1 using public domain basemap data from  
511 <https://naturalearthdata.com>.

512

513 Figure 2. Bayesian networks illustrating the relationship between the target variable *Locations*,  
514 representing the set of caribou telemetry locations (obs) and random locations, and the habitat  
515 predictor variables for the Bistcho caribou range and Trout Lake study area. By convention, arcs  
516 are directed from the target variable to predictors and direction is arbitrary among predictors.  
517 Labels on arcs indicate the relative mutual information shared between each predictor and the  
518 target node, expressed as a percentage of the total mutual information shared between all of the  
519 predictors and the target.

520

521 Figure 3. Receiver Operator Characteristic (ROC) curves for the Bayesian network habitat  
522 selection models for the Bistcho and Trout Lake study areas. The true positive rate in this study  
523 was the proportion of telemetry points predicted by the model to be telemetry points out of all of  
524 the points (telemetry and random) predicted to be telemetry points. The false positive rate was  
525 the proportion of random locations predicted by the model to be telemetry points out of all of the  
526 points predicted to be random points.

527

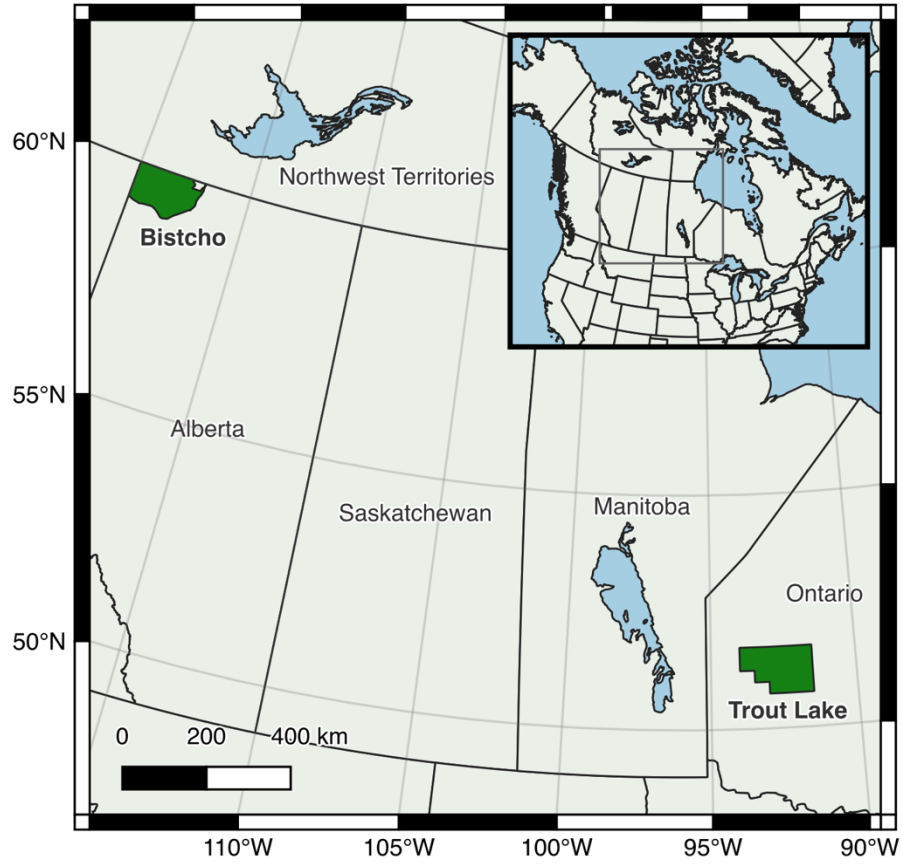
528 Figure 4. Direct effects of predictor habitat variables on the probability of a location being an  
529 observation. Y-axis values  $>0.5$  (above the horizontal line) represent habitat selection for a  
530 predictor, adjusting for all other predictors by likelihood matching. The conifer basal area  
531 relationship was restricted to forested stands only to omit stands with no basal area. Understorey

532 density is estimated by the number of steps required to traverse study plots. Understorey density,  
533 elevation, and canopy height were standardized to allow for relative comparisons between study  
534 areas. See Figure 2 for the relative strengths of these variables in contributing to the habitat  
535 selection behaviour of caribou in the two study areas.

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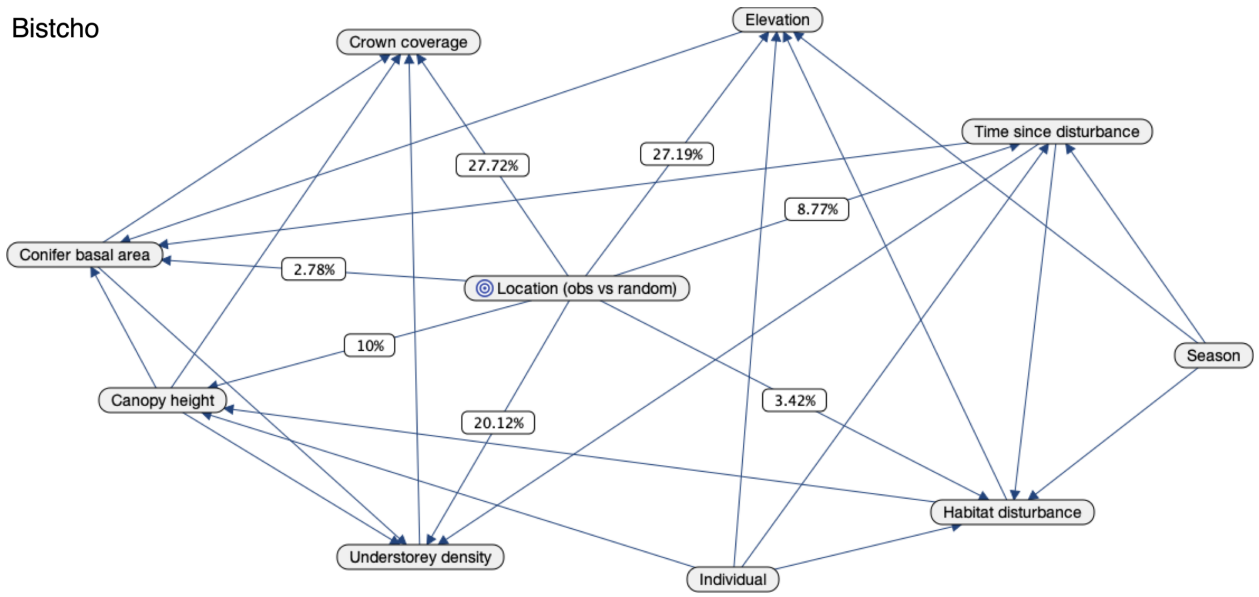
537 Figure 5. Selection by caribou of habitats with different disturbance causes and times since  
538 habitat disturbance. Values  $>0.5$  (above the horizontal line) represent habitat selection for the  
539 class.

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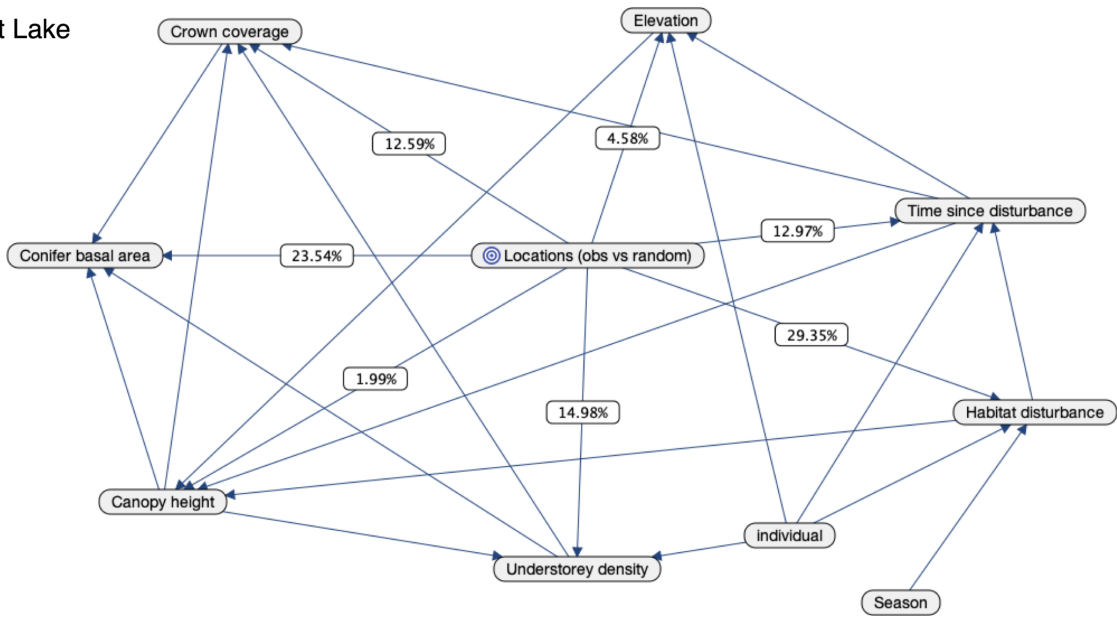


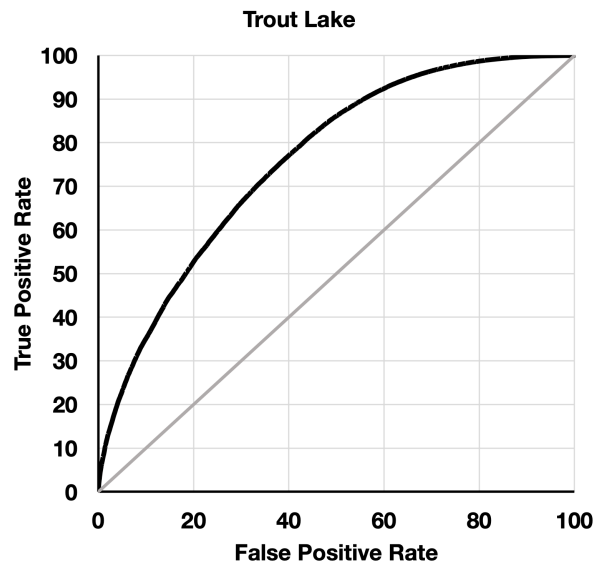
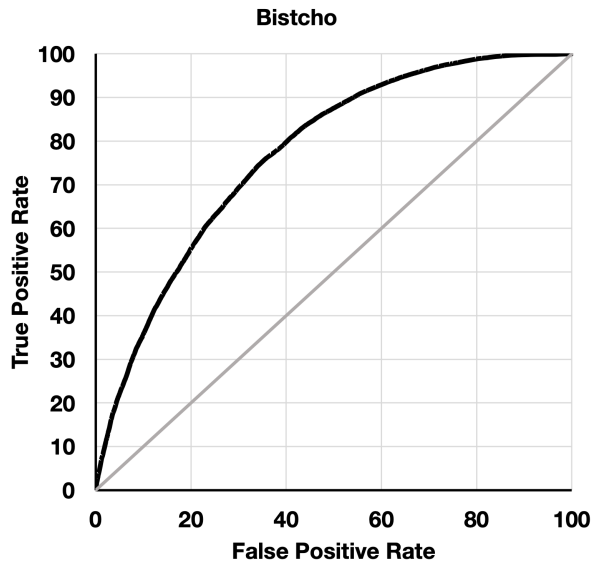
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Bistcho

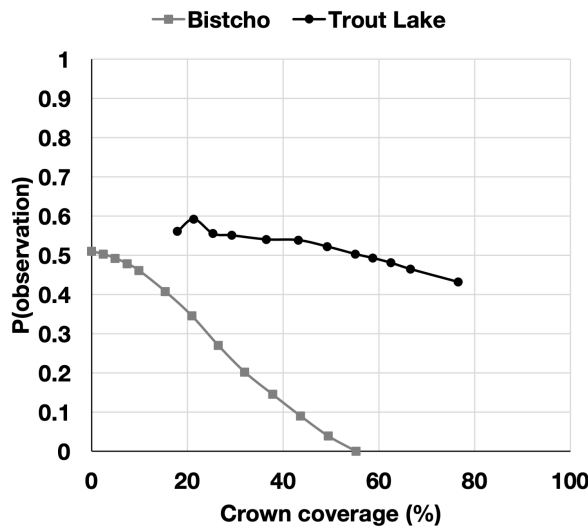
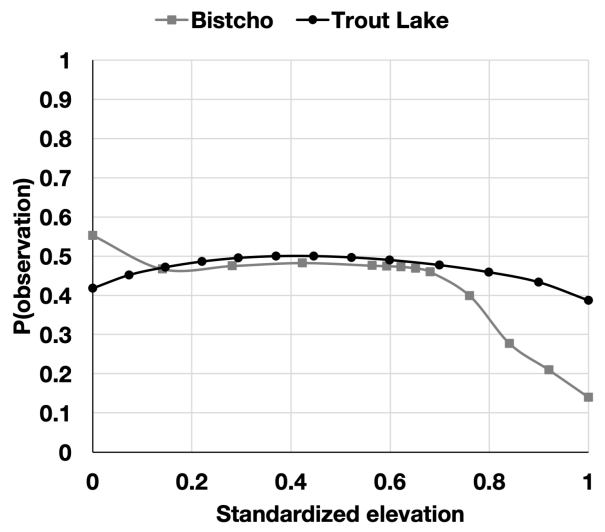
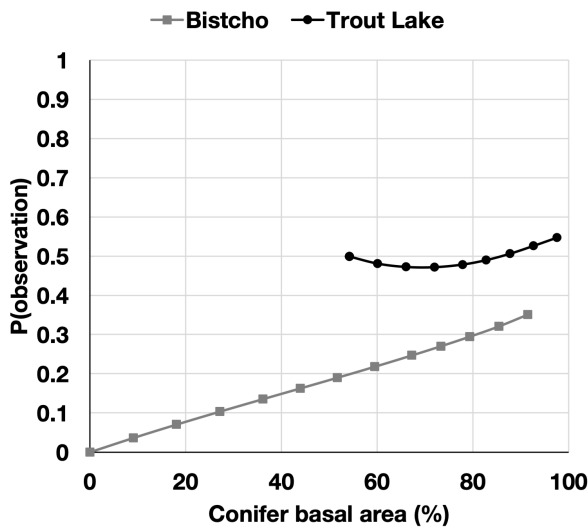
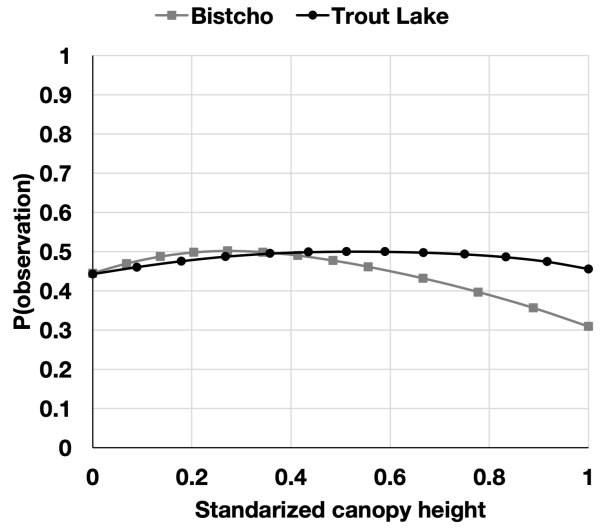
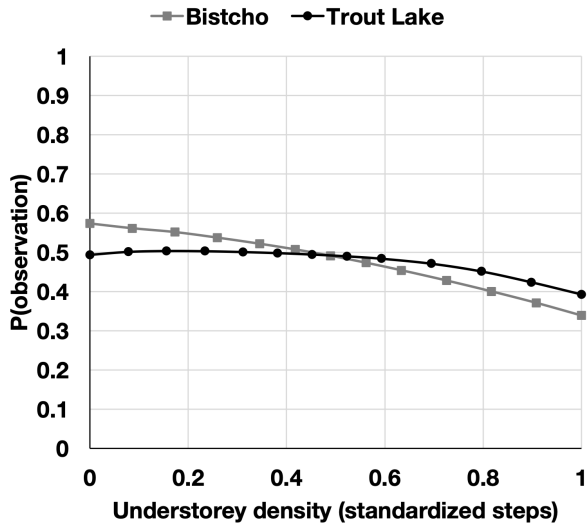


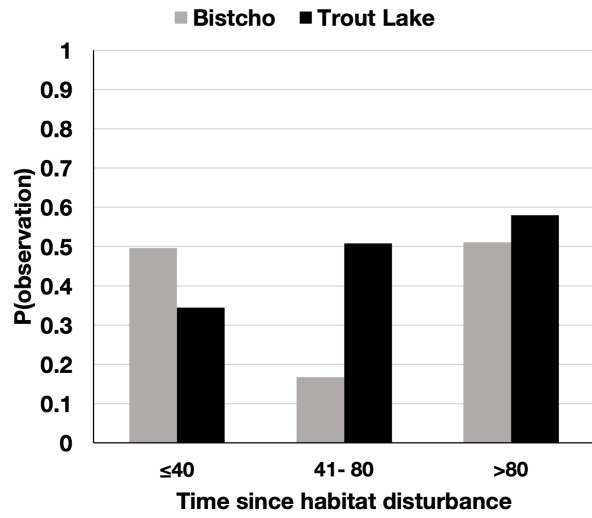
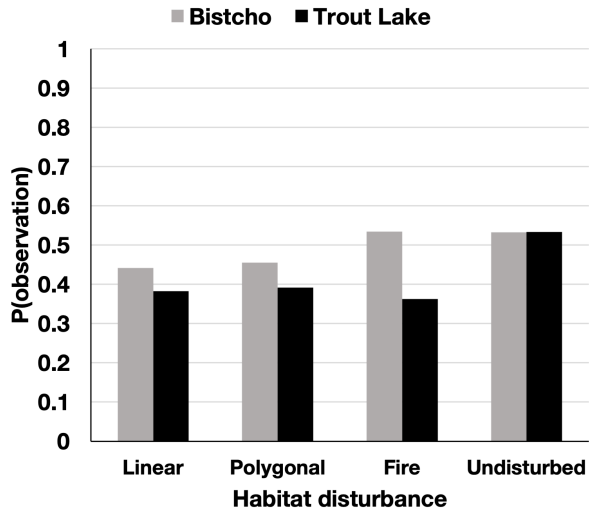
Trout Lake





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