Socio-ecological factors shape the distribution of a cultural keystone species (bearded pig, *Sus barbatus*) in Malaysian Borneo

Authors

David J. Kurz[1,2,*], Thomas Connor[1], Jedediah F. Brodie[3,4], Esther Lonnie Baking[5], Sabrina H. Szeto[6,7], Andrew Hearn[8], Penny C. Gardner[9,10], Oliver R. Wearn[11], Mairin Deith[12], Nicolas J. Deere[13], Ahmad Ampeng[14], Henry Bernard[4], Jocelyn Goon Ee Wern[10], Alys Granados[15], Olga Helmy[3], Hong-Ye Lim[10,16], Matthew Scott Luskin[17], David Macdonald[8], Joanna Ross[8], Boyd Simpson[18], Matthew Struebig[13], Jayasilan Mohd-Azlan[4], Matthew D. Potts[1], Benoit Goossens[10,19,20,21], and Justin S. Brashares[1]

Affiliations

[1]Department of Environmental Science, Policy, and Management, University of California, Berkeley, Berkeley, CA 94720, USA

[2]Environmental Science Program, Trinity College, Hartford, CT 06106, USA

[3]Division of Biological Sciences & Wildlife Biology Program, University of Montana, Missoula, MT 59812, USA

[4]Institute of Biodiversity and Environmental Conservation, Universiti Malaysia Sarawak, 94 300 Kota Samarahan, Sarawak, Malaysia

[5]Institute for Tropical Biology and Conservation, Universiti Malaysia Sabah, 88450 Kota Kinabalu, Sabah, Malaysia

[6]School of the Environment, Yale University, New Haven, CT 06511, USA

[7]Sabrina Szeto Consulting, Isen, Bavaria 84424, Germany

[8]Wildlife Conservation Research Unit, Department of Biology, The Recanati-Kaplan Centre, University of Oxford, Oxford, UK

[9]RSPB UK Headquarters, The Lodge, Sandy, Bedfordshire, SG19 2DL

[10]Danau Girang Field Centre, c/o Sabah Wildlife Department, Wisma Muis, Sabah, Malaysia

[11]Fauna & Flora International, Vietnam Programme, Hanoi, Vietnam

[12]Department of Zoology, University of British Columbia, #4200-6270 University Boulevard 15, Vancouver, British Columbia, Canada V6T 1Z4

[13]Durrell Institute of Conservation and Ecology, School of Anthropology and Conservation, University of Kent, Canterbury, UK

[14]Forest Department Sarawak, Level 15, East Wing, Bangunan Baitul Makmur II, Medan Raya, Petra Jaya, 93050 Kuching, Sarawak, Malaysia

[15]Department of Forest Resources Management, University of British Columbia, 2424 Main Mall, Vancouver, BC V6T 1Z4 Canada

[16]Forever Sabah, H30, Gaya Park, Lorong Muntahan 1C, Jalan Penampang, 88300 Kota Kinabalu, Sabah, Malaysia
[17]School of Biological Sciences, University of Queensland, St. Lucia, QLD, Australia
[18]Copenhagen Zoo, Department of Research & Conservation, Frederiksberg, Denmark
[19]Organisms and Environment Division, School of Biosciences, Cardiff University, Cardiff, UK
[20]Sabah Wildlife Department, Wisma Muis, Sabah, Malaysia
[21]Sustainable Places Research Institute, Cardiff University, Cardiff, UK

*Corresponding author: David J. Kurz, david.kurz@berkeley.edu

Abstract

Biophysical and socio-cultural factors have jointly shaped the distribution of global biodiversity, yet relatively few studies have quantitatively assessed the influence of social and ecological landscapes on wildlife distributions. We sought to determine whether social and ecological covariates shape the distribution of a cultural keystone species, the bearded pig (*Sus barbatus*). Drawing on a dataset of 295 total camera trap locations and 25,755 trap days across 18 field sites and three years in Sabah and Sarawak, Malaysian Borneo, we fitted occupancy models incorporating socio-cultural covariates and environmental covariates hypothesized to be associated with bearded pig occupancy. We found that all competitive occupancy models included both socio-cultural covariates and ecological covariates. Moreover, we found quantitative evidence supporting Indigenous pig hunting rights: high levels of Indigenous pig-hunting groups were positively associated with pig occupancy in low-accessibility areas, and medium and low levels of Indigenous pig-hunting groups were positively associated with pig occupancy in high-accessibility areas. These results suggest that bearded pig populations in Malaysian Borneo should be managed with context-specific strategies, promoting Indigenous pig hunting rights. We also provide important baseline information on bearded pig occupancy levels prior to the 2020-2021 outbreak of African Swine Fever (ASF), which caused social and ecological concerns after mass dieoffs of bearded pigs in Borneo.

Introduction

Management of socio-ecological systems must accommodate both human needs and wildlife persistence (Kays et al. 2016). Socio-cultural and biophysical landscapes are fundamentally connected, but our empirical understanding of the links between them is still limited. While researchers often quantitatively examine links between biophysical factors and wildlife, far less quantitative work has been carried out on the influence of socio-cultural factors on wildlife distributions. Yet socio-cultural factors—such as ethnic identity, culturally-distinctive hunting practices, armed conflict, recreation, feasts, traditions, and value systems—have been shown to have far-reaching implications for animal behavior, wildlife distributions, and conservation efforts (e.g., Heberlein & Ericsson 2006, Wong et al. 2009, Riley 2010, Gaynor et al. 2016, Kurz et al. 2021). As such, social and cultural practices, tolerances, affinities, and other socio-cultural factors require more attention as important predictor variables, alongside ecological variables, for determining occurrence patterns of many wildlife species (Karanth et al. 2009).

A primary challenge has been integrating nuanced quantitative measures of socio-cultural factors into wildlife distribution modeling. Recently, socio-cultural covariates have begun to move beyond broader indices of human disturbance or footprint (e.g. Barber-Meyer et al. 2013, Linkie et al. 2013) to include culturally-shaped metrics, such as hunting accessibility, social carrying capacity for development, or religious practices (e.g., Bettigole et al. 2014, Stahlecker et al. 2017, Deith & Brodie 2020). These more recently adopted metrics reflect an emerging understanding of the conceptual complexity of human-wildlife interactions, which take place within nested social and ecological systems (Lischka et al. 2018). While the number of socio-ecological studies has increased dramatically in recent years (Guerrero et al. 2018), greater attention is needed to the integration of social and ecological variables in wildlife modeling (Behr et al. 2017). However, relatively few robust case studies have leveraged the social and environmental data needed to quantify their influence on species distributions.

The bearded pig, *Sus barbatus*, is an ideal species for assessing the relative contribution of socio-cultural and ecological variables to wildlife distributions. Bearded pigs are sensitive to social factors, such as hunting practices, which can influence their local distribution and behavior (Bennett et al. 2000, Kurz et al. 2021). The bearded pig is also the most favored terrestrial game species for many non-Muslim Indigenous communities in Sabah and Sarawak, accounting for up to 54-97% of their wild meat (Bennett et al. 2000, Chin 2001). Within these communities, the bearded pig is still hunted using both historical approaches (e.g. dogs and spear, snare, nets) and contemporary methods (e.g. on foot with a gun, drive hunt) (Yi and Mohd-Azlan 2020, Kurz et al. 2021). Moreover, the species plays a central role in a variety of Indigenous ceremonial practices and celebrations (Janowski 2014), including gifting of the meat to others, and the pig is also hunted for additional reasons, such as pest control, sport, and sale (Yi and Mohd-Azlan 2020, Kurz et al. 2021). However, among one pool of Indigenous pig hunters, only about a quarter reported hunting bearded pigs for sale, and several hunters felt that hunting for sale was unnecessary or irresponsible (Kurz et al. 2021). In Sabah and Sarawak, the largest Indigenous pig-hunting group makes up roughly 21% and 29% of the total population of each state, respectively (Malaysian Department of Statistics 2011), accounting for a substantial proportion of the population of each state.

In our study area, bearded pig hunting has been a particularly salient cultural force for thousands of years (Harrisson et al. 1961, Kurz et al. 2021). This hunting relationship is given further texture by the role of religion in shaping cultural affinities toward bearded pigs. In Sabah and Sarawak, religion and ethnicity are closely linked. According to census data, 100% of Malays in both states are registered as Muslim, and roughly 75% of Kadazandusun-Muruts (KDM, an abbreviation used locally; KDMs are the primary pig-hunting Indigenous group in Sabah) and Ibans (the primary pig-hunting Indigenous group in Sarawak) identify as Christians (Malaysia Department of Statistics 2011). Wild pork is highly favored by Christian KDM and Iban communities in Sabah and Sarawak (Bennett et al. 2000, Kurz et al. 2021), but actively avoided in Muslim communities in Malaysia generally (Yusof et al. 2012). Therefore, the role of bearded pig hunting in contemporary Malaysian Borneo also speaks to the role of religion in shaping socio-ecological interactions. Given these dynamics as well as the potential threat to pig populations posed by hunting (Bennett et al. 2000, Luskin et al. 2018), it is critical to better understand how ethnicity, and religious factors tied to ethnicity, may be related to hunting pressure and bearded pig distributions in Malaysian Borneo. It is also critical to understand the role of landscape accessibility and population density on bearded pig distributions, as both are likely to influence hunting pressure in the region (Deith & Brodie 2020). Considering the long history of sustained pig hunting in Sabah (Medway 1964), it is possible that hunting practices in some areas may be neutral or positively associated with pig occupancies. However, with modern hunting techniques and extensive land-use change in Sabah and Sarawak, bearded pigs have experienced local declines and behavior change over time, complicating this question (Bennett et al. 2000, Kurz et al. 2021).

Physical ecological factors are also likely to shape the distribution of the bearded pig. The species is dependent on forest habitat for several of its behaviors, such as wallowing, nest building, and mud scraping (Love et al. 2018). Additionally, its natural history is closely linked to Southeast Asian forest phenology due to its local and long-distance movements to track Dipterocarp fruit during mast fruiting events (Curran & Leighton 2000, Luskin & Ke 2018, Granados et al. 2019, Kurz et al. in press). Loss of Dipterocarp forests has not only led to fewer reports of nomadic movements in places (Kurz et al. 2021), but has also led to an estimated 23% decline in bearded pig habitat in Borneo (Ke & Luskin 2019). As forests have declined, fruit provision from Dipterocarps has been replaced in many areas by subsidies from oil palm plantations that fruit throughout most of their lifecycle (Gaveau et al. 2016, Luskin et al. 2017, Love et al. 2018, Luskin et al. 2018). At fine scales, bearded pigs are known to be capable of sustaining populations in heavily logged areas with oil palm fruit subsidies (Love et al. 2018, Davison et al. 2019), but it is unclear how forest and oil palm patchworks are shaping pig distribution at broad scales. As such, understanding the ecological correlates of pig distributions will help plan large landscape configurations that sustain healthy bearded pig populations.

Here, we integrate socio-cultural and ecological covariates to quantify their collective influence on the distribution of a cultural keystone species, the bearded pig. Specifically, we investigate the influence of two Indigenous, predominantly Christian pig-hunting groups on distributions of bearded pigs. By considering this social covariate, alongside a metric of landscape accessibility and population density, we provide a first step toward understanding a broader suite of socio-cultural covariates on wildlife distributions. Selecting Malaysian Borneo as our study area allowed us to investigate socio-ecological dynamics by drawing from extensive camera trap surveys, a published hunting metric, and census data for Sabah and Sarawak. Analyzing data from 295 camera locations distributed across land-use, management, and socio-cultural contexts, we investigate how social and ecological factors together shape bearded pig occupancy. We ran occupancy models in a multi-model approach with top models ranked by AICc. Our results: (a) provide novel empirical associations between socio-ecological factors and the distribution of a large-bodied game species; (b) provide area-specific baseline evidence of bearded pig occupancies before the 2020-2021 African Swine Fever (ASF) outbreak; and (c) quantitatively show that Indigenous hunting in our study area can be compatible with high pig occupancies. In light of these findings, we discuss the implications of our results for context-specific bearded pig management in Sabah and Sarawak.

Results

Our findings show that bearded pig distributions are associated with both social and ecological covariates in Malaysian Borneo. Additionally, bearded pig distributions are significantly negatively associated with one ecological covariate (distance to water) and one social covariate (the interaction between pig-hunting group and hunting accessibility). Pig occupancy showed a positive association with a high level of pig-hunting group for low accessibility areas, and a positive association with a medium to low level of pig-hunting group for high accessibility areas. Additionally, estimated pig occupancies were relatively high across most study sites.

Bearded pig occupancy associations with socio-cultural and ecological factors

All top-ranking occupancy models included both socio-cultural and ecological covariates (Table 1). Pig occupancy was associated with four ecological covariates in competitive models: distance to forest edge, distance to water, slope, and tree cover. Pig occupancy was significantly negatively associated with distance to water in the model average of top models (Figure 1). Occupancy probability was associated with three socio-cultural covariates in competitive models: hunting accessibility, the proportion of the district population composed of the KDM or Iban group (hereafter, “pig-hunting group”), and their interaction (Table 1). Pig occupancy was significantly negatively associated with the interaction between pig-hunting group and hunting accessibility in the model average of top models (Figure 1). Elevation and protected area status were not present in the top models.

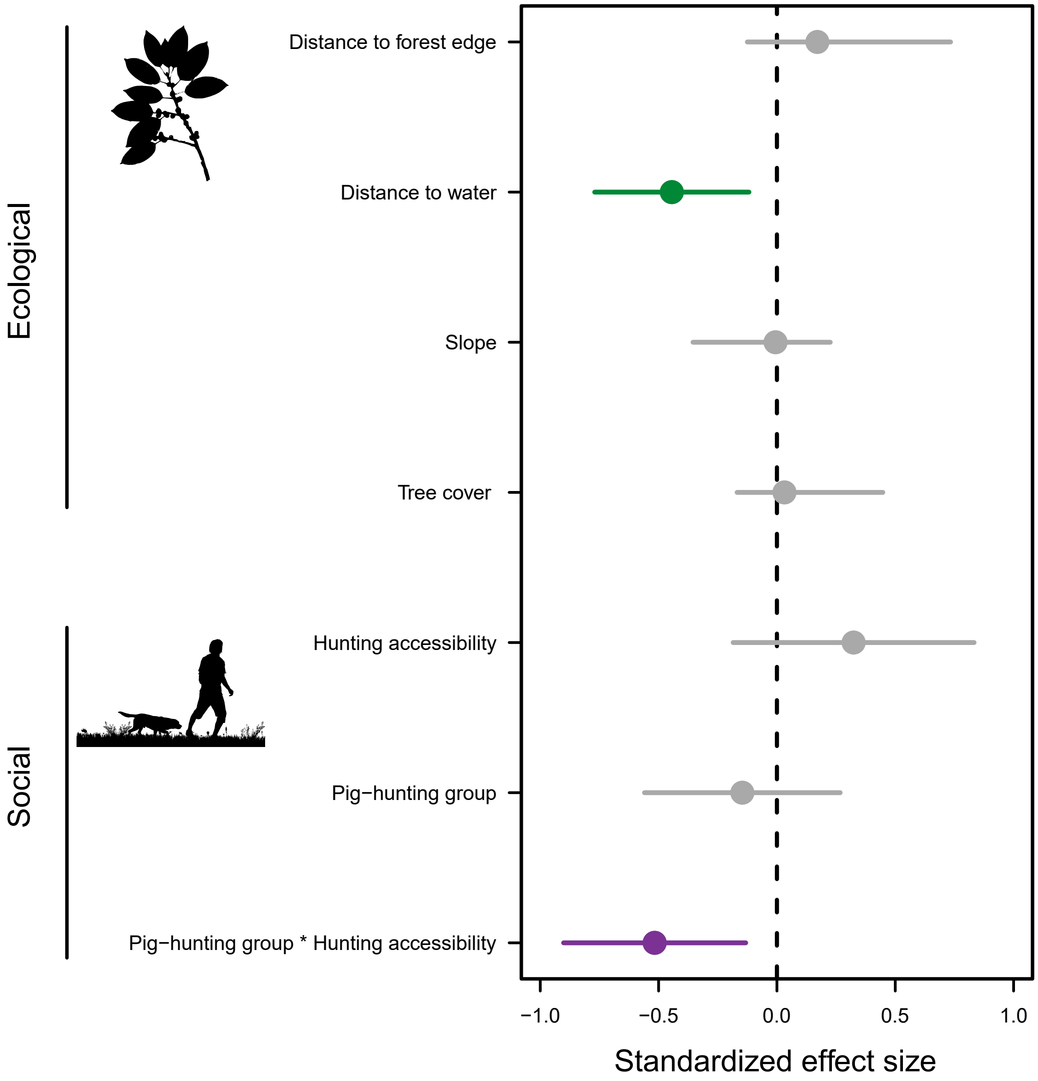

Figure 1. Standardized effect size median values and 95% confidence intervals (CI) for all occupancy covariates in model average of top models with <2 ∆AICc of the top model. Pig occupancy is associated with all covariates in the model-averaged model; however, the influence of covariates with 95% CI that do not overlap with zero can be considered significant.

Table 1. Occupancy models with <2 ΔAICc of the top ranked model; "p" indicates detection probability and "ψ" indicates occupancy probability. "Model" shows the variables present in the model, "W" indicates model weight, "AICc" indicates corrected Akaike Information Criterion, and ΔAICc indicates the difference in the AICc between the model and the top model. [Covariate abbreviations are: Access = Hunting accessibility; dFE = distance to forest edge; dWat = distance to water; eff = sampling effort; NT = non-tree vegetation cover; PHgrp = Pig-hunting group; TC = Tree cover; slp = Slope; PHgrp*Access = Interaction of pig-hunting group and hunting accessibility]

Model	W	AICc	ΔAICc
p(eff) + ψ(dWat) + ψ(dFE) + ψ(PHgrp) + ψ(Access) + ψ(PHgrp*Access)	0.066	4226.8	0.00
p(eff) + ψ(dWat) + ψ(PHgrp) + ψ(Access) + ψ(PHgrp*Access)	0.059	4227.0	0.23
p(eff) + p(NT) + ψ(dWat) + ψ(dFE) + ψ(PHgrp) + ψ(Access) + ψ(PHgrp*Access)	0.035	4228.1	1.30
p(eff) + ψ(dWat) + ψ(PHgrp) + ψ(Access) + ψ(TC) + ψ(PHgrp*Access)	0.033	4228.2	1.40
p(eff) + ψ(dWat) + ψ(dFE) + ψ(PHgrp) + ψ(Access) + ψ(TC) + ψ(PHgrp*Access)	0.031	4228.3	1.50
p(eff) + p(NT) + ψ(dWat) + ψ(PHgrp) + ψ(Access) + ψ(PHgrp*Access)	0.031	4228.3	1.52
p(eff) + ψ(dWat) + ψ(dFE) + ψ(PHgrp) + ψ(Access) + ψ(slp) + ψ(PHgrp*Access)	0.025	4228.7	1.94

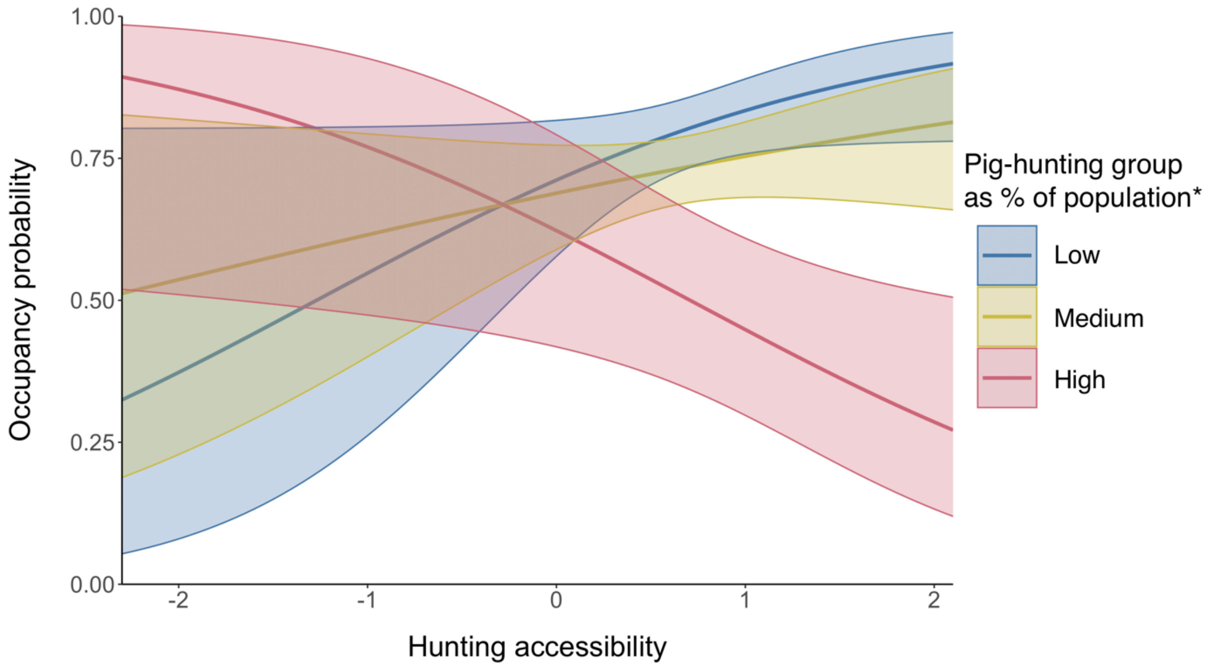

Figure 2. Plot showing a prediction, based on the model-averaged results, of the interaction between pig-hunting group and hunting accessibility, while holding other covariates constant. Bands show 95% confidence intervals. *Pig-hunting group population percentage is relative to the district population level, and percent is standardized.

Table 2. Estimated occupancy probability for bearded pigs at study sites across Sabah and Sarawak, Malaysian Borneo in 2010, 2012, and 2014, prior to the outbreak of African Swine Fever in 2021. Occupancy probability predictions are based on model-averaged results.

Site	Pred. occupancy	95% CI
Crocker Range National Park	0.35	[0.21, 0.52]
Danum Valley Conservation Area	0.92	[0.74, 0.97]
Gunung Mulu National Park	0.51	[0.32, 0.71]
Hose Mountains	0.66	[0.37, 0.86]
Lower Kinabatangan Wildlife Sanctuary	0.85	[0.76, 0.92]
Madai Baturong	0.90	[0.78, 0.96]
Maliau Basin Conservation Area	0.76	[0.64, 0.84]
Malua Forest Reserve	0.77	[0.61, 0.87]
Pulong Tau National Park	0.74	[0.61, 0.83]
SAFE Project	0.73	[0.57, 0.85]
Sapulut	0.37	[0.23, 0.54]
Silabukan	0.80	[0.69, 0.88]
Sipitang	0.77	[0.66, 0.86]
Tabin Wildlife Reserve	0.78	[0.64, 0.87]
Tawau Hills Park	0.79	[0.64, 0.88]
Ulu Baram	0.61	[0.46, 0.74]
Ulu Padas	0.70	[0.60, 0.79]
Ulu Trusan	0.78	[0.68, 0.85]

Interaction between Indigenous pig-hunting group and hunting accessibility

The interaction between pig-hunting group and hunting accessibility significantly negatively influenced pig occupancy in the model-averaged model (Figure 1). At low levels of hunting accessibility, a prediction based on model-averaged results showed that pig occupancy was positively associated with a high level of pig-hunting group (Figure 2). At high levels of hunting accessibility, the prediction based on model-averaged results showed that pig occupancy was positively associated with medium and low levels of pig-hunting group (Figure 2).

Estimated occupancy values and detection associations

Our results show relatively high occupancy estimates of bearded pigs across most sites within our study area in Sabah and Sarawak (Table 2). Predicted average bearded pig occupancy across all study sites from our model-averaged occupancy models was 0.70 (95% CI [0.57, 0.81]). Predicted average bearded pig occupancy per site ranged from 0.92 (95% CI [0.74, 0.97]) at Danum Valley Conservation Area to 0.35 (95% CI [0.21, 0.52]) at Crocker Range National Park (Table 2). Notably, relatively high and low estimated pig occupancies were present in Sabah and

Sarawak, in protected and unprotected areas, in primary and secondary forests, and in hill and montane forests.

Bearded pig detection was associated with two covariates in the model average of top models: sampling effort and non-tree vegetation cover. Detection was significantly positively associated with sampling effort in the model average of competitive models (standardized effect size = 0.296, 95% CI [0.202, 0.389]). Non-tree vegetation cover was present in only two top models (Table 1), and pig detection was not significantly associated with non-tree vegetation cover in the model average of top models (standardized effect size = -0.009, 95% CI [-0.122, 0.045]).

Discussion

Our results provide robust, quantitative evidence that socio-cultural and ecological factors underpin the spatial distribution of a large-bodied game species. By demonstrating the importance of socio-ecological drivers in wildlife distribution modeling for this species, our approach goes beyond many conventional modeling frameworks that consider ecological factors in isolation or that use broad indices of human footprint. Furthermore, we quantitatively show that Indigenous pig hunting is potentially compatible with high bearded pig occupancy in both high and low accessibility areas. Relatively high pig occupancies across many sites, from data collected prior to the ASF outbreak, suggest the potential for sustainable Indigenous pig hunting and sustainable pig populations after bearded pig recovery from ASF.

Influence of hunting accessibility and Indigenous hunting practices

Socio-cultural covariates were predictors of bearded pig occupancy, highlighting the relevance of cultural practices for wildlife distributions. While conceptual models of socio-ecological systems are becoming increasingly common (e.g. Lischka et al. 2018), and urban ecology has embraced human demographics and cultural practices in wildlife distribution assessments (e.g. Alberti et al. 2003, Kumar et al. 2018), we assert that it is critical to normalize more fully the integration of socio-cultural practices into wildlife ecology and conservation. Published accessibility maps (e.g. Weiss et al. 2018), census data, and government-collected social data provide opportunities for such quantitative integration with wildlife data from camera traps, surveys, acoustic data recorders, citizen science datasets, and integrated databases (e.g. Hudson et al. 2017). In our study context, social landscapes and ecological landscapes share important intersections. Deith & Brodie (2020) show that hunter movements are strongly associated with landscape characteristics across Malaysian Borneo. Our results build on this link by showing that the interaction between hunting accessibility and pig-hunting cultural groups is associated with a game species response.

Pig-hunting group and hunting accessibility showed a strong interaction, with important lessons for socio-ecological theory and wildlife management. At low levels of hunting accessibility, a predicted high level of pig-hunting group was positively associated with pig occupancy. At high levels of hunting accessibility, a high level of pig-hunting group was negatively associated with pig occupancy. However, medium and low levels of pig-hunting group were positively associated with pig occupancy in high accessibility areas. The role and relevance of ethnicity and religion, including cultural practices, in empirical and theoretical studies of wildlife distribution has been too often overlooked. Robust cultural traditions have important implications for conservation value formation (Van Houtan 2006), and social practices shape wildlife utilization patterns across the globe (Hunt & Ditton 2002, Amador et al. 2015, Kurz et al. 2021). Our findings validate this theoretical consideration by empirically demonstrating the connections between pig-hunting group, hunting accessibility, and bearded pig occupancies in Malaysian Borneo. On a management level, this finding tracks with other studies highlighting opportunities for integration of Indigenous harvesting practices and sustainable management of game species, such as moose (Popp et al. 2019), primates (Shaffer et al. 2018), and whales (Breton-Honeyman et al. 2021). The positive relationship between Indigenous pig-hunting group and pig occupancy in low-accessibility areas suggests that there continue to be pathways for sustainable Indigenous hunting of bearded pigs. In high-accessibility areas, the positive association between medium and low levels of pig-hunting groups and pig occupancy suggests the need for adaptive, context-specific management that incorporates a certain threshold of hunting as a valid land use (Neumann et al. 2022).

In our study region, ethnic identification is also tightly linked to religion, leading to additional cultural implications for bearded pig hunting. The majority of KDM and Iban communities identify as Christian (although a minority are Muslim), and the vast majority of Malays identify as Muslim (Malaysia Department of Statistics 2011). These tight ethno-religious identifications add an additional layer of group identity to pig hunting practices and dietary choices. In fact, food practices rooted in religion and ethnicity are so strong in our study area that a “pig line” has been recorded in Sarawak between Muslim fishing communities along the coast and Christian pig-hunting communities in the interior (Bolton et al. 1972). Our results show that two particularly Indigenous, Christian cultural hunting traditions—by Iban and KDM groups—shape bearded pig distributions, and may continue to be compatible with bearded pig conservation. While ethno-religious traditions have been linked to harvest of wild animals and plants (e.g., Wadley et al. 1997, Golden & Comaroff 2015, Pieroni & Sõukand 2019), relatively few studies have quantitatively explored these connections. Therefore, by showing the influence of ethnicity, including religious dimensions of ethnicity, on bearded pig occupancy, our findings make a compelling case for the wider relevance of cultural factors on wildlife consumption patterns generally.

The significance of density estimates for bearded pigs

Our study focused on the influence of socio-ecological factors on occupancy, and did not include density estimates within the scope of our analysis. Bearded pigs have a number of population states of varying densities under different environmental conditions (Caldecott et al. 1993), presumably resulting in varying levels of resilience to hunting pressure. In areas with plentiful food resources and high pig densities, bearded pig resilience to hunting could be high; indeed, during resource-rich mast fruiting periods, female bearded pigs can give birth to 10-30 piglets in a single year (Luskin & Ke 2018). In Sabah, older pig hunters recall pig mass movements more than younger hunters (Kurz et al. 2021); this pattern may point to fewer large-scale Dipterocarp mast fruiting events accompanied by bearded pig herding behavior (Caldecott & Caldecott 1985; Caldecott et al. 1993). Additionally, in contemporary Malaysian Borneo, oil palm plantations provide year-round food subsidies to bearded pigs in many areas (Davison et al. 2019, Kurz et al. 2021); for example, in one mixed landscape, bearded pigs were found in 80% of oil palm transects adjacent to forest (Love et al. 2018). Therefore, the relationship between high bearded pig densities and hunting accessibility deserves further study, and could hold important implications for context-specific hunting management.

Baseline benchmarks for bearded pig recovery from African Swine Fever

The outbreak of African Swine Fever (ASF) has led to the collapse of bearded pig populations across Sabah, and is a threat to populations in Sarawak (Kurz et al. in press). ASF case fatality rates of 47.7 to 100% have been recorded in wild and domestic pigs (Liu et al. 2020, FAO 2021). Our site-specific findings—from data collected prior to the ASF outbreak—therefore provide a critical baseline of bearded pig occupancy in a number of areas within Malaysian Borneo. At the time of publication, bearded pig populations are almost certainly much lower than the estimates reflected here, due to widespread local population collapses due to ASF (SE Asian Wildlife Research group, pers. comm.). As populations recover, further research can compare future bearded pig occupancy relative to these baselines for specific areas as part of ongoing monitoring efforts for this IUCN-listed Vulnerable species (Luskin et al. 2018). For example, ongoing camera trap surveys across protected areas could help managers and decision makers assess the utility of hunting policies, movement control orders, law enforcement, and other regulatory mechanisms that could assist in bearded pig recovery. Protecting and expanding existing parks and wildlife reserves may also aid in the distribution of numerous bearded pig sub-populations, providing contexts for pig populations to safely recover and bolstering resilience for future disease-related threats.

Context-specific management of bearded pigs in Malaysian Borneo

Our results support context-specific management of bearded pigs in Sabah and Sarawak, Malaysian Borneo. Average estimated occupancy probability for bearded pigs across all sites was relatively high, at 0.70 (95% CI [0.57, 0.81]). However, spatial variation in predicted occupancy, which varied widely across field sites (0.35 - 0.92), suggests a range of management needs for bearded pig populations, which are known to have widely varying movement patterns and population growth rates that correspond to different population states (Caldecott et al. 1993). Bearded pig natural history fits many characteristics of a generalist species—e.g., dietary flexibility, broad habitat use, and high dispersal ability (Luskin & Ke 2017, Davison et al. 2019), and indeed our results showed high estimated occupancy probabilities in primary, secondary, protected, unprotected, hill, montane, and lowland forests in our study. However, the species still seems to require forest cover for safety, thermoregulation, and nesting (Luskin & Potts 2011, Love et al. 2018, Kurz et al. in press). Widespread habitat destruction in recent decades has led to losses of large portions of its range and habitat in Borneo as well as in Peninsular Malaysia and Sumatra (Ke & Luskin 2019). As such, management strategies that preserve and protect contiguous forest areas are of critical long-term importance.

Moreover, our results show a significant, negative relationship between bearded pig occupancy and distance to water. This finding suggests the importance of proximity to water for high bearded pig occupancies. Our data do not allow us to conclude which aspect of proximity to water is particularly important for bearded pig populations, but there are several potential hypotheses. One possibility is that nearby water availability provides ready access to thermoregulation, as bearded pigs need to wallow regularly to cool themselves in warm tropical conditions (Kurz et al. in press). Another hypothesis is that *Ficus* sp. (fig) trees are more abundant or of greater diversity in riparian areas in our study area, providing a steady food source to sustain pig populations. Figs are thought to be a key food resource for bearded pigs during inter-mast intervals between Dipterocarp fruit pulses (Wong et al. 2005). Additionally, some *Ficus* species are particularly prominent in riparian areas in Southeast Asia (Corner 1969, Pothasin 2014), suggesting that this may hold true for our study areas. Management policies could consider protecting areas with Dipterocarps, *Ficus* sp. trees, *Fagaceae* sp. (oak) trees, and other food sources for bearded pigs, thereby providing variation in fruiting cycles that can sustain bearded pigs through lean periods (Caldecott et al. 1993, Wong et al. 2005). Large protected areas with abundant Dipterocarps may also provide one of the last contexts for conserving the unique migratory ecology of bearded pigs, which are thought to historically have traveled for hundreds of kilometers tracking supraannual mast fruiting bounties (Caldecott & Caldecott 1985). Large tracts of protected forest in Borneo and Sumatra are likely the only places remaining with sufficiently high masting activity (e.g. Granados et al. 2019), intact migration corridors, low hunting risk, and reduced influence of oil palm fruit subsidies – together potentially establishing conditions for long-range bearded pig nomadic movements. In Borneo, continued tri-lateral cooperation between the governments of Brunei, Malaysia, and Indonesia through the Heart of Borneo initiative will be essential to protect large intact forest amidst infrastructure development, logging, mining, and oil palm expansion (Declaration on the Heart of Borneo Initiative 2007, Keong and Onuma 2021).

Finally, our results point to a need to manage bearded pigs to provide for sustainable Indigenous hunting practices and resilient bearded pig populations. Our results indicate that in both high and low hunting accessibility areas, there may be pathways for sustainable bearded pig hunting. In low accessibility areas, we found that pig occupancy was positively associated with Indigenous pig-hunting groups, suggesting that customary Indigenous pig hunting rights can continue to co-exist with sustainable management of pig populations, as they have for thousands of years (Harrisson et al. 1961). In high accessibility areas, medium and low levels of Indigenous pig-hunting groups were positively associated with pig occupancy (Figure 2). Forest-oil palm mixed landscapes, which are generally in high accessibility areas (Gaveau et al. 2016, Deith & Brodie 2020), are known to support high levels of bearded pig utilization (e.g. Love et al. 2018, Davison et al. 2019). Prior to the ASF outbreak, bearded pig hunting was legal and common in both Sabah and Sarawak outside of protected areas and, in the case of Sabah, with a hunting license (Wildlife Conservation Enactment 1997, Wildlife Protection Ordinance 1998, Wilting et al. 2006, Yi and Mohd-Azlan 2020). After bearded pig populations recover to pre-ASF levels, we recommend preserving Indigenous pig hunting rights while also ensuring the protection of pig populations in protected areas. Our results suggest that in low accessibility areas, higher levels of Indigenous hunting may be compatible with high pig occupancies, whereas in high accessibility areas, moderate levels of Indigenous hunting may be compatible with high pig occupancies. In particular, regulated hunting in and around industrial and small-scale oil palm plantations could provide pest control benefits in plantations while also offering limited hunting opportunities for KDM and Iban hunters, providing meat provision and cultural benefits (Yi and Mohd-Azlan 2020, Kurz et al. 2021). Relatively high pig occupancy estimates across most of our study sites and high bearded pig reproductive capacity (Luskin & Ke 2017) suggest that balancing these goals is possible. However, further research on pig densities, up-to-date records of hunting rates, and ongoing monitoring of pig populations are needed before specific regulations could be most accurately developed.

Nuanced solutions will be critical so that Indigenous communities in Sabah and Sarawak can continue to sustainably hunt bearded pig populations, as they have for millennia (Harrisson et al. 1961, Medway 1964). For integrated biocultural conservation goals, it is critical to preserve the cultural and dietary importance of bearded pig hunting for Indigenous groups while also limiting hunting to sustainable levels for bearded pig populations (Kurz et al. 2021). Our results highlight the tensions and opportunities of these twin goals.

Methods

Study region and data collection

We collated bearded pig detections and non-detections from 18 camera trap surveys conducted in 2010, 2012, or 2014 across the Malaysian Bornean states of Sabah and Sarawak (Figure 3). We followed institutional guidelines for research and obtained local research permissions from the Sabah Forestry Department, Forest Department Sarawak, Sabah Biodiversity Centre, and the Sarawak Biodiversity Centre (permit #s: JKM/MBS.1000-2/12(156) and JKM/ MBS.1000-2/2 JLD.3 (18)).

From these surveys, we integrated data from a total of 295 camera trap locations across 18 sites, comprising 10,462 bearded pig detections across 25,755 trap days. Camera surveys were designed to target multiple mammal species and were therefore suitable for detecting bearded pigs, one of the most common mammals in many camera surveys in our study area (e.g. Bernard et al. 2013, Mohd-Azlan et al. 2019). We spatially filtered camera locations to ensure at least 1 km between locations. For studies using paired cameras at each sampling location, we applied a random number generator to randomly select one camera from the pair.

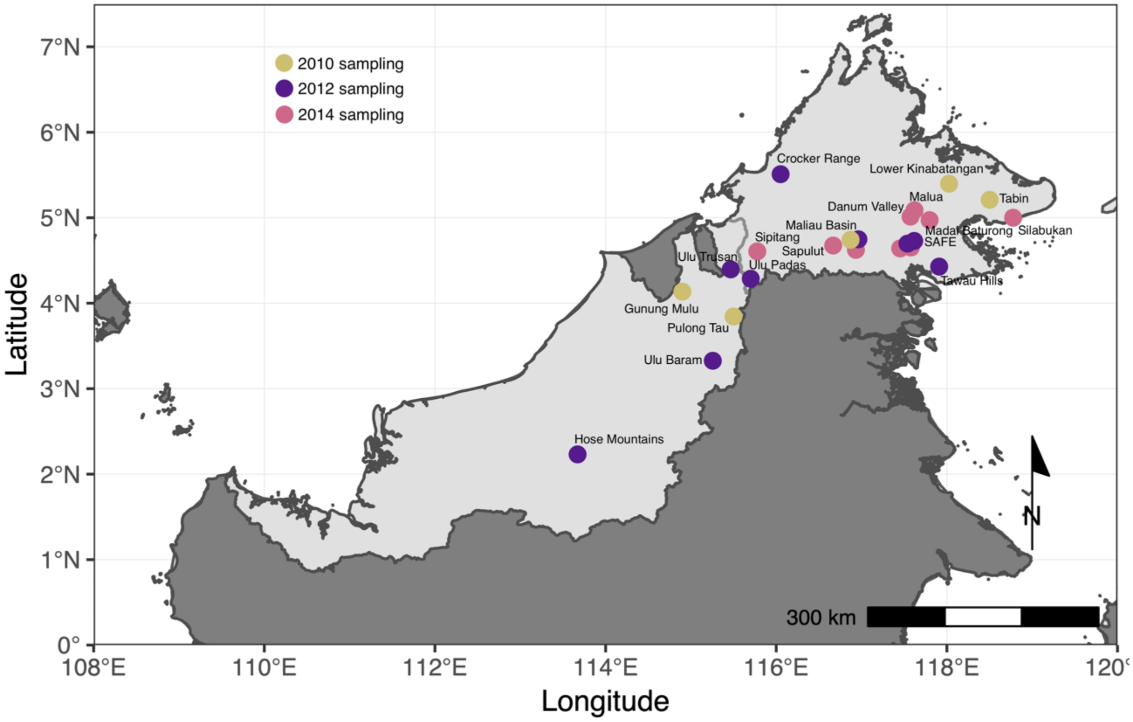

Figure 3. Location of 18 wildlife camera surveys conducted in 2010, 2012, and 2014.

Camera trap surveys were carried out within lowland, hill, and montane primary forests (within protected areas) and previously logged lowland, hill, and montane secondary forests (including both protected and non-protected areas) (Miettinen et al. 2012, Miettinen et al. 2016). Landscape accessibility across the study area varied considerably; our survey data included low and high accessibility areas (Deith & Brodie 2020). Within arrays, cameras were active for varying lengths of time (Supplemental Table 1).

Ecological covariates

We chose covariates that we hypothesized to be strong predictors of bearded pig occupancy (Table 3). Ecological covariates (extracted from Earth Engine Data Catalog) included in analyses were distance to water (Pekel et al. 2016), distance to forest edge (Gaveau et al. 2016), protected area status (UNEP-WCMC and IUCN 2019), elevation (Farr et al. 2007), slope (Farr et al. 2007), and tree cover (Dimiceli et al. 2015) (Table 3, Supplemental Table 2). To match ecological variables with the camera trap data, we either used the closest year of ecological data or an average of multiple years. We removed highly correlated continuous covariates until Pearson's correlation coefficients among all pairs were less than |0.6| (Zuur et al. 2009, R Core Team 2019). To allow for ready comparison among covariate coefficients, we standardized all ecological covariates, except for protected area status (which is a factor).

Table 3. Social and ecological covariates included in occupancy models. KDM is a locally-used abbreviation for the Kadazandusun-Murut Indigenous group.

Model covariate	Hypothesized relationship with occupancy	Covariate type
Hunting accessibility	—	social
KDM or Iban Indigenous group	—	social
Hunting accessibility*Indigenous grp.‡	—	social
Distance to water (m)	—	ecological
Distance to forest edge (m)	+	ecological
Protected area status	+	ecological
Elevation (m)	+	ecological
Slope (deg)	—	ecological
Tree cover (%)	+	ecological

‡ The * represents an interaction effect. An interaction effect occurs when the level of one covariate influences the effect of another covariate on the dependent variable, which in this case is pig occupancy.

Social covariates

We chose social variables that we hypothesized to influence bearded pig distributions (Table 3). The social covariates included in the models were: pig-hunting group (proportion of the district population composed of the predominant Indigenous pig-hunting group), hunting accessibility, and their interacting effects (Table 3, Supplemental Table 2). For the pig-hunting group covariate, we used published census data to calculate the proportion of the district population composed of the KDM peoples in Sabah and the Iban peoples in Sarawak (Bennett et al. 2000, Malaysian Department of Statistics 2011, Kurz et al. 2021).

Our study was not set up to explore hunting pressure directly, e.g. through separate covariates for raw numbers of total hunters per district and physical accessibility of the landscape. Instead, we used a published metric—which we refer to as "hunting accessibility"—that shows landscape accessibility adjusted for coarse estimates of plausible hunter density (Deith & Brodie 2020). A higher metric score indicates higher hunting accessibility, and vice versa (Deith & Brodie 2020). Despite measures of physical accessibility incorporated into the hunting accessibility metric, the metric was not highly correlated with any of our ecological covariates at a level above |0.4| (Zuur et al. 2009). While this metric also accounts for the relative human population in a given area (Deith & Brodie 2020), it is unable to capture cultural influences within the local populations engaging in hunting in that area, which is influenced by ethnicity and religion (Kurz et al. 2021). Alongside hunting accessibility, we included the covariate "pig-hunting group" (see above), thereby incorporating Indigenous, Christian pig-hunting practices into our models. We also included an interaction term for pig-hunting group and hunting accessibility, which tests whether the level of one of these covariates influences the impact of the other on pig occupancy. As with ecological covariates, we standardized social covariates to allow for ready comparison of coefficients.

Detection covariates

We used percent of non-tree vegetation cover and effort as our predictors for detection. Non-tree vegetation cover can block a pig from view and thereby inhibit the ability of a given camera to detect a passing pig. Including effort as a detection covariate in occupancy models allowed us to correct for variable sampling effort, per trap and occasion window. As with ecological and social covariates, we standardized detection covariates to allow for ready comparison of coefficients.

Occupancy modeling approach

We fitted one single-season occupancy model for all the data from 2010, 2012, and 2014 (Davidson 2020). However, given the long-distance movements for which the bearded pig is known (Caldecott & Caldecott 1985, Curran & Leighton 2000, Luskin & Ke 2018), detections of bearded pigs cannot be considered independent and their populations cannot be considered closed. Therefore, model-estimated occupancy results for species in this situation should be interpreted as probability of site use rather than true occupancy (sensu Petracca et al. 2018, Gould et al. 2019).

We ran models using the packages ‘unmarked’ (Fiske & Chandler 2011) and ‘camtrapR’ (Niedballa et al. 2016) in R version 3.6.0 (R Core Team 2019). We used a seven-day sampling occasion window, corrected for sampling effort per trap and per occasion within the ‘camtrapR package’ (Niedballa et al. 2016). We then used the ‘dredge’ function (Bartoń 2009) in R version 3.6.0 to identify top models (R Development Core Team 2019) according to Akaike’s Information Criterion values, adjusted for small sample sizes (AICc). Models within <2 ΔAICc of the top model were considered to be competing models (Ainley et al. 2006). We model-averaged competing models. To visualize the interaction between pig-hunting group and hunting accessibility, we created an interaction plot using packages ‘ggplot2’ and ‘AICcmodavg’ (Wickham and Chang 2014, Mazerolle 2020).

Acknowledgements

We thank Dr. Mar Cabeza and two anonymous reviewers for comments that helped greatly improve the manuscript. Thank you to Owen Bidder and Lindsey Rich for initial advice on the modeling approach, and to Claire Kremen, Jimmy McGuire, the Brashares Lab, Justin Brashares’ Writing Seminar, the UC Berkeley Conservation Community, and the WildCo Lab at UBC for support and feedback at various stages of the project. Thank you also to the Sabah Biodiversity Centre, the Sarawak Biodiversity Centre, the Sabah Forestry Department, and the Forest Department Sarawak for permissions for field research and data collection (permit #s: JKM/MBS.1000-2/12(156) and JKM/ MBS.1000-2/2 JLD.3 (18)). Work at SAFE was supported by the UK Natural Environment Research Council Human Modified Tropical Forest Programme. DJK was supported in this work by the Thomas McKenna Meredith ’48 Postdoctoral Fellowship in Environmental Science at Trinity College, a Philomathia Graduate Student Fellowship in the Environmental Sciences, a Continuing Fellowship through the UC Berkeley College of Natural Resources, the Hannah M. and Frank Schwabacher Memorial Scholarship Fund, the Howard William Siggins Fellowship, and the S.J. Hall Fellowship.

Author Contributions Statement

JSB conceived the study. DJK, TC, and JSB designed the research approach. JFB, AH, PCG, OW, AA, ELB, HB, NJD, JGEW, AG, OH, H-YL, MSL, DM, JM-A, JR, BS, MS, and BG contributed to field work. SHZ led collation of covariates and DJK led collation of camera data. TC led the analysis, and DJK, SHZ, JFB, OW, MD, NJD, AG, MSL, MS, MDP, and JSB contributed to the analysis. DJK led writing, and all authors contributed to writing, review, and approval of the manuscript.

Competing Interests Statement

The authors declare no competing interests.

Data Availability Statement

The census data and environmental variable data used in this study are publicly available (see Supplemental Table 2). Data sharing for the camera trap data is not applicable to this study, as no new camera data were created for this study. Requests for camera data may be made to the authors of the original camera studies for which those data were collected.

References

1. Alberti, M., J.M. Marzluff, E. Shulenberger, G. Bradley, C. Ryan, and C. Zumbrunnen. 2003. Integrating humans into ecology: opportunities and challenges for studying urban ecosystems. *BioScience* 53, 1169-1179.

2. Ainley, D.G., G. Ballard, and K.M. Dugger. 2006. Competition among penguins and cetaceans reveals trophic cascades in the Western Ross Sea, Antarctica. *Ecology* 87(8), 2080-2093.

3. Amador, R., C. Casanova, and P. Lee. 2015. Ethnicity and perceptions of bushmeat hunting inside Lagoas de Cufada Natural Park (LCNP), Guinea-Bissau. *Journal of Primatology* 3, 121.

4. Barber-Meyer, S.M., S.R. Jnawali, J.B. Karki, P. Khanal, S. Lohani, B. Long, D.I. MacKenzie, B. Pandav, N.M.B. Pradhan, R. Shrestha, N. Subedi, G. Thapa, K. Thapa, and E. Wikramanayake. 2012. Influence of prey depletion and human disturbance on tiger occupancy in Nepal. *Journal of Zoology* 289, 10-18.

5. Bartoń, K. 2020. MuMIn: multi-model inference. R package version 1.43.17.

6. Behr, D.M., A. Ozgul, and G. Cozzi. 2017. Combining human acceptance and habitat suitability in a unified socio-ecological suitability model: a case study of the wolf in Switzerland. *Journal of Applied Ecology* 54, 1919-1929.

7. Bennett, E.L., A.J. Nyaoi, and J. Sompud. 2000. Saving Borneo's bacon: the sustainability of hunting in Sarawak and Sabah. In 'Hunting for Sustainability in Tropical Forests'. (Eds. J.G. Robinson and E.L. Bennett.) pp. 305-324. Columbia University Press, New York.

8. Bernard, H., A.H. Ahmad, J. Brodie, A.J. Giordano, M. Lakim, R. Amat, …, and W. Sinun. 2013. Camera-trapping survey of mammals in and around Imbak Canyon Conservation Area in Sabah, Malaysian Borneo. *The Raffles Bulletin of Zoology* 61, 861-870.

9. Bettigole, C.A., T.M. Donovan, R. Manning, J. Austin, and R. Long. 2014. Acceptability of residential development in a regional landscape: potential effects on wildlife occupancy patterns. *Biological Conservation* 169, 401-409.

10. Bolton, J.M., M.R.C.S., L.R.C.P., D.T.M., and H., D. Obst. R.C.O.G. 1972. Food taboos among the Orang Asli in West Malaysia: a potential nutritional hazard. *The American Journal of Clinical Nutrition* 25, 789-799.

11. Breton-Honeyman, K., H.P. Huntington, M. Basterfield, K. Campbell, J. Dicker, T. Gray, A.E.R. Jakobsen, F. Jean-Gagnon, D. Lee, R. Laing, L. Loseto, P. McCarney, J Nokasana Jr., T. Palliser, L. Ruben, C. Tartak, J. Townley, and E. Zdor. 2021. Beluga whale stewardship and collaborative research practices among Indigenous peoples in the Arctic. *Polar Research* 40, 5522.

12. Caldecott, J., and S. Caldecott. 1985. A horde of pork. *New Scientist* 110:32-35.

13. Caldecott, J.O., R.A. Blouch, and A.A. Macdonald. 1993. The bearded pig (*Sus barbatus*). In 'Pigs, Peccaries and Hippos: Status Survey and Conservation Action Plan'. (Ed. W. L. oliver.) pp. 136-145. Gland, IUCN.

14. Chin, C. 2001. Pig in the pot: comments on Sus barbatus in the hunting lifestyle of the Penan in Sarawak (Borneo). *Asian Wild Pig News* 1, 10-12.

15. Corner, E.J.H. 1969. The complex of Ficus deltoidea; a recent invasion of the Sunda shelf. *Philosophical Transactions of the Royal Society of London. Series B, Biological Sciences* 256, 281-317.

16. Curran, L.M. and M. Leighton. 2000. Vertebrate responses to spatiotemporal variation in seed production of mast-fruiting Dipterocarpaceae. *Ecological Monographs* 70, 101-128.

17. Davison, C.W., P.M. Chapman, O.R. Wearn, H. Bernard, and R.M. Ewers. 2019. Shifts in the demographics and behavior of bearded pigs (*Sus barbatus*) across a land-use gradient. *Biotropica* 51, 938-948.

18. Declaration on the Heart of Borneo Initiative. 2007. Brunei Darussalam, Indonesia, and Malaysia. 16 September 2022. https://wwf.panda.org/discover/knowledge_hub/where_we_work/borneo_forests/about_borneo_forests/declaration/

19. Deith, M.C.M, and J.F. Brodie. 2020. Predicting defaunation: accurately mapping bushmeat hunting pressure over large areas. *Proceedings of the Royal Society B* 287, 20192677

20. Dimiceli, C., Carroll, M., Sohlberg, R., Kim, D.H., Kelly, M., Townshend, J.R.G. (2015). MOD44B MODIS/Terra Vegetation Continuous Fields Yearly L3 Global 250m SIN Grid V006 [Data set]. NASA EOSDIS Land Processes DAAC. doi: 10.5067/MODIS/MOD44B.006

21. Farr, T.G., Rosen, P.A., Caro, E., Crippen, R., Duren, R., Hensley, S., Kobrick, M., Paller, M., Rodriguez, E., Roth, L., Seal, D., Shaffer, S., Shimada, J., Umland, J., Werner, M., Oskin, M., Burbank, D., and Alsdorf, D.E., 2007, The shuttle radar topography mission: Reviews of Geophysics, v. 45, no. 2, RG2004, at https://doi.org/10.1029/2005RG000183.

22. Food and Agriculture Organization of the United Nations. 2021. Agriculture and Consumer Protection Department. ASF situation in Asia & Pacific update. 27 May 2021. http://www.fao.org/ag/againfo/programmes/en/empres/ASF/situation_update.html

23. Fiske, I., and R. Chandler. 2011. unmarked: an R package for fitting hierarchical models of wildlife occurrence and abundance. *Journal of Statistical Software* 43, 1–23. http://www.jstatsoft.org/v43/i10/.

24. Francis, C. M. (2008). *A Guide to the Mammals of Southeast Asia.* Princeton: Princeton University Press.

25. Gaveau, D.L.A., M. Salim, S. Arjasakusuma. 2016, "Deforestation and industrial plantations development in Borneo", https://doi.org/10.17528/CIFOR/DATA.00049, Center for International Forestry Research (CIFOR), V2

26. Gaynor, K.M., K.J. Fiorella, G.H. Gregory, D.J. Kurz, K.L. Seto, L.S. Withey, and J.S. Brashares. 2016. War and wildlife: linking armed conflict to conservation. *Frontiers in Ecology and the Environment* 14, 533-542.

27. Golden, C.D., and J. Comaroff. 2015. Effects of social change on wildlife consumption taboos in northeastern Madagascar. *Ecology and Society* 20, 41

28. Gould, M.J., W.R. Gould, J.W. Cain III, and G.W. Roemer. 2019. Validating the performance of occupancy models for estimating habitat use and predicting the distribution of highly-mobile species: a case study using the American black bear. *Biological Conservation* 234, 28-36.

29. Granados, A., H. Bernard, and J.F. Brodie. 2019. The influence of logging on vertebrate responses to mast fruiting. *Journal of Animal Ecology* 88, 892-902.

30. Guerrero, A.M., N.J. Bennett, K.A. Wilson, N. Carter, D. Gill, M. Mills, C.D. Ives, M.J. Selinske, C. Larrosa, S. Bekessy, F.A. Januchowski-Hartley, H. Travers, C.A. Wyborn, and A. Nuno. 2018. Achieving the promise of integration in social-ecological research: a review and prospectus. *Ecology and Society* 23, 38.

31. Harrisson, T., D. A. Hooijer, and L. Medway. 1961. An extinct giant pangolin and associated mammals from Niah Cave, Sarawak. Nature 189:166.

32. Heberlein, T.A., and G. Ericsson. 2006. Ties to the countryside: accounting for urbanites attitudes toward hunting, wolves, and wildlife. *Human Dimensions of Wildlife* 10, 213-227.

33. Hudson, L.N., T. Newbold, S. Contu, S.L.L. Hill, I. Lysenko, A. De Palma…and A. Purvis. The database of the PREDICTS (Projecting Responses of Ecological Diversity In Changing Terrestrial Systems) project. *Ecology and Evolution* 7, 145-188.

34. Hunt, K.M., and R.B. Ditton. 2002. Freshwater fishing participation patterns of racial and ethnic groups in Texas. *North American Journal of Fisheries Management* 22, 52-65.

35. Janowski, M. 2014. Pigs and people in the Kelabit Highlands, Sarawak. *Indonesia and the Malay World* 42, 88-112.

36. Karanth, K.K., J.D. Nichols, J.E. Hines, K.U. Karanth, and N.L. Christensen. 2009. Patterns and determinants of mammal species occurrence in India. *Journal of Applied Ecology* 46, 1189-1200.

37. Kays, R., A.W. Parsons, M.C. Baker, E.L. Kalies, T. Forrester, R. Costello, … and W.J. McShea. 2016. Does hunting or hiking affect wildlife communities in protected areas? *Journal of Applied Ecology* 54, 242-252.

38. Ke, A., and M.S. Luskin. 2019. Integrating disparate occurrence reports to map data-poor species ranges and occupancy: a case study of the Vulnerable bearded pig *Sus barbatus*. Oryx 53, 377-387.

39. Keong, C.Y., and A. Onuma. 2021. Transboundary ecological conservation, environmental value, and environmental sustainability: lessons from the Heart of Borneo. *Sustainability* 13, 9727.

40. Kumar, N., U. Gupta, Y.V. Jhala, Q. Qureshi, A.G. Gosler, and F. Sergio. 2018. Habitat selection by an avian top predator in the tropical megacity of Delhi: human activities and socio-religious practices as prey-facilitating tools. *Urban Ecosystems* 21, 339-349.

41. Kurz, D.J., T.P. Malim, and B. Goossens. In press. Bearded pigs. In: Wildlife Atlas of Sabah. Eds: G. Davies, A. Shabrani, A. Lim, and D. Arnielia.

42. Kurz, D., F.H. Saikim, V.T. Justine, J. Bloem, M. Libassi, M. Luskin, L.S. Withey, B. Goossens, J.S. Brashares, and M.D. Potts. 2021. Transformation and endurance of Indigenous hunting: Kadazandusun-Murut bearded pig hunting practices amidst oil palm expansion and urbanization in Sabah, Malaysia. *People and Nature* https://doi.org/10.1002/pan3.10250

43. Linkie, M., G. Guillera-Arroita, J. Smith, A. Ario, G. Bertagnolio, F. Cheong, … and Zulfahmi. 2013. Cryptic mammals caught on camera: assessing the utility of range wide camera trap data for conserving the endangered Asian tapir. *Biological Conservation* 162, 107-115.

44. Lischka, S.A., T.L. Teel, H.E. Johnson, S.E. Reed, S. Breck, A.D. Carlos, and K.R. Crooks. 2018. A conceptual model for the integration of social and ecological information to understand human-wildlife interactions. *Biological Conservation* 225, 80-87.

45. Liu, J., Liu, B., Shan, B., Wei, S., An, T., Shen, G., and Chen, Z. (2020). Prevalence of African Swine Fever in China, 2018-2019. *Journal of Medical Virology 92,* 1023-1034. https://doi.org/10.1002/jmv.25638

46. Love, K., D.J. Kurz, I.P. Vaughan, A. Ke, L.J. Evans, and B. Goossens. 2018. Bearded pig (*Sus barbatus*) utilisation of a fragmented forest-oil palm landscape in Sabah, Malaysian Borneo. *Wildlife Research* 44, 603-612.

47. Luskin, M.S., J.S. Brashares, K. Ickes, I-F. Sun, C. Fletcher, S.J. Wright, and M.D. Potts. 2017. Cross-boundary subsidy cascades from oil palm degrade distant tropical forests. *Nature Communications* 8:2231.

48. Luskin, M.S., and A. Ke. 2018. Bearded pig *Sus barbatus* (Muller, 1838). In ‘Ecology, Conservation and Management of Wild Pigs and Peccaries’. (Eds. M. Melletti and E. Meijaard.) pp. 175-183. Cambrige University Press, Cambridge.

49. Luskin, M.S., A. Ke, M. Linkie, and E. Meijaard. 2018. Sus barbatus. The IUCN Red List of Threatened Species: e.T41772A10559190.

50. Luskin, M.S., and M.D. Potts. 2011. Microclimate and habitat heterogeneity through the oil palm lifecycle. *Basic and Applied Ecology* 12, 540-551.

51. Malaysia Department of Statistics. 2011. Taburan penduduk dan ciri-ciri asas demografi.

52. Mazerolle, M.J. 2020. Model selection and multimodel inference using the AICcmodavg package. Accessed 26 September 2022. https://mirror.marwan.ma/cran/web/packages/AICcmodavg/vignettes/AICcmodavg.pdf

53. Medway, L. 1964. Post-Pleistocene changes in the mammalian fauna of Borneo: archaeological evidence from the Niah caves. *Studies in Speleology* 1, 33-37.

54. Miettinen, J., C. Shi, W.J. Tan, and S.C. Liew. 2012. 2010 Land cover map of insular Southeast Asia in 250-m spatial resolution. *Remote Sensing Letters* 3, 11-20.

55. Miettinen, J., C. Shi, and S.C. Liew. 2016. 2015 Land cover map of Southeast Asia at 250 m spatial resolution. *Remote Sensing Letters* 7, 701-710.

56. Mohd-Azlan, J., M.C.K. Yi, B. Lip, and J. Hon. 2019. Camera trapping of wildlife in the newly established Baleh National Park, Sarawak. *Journal of Sustainability Science and Management* 14, 38-51.

57. Neumann, W., C. Levers, F. Widemo, N.J. Singh, J.P.G.M. Cromsigt, and T. Kuemmerle. 2022. Hunting as land use: understanding the spatial associations among hunting, agriculture, and forestry. *Ecology and Society* 27, 2.

58. Nguyen, T.V., A. Wilting, J. Niedballa, A. Nguyen, B.M. Rawson, A.Q.H. Nguyen, T.T. Cao, O.R. Wearn, A.C. Dao, and A. Tilker. 2022. Getting the big picture: landscape-scale occupancy patterns of two Annamite endemics among multiple protected areas. *Conservation Science and Practice* 4, e620.

59. Niedballa, J., R. Sollmann, A. Courtiol, and A. Wilting. 2016. camtrapR: an R package for efficient camera trap data management." *Methods in Ecology and Evolution* 7, 1457-1462.

60. Pekel, J.-F., A. Cottam, N. Gorelick, and A.S. Belward. 2016. High-resolution mapping of global surface water and its long-term changes. *Nature* 540, 418-422.

61. Petracca, L.S., J.L. Friar, J.B. Cohen, A.P. Calderón, J. Carazo-Salazar, F. Castañeda, D. Corrales-Gutiérrez, R.J. Foster, B. Harmsen, S. Hernández-Potosme, et al. 2018. Robust inference on large-scale species habitat use with interview data: the status of jaguars outside protected areas in Central America. *Journal of Applied Ecology* 55, 723-734.

62. Pieroni, A., and R. Sõukand. 2019. Ethnic and religious affiliations affect traditional wild plant foraging in Central Azerbaijan. *Genetic Resources and Crop Evolution* 66, 1495-1513.

63. Popp, J.N., P. Priadka, and C. Kozmik. 2019. The rise of moose co-management and integration of Indigenous knowledge. *Human Dimensions of Wildlife* 24, 159-167.

64. Pothasin, P., S.G. Compton, and P. Wangpakapattanawong. 2014. Riparian *Ficus* tree communities: the distribution and abundance of riparian fig trees in Northern Thailand. *PloS ONE* 9: e108945

65. R Development Core Team. 2019. R: A Language and Environment for Statistical Computing. Vienna, Austria: R Foundation for Statistical Computing.

66. Rademaker, M., E. Meijaard, G. Semiadi, S. Blokland, E.W. Neilson, and E.J. Rode-Margono. 2016. First ecological study of the Bawean Warty Pig (*Sus blouchi*), one of the rarest pigs on earth. *PloS one* 11, e0151732

67. Riley, E.P. 2010. The importance of human-macaque folklore for conservation in Lore Lindu National Park, Sulawesi, Indonesia. *Oryx* 44, 235-240.

68. Sadovy de Mitcheson, Y., and B.E. Erisman. 2011. Fishery and biological implications of fishing spawning aggregations, and the social and economic importance of aggregating fishes. In Reef fish spawning aggregations: biology, research, and management. Edited by Y. Sadovy and P.L. Colin. Springer, New York.

69. Semiadi, G., and E. Meijaard. 2006. Declining populations of the Javan warty pig *Sus verrucosus*. *Oryx* 40, 50-56.

70. Shaffer, C.A., M.S. Milstein, P. Suse, E. Marawanaru, C. Yukuma, T.M. Wolf, and D.A. Travis. 2018. Integrating ethnography and hunting sustainability modeling for primate conservation in an Indigenous reserve in Guyana. *International Journal of Primatology* 39, 945-968.

71. Stahlecker, D.W., Z.P. Wallace, D.G. Mikesic, and C.S. Smith. 2017. Does Hopi religious harvest of eaglets affect golden eagle territory occupancy and reproduction on the Navajo Nation? *Journal of Raptor Research* 51, 305-318.

72. Todorov, S.D., and M.L. Chikindas. 2020. Lactic acid bacteria bacteriocins and their impact on human health. In *Lactic Acid Bacteria*. *A Functional Approach*. Taylor & Francis Group, Boca Raton.

73. UNEP-WCMC and IUCN (2019), Protected Planet: The World Database on Protected Areas (WDPA) [On-line], [03/2019], Cambridge, UK: UNEP-WCMC and IUCN Available at: www.protectedplanet.net.

74. Van Houtan, K.S. 2006. Conservation as virtue: a scientific and social process for conservation ethics. *Conservation Biology* 20, 1367-1372.

75. Wadley, R.L., Colfer, C.J.P., & Hood, I.G., 1997. Hunting primates and managing forests: the case of Iban forest farmers in Indonesian Borneo. *Human Ecology*, 25(2), 243-271.

76. Weiss, D.J., A. Nelson, H.S. Gibson, W. Temperley, S. Peedell, A. Lieber…and P.W. Gething. 2018. A global map of travel time to cities to assess inequalities in accessibility in 2015. *Nature* 553, 333-336.

77. Wickham, H., and W. Chang. 2014. Package 'ggplot2'. Accessed 26 September 2022. https://cran.microsoft.com/snapshot/2015-01-06/web/packages/ggplot2/ggplot2.pdf

78. Wildlife Conservation Enactment 1997. No. 6 of 1997. State of Sabah.

79. Wild Life Protection Ordinance, 1998. Chapter 26. Laws of Sarawak.

80. Wilting, A., F. Fischer, S. A. Bakar, and K. E. Linsenmair. 2006. Clouded leopards, the secretive top carnivore of South-East Asian rainforests: their distribution, status and conservation needs in Sabah, Malaysia. *BMC Ecology* 6, 16.

81. Wong, S.T., C. Servheen, L. Ambu, and A. Norhayati. 2005. Impacts of fruit production cycles on Malayan sun bears and bearded pigs in lowland tropical forest of Sabah, Malaysian Borneo. *Journal of Tropical Ecology* 21, 627-639.

82. Wong, J.Z., S. Etoh, and A.B. Sujang. 2009. Towards sustainable community-based fishery resources management: the tagal system of Sabah, Malaysia. *Southeast Asian Fisheries Development Center* 7, 18-23.

83. Yi, M.C.K., and J. Mohd-Azlan. 2020. Wildlife hunting and utilization in Ulu Baleh, Sarawak, Malaysian Borneo. *Ethnobiology Letters* 11, 76-84.

84. Yusof, N.M. 2012. Study of social interaction among students of Vision Schools in Malaysia. *Asian Ethnicity* 13, 47-73.

85. Zuur, A., E. N. Ieno, N. Walker, A. A. Saveliev, and G. M. Smith. 2009. Mixed effects models and extensions in ecology with R. Springer, New York, New York, USA. 574 pp.